- Rise and fall of island butterfly diversity. Understanding population diversification and
   extinction in a highly diverse Archipelago.
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#### 50 ABSTRACT

51 Aim

We described fine-scale diversity patterns of the entire butterfly fauna occurring on the Tuscan Archipelago. By assessing the traits associated with population diversification, haplotype uniqueness and extinction, we aimed to identify the factors determining the origin and the maintenance of genetic diversity and population vulnerability to environmental changes.

## 57 Location

58 Tuscan Archipelago, Sardinia, Tuscany (Italy) and Corsica (France).

## 59 Methods

We built a mtDNA dataset (1303 COI sequences) for the 52 butterfly species reported in the Archipelago also including specimens from neighbouring areas and compiled data on 12 species traits and on the apparent extinction of butterfly species from the main islands. We calculated different indices indexes of genetic differentiation, and using stepwise phylogenetic regressions we evaluated the relationships between these indices and species traits. Finally, we inferred which traits are associated with apparent extinction of species on individual islands using phylogenetic regression.

## 67 **Results**

68 The overall spatial pattern of genetic diversity corresponded with the proximity of the areas,

69 but strong contrasts were also identified between geographically close areas. Together with

the island endemics, several common and widespread species revealed a high genetic

diversification among islands and mainland. Phylogenetic regressions showed that smallersized, more specialized species, with a preference for drier regions, displayed greater genetic
structure and/or haplotype uniqueness. Capraia has particularly suffered loss of diversity,
which significantly affected species with shorter flight periods.

# 75 Main conclusions

Tuscan island butterflies are characterized by strong genetic contrasts and species differ in their contribution to the overall genetic diversity. By ranking the species for their contribution to genetic diversity and by identifying the traits linked to the emergence and maintenance of diversity, we provided a valuable tool to prioritize the populations as targets for monitoring and conservation actions. The large dataset we provided represents a resource to test biogeographical hypotheses.

#### 82 INTRODUCTION

The worldwide biodiversity crisis calls for the identification, prioritization and protection of biodiversity hotspots, and understanding how biodiversity is generated and lost determines the success of this endeavour (Dirzo et al., 2014; Venter et al., 2014). The wealth of information now available in the "big data era" greatly facilitates these efforts, especially the ability to work at an unprecedented resolution (Hampton et al., 2013). For conservation biogeography this includes the increased availability of molecular, occurrence and trait data for various organisms (Ladle & Whittaker, 2011; Fernández-Palacios et al., 2015).

90 Islands host a disproportionate fraction of global biodiversity, often characterized by distinctive faunas (Whittaker & Fernández-Palacios, 2007), and therefore represent a model 91 system to better understand general patterns in biogeography. Insular populations are typically 92 exposed to high risks of extinction since restricted habitats can easily become unsuitable 93 because of stochastic events or human disturbances acting at both local (habitat 94 fragmentation, alteration, invasion by alien species) and global scales (e.g. climate change) 95 (Fordman et al., 2010). Understanding the factors that drive the emergence, maintenance and 96 loss of island biodiversity is crucial for planning and implementing evidence-based 97 98 conservation prioritization and protection measures (Ladle & Whittaker, 2011).

99 Remote oceanic islands experience infrequent colonization events followed by genetic drift 100 and adaptive radiation (Rosindell & Phillimore, 2011). In contrast, biotas on less isolated 101 islands are mostly assembled following frequent events of colonization and extinction, which 102 produce nested communities of the source species with few endemics (Whittaker & 103 Fernández-Palacios, 2007). However, a pace of extinction and colonization on islands 104 (turnover) slower than the dynamics occurring in the surrounding regions can generate 105 intraspecific diversification and relictuality, which creates community distinctiveness among

islands and from neighbouring mainlands (Masini et al., 2008; Dapporto et al., 2012). Such
diversity is usually represented by cryptic species or genetic lineages that can be detected only
after in-depth molecular and/or morphologic analyses (Hernández-Roldán et al., 2016; Vodă
et al., 2015a, 2016).

The occurrence of endemic and relict populations can be the result of deterministic processes. 110 largely affected by species characteristics. According to a widely accepted paradigm of island 111 112 conservation biogeography, genetic diversification and extinction probability of island populations are inversely related to their degree of mobility and generalism (Burney & 113 Brumfield, 2009; Dennis et al., 2011; Salisbury et al., 2012; Dawson et al., 2014 but see also 114 Kobayashi & Sota, 2016 for different patterns). This hypothesis has profound implications for 115 conservation ecology because populations with unique genetic fingerprints are clearly 116 117 irreplaceable but potentially suffer from high extinction probability (Ricklefs, 2009). Nevertheless, this has been rarely tested (Burney & Brumfield, 2009; Vodă et al., 2016), 118 119 probably because of the absence of integrative datasets combining molecular and occurrence 120 data with species traits for entire species-rich taxonomic groups and from a large geographic 121 area.

Here we tested if species having low dispersal capacities and restricted ecological 122 requirements tend to colonize islands at a slower rate, experience reduced gene flow and thus 123 have i) higher diversification rates and ii) higher likelihood of local extinction compared with 124 more mobile and generalist species. We designed an integrated framework (Fig. 1) based on a 125 dataset comprising: i) a revised distribution list of the 52 butterfly species reported for the 126 127 Tuscan Archipelago and the apparent disappearance of several species in the last 115 years, ii) 1303 cytochrome c oxidase subunit I (COI) sequences (658 bp) for all the species, including 128 populations from the surrounding areas (Sardinia, Corsica and Tuscany), and iii) 12 species 129

traits related to mobility, phenology, trophic generalism and climatic preferences. The 130 analysis of this dataset allowed us to: i) describe the patterns of population diversification 131 (genetic diversity and haplotype uniqueness) among islands and neighbouring areas, ii) 132 identify the functional traits that are correlated with the emergence of diversification and 133 species disappearance on islands. Finally, iii) we ranked the species according to their 134 contribution to the overall island diversity. Our integrated framework (Fig. 1) allowed us to 135 provide evidence-based guidelines for butterfly conservation in a well-known Mediterranean 136 diversity hotspot (Dennis et al., 2008). 137

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139 METHODS

## 140 Study area and occurrence data

141 We carried out annual surveys of the butterflies of the Tuscan islands and the neighbouring areas of Sardinia, Corsica and Tuscany between 2000 and 2015 (Fig. 1a). We sampled the 142 main biotopes from early spring to late autumn, with different tools (mostly Malaise traps and 143 insect nets). We compared occurrence data from our surveys with published records dating 144 back to 1900 (Appendix S1). Sampling was not done constantly throughout the 15 years but 145 our extensive collections represent the most intensive sampling effort done on the butterflies 146 of these islands, therefore it is reasonable to assume that a species reported in the past but not 147 148 during the last decade has either become extinct or has significantly declined. We restricted 149 our analysis of species disappearance to the three largest islands: Elba, Capraia and Giglio, for which sufficient historical data were available. 150

#### 152 *Genetic assessment*

Using standard sequencing procedures (deWaard et al., 2008), we obtained 1303 COI sequences for specimens belonging to the 52 butterfly species reported for the Tuscan islands between 1900 and 2015, including specimens from Sardinia, Corsica, Argentario and four mainland areas (Fig. 1a). We also used 2940 sequences from other parts of Europe and Asia. Information about specimens and sample size for each population is available in the "dataset.csv" file (Appendix S2). All sequences were aligned in Geneious 6.0.6 (www.geneious.com).

160 The butterfly species currently recognized by taxonomists, and for which trait data are available, show varying levels of intraspecific genetic divergence, including potential cases of 161 cryptic taxa (Dincă et al., 2015). We considered as separate units most species recognized by 162 the widely accepted checklist of the Fauna Europaea Project (Karsholt & Van Nieukerken, 163 2013; www.faunaeur.org) and also considered as single units taxa displaying a minimum COI 164 165 p-distance lower than 3%. In butterflies, this threshold separates more than 90% of the recognized species (Hebert et al., 2003) and a recent study on Sicilian islands confirmed this 166 observation (Vodă et al., 2016). We also repeated all the analyses without setting any distance 167 168 threshold, and considered as single units only the species recognized by Fauna Europaea.

We calculated the genetic uncorrected p-distances among all sequenced specimens for each species by using the function "dist.dna" of the "ape" R package. We preferred p-distance to tree-based genetic distances because, at the intraspecific level, coalescence has not taken place and distances calculated on branch lengths of bifurcating phylogenetic trees do not properly reflect the reticulated evolutionary processes (e.g. Posada et al., 2001). Moreover, recent reviews indicate that p-distances are the best option in the analysis of COI data compared to other indices (Srivathsan and Meier, 2012). Based on p-distances, we obtained two measures for population differentiation: Dst and Gst (Nei, 1987). We also calculated the
pairwise Gst among all pairs of populations for each species (see Appendix S1 for a
description of the indices).

Based on the Gst pairwise matrices we produced for each species the mean Gst matrix, 179 representing the degree of differentiation among areas based on all species. A Principal 180 181 Coordinates Analysis (PCoA) was applied to this matrix to obtain the overall diversity pattern 182 among areas. Subsequently, we aligned this configuration with the geographic location of the areas by using the "procrustes" analysis from the "vegan" R package and tested the correlation 183 184 between the PCoA configurations of Gst and the spatial location by using the vegan function "protest". To visualize the pattern of similarity among islands in the geographic space, we 185 projected the PCoA configuration in RGB space using the R package "recluster" (Dapporto 186 et al., 2014). The colour resemblance of the resulting configuration is directly proportional to 187 the genetic similarity among the communities. 188

Four species (*Leptidea sinapis*, *Aglais urticae*, *Nymphalis polychloros*, *Argynnis paphia*) recorded for Tuscan islands in the last century but not during our surveys have also been analysed by computing the Dst and Gst between Tuscany, Sardinia and Corsica to estimate the fraction of diversity that has presumably disappeared.

193 Island haplotype uniqueness for each species was calculated as:

194 
$$Uni = \prod_{h=1}^{n} \frac{\min(D_{h,m})}{occ}$$

where h are the haplotypes found on the Tuscan islands and not recorded on mainland areas of Europe and Asia,  $min(D_{h,m})$  is the p-distance between the h<sup>th</sup> haplotype and the genetically closest haplotype from mainland, and occ is the number of islands (Tuscan islands, Sardinia and Corsica) where the h<sup>th</sup> haplotype has been found. Uniqueness for a given species is higher when: 1) there are many insular endemic haplotypes, and/or 2) they show high divergence
with respect to the closest mainland haplotype and/or 3) they are found in fewer islands.
Uniqueness values (Uni) were calculated for each species occurring on any island except for
the island endemics that do not have mainland populations (*Hipparchia aristaeus* and *H. neomiris*).

To assess the importance of each species for the diversity of the Tuscan Archipelago we ranked species according to Gst, Dst and Uni and subsequently summed the ranks. The patterns of genetic variation were also analysed by inferring maximum parsimony haplotype networks using the program TCS 1.21, with a 95% connection limit for all species except *H. neomiris* (94% connection limit) and *Coenonympha corinna* for which we used a fixed connection limit of 23 steps (Clement et al., 2000).

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211 Species traits and phylogenetic regressions

212 Review studies suggest that only morphological, physiological or phenological features that 213 can be measured on individual organisms and without reference to the environment or any other level of organization should be considered as functional traits (Violle et al., 2007; 214 Moretti et al., 2016). Moretti et al. (2016) identified a series of 29 functional traits to cover 215 216 the primary functions of invertebrates, divided into five major groups: morphology, feeding, life history, physiology and behaviour. Measuring the 29 traits on individuals for all the 217 studied species would have entailed a major long-term effort that was beyond the scope of 218 219 this study. Thus, based on literature data and personal observations, we assessed 12 species traits representing four of the five groups (excluding behaviour) identified by Moretti et al. 220 221 (2016): but encompassing morphology, feeding, life history and physiology. For each trait we formulated functional hypotheses (Table 1): a) Trophic generalism (feeding), was identified 222

223 as i) the number of host plant genera reported in the literature; b) Mobility was assessed with 224 a morphological traits represented by ii) wingspan and obtained as the average between 225 minimum and maximum size reported in the literature; c) Phenology (life history traits) was identified as iii) the length of the flight period, iv) the first month when adults emerge, v) the 226 last month when adults fly, and vi) voltinism. Finally, d) climatic preference and tolerance 227 (physiology) were assessed by proxy variables for eco-physiological responses to 228 environmental conditions. These variables have been calculated by Schweiger et al. (2014) by 229 230 modelling species distribution in Europe based on occurrence data, and then by averaging temperature and precipitation among the spatial cells where each species is predicted to occur. 231 Although these indices cannot be considered as strict functional traits since they are obtained 232 233 from the geographic distribution of the species (Violle et al., 2007), they are widely recognized as proxies for the traits responsible for eco-physiological responses to climate (e.g. 234 Devictor et al., 2012). The variables we included are: vii) mean annual temperature viii) and 235 precipitation, ix) standard deviations of the temperature mean and x) and precipitation, xi) 236 upper 95% confidence limit of temperature mean, and xii) lower 95% confidence limit of 237 238 precipitation mean.

239 Butterfly traits are usually highly inter-correlated but they can be conveniently reduced to factors by using ordination methods (Carnicer et al., 2013; Dapporto & Dennis, 2013). For 240 morphology, life history and physiology traits we applied a Principal Component Analysis 241 (PCA), using the R function "rda" and the components with eigenvalues higher than one have 242 been used as variables for successive analyses. Some literature sources did not report the 243 wingspan for all the studied species and we imputed the missing values by using the "mice" 244 function of the "mice" R package (see Appendix S2). The algorithm imputes an incomplete 245 variable by generating plausible values based on other variables in the data by Multivariate 246

247 Imputations by Chained Equations (MICE) (Van Buuren & Groothuis-Oudshoorn, 2011). The existence of a phylogenetic signal for the variables of each trait following the PCA 248 249 ordination and for Dst, Gst and Uni was tested with Pagel's lambda index by applying the "phylosig" R function of the "phytools" package. Gst, Dst and uniqueness have been mapped 250 onto the phylogenetic tree by using the "contMap" function of the "phytools" package. 251 252 Character mapping is accomplished by estimating states at internal nodes using maximum likelihood and then by interpolating the states along each edge (Ravell, 2013). The 253 254 relationships between the variables and Dst, Gst and Uni have been assessed using phylogenetic stepwise regressions. We also employed Pagel's lambda as a model for the 255 phylogenetic covariance of residuals and applied a two-way selection of variables based on 256 257 the Akaike Information Criterion (AIC) as implemented in the function "phylostep" of the package "phylolm". From the phylogenetic regressions we removed the species not recorded 258 for the Tuscan Archipelago during the study period since no DNA sequences were available. 259 We square-root transformed Dst and Uni to improve their normality and standardized the 260 values of the traits with zeta-scores to provide a balanced contribution to the phylogenetic 261 regression. 262

The importance of traits in explaining possible extinctions of butterflies on two islands (Elba and Capraia) has been assessed with a logistic phylogenetic Generalized Linear Model using the function "phyloglm" of the package "phylolm", in which species found during our surveys and species that have not been confirmed, represented the binary response variable, and the trait variables the predictors. Logistic GLM was not performed for Giglio as only three species disappeared on this island. "Phylolm" function was used to assess if species disappeared in at least one island showed higher values of Gst, Dst and Uni

As a reference phylogeny, we used the Maximum Likelihood (ML) phylogenetic tree based 271 on COI sequences for all the western Mediterranean butterflies, freely available in the 272 package "recluster" (Dapporto et al., 2013). The tree was inferred with topological constraints 273 at family and subfamily levels following the butterfly phylogeny (see Appendix S1 for 274 details). ML analyses were performed using RAxML BlackBox (Stamatakis et al. 2008). A 275 GTR+Gamma+I model was selected and node supports were assessed through 100 rapid 276 bootstrap replicates. Effect size for models has been evaluated by plots of observed vs fitted 277 278 values associated with Spearman rho correlation.

