- 1 GLOWORM-FL: a simulation model of the effects of climate and climate
- 2 change on the free-living stages of gastro-intestinal nematode parasites of

# 3 ruminants

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#### 14 Abstract

Gastrointestinal nematodes are important parasites of livestock and wildlife 15 worldwide, causing mortality and morbidity, regulating host populations and 16 threatening food security through reduced productivity of ruminant livestock. A 17 significant part of the life-cycle of most GINs is completed outside of the host. GINs 18 are therefore susceptible to changes in climate, and evidence of climate-driven 19 changes in the phenology of GINs and the seasonal incidence of disease already 20 21 exists. A modelling framework, GLOWORM-FL was developed to predict changes in the seasonal dynamics of the free-living stages of trichostrongylid GINs on pasture 22 as a first step towards evaluating potential mitigation strategies. The general model 23 framework was parameterised and validated for three GIN species that infect a 24 range of ruminants worldwide: Haemonchus contortus, Teladorsagia circumcincta 25 26 and Ostertagia ostertagi. The model builds significantly on previous models of GIN population dynamics by incorporating the behaviour of nematodes in response to 27 28 climate variability, facilitated by recent advances in our understanding of the ecology 29 of GINs. Simulations using historical and predicted future climatic data for a temperate region reveal the potential for an increase in annual infection pressure of 30 H. contortus and T. circumcincta in small ruminants as increasing temperatures 31 32 accelerate development and remove constraints on the development of *H. contortus* during the winter months. In contrast, a significant decrease in annual infection 33 pressure is predicted for O. ostertagi in cattle due to accelerated development being 34 offset by rapid mortality at higher temperatures. A similar trade-off is predicted during 35 the summer months for *H. contortus* and *T. circumcincta* resulting in complex 36 37 seasonal dynamics of the availability of infective stages on pasture. These changes could have significant impacts on the seasonal incidence and pathology of infection 38

- by GINs. GLOWORM-FL therefore provides an important tool to predict the seasonal
- 40 risk of transmission of GINs and will aid in the design of climate-driven, risk-based
- 41 GIN control strategies.
- 42 **Keywords:** gastrointestinal nematode; climate change; population dynamics;
- 43 nematode behaviour; ruminant; parasite ecology

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#### 46 **1. Introduction**

Trichostrongyloid gastrointestinal nematodes (GINs) are a major cause of mortality
and morbidity in livestock (e.g. Allonby and Urquhart, 1975), threatening food
security via constraints on productivity (Fitzpatrick, 2013). Costs of GINs have been
estimated at 84 million pounds sterling (105 million Euros) annually for the sheep
industry in the UK alone (Nieuwhof and Bishop, 2007), although effects of infection
on farm economics can be complex and difficult to estimate (van der Voort et al.,
2013).

Adult trichostrongylid GINs inhabit the gastrointestinal system of a range of host 54 species including ruminants (Allonby and Urguhart, 1975; Morgan et al., 2005), 55 lagomorphs (Newey et al., 2005) and birds (Hudson et al., 1998). Eggs are 56 deposited in the environment in faeces, where they develop to infective larvae, which 57 then move onto herbage. Larvae are ingested by the host during grazing and 58 complete their life-cycle in the host (Anderson, 2000). The development, survival and 59 60 behaviour of the free-living stages and thus the availability of infective stages for transmission is highly dependent on weather and micro-climatic conditions (Khadijah 61 et al., 2013a; Morgan and van Dijk, 2012; O'Connor et al., 2008, 2007, 2006; 62 Reynecke et al., 2011; Rose, 1963, 1961; van Dijk and Morgan, 2008, 2011). There 63 is evidence that recent increases in temperature in the UK have resulted in changes 64 in the phenology of GINs on pasture (Gethings et al. submitted) and in the incidence 65 of disease due to GIN infection (parasitic gastroenteritis) in livestock (van Dijk et al., 66 2008). As a result, the potential impact of climate change on GIN-host dynamics is of 67 increasing concern (Rose et al., 2014; van Dijk and Morgan, 2010; van Dijk et al., 68 2010). 69

70 Predicting climate-driven changes in the seasonal availability of free-living GIN infective stages is the first step to evaluating the potential impact of climate change 71 on GIN infections in livestock and wildlife and developing sustainable strategies to 72 control GINs and mitigate any increased transmission risk. These baseline 73 predictions of infection pressure can then be integrated with patterns of host 74 availability to evaluate the seasonal risk of transmission (e.g. Morgan et al., 2006). 75 76 However, predicting the response of GINs to climate change is complicated by nonlinear relationships and interactions between climate, development and survival 77 78 (Molnár et al., 2013), and the system necessitates parsimonious predictive models that balance sufficient biological detail with experimentally verifiable parameters 79 (Morgan, 2013) and computational efficiency. 80

Numerous gastrointestinal nematode models have been developed over previous 81 82 decades (reviewed elsewhere by Cornell, 2005; Roberts, 1995; Smith and Grenfell, 1994). Many are deliberately simple in order to explore model behaviour and system 83 dynamics (Cornell et al., 2004; Grenfell, 1992; Louie et al., 2005; Roberts and 84 Heesterbeek, 1995). Others include more biological detail in order to address 85 specific questions (Grenfell et al., 1987; Laurenson et al., 2011; Learmount et al., 86 87 2006; Leathwick et al., 1995, 1992; Smith et al., 1987). However, climate-dependent life-history parameters that determine the availability of infective stages on pasture 88 are often set at a constant rate (Laurenson et al., 2013). Furthermore, although 89 90 many models incorporate climate-dependence (Grenfell et al., 1987; Molnár et al., 2013) and stage-specific mortality and development rates, to the authors' knowledge 91 no model explicitly incorporates movement of infective larvae between soil and 92 herbage nor addresses moisture-limitations on migration between faeces and 93 pasture (herbage and soil combined). Detail such as this will become increasingly 94

important as increases in the frequency of extreme events such as drought and
heavy rainfall are predicted into the late 21<sup>st</sup> century (IPCC, 2013).

The model framework presented here, GLOWORM-FL, builds on the work of 97 Grenfell et al. (1987) and Smith (1990) by incorporating recent advances in our 98 understanding of the behaviour and ecology of GINs on pasture to predict the 99 climate-dependent seasonal dynamics of GIN infection pressure. The model 100 101 provides a generic framework that can be applied to a range of GIN species. To demonstrate the flexibility of the framework and methods for data-driven parameter 102 estimation, the model is parameterised and validated for three trichostrongylid GIN 103 104 species - Haemonchus contortus, Teladorsagia circumcincta and Ostertagia ostertagi - and used to simulate the seasonal dynamics of the availability of infective 105 stages on pasture under scenarios of likely climate change, independent of host 106 107 factors.

108 The three species of GIN chosen here are of economic importance to the ruminant 109 livestock industry worldwide, but also have a broad host range and infect freeranging ruminants. The haematophagous abomasal nematode, *H. contortus* is highly 110 pathogenic. Chronic infections in sheep may result in anaemia and death (Allonby 111 and Urquhart, 1975). The abomasal nematodes T. circumcincta and O. ostertagi are 112 responsible for significant production losses in the ruminant livestock industry 113 (Charlier et al., 2009; Nieuwhof and Bishop, 2007). Anthelmintic resistance is 114 increasingly widespread in all three species in livestock (De Graef et al., 2013; 115 Kaplan and Vidyashankar, 2012; Papadopoulos et al., 2012; Sutherland and 116 Leathwick, 2011) and has been recorded in *H. contortus* in wild deer (Chintoan-Uta 117 et al., 2014). A better understanding of the population ecology of these parasites is 118

- therefore needed to underpin the development of alternative control strategies
- against the backdrop of climate change and anthelmintic resistance.

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#### 122 **2. Materials and Methods**

### 123 **2.1 GLOWORM-FL model framework**

124 The model framework is based on the general life-cycle of the free-living stages of

trichostrongylid gastrointestinal nematodes (Figure 1). Eggs (*E*) develop to third

stage larvae in the faeces (*L3f*) via the pre-infective larval stages (*L*), and are subject

to stage-specific mortality rates ( $\mu_i$ ). As pre-infective larval stages (first stage larvae,

L1, and second-stage larvae, L2) are not separated, the model can be applied to

trichostrongylids that hatch as first stage larvae (e.g. Haemonchus spp.,

130 *Teladorsagia* spp. and *Ostertagia* spp.; Anderson, 2000), or second stage larvae

131 (*Marshallagia marshalli*; Carlsson et al., 2013).

$$\frac{dE}{dt} = -(\mu_1 + 2\delta)E + E_{new}C \tag{1}$$

$$\frac{dL}{dt} = -(\mu_2 + 2\delta)L + 2\delta E \tag{2}$$

$$\frac{dL3_f}{dt} = -(\mu_3 + m_1)L3_f + 2\delta L$$
(3)

As development from egg to L3 in faeces is divided over two stages, the development rate ( $\delta$ ) is doubled. The framework tracks numbers of overlapping cohorts of nematodes, and so new eggs deposited on pasture ( $E_{new}$ ) join the pool of existing eggs.

