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## Five primary sources of organic aerosols in the urban atmosphere of Belgrade (Serbia)



Roberta Zangrando <sup>a,\*</sup>, Elena Barbaro <sup>a,b</sup>, Torben Kirchgeorg <sup>a,b,c</sup>, Marco Vecchiato <sup>a,b</sup>, Elisa Scalabrin <sup>b</sup>, Marta Radaelli <sup>b</sup>, Dragana Đorđević <sup>d</sup>, Carlo Barbante <sup>a</sup>, Andrea Gambaro <sup>a,b</sup>

<sup>a</sup> Institute for the Dynamics of Environmental Processes CNR, Via Torino 155, 30170 Mestre, (VE), Italy

<sup>b</sup> Department of Environmental Sciences, Informatics and Statistics, Ca' Foscari University of Venice, Via Torino 155, 30170 Mestre, (VE), Italy

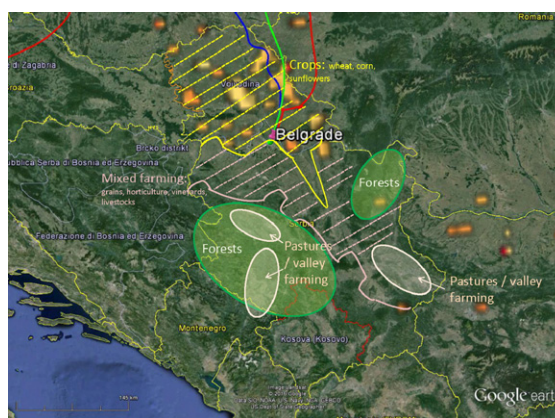
<sup>c</sup> Institute of Sustainable and Environmental Chemistry, Leuphana University of Lüneburg, Lüneburg 21335, Germany

<sup>d</sup> Institute of Chemistry, Technology and Metallurgy – Centre of Chemistry, University of Belgrade, 11158 Belgrade, Serbia

### HIGHLIGHTS

- Belgrade aerosols are affected by five primary sources.
- The increase of the levoglucosan/mannosan ratio characterizes the open fire period.
- The enantiomeric ratio (D/L) value for amino acids is higher in open combustions.
- Tracers from wood and other matrices involved in combustions help discerning events.

### GRAPHICAL ABSTRACT



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### ABSTRACT

Biomass burning and primary biological aerosol particles (PBAPs) represent important primary sources of organic compounds in the atmosphere. These particles and compounds are able to affect climate and human health. In the present work, using HPLC-orbitrapMS, we determined the atmospheric concentrations of molecular markers such as anhydrosugars and phenolic compounds that are specific for biomass burning, as well as the concentrations of sugars, alcohol sugars and D- and L-amino acids (D-AAAs and L-AAAs) for studying PBAPs in Belgrade (Serbia) aerosols collected in September–December 2008. In these samples, high levels of all these biomarkers were observed in October. Relative percentages of vanillic (V), syringic compounds (S) and p-coumaric acid (PA), as well as levoglucosan/mannosan (L/M) ratios, helped us discriminate between open fire events and wood combustion for domestic heating during the winter. L-AAAs and D-AAAs (1% of the total) were observed in Belgrade aerosols mainly in September–October. During open fire events, mean D-AA/L-AA (D/L) ratio values of aspartic acid, threonine, phenylalanine, alanine were significantly higher than mean D/L values of samples unaffected by open fire. High levels of AAs were observed for open biomass burning events.

Thanks to four different statistical approaches, we demonstrated that Belgrade aerosols are affected by five sources: a natural source, a source related to fungi spores and degraded material and three other sources linked

\* Corresponding author.

E-mail address: [roberta.zangrando@idpa.cnr.it](mailto:roberta.zangrando@idpa.cnr.it) (R. Zangrando).

to biomass burning: biomass combustion in open fields, the combustion of grass and agricultural waste and the combustion of biomass in stoves and industrial plants. The approach employed in this work, involving the determination of specific organic tracers and statistical analysis, proved useful to discriminate among different types of biomass burning events.

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## 1. Introduction

Primary sources emit water-soluble organic compounds directly into the atmosphere; the contributions of biomass burning and biological material are among the most relevant ones.

Biomass burning represents the largest source of primary fine carbonaceous matter, the second most important source of gases (Akagi et al., 2011; Calvo et al., 2013) and an important emitter of toxic compounds in the atmosphere (Lemieux et al., 2004; Naeher et al., 2007; Torres-Duque et al., 2008). In addition, biomass burning aerosols influence the climate by affecting the solar balance of the Earth (Hobbs et al., 1997; Carmichael et al., 2009), by influencing snow albedo (Stocker et al., 2013; Flanner et al., 2007; Ramanathan and Carmichael, 2008), and by acting as cloud condensation nuclei (Novakov and Corrigan, 1996; Vestin et al., 2007).

Biomass burning includes both natural and anthropogenic (Akagi et al., 2011; Calvo et al., 2013) sources: open fires in forests, savannas, deforestation, shifting cultivation, and burning of agricultural waste. Although volcanic activities and lightning may cause the ignition of forest fires, humans are generally the starting source of ignition, whether deliberately or accidentally (Taylor, 2010). Moreover, biomass represents an energy source exploited by humans for domestic heating and cooking.

In the atmosphere, sources of primary biological aerosol particles (PBAPs) include biological materials such as microorganisms, fungal spores, bacteria, viruses, pollen, dispersal units, fragments of plants, cells and animals, and excretions from biological organisms (Despres et al., 2012). PBAPs are ubiquitous and are emitted directly into the atmosphere. Bioaerosol can contribute 20–30% of the total atmospheric particulate matter (>0.2 μm) (Chen et al., 2013) and, in the Amazonian region, it can represent 85% of the mass of coarse particles (Poeschl et al., 2010).

PBAPs influence the climatic system by acting as cloud condensation nuclei and as ice nuclei. There are also indications of their importance in the cloud water cycle (Manninen et al., 2014) (and references therein). Bioaerosols are aerodynamically buoyant and can be transported over long distances (Despres et al., 2012), thereby contributing to the diffusion of trace elements and pathogens far from their original location (Yang et al., 2012) (and references therein).

PBAPs also affect humans, animals and vegetation. Pollen and fungal spores are responsible for allergies in humans (Manninen et al., 2014), especially in the spring. The composition and abundance of PBAPs in the atmosphere are related to season, meteorological conditions (temperature and humidity are the most important factors), geographical location and human activities (Manninen et al., 2014; Yang et al., 2012) (and references therein).

In the present work, we determined the atmospheric concentration of organic tracers that are specific of biomass burning: phenolic compounds (syringic acid (SyA), isovanillic acid, homovanillic acid (HA), *p*-coumaric acid (PA), coniferyl aldehyde (CAH), vanillic acid (VA), vanillin (VAN), syringaldehyde (SyAH) and ferulic acid (FA)) (Simoneit, 2002) and anhydrosugars (levoglucosan, mannosan and galactosan). We also determined other compounds related to PBAPs as sugars (arabinose, mannose, xylose, galactose, glucose, fructose and sucrose), alcohol sugars (xylitol, arabitol, ribitol, sorbitol and galactitol, mannitol, glycerol, erythritol, maltitol) (Medeiros et al., 2006; Simoneit et al., 2004) and D- and L-AAAs (L-Ala, L-Asp, L-Asn, L-Arg, L-Glu, L-Phe, L-

Pro, L-Tyr, L-Thr, L-Hys, L-Lys, L-Leu/Ile, L-Orn, L-Ser, L-Gln, L-Val, Gly, D-Ala, D-Asp, D-Phe, D-Ser, and D-Thr) (Ge et al., 2011). Although biomass combustion often produces compounds such as sugars and alcoholsugars, these are also related to PBAPs (Jia et al., 2010b) (and references therein).

Phenolic compounds (PCs) are produced during biomass burning processes based on lignin combustion (Simoneit, 2002). Lignin is a biopolymer that is composed of three different aromatic alcohols: *p*-coumaryl, coniferyl and sinapyl alcohols. Their proportions differ among the major plant classes. The degradation products from the oxidation or burning of lignin are classified as coumaryl, vanillyl and syringyl moieties (Simoneit, 2002). In atmospheric aerosols, PCs may indicate the types of burned plants. Softwoods (gymnosperms) (Oros and Simoneit, 2001a) contain high proportions of coniferyl alcohol and minor proportions of sinapyl alcohol, and wood combustion produces primarily vanillyl moieties. The dominant PCs produced are vanillin, homovanillic acid, vanillic acid, and homovanillyl alcohol. Hardwood (angiosperm) lignin (Oros and Simoneit, 2001b) is enriched in sinapyl alcohol precursors, the combustion of these plants principally produces syringyl and vanillyl moieties. In deciduous tree smoke, the main methoxy phenols produced include homovanillyl alcohol, vanillic acid, vanillin, and syringic acid. In grasses (*gramineae*) (Oros et al., 2006), *p*-coumaryl alcohol is the prevalent lignin unit. Other significant products from burning grasses are acetosyringone, syringic acid, vanillin and vanillic acid.