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## 280 RESULTS

Based on the taxonomy proposed by Fauna Europaea and by applying a 3% threshold of COI divergence, we identified 52 units among the taxa reported in literature for the Tuscan Islands (hereafter 'species', Table 2). During our surveys on these islands we recorded a total of 46 species. The comparison between observations during the last decade and literature data from 1900 to 2000 for Elba, Giglio and Capraia, identifies those species (see Table 2) that probably became extinct or strongly declined on these islands (respectively six, three and seven on Elba, Giglio and Capraia).

Dst was correlated with both Gst (Spearman rank test: rho 0.836, P<0.001) and Uni (rho 0.460, P<0.001), while Gst was not correlated to Uni (rho 0.192, P=0.213). Gst values showed an almost bimodal distribution (14 species with Gst<0.25 and 10 species with Gst>0.75, Table 2). Twenty-four species had haplotypes not recorded on the mainland and most of these species did not belong to endemic taxa (Table 2). Gst and Uni did not have a significant phylogenetic signal (lambda 0.218; P=0.314 and lambda <0.001; P=1.000, Fig. 2), while Dst

had a significant effect (lambda 0.334; P=0.034; Fig. 2). Ordering species by the sum of ranks
of the three indices (Dst, Gst and Uni) showed that, together with endemics, several common
and widespread species provided a large contribution to diversity. The first quartile of the top
ranking species comprised four species/groups with endemic elements in the Tuscan islands
(*C. corinna, Lasionmata megera/paramegaera, H. neomiris, Aglais urticae/ichnusa*) and ten
widespread species, most of them ubiquitous in Europe and not included in any protection list
(Table 1).

The wingspan measures reported in the four literature sources used were highly correlated 301 (Pearson R>0.9 for all pairs) and the PCA identified only one component with an eigenvalue 302 303 higher than one (Table 1, Figure S46). For life history and physiologic traits two components were considered (Table 1, Figure S47). The first phenological component was mainly linked 304 305 to the length of the flight period (voltinism, number of months when adults occur), while the 306 second was mainly linked to seasonality (first and last month of emergence). The first component for physiologic traits ordered species from those experiencing high temperatures 307 and low precipitation to those living in colder and wetter areas, while the second component 308 309 ordered species mostly according to their precipitation tolerance (Table 1, Figure S48). The six resulting variables showed a lower correlation among each other with Pearson correlation 310 values always lower than 0.400 (Table S1). Among the six resulting variables, the number of 311 host plants, phenology PC1, physiology PC1 and PC2 did not show a phylogenetic signal 312 (lambda 0.282; P=0.228; lambda<0.001; P=1.000; lambda 0.062; P=0.711; lambda<0.001; 313 P=1.000, respectively), while wingspan PC1 and phenology PC2 showed a significant effect 314 (lambda 1.187; P<0.001; lambda 1.083; P<0.001, respectively). 315

The overall spatial pattern of genetic variation based on Gst corresponded to the proximity of the areas but with a rather low level of correlation (protest correlation 0.592, P=0.015, Fig. 318 3a,b).

The AIC procedure for the stepwise phylogenetic regression for Gst selected a model with 319 four variables, but only mobility and trophic generalism had a significant effect (Table 3), 320 meaning that smaller-sized and more generalist species had a higher Gst. For Dst, four 321 322 variables entered the model - three were significant and showed that smaller-sized, more generalist species and those experiencing less annual precipitation had a higher variation 323 324 (Table 3). Two variables entered the Uni model showing that species living in drier areas significantly had higher haplotype uniqueness (Table 3), while trophic generalism entered the 325 model but without significant effect. Plots for observed vs fitted values of the three models 326 (Fig. S53) showed large residuals indicating that Gst, Dst and Uni are only weakly explained 327 by the measured traits (Spearman rho: Gst 0.537, Dst 0.430. Uni 0.368). 328

329 Species that had disappeared from at least one island showed significantly higher values of330 Dst, while no differences in Gst and Uni were found (Table 3).

According to the logistic phylogenetic GLM, species that disappeared from Capraia had lower values in PC1 for phenology, corresponding to shorter flight periods (Table 4). A plot for observed vs fitted values (Fig. S54) revealed a good fit for this analysis (Spearman rho 0.798). For Elba we found no significant effect explaining the disappearance of the six species (Table 4).

The analyses in which we used only the taxonomy from Fauna Europaea returned very similar results to the ones in which we used a 3% threshold for species identification (see Table S2-S5 and Figure S55 in Appendix S1).

## 340 DISCUSSION

The integration of an updated taxonomic list, species occurrence spanning across 115 years (1900-2015), mitochondrial DNA sequences and species traits, allowed us to characterize the butterfly diversity in the Tuscan Archipelago at an unprecedented resolution and to infer which species traits explain the rise and decline of butterfly diversity in this archipelago.

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## 346 An area of biogeographic contrasts

The Tuscan Archipelago is an insular hotspot for butterfly diversity and stands out among European islands for hosting far more endemics than would be expected based on their geography (e.g. area and isolation) (Dennis et al., 2008), resulting in the highest priority for butterfly conservation among circum-Italian islands (Dapporto & Dennis, 2008). Its unexpected level of endemicity and richness is a consequence of its intermediate location between Tuscany and the Sardo-Corsican region, generating a double filtering effect (Dapporto & Cini, 2007; Fattorini, 2009).

We found that a main determinant for the genetic make-up of populations is island location (see the protest analysis), the same as for the community composition at species level (Dapporto & Cini, 2007). Nevertheless, a model based only on a double filtering effect is not sufficient to explain the observed degree of genetic diversity. In fact, we identified strong contrasts between geographically close areas, such as between Montecristo (very similar to the Sardo-Corsican region) and Pianosa (more similar to Elba and the Italian Peninsula), which are separated by 30 km; between Capraia and Elba (same pattern as before, distance 33

km) and between Giglio and Argentario, separated by only 14 km (Fig. 3). Striking 361 divergence among populations from nearby areas is unexpected in butterflies characterized by 362 a high mobility resulting in a high capacity to track suitable environments (Wilson et al., 363 2010; Waters, 2011; Devictor et al., 2012). However, in the western Mediterranean (Tuscan 364 islands included), chequered distributions of sister species and genetic lineages are a common 365 phenomenon, probably due to the combination of several historical determinants and 366 contemporary ecological forces (e.g. the connection between land masses during the Last 367 368 Glacial Maximum, density-dependent phenomena, differences in climatic and environmental features, Vodă et al., 2015a,b, 2016). Accordingly, more than 30% of the examined species, 369 comprising both endemic and widespread taxa had a Gst value higher than 0.5. 370

Endemic taxa also have unexpected patterns of intraspecific genetic diversity among islands. 371 Coenonympha corinna/elbana is highly divergent among the three clades they form in: i) 372 373 Tuscany, Elba and Giannutri; ii) Capraia and Corsica; and iii) Sardinia. This pattern only partially supports the current taxonomic separation into two species or subspecies (C. 374 corinna: Sardinia, Corsica and Capraia; C. elbana: Elba, Tuscany and Giannutri) (Fig. 3c). 375 Hipparchia neomiris, an endemic species from Sardinia, Corsica and Elba, also displayed 376 notable intraspecific divergence. COI sequences from Elba and Corsica are differentiated by 377 378 at least 2% compared to conspecific individuals from Sardinia. By contrast, no genetic diversification was detected in the *Plebejus idas* group, since the endemic *P. bellieri* from 379 Sardinia and Corsica (recognized as a good species in Fauna Europea) shared COI barcodes 380 381 with the population from Elba, treated as a species in some works, and with the mainland populations, elsewhere reported as as P. idas or P. abetonicus (Balletto et al. 2015). 382

Phylogenetic regressions suggest that the species adapted to a dry climate, the small-sizedones and the host plant specialists showed a higher degree of island uniqueness and of genetic

diversification among islands and populations. According to our hypotheses, these species traits can facilitate the emergence and maintenance of these genetic contrasts since typical Mediterranean species thrive on islands, while species with reduced dispersal and poor colonization capabilities probably experience reduced gene flow.

Several species previously recorded on three Tuscan islands were not observed in the last 10 years, suggesting that they became extinct or declined considerably (Table 2). On Giglio only three species have not been recorded during the last decade, but they include *C. corinna*, the most emblematic species for the archipelago, as well as *A. agestis* and *Polyommatus icarus*, both showing high levels of population diversification (Table 2).

Elba has apparently lost six species. In this case as well, some of the taxa have diversified populations in the study area (*Aglais urticae/ichnusa, Leptidea sinapis* and *Nymphalis polychloros* showed a Gst higher than 0.4). For this island we found no species traits correlated with disappearance and the fraction of likely extinct species (11.5%) was lower than on Capraia 30% of the species reported in the past have disappeared.

Two species disappeared from Capraia were insular endemics (*H. neomiris* and *H. aristaeus*) 399 400 and others showed divergent populations in the study area (*P. cecilia*, *L. phlaeas*, *M. jurtina*). Disappeared species represented a fraction of the fauna with a short flight period. As the 401 length of the flight period is correlated to inter-island dispersal in the Tuscan Archipelago 402 (Dapporto et al., 2012), species with a short flight period have a lower probability of re-403 colonizing from surrounding areas. Interestingly, there are no typical springtime species on 404 Capraia and taxa with short flight periods are typically monovoltine, with adults emerging at 405 the beginning of summer, aestivating during the hottest weeks and laying eggs in 406 September/October. Aestivation in the Mediterranean region is known for *M. jurtina* (Scali, 407

408 1971) and Hipparchia semele (García-Barros, 1988), a species closely related to H. aristaeus. Both of them have disappeared from Capraia, together with two other Satyrinae (P. cecilia 409 410 and *H. neomiris*), which tend to be frequent in woodlands and scrub/maquis during the hottest and driest months. Because there are barely any remaining woods on Capraia, aestivation may 411 represent an important stress period with current temperature increases (Shreeve et al., 2009; 412 Cerrato et al., 2016). Compared to Elba and Giglio, Capraia is more isolated and this could 413 have hampered a rescue effect for many species. Moreover, Elba has the highest mountain 414 415 peak among the small Italian islands (Monte Capanne, 1019m) and woodlands are common on both Giglio and Elba. The higher environmental heterogeneity of Elba and Giglio could 416 have provided a wider range of suitable areas for many species under environmental stress 417 418 and climatic oscillations.

Extinction events over long periods are expected on islands based on the equilibrium theory (MacArthur & Wilson, 1967), but they should be paralleled by colonization events, which was not the case for the Tuscan islands. On Elba only *Anthocharis cardamines* has been discovered after 1950; on Capraia only *P. aegeria* and *A. agestis* have been found after intensive field research between 1970-1980, when almost all the seven extinct species were present; only *C. rubi* has been recently discovered on Giglio.

425 Guidelines for the conservation of the Tuscan Archipelago butterflies

Ranking species according to their contribution to genetic diversity (Table 2) shows that butterfly diversity in the Tuscan Archipelago is encompassed both by insular endemic taxa as well as by widespread species (*A. agestis, C. pamphilus, C. alceae, Melitaea nevadensis, M. jurtina, P. aegeria* and *Zerynthia cassandra*). While these latter species are currently treated as being of 'Least Concern' in the European and Italian Red Lists (Van Swaay et al., 2010; 431 Balletto et al., 2015), some of them disappeared from some islands resulting in a loss of faunistic and genetic diversity. Some of the populations that have apparently disappeared in 432 the last 10 years may be still rediscovered following dedicated field research, as occurred for 433 Zerynthia cassandra, discovered on Elba in 1932 but apparently disappeared before our 434 intensive collection effort (Appendix S1). Species that disappeared from at least one island 435 showed a higher overall population diversification (Dst) compared to persisting species. In 436 fact, our analyses revealed that species accumulated genetic diversification because of their 437 438 reduced migration and colonization capabilities; but these characteristics can also produce higher extinction risk due to reduced gene flow and rescue effect. We showed that one of the 439 main peculiarities of Tuscan islands is the occurrence of strong genetic contrasts among 440 441 nearby areas. If an insular relict or a genetically endemic population goes extinct, it is likely that would be replaced by conspecific propagules from the nearest source, thus lowering the 442 ancestral genetic diversification. There is also evidence for Mediterranean butterflies that the 443 presence of endemic and relict island populations limits the colonization by mainland 444 populations probably due to density-dependent phenomena (Dapporto et al. 2012, Vodă et al. 445 446 2015b). The establishment of the Tuscan Archipelago National Park in 1996 represented a fundamental step for the broad-scale protection of island communities. Nevertheless, specific 447 conservation strategies tailored on particularly valuable species are still lacking because 448 449 evidence-based information are missing. Our integrated approach, by prioritizing species according to their contribution to genetic diversity and by identifying the impact of ecological 450 drivers on the emergence and extinction of differentiated populations, allows conservation 451 452 priorities to be established, with a necessity for regular monitoring schemes. Such schemes 453 should evaluate population consistence and health, particularly focusing on population size and trends, genetic load and persistence of habitat suitability. 454

Our study also suggested that reduction of shady areas might be a driver of species loss in the 455 Archipelago. Historically, wooded and shrub areas have been considerably reduced, mostly 456 457 on Capraia and Montecristo. It has been already reported that deforestation can be a main driver for butterfly extinction in Mediterranean islands. For example, the strong reduction of 458 shady areas has been linked with post-glacial reduction of butterflies on Malta with the recent 459 extinction of most relict elements (Vodă et al., 2016). We thus indicate as a specific 460 conservation action to adopt environmental management procedures aimed at preserving and 461 462 favouring environmental heterogeneity, thus increasing resources availability and suitable habitats for a larger number of butterfly species (Dennis, 2010). The increase of temperature 463 predicted by recent climate change scenarios suggests that environmental heterogeneity will 464 play a pivotal role in buffering increasing thermal and drought stress. 465

In conclusion, this study shows that the integration of molecular and trait data with long-term 466 occurrence records allows the identification of the eco-evolutionary processes underlying the 467 high butterfly diversity in this Mediterranean diversity hotspot (Dennis et al., 2008). 468 Evidence-based priorities for future conservation actions have been provided following the 469 theory of conservation biogeography (Ladle & Whittaker, 2011). Public institutions, such as 470 the NGO Legambiente and the Tuscan Archipelago National Park that have collaborated in 471 this project, have already used some of these results to raise awareness for island diversity 472 protection and to implement key conservation measures for butterflies. 473

474

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483

## 484 Biosketch

- LD, RaV, AC, MM and RoV conceived the idea; LD, RaV, AC, MM, RoV, VD, LF, HB,
- 486 LPC, SS, FZ, UM and LV collected the data and the specimens in the study area; LD, RaV,

487 VD and RoV assessed the taxonomy of the study species; LD, MM, AC and SS compiled the

- 488 trait data; LD, VD, JCH, RaV, MM, AC, RoV, FZ, UM and LV managed and obtained the
- 489 COI sequences; LD, HB and EB gathered the occurrence data; LD, RaV, MM, JCH and AC
- 490 performed the data analysis; all the authors discussed the results and participated in writing

491 the paper.