Previous models of GIN free-living stages either model total L3 on pasture and implicitly assume that, once developed, L3 are available for transmission (Grenfell et al., 1987; Learmount et al., 2006), or separate L3 in faeces from L3 on pasture and apply a constant horizontal migration rate (Grenfell et al., 1986; Smith, 1990). Here, 140 L3 in faeces actively migrate from faeces onto pasture at a climate-dependent 141 horizontal migration rate  $(m_1)$ .

Once on pasture, L3 can be recovered from both soil/mat layer and herbage. Although 142 Grenfell et al. (1986) included losses of L3 in soil due to soil moisture deficit or wash-143 down during rainfall in their model, movement between soil and herbage was not 144 considered. Experiments have demonstrated the potential for bi-directional movement 145 of trichostrongylid L3 between soil and herbage (Krecek and Murrell, 1988; Rose and 146 Small, 1985) and that there is random movement between the soil and herbage (van 147 Dijk and Morgan 2011). Therefore, L3 on pasture (*L3p*) are assumed to reside in either 148 149 the soil and vegetation mat layer (L3s) or on herbage (L3h). In order to simulate random, bi-directional movement between herbage and the soil reservoir, substrate-150 specific mortality rates ( $\mu_4$ ,  $\mu_5$ ) are applied to the proportion of larvae estimated to 151 reside in soil and on herbage respectively, dependant on a vertical migration 152 parameter  $(m_2)$ . 153

$$\frac{dL3_p}{dt} = -\mu_4 \left( (1 - m_2)L3_p \right) - \mu_5 (m_2 L3_p) + m_1 L3_f$$
(4)

State variables and parameter definitions are listed in Table 1. The model was 154 implemented in R (R Core Team, 2013) using the Isoda function in the deSolve 155 package (Soetaert et al., 2010). *Isoda* uses an Adams-BDF (backward differentiation 156 formulae) adaptive integration method that detects the stiffness of the problem 157 throughout the simulation and switches between Adams and BDF integration 158 accordingly (Soetaert et al., 2010). The model returns daily output but the time steps 159 used for integration are not known prior to simulation when using the Adams-BDF 160 integration method. Therefore, time-series of variable climate-dependent rates e.g. 161 temperature-dependent development rates, were generated prior to simulation and 162

introduced by interpolation using the *approxfun* function (Soetaert et al., 2012). New
eggs were deposited using the "events" argument of the *lsoda* function (Soetaert et
al., 2012). Model output is numbers of individuals per unit area e.g. per hectare, and
is therefore independent of herbage density.

#### 167 **2.2 Parameter estimates**

The model was parameterised for three trichostrongylid GINs that infect ruminants: *Haemonchus contortus, Teladorsagia circumcincta* and *Ostertagia ostertagi* (Figures
2-4; Table 2).

### 171 **2.2.1 Temperature dependent development and mortality**

Temperature-dependent instantaneous daily rates were estimated for development from egg to L3 in faeces and stage- and substrate-specific mortality using data from experiments that reported the proportions of individuals developed (for development rates) or surviving (for mortality rates) at discrete intervals and at a range of constant temperatures (Table 2).

Instantaneous daily rates were first estimated for each constant temperature from 177 the reported time to 50% development (D50) or time to 50% mortality (M50) 178 as -ln(0.5/D50 or -ln(0.5)/M50. If these data were unavailable, rates were estimated 179 in one of three ways: 1) using the proportion remaining at a single sampling interval, 180 181 as -In(proportion remaining)/days; 2) using the mean of the minimum and maximum development or mortality times, or; 3) by linear regression of the transformed 182 proportions of individuals developed or surviving over time as described by Azam et 183 al. (2012). Where 100% mortality was observed within 24 hours, an instantaneous 184 mortality rate of 1 was applied. 185

Linear models were then fitted to the instantaneous daily rates at a range of temperatures, yielding a regression equation that could be used to estimate daily rates dependent on time-series of observed temperature data (Table 2). For development rates, which increase linearly as a function of temperature, simple linear regression was used. Mortality rates are highest at extreme high and low temperatures, therefore polynomial models were fitted to the log-transformed instantaneous mortality rates. Rates were limited between 0 and 1 where necessary.

Data were only used to estimate the mortality rates of pre-infective stages if the
temperatures were low enough to preclude development, or high enough that
mortality occurred prior to development to the next larval stage (e.g. Todd et al.,
1976). Development of L3 is arrested until they are ingested by the host. Therefore a
range of temperatures could be used to estimate the substrate-specific mortality
rates of L3.

199 The mortality rate of L3 in soil ( $\mu_4$ ) for all GIN species was estimated using

observations on the mortality of L3 in water, which for *H. contortus* and *T.* 

201 *circumcincta*, provided point estimates of instantaneous daily mortality rates similar

to those reported by van Dijk and Morgan (2011) in soil at 20-24°C (Figures 2-3).

Exposure to UV irradiation increases the mortality of trichostrongyloid L3 in water (van Dijk et al., 2009) and the estimates of mortality in soil (Table 2; van Dijk and Morgan, 2011) are considerably lower than estimates of mortality on pasture (Grenfell et al., 1986). Therefore it is assumed that the mortality rate of L3 on herbage is higher than in soil, and the mortality of L3 in faeces ( $\mu_3$ ) was used as a proxy for L3 mortality on herbage ( $\mu_5$ ; Table 2). 209 No data were available to estimate the mortality of pre-infective larvae ( $\mu_2$ ) and L3 in faeces ( $\mu_3$ ) for T. circumcincta and O. ostertagi, Therefore the same mortality rate 210 was used for eggs ( $\mu_1$ ) and pre-infective larvae ( $\mu_2$ ; Table 2). To estimate L3 211 mortality in faeces the temperature-dependent mortality of O. ostertagi in water 212 (used to estimate  $\mu_4$ ) was compared with point estimates of mortality of O. ostertagi 213 and Cooperia oncophora in cow manure (Persson, 1974a). The instantaneous daily 214 mortality rates at 20°C and 3°C were estimated using Persson's data. As there was 215 significant variability between sampling intervals the instantaneous mortality rate was 216 217 calculated for each sampling interval and the mean estimated from these rates. Based on these analyses L3 mortality in soil is 4.9-18.5 times lower than in faeces. 218 Therefore, in the absence of data to directly estimate the mortality rate of *T*. 219 220 *circumcincta* and *O. ostertagi* L3 in faeces, it is estimated that  $\mu_3 = 10\mu_4$ , within the limits of 0 and 1 (Figures 3-4). 221

### 222 2.2.2 Moisture limitations and differences between host species

223 Moisture limitations on the availability of GIN infective stages are primarily mediated through changes in faecal moisture content (FMC; Mauleon and Gruner, 1984; 224 Rossanigo and Gruner, 1995). There are significant differences in faeces structure 225 and drying rates between host species. Sheep faecal pellets tend to dry rapidly 226 following deposition, whereas the decrease in cow pat FMC is more gradual 227 (Mauleon and Gruner, 1984). It is therefore necessary to not only parameterise the 228 model for different nematode species, but also different host species. Here, we 229 consider moisture limitations on *H. contortus* and *T. circumcincta* infecting sheep or 230 other ruminants with a similar faecal pellet structure, and O. ostertagi infecting cattle. 231

### 232 2.2.2.1 Moisture limitations on development success

In addition to temperature limitations, development success (the proportion of eggs 233 that develop to L3) is also a function of faecal moisture content (Rossanigo and 234 Gruner, 1995). To impose faecal moisture limitations on development and mortality 235 of *H. contortus* and *T. circumcincta* without explicitly modelling FMC, cumulative 236 precipitation divided by cumulative potential evapotranspiration (referred to as 237 cumulative P/E) is estimated for a species-specific critical period following deposition 238 of eggs after O'Connor et al. (2008), who observed a strong positive relationship 239 between cumulative P/E and the FMC of sheep faecal pellets. If cumulative P/E<1 240 241 then the number of new eggs ( $E_{new}$ ) entering the pool of eggs in faeces is reduced by an amount specified by the correction factor parameter, C. 242

O'Connor et al. (2008) observed a significant decrease in the development success 243 of *H. contortus* where cumulative P/E fell below 1 within 4 days of deposition of eggs. 244 245 Khadijah et al. (2013a) recovered maximum H. contortus L3 from faecal pellets and soil when simulated rainfall was applied between -1 and 2 days post deposition of 246 faeces containing eggs and concluded that faecal moisture 48-72 hours post 247 deposition was important for development success. No L3 were recovered from un-248 watered controls. Similar data were not available for *T. circumcincta*. However, 249 250 Khadijah et al. (2013a) note that for *Trichostrongylus colubriformis*, faecal moisture in the period 72-96h post deposition is important for development success. This 251 period is likely to be extended for *T. circumcincta* which is more resistant to 252 desiccation than T. colubriformis. A lower faecal moisture content (FMC) threshold 253 was observed for T. circumcincta development (yielding  $\geq 1$  L3 per 100 eggs) than T. 254 colubriformis (25% and 35% respectively; Rossanigo and Gruner, 1995). The critical 255 periods for H. contortus and T. circumcincta were therefore identified as 4 days and 256 7 days post deposition of eggs respectively (Khadijah et al., 2013b; O'Connor et al., 257

258 2008). The cumulative P/E for development success is referred to as P/E<sub>4</sub> for *H*. 259 *contortus* and *P/E*<sub>7</sub> for *T. circumcincta*.