Sugars can derive from numerous sources. Cellulosic material combustion produces anhydrosugars, which are specific markers of biomass burning (Simoneit, 2002) in the atmosphere. The combustion of hemicellulose generates high quantities of levoglucosan (from cellulose) and lower amounts of mannosan and galactosan.

Primary saccharides are produced from microorganisms, plants, animals, lichens, and bacteria (Dahlman et al., 2003; Simoneit et al., 2004; Yttri et al., 2007), while alcohol sugars come from fungal spores, bacteria, and lower plants (Bauer et al., 2008; Medeiros et al., 2006). Biomass burning is a source of saccharides (Di Filippo et al., 2013; Medeiros and Simoneit, 2008; Pio et al., 2008) due to the breakdown of polysaccharides and to the hydrolysis of anhydrosugars (Pio et al., 2008). It was also proposed as potential origin of fungal spores in urban areas. Sugars from soils and associated biota can be emitted in the atmosphere through re-suspension, erosion and agricultural activities (Jia et al., 2010a; Medeiros and Simoneit, 2007; Simoneit et al., 2004). Glucose, fructose and sucrose may derive from plant pollen and developing leaves (Fu et al., 2012; Graham et al., 2003; Pashynska et al., 2002). Arabitol and mannitol are described as molecular markers for fungal spores (Bauer et al., 2008; Burshtein et al., 2011; Di Filippo et al., 2013; Elbert et al., 2007) reflecting the contribution of microbially degraded material during the leaf senescence period and the fungal reproduction season (Medeiros et al., 2006; Pashynska et al., 2002). Their concentrations can be enhanced by biomass burning (Di Filippo et al., 2013).

AAAs are an important class of compounds originating from biological, terrestrial and marine organisms (Ge et al., 2011). In general, L-AAAs are associated to animals, terrestrial plants and phytoplankton (Cowie and Hedges, 1992), and are used as primary production molecular markers. Gly is one of the most stable AAAs (McGregor and Anastasio, 2001). In the atmosphere, Gly suggests a more aged aerosol (Samy et al., 2013) correlated to long-range transport (Barbaro et al.,

2011; Samy et al., 2013). D-AAs are used as markers of bacterial material both in terrestrial and marine environments (bacterioplankton) (Friedman, 2010; Kaiser and Benner, 2008; McCarthy et al., 1998). Enantiomeric D/L ratios are used as indicators of bacterial origin. Indeed, peptidoglycan in microbial cellular walls contains a high proportion of D-Ala, D-Asp, D-Glu and D-Ser (Dittmar et al., 2001; McCarthy et al., 1998). Due to their low volatility, AAs have been observed in condensed phases such as particulate matter (Barbaro et al., 2011; Barbaro et al., 2015; Di Filippo et al., 2014; Matsumoto and Uematsu, 2005; Scalabrin et al., 2012), dew (Scheller, 2001), rain (Mace et al., 2003b; Mace et al., 2003c), fog (Zhang and Anastasio, 2003), microlayer (Kuznetsova et al., 2004; Matrai et al., 2008), lakes (Barbaro et al., 2014) and marine waters (Kuznetsova et al., 2004; Matrai et al., 2008; Sommerville and Preston, 2001).

In this study we developed two new analytical methods to determine anhydrosugars and PCs using HPLC-orbitrap MS. The aim of this work was to better understand how primary sources, such as biomass burning and PBAPs, affect the chemical composition of particulate matter during the seasonal transition from late summer to early winter in an urban environment. We performed this task by determining organic biomarkers. Moreover, we performed a source apportionment in order to assess the contribution of the studied primary sources and to highlight possible differences in the origins of the contributors to the particulate matter. We performed this study in Belgrade, an urban site where information on water-soluble organic compounds are scarce despite the importance of a source such as biomass burning (Glavonjic, 2011; Glavonjic and Oblak, 2012).

## 2. Experimental

### 2.1. Aerosol sampling

Because the aim of this work was to study biomass burning, PBAPs sources and their changes during the transition between last summer and winter, aerosols samples were collected between September and December 2008 in the urban area of Belgrade, Serbia (44°49'14"N, 20°27'44"E) in order to cover the seasonal change. The sampling was performed over a 24 h period, every 6 days, from 1 September to 30 November. From 1 to 12 December, the sampling was performed every 12 h. In order to compare December samples to those collected from September to November, the compound concentrations in the two samples collected each day were added together considering a sampling time of 24 h.

The temperature ranged between 3 °C (10 December) and 28 °C (7 September), with an average temperature of about 13 °C. The average pressure was 1016.18 hPa and the average humidity was 67.9%. The prevailing winds blew from NW, with an average wind speed of 3.5 m s<sup>-1</sup>. Daily meteorological data are in Table S3.

Total suspended particles (TSP) were collected on a pre-combusted (4 h at 400 °C in a muffle furnace) circular quartz fiber filter (QFF) (SKC Inc., Eighty Four, To-13 model) using a TE 5000 High Volume Air Sampler (Tisch Environmental Inc., OH, USA). In order to check for possible contamination, blank samples were collected by placing filters in the holder when the instrument was turned off. Samples and field blanks were stored at -20 °C in aluminum foil envelopes until analysis. The list of materials used in this work is reported in the supplementary information.

### 2.2. Sample processing

All sample preparations were performed under a class-1000 clean bench to avoid contamination. The pre-analytical protocol for the determination of anhydrosugars, PCs, AAs and sugars was performed as previously described (Perrone et al., 2012). Briefly, a 3/8 portion of each QFF was broken up into small pieces and placed in a 50 mL conical flask using steel tweezers, and the internal standard solution was spiked

to the filter. The internal standard solution contained 70 ng of <sup>13</sup>C<sub>6</sub> vanillin, 1500 ng of <sup>13</sup>C<sub>6</sub> levoglucosan and 200–300 ng of <sup>13</sup>C AA mixture (Ala\*, Arg\*, and Glu\*). The filter was ultrasonically extracted (2 cycles of 15 min with 5 mL and 2 mL of ultrapure water) and filtered through a 0.45 µm PTFE filter (Phenomenex) before instrumental analysis.

### 2.3. Instrumental methods

Anhydrosugars were determined using HPAEC-(–)-APCI-OrbitrapMS; PCs, using HPLC-(–)-ESI-OrbitrapMS; AAs, using HPLC-(+)-ESI-orbitrapMS; and mono-, di-saccharides and alcohol sugars were determined using HPAEC-(–)-ESI-MS. All the details regarding instrumental methods and quality control are reported in the supplementary information.

### 2.4. Back-trajectory calculation

Back-trajectories were calculated using the Hybrid Single Particle Lagrangian Integrated Trajectory (HYSPLIT) transport and dispersion model (Draxler and Rolph, 2012). All back-trajectories (24 h) used GDAS Meteorological Data and vertical velocity model and were calculated at 200, 500, 1000 m AGL. Moreover, in order to evaluate long-range transport back-trajectories, 120 h were calculated at 1000, 1500 and 2000 m AGL for each sample.

## 3. Results

### 3.1. Phenolic compounds

A limited number of papers report PC concentrations in the atmosphere, usually in urban environments and during smoke events. Mean concentrations for SyAH, SyA, VA and VAN were 4.4, 1.8, 1.3 and 1.3 ng m<sup>-3</sup> respectively. These results are similar to those reported in the literature (Table 1 and references therein). During the sampling period, the major compounds determined were VA (11%), VAN (11%), SyA (16%), SyAH (38%) and CAH (10%). PA, FA, HA and IVA accounted for 14%. PA (2%) was observed only in fall samples (Table S4).