## 494 **References**

- Balletto, E. & Kudrna, O. (1985) Some aspects of the conservation of butterflies in Italy, with
  recommendations for a future strategy. *Bollettino della Società Entomologica Italiana*, 117, 39-59.
- Balletto, E., Bonelli, S., Barbero, F., Casacci, L.P., Sbordoni, V., Dapporto, L., Scalercio, S.,
  Zilli, A., Battistoni, A., Teofili, C. & Rondinini, C. (2015) Lista Rossa IUCN delle
  Farfalle Italiane Ropaloceri. Comitato Italiano IUCN Ministero dell'Ambiente,
  Roma.
- Burney, C.W. & Brumfield, R.T. (2009) Ecology predicts levels of genetic differentiation in
  Neotropical birds. *The American Naturalist*, **174**, 358-368.
- Cardoso, P., Erwin, T.L., Borges, P.A. & New, T.R. (2011) The seven impediments in
  invertebrate conservation and how to overcome them. *Biological Conservation*,
  144, 2647-2655.
- 507 Carnicer, J., Stefanescu, C., Vila, R., Dincă, V., Font, X. & Peñuelas, J. (2013) A unified
   508 framework for diversity gradients: the adaptive trait continuum. *Global Ecology* 509 *and Biogeography*, 22, 6-18.
- 510 Cerrato, C., Lai, V., Balletto, E. & Bonelli, S. (2016) Direct and indirect effects of weather
  511 variability in a specialist butterfly. *Ecological Entomology*, 41, 263-275.
- 512 Clement, M., Posada, D. & Crandall, K.A. (2000) TCS: a computer program to estimate gene
  513 genealogies. *Molecular Ecology*, 9, 1657–1660.

514	Dapporto,	, L. & Cini, A. (2007) Faunal patterns in Tuscan archipelago butterflies: the
515		dominant influence is recent geography not paleogeography. European Journal of
516		Entomology, 104, 497-503.
517	Dapporto,	, L. & Dennis, R.L.H. (2008) Island size is not the only consideration. Ranking
518		priorities for the conservation of butterflies on Italian offshore islands. Journal of
519		Insect Conservation, 12, 237-249.
520	Dapporto,	, L. (2010) Satyrinae butterflies from Sardinia and Corsica show a kaleidoscopic
521		intraspecific biogeography (Lepidoptera, Nymphlidae). Biological Journal of the
522		Linnean Society, 100, 195-212.
523	Dapporto,	, L., Bruschini, C., Dincă, V., Vila, R. & Dennis, R.L.H. (2012) Identifying zones of
524		phenetic compression in West Mediterranean butterflies (Satyrinae): refugia,
525		invasion and hybridization. Diversity and Distributions, 18, 1066–1076.
526	Dapporto,	, L. & Dennis, R.L. (2013) The generalist-specialist continuum: testing predictions
527		for distribution and trends in British butterflies. Biological Conservation, 157, 229-
528		236.
529	Dapporto,	, L., Ramazzotti, M., Fattorini, S., Talavera, G., Vila, R. & Dennis, R.L.H. (2013)
530		recluster: an unbiased clustering procedure for beta-diversity turnover. <i>Ecography</i> ,
531		<b>36</b> , 1070-1075.
532	Dapporto,	, L., Fattorini, S., Vodă, R., Dincă, V. & Vila, R. (2014) Biogeography of western
533		Mediterranean butterflies: combining turnover and nestedness components of
534		faunal dissimilarity. Journal of Biogeography, 41, 1639-1650.
535	Dawson,	M. N., Hays, C. G., Grosberg, R. K. & Raimondi, P. T. (2014) Dispersal potential
536		and population genetic structure in the marine intertidal of the eastern North
537		Pacific. Ecological Monographs 84, 435–456.

- 538 Dennis, R.L.H. (2010) A resource-based habitat view for conservation: butterflies in the
  539 British landscape. Wiley-Blackwell.
- 540 Dennis, R.L.H., Williams, W.R. & Shreeve, T.G. (1991) A multivariate approach to the
  541 determination of faunal structures among European butterfly species (Lepidoptera:
  542 Rhopalocera). *Zoological Journal of the Linnean Society*, **101**, 1-49.
- 543 Dennis, R.L.H., Dapporto, L., Shreeve, T.G., John, E., Coutsis, J.G., Kudrna, O., Saarinen,
- 544K., Ryrholm, N. & Williams, W.B. (2008) Butterflies of European islands: the545implications of the geography and ecology of rarity and endemicity for

546 conservation. *Journal of Insect Conservation*, **12**, 205–236.

- 547 Dennis, R.L.H., Dapporto, L., Fattorini, S. & Cook, L.M. (2011) The generalism-specialism
  548 debate: the part played by generalists in the life and death of species. *Biological*549 *Journal of the Linnean Society*, **104**, 725–737.
- Dennis, R.L., Hardy, P.B. & Dapporto, L. (2012) Nestedness in island faunas: novel insights
  into island biogeography through butterfly community profiles of colonization
  ability and migration capacity. *Journal of Biogeography*, **39**, 1412-1426.
- 553 Devictor, V., van Swaay, C., Brereton, T., Chamberlain, D., Heliölä, J., Herrando, S., Julliard,
- 554R., Kuussaari, M., Lindström, Å., Roy, D.B. & Schweiger, O. (2012) Differences in
- the climatic debts of birds and butterflies at a continental scale. *Nature Climate Change*, 2, 121-124.
- deWaard J.R., Ivanova N.V., Hajibabaei M. & Hebert P.D.N. (2008) Assembling DNA
  Barcodes: Analytical Protocols. Pp. 275-293. In: Cristofre M. (Hrsg.), *Methods in*
- 559 *Molecular Biology: Environmental Genetics*. Humana Press Inc., Totowa, USA,
  - 560 364 pp.

561	Dincă, V., Montagud, S., Talavera, G., Hernández-Roldán, J., Munguira, M.L., García-
562	Barros, E., Hebert, P.D. & Vila, R. (2015) DNA barcode reference library for
563	Iberian butterflies enables a continental-scale preview of potential cryptic
564	diversity. Scientific Reports, 5, 12395; doi: 10.1038/srep12395.
565	Dirzo, R., Young, H.S., Galetti, M., Ceballos, G., Isaac, N.J. & Collen, B. (2014) Defaunation
566	in the Anthropocene. Science, 345, 401-406.
567	Fattorini, S. (2009) Both Recent and Pleistocene geography determine animal distributional
568	patterns in the Tuscan Archipelago. Journal of Zoology, 277, 291-301.
569	Fernández-Palacios, J.M., Kueffer, C. & Drake, D. (2015) A new golden era in island
570	biogeography. Frontiers of Biogeography, 7(1).
571	Fordham, D.A. & Brook, B.W. (2010) Why tropical island endemics are acutely susceptible
572	to global change. Biodiversity and Conservation, 19, 329-342.
573	García-Barros, E. (1988) Delayed ovarian maturation in the butterfly Hipparchia semele as a
574	possible response to summer drought. <i>Ecological Entomology</i> , <b>13</b> , 391-398.
575	Hampton, S.E., Strasser, C.A., Tewksbury, J.J., Gram, W.K., Budden, A.E., Batcheller, A.L.,
576	Duke, C.S. & Porter, J.H. (2013) Big data and the future of ecology. Frontiers in
577	Ecology and the Environment, 11, 156-162.
578	Hebert, P.D.N., Ratnasingham, S,. de Waard, J.R. (2003) Barcoding animal life: cytochrome c
579	oxidase subunit 1 divergences among closely related species. Proceedings of the
580	Royal Society B, 270, (Suppl 1), S96-S99.
581	Hernández-Roldán L, Dapporto L, Dincă V, Vicente JC, Hornett EA, Šíchová J, Lukhtanov
582	VA, Talavera G, Vila R (2016) Integrative analyses unveil speciation linked to host
583	plant shift in Spialia butterflies. Molecular Ecology, 25, 4267-4284.

- Higgins, L.G. & Riley, N.D. (1970) A field guide to the butterflies of Britain and
  Europe. Collins, London.
- 586 Karsholt, O. & van Nieukerken, E.J. (2012) Fauna Europaea: Lepidoptera. Fauna Europaea
  587 version 2.6.2., www.faunaeur.org
- Kisel, Y. & Barraclough, T.G. (2010) Speciation has a spatial scale that depends on levels of
  gene flow. *The American Naturalist*, **175**, 316-334.
- Kobayashi, T. & Sota, T. (2016) Distance decay of similarity in fungivorous insect
  communities: assessing dispersal limitation using genetic data. *Ecosphere*, 7(6).
- 592 Ladle, R.J. & Whittaker, R.J. (2011). Conservation biogeography. John Wiley & Sons.
- 593 Lafranchis, T. (2000) Les Papillons de jour de France, Belgique et Luxembourg et leurs
  594 chenilles. Biotope Éditions, France.
- 595 Lafranchis, T. (2007) Papillons d'Europe. Editions Diatheo.
- MacArthur, R.H. & Wilson, E.O. (1967) The theory ofisland biogeography. *Monographs in Population Biology*, 1.
- Masini, F., Petruso, D., Bonfiglio, L. & Mangano, G. (2008) Origination and extinction
  patterns of mammals in three central Western Mediterranean islands from the Late
  Miocene to Quaternary. *Ouaternary International*, 182, 63–79.
- 601 Moretti, M., Dias, A. T. C., de Bello, F., Altermatt, F., Chown, S. L., Azcárate, F. M., Bell, J.
- 602 R., Fournier, B., Hedde, M., Hortal, J., Ibanez, S., Öckinger, E., Sousa, J. P., Ellers,
- J. and Berg, M. P. (2016). Handbook of protocols for standardized measurement of
  terrestrial invertebrate functional traits. *Functional Ecology*, **31**, 558–567
- 605 Papadopoulou, A., Anastasiou, I., Spagopoulou, F., Stalimerou, M., Terzopoulou, S., Legakis,
- A. & Vogler, A.P. (2011) Testing the species–genetic diversity correlation in the

- Aegean archipelago: toward a haplotype-based macroecology?. *The American Naturalist*, **178**, 241-255.
- Pamperis, L.N. (2009) The Butterflies of Greece. Second edition. Editions Pamperis, Greece.
- Posada, D,. Crandall, K.A. (2001) Intraspecific gene genealogies: trees grafting into networks.
   *Trends in Ecology and Evolution*, 16: 37–45
- Revell, L. J. (2013) Two new graphical methods for mapping trait evolution on phylogenies. *Methods in Ecology and Evolution*, 4, 754-759.
- Ricklefs, R.E., 2009. Dynamics of colonization and extinction on islands. The theory of island
  biogeography revisited, p.388.
- Rosindell, J. and Phillimore, A.B. (2011) A unified model of island biogeography sheds light
  on the zone of radiation. *Ecology Letters*, 14, 552-560.
- Salisbury, C.L., Seddon, N., Cooney, C.R. & Tobias, J.A. (2012) The latitudinal gradient in
  dispersal constraints: ecological specialisation drives diversification in tropical
  birds. *Ecology Letters*, 15, 847-855.
- Scali, V. (1971) Imaginal diapause and gonadal maturation of *Maniola jurtina* (Lepidoptera:
  Satyridae) from Tuscany. *Journal of Animal Ecology*, 40, 467-472.
- Schweiger, O., Harpke, A., Wiemers, M. & Settele, J. (2014) CLIMBER: Climatic niche
  characteristics of the butterflies in Europe. *ZooKeys*, 367, 65-84.
- 625 Srivathsan, A. and Meier, R., 2012. On the inappropriate use of Kimura-2-parameter (K2P)
  626 divergences in the DNA-barcoding literature. Cladistics, 28(2), pp.190-194.
- 627 Sekar, S. (2012) A meta-analysis of the traits affecting dispersal ability in butterflies: can
- 628 wingspan be used as a proxy? *Journal of Animal Ecology*, **81**, 174-184.

629	Sheffer, E. (2012) A review of the development of Mediterranean pine-oak ecosystems after
630	land abandonment and afforestation: are they novel ecosystems? Annals of Forest
631	Science, <b>69</b> , 429-443.
632	Shreeve, T., Konvička, M. & Van Dyck, H. (2009) Functional significance of butterfly wing
633	morphology variation. Ecology of Butterflies in Europe (ed. by J. Settele, T.
634	Shreeve, M. Konvička and H. Van Dyck), pp.171-188. Cambridge University
635	Press, Cambridge UK.
636	Thiel-Egenter, C., Gugerli, F., Alvarez, N., Brodbeck, S., Cieślak, E., Colli, L., Englisch, T.,
637	Gaudeul, M., Gielly, L., Korbecka, G. & Negrini, R. (2009) Effects of species traits
638	on the genetic diversity of high-mountain plants: a multi-species study across the
639	Alps and the Carpathians. Global Ecology and Biogeography, 18, 78-87.
640	Tolman, T. & Lewington, R. (2008) Collins butterfly guide. Harper Collins UK.
641	Tshikolovets, V.V. (2011) Butterflies of Europe and the Mediterranean area. Tshikolovets
642	Publications, Czech Republic.
643	Venter, O., Fuller, R.A., Segan, D.B., Carwardine, J., Brooks, T., Butchart, S.H., Di Marco,
644	M., Iwamura, T., Joseph, L., O'Grady, D., Possingham, H.P. (2014) Targeting
645	global protected area expansion for imperiled biodiversity. PLoS Biology, 12,
646	1001891.
647	Violle, C., Navas, ML., Vile, D., Kazakou, E., Fortunel, C., Hummel, I. and Garnier, E.
648	(2007) Let the concept of trait be functional! Oikos, 116, 882–892.
649	Vodă, R., Dapporto, L., Dincă, V. & Vila, R. (2015a) Cryptic matters: overlooked species
650	generate most butterfly beta-diversity. <i>Ecography</i> , <b>38</b> , 405-409.
651	Vodă, R., Dapporto, L., Dincă, V. & Vila, R (2015b) Why do cryptic species tend not to co-
652	occur? A case study on two cryptic pairs of butterflies. PLoS one, 10, 0117802.

653	Vodă, R., Dapporto, L., Dincă, V., Shreeve, T.G., Khaldi, M., Barech, G., Rebbas, K.,
654	Sammut, P., Scalercio, S., Hebert, P.D.N. & Vila, R. (2016) Historical and
655	contemporary factors generate unique butterfly communities on islands. Scientific
656	Reports, 6, 28828; doi: 10.1038/srep28828 (2016).
657	Van Buuren, S., Groothuis-Oudshoorn, K. (2011) mice: Multivariate Imputation by Chained
658	Equations in R. Journal of Statistical Software, 45, 1-67.
659	Van Swaay, C., Cuttelod, A., Collins, S., Maes, D., Munguira, M.L., Šašić, M., Settele, J.,
660	Verovnik, R., Verstrael, T., Warren, M., Wiemers, M. & Wynhof, I. (2010)
661	European Red List of Butterflies. Luxembourg: Publications Office of the European
662	Union.
663	Waters, J.M. (2011) Competitive exclusion: phylogeography's 'elephant in the
664	room? <i>Molecular Ecology</i> , <b>20</b> , 4388-4394.
665	Whittaker, R.J. & Fernández-Palacios, J.M. (2007) Island Biogeography: Ecology, Evolution,
666	and Conservation, Oxford University Press.
667	Wilson, R.J., Davies, Z.G. & Thomas, C.D. (2010) Linking habitat use to range expansion
668	rates in fragmented landscapes: a metapopulation approach. <i>Ecography</i> , <b>33</b> , 73-82.
669	
670	
671	Data accessibility. The complete dataset, together with R scripts to replicate the analyses
672	published in the paper, is available in the Supporting Information. The COI sequences are also
673	available in the following repositories (Genbank, BOLD and IO databases, accession codes
674	for Genbank, XXXXX-XXXXX; BOLD, XXXXX).