A protective surface crust forms on cow pats soon after deposition. It is therefore assumed that moisture is not limiting for GIN development within cattle faeces at lower FMCs (Rose, 1961).

### 263 **2.2.2.2 Moisture limitations on the translation of L3 onto pasture**

Laboratory observations on the migration of H. contortus and T. colubriformis L3 264 indicate that, similar to development success, moisture limitations on horizontal 265 migration are mediated by faecal moisture content, which varies as a result of 266 interacting microclimatic factors (Khadijah et al., 2013a, 2013b; O'Connor et al., 267 2008, 2007; van Dijk and Morgan, 2011). Few data were available to estimate the 268 temperature- and moisture-dependent horizontal migration rate of infective larvae 269 from faeces onto pasture. Therefore, data in the published literature were 270 supplemented with laboratory experiments to derive heuristic estimates for horizontal 271 migration under: 1) optimal moisture conditions (sufficient rainfall); 2) sub-optimal 272 moisture conditions (insufficient rainfall but sufficient FMC) and; 3) moisture-limiting 273 274 conditions (low FMC and insufficient rainfall).

Horizontal migration of GINs has been observed from cow pats following 1.6mm of simulated rainfall (Grønvold and Høgh-Schmidt, 1989) and from sheep faecal pellets following 2mm of simulated rainfall (this study). Furthermore, horizontal migration rates of *H. contortus* were not significantly influenced by the amount of rainfall between 4mm and 8mm (Wang et al., 2014). Therefore, optimal moisture was defined as days where total precipitation  $\geq$  2mm. A daily horizontal migration rate of 0.06 (S.D. 0.057) applied for *O. ostertagi* based on the number of L3 recovered by Grønvold and Høgh-Schmidt (1989) from within and outside of cow pats after 1.6-1.7mm simulated rainfall was applied to pats with FMCs of 54-66%.

To estimate the daily horizontal migration rate for *H. contortus* and *T. circumcincta* 285 faeces containing either H. contortus or T. circumcincta eggs provided by Moredun 286 Research Institute, Edinburgh, UK were incubated at 20°C for 7 days and then 287 allowed to dry at room temperature for varying amounts of time to obtain pellets with 288 variable initial faecal moisture content (FMC). Three replicates of 3g (~6 pellets) 289 290 were subjected to approximately 2mm simulated rainfall and after 24 hours L3 that had migrated out of faeces and L3 remaining in the faeces were recovered and 291 enumerated (Wang et al., 2014). The recovery efficiency of extra-pellet and intra-292 pellet L3 were determined to be 84% (S.D. 3%) and 74% (S.D. 7%) respectively by 293 placing a known number of L3 in the cup used to contain L3 that had migrated out of 294 faeces (Wang et al., 2014) and seeding faeces with a known number of L3. The 295 weight of each pellet was recorded before and immediately after the rainfall event, 296 and after drying in an oven, to estimate the FMC at each stage. 297

The proportion of L3 that had migrated out of faeces was calculated as:

extra-pellet L3/(extra-pellet L3 + intra-pellet L3), corrected for recovery efficiency.

Mean daily horizontal migration rates of 0.25 (S.D. 0.11) and 0.21 (S.D. 0.44) were

301 observed for *H. contortus* and *T. circumcincta*. FMCs prior to rainfall were 3-61% (*H.* 

302 contortus) and 7-34% (T. circumcincta), increasing to 45-73% (H. contortus) and 39-

303 56% (*T. circumcincta*) after simulated rainfall.

304 O. ostertagi L3 were only observed on the surface of experimental pats that had been watered (Grønvold and Høgh-Schmidt, 1989) suggesting that rainfall is 305 required to moisten the protective surface crust sufficiently to allow migration. 306 307 Therefore, estimates of horizontal migration of *O. ostertagi* under sub-optimal moisture conditions were not considered. In contrast, small numbers of extra-pellet 308 H. contortus and T. colubriformis L3 have been recovered in the absence of rainfall 309 (O'Connor et al., 2008; Wang et al., 2014). Therefore a 4- and 7-day trailing 310 cumulative P/E rule was applied to H. contortus and T. circumcincta respectively to 311 312 characterise sub-optimal conditions, extrapolated from the observations of O'Connor et al. (2008) and Khadijah et al. (2013b) on the effect of cumulative P/E on FMC and 313 development success. Sub-optimal days were defined as days where total 314 precipitation<2mm, and trailing cumulative P/E≥1. The species specific trailing 315 cumulative P/E values for migration are referred to as P/E-4 for H. contortus and 316 P/E-7 for *T. circumcincta*. 317

An estimated horizontal migration rate of 0.051 for *H. contortus* under sub-optimal moisture conditions was derived from observations by O'Connor et al. (2008), where 30% of L3 migrated out of faeces within a 7 day period in the absence of rain but following a period of simulated rainfall in the preceding 7 days. This is consistent with the observed mean migration rate of 0.057 (S.D. 0.027) for *H. contortus* maintained under high relative humidity of 98% and FMC of ~60% (Wang et al., 2014).

To estimate the corresponding rate for *T. circumcincta* the instantaneous daily migration rate of *H. contortus* L3 estimated from data provided by Wang et al. (2014) was compared with the instantaneous daily migration rate of *T. circumcincta* estimated from an unpublished experiment conducted concurrently with the experiment of Wang et al. (2014) and using identical methods. These experiments show that under optimal moisture conditions the instantaneous daily migration rate of *T. circumcincta* is 49% that of *H. contortus*. Ninety-nine percent (S.D. 0.4) of *H. contortus* L3 had migrated out of faeces within 7 days (Wang et al. 2014), compared with 91% (S.D. 6.8) of *T. circumcincta* L3, giving instantaneous daily horizontal migration rates of 0.71 and 0.35 respectively. Thus the estimated instantaneous daily migration rate for *T. circumcincta* where moisture is sub-optimal is 0.051\*0.49 = 0.025 (Table 2).

Finally, a horizontal migration rate of 0 was applied for all GIN species on moisture-

337 limited days where P/E<1 and total precipitation<2mm.

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### 339 2.2.3 Migration between soil and herbage

Crofton (1948) observed seasonal patterns in the vertical migration of 340 Trichostrongylus retortaeformis L3 on pasture where fewer L3 were recovered from 341 the upper herbage layer in winter than in summer. It is likely that interacting climatic 342 and other abiotic variables including temperature, moisture, biomass composition 343 and light drive this seasonality (Amaradasa et al., 2010; Callinan and Westcott, 344 1986; Crofton, 1948; Dusenbery, 1989; Ogbourne, 1973; Rees, 1950; Saunders et 345 al., 2000; Silangwa and Todd, 1964; van Dijk et al., 2009). However, the majority of 346 studies have sampled only a superficial layer of soil (e.g. Crofton, 1948; Rees, 1950) 347 and therefore could underestimate the proportion of pasture L3 in soil relative to L3 348 349 on herbage. Mesocosm experiments (e.g. Callinan and Westcott, 1986; Knapp-Lewitzke et al. in preparation) offer an alternative to ensure more complete sampling 350 of L3 in soil. The temperature-dependent proportion of trichostrongylid L3 expected 351 on herbage and in soil was estimated by fitting a second order polynomial regression 352

353 to the log transformed proportion of total *Teladorsagia* and *Trichostrongylus* spp. L3 recovered from herbage (Callinan and Westcott, 1986). In the absence of suitable 354 species-specific data, the same estimate was used for all trichostrongylid GIN 355 species, subject to validation. The observation at 20°C was omitted from analysis as 356 the decrease in L3 recovered from herbage was inconsistent with observations of L3 357 availability on pasture where the percentage of L3 recovered from herbage tends to 358 359 increase with increasing mean soil temperature between approximately 8-22°C (Callinan, 1979, 1978a, 1978b). 360

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### 362 2.3 Model validation

Laboratory observations on the development success of *T. circumcincta* and *O.* 363 364 ostertagi (Rossanigo and Gruner, 1995) and field observations on the development time and development success of *H. contortus* (Rose, 1963) and *O. ostertagi* (Rose, 365 1961) in a temperate region were used to validate model predictions of development 366 and survival in faeces. Rossanigo and Gruner (1995) recorded the development 367 success of various GIN species in faeces incubated at a range of constant 368 369 temperature while maintaining optimal moisture conditions. Rose (1963, 1961) recovered L3 from faeces containing eggs that had been deposited at monthly 370 intervals on pasture in South East England where the GINs were exposed to variable 371 temperatures and moisture conditions. The model was initialised with 100 eggs and 372 simulations were run using either the constant temperatures tested (Rossanigo and 373 Gruner, 1995) or time series of daily air temperatures obtained for the study location. 374 375 Meteorological data from Wisley weather station (Ordnance Survey grid reference: TQ062579), approximately 10km from the Central Veterinary Laboratories, 376