### 3.2. Anhydrosugars, mono- disaccharides and alcohol sugars

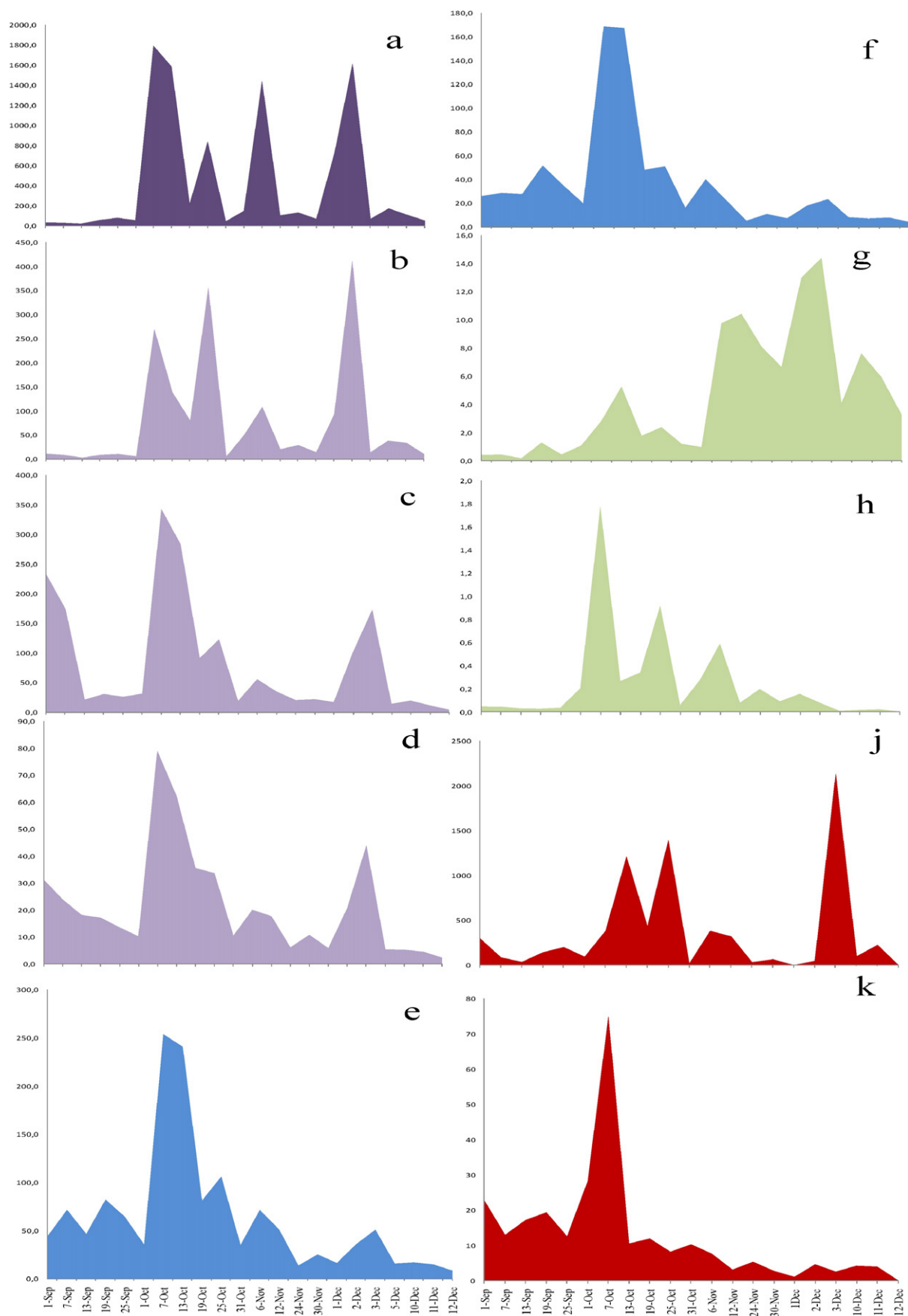
In Belgrade, anhydrosugars in urban aerosols (Table S4) represented a mean percentage of 44% of the total sugar concentration. From 1 September to 12 December, levoglucosan was at a mean concentration of 424.9 ng m<sup>-3</sup> (range 18.8–1791.4 ng m<sup>-3</sup>). Mannosan and galactosan were at 56.3 ng m<sup>-3</sup> (9.6–275.1 ng m<sup>-3</sup>) and 25.3 ng m<sup>-3</sup> (1.9–109.6 ng m<sup>-3</sup>). Levoglucosan represented 84% of the total anhydrosugar concentration, mannosan and galactosan 11% and 5%. Mean levels of levoglucosan were comparable to those reported in the literature and ranging, in European cities, between 40 and 2300 ng m<sup>-3</sup> (Table 1 and references therein).

The atmospheric concentrations of mono-, disaccharides and alcohol sugars reported in previous studies of urban environments (Table 1 and references therein) are similar to those determined in Belgrade (Table S4). During the sampling period, the mean concentration of glucose was 21.6 ng m<sup>-3</sup>, that of fructose 83.6 ng m<sup>-3</sup>, that of galactose 8.6 ng m<sup>-3</sup>, and that of sucrose 10.8 ng m<sup>-3</sup>. Among monosaccharides and disaccharides, the prevailing compounds were fructose (9% of total sugar concentration), xylose (8.3%), arabinose (3.4%) and glucose (2.3%). Galactose, mannose and sucrose represented 0.9%, 0.7% and 1.2%, respectively.

Considering polyols (Table S4), the most important compounds were arabitol (6.7% total mean percentage), mannitol (3.9%) and glycerol (7.1%). Xylitol, ribitol, sorbitol + galactitol, erythritol and maltitol represented collectively 2%. In Belgrade the mean concentration of arabitol in fall–winter was of 62.5 ng m<sup>-3</sup> and mannitol was 35.9 ng m<sup>-3</sup>, while

**Table 1**  
Comparison between atmospheric concentrations ( $\text{ng m}^{-3}$ ) of vanillic acid (VA), syringic acid (SyA), vanillin (VAN), syringaldehyde (SyAH), levoglucosan, glucose, galactose, fructose sucrose, arabitol and mannitol reported in literature and in this study.