**Table 1** Species traits used in the study with the description of the type of trait (sensu Moretti et al. 2016), in bold and the relative functional hypothesis in italics; the trait(s) measured; a description of the trait(s), the literature sources and the weights obtained by each trait in the first two Principal Components (in the cases when the analysis has been carried out). PC1 and PC2 represent the weights of the traits in the first two components obtained after principal component analysis (PCA) of each type of trait. PCA on the single variable of host plants has not been carried out and for wingspan only the first component returned an eigenvalue higher

682 than 1.

Type of trait	Functional hypothesis	Trait measured and description	Sources	PC1	PC2
Feeding	Species feeding on a larger number of plants have a wider niche, thus a higher potential to colonize islands (Dennis et al. 2012)	Number of host plant genera used by larvae as reported in two literature sources	Lafranchis (2007) Tolman & Lewington (2008)	-	-
	Larger species are characterized by higher mobility (Sekar, 2012), thus more probabilities to cross	Wingspan : mean between minimum and maximum size reported in four main sources for European butterflies.	Higgins & Riley (1970) Lafranchis (2000) Pamperis (2009)	0.448 0.448 0.446	-
Morphology	sea barriers (Dennis et al. 2012)	Tshikolovets (2011) reported size for males and females	Tshikolovets (2011) males	0.448	-
			Tshikolovets (2011) females	0.448	-
Life history	Phenological attributes determine the length of the most mobile life stage of winged adults. The period of the year when it is expressed.	Month of adult first emergence. It ranges from the coldest, January (1), to the warmest, August (8). No butterfly species has a later first emergence in the	Authors' collection data	-0.293	-0.822

	These characteristics affect	study area			
	barriers (Dapporto et al., 2012; Dennis et al. 2012) and can interact with climatic	Last month when adults fly, ranging from January (1) to December (12)	Authors' collection data	0.533	-0.482
	changes in determine extinction probabilities.	Length of the flight period: number of months when the adults occur in the study area	Authors' collection data	0.587	-0.189
		<b>Voltinism.</b> Number of generations/year in the study area	Authors' collection data and Tolman & Lewington (2008)	0.535	0.237
	Mean climatic conditions of the areas inhabited by a species are considered as good proxies for their econhysiological response to	Mean temperature occurring in the 50×50 km spatial cells where the species has been modeled to occur	Schweiger et al. (2014)	0.334	-0.387
	climate (Devictor et al., 2012). They can affect the	Mean precipitation in the same spatial cells as above	Schweiger et al. (2014)	-0.292	-0.607
gy	probability for species persistence in the warm and dry Mediterranean climate of Tuscan islands.	Maximum temperature tolerance. Upper 95% confidence interval for temperature mean	Schweiger et al. (2014)	0.334	-0.387
Physiolog		Minimum precipitation tolerance. Lower 95% confidence interval for precipitation mean	Schweiger et al. (2014)	-0.332	-0.413
		<b>Overall temperature tolerance.</b> Standard deviation for temperature mean	Schweiger et al. (2014)	-0.332	0.313
		<b>Overall precipitation tolerance.</b> Standard deviation for precipitation mean	Schweiger et al. (2014)	-0.337	-0.056

**Table 2** Species ranked for Gst, Dst and Uni, with information on single island uniqueness
 690 and documented recent extinction events. Dis, species that disappeared from Elba (E), Giglio 691 692 (G) or Capraia (C). The last seven columns represent the uniqueness values of each species on each island. "-" means that the species has not been reported on that particular island, "NA" 693 means that the species has been reported but that the population was not included in this study 694 because of lack of genetic data. Taxa endemic to the insular region (Sardinia, Corsica and 695 Tuscan islands) are written in bold, while the six species that were not recorded on islands 696 during our surveys in the last 10 years are highlighted in grey. 697

Species	Gst	Dst	Uni	Dis	Gor	Cap	Elb	Pia	Mon	Gig	Gia
Aglais urticae/ichnusa	0.937	0.449	NA	Е	-	-	NA	-	-	-	-
Anthocharis cardamines	0	0	0.051		NA	NA	0.051	NA	NA	NA	NA
Argynnis pandora	0.171	0.15	0.076		-	-	0.076	-	-	0.114	-
Argynnis paphia	NA	0	0	Е	-	-	NA	-	-	-	-
Aricia agestis/cramera	0.738	0.984	0.094	G	-	0.152	0.095	0.046	-	NA	NA
Callophrys rubi	0.846	0.279	0		-	-	0	-	-	0	-
Carcharodus alceae	0.779	0.258	0.095		0	NA	0	0.456	NA	0.152	-
Celastrina argiolus	0.114	0.011	0.076		0	0	0.051	0	0	0.076	-
Charaxes jasius	0	0	0		NA	0	0	NA	-	0	0
Coenonympha corinna	0.943	2.194	0.19	G	-	0.456	0.101	-	-	NA	0
Coenonympha pamphilus	0.864	0.86	0.152		-	-	0.152	-	-	-	-
Colias croceus	NA	0	0		0	0	0	0	NA	0	NA
Euchloe ausonia	NA	0	0		-	-	0	-	-	-	-
Favonius quercus	0.393	0.053	0.051		-	-	0.152	-	-	-	-
Gegenes pumilio	NA	0	0		-	-	0	-	-	0	-
Glaucopsyche alexis	0	0	0		-	-	0	-	1	-	-
Gonepteryx cleopatra	0.385	0.004	0		0	0	0	0	-	0	0
Gonepteryx rhamni	0.323	0.237	NA	Е	-	-	NA	-	-	-	-
Hipparchia aristaeus	0.119	0.019	NA	С	NA	NA	NA	-	-	NA	-
Hipparchia neomiris	0.871	0.968	End	С	-	NA	NA	-	-	-	-
Hipparchia statilinus	0.281	0.053	0		-	-	0	-	-	-	-
Iphiclides podalirius	0.701	0.063	0		-	-	0	-	-	-	NA
Issoria lathonia	0.431	0.014	0		-	-	0	-	-	0	-
Lampides boeticus	0.044	0.005	0.038		0	0.152	0	0	0	0	NA

Lasiommata megera/paramegaera	0.966	0.615	0.37		0	0.798	0.076	0	0.152	0	-
Leptidea sinapis	0.501	0.081	0	Е	-	-	NA	-	-	-	-
Leptotes pirithous	0.298	0.048	0		-	0	0	NA	0	0	0
Limenitis reducta	0.185	0.015	0		-	-	0	-	-	-	-
Lycaena phlaeas	0.762	0.098	0.051	С	-	NA	0	0	0.152	0	-
Maniola jurtina	0.603	0.298	0.085	С	-	NA	0.092	0.076	-	0	-
Melitaea cinxia	0.379	0.54	NA	Е	-	-	NA	I	-	-	-
Melitaea didyma	0.344	0.07	0		-	-	0	-	-	-	-
Melitaea nevadensis	0.455	0.492	0.988		-	-	0.988	-	-	-	-
Nymphalis polychloros	0.415	0.043	NA	Е	-	-	NA	-	-	-	-
Papilio machaon	0	0	0.038		0	NA	0	0	-	0.152	-
Pararge aegeria	0.88	0.636	0.076		NA	0.076	NA	0.076	NA	NA	NA
Pieris brassicae	0	0	0.076		NA	NA	0	0	NA	0.101	NA
Pieris mannii	0	0	0		-	-	0	-	-	-	-
Pieris napi	0.432	0.133	0		NA	-	0	-	-	-	0
Pieris rapae	0.118	0.05	0.054		0.157	0.051	0	0.038	NA	0	0
Plebejus bellieri/idas	0.616	0.134	0.051		-	1	0.152	I	-	-	-
Polyommatus icarus	0.216	0.018	0.076	G	-	0.076	0	0	-	NA	-
Pontia edusa	0.079	0.015	0.051	С	-	*	0.076	NA	-	0	NA
Pyrgus armoricanus	NA	0	0		-	-	0	-	-	-	-
Pyronia cecilia	0.26	0.224	0.091	С	-	NA	0	0.152	-	0.076	-
Pyronia tithonus	0.26	0.015	0		-	-	0	-	-	-	-
Satyrium ilicis	NA	0	0		-	1	0	I	-	-	-
Spialia sertorius	0.417	0.009	0		-	-	0	-	-	-	-
Thymelicus acteon	0.432	0.067	0		-	-	0	-	-	-	-
Vanessa atalanta	0.127	0.009	0.051		NA	0.051	0	NA	0	0	NA
Vanessa cardui	0.032	0.007	0.038		0	0	0.076	0	NA	0	0.076
Zerynthia cassandra	0.293	0.053	0.152		-	-	0.228	-	-	-	-
701 **Table 3** Trait variables entered in the AIC phylogenetic regressions for Gst, Dst and

702 Uniqueness (Uni). And the differences in Gst, Dst and Uniqueness between species that have

disappeared at least from one island, compared to species that have not disappeared

	Trait variables	Estimate	StdErr	t.value	p.value
	Host Plants	-0.100	0.045	-2.226	0.032
Gst	Mobility PC1	-0.141	0.046	-3.031	0.004
	Phenol PC1	0.058	0.043	1.363	0.181
	Ecophy PC2	0.068	0.040	1.713	0.095
	Host Plants	-0.110	0.043	-2.550	0.015
Dst	MobilityPC1	-0.126	0.047	-2.660	0.011
	Phenol PC1	0.060	0.042	1.454	0.153
	Ecophy PC2	0.119	0.039	3.042	0.004
Uni	Host Plants	-0.045	0.028	-1.643	0.108
	Ecophy PC2	0.102	0.033	3.112	0.003

**Table 4** The effects of trait variables in the logistic phylogenetic regressions for species
disappearance in Elba and Capraia.

Island	PC1/PC2	Estimate	StdErr	z.value	p.value
	Host_Plants	-0.380	0.469	-0.810	0.418
	Mobility PC1	0.412	0.482	0.855	0.393
Elba	Phenol PC1	-0.382	0.544	-0.702	0.483
	Phenol PC2	-0.522	0.437	-1.194	0.233
	Ecophy PC1	-0.813	0.584	-1.393	0.164
	Ecophy PC2	0.766	0.561	1.367	0.172
	Host_Plants	-0.969	0.972	-0.997	0.319
	Mobility PC1	1.605	1.206	1.331	0.183
Capraia	Phenol PC1	-6.375	3.002	-2.124	0.034
-	Phenol PC2	2.835	1.869	1.517	0.129
	Ecophy PC1	-1.820	1.196	-1.522	0.128
	Ecophy PC2	2.753	1.783	1.544	0.123

Figure 1 (a) The study region where the islands and mainland areas are highlighted with the
same colours obtained in the RGB projection of the Principal Coordinate Analyses (PCoA) in
figure 3; (b) the workflow of the protocol used for the analyses.



Figure 2 Phylogenetic tree based on cytochrome c oxidase subunit 1 (COI) sequences of the 715 butterfly species occurring in the Tuscan Archipelago, with their Gst, Dst, Uni and ecological 716 717 traits. Square root Dst values are mapped over the tree, while Gst and Uni are reported as 718 coloured squares (character mapping on the tree for Gst and Uni is available in Appendix S1). Host plants represent the number of plant genera on which the larva of a given species has 719 been reported: one leaf - one genus; two leaves two to four genera, three leaves more than 720 four genera. The sizes of the butterfly silhouettes are directly correlated with the species size 721 722 (wingspan). Ecophy 1 represents the first PC of physiology traits mostly represent mean temperature and is reported with colours representing quartiles of values (red, preference for 723 warm temperatures; blue, preference for cold temperatures). Ecophy 2 represents the second 724 725 PC of physiology traits mostly correlated with precipitation tolerance and represented in 726 quartile from small (high tolerance to drought) to large drops (low tolerance to drought). Phenol 1 represent quartiles the first PC of phenology correlated with the length of flying 727 period from shortest (one black sector) to longest (four black sectors). Phenol 2, mostly linked 728 to the period of emergence, represent spring species appearing early in the year (black sector 729 730 right-top) to species with a later summer-autumn appearance (black sector left-top).



735	Figure 3 Overall genetic patterns obtained after comparisons among islands based on Gst. (a)
736	The colours obtained in the RGB projection of the Principal Coordinate Analyses (PCoA) are
737	more similar among more genetically similar island communities. When the colours are
738	reported on a map (b) of the studied region they show similarity and contrast among areas.
739	The locality codes are: Arg, Argentario; Cap, Capraia; Cor, Corsica; Elb, Elba; Gia,
740	Giannutri; Gig, Giglio; Gor, Gorgona; Mon, Montecristo; Pia, Pianosa; Sar, Sardinia; Tus,
741	Tuscany; T_C central Tuscany coast; T_N, northern Tuscany coast; T_S, southern Tuscany
742	coast. (c) Haplotype networks based on the COI gene for nine species that exemplify different
743	patterns of genetic variation, obtained by comparing haplotypes from the study area with
744	other regions of Europe (c). A molecular assessment for all the species is available in the
745	Supporting Information (see Appendix S1).



747

## 748 Supporting Information

Additional Supporting Information is available in the online version of this study:

750

Appendix S1 {Supplementary methods and results containing the taxonomic assessment for each species, the occurrence data on Tuscan islands from 1900 to 2015 and the COI assessment for all the species. Supplementary results for PCA and Phylogenetic Models are also provided}

755	Appendix S2 {The dataset in the form of a fasta file (sequences_TA.fas), the information
756	about the specimens used in the study (dataset.txt), the table containing the
757	ecological traits for the 52 species (selected.traits.txt) and the R scripts used to
758	carry out the analyses (Dapporto_et_al_script.R)}
759	
760	Conflict of Interest:
761	The authors declare no conflicts of interest
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# 776 Appendix S1

# Rise and fall of island butterfly diversity. Understanding population diversification and extinction in a highly diverse Archipelago.

Dapporto L., Cini A., Menchetti M., Vodă R., Bonelli S, Casacci L.P., Dincă V., Scalercio S.,
Hinojosa J.C., Biermann H., Forbicioni L., Mazzantini U., Venturi L., Zanichelli F., Balletto E.,
Shreeve T.G., Dennis R.L.H., Vila R.

788 Calculation of Dst and Gst indexes (Nei, 1987).

789

790 Dst is defined as:

791 Dst = Ht - Hs

where Ht represents the average intraspecific p-distances for all specimens of a given species, and Hs is the average of the intra-population p-distances. Thus, Dst represents the average genetic differentiation among populations in p-distance units.

795

The second measure (Gst) is a standardized index (Nei, 1987) defined as:

797 Gst = Dst/Ht

This index ranges from negative values to 1 (complete differentiation). Negative values (intra-area differentiation higher than inter-area differentiation) can have different subtle meanings, but most often are a bias due to relatively small sample sizes; usually they are set to zero (Meirmans & Hedrick, 2011) and for several cases in our study we set the negative values to zero.