Weybridge, where Rose's (1963) observations were made, were obtained from the 377 British Atmospheric Data Centre (badc.nerc.ac.uk). The same data were not 378 available for the time period of observations made by Rose (1961). Therefore, 379 380 meteorological data were obtained from the E-OBS gridded dataset (lat/lon: 51.355°N, -0.496°E; Haylock et al., 2008). Potential evaporation (mm/day) was 381 estimated from mean air temperatures using the Hamon method (Xu and Singh, 382 383 2001). Two simulations were run for each monthly deposit using mean daily temperature and linear fluctuations between minimum and maximum daily 384 385 temperature to determine whether mean air temperature is sufficient to predict development when temperatures are close to the minimum threshold for 386 development. Horizontal migration,  $m_1$ , was set to 0 to prevent migration out of 387 faeces. 388

389 Field observations of *H. contortus* and other trichostrongylid

(Trichostrongylus/Teladorsagia spp.) L3 over winter on naturally contaminated 390 391 pasture in the absence of continued grazing by livestock (Wilkie et al. submitted) were used to validate the predicted dynamics of *H. contortus* and *T. circumcincta* L3 392 availability on herbage. Temperature data for the observation period were obtained 393 from Yeovilton weather station (Ordnance survey grid reference: ST549231), 394 approximately 60km from the farm where observations were made from the British 395 Atmospheric Data Centre. The initial number of L3 recovered from herbage at the 396 397 start of the observation period and the vertical migration parameter,  $m_2$ , were used to estimate the corresponding initial number of L3 expected in soil. All other initial 398 values were set to 0. For each simulation the daily number of L3 on herbage is a 399 product of the daily number of L3 on pasture, L3p, and the vertical migration 400 parameter,  $m_2$ . 401

To determine whether the additional complexity of the pasture component of the GLOWORM-FL model was justified, simulations using an existing model for *H. contortus* (Smith, 1990) were also validated using Wilkie's (submitted) data as described above.

For each validation dataset and corresponding simulations, model fit was assessed 406 using the residual sum of squares (RSS; Mayer and Butler, 1993) and linear 407 regression through the origin. An intercept of 0 and slope of 1 would indicate perfect 408 correspondence between model output and observations, therefore a regression 409 through the origin with a slope that is not statistically significantly different from 1 410 411 indicates a good model fit. It is assumed that the slope is not statistically significantly different from 1 if the 95% confidence interval (estimated as the coefficient  $\pm$ (2 x 412 standard error of the coefficient) includes 1. 413

#### 414 **2.4 Climate change simulations**

The validated model was run using mean daily temperature and total daily 415 precipitation data from the atmospheric dataset provided by the Coupled Model 416 Intercomparison Project Phase 5 (CMIP5; Taylor et al., 2012) to predict the potential 417 418 impact of current climate change predictions on the seasonal availability of L3 on pasture (infection pressure). Simulations ran for 30-year time periods using either 419 historical climatic data for the period 01/12/1969-30/11/1999 (representative of 420 current climate) or a high emissions scenario (Representative Concentration 421 422 Pathway 8.5; RCP8.5) for the period 01/12/2070-30/11/2100, from the HadGEM2-ES model output (ensemble r1i1p1) developed and run by the Met Office Hadley Centre 423 424 (Collins et al., 2011; Martin et al., 2011). Characteristics of the RCP8.5 scenario include high greenhouse gas emissions, a high rate of population growth, a 425

dependence on fossil fuel and global CO<sub>2</sub> concentrations of ~950ppm by 2100 (van 426 Vuuren et al., 2011). For comparison, record CO<sub>2</sub> concentrations of over 400ppm 427 were observed at the Mauna Loa observatory in May 2014 (Tans, 2014). 428 Time series of mean daily air temperature and total daily precipitation were extracted 429 for a grid cell in North Somerset in South West England, UK. This area is of 430 particular interest as recent climate change has been associated with an increase in 431 diagnoses of parasitic gastroenteritis in the region (van Dijk et al., 2008). 432 433 One hundred new eggs (Enew) were added daily to simulate a scenario of continuous grazing and host infection without making assumptions about management or 434 seasonal changes in intensity of infection/nematode egg output. The first year of 435 simulation was discarded as L3 accumulated on pasture throughout the first year. 436 Output is presented as annual time series of daily mean numbers of L3 on pasture, 437 calculated using the remaining 29-year output disaggregated into annual time series. 438 The area under curve (AUC) was calculated for each year using a trapezoid function 439 in R to estimate the annual infection pressure under the historical and future climate 440 scenarios. Wilcoxon rank sum tests were used to compare scenarios for each GIN 441 442 species (Figure 5).

443

#### 444 **3. Results**

#### 445 **3.1 Model validation**

446 Overall the model was able to reproduce the observed development times,

development success, and dynamics of L3 on pasture (Table 3), demonstrating the
potential for a generic framework such as GLOWORM-FL to be adapted to suit
different GIN and host species.

A slope marginally greater than 1 suggested that there was a tendency to

under-predict the development success of *T. circumcincta* and *O. ostertagi* at

452 constant temperatures between 5°C and 35°C compared with laboratory

453 observations (Table 3).

The model performed well when tested against field observations of *H. contortus* and *O. ostertagi* development and survival in faeces (Table 3) demonstrating that the models, parameterised using laboratory data, transferred well onto conditions observed in the field. The range of mean air temperatures during the observation periods was -3.9 to 24.7°C (Rose, 1963) and -2.5 to 23.2°C (Rose, 1961). The range of total daily precipitation during the observation periods was 0-31.2mm (Rose, 1963) and 0-23.2mm (Rose, 1961).

The predicted dynamics and numbers of *H. contortus* and *T. circumcincta* L3 on
pasture fitted observations well, replicating the initial decrease in L3 density on
herbage followed by an increase, despite no further contamination of pasture (Figure
6; Table 3). This seasonal variability in L3 on herbage can be explained by the
vertical migration of L3 between soil and herbage; a model using only a temperaturedependent mortality rate and not considering movement between the soil and

herbage was not able to replicate these dynamics. The range of mean daily air
temperatures during the observation period was -3.05 to13.7°C.

The GLOWORM-FL model, validated using observed numbers of *H. contortus* on
pasture, outperformed a previously published, less complex model (Table 3; Figure
6).

The performance of models using minimum and maximum or mean daily
temperatures varied dependent on the validation dataset (Table 3) and in most
cases both models gave similar output. However, models using minimum and
maximum daily temperatures performed poorly against observation of *O. ostertagi*development in the field (Table 3). Therefore mean temperatures were used for all
subsequent simulations.

#### 478 **3.2 Climate change simulations**

At the chosen test location in South West England, UK, the HadGEM-ES model 479 predicts warmer wetter winters and warmer, drier summers during 2070-2100 under 480 the RCP8.5 high emissions scenario, compared with the historical period of 1969-481 1999. A mean increase in mean air temperature of 4.57°C (S.D. 1.91°C) is predicted 482 by 2070-2100. The increase is greatest during the summer months with a maximum 483 of 8.65°C increase predicted during July and a minimum of 1.49°C increase 484 predicted during March. A mean decrease of 0.03mm (S.D. 1.09mm) in mean daily 485 rainfall is predicted under the RCP8.5 scenario, with an increase of up to 3.02mm 486 during the winter period and a decrease of up to 3.06mm during the summer period. 487 The change in seasonal temperatures and rainfall resulted in an increase in 488 predicted development rate throughout the year whereas mortality rates decreased 489 during the winter and increased during the summer. The pattern of moisture-490

limitation on development success and horizontal migration of *T. circumcincta* and *H. contortus* was similar under both scenarios. Although the patterns of change in lifehistory parameters were similar, the magnitude of change was species dependent,
resulting in differing seasonal patterns of L3 on pasture.

There was a significant predicted increase in annual infection pressure for both H. 495 contortus (W=47, p<0.001) and T. circumcincta (W=95, p<0.001; Figure 5) under the 496 RCP8.5 scenario compared with historical climatic data. Mean air temperature was 497 regularly higher than the predicted lower threshold for development of T. 498 circumcincta (4.46°C) when both historical and RCP8.5 data were used and 499 500 development was possible year round. However, the number of days where development was possible increased from 328 to 360 (the HadGEM-ES model was 501 run on a 360 day year and therefore 360 represents the entire year). The increase in 502 503 temperatures predicted under RCP8.5 therefore resulted in increased development rates year-round for T. circumcincta. In contrast, very little H. contortus development 504 505 is completed over winter when using historical climatic data as the mean air temperature rarely rises above the predicted threshold for development of 9.17°C. 506 Therefore, the increase in temperatures predicted under RCP8.5 not only results in 507 508 an increase in development rate but also a lengthening of the season during which development is possible. The period during which mean daily temperatures 509 exceeded the development threshold for H. contortus was extended by 3.3 months 510 from 188 days between March and September under historical climate to 258 days 511 between February and December under RCP8.5. A corresponding decrease in 512 mortality rates during the winter results in an overall increase in infection pressure 513 over the winter period, extending into early summer. Further increases in 514 temperatures during the summer result in increased mortality which offsets the 515

increased development rate and results in a decrease in the number of L3 onpasture, below numbers predicted using historical data (Figure 7).