Compound	Site	Season	Concentration ( $\text{ng m}^{-3}$ )	Ref
<b>SyAH</b>	Dettenhausen (D), urban	Fall–winter	15.19	(Bari et al., 2009)
	Libby (MT), urban	Winter 2003–2004	$5.8 \pm 2.7$	(Ward et al., 2011)
	Denver (CO), urban	Yearly	1.2	(Simpson et al., 2005)
	Seattle (WA), urban	Oct–Jan	0.09–0.27	(Simpson et al., 2005)
	Missoula (MT), urban	Smoke events	0.23–4.03	(Ward et al., 2006)
	Rondonia (BR), pasture	October	25.2	(Graham et al., 2002)
	Rondonia (BR), forest	October	3.1	(Graham et al., 2002)
	Belgrade (SRB)	Fall–winter	4.4	This study
	Libby (MT), urban	Winter 2003–2004	$1.4 \pm 0.4$	(Ward et al., 2011)
	Libby (MT), urban	Winter 2008–2009	$0.5 \pm 0.5$	(Ward et al., 2011)
<b>SyA</b>	Belgrade (SRB)	Fall–winter	1.8	This study
	Dettenhausen (D), urban	Fall–winter	0.33	(Bari et al., 2009)
	Denver (CO), urban	Yearly	2.44	(Simpson et al., 2005)
	Seattle (WA), urban	Oct–Jan	0.18–0.29	(Simpson et al., 2005)
<b>VAN</b>	Missoula (MT), urban	Smoke events	2.27–18.9	(Ward et al., 2006)
	Rondonia (BR), pasture	October	3.2	(Graham et al., 2002)
	Rondonia (BR), forest	October	1.2	(Graham et al., 2002)
	Belgrade (SRB), urban	Fall–winter	1.3	This study
<b>VA</b>	Libby (MT)	Winter 2003–2004	$55.5 \pm 12.4$	(Ward et al., 2011)
	Libby (MT)	Winter 2008–2009	$68.4 \pm 36.5$	(Ward et al., 2011)
	Belgrade (SRB)	Fall–winter	1.3	This study
	Oporto (P), urban	Winter	120	(van Drooge et al., 2014)
	Copenhagen (DK), urban	Winter	40	(van Drooge et al., 2014)
	Amsterdam (NL), urban	Winter	360	(van Drooge et al., 2014)
	Athens (G)	Winter	45	(van Drooge et al., 2014)
	Vienna (A), urban	Annual	120–160	(Caseiro et al., 2009)
	Graz (A), urban	Annual	250–480	(Caseiro et al., 2009)
	Salzburg (A), urban	Annual	150–220	(Caseiro et al., 2009)
<b>Levoglucosan</b>	Slapanile (CZ), urban	Winter	572	(Krumal et al., 2010)
	Brno (CZ), urban	Winter	326	(Krumal et al., 2010)
	Oslo (N), urban	November–December	166	(Yttri et al., 2005)
	Turin (I), urban	Fall–winter	680	(Piazzalunga et al., 2013)
	Milan (I), urban	Fall–winter	390–1210	(Piazzalunga et al., 2011)
	Florence (I)	Winter	371	(Giannoni et al., 2012)
	Belgrade (SRB), urban	Fall–winter	424.9	This study
	Bologna (I)	Fall	6.9	(Pietrogrande et al., 2014)
	Bologna (I)	Winter	5.6	(Pietrogrande et al., 2014)
	Oslo (N), urban	Fall	47	(Yttri et al., 2007)
<b>Glucose</b>	Elverum (N), suburban	Winter	22	(Yttri et al., 2007)
	Ghent (B)	Winter	73	(Pashynska et al., 2002)
	Belgrade (SRB), urban	Fall–winter	21.6	This study
	Bologna (I)	Fall	2.8	(Pietrogrande et al., 2014)
<b>Galactose</b>	Bologna (I)	Winter	3.0	(Pietrogrande et al., 2014)
	Belgrade (SRB), urban	Fall–winter	8.6	This study
<b>Fructose</b>	Oslo (N), urban	Fall	42	(Yttri et al., 2007)
	Elverum (N), suburban	Winter	11	(Yttri et al., 2007)
	Ghent (B)	Winter	37	(Pashynska et al., 2002)
	Belgrade (SRB), urban	Fall–winter	83.6	This study
<b>Sucrose</b>	Bologna (I)	Fall	4.6	(Pietrogrande et al., 2014)
	Bologna (I)	Winter	3.1	(Pietrogrande et al., 2014)
	Oslo (N), urban	Fall	150	(Yttri et al., 2007)
	Elverum (N), suburban	Winter	95	(Yttri et al., 2007)
<b>Arabitol</b>	Ghent (B)	Winter	48	(Pashynska et al., 2002)
	Belgrade (SRB), urban	Fall–winter	10.8	This study
	Oslo (N), urban	Fall	5.3	(Yttri et al., 2007)
	Elverum (N), suburban	Winter	5.3	(Yttri et al., 2007)
	Rome (I), urban	Fall	4.89	(Di Filippo et al., 2013)
	Rome (I), urban	Winter	20.83	(Di Filippo et al., 2013)
	Bologna (I), urban	Fall	3.4	(Pietrogrande et al., 2014)
	Bologna (I), urban	Winter	3.83	(Pietrogrande et al., 2014)
	Ghent (B), urban	Winter	26	(Pashynska et al., 2002)
	Vienna (A), urban	Fall	7–63	(Bauer et al., 2008)
<b>Mannitol</b>	Rehovot (IL), urban	Fall	18.90	(Burshtein et al., 2011)
	Rehovot (IL), urban	Winter	8.43	(Burshtein et al., 2011)
	Belgrade (SRB), urban	Fall–winter	62.5	This study
	Oslo (N), urban	Fall	8.1	(Yttri et al., 2007)
	Elverum (N), suburban	Winter	4.2	(Yttri et al., 2007)
	Rome (I), urban	Fall	22.4	(Di Filippo et al., 2013)
	Rome (I), urban	Winter	13.47	(Di Filippo et al., 2013)
	Ghent (B), urban	Winter	26	(Pashynska et al., 2002)
	Vienna (A), urban	Fall	8.9–83.0	(Bauer et al., 2008)
	Rehovot (IL), urban	Fall	49.15	(Burshtein et al., 2011)
Rehovot (IL), urban	Winter	21.94	(Burshtein et al., 2011)	
Belgrade (SRB), urban	Fall–winter	35.9	This study	



**Fig. 1.** Temporal trends of levoglucosan (a), xylose (b), fructose (c), glucose (d), arabitol (e), mannitol (f), syringaldehyde (g), p-coumaric acid (h), concentrations ( $\text{ng m}^{-3}$ ). Glycine (j) and L-arginine (k), concentrations ( $\text{pmol m}^{-3}$ ), in Belgrade atmosphere.

**Table 2**

Mean concentrations determined in September, October, November and December 2008 for PCs, total anhydrosugar concentration and levoglucosan, mannosan, galactosan, mono-saccharides, disaccharides, alcohol sugars ( $\text{ng m}^{-3}$ ). In brackets, the percentage of class of sugars with respect to total sugar concentrations.

	September	October	November	December
<b>PCs</b>	1.9	10.1	17.6	17.7
<b>Sugars total</b>	355.6	1656.5	772.1	769.2
<b>Anhydrosugars</b>	65.1 (18.3%)	875.1 (52.8%)	521.0 (67.5)	457.5 (62.5)
Levoglucosan	42.0	753.7	454.9	399.3
Mannosan	18.1	83.7	41.7	68.5
Galactosan	5.4	37.7	24.4	29.7
<b>Monosaccharides</b>	140.3 (39.5%)	418.2 (25.2%)	123.4 (16.0%)	186.2 (23.4%)
<b>Disaccharides</b>	12.5 (3.5%)	21.4 (1.3%)	4.6 (0.6%)	5.8 (0.7%)
<b>Alcohol sugars</b>	140.2 (39.4%)	345.2 (26.8%)	123.0 (15.9%)	106.7 (13.4%)

the literature data range between  $3.4$  and  $63 \text{ ng m}^{-3}$  and  $4.2$  and  $83 \text{ ng m}^{-3}$  (Table 1 and references therein).

### 3.3. D-L amino acids

In the particulate matter sampled in Belgrade, the total mean concentration of free amino acids (FAAs) was  $564.2 \text{ pmol m}^{-3}$  ( $69.4 \text{ ng m}^{-3}$ ) for L-AAAs and  $35.7 \text{ pmol m}^{-3}$  ( $3.5 \text{ ng m}^{-3}$ ) for D-AAAs, while the total mean concentration of (L + D)-AAAs was  $599.9 \text{ pmol m}^{-3}$  ( $72.9 \text{ ng m}^{-3}$ ). In urban environments, the literature reports atmospheric concentrations of L-FAAs ranging between  $334$  and  $2470 \text{ pmol m}^{-3}$  (Barbaro et al., 2011; Di Filippo et al., 2014; Yang et al., 2005; Zhang and Anastasio, 2003). In marine and marine remote aerosols, lower levels ranging from  $1.5$  to  $648 \text{ pmol m}^{-3}$  were observed (Barbaro et al., 2015; Mace et al., 2003b; Mace et al., 2003c; Matsumoto and Uematsu, 2005; Wedyan and Preston, 2008). At a Polar coastal site in the Arctic (Ny Alesund, Svalbard Islands) (Scalabrin et al., 2012),

during the summer, FAAs ranged from  $0.59$  to  $4.7 \text{ pmol m}^{-3}$ . On a coastal site in Antarctica (Mario Zucchelli Station, Victoria Land), the mean level was  $11 \text{ pmol m}^{-3}$ . On the Antarctic plateau, the mean levels were  $0.8 \text{ pmol m}^{-3}$  in 2011–2012 and  $0.7 \text{ pmol m}^{-3}$  in 2012–2013 (Barbaro et al., 2015).