803

We also calculated the pairwise Gst among all pairs of populations for each species, using the following formula:

806  $Gst_{i,j} = Dst_{i,j}/Ht$ 

This represents the specific fraction of the overall genetic diversity (Ht) expressed as the inter-population diversification (Dst<sub>i,j</sub>) between a given pair of areas (i and j).

809

In the following section distributional, taxonomic and historical notes for all the species are reported together with their assessment for population diversification. The genetic distances among haplotypes have been reduced to two dimension by using Principal Component Analysis. The resulting configuration has been plotted in RGB space then the resulting colours for specimens reported in the map with pie charts.

816

- 817 Carcharodus alceae (Esper, 1780)
- 818

### 819 **1900-2005**

- 820 Elba: 1908, 1916, 1921, 1951-1961, 1963, 1980 (Biermann & Hesch, 1982, Balletto et al.,
- 821 2007), 2000-2003 (authors).
- 822 Giglio: 1980 (Biermann & Hesch, 1982).
- Pianosa: 1998 (Dapporto et al. 1999).
- 824 **2006-2015**
- 825 Elba: 2008-2015 (authors).
- 826 Giglio: 2010, 2013, 2014 (authors).
- 827 Capraia: 2012 (authors).
- 828 Gorgona: 2015 (authors).

- 830 This species showed some diversification among haplotypes (maximum p-distance 0.9%) and
- a relatively high population diversification (Dst 0.258 and Gst 0.779). Some haplotypes from
- Pianosa and Giglio were not found on the mainland resulting in a uniqueness of 0.095.



Figure S1. PCoA projection (left) of genetic distances among specimens (dots) in the bidimensional RGB colour space. Individual colours have been subsequently plotted on the map (right). The first two PCoA axes explained 99.43% of variance.

839 Spialia sertorius (Hoffmannsegg, 1804)

840

The taxon living in Sardinia and Corsica was identified in the past as *Spialia sertorius therapne*. Recent molecular studies showed that *S. therapne* must be considered as a good species with a diversification higher than 3% respect to *S. sertorius* (Hernández-Roldán et al, 2016). For this reason we did not include Sardo-Corsican populations of *S. therapne* as comparison for *S. sertorius*.

846

847 **1900-2005** 

848 Elba: 1908, 1916, 1921, 1963 (Biermann & Hesch, 1982, Balletto et al., 2007).

849 **2006-2015** 

850 Elba: 2009, 2010, 2013, 2015 (authors).

851

This species showed a low diversification among haplotypes (maximum p-distance 0.2%),

and a rather low population diversification (Dst 0.009 and Gst 0.417). All the haplotypes

found on Elba island were also found on the mainland.

855



Figure S2. PCoA projection (left) of genetic distances among specimens (dots) in the bidimensional RGB colour space. Individual colours have been subsequently plotted on the map (right). The first two PCoA axes explained 100.00% of variance.

862	Pyrgus armoricanus (Oberthür, 1910)
863	
864	1900-2005
865	Elba: 1916, 1956, 1957, 1963 (Biermann & Hesch, 1982 , Balletto et al., 2007).
866	2006-2015
867	Elba: 2013 (authors).
868	
869 870	This species showed a single haplotype in the 13 specimens we analyzed from Tuscany, Elba and Corsica.
871	
872	Thymelicus acteon (von Rottemburg, 1775)
873	
874	1900-2005
875 876	Elba: 1908, 1916, 1921, 1963, 1976 (Biermann & Hesch, 1982, Balletto et al., 2007), 2000-2004 (authors).
877	2006-2015
878	Elba: 2008, 2010, 2012, 2015 (authors).
879	

This species showed a low diversification among haplotypes (maximum p-distance 0.5%), and a medium population diversification (Dst 0.067 and Gst 0.432) mostly due to the

occurrence on Elba of a single haplotype very uncommon on the Tuscan mainland.



Figure S3. PCoA projection (left) of genetic distances among specimens (dots) in the bidimensional RGB colour space. Individual colours have been subsequently plotted on the map (right). The first two PCoA axes explained 100% of variance.

888

890	Gegenes pumilio (Hoffmannsegg, 1804)
891	
892	1900-2005
893 894	Elba: 1916, 1951, 1956, 1957, 1963, 1976 (Biermann & Hesch, 1982 , Balletto et al., 2007), 2000 (authors).
895	2006-2015
896	Elba: 2008, 2013, 2015 (authors).
897	Giglio: 2013 (authors).
898	
899 900	This species showed a single haplotype in the 5 specimens we analyzed from Tuscany, Elba, Giglio and Sardinia.
901	
902	Zerynthia cassandra (Geyer, 1828)
903	
904	1900-2005
905	Elba: 1932 (Bryk, 1932).
906	2006-2015
907	Elba: 2008-2011, 2013, 2014 (authors).

This species showed some diversification among haplotypes (maximum p-distance 0.8%), and a rather low population diversification (Dst 0.053 and Gst 0.293) mostly due to the occurrence on Elba of a single haplotype not occurring on the Tuscan mainland. This haplotype also produced a uniqueness of 0.152.



Figure S4. PCoA projection (left) of genetic distances among specimens (dots) in the bidimensional RGB colour space. Individual colours have been subsequently plotted on the map (right). The first two PCoA axes explained 73.84% of variance. 

920 *Iphiclides podalirius* (Linnaeus, 1758)

921

#### 922 **1900-2005**

- 923 Elba: 1908, 1916, 1921, 1963, 1976 (Biermann & Hesch, 1982, Balletto et al., 2007), 1998
- 924 (Biermann, 1999).
- 925 Giglio: no collection year (Balletto et al. 2007).

#### 926 **2006-2015**

927 Elba: 2014-2015 (authors).

928

This species showed a low diversification among haplotypes (maximum p-distance 0.3%), and a low Dst (0.063). Nevertheless it showed a high Gst of 0.701, mostly due to the occurrence on Corsica of a single haplotype not occurring in the rest of the study area. All the island haplotypes have been found in other areas of the European mainland.





934

Figure S5. PCoA projection (left) of genetic distances among specimens (dots) in the
bidimensional RGB colour space. Individual colours have been subsequently plotted on the
map (right). The first two PCoA axes explained 100% of variance.

941 Papilio machaon Linnaeus, 1758

942

#### 943 **1900-2005**

- 944 Capraia: 1914 (Razzauti, 1917).
- 945 Elba: 1908, 1916, 1921, 1963, 1976, 1980 (Biermann & Hesch, 1982, Balletto et al., 2007),
- 946 1998 (Biermann, 1999), 2000 (authors).
- 947 Pianosa: 1998 (Dapporto et al. 1999).
- 948 Gorgona: 1999-2002 (Dapporto, 2002).
- 949 **2006-2015**
- 950 Elba: 2009-2011, 2013-2015 (authors).
- 951 Giglio: 2007, 2010-2014 (authors).
- 952 Gorgona: 2015 (authors).
- 953 Pianosa: 2014 (authors).
- 954
- 955 Not recorded on Capraia in the last 10 years.

956

This species showed a high diversification among haplotypes (maximum p-distance 1.7%), but their distribution was not spatially structured resulting in no population diversification (Dst and Gst 0). The detection on Giglio of a haplotype not recorded on the mainland resulted in a uniqueness value of 0.038.



Figure S6. PCoA projection (left) of genetic distances among specimens (dots) in the bidimensional RGB colour space. Individual colours have been subsequently plotted on the map (right). The first two PCoA axes explained 89.53% of variance.

968 *Pieris brassicae* (Linnaeus, 1758)

969

#### 970 **1900-2005**

- 971 Gorgona: 1999-2002 (Dapporto, 2004).
- 972 Capraia: 1915 (Razzauti, 1917), 2000 (authors).

Elba: 1908, 1916, 1921, 1963 (Biermann & Hesch, 1982, Balletto et al., 2007), 1998
(Biermann, 1999), 2000-2003 (authors).

- 975 Montecristo: 2001 (authors).
- 976 **2006-2015**
- 977 Elba: 2008-2015 (authors).
- 978 Giglio: 2007, 2010, 2013 (authors).
- 979 Capraia: 2006 (authors).
- 980 Pianosa: 2008, 2014 (authors).

981

This species showed a diversification among haplotypes (maximum p-distance 0.9%), but their distribution was not spatially structured resulting in no population diversification (Dst and Gst 0). The detection on Giglio island of a haplotype not found on the mainland resulted in a uniqueness value of 0.076.



Figure S7. PCoA projection (left) of genetic distances among specimens (dots) in the bidimensional RGB colour space. Individual colours have been subsequently plotted on the map (right). The first two PCoA axes explained 92.72% of variance. 

993 Pieris mannii (Mayer, 1851)

994

- 995 **1900-2005**
- 996 Elba: 1908, 1915, 1921, 1956, 1963, 1963 (Biermann & Hesch, 1982, Balletto et al., 2007).
- 997 Gorgona: 1999-2002 (Dapporto, 2004).
- 998 **2006-2015**
- 999 Elba: 2008-2010, 2012 (authors).

1000

This species is uncommon in the study area and we only examined five specimens. They showed a very low diversification among haplotypes (maximum p-distance 0.5%), and their distribution was not spatially structured resulting in no population diversification (Dst and Gst 0). The haplotype found on Elba was also found on the mainland.

1005



. . . . .

1007

Figure S8. PCoA projection (left) of genetic distances among specimens (dots) in the bidimensional RGB colour space. Individual colours have been subsequently plotted on the map (right). The first two PCoA axes explained 100% of variance.

1012 *Pieris rapae* (Linnaeus, 1758)

1013

#### 1014 **1900-2005**

- 1015 Elba: 1908, 1916, 1921, 1963, 1980 (Biermann & Hesch, 1982, Balletto et al., 2007), 1998
  1016 (Biermann, 1999), 2000-2003 (authors).
- 1017 Giglio: 1908 (Rocci and Turati, 1925), 1980 (Biermann & Hesch, 1982), 2003, 2004 (authors).
- 1018 Gorgona: 1999-2002 (Dapporto, 2004).
- 1019 Capraia: 1915 (Razzauti, 1917), 1980 (Biermann & Hesch, 1982), 2000-2003 (authors).
- 1020 Pianosa: 1998 (Dapporto et al. 1999).
- 1021 Montecristo: 1983 (Raineri, 1986).
- 1022 **2006-2015**
- 1023 Elba: 2008-2015 (authors).
- 1024 Giglio: 2007, 2010, 2013, 2014 (authors).
- 1025 Gorgona: 2015 (authors).
- 1026 Capraia: 2006-2008, 2010, 2013, 2014 (authors).
- 1027 Giannutri: 2014 (authors).
- 1028 Pianosa: 2011, 2013 (authors).

1029

1030 This species showed a high diversification among haplotypes (maximum p-distance 1.4%), 1031 but their distribution was not spatially structured resulting in low population diversification 1032 (Dst 0.050 and Gst 0.112). The detection on Tuscan islands of haplotypes not found on the 1033 mainland resulted in a uniqueness value of 0.054.



1037 Figure S9. PCoA projection (left) of genetic distances among specimens (dots) in the 1038 bidimensional RGB colour space. Individual colours have been subsequently plotted on the 1039 map (right). The first two PCoA axes explained 75.52% of variance.

1042 Pieris napi (Linnaeus, 1758)

1043

#### 1044 **1900-2005**

1045 Elba: 1908, 1916, 1921, 1979, 1980 (Biermann & Hesch, 1982, Balletto et al., 2007), 1998
1046 (Biermann, 1999), 2000-2003 (authors).

1047 Gorgona: 1999-2002 (Dapporto, 2004).

#### 1048 **2006-2015**

1049 Elba: 2008-2011, 2015 (authors).

1050 Giannutri: 2014 (authors).

1051

1052 This species showed a diversification among haplotypes (maximum p-distance 1.1%), and a 1053 discrete population diversification (Dst 0.133 and Gst 0.432) mostly due to the occurrence 1054 on Corsica of a series of haplotypes uncommon on the Tuscan mainland, Elba and Giannutri. 1055 All island haplotypes have been also found on the mainland.







Figure S10. PCoA projection (left) of genetic distances among specimens (dots) in the bidimensional RGB colour space. Individual colours have been subsequently plotted on the map (right). The first two PCoA axes explained 98.41% of variance.

1062

#### 1064 Pontia edusa (Fabricius, 1777)

1065

1066 It has been recognized since decades that Pontia edusa and Pontia daplidice represent two 1067 highly differentiated cryptic species showing 7% of COI divergence and differences in 1068 allozymes (revised by John et al. 2013). For this reason, they have been considered as 1069 distinct entities in this study. In Sardinia and Corsica only P. daplidice has been found, while 1070 on Tuscan islands only P. edusa has been identified so far. It is unknown which species was 1071 found in Capraia between 1968 and 1970 (Gross, 1970). We attributed the population to P. 1072 edusa for the extinction analysis and it is unlikely that this decision have affected the results 1073 of the phylogenetic regression since the traits of the two species are almost identical.

1074

#### 1075 **1900-2005**

- 1076 Elba: 1908, 1916, 1921, 1963, 1980 (Biermann & Hesch, 1982, Balletto et al., 2007).
- 1077 Pianosa: 1998 (Dapporto et al. 1999), 2000, 2003 (authors).
- 1078 Giglio: 1908 (Rocci & Turati, 1908).
- 1079 Capraia: 1968-1970 (Gross, 1970).
- 1080 **2006-2015**
- 1081 Elba: 2008-2013 (authors).
- 1082 Giglio: 2007, 2010, 2013, 2014 (authors).
- 1083
- 1084 Not recorded on Capraia in the last 10 years.

1085

1086 This species showed some diversification among haplotypes (maximum p-distance 0.8%), 1087 but their distribution was not spatially structured resulting in low population diversification 1088 (Dst 0.015 and Gst 0.079). The detection on Elba of a haplotype not found on the mainland 1089 resulted in a uniqueness value of 0.051.



Figure S11. PCoA projection (left) of genetic distances among specimens (dots) in the bidimensional RGB color space. Individual colours have been subsequently plotted on the map (right). The first two PCoA axes explained 100% of variance.

1097 Euchloë ausonia (Hübner, 1804)

1098

1099 In Sardinia and Corsica this species is replaced by the endemic taxon *E. insularis*. *E. ausonia* 1100 and *E. insularis* are not sister species and their p-distance in COI sequences is about 5% 1101 (unpublished data). For this reason, we did not include *E. insularis* in the analysis.

1102

#### 1103 **1900-2005**

- 1104 Elba: 1917 (Biermann & Hesch, 1982, Balletto et al., 2007), 1997, 1998 (Biermann, 1999).
- 1105 **2006-2015**
- 1106 Elba: 2008, 2009, 2013, 2014 (authors).
- 1107
- 1108 This species showed a single haplotype in the 13 specimens we analyzed from Tuscany, 1109 Argentario and Elba.
- 1110
- 1111
- 1112 Anthocharis cardamines (Linnaeus, 1758)
- 1113
- 1114 **1900-2005**
- 1115 Elba: 1998 (Biermann, 1999).
- 1116 **2006-2015**
- 1117 Elba: 2008, 2010, 2011, 2013, 2014 (authors).

1118

1119 This species showed a low diversification among haplotypes (maximum p-distance 0.5%), 1120 and their distribution was not spatially structured resulting in no population diversification 1121 (Dst and Gst 0). The detection on Elba of a haplotypes not found on the mainland resulted in 1122 a uniqueness value of 0.051.