A similar pattern of summer mortality is predicted for *O. ostertagi* (Figure 7). When 518 using historical climatic data, there is a small increase in L3 on pasture during the 519 spring as temperatures exceed the predicted threshold for development of 7.44°C, 520 but the large number of L3 predicted on pasture is fairly consistent throughout the 521 year due to low mortality rates. However, there is a significant decrease in annual 522 infection pressure when using the RCP8.5 climatic data, compared with predictions 523 using historical data (W=95, p<0.001). The period during which mean daily 524 525 temperatures exceeded the development threshold for *O. ostertagi* was extended by 4.6 months from 216 days between March and October under historical climate to 526 347 days throughout the year under RCP8.5, but this is offset by significant 527 increases in mortality rates between May and November depleting the reservoir of 528 L3 in faeces and soil. 529

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531

#### 532 **4. Discussion**

533 The model presented here consolidates advances in our understanding of the ecology and behaviour of gastrointestinal nematode free-living stages with the 534 numerous existing models developed to simulate the population dynamics of GINs in 535 ruminant livestock. Previous models were species-specific (Grenfell et al., 1986; 536 Smith, 1990), restricted to livestock ruminants (Learmount et al., 2006), and 537 constrained by the available data at the time they were developed. GLOWORM-FL 538 builds significantly on these models to incorporate the active movement of GINs 539 between substrates, and substrate-specific mortality rates, in addition to explicitly 540 541 climate-dependent life-history parameters. Comparison of the output of GLOWORM-FL with an example of a preceding model that did not consider nematode behaviour 542 with field observations confirmed that the additional complexity significantly improved 543 model predictive performance. This is probably because the majority of L3 are 544 sequestered in the soil at any one time (Callinan and Westcott, 1986; Silangwa and 545 Todd, 1964; van Dijk and Morgan, 2011), and emerge onto herbage when climatic 546 conditions allow. Therefore, soil should not be overlooked as a significant source of 547 infection, acting as a reservoir for L3 that can recolonize herbage (van Dijk and 548 549 Morgan, 2011). For the same reason, absence of L3 on herbage should not be interpreted as evidence for absence of GINs. 550

551 An additional motivation for the development of the model was that characteristics 552 relevant to the epidemiology of GINs are similar between different livestock systems 553 and wild ruminants (Rose et al., 2014) and there is evidence of transmission where 554 livestock and wildlife meet or share ranges (Chintoan-Uta et al., 2014; Morgan et al., 555 2007). There is therefore a need for a common framework that can be applied to a 556 range of GIN and host species. The GLOWORM-FL model framework was

parameterised and successfully validated using data available in the literature for 557 three GIN species that are economically important parasites of livestock worldwide 558 but also infect free-living ruminants (Morgan et al., 2005). Due to their economic 559 importance, research on these species spans decades, thus providing sufficient data 560 for parameter estimation. There were some gaps in the available data and therefore 561 these species also provided an opportunity to demonstrate how the model can be 562 563 successfully adapted by drawing on similarities between GIN species and robust validation exercises. 564

GINs are a global constraint on livestock production (Nieuwhof and Bishop, 2007; 565 566 Perry and Grace, 2009). The increasing prevalence of anthelmintic resistance worldwide (Kaplan and Vidyashankar, 2012) and the threat of altered seasonal 567 patterns of transmission due to climate change (Gethings et al. submitted; Molnár et 568 al., 2013; van Dijk and Morgan, 2010; van Dijk et al., 2008) necessitate the 569 development of alternative control strategies (Krecek and Waller, 2006). It may be 570 possible to control the magnitude of exposure to GINs and therefore intensity of 571 infection and production losses, for example by altering management practices to 572 avoid grazing during periods of high risk or targeting treatments according to risk of 573 574 exposure or suitability for development of free-living stages. To do this, understanding the population ecology of GINs and predicting the seasonal dynamics 575 of infection pressure is fundamental. 576

577 GLOWORM-FL provides a tool to aid in the development of climate-based GIN 578 control methods. The model can be used to track pasture contamination and 579 evaluate the resultant climate-dependent infection pressure under a range of 580 management and climate scenarios. Here, its use is demonstrated using climatic 581 data representative of recent historical climate and climate expected under the

IPCC's RCP8.5 scenario. Climate-driven changes in the seasonal availability of L3 582 on pasture are likely to become increasingly important in the dynamics of GIN 583 infection, particularly where host behaviour or farm management is slow to adapt in 584 response to the change. In some cases, this may lead to an increase in disease (van 585 Dijk et al., 2008) whereas under different circumstances climate-driven changes may 586 decrease exposure to infection, as has apparently been the case for Nematodirus 587 battus infections in lambs in some parts of the UK (Gethings et al. submitted). A 588 better understanding of the seasonal dynamics of infection pressure will be key to 589 590 the future of sustainable GIN control in livestock and could also benefit the management and conservation of wild ruminants. 591

Using historical climatic data, large numbers of *O. ostertagi* L3 were predicted year 592 round on pasture due to low mortality rates over winter and a turnover of L3 between 593 594 April and November when development rates increase and compensate for losses due to the increased mortality rate. This suggests that the observed patterns of 595 ostertagiasis in calves in Europe (Williams et al., 1993), where peak worm burdens 596 are observed towards the end of the grazing season, are driven by cumulative 597 exposure to L3 on pasture and management or host factors as opposed to seasonal 598 599 variability in infection pressure (Höglund et al., 2013; Roberts and Grenfell, 1992). Under the RCP8.5 climate scenario and a constant input of eggs, a decrease in O. 600 ostertagi infection pressure is predicted throughout the year due to significant 601 602 increases in predicted mortality rates depleting the reservoirs of infective stages in faeces and on pasture. Although a reduction in the magnitude of exposure to 603 infective stages is favourable, there may also be an adverse impact on the 604 development of immunity through reduced exposure to L3 (Ploeger et al., 1995). 605 However, since the epidemiology of *O. ostertagi* infection is largely driven by 606

management and host factors (Höglund et al., 2013; Roberts and Grenfell, 1992),
altered management strategies in response to climate change may negate the
change in seasonal availability of L3 on pasture predicted here.

The seasonal incidence of *H. contortus* infection is primarily climate-driven, and the 610 pattern predicted here for South West England using the historical climatic data 611 broadly mirrors the seasonal diagnoses of haemonchosis in sheep in the region (van 612 613 Dijk et al., 2008). The implications of this for the control of haemonchosis in livestock are that predicted changes in L3 on pasture are likely to result in similar changes in 614 the seasonal incidence of haemonchosis. The predicted pattern of infection pressure 615 616 for T. circumcincta when using historical climatic data also reflects patterns of seasonal diagnoses (van Dijk et al., 2008), indicating a degree of climate-617 dependence in the epidemiology of *T. circumcincta* infection in sheep in South West 618 England. 619

620 An increase in temperature during the winter months was predicted, resulting in an 621 increase in infection pressure for both *H. contortus* and *T. circumcincta* due to a corresponding increase in development rates. Development of T. circumcincta is 622 possible throughout the year in South West England. In temperate regions H. 623 contortus survival on pasture over winter is poor and there is very little development 624 of eggs deposited on pasture as temperatures fall below the development threshold. 625 However, *H. contortus* is able to survive the winter period as arrested larvae within 626 the host (Waller et al., 2004). The increase in temperatures predicted here for South 627 West England, extends the period where the development of *H. contortus* is possible 628 and could have significant short- and long-term impacts on the epidemiology of H. 629 contortus in temperate regions. In the short-term, the increase in infection pressure 630 631 throughout the year will result in year-round transmission. In the long-term, H.

632 contortus may adapt in response to the reduced selection pressure for arrested development (hypobiosis) in the host, potentially resulting in a decreased propensity 633 to arrest. Using a series of mathematical models, Dobson and Hudson (1992) 634 635 showed that hypobiosis decreases the basic reproductive rate (R<sub>0</sub>) of trichostrongylid nematodes. Gaba and Groubière (2008) built on the work of Dobson and Hudson 636 and further demonstrated the potentially destabilising effect of hypobiosis on GIN 637 population dynamics as the mortality rate of the free-living stages is decreased. 638 Therefore, hypobiosis would not be favoured when climatic conditions render it 639 640 unnecessary.