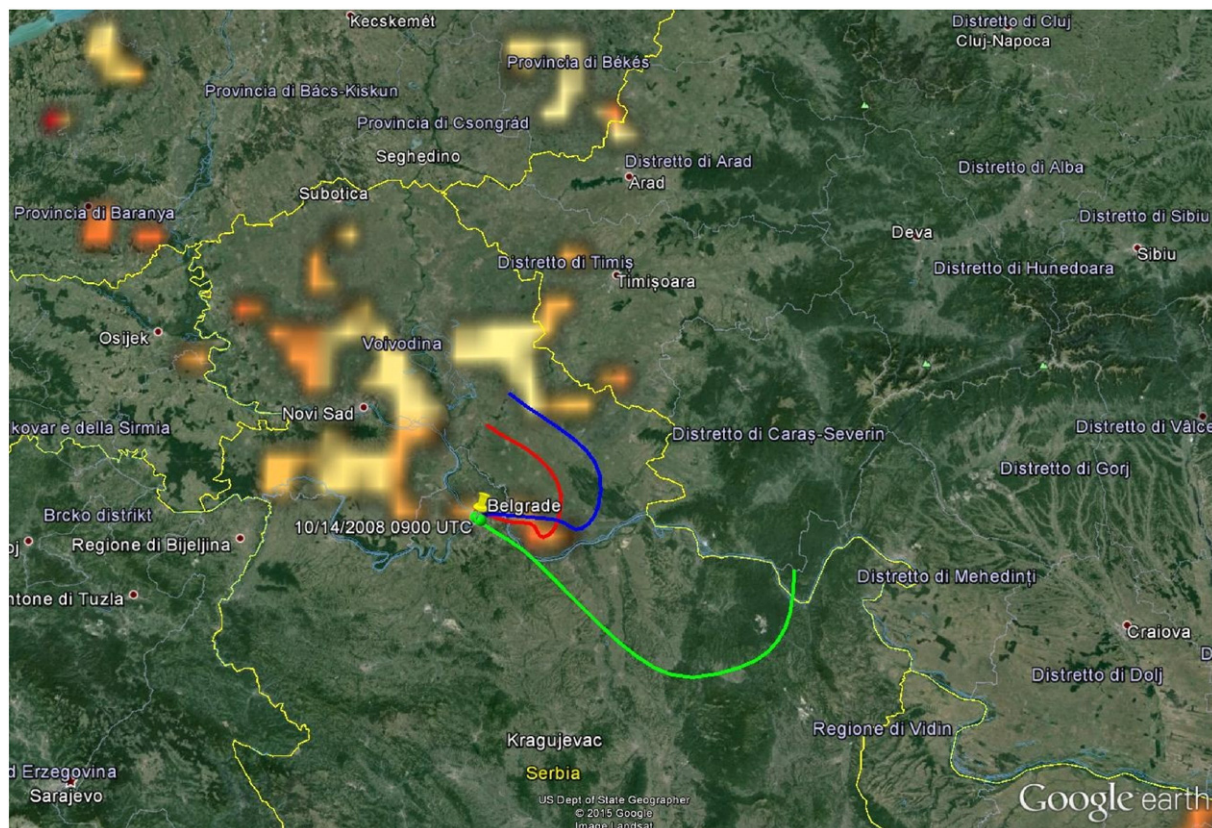
In Belgrade aerosols, between September and December (Table S4), the most important AA was Gly, which represented 46.7% of total the FAAs, followed by L-Ala (8.6%), L-Glu (7.1%), L-Arg 5.9%, L-Val (5.5%) and L-Asp (3%). All D-AAAs represented <1% of the total. These results are in good agreement with the observations of other authors, according to whom Gly is the most abundant AA followed by lower percentages of L-Ser, L-Ala, L-Asp, L-Arg, and L-Glu in both urban (Barbaro et al., 2011; Di Filippo et al., 2014; Zhang and Anastasio, 2003) and marine aerosols (Barbaro et al., 2015; Mace et al., 2003c; Mandalakis et al., 2011; Matsumoto and Uematsu, 2005; Scalabrin et al., 2012).

## 4. Discussion

The temporal trends studied in Belgrade's atmosphere (Table S4) are characterized as follows: higher concentrations are registered for the samples of 7, 13, 25 October, 12 November and 3 December for anhydrosugars (Figs. 1 and S3), mono- and disaccharides, (Fig. S4), alcohol sugars (Figs. 1, S5 and S6) and AAs (Figs. 1, S7\_S13) in comparison to the levels registered in September, and to the other samples collected in November and December.

Relevant concentrations of PCs were observed in October, although the highest ones were observed in December. PC concentrations increased from fall to winter. (Table 2, S4, Figs. 1, S14).

Seasonal trends for anhydrosugars (Fig. S3, Table 2) show lower values in September. In October, a sudden increase in concentrations of biomass combustion tracers levoglucosan, mannosan and galactosan was observed due to open fire events. The overlap of HYSPLIT (Draxler



**Fig. 2.** MODIS satellite image (NEO Active Fires) of 13 October 2008 and HYSPLIT 24 h back trajectories 200 m (red), 500 m (blue), 1000 m (green).

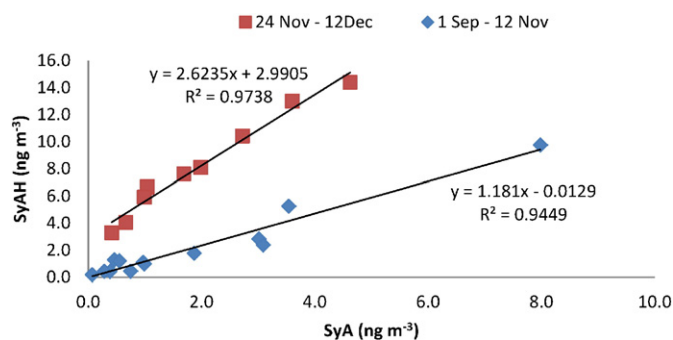


Fig. 3. Plot reporting SyA versus SyAH. Blue diamonds represented fall samples while red squares winter samples.

and Rolph, 2012) 24 h back trajectories and Active Fires from NEO (NASA) shows that the air masses that reached Belgrade on 7, 13, 19, 25 October had passed above fire areas (see Figs. 2 and S15). We considered the possibility these high concentrations could be from long-range transport. 120 h back-trajectories did not show intense fires along the air masses journey (Figs. S16–S18) except for fires in the North of Serbia, demonstrating that these high levels were caused by local events. These observations agree with Vojkan & Bojana (Vojkan and Bojana, 2013), who regularly observed open fire phenomena in Serbia in the month of October every year from 2000 to 2013, due to the burning of agricultural waste.

In November and December, similar mean concentrations of levoglucosan, mannosan and galactosan were registered due to the exploitation of wood for energy production in households (Glavonjic, 2011), industry, and commercial and public facilities (Glavonjic and Oblak, 2012).

These events also affected the atmospheric concentrations of sugars (Figs. S4–S6). In the September samples (Table 2), collected at the end of the growing season, the mean total sugar concentration was  $355.6 \text{ ng m}^{-3}$ ; in October, due to the increase of agricultural waste combustion, there was a general increase in atmospheric sugar concentrations to a total mean of  $1656.5 \text{ ng m}^{-3}$ , consisting principally of anhydrosugars. Because biomass burning is a source of sugars (Medeiros et al., 2006), monosaccharides, disaccharides and alcohol sugars increased as well. In November and December, the sugar levels decreased, but remained higher than in September. The same occurred for monosaccharide disaccharide and alcohol sugar levels. However, due to residential wood burning (Glavonjic, 2011), anhydrosugars contributed to most of the total sugar concentrations detected in the Belgrade aerosols in November at a percentage of 67.5% and December of 62.5% (Table 2).

Among the most abundant sugars detected, the monthly mean percentages of glucose, fructose and sucrose decreased progressively from September to December (Tables 2, S4). This data is in agreement with Medeiros et al. (Medeiros et al., 2006), who observed higher concentrations in the spring and lower ones in the summer, as the growing season progresses. Sucrose is relevant only in the spring as it is particularly

important for the development of flowers buds (Medeiros et al., 2006). Arabinose and galactose are produced from the combustion of leaves and needles, which are rich in pectin (Medeiros et al., 2006). The highest concentrations of these compounds were observed in the samples collected in October (Table S4).

From September to December, the mean concentrations of arabitol and mannitol decreased from  $61.5 \text{ ng m}^{-3}$  to  $22.5 \text{ ng m}^{-3}$  and from  $33.6 \text{ ng m}^{-3}$  to  $10.6 \text{ ng m}^{-3}$ , respectively. The highest concentrations were registered in October at  $124.9 \text{ ng m}^{-3}$  and  $78.4 \text{ ng m}^{-3}$ , respectively. These values may be due to high fungal growth and open fire events in this season.

Different kinds of biomass burning events, such as wildfire, burning of agricultural waste and the exploitation of wood as an energetic source, can contribute to the total aerosol load in the atmosphere. A first indication of different biomass burning sources can derive from the observation of the temporal trends of PCs (Figs. 1 and S14). The PC trend shows a sudden increase of atmospheric concentrations of SyAH, CAH and SyA on 12 November. It also shows a marked difference between the samples collected from September to 12 November and those collected from 24 November to 12 December. This is highlighted in Fig. 3, where SyA was plotted against SyAH, showing a good separation between fall and winter concentrations.

In order to better understand the different kinds of combustion sources during the sampling, we studied the sampling period while considering the monthly mean percentage of PCs and the L/M ratio.

Among the PCs, we considered V (sum of VA and VAN), S (SyA + SyAH) and PA as indicators of the type of wood combusted, as these compounds are respectively present in softwood, hardwood and grasses (Kuo et al., 2011; Opsahl and Benner, 1998). In September and October, the percentages of V (41% and 36%) and S (56% and 58%) were similar, while PA increased from 3% to 6%, due to a surge of grass fires. In November, PA contributed 3%, indicating residual open fires; V decreased to 29%; and S increased to 68%, probably due to the start of domestic heating that uses firewood as fuel. In December, we observed a further decrease in V (23.5%) and an increase in S to 76%, indicating the use of hardwood (Oros and Simoneit, 2001b) in house heating.