Figure S12. PCoA projection (left) of genetic distances among specimens (dots) in the 1126 bidimensional RGB colour space. Individual colours have been subsequently plotted on the 1127 map (right). The first two PCoA axes explained 73.44% of variance. 1128

1129

1130

Colias croceus (Geoffroy, 1785) 1131

1132

#### 1133 1900-2005

1134 Elba: 1908, 1916, 1921, 1963, 1976, 1980 (Biermann & Hesch, 1982, Balletto et al., 2007),

1998 (Biermann, 1999), 2000-2003 (authors). 1135

- Capraia: 1968-1970 (Gross, 1970), 1996 (Jutzeler & de Bros, 1997), 2000 (authors). 1136
- 1137 Giglio: 1980 (Biermann & Hesch, 1982), 2003 (authors).
- 1138 Montecristo: 1983 (Raineri, 1986).
- 1139 Gorgona: 1999-2002 (Dapporto, 2004)
- 1140 Pianosa: 1998 (Dapporto et al. 1999).
- 2006-2016 1141

- 1142 Elba:, 2008-2015 (authors).
- 1143 Giglio: 2007, 2010, 2012-2014 (authors).
- 1144 Gorgona: 2015 (authors).
- 1145 Capraia: 2007, 2008, 2010, 2012, 2014 (authors).
- 1146 Pianosa: 2011, 2014 (authors).
- 1147
- 1148 This species showed a single haplotype in the 32 specimens we analyzed from several areas.
- 1149
- 1150

1151 *Gonepteryx cleopatra* (Linnaeus, 1767)

1152

#### 1153 **1900-2005**

- Elba: 1908, 1916, 1921, 1963, 1976, 1980 (Biermann & Hesch, 1982, Balletto et al., 2007),
  2000, 2003 (authors).
- 1156 Capraia: 1914 (Razzauti, 1917), 1978 (Biermann & Hesch, 1982), 1996 (Jutzeler & de Bros,
  1157 1998), 2002 (authors).
- 1158 Giglio: 1980 (Biermann & Hesch, 1982).
- 1159 **2006-2015**
- 1160 Elba: 2008-2010, 2012-2015 (authors).
- 1161 Giglio: 2007, 2008, 2010, 2012, 2013 (authors).
- 1162 Giannutri: 2014 (authors).
- 1163 Gorgona: 2015 (authors).
- 1164 Capraia: 2008, 2010, 2012-2014 (authors).
- 1165 Pianosa: 2011, 2014 (authors).

1166

This species showed only two haplotypes differentiated for a single mutation in the 32 specimens we analyzed from several areas. One of the two haplotypes has only been found on Sardinia and on other mainland regions outside the study area. This resulted in some Gst (0.385) in a low Dst (0.004) and in a zero value for Uni.



Figure S13. PCoA projection (left) of genetic distances among specimens (dots) in the bidimensional RGB colour space. Individual colours have been subsequently plotted on the map (right). The first two PCoA axes explained 100% of variance.
```
1179 Gonepteryx rhamni (Linnaeus, 1758)
1180
1181 1900-2005
1182 Elba: 1908 (Biermann & Hesch, 1982, Balletto et al., 2007).
1183
```

1184 In the study area this species showed a series of haplotypes with a high diversification 1185 (maximum p-distance 1.4%). However, they revealed to be only partially spatially structured 1186 resulting in low values of population differentiation (Dst 0.237, Gst 0.323).

- 1187
- 1188 Not recorded on Elba in the last 10 years.
- 1189



1190

1191

Figure S14. PCoA projection (left) of genetic distances among specimens (dots) in the bidimensional RGB colour space. Individual colours have been subsequently plotted on the map (right). The first two PCoA axes explained 87.53% of variance.

1196 Leptidea sinapis (Linnaeus, 1758)

1197

1198 It has been recently showed that in Europe the sinapis group is composed by a triplet of 1199 species (*L. sinapis, L. reali* and *L. juvernica*). Over the study area (Sardinia, Corsica, Tuscan 1200 islands and Tuscany mainland), only *L. sinapis* has been found so far. The species was 1201 probably common on Elba island in the past and a large series of specimens is preserved in 1202 the Roger Verity collection in MZUF. The last report for this species belongs to 1980, 1203 thereafter it apparently disappeared from the island.

1204

### 1205 **1900-2005**

1206 Elba: 1908, 1909, 1916, 1917, 1921, 1963, 1976, 1980 (Biermann & Hesch, 1982, Balletto et1207 al., 2007).

- 1208
- 1209 Not recorded on Elba in the last 10 years.
- 1210

We didn't collect any specimen of this species in Tuscan islands. Over the study area *L. sinapis* showed a low diversification among haplotypes (maximum p-distance 0.3%), but their distribution revealed a good population diversification (Dst 0.081 and Gst 0.501) since the haplotypes occurring on Sardinia and Corsica differ from those detected on the Tuscan mainland.



Figure S15. PCoA projection (left) of genetic distances among specimens (dots) in the bidimensional RGB colour space. Individual colours have been subsequently plotted on the map (right). The first two PCoA axes explained 100% of variance.

1222 Lycaena phlaeas (Linnaeus, 1761)

1223

# 1224 **1900-2005**

- 1225 Elba: 1908, 1916, 1921, 1963, 1976 (Biermann & Hesch, 1982, Balletto et al., 2007), 1998
- 1226 (Biermann, 1999), 2001-2004 (authors).
- 1227 Capraia: 1979 (Biermann & Hesch, 1982).
- 1228 Giglio: 1908 (Rocci & Turati, 1925), 1980 (Biermann & Hesch, 1982).
- 1229 Pianosa: 1998 (Dapporto et al. 1999).
- 1230 **2006-2015**
- 1231 Elba:, 2008-2013 (authors).
- 1232 Giglio: 2007, 2008, 2010, 2014 (authors).
- 1233 Montecristo: 2014 (authors).
- 1234 Pianosa: 2011 (authors).
- 1235
- 1236 Not recorded on Capraia in the last 10 years.
- 1237
- 1238 This species showed a low diversification among haplotypes (maximum p-distance 0.5%), but
- their distribution revealed a high population diversification (Dst 0.098 and Gst 0.762) for the
- 1240 occurrence on Corsica and Montecristo of haplotypes not detected on the Tuscan mainland.
- 1241 These haplotypes were not found elsewhere in Europe resulting in a uniqueness of 0.051.



Figure S16. PCoA projection (left) of genetic distances among specimens (dots) in the bidimensional RGB colour space. Individual colours have been subsequently plotted on the map (right). The first two PCoA axes explained 99.60% of variance. 

1250 Favonius quercus (Linnaeus, 1758)

1251

# 1252 **1900-2005**

1253 Elba: 1908, 1916, 1921, 1963, 1980 (Biermann & Hesch, 1982 and Balletto et al. 2007), 20031254 (authors).

1255 **2006-2015** 

0.051.

1256 Elba: 2008, 2010, 2012, 2015 (authors).

1257

This species showed a low diversification among haplotypes (maximum p-distance 0.5%), but their distribution revealed a high population diversification (Dst 0.053 and Gst 0.393). The haplotype detected on Elba was not found elsewhere in Europe resulting in a uniqueness of

1261

1262



1263

1264

Figure S17. PCoA projection (left) of genetic distances among specimens (dots) in the bidimensional RGB colour space. Individual colours have been subsequently plotted on the map (right). The first two PCoA axes explained 100% of variance.

1270 Callophrys rubi (Linnaeus, 1758)

1271

- 1272 **1900-2005**
- 1273 Elba: 1998 (Biermann, 1999).
- 1274 **2006-2015**
- 1275 Elba: 2008, 2009, 2011, 2013, 2014 (authors).
- 1276 Giglio: 2010 (authors).

1277

This species showed some diversification among haplotypes (maximum p-distance 0.9%), and their distribution revealed a high population diversification (Dst 0.279 and Gst 0.846) for the occurrence on all the studied islands of a group of haplotypes not detected on the Tuscan mainland. However, since these haplotypes also occur in western Europe (Iberia) this species did not score island uniqueness.

1283



1284

Figure S18. PCoA projection (left) of genetic distances among specimens (dots) in the bidimensional RGB colour space. Individual colours have been subsequently plotted on the map (right). The first two PCoA axes explained 99.19% of variance.

- 1289
- 1290 Satyrium ilicis (Esper, 1779)

- 1292 **1900-2005**
- 1293 Elba: 1908, 1916, 1921, 1963, 1968, 1978, 1980 (Biermann & Hesch, 1982; Balletto et al. 1294 2007).
- 1295 **2006-2015**
- 1296 Elba: 2008, 2010, 2015 (authors).

1297

1298 This species showed a single haplotype in the 11 specimens we analyzed from Tuscany 1299 mainland, Elba and Argentario.

1300

1301

1303 *Lampides boeticus* (Linnaeus, 1767)

1304

### 1305 **1900-2005**

- 1306 Elba: 1908, 1916, 1921, 1963, 1968 (Biermann & Hesch, 1982, Balletto et al. 2007), 1998
- 1307 (Biermann, 1999), 2000-2003 (authors).
- 1308 Gorgona: 1999-2002 (Dapporto, 2004).
- 1309 Capraia: 1968-1970, 1996 (revised by Balletto et al. 2007).
- 1310 Giglio: 1980 (Biermann & Hesch, 1982), 2004 (authors).
- 1311 **2006-2015**
- 1312 Elba: 2008-2010, 2012 (authors).
- 1313 Capraia: 2007, 2008, 2012 (authors).
- 1314 Montecristo: 2012 (authors).
- 1315 Giglio: 2007, 2008, 2010, 2013, 2014 (authors).
- 1316 Gorgona: 2015 (authors).
- 1317 Pianosa: 2014 (authors).

1318

- 1319 This species showed a low diversification among haplotypes (maximum p-distance 0.6%), 1320 and their distribution was not spatially structured resulting in low population diversification
- 1321 (Dst 0.005 and Gst 0.045). The detection on Capraia of a haplotype not found on the
- 1322 mainland resulted in a uniqueness value of 0.038.



Figure S19. PCoA projection (left) of genetic distances among specimens (dots) in the bidimensional RGB colour space. Individual colours have been subsequently plotted on the map (right). The first two PCoA axes explained 55.88% of variance. 

1331 *Leptotes pirithous* (Linnaeus, 1767)

1332

### 1333 **1900-2005**

- Elba: 1908, 1916, 1921, 1951, 1963, 1980 (Biermann & Hesch, 1982, Balletto et al. 2007),
  2000, 2002, 2004 (authors).
- 1336 Capraia: 1914 (Razzauti 1917); 1979 (Biermann & Hesch, 1982), 2000 (authors).
- 1337 Gorgona: 1999-2002 (Dapporto, 2004).
- 1338 Giglio:

### 1339 **2006-2015**

- 1340 Elba: 2008, 2010, 2013, 2014 (authors).
- 1341 Giglio: 2010, 2012, 2013, 2014 (authors).
- 1342 Capraia: 2006, 2010 (authors).
- 1343 Montecristo: 2012 (authors).
- 1344 Giannutri: 2014 (authors).
- 1345

1346 This species showed a low diversification among haplotypes (maximum p-distance 0.6%),

and their spatially structure was low (Dst 0.048 and Gst 0.298). All the haplotypes found on

1348 Tuscan islands also occurred on mainland.



Figure S20. PCoA projection (left) of genetic distances among specimens (dots) in the bidimensional RGB colour space. Individual colours have been subsequently plotted on the map (right). The first two PCoA axes explained 99.79% of variance.

1355

1357 *Celastrina argiolus* (Linnaeus, 1758)

1358

### 1359 **1900-2005**

- 1360 Elba: 1908, 1916, 1963, 1980 (Biermann & Hesch, 1982, Balletto et al. 2007), 1998
  1361 (Biermann, 1999), 2002, 2003 (authors).
- 1362 Capraia: 1968-1970, 1978-1980, 1996 (revised by Biermann & Hesch, 1982; Balletto et al.1363 2007), 2001, 2002 (authors).
- 1364 Giglio: 1908, 1979 (revised by Biermann & Hesch, 1982; Balletto et al. 2007).
- 1365 Gorgona: 1999-2002 (Dapporto, 2004).

# 1366 **2006-2015**

- 1367 Elba: 2008-2010, 2013-2015 (authors).
- 1368 Capraia: 2006-2008, 2010-2014 (authors).
- 1369 Giglio: 2007, 2008, 2012, 2013, 2014 (authors).
- 1370 Gorgona: 2015 (authors).
- 1371 Montecristo: 2014 (authors).
- 1372 Pianosa: 2011 (authors).

1373

- 1374 This species showed a low diversification among haplotypes (maximum p-distance 0.3%),
- 1375 and their distribution was not spatially structured resulting in low population diversification
- 1376 (Dst 0.010 and Gst 0.114). The detection on Elba and Giglio of a haplotype not found on the
- 1377 mainland resulted in a uniqueness value of 0.076.



Figure S21. PCoA projection (left) of genetic distances among specimens (dots) in the bidimensional RGB colour space. Individual colours have been subsequently plotted on the map (right). The first two PCoA axes explained 99.98% of variance. 

1386 Glaucopsyche alexis (Poda, 1761)
1387
1388 1900-2005
1389 Elba: 1916 (revised by Biermann & Hesch, 1982).
1390 2006-2015

1391 Elba: 2008, 2009, 2013, 2014 (authors).

1392

- 1393 This species showed a high diversification among haplotypes (maximum p-distance 2.0%),
- but their distribution was not spatially structured resulting in no population diversification
- 1395 (Dst and Gst were 0). The haplotype found on Elba also occurs on the European mainland.

1396



1397

1398

Figure S22. PCoA projection (left) of genetic distances among specimens (dots) in the bidimensional RGB colour space. Individual colours have been subsequently plotted on the map (right). The first two PCoA axes explained 100% of variance.

1402

1404 Plebejus idas/bellieri

1405

The taxonomy of this group is still debated. The Sardo-Corsican populations as referred as *P*. *bellieri*, the Elba population is referred by some authors as *P. villai* or *P. bellieri villai* and the Tuscan populations as *P. idas* or *P. abetonica*). According to COI there clear distinction between insular and mainland populations populations. For this reason, they have been analyzed together.

1411

### 1412 **1900-2005**

1413 Elba: 1958, 1963, 1973, 1976, 1978, 1979, 1980, 1993; 1998; 2002 (Biermann & Hesch, 1982,

- 1414 Balletto et al. 2007).
- 1415 **2006-2015**
- 1416 Elba: 2008, 2010, 2013 (authors).
- 1417
- 1418 This species showed some diversification among haplotypes (maximum p-distance 0.8%),
- and their distribution showed a good spatial structure resulting in Dst equal to 0.134 and Gst

1420 equal to 0.616. The detection on Elba of a haplotype not found on the mainland resulted in a

1421 uniqueness value of 0.051.

1422



Figure S23. PCoA projection (left) of genetic distances among specimens (dots) in the bidimensional RGB colour space. Individual colours have been subsequently plotted on the map (right). The first two PCoA axes explained 82.89% of variance.

1428

1430 Aricia agestis/cramera

1431

1432 In Sardinia, *A. agestis* is replaced by the South-Western taxon *A. cramera*. The two taxa show 1433 a rather low COI divergence (around 2.4% depending on different haplotypes) (Sañudo-1434 Restrepo et al., 2013). For this reason, *A. agestis* from Tuscany, Tuscan islands and Corsica 1435 and *A. cramera* from Sardinia have been used together in the analyses.