641 The predicted increase in the availability of *H. contortus* and *T. circumcincta* L3 on pasture during the spring under the RCP8.5 scenario is a concern as this coincides 642 with peak lambing/kidding and peak parturition in many wild ruminants. Therefore, 643 644 naïve individuals may experience a much greater challenge early in the grazing season under this scenario of climate change. Ewes experiencing a breakdown in 645 immunity to gastrointestinal nematodes during pregnancy and lactation (Houdijk et 646 al., 2001) will also experience a greater challenge. For example, an increase in H. 647 contortus infection pressure during the spring could result in more acute 648 649 haemonchosis in naïve individuals and increased pasture contamination early in the grazing season. These effects may be magnified by management and host factors. 650 The current model considers a scenario of constant pasture contamination. 651 However, pasture contamination may increase during the spring reproductive period 652 due to the periparturient rise (PPR) in faecal egg counts observed in reproducing 653 animals, which is due to the maturation of hypobiotic larvae resulting from a complex 654 of factors thought to result in a reduction in immunity during pregnancy and lactation 655 (Falzon et al., 2013; Gibbs and Barger, 1986; Houdijk et al., 2001). 656

A decrease in infection pressure is predicted during the summer months for all 657 species tested as a result of a trade-off between increased development and 658 increased mortality rates. As discussed for *O. ostertagi*, reduced exposure to L3 may 659 impact on the development of immunity. However, these reductions may be 660 dampened by increasing worm burdens in the host and therefore increased pasture 661 contamination throughout the grazing season. Furthermore, there is potential for 662 663 parasite adaptation in response to decreased transmission and the impact this has on host immunity. For example, nematode fecundity may be negatively associated 664 665 with host immune response as suggested by the negative correlation between adult *T. circumcincta* length and immune response (mucosal and serum IgA against L3) 666 and L4), and the positive correlation between worm length and number of eggs in 667 (nematode) utero in artificially infected lambs (Stear et al., 1995). Integrating the 668 GLOWORM-FL framework with models of the parasitic stages, host immunity (e.g. 669 Grenfell et al., 1987) and parasite adaptation will allow the impact of changes in the 670 seasonal exposure to L3 on the potential pathogenesis of infection and subsequent 671 population dynamics of parasites on pasture to be explored. 672

Validation was successful for all three species tested, not only validating the model
structure but also demonstrating that gaps in parameter estimates can be addressed
using data from other species, that parameter estimates derived from laboratory
observations perform well under conditions experienced in the field, and that data
obtained from the nearest weather station can be used in the absence of local
meteorological observations.

In some cases, linear regression of observations against model predictions was
significant, but the slope of the regression was marginally different from 1, indicating
systematic bias in the output. However, in most cases the error was within the range

expected from factors such as trait variation (Troell et al., 2006; van Dijk and
Morgan, 2010), measurement error (Persson, 1974b) and uncertainty arising from
model structure.

Simulations using mean daily temperature data outperformed those using minimum 685 and maximum data. Minimum and maximum daily temperature data were used to 686 test whether fluctuations above and below the development thresholds were 687 important in predicting the population dynamics free-living GINs. At certain times of 688 year the mean temperature may be above the threshold for development, but if the 689 minimum temperature falls below the threshold there is potential to over-predict the 690 691 development rate using only mean temperatures. Conversely, if the mean temperature is below the threshold but the maximum falls above the threshold, then 692 models may fail to predict development at all. It was therefore surprising that 693 allowing temperatures to fluctuate between the minimum and maximum daily values 694 did not improve model performance. 695

696 Discrepancies between meteorological observations and microclimatic conditions may account for the superior performance of simulations using mean daily air 697 temperatures. Recent studies have demonstrated the importance of microclimatic 698 factors in determining GIN abundance under controlled conditions (Khadijah et al., 699 2013b; O'Connor et al., 2008; Wang et al., 2014). In the field, temperature and 700 701 moisture fluctuations in faeces may be buffered by the soil beneath and surrounding herbage. This buffering effect may also explain discrepancies between model 702 703 predictions and observations e.g. the model underestimated the time to development of H. contortus L3 observed by Rose (1963) during April and October regardless of 704 whether minimum-maximum or mean temperatures were used. Soil temperature 705 706 data were not available for use in these validation exercises but may better reflect

the microclimate around faeces. Where possible, further validation should alsoinclude soil temperature.

The model was validated using observations made in a temperate region, with 709 temperatures ranging between - 4°C and 25°C. However, temperatures of up to 710 39.2°C were predicted under the RCP8.5 scenario and therefore some simulations 711 projected beyond the range of the conditions used for validation. H. contortus and T. 712 713 *circumcincta* simulations using climatic data for the RCP8.5 scenario showed that high temperatures may result in counter-intuitive decreases in the availability of L3 714 due to a trade-off between increased mortality and development rates. Uncertainty in 715 716 climate change simulations due to projections outside of the range of observed data could be reduced by repeating validation using data from regions with current 717 climatic conditions similar to those predicted under the chosen climate change 718 scenario. 719

720 The development of parsimonious mechanistic models is inevitably a compromise 721 between biological realism, complexity and the availability of data for parameter estimation. Here we have used observations on mortality of L3 in water to estimate 722 mortality rates for L3 in soil. Although these estimates are similar to published 723 observations of L3 mortality in soil (van Dijk and Morgan, 2011) and the model was 724 able to predict the survival of L3 on pasture in a temperate region well, site-specific 725 and temporal variations in soil conditions such as moisture content, pH and the 726 presence of nematophagous fungi may affect the observed mortality rates and 727 increase model uncertainty. Variations in faecal moisture and structural integrity of 728 faeces in the field may also affect the population dynamics of GINs. Diarrhoea is 729 commonly associated with infection by GINs such as *T. circumcincta* and *O.* 730 731 ostertagi but can also be attributed to a range of other causative agents such as diet and other gastrointestinal infections. As such it is difficult to characterise faecal
consistency for inclusion in mechanistic models, but this potential source of variation
should be noted, especially in the context of development success and horizontal
migration of L3 between faeces and pasture.

### 736 5. Conclusion

737 A general model framework was developed to simulate the climate-driven population dynamics of the free-living stages of trichostrongylid GINs. Simulations using 738 historical and future climatic data predicted significant changes in seasonal and 739 annual infection pressure in the absence of host management, including a surprising 740 decrease in infection pressure for O. ostertagi. Integration with management data, 741 host behaviour and models developed to simulate the parasitic stages of these 742 species, will enable the evaluation of GIN control options under a range of climate 743 scenarios to identify long-term sustainable strategies. 744

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Figure 1. Conceptual diagram of the GLOWORM-FL model framework. Parameter(lower case) and state variable (upper case) definitions are given in Table 1.

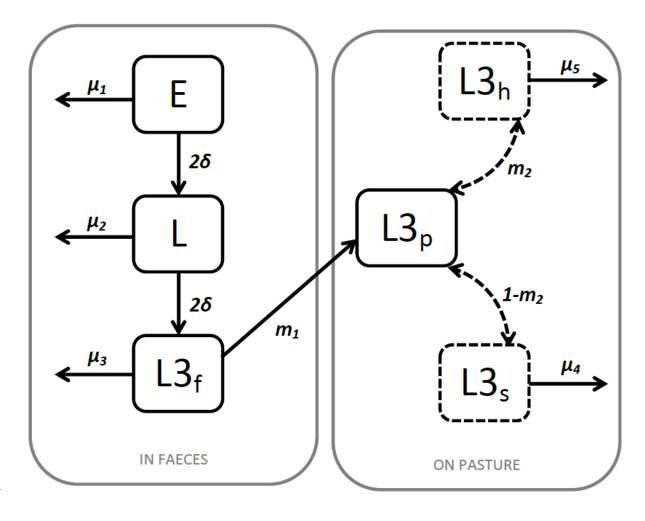
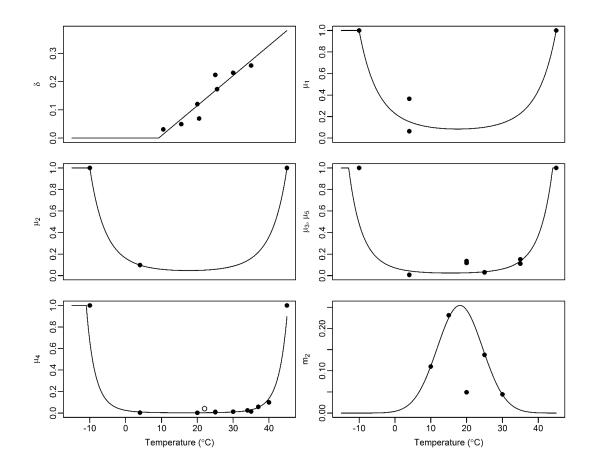


Figure 2. Estimates of temperature-dependent life-history parameters for H. 1047 contortus (lines) based on analysis of data in the literature (closed circles). 1048 Parameter definitions are given in Table 1. Statistical output for linear models is 1049 provided in Table 2. Mortality of L3 in soil ( $\mu_4$ ) was estimated from observations of L3 1050 mortality in water. A point estimate of L3 mortality in desiccated soil at 20-24°C (van 1051 Dijk and Morgan, 2011) is superimposed (open circle) for comparison. Data were not 1052 available to estimate the mortality of L3 on herbage ( $\mu_5$ ), which was therefore 1053 estimated using the mortality rate of L3 in faeces ( $\mu_3$ ). The data point at 20 degrees 1054 was omitted from analysis of the vertical migration parameter  $(m_2)$  but is shown here. 1055 A minimum threshold for development of 9.17°C is predicted. 1056



1057

Figure 3. Estimates of temperature-dependent life-history parameters for T. 1060 circumcincta (lines) based on analysis of data provided in the literature (closed 1061 circles). Parameter definitions are given in Table 1. Statistical output for linear 1062 models is provided in Table 2. Mortality of L3 in soil ( $\mu_4$ ) was estimated from 1063 observations of L3 mortality in water. A point estimate of L3 mortality in desiccated 1064 soil at 20-24°C (van Dijk and Morgan, 2011) is superimposed (open circle) for 1065 comparison. Data points are not shown for the mortality rates of L3 in faeces ( $\mu_3$ ) 1066 and on herbage ( $\mu_5$ ) as no data were available to directly estimate these parameters. 1067 These rates were therefore estimated from the mortality rate of L3 in soil ( $\mu_4$ ). A 1068 minimum threshold for development of 4.46°C is predicted. The vertical migration 1069 1070 parameter is as shown in Figure 2.