In previous studies, L/M ratios were used as indicators of the kind of biomass burned, although these ratios are quite variable (Fabbri et al., 2009). In the smoke L/M values of 2.5 were observed for briquettes, of 3.6–3.9 for softwood and of 14.4–14.8 for hardwood (Schmidl et al., 2008b). Regarding grass combustion, L/M ratios ranged between 2 and 33 (Oros et al., 2006); for leaves combustion, the mean value was 5.5 (Schmidl et al., 2008a); and for crop combustion, the mean values were 32.6 (12.7–55.7) (Sang et al., 2013). In atmospheric aerosols, L/M mean values ranged between 3.5 and 75 (Fabbri et al., 2009). In Belgrade aerosols, the mean L/M ratio was 2.3 (range 1.5–3.0) in September. In October, due to a bonfire increase, the ratios increased to 7.6 (range 2.5–16.7). In November, residual open fires and domestic heating resulted in a mean value of 8.0 (range 3.5–18.6). In December, a value of 5.0 (3.4–8.2) was observed.

In this study V, S and PA proved useful to indicate the type of wood fuel involved in the combustion, while the increase of L/M ratio distinguished the period in which the open fires occurred.

Table 3

D/L AA ratios reported in literature and determined in Belgrade aerosols.

	Ala	Ser	Glu	Asp	Thr	Phe	Sampling site	Reference
Aerosol	0.35	0.11					Atlantic Ocean	Wedyan and Preston, 2008
Aerosol	0.06–0.16	0.02–0.13	0.06–0.07	0.09–0.15			Mediterranean Sea	Kuznetsova et al., 2005
Aerosol	0.04–0.07	0.02–0.05	0.01–0.20	0.03–0.12			Long Island Sound	Kuznetsova et al., 2005
Rainwater	0.61	0.21	0.26	0.34			Seoul (urban)	Yan et al., 2015
Rainwater	0.31	0.09	0.11	0.18			Uljin (coastal rural / marine)	Yan et al., 2015
Aerosol	0.38	0.18		0.08	0.72	0.42	Belgrade Mean sampling campaign	this work
Aerosol	0.88			0.23	1.42	1.57	Belgrade Open fire samples	this work
Aerosol	0.23			0.03	0.49	0.07	Belgrade No open fire samples	this work

The impact of biomass burning on aerosol concentrations in the atmosphere of Belgrade can be evaluated using the glucose/levoglucosan (Samburova et al., 2013) ratio, where a value of  $\approx 4.5$  indicates smoke-free samples, and 0.9 indicates samples affected by smoke. This ratio was calculated by considering that the mean concentrations of glucose and levoglucosan reported for atmospheric urban aerosols in fall–winter ranged between 0.02 and 0.2 (Pashynska et al., 2002; Pietrogrande et al., 2014; Yttri et al., 2007). Throughout the sampling campaign in Belgrade, a comparable value of 0.02 (range 0.98–0.01) was calculated.

The temporal trends of AAs (Figs. 1, S7–S13) show higher levels in the first part of September. An abrupt increase in October and early November occurred in coincidence with bonfires (7, 13, 19, 25 October, 12 November). A decrease was observed in November and December, except for Gly, which showed high levels also on 3 December (Fig. 1). The literature reports a possible connection between high FAA concentrations and biomass burning (Mace et al., 2003a; Samy et al., 2013) and our data seem to support the same connection for open fire.

Very few studies report levels of D-AA in atmospheric aerosols. Wedyan & Preston (Wedyan and Preston, 2008) observed a strong enantiomeric excess of L-AAs in marine aerosols, indicating a relevant proportion of fresh rather than microbially degraded material. In addition, they reported the presence of D-AAs, among which D-Ala was the most abundant AA. The mean D/L ratio for hydrolyzable AAs was 0.35 for Ala and 0.11 for Ser. Kuznetsova et al. (Kuznetsova et al., 2005) studied the presence of total AAs in the Mediterranean Sea and in Long Island Sound. They reported the presence of D-AAs, with a D/L ratio for Asp, Glu, Ser, Ala generally below 0.2 (see Table 3). Yan et al. (Yan et al., 2015) discussed the presence of D-AAs in rainwater samples collected in Seoul (urban site) and Uljin (coastal rural site with marine input). The comparison of mean D/L ratios for Asp, Glu, Ser, Ala (Table 3) between the urban and rural sites showed higher values for Seoul than for Uljin, suggesting that the former site presents a higher level of microbial degradation and that D/L ratios can be used as indicators of such degradation.

In Belgrade, we detected D-Ala in all the samples (Fig. S11) from September to November and in a few samples in December. D-Asp, D-Thr, D-Phe and D-Ser were observed sporadically, except for samples from 7, 13, 19, 25 October and 12 November, which were affected from open fires (Table S4). The mean D/L ratios from September to December for Asp, Thr, Phe, Ser and Ala were similar to those reported in the literature (Table 3). Interestingly, the Student *t*-test proved that the mean values of the D/L ratio for samples affected by open fire (7, 13, 19, 25 October and 12 November) were significantly higher than the mean values calculated for the samples unaffected by fire: for Asp D/L, the mean value for samples affected by fire was 0.23, and for unaffected samples 0.03,  $p = 0.00083$ , Thr 1.42, 0.49,  $p = 0.028$ , Phe 1.57, 0.07,  $p = 0.0027$  and Ala 0.88, 0.23,  $p = 0.00068$ . These higher values observed in samples affected by fire can be explained by the atmospheric input during combustion-related soil re-suspension. Soils include bacterial material (Amelung et al., 2006; Simoneit et al., 2004; Vranova et al., 2012) and degraded material in vegetal waste.

#### 4.1. Primary sources of particulate matter in the Belgrade atmosphere

In order to identify the different sources contributing to the total composition of the samples, different statistical approaches were applied to these data. In the matrix of data obtained from Table S4, the compounds that were mainly below detection limit (BDL) were excluded from statistical analysis. In the samples where variables BLD were <50%, the value corresponding to half the limit of detection was substituted to BDL. The auto-scaled data matrix was used in statistical analysis.

Hierarchical Cluster Analysis (CA) performed using Statistica 8.0 (Ward's method, Squared Euclidean Distance) produced a tree diagram whose variables were divided in two macro-clusters A and B. In macro-cluster A (Fig. 4) the compounds related to biomass burning: phenolic

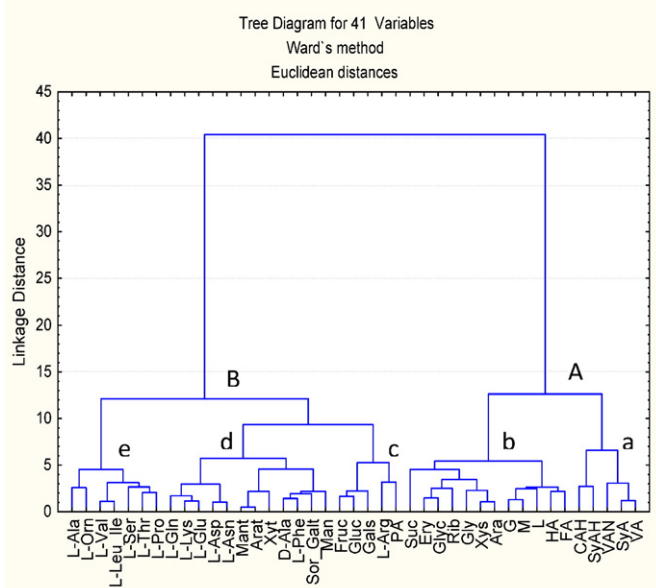


Fig. 4. Result of Hierarchical Cluster Analysis.

compounds (lignin combustion) (cluster (a)). In the cluster (b), two sub-clusters are present the first contains levoglucosan, mannosan and galactosan (cellulose and hemicellulose combustion) HA and FA,

Table 4  
Factor loadings in factor analysis.