1436

# 1437 **1900-2005**

- Elba: 1908, 1916, 1920, 1921, 1951, 1957, 1963, 1968, 1976, 1980 (Biermann & Hesch, 1982;
  Balletto et al. 2007), 2000, 2002, 2003 (authors).
- 1440 Giglio: 1908 (Rocci and Turati, 1925), 1979 (Biermann & Hesch, 1982).
- 1441 Pianosa: 1998 (Dapporto et al. 1999).
- 1442 **2006-2015**
- 1443 Elba: 2008-2010, 2012-2015 (authors).
- 1444 Capraia: 2013 (authors).
- 1445
- 1446 Not recorded on Giglio in the last 10 years.

1447

As reported above, this species showed a high diversification among two taxa (maximum pdistance 3.2%), and their distribution revealed a high population diversification (Dst 0.984 and Gst 0.738) for the occurrence on all the studied islands of a group of haplotypes not detected on the Tuscan mainland. Some haplotypes were not recorded on the mainland and this species scored a uniqueness of 0.094.



Figure S24. PCoA projection (left) of genetic distances among specimens (dots) in the bidimensional RGB colour space. Individual colours have been subsequently plotted on the map (right). The first two PCoA axes explained 96.73

% of variance. 

1462 *Polyommatus icarus* (Rottemburg, 1775)

1463

- 1464 It has been recently found that the populations from Sardinia belong to *P. celina*, a S-W 1465 Mediterranean taxon morphologically almost identical to *P. icarus*. The COI divergence is
- high (6%) and, for this reason, *P. celina* has been removed from the analysis of *P. icarus*.
- 1467

# 1468 **1900-2005**

- 1469 Elba: 1908, 1916, 1921, 1963, 1968, 1978, 1980 (Biermann & Hesch, 1982; Balletto et al.
  1470 2007), 2000, 2002, 2003 (authors).
- 1471 Capraia: 1979 (Biermann & Hesch, 1982).
- 1472 Giglio: 1908, 1979 (Biermann & Hesch, 1982; Balletto et al. 2007).
- 1473 Pianosa: 1998 (Dapporto et al. 1999).
- 1474 **2006-2015**
- 1475 Elba: 2008-2010, 2012-2015 (authors).
- 1476 Capraia: 2013, 2014 (authors).
- 1477 Pianosa: 2011 (authors).
- 1478
- 1479 Not recorded on Giglio in the last 10 years.

1480

- 1481 This species showed a low diversification among haplotypes (maximum p-distance 0.5%) and
- in a rather low population diversification (Dst 0.018 and Gst 0.216). The occurrence on
- 1483 Capraia of a haplotype not found on the mainland resulted in an island uniqueness of 0.076.



Figure S25. PCoA projection (left) of genetic distances among specimens (dots) in the bidimensional RGB colour space. Individual colours have been subsequently plotted on the map (right). The first two PCoA axes explained 46.86% of variance.

1491	Argynnis paphia (Linnaeus, 1758)
1492	
1493	1900-2005
1494	Elba: 1908, 1916, 1921 (Biermann & Hesch, 1982; Balletto et al. 2007).
1495	
1496	Not recorded on Elba in the last 10 years.
1497	
1498 1499	We didn't collect any specimen of this species in Tuscan islands. Over the study area (Tuscany, Corsica and Sardinia) this species showed a single haplotype.
1500	
1501	Argynnis pandora (Denis & Schiffermüller, 1775)
1502	
1503	1900-2005
1503 1504 1505	<b>1900-2005</b> Elba: 1908, 1909, 1916, 1921, 1961, 1963, 1978, 1980 (Biermann & Hesch, 1982; Balletto et al. 2007), 2002 (authors).
1503 1504 1505 1506	1900-2005         Elba: 1908, 1909, 1916, 1921, 1961, 1963, 1978, 1980 (Biermann & Hesch, 1982; Balletto et al. 2007), 2002 (authors).         Giglio: 1908 (Rocci & Turati, 1925), 1979 (Biermann & Hesch, 1982).
1503 1504 1505 1506 1507	1900-2005         Elba: 1908, 1909, 1916, 1921, 1961, 1963, 1978, 1980 (Biermann & Hesch, 1982; Balletto et al. 2007), 2002 (authors).         Giglio: 1908 (Rocci & Turati, 1925), 1979 (Biermann & Hesch, 1982).         2006-2015
1503 1504 1505 1506 1507 1508	1900-2005         Elba: 1908, 1909, 1916, 1921, 1961, 1963, 1978, 1980 (Biermann & Hesch, 1982; Balletto et al. 2007), 2002 (authors).         Giglio: 1908 (Rocci & Turati, 1925), 1979 (Biermann & Hesch, 1982).         2006-2015         Elba: 2008, 2009, 2010, 2012, 2013, 2015 (authors).
1503 1504 1505 1506 1507 1508 1509	1900-2005         Elba: 1908, 1909, 1916, 1921, 1961, 1963, 1978, 1980 (Biermann & Hesch, 1982; Balletto et al. 2007), 2002 (authors).         Giglio: 1908 (Rocci & Turati, 1925), 1979 (Biermann & Hesch, 1982).         2006-2015         Elba: 2008, 2009, 2010, 2012, 2013, 2015 (authors).         Giglio: 2008, 2010, 2012, 2013, 2014 (authors).
1503 1504 1505 1506 1507 1508 1509 1510	<ul> <li>1900-2005</li> <li>Elba: 1908, 1909, 1916, 1921, 1961, 1963, 1978, 1980 (Biermann &amp; Hesch, 1982; Balletto et al. 2007), 2002 (authors).</li> <li>Giglio: 1908 (Rocci &amp; Turati, 1925), 1979 (Biermann &amp; Hesch, 1982).</li> <li>2006-2015</li> <li>Elba: 2008, 2009, 2010, 2012, 2013, 2015 (authors).</li> <li>Giglio: 2008, 2010, 2012, 2013, 2014 (authors).</li> </ul>
1503 1504 1505 1506 1507 1508 1509 1510 1511 1512 1513 1514	<ul> <li>1900-2005</li> <li>Elba: 1908, 1909, 1916, 1921, 1961, 1963, 1978, 1980 (Biermann &amp; Hesch, 1982; Balletto et al. 2007), 2002 (authors).</li> <li>Giglio: 1908 (Rocci &amp; Turati, 1925), 1979 (Biermann &amp; Hesch, 1982).</li> <li>2006-2015</li> <li>Elba: 2008, 2009, 2010, 2012, 2013, 2015 (authors).</li> <li>Giglio: 2008, 2010, 2012, 2013, 2014 (authors).</li> <li>This species showed a high diversification among haplotypes (maximum p-distance 2.6%), but their distribution was not highly spatially structured resulting a low population Gst (0.171) and in a medium Dst (0.150). One haplotype detected on Giglio and Elba was not found on the mainland resulting in a uniqueness value of 0.076.</li> </ul>



Figure S26. PCoA projection (left) of genetic distances among specimens (dots) in the bidimensional RGB colour space. Individual colours have been subsequently plotted on the map (right). The first two PCoA axes explained 98.05 % of variance.

1521

1523 Issoria lathonia (Linnaeus, 1758)

1524

### 1525 **1900-2005**

- 1526 Elba: 1908, 1916, 1921, 1963, 1978, 1980 (Biermann & Hesch, 1982; Balletto et al. 2007),
- 1527 1998 (Biermann, 1999).
- 1528 Giglio: 1908 (Rocci & Turati, 1925), 1979 (Biermann & Hesch, 1982).
- 1529 **2006-2015**
- 1530 Elba: 2008-2010, 2012-2015 (authors).
- 1531 Giglio: 2014 (authors).

1532

- 1533 This species showed only three haplotypes only slightly differentiated for a single mutation.
- 1534 One of them has been only found in Corsica resulting in some Gst (0.431) in a low Dst (0.014) 1535 and in a zero value for island uniqueness.
- 1536



Figure S27. PCoA projection (left) of genetic distances among specimens (dots) in the bidimensional RGB colour space. Individual colours have been subsequently plotted on the map (right). The first two PCoA axes explained 100% of variance.

1544 Aglais urticae/ichnusa

1545

1546 In Sardinia and Corsica, *A. urticae* is replaced by the island endemic *A. ichnusa*, showing a 1547 low COI divergence (around 1.2% in our sample) (see also Vandewoestijne et al., 2004). For 1548 this reason, the two species have been analyzed together. The two taxa differ for wing 1549 pattern and the specimen collected on Elba island by Roger Verity (still preserved in his 1550 collection) revealed a typical *A. urticae* wing pattern.

- 1551
- 1552 **1900-2005**
- 1553 Elba: 1908, 1916 (Biermann & Hesch, 1982).
- 1554
- 1555 Not recorded on Elba in the last 10 years.
- 1556

According to the existence of two taxa, *A. urticae* and *A ichnusa* showed a high diversification between Tuscany and Corsica-Sardinia (maximum divergence 1.2%, Gst (0.934) and Dst (0.449).



Figure S28. PCoA projection (left) of genetic distances among specimens (dots) in the bidimensional RGB colour space. Individual colours have been subsequently plotted on the map (right). The first two PCoA axes explained 100% of variance.

1566

1568 Nymphalis polychloros (Linnaeus, 1758)

1569

1570 **1900-2005** 

1571 Elba: 1908, 1916 (Biermann & Hesch, 1982; Balletto et al. 2007)

1572

1573 In the study area this species showed a series of haplotypes with a low level of diversification 1574 (maximum p-distance 0.3%). However, they revealed some spatial structured resulting in a 1575 rather low value of Dst due to the low diversification (0.043) but in a medium values of Gst 1576 (0.415).

- 1577
- 1578 Not recorded on Elba in the last 10 years.

1579



1580

1581

Figure S29. PCoA projection (left) of genetic distances among specimens (dots) in the bidimensional RGB color space. Individual colours have been subsequently plotted on the map (right). The first two PCoA axes explained 100% of variance.

1585

1587 Vanessa atalanta (Linnaeus, 1758)

1588

### 1589 **1900-2005**

- 1590 Elba: 1908, 1916, 1921, 1956, 1963, 1978, 1980, 2007 (Biermann & Hesch, 1982, Balletto et
  1591 al. 2007), 1998 (Biermann, 1999).
- 1592 Capraia: 1915 (Razzauti, 1917), 1978 (Biermann & Hesch, 1982), 2002 (authors).
- 1593 Giglio: 1908, 1979 (Biermann & Hesch, 1982; Balletto et al. 2007), 2003 (authors).
- 1594 Montecristo: 1983 (Raineri, 1986), 2003 (authors).
- 1595 Gorgona: 1999-2002 (Dapporto 2004).
- 1596 Pianosa: 1998 (Dapporto et al., 1999).
- 1597 **2006-2015**
- 1598 Elba: 2008-2015 (authors).
- 1599 Capraia: 2006-2008, 2010, 2012-2014 (authors).
- 1600 Giglio: 2007, 2008, 2010, 2012, 2013 (authors).
- 1601 Montecristo: 2012, 2014 (authors).

1602

This species showed a low diversification among haplotypes (maximum p-distance 0.5%), and their distribution was not spatially structured resulting in low population diversification (Dst 0.009 and Gst 0.126). The detection on Capraia and Elba of two haplotypes not found on the mainland resulted in a uniqueness value of 0.051.



Figure S30. PCoA projection (left) of genetic distances among specimens (dots) in the bidimensional RGB colour space. Individual colours have been subsequently plotted on the map (right). The first two PCoA axes explained 82.39% of variance.

- 1615 Vanessa cardui (Linnaeus, 1758)
- 1616

### 1617 **1900-2005**

- 1618 Elba: 1908, 1916, 1921, 1963, 1978, 1980 (Biermann & Hesch, 1982, Balletto et al. 2007),
  1619 1998 (Biermann, 1999), 2000-2003 (authors).
- 1620 Capraia: 1915 (Razzauti, 1917), 1968-1970 (Gross, 1970), 1978 (Biermann & Hesch, 1982),
  1621 2001, 2003 (authors).
- 1622 Giglio: 1908 (Rocci & Turati, 1925), 1979 (Biermann & Hesch, 1982).
- 1623 Pianosa: 1998 (Dapporto et al., 1999).
- 1624 Montecristo: 1979 (Fanfani & Groppali, 1979).
- 1625 **2006-2015**
- 1626 Elba: 2006, 2008-2010, 2012-2015 (authors).
- 1627 Capraia: 2006-2008, 2012, 2014 (authors).
- 1628 Giglio: 2007, 2013 (authors).
- 1629 Montecristo: 2012, 2014 (authors).
- 1630

1631 This species showed a low diversification among haplotypes (maximum p-distance 0.8%), 1632 and their distribution was not spatially structured resulting in very low population 1633 diversification (Dst 0.007 and Gst 0.032). The detection on Gorgona and Giannutri of a 1634 haplotype not found on the mainland resulted in a uniqueness value of 0.038



Figure S31. PCoA projection (left) of genetic distances among specimens (dots) in the bidimensional RGB colour space. Individual colours have been subsequently plotted on the map (right). The first two PCoA axes explained 76.06% of variance. 

1643 Melitaea didyma (Esper, 1778)

1644

# 1645 **1900-2005**

- 1646 Elba: 1908, 1916, 1917, 1921, 1924, 1956, 1957, 1961, 1963, 1968, 1978, 1980 (Biermann &
  1647 Hesch, 1982, Balletto et al. 2007).
- 1648 **2006-2015**
- 1649 Elba: 2012, 2015 (authors).

1650

- 1651 This species showed a low diversification among haplotypes (maximum p-distance 0.6%) but
- their distribution revealed some spatial structure (Dst 0.070 and Gst 0.344). The haplotype
- 1653 from Elba has been also found on the mainland.

1654



1655

1656

Figure S32. PCoA projection (left) of genetic distances among specimens (dots) in the bidimensional RGB colour space. Individual colours have been subsequently plotted on the map (right). The first two PCoA axes explained 93.13% of variance.

1660

1662 *Melitaea nevadensis* (Oberthür, 1904)

1663

# 1664 **1900-2005**

1665 Elba: 1908, 1916, 1917, 1921, 1978 (Biermann & Hesch, 1982; Balletto et al. 2007), 2002 1666 (authors).

- 1667 **2006-2015**
- 1668 Elba: 2015 (authors).

1669

1670 This species revealed a high intra-specific diversification in the study area (2.1%). The 1671 existence of two closely related haplotypes from Elba not found elsewhere resulted in high 1672 population differentiation (Dst 0.492, Gst 0.454 and Uni 0.989). The haplotypes found on 1673 Elba resulted in the highest uniqueness measured in this study.

1674



1675

1676

1677 Figure S33. PCoA projection (left) of genetic distances among specimens (dots) in the 1678 bidimensional RGB colour space. Individual colours have been subsequently plotted on the 1679 map (right). The first two PCoA axes explained 91.39% of variance.
1682 Limenitis reducta Staudinger, 1901

1683

1684 **1900-2005** 

1685 Elba: 1908, 1916, 1921, 1963, 1978, 1980 (Biermann & Hesch, 1982; Balletto et al. 2007).

- 1686 **2006-2015**
- 1687 Elba: 2006-2013 (authors).