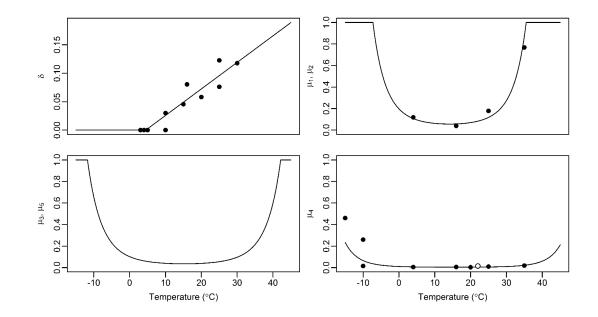
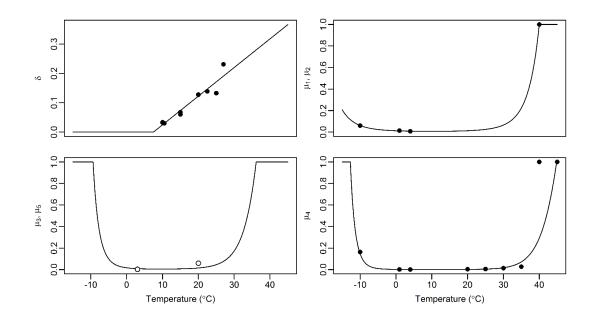
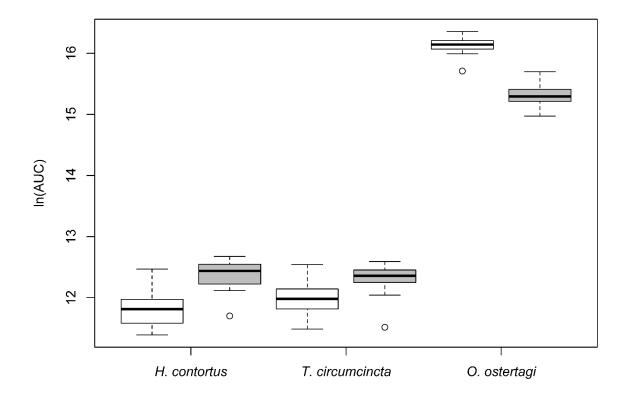


Figure 4. Estimates of temperature-dependent life-history parameters for O. ostertagi 1073 (lines) based on analysis of data provided in the literature (closed circles). Parameter 1074 definitions are given in Table 1. Statistical output for linear models is provided in 1075 Table 2. Data points are not shown for the mortality rates of L3 in faeces ( $\mu_3$ ) and on 1076 herbage ( $\mu_5$ ) as no data were available to directly estimate these parameters. These 1077 rates were therefore estimated from the mortality rate of L3 in soil ( $\mu_4$ ). Estimates of 1078 mortality in faeces ( $\mu_3$ ) based on analysis of observations on Cooperia oncophora 1079 1080 and O. ostertagi mixed infections (open circles) are superimposed for comparison 1081 (Persson, 1974a). A minimum threshold for development of 7.44°C is predicted. The vertical migration parameter is as shown in Figure 2. 1082



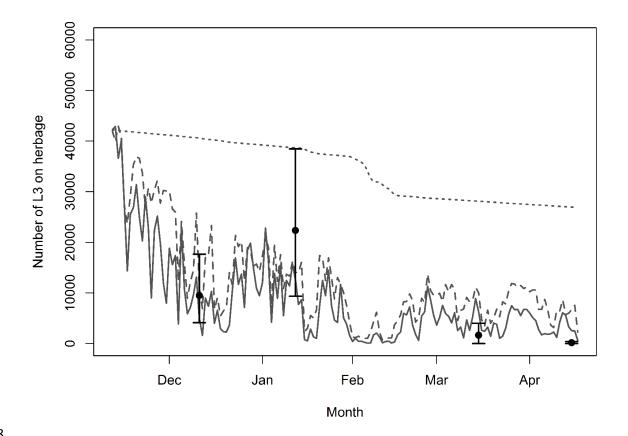
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Figure 5. Estimated annual AUC (Area Under the Curve) of the predicted numbers of L3 on pasture for *H. contortus, T. circumcincta* and *O. ostertagi* when using historical climatic data for the period 1969-1999 (white) and climatic data based on the RCP8.5 high emissions climate change scenario for the period 2070-2100 (grey).



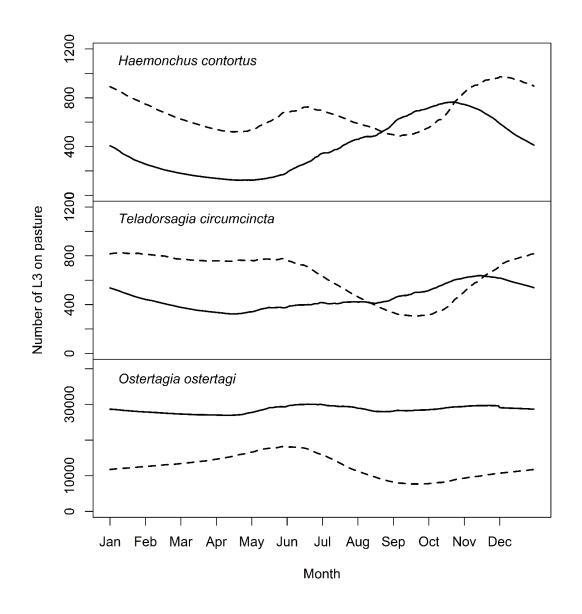
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Figure 6. The number of H. contortus L3 on herbage (L3 kg DM-1) predicted using the GLOWORM-FL model and mean daily air temperature (dashed line) or fluctuating between the daily minimum and maximum temperature (dotted line) and the number (L3 kg DM-1) observed by Wilkie et al. (submitted; points and error bars show mean and 95% confidence interval). Predicted numbers of L3 on herbage using a single mortality rate for L3 on pasture and no vertical migration (dotted line; Smith, 1990) are superimposed for comparison.





1100 Figure 7. The number of L3 on pasture (soil and herbage combined) predicted for Haemonchus contortus (top panel), Teladorsagia circumcincta (middle panel) and 1101 Ostertagia ostertagi (bottom panel) when using historical climatic data for the period 1102 1969-1999 (solid line) and climatic data based on the RCP8.5 high emissions climate 1103 change scenario for the period 2070-2100 (broken line). Data shown are the 1104 1105 disaggregated annual means from the thirty year time-series. The first year of each time series was discarded. One hundred new eggs were input daily. Therefore no 1106 assumptions were made regarding management or intensity of infection in the host 1107 1108 and the predicted dynamics are entirely climate-driven.





State	Definition	Units			
variable/					
Parameter					
Ε	Eggs	-			
L	First stage (L1) and second stage (L2) larvae	-			
L3 <sub>f</sub>	Third stage infective larvae (L3) in faeces	-			
L3p	Total L3 on pasture (soil and herbage	-			
	combined)				
L3s	L3 in soil	-			
L3 <sub>h</sub>	L3 on herbage	-			
σ	Development rate from egg to L3	Instantaneous daily			
		rate			
$\mu_1$	Egg mortality rate	Instantaneous daily			
		rate			
$\mu_2$	L1 and L2 mortality rate	Instantaneous daily			
		rate			
$\mu_3$	L3 mortality rate in faeces	Instantaneous daily			
		rate			
$\mu_4$	L3 mortality rate in soil	Instantaneous daily			
		rate			
$\mu_5$	L3 mortality rate on herbage	Instantaneous daily			
		rate			
$m_1$	Horizontal migration (translation) of L3 onto	Instantaneous daily			
-	pasture	rate			
$m_2$	Proportion of total pasture L3 on herbage	Proportion			
C	Development success correction factor	Proportion			

1112 Table 1. State variable and parameter definitions

- 1115 Table 2. Parameter estimates derived from data in the literature and additional
- 1116 laboratory experiments. ANOVA results are shown for the linear models fitted to
- data from the literature to estimate temperature-dependent rates (see text).