	Factor 1	Factor 2	Factor 3	Factor 4	Factor 5
VA	−0.172	0.585	0.142	0.612	0.136
VAN	−0.128	0.301	0.525	0.697	0.068
SyA	−0.189	0.497	0.160	0.682	0.146
SyAH	−0.300	0.332	−0.140	0.833	−0.111
PA	0.234	0.332	0.846	0.072	0.145
FA	−0.053	0.822	0.039	0.484	−0.015
HA	−0.148	0.860	0.041	0.394	0.079
CAH	−0.187	0.157	−0.164	0.716	−0.234
L	0.102	0.675	0.364	0.467	0.385
M	−0.016	0.769	−0.027	0.352	0.386
G	−0.045	0.856	0.073	0.386	0.302
Ara	0.090	0.933	0.238	0.113	0.191
Man	0.106	0.502	−0.017	−0.094	0.835
Xys	0.043	0.935	0.295	0.143	−0.038
Gals	0.354	0.647	0.521	0.081	0.228
Gluc	0.481	0.550	0.429	0.044	0.476
Fruc	0.726	0.451	0.225	0.108	0.412
Suc	0.202	0.788	−0.305	−0.294	0.141
Xyt	0.444	0.259	0.333	−0.100	0.704
Arat	0.336	0.349	0.499	−0.111	0.698
Rib	0.189	0.687	0.325	0.044	0.485
Sor_Galt	0.357	0.343	0.026	0.036	0.833
Mant	0.323	0.286	0.484	−0.079	0.742
Glyc	−0.020	0.899	0.103	0.096	0.385
Ery	0.133	0.788	0.289	0.079	0.504
L-Asn	0.666	0.100	−0.072	−0.050	0.727
L-Asp	0.582	0.181	−0.032	−0.052	0.759
L-Glu	0.644	0.096	0.168	−0.085	0.723
L-Lys	0.782	0.026	0.253	−0.059	0.548
L-Pro	0.947	−0.152	−0.113	−0.085	0.099
L-Thr	0.889	−0.011	0.116	−0.045	0.391
L-Arg	0.524	−0.088	0.764	−0.100	0.141
L-Leu_Ile	0.875	0.006	0.105	−0.262	0.245
Gly	0.008	0.961	−0.111	−0.008	0.127
L-Orn	0.678	0.045	0.306	−0.316	0.368
L-Phe	0.464	0.154	0.065	0.072	0.842
L-Ser	0.833	0.045	0.287	−0.252	0.142
L-Ala	0.533	0.225	0.140	−0.521	0.374
L-Gln	0.624	0.024	0.373	−0.235	0.589
L-Val	0.845	−0.050	0.145	−0.225	0.342
D-Ala	0.374	0.395	0.094	0.037	0.804



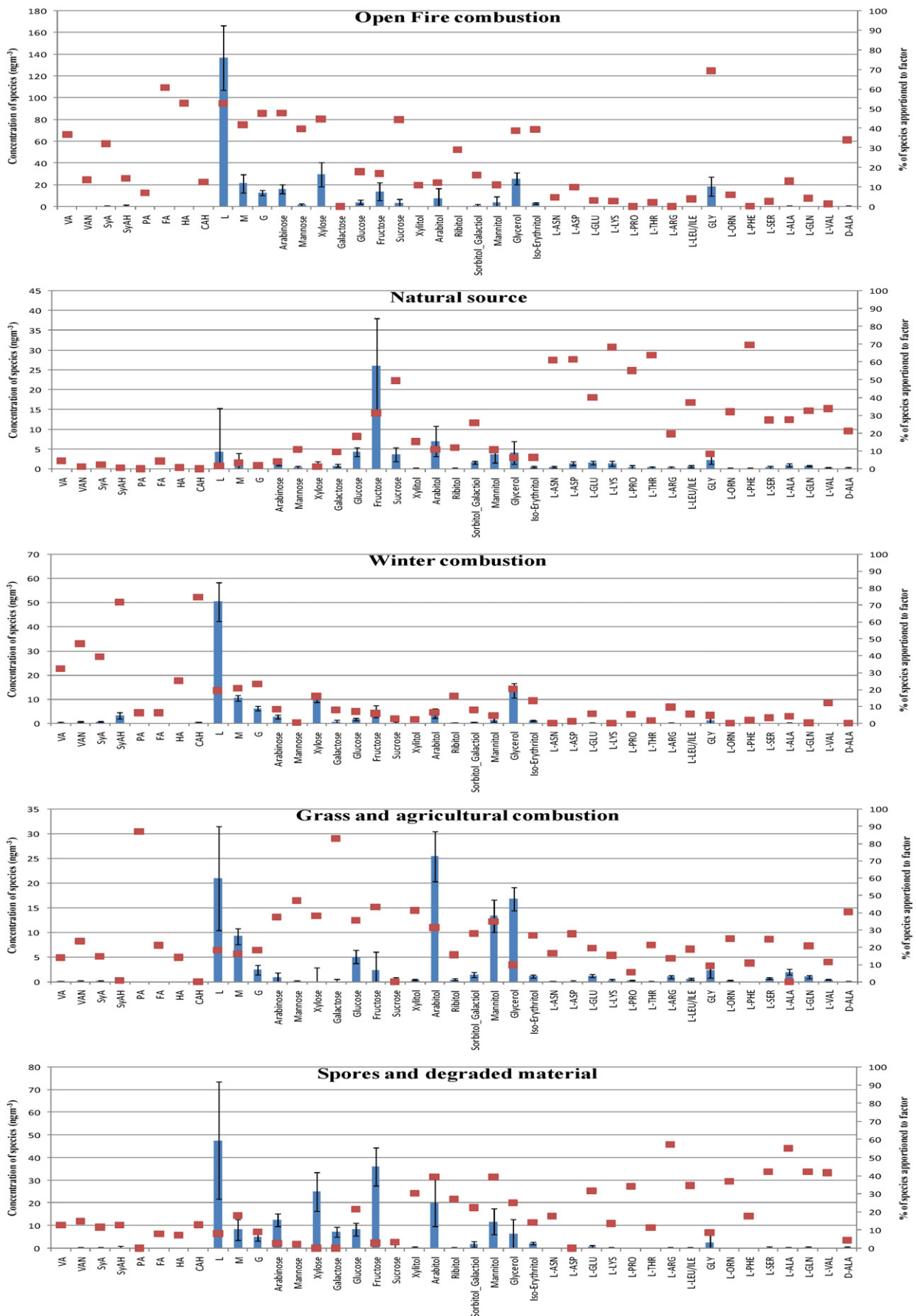


Fig. 5. Source profiles obtained by PMF analysis: contribution of each species to the chemical profile composition of each source ( $\text{ng m}^{-3}$ , blue bars) and percentage contribution of each source to the concentration of each species (% red squares).

while the second includes sugars and sugar alcohol from vegetal combustion. The second macro-cluster (B) is separated in two clusters, the first cluster (c) contain sugars such as glucose, fructose and galactose (vegetal material), L-Arg (an AA linked to urea cycle (Zhang and Anastasio, 2003)) and PA (grass combustion (Oros et al., 2006)). The second cluster (d) consists of mannitol and arabitol (fungal spore (Bauer et al., 2008)) and D-Ala (from bacteria). The second macro-cluster contains also L-AAs as markers of biological fresh material (e).

PCA and factor analysis (Fca) aim to reduce the dimensionality of a set of data using different approaches (Joliffe, 2002). PCA and Fca provide different insights and can be informative when jointly applied to the same data set. PCA and Fca were performed using Statistica 8.0 and the five factors considered explained 91.4% of the cumulative variance. PCA showed high negative factor loadings for sugars and amino acids in PC1 (attributable to a “natural source”), while PC2 presented high positive values for anhydrosugars and phenolic compounds indicating a generic “biomass burning” source for this component (data not shown).

For Fca, varimax rotation was applied. High values of factor loadings (Table 4) were shown in Factor 1 for AAs (“natural” factor); in Factor 2 factor loadings were high for compounds from biomass combustion such as anhydrosugars, PCs, sugars as glycerol and erythritol and Gly (open fire combustion factor). In Factor 3 high factor loadings were for PA and L-Arg we can therefore infer that this Factor is linked to the combustion of grass and biological material (grass combustion). Factor 4 presents high loading values for PCs such as VA, VAN, SyAH, SyA and CAH, indicating a different kind of combustion typical of the winter period (winter combustion factor); Factor 5 presents high values of compounds connected to degraded organic material affected by fungi (mannitol and arabitol, sugar alcohol), bacteria (D-Ala), and other L-AAs, indicating the combustion of degraded material.

In order to obtain more detailed information from our samples, we tried to apply a Positive Matrix Factorization (PMF) (Paatero, 1997; Paatero and Tapper, 1994), although our data set consisted of only 19 samples. This statistical technique has been widely applied in literature with the aim to assess the sources that contribute to atmospheric aerosol composition. Handling PMF data required discriminating between “bad”, “weak” and “strong” variables according to signal-to-noise ratio criteria (Paatero and Hopke, 2003). We set all the variables “strong”. The choice of the solution was based on the results obtained by the model from 3 to 10 factors. The resulting parameters obtained from the scaled residual matrix, IM (the maximum individual column mean), and IS (the maximum individual column standard deviation), together with Q values, were examined to identify the most reasonable solution (Lee et al., 1999; Viana et al., 2008). For Q values, the solution to the system was the point where the slope of the curve showed a marked change. When the number of factors increased to a critical value, the IM and IS parameters experienced a very marked drop. In the present work, the Q, IM and IS values showed a reasonable solution in correspondence with 5 factors. Considering the total concentration of all the compounds that we determined and used in the PMF elaboration, the linear regression between the real value determined experimentally and those calculated from the model showed a  $R^2$  of 0.9668.