Parameter	<b>Species</b> <sup>a</sup>	Estimate <sup>b</sup>	Data source
δ	Hc	-0.09746 + 0.01063T (F <sub>1,6</sub> =43.5, p<0.001, R <sup>2</sup> =0.88, R <sup>2</sup> <sub>adj</sub> =0.86)	Hsu and Levine, 1977; Rose, 1963
	Тс	-0.02085 + 0.00467T (F <sub>1,10</sub> =76.57, p<0.001, R <sup>2</sup> =0.88, R <sup>2</sup> <sub>adj</sub> =0.87)	Crofton and Whitlock, 1965; Crofton, 1965; Pandey et al., 1989; Salih and Grainger, 1982; Young et al., 1980a
	Оо	-0.07258 + 0.00976T (F <sub>1,8</sub> =76.14, p<0.001, R <sup>2</sup> =0.90, R <sup>2</sup> <sub>adj</sub> =0.89)	Pandey, 1972a; Rose, 1961; Young et al., 1980b
$\mu_1$	Нс	$\begin{array}{l} \exp(-1.47135-0.11444T\\ +\ 0.00327T^2)\\ (F_{2,3}\!=\!4.65,p\!=\!0.12,R^2\!=\!0.76,\\ R^2_{adj}\!=\!0.59) \end{array}$	Todd et al., 1976a
	Тс	$exp(-1.62026 - 0.17771T + 0.00629T^2)$ (F <sub>2,2</sub> =6.27, p=0.27, R <sup>2</sup> =0.93, R <sup>2</sup> adj=0.78)	Pandey et al., 1993, 1989
	Оо	$exp(-4.38278 - 0.10640T + 0.00540T^{2})$ (F <sub>2,1</sub> =6.27, p=0.06, R <sup>2</sup> =0.99, R <sup>2</sup> adj=0.99)	Pandey, 1972
μ <sub>2</sub>	Нс	$\begin{array}{l} \exp(-1.82300-0.14180T\\ +\ 0.00405T^2)\\ (F_{2,1}=1.723^{31},p{<}0.001,R^2{=}1,\\ R^2{}_{adj}{=}1)^c \end{array}$	Todd et al., 1976a
	Тс, Оо	As $\mu_1$ above	-
$\mu_3$	Нс	$exp(-2.63080 - 0.14407T + 0.00463T^2)$ (F <sub>2,9</sub> =8.48, p=0.008, R <sup>2</sup> =0.65, R <sup>2</sup> <sub>adj</sub> =0.58)	Todd et al., 1976a, 1976b

	Tc, Oo	$10 * \mu_4$	Pandey, 1972; Persson, 1974a		
$\mu_4$	Нс	$exp(-3.68423 - 0.25346T + 0.00740T^2)$ (F <sub>2,8</sub> =50.76, p<0.001, R <sup>2</sup> =0.93,	Jehan and Gupta, 1974; Todd et al., 1976b		
	Тс	$\begin{array}{l} R^{2}_{adj} = 0.91) \\ \exp(-4.58817 - 0.13996T \\ + 0.00461T^{2}) \\ (F_{2,12} = 43.55,  p < 0.001,  R^{2} = 0.88, \\ R^{2}_{adj} = 0.86) \end{array}$	Gruner and Suryahadi, 1993; Jasmer et al., 1987; Pandey et al., 1993; Rossanigo and Gruner, 1996		
	Оо	$\begin{split} \exp(-6.388 - 0.2681T + 0.01633T^2 \\ &- 0.00016T^3) \\ (F_{3,5} = 28.81,  p = 0.001,  R^2 = 0.95, \\ R^2_{adj} = 0.91) \end{split}$	Pandey, 1972		
$\mu_5$	Нс, Тс, Оо	As $\mu_3$ above	Grenfell et al., 1986; van Dijk and Morgan, 2011; van Dijk et al., 2009		
<i>m</i> <sub>1</sub>	Нс	$\begin{cases} 0.25, & P \ge 2\\ 0, & P < 2 \text{ AND } \sum_{i=-4}^{t} P_i / E_i < 1\\ 0.051, & P < 2 \text{ AND } \sum_{i=-4}^{t} P_i / E_i \ge 1 \end{cases}$	Present study; O'Connor et al., 2008; Wang et al., 2014		
	Тс	$\begin{cases} 0.21, & P \ge 2\\ 0, & P < 2 \text{ AND } \sum_{i=-7}^{t} P_i / E_i < 1\\ 0.025, & P < 2 \text{ AND } \sum_{i=-7}^{t} P_i / E_i \ge 1 \end{cases}$	Present study; O'Connor et al., 2008; Wang et al., 2014		
	Оо	$\begin{cases} 0.06, & P \ge 2 \\ 0, & P < 2 \end{cases}$	Grønvold and Høgh-Schmidt, 1989		

$$\begin{array}{ccc} C & Hc \\ \begin{cases} 0.1, & \sum_{i=4}^{t} P_i/E_i < 1 \\ 1, & \sum_{i=4}^{t} P_i/E_i \ge 1 \\ \end{array} \\ Tc \\ \begin{cases} 0.1, & \sum_{i=7}^{t} P_i/E_i < 1 \\ 1, & \sum_{i=7}^{t} P_i/E_i \ge 1 \end{array} \end{array}$$

- <sup>a</sup> Hc = Haemonchus contortus, Tc = Teladorsagia circumcincta, Oo = Ostertagia
   ostertagi
- <sup>b</sup> T = temperature (°C), P = total daily precipitation (mm), E = total daily
   evapotranspiration (mm)
- <sup>c</sup> Note that the statistical significance here is an artefact of overfitting

Table 3. Validation of simulations using data provided in the literature. Models are considered a good fit if regression through the origin is significant and the slope is not significantly different from 1. The error and R<sup>2</sup> are used to compare competing models.

Data source	Specie s <sup>a</sup>	Model compone nt tested	Temperat ure data used	Error (residu al sum of square s)	Linear regressi on	R <sup>2</sup> (R <sup>2</sup> adjust ed)	Slop e (95 % Cl <sup>b</sup> )
Rose, 1963	Hc	Faeces (eq. 1-3; D50)	Min - Max	1093.5	F <sub>1,6</sub> = 12.21, p=0.013	0.67 (0.62)	2.18 (0.9 3 – 3.43 )
Rose, 1963	Нс	Faeces (eq. 1-3; D50)	Mean	1097.5	F <sub>1,6</sub> = 9.561, p=0.021	0.61 (0.55)	, 1.93 (0.6 8 – 3.17 )
Rose, 1963	Hc	Faeces (eq. 1-3; developm ent success)	Min - Max	31.28	F <sub>1,11</sub> = 28.53, p<0.001	0.72 (0.70)	, 1.28 (0.8 0 – 1.76 )
Rose, 1963	Hc	Faeces (eq. 1-3; developm ent success)	Mean	105.74	F <sub>1,11</sub> = 11.91, p=0.005	0.52 (0.48)	) (0.49 (0.2 0 – 0.77
Wilkie et al. submitte d	Hc	Pasture (eq. 4)	Min - Max	1.83 x 10 <sup>8</sup>	F <sub>1,3</sub> = 6.78, p=0.080	0.69 (0.59)	) 0.94 (0.2 2- 1.67
Wilkie et al. submitte d	Нс	Pasture (eq. 4)	Mean	1.43 x 10 <sup>8</sup>	F <sub>1,3</sub> = 16.72, p=0.026	0.85 (0.80)	) 1.48 (0.7 6- 2.20 )
Rossani go and Gruner, 1995	Тс	Faeces (eq. 1-3; developm ent success)	Constant	1679.1 0	F <sub>1,9</sub> = 86.9, p<0.001	0.91 (0.90)	1.54 (1.2 1 - 1.86 )
Wilkie et al. submitte d	Тс	Pasture (eq. 4)	Min - Max	1.03 x 10 <sup>10</sup>	F <sub>1,3</sub> = 9.91, p=0.051	0.77 (0.69)	1.26 (0.4 6- 2.07 )

Wilkie et al. submitte d	Тс	Pasture (eq. 4)	Mean	1.17 x 10 <sup>10</sup>	F <sub>1,3</sub> = 17.84, p=0.024	0.86 (0.81)	1.75 (0.9 2- 2.58 )
Rossani go and Gruner, 1995	Oo	Faeces (eq. 1-3; developm ent success)	Constant	2050.2 2	F <sub>1,5</sub> = 36.86, p=0.002	0.88 (0.86)	0.84 (0.5 7- 1.12 )
Rose, 1961	00	Faeces ( (eq. 1-3; D50)	Min – Max		F <sub>1,11</sub> = 26.9, p<0.001	0.71 (0.68)	, 21.0 3 (12. 9 – 29.1 )
Rose, 1961	00	Faeces (eq. 1-3; D50)	Mean		F <sub>1,11</sub> = 98.51, p<0.001	0.90 (0.89)	0.62 (0.5 0 – 0.75 )
Wilkie et al. submitte d	Нс	Smith (1990)	Mean	2.66 x 10 <sup>9</sup>	F <sub>1,3</sub> = 4.73, p=0.12	0.61 (0.48)	0.28 (0.0 2- 0.53 )

<sup>a</sup> Hc = Haemonchus contortus, Tc = Teladorsagia circumcincta, Oo = Ostertagia 

ostertagi <sup>b</sup> 95% confidence intervals were estimated as 2 x the standard error of the slope 

coefficient.