The results obtained by PMF are similar to those obtained by Fca. In PMF factor profiles (Fig. 5), high percentages of levoglucosan, mannosan and galactosan are reported for factors F1, F3, and F4, indicating three factors connected to different kinds of biomass burning: PMF F1 presented the highest percentages of levoglucosan, mannosan, galactosan, VA, FA, HA, glycerol, erythritol and Gly (this last compound is related to long-range transport), good percentages of methoxy phenols (SyA, SyAH), and a relevant presence of D-Ala. This factor is correlated to Factor 2 in Fca (Pearson  $r = 0.91$ ) with higher contribution in October individuating an open fire source. PMF F2 showed high percentages of sucrose and L-AAs, showing a correlation ( $r = 0.71$ ) to Factor 1 in Fca. PMF F2 can be related to a “natural” source. PMF F3 showed the highest contribution of VAN, SyA, SyAH, CAH and a good percentage of

levoglucosan, mannosan and galactosan. This factor is correlated to Factor 4 in Fca ( $r = 0.85$ ), with higher contribution in winter samples (winter combustion) when firewood is burned for house heating. In PMF F4 the most relevant feature is the highest percentage of PA attributable to grass combustion; galactose can be attributed to the combustion of pectin present in leaves and needles (Medeiros et al., 2006); sugars (glucose, fructose and sucrose) are connected to vegetation; mannose to soil resuspension (Simoneit et al., 2004) and D-Ala (bacteria). Levoglucosan, mannosan and galactosan (tracers of biomass burning) are present in PMF F4 in good percentages; so are mannitol and arabitol, which are linked to vegetation and mature leaves (Pashynska et al., 2002) as well as to fungi spores. We can consider PMF F4 to be primarily associated to the combustion of grass and agricultural waste. This factor contributes remarkably to the sample collected on 7 October (Table 5). PMF F5 is characterized by high percentages of mannitol and arabitol markers of fungi spore; L-AAs, especially L-Orn and L-Arg, are linked to animal origins such as livestock (Zhang and Anastasio, 2003); the good percentage of alcohol sugars indicates a relation to fungal spores and degraded material.

PMF contributions showed high values of F2 (natural factor) for the samples collected in the summer on 1 and 7 September (Table 5). The sample from 7 October shows a high contribution of F4 (grass combustion). The 13 October sample presented high values for F1 (open combustion) with a relevant contribution of F2 (natural). The different contributions to these two samples are drawn from temporal trends (Figs. S1–S13) and they are confirmed by statistical elaboration. An additional explanation comes from HYSPLIT back-trajectories, which demonstrate the different geographical origins of the air masses. On 7 October, the air masses came from Vojvodina, while on the 13 October they came from the South and South-East (Figs. S15 and 2).

For the samples collected on 19 and 25 October, the contribution came from open-combustion and grass-combustion factors (F1 and F4); on 19 October, the contribution came also from degraded material. The 12 November sample was affected mainly by open fire combustion (F1) and winter combustion (F3). In general, all the samples collected from 12 November to December indicated a relevant contribution of winter combustion (F3). The sample collected on 3 December, in particular, presented a relevant contribution also from open combustion (F1).

## 5. Conclusions

In this work we applied two new methods. A HPAEC(–)-APCI-OrbitrapMS for the determination of levoglucosan, mannosan and galactosan and a HPLC(–)-ESI-OrbitrapMS for PCs in particulate matter. We also applied two additional, previously validated methods for the determination of sugars, alcohol sugars and D and L-AAs using these compounds as tracers of primary sources in Belgrade aerosols collected from September to December 2008 during the seasonal transition.

This study contributes to improving our currently scarce knowledge about the atmospheric concentration of water-soluble organic compounds in the Belgrade area.

The sampling campaign was characterized by the presence of different kinds of biomass burning sources. Particularly high values of biomarkers were observed in October in conjunction with the frequent wildfire and bonfire events that affected the Belgrade area. The temporal trends of PCs, and in particular the high atmospheric concentration of SyAH and CAH in November–December, indicated a different biomass combustion source, probably linked to domestic heating.

The indications from temporal trends were confirmed by studying the relative percentage abundance of vanillic V (VA + VAN) and syringic S (SyA + SyAH) compounds and of p-coumaric acid (PA) in September, October, November and December. These parameters identified the presence of grass combustion in the fall and the employment of hardwood in domestic heating in December. The L/M ratios calculated during the sampling period highlight open fires in October and

**Table 5**  
Factor contribution in PMF analysis ( $\text{ng m}^{-3}$ ).

	F1 Open fire combustion	F2 Natural contribution	F3 Winter combustion	F4 Grass and agricultural combustion	F5 Fungi spores and degraded material
1- Sep	0	514.3	17.4	47.5	0
7- Sep	0	269.2	0	86.1	136.2
13- Sep	0	52.2	0	61.6	151.7
19- Sep	0	0	0	108.4	320.4
25- Sep	71.7	18.7	0	63.4	203.0
1-Oct	55.5	74.8	15.4	0	141.7
7- Oct	0	0	19.0	1756.3	461.6
13- Oct	1399.4	426.8	26.8	374.8	5.4
19- Oct	346.2	0	0	387.3	189.9
25- Oct	1280.9	0	0	506.2	103.7
31- Oct	0	20.1	28.5	46.28	132.4
6- Nov	228.0	37.3	0	227.0	157.3
12- Nov	629.9	4.5	253.7	0	106.7
24- Nov	0	27.1	301.9	0	5.5
30- Nov	20.9	0	232.2	90.1	35.8
2- Dec	0	0	571.4	230.0	0
3- Dec	1969.6	0	368.5	0	0
10- Dec	61.8	4.6	318.4	9.5	6.9
11- Dec	108.3	14.9	178.6	0	35.8

November. The simultaneous application of these parameters proved useful for the discrimination of different biomass burning events.

The detected L-AAAs were mainly L-Ala, L-Glu, L-Arg, L-Val and L-Asp. D-AAAs represented <1% of the total AAAs concentration. D-Ala was observed in all the samples from September to November, while other D-AAAs were observed only sporadically, except for the samples collected during biomass burning events. The Student *t*-test confirmed that means of D/L ratio values for Asp, Thr, Phe, Ala were significantly higher in samples affected by wildfire or bonfire than those unaffected by such events. Interestingly, we observed high FAA levels during open biomass burning events.

The study of the sources affecting Belgrade aerosols was performed by means of four statistical approaches: CA, PCA, FcA, PMF. All proved to be in good agreement. Better information was obtained by FcA and PMF, which identified five contributing sources: a natural source present in September; a source related to fungi spores and degraded material, mostly present in September–October; and other 3 factors linked to biomass burning: open fire sources (e.g., wildfire); the combustion of grass and agricultural waste (bonfire of leaves, agricultural cuttings) principally present in October and November; and the combustion of biomass in stoves and industrial plants (November and December).

The discrimination of tree different sources of biomass burning, two of them concomitant in October, is an important result. This work suggests an approach to discriminate between concomitant biomass burning events. We used a wide range of specific tracers for wood combustion (anhydro sugars and PCs), fresh vegetal material (e.g., fructose and glucose), fungi (mannitol and arabitol) and degraded material and bacteria (D-AA) present also in soils. These tracers represent a series of compounds from the diverse matrices involved in biomass burning events. Statistical analysis allowed the discrimination of different types of biomass combustion in which the vegetal component is affected by the presence of other matrices.

## Acknowledgments

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## Appendix A. Supplementary data

Materials, description of the analytical methods for PCs, anhydrosugars, L-AAAs, D-AAAs, sugars determinations and quality control are available. Chromatographic separations of PCs and anhydrosugars are presented in Fig. S1 and S2. Temporal trends for target compounds analyzed are reported in Fig. S3–S14 and back-trajectories in Fig. S15–S18. Table S1 presents the summary of theoretical masses of the PCs and anhydrosugars studied; Table S2 reports the analytical performance of the developed methods; Table S3 reports the meteorological data for the sampling campaign and Table S4, the atmospheric concentration for all analytes determined in Belgrade's atmosphere. Supplementary data associated with this article can be found in the online version, at <http://dx.doi.org/10.1016/j.scitotenv.2016.06.188>.

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