1688

1689 This species showed a low diversification among haplotypes (maximum p-distance 0.3%) and 1690 their distribution showed a very low spatial structure (Dst 0.015 and Gst 0.185). The 1691 haplotypes from Elba have been also found on the mainland.

1692



1693

1694

Figure S34. PCoA projection (left) of genetic distances among specimens (dots) in the bidimensional RGB colour space. Individual colours have been subsequently plotted on the map (right). The first two PCoA axes explained 79.48% of variance.

1698

- 1700 Charaxes jasius (Linnaeus, 1767)
- 1701
- 1702 **1900-2005**
- 1703 Capraia: 1996 (Jutzeler & de Bros, 1997), 2000, 2001 (authors).
- 1704 Elba: 1908, 1916, 1920, 1921, 1963, 1980 (Biermann & Hesch, 1982; Balletto et al. 2007),
- 1705 2000 (authors)
- 1706 Giglio: 1908 (Rocci & Turati, 1925), , 1979 (Biermann & Hesch, 1982), 2004 (authors).
- 1707 Gorgona: 2004 (Dapporto & Cini, 2007) [without collection date]
- 1708 Pianosa: 2005 (Dapporto & Cini, 2007) [without collection date]

# 1709 **2006-2015**

- 1710 Elba: 2006-2009, 2011, 2013, 2014 (authors).
- 1711 Giglio: 2010, 2013 (authors).
- 1712 Capraia: 2007, 2010, 2014 (authors).
- 1713 Giannutri: 2014 (authors).
- 1714

This species showed three haplotypes only differentiated for a few mutations and with a maximum diversification of 0.2%. One of the haplotypes has been only found on Sardinia resulting in negative Gst and Dst values which have been set to zero. The haplotype from Tuscan islands has been also found on the mainland.



Figure S35. PCoA projection (left) of genetic distances among specimens (dots) in the bidimensional RGB colour space. Individual colours have been subsequently plotted on the map (right). The first two PCoA axes explained 100% of variance. 

1727 *Pararge aegeria* (Linnaeus, 1758)

1728

## 1729 **1900-2005**

- 1730 Elba: 1908, 1916, 1919, 1921, 1968, 1980 (Biermann & Hesch, 1982; Balletto et al. 2007),
- 1731 1998 (Biermann, 1999), 2002, 2003 (authors)
- 1732 Montecristo: 1979 (Fanfani & Groppali 1979).
- 1733 Giglio: 1908 (Rocci & Turati 1925), 1979 (Biermann & Hesch, 1982).

## 1734 **2006-2015**

- 1735 Elba: 2008-2015 (authors).
- 1736 Capraia: 2008, 2010, 2014 (authors).
- 1737 Giglio: 2007, 2010, 2014 (authors)
- 1738 Giannuutri: 2014 (authors)
- 1739 Pianosa: 2011 (authors).
- 1740 Montecristo: 2014 (authors).
- 1741 Gorgona: 2015 (authors).

1742

1743 It is well known that this species shows two diverging clades between North Africa and 1744 Europe (Weingartner et al. 2006; Vodă et al. 2016) (2.1% of maximum divergence in our 1745 dataset). We found that the North African lineage also occurs on Sardinia and a single 1746 specimen has been also found on Corsica. All the specimens from Tuscan islands revealed to 1747 belong to the European clade. According to this diversification the species revealed a high 1748 population differentiation (Dst 0.636, Gst 0.880); the occurrence on Capraia and Pianosa of a 1749 haplotype not found on the mainland resulted in a uniqueness value of 0.076.

1750



Figure S36. PCoA projection (left) of genetic distances among specimens (dots) in the bidimensional RGB colour space. Individual colours have been subsequently plotted on the map (right). The first two PCoA axes explained 96.10% of variance.

1759 Lasiommata megera/paramegaera

1760

*Lasiommata megera* is replaced on Sardinia, Corsica and Capraia by the insular endemic taxon *L. paramegaera*, differentiated in wing pattern, genitalia shape and composition of cuticular lipids (e.g. Kudrna 1977; Dapporto 2007, 2008). To our knowledge no COI sequence was available in literature for *L. paramegaera*. Based on our sequences the genetic distance in COI between *L. megera* and *L. paramegaera* is lower than 3% (1.5% of maximum divergence). For this reason we analysed the two taxa together.

1767

### 1768 **1900-2005**

1769 Elba: 1908, 1916, 1921, 1963, 1968, 1980 (Biermann & Hesch, 1982; Balletto et al. 2007),

1770 1998 (Biermann, 1999), 2000-2003 (authors)

- 1771 Montecristo: 1979 (Fanfani & Groppali 1979), 2001 (authors).
- 1772 Giglio: 1908 (Rocci & Turati 1925), 1979 (Biermann & Hesch, 1982), 2003, 2004 (authors).
- 1773 Gorgona: 1999-2002 (Dapporto 2004).
- 1774 Pianosa: 1998 (Dapporto et al., 1999).
- 1775 Capraia: 1980, 1996 (Biermann & Hesch, 1982), authors (2000-2002).
- 1776 **2006-2015**
- 1777 Elba: 2006-2015 (authors).
- 1778 Capraia: 2006-2008, 2010, 2012-2014 (authors).
- 1779 Giglio: 2007, 2008, 2010, 2012-2014 (authors).
- 1780 Pianosa: 2011, 2014 (authors).
- 1781 Montecristo: 2010, 2012, 2014 (authors).
- 1782 Gorgona: 2015 (authors).

1783

We revealed a strong correlation between the COI structure and the identification of populations to *L. megera* and *L. paramegaera* (Dapporto 2008) with the individuals belonging to Capraia and Montecristo clustering with those from Sardinia and Corsica. According to this diversification this complex of taxa showed a high populationdifferentiation (Dst 0.615, Gst 0.966, Uni 0.370).



1790

Figure S37. PCoA projection (left) of genetic distances among specimens (dots) in the bidimensional RGB color space. Individual colours have been subsequently plotted on the map (right). The first two PCoA axes explained 99.79% of variance.

1794

1796 *Coenonympha corinna* (Hübner, 1804)

1797

1798 The taxonomic status of Coenonympha corinna (Sardinia, Corsica, Capraia) and 1799 Coenonympha elbana (Elba, Giglio, Giannutri, Argentario and Tuscany coast) has been highly 1800 debated in last years resulting in different authors considering them as different taxa (e.g. 1801 Kodandaramaiah et al. 2009; Balletto et al. 2007) or as a single one (e.g. Dapporto & Strumia 1802 2008; Kudrna et al. 2015). Kodandaramaiah et al. (2009) compared specimens from Sardinia 1803 and Elba for genetic markers finding some diversification while Dapporto and Strumia (2008) 1804 did not reveal any constant morphological diversification in genitalia and a continuous cline 1805 in ocellation.

1806

### 1807 **1900-2005**

- 1808 Elba: 1908, 1917, 1920, 1921, 1956, 1957, 1963, 1968, 1980 (Biermann & Hesch, 1982;
  1809 Balletto et al. 2007), 1998 (Biermann, 1999), 2000-2003 (authors)
- 1810 Giglio: 1980 (Biermann & Hesch, 1982), 2000 (Balletto et al. 2007).
- 1811 Capraia: 1915, 1968, 1978-1980, 1995, 1996 (Biermann & Hesch, 1982; Balletto et al. 2007),
- 1812 authors (2000-2003).
- 1813 Giannutri: 1994 (Baletto et al. 2007).
- 1814 **2006-2015**
- 1815 Elba: 2008-2010, 2012-2015 (authors).
- 1816 Capraia: 2006-2008, 2010, 2012-2014 (authors).
- 1817 Giannutri: 2007, 2014 (authors).
- 1818
- 1819 Not recorded on Giglio in the last 10 years.

1820

The analysis of a complete COI dataset revealed a complex pattern of diversification with the existence of three main clades (I. Sardinia, II. Corsica+Capraia, III. Elba+Giannutri+Tuscan mainland). This diversification did not correlate with the supposed taxonomic diversification between Sardinia-Corsica-Capraia and Elba-Giannutri-Tuscany coast populations. The haplotype diversification appeared to be high (with a maximum divergence of 5.2%) and it revealed a high population diversification and uniqueness (Dst 2.194, Gst 0.943, Uni 0.190).

1827



1829

Figure S38. PCoA projection (left) of genetic distances among specimens (dots) in the bidimensional RGB colour space. Individual colours have been subsequently plotted on the map (right). The first two PCoA axes explained 97.41% of variance.

1833 Coenonympha pamphilus (Linnaeus, 1758)

1834

1835 It has been suggested that populations from North Africa, Southern Spain, Balearics and 1836 Sardinia belong to a different species (*Coenonympha lyllus*) (Boillat, 2002). Actually, 1837 populations from North Africa and Balearics differ from those belonging to Europe but the p-1838 distance in COI are low (Dincă et al. 2015; Vodă et al. 2016). We confirmed here that the 1839 Sardinian specimens belong to the North African clade and that the maximum divergence 1840 was 2.4%, for this reason we analysed the two taxa together.

1841

#### 1842 **1900-2005**

1843 Elba:1908, 1916, 1921, 1963, 1968, 1980 (Biermann & Hesch, 1982; Balletto et al. 2007),
1844 1998 (Biermann, 1999).

- 1845 **2006-2015**
- 1846 Elba: 2008-2015.

1847

We revealed a strong correlation between the COI structure and the identification of populations to *C. pamphilus* and *C. lyllus* (Boillat 2002) with the individuals from Sardinia representing a different genetic clade from those belonging to other localities. According to this diversification this taxon showed a high population differentiation (Dst 0.856, Gst 0.864). One haplotype from Elba was not recorded on the mainland resulted in a uniqueness of 0.152.



Figure S39. PCoA projection (left) of genetic distances among specimens (dots) in the bidimensional RGB colour space. Individual colours have been subsequently plotted on the map (right). The first two PCoA axes explained 98.74% of variance.

1862 Pyronia tithonus (Linnaeus, 1771)

1863

1864 **1900-2005** 

1865 Elba:1908, 1921, 1963, 1968, 1980 (Biermann & Hesch, 1982; Balletto et al. 2007), 20031866 (authors).

- 1867 **2006-2015**
- 1868 Elba: 2008, 2010, 2012, 2015 (authors).

1869

- 1870 This species showed a low diversification among haplotypes (maximum p-distance 0.3%),
- 1871 and their distribution was not spatially structured resulting in low population diversification
- 1872 (Dst 0.015 and Gst 0.260). The haplotype from Elba also occurs on the mainland.

1873



1874

1875

Figure S40. PCoA projection (left) of genetic distances among specimens (dots) in the bidimensional RGB colour space. Individual colours have been subsequently plotted on the map (right). The first two PCoA axes explained 97.04% of variance.

1879

1881 Pyronia cecilia (Vallantin, 1894)

1882

## 1883 **1900-2005**

- 1884 Elba: 1908, 1916, 1921, 1956, 1957, 1963, 1968, 1980 (Biermann & Hesch, 1982; Balletto et
  1885 al. 2007), 2000-2002 (authors).
- 1886 Capraia: 1961, 1978, 1996 (Biermann & Hesch, 1982; Balletto et al. 2007), 2001 (authors).
- 1887 Pianosa: 1998 (Dapporto et al., 1999).
- 1888 Giglio: 1980 (Biermann & Hesch, 1982), 2003, 2004 (authors).
- 1889 **2006-2015**
- 1890 Elba: 2008, 2010, 2012, 2013, 2015 (authors).
- 1891 Pianosa: 2011 (authors).
- 1892 Giglio: 2008, 2010 (authors).
- 1893
- 1894 Not recorded on Capraia in the last 10 years.

1895

1896 As showed by Vodă et al. (2016) this species shows two highly diverging clades, the first 1897 belongs to North Africa, the second to the Italian mainland. The two clades are sympatric in Sicily. We confirmed the existence of two highly diverging clades also in the study area 1898 1899 (maximum diversification 2.9%). The clades from S-W Mediterranean is the only occurring on the studied islands, while on the mainland it coexists with the clade typical of the Italian 1900 1901 peninsula. According to this pattern the species showed a high Dst (0.224) but a medium Gst (0.260). The occurrence on Pianosa and Giglio of haplotypes not found on the mainland 1902 1903 resulted in an island uniqueness value of 0.091.



Figure S41. PCoA projection (left) of genetic distances among specimens (dots) in the bidimensional RGB colour space. Individual colours have been subsequently plotted on the map (right). The first two PCoA axes explained 85.40% of variance.

1912 Maniola jurtina (Linnaeus, 1758)

1913

#### 1914 **1900-2005**

- 1915 Elba: 1908, 1916, 1919, 1921, 1963, 1967-1970, 1972, 1980, 2000 (Biermann & Hesch, 1982;
  1916 Balletto et al. 2007), 2002 (authors).
- 1917 Capraia: 1915, 1978 (Biermann & Hesch, 1982; Balletto et al. 2007), 1968-1970 (Gross,1918 1970).
- 1919 Pianosa: 1998 (Dapporto et al., 1999).

1920 Giglio: 1908, 1967, 1969, 1970, 1979 (Biermann & Hesch, 1982, Balletto et al. 2007), 20011921 (authors).

#### 1922 **2006-2015**

- 1923 Elba: 2008, 2010, 2012, 2013, 2015 (authors).
- 1924 Pianosa: 2011, 2014 (authors).
- 1925 Giglio: 2007, 2008, 2010, 2012, 2014 (authors).

1926

It has been recognized since decades that this species shows two main lineages in Europe 1927 differing for morphological and genetic characters (e.g. Thomson 1987; Schmitt et al. 2005; 1928 1929 Dapporto et al. 2009; 2014). The study area is one of the contact and hybrid zones for the two lineages (Thomson 1987; Dapporto et al. 2014). We confirmed this observation showing 1930 1931 that this species has a maximum divergence of 1.2% in the study area and that the two 1932 lineages coexist on Elba and Corsica with similar frequencies. Accordingly, the parameters 1933 for populations differentiation were high (Dst 0.298, Gst 0.603). The occurrence on Elba and 1934 Pianosa of two haplotypes not found on the mainland resulted in an island uniqueness value 1935 of 0.085.



Figure S42. PCoA projection (left) of genetic distances among specimens (dots) in the bidimensional RGB colour space. Individual colours have been subsequently plotted on the map (right). The first two PCoA axes explained 93.34% of variance.

1944 *Hipparchia neomiris* (Godart, 1822)

1945

1946 **1900-2005** 

1947 Capraia: 1993 (Terzani, 1994).

1948 Elba: 1916, 1917, 1920, 1921, 1963, 1968, 1978, 1979, 1980, 1993, 2002 (Biermann & Hesch,
1949 1982, Balletto et al. 2007), 2002 (Favilli et al., 2003).

- 1950 **2006-2015**
- 1951 Elba: 2008-2010, 2012.

1952

1953 Not recorded in Capraia in the last 10 years.

1954

This species, endemic of the study area, showed a large diversification (maximum p-distance 2.3%) between two lineages, the first occurring on Sardinia and the second on Corsica and Elba. A previous analysis on genitalic shape also revealed a diversification among the same populations (Dapporto 2010), suggesting the need of further studies to understand the taxonomic status of the two lineages. According to such a clear spatial pattern, the species showed a high population differentiation (Dst 0.968, Gst 0.871). The status of island endemic did not allow the calculation of uniqueness.



Figure S43. PCoA projection (left) of genetic distances among specimens (dots) in the bidimensional RGB colour space. Individual colours have been subsequently plotted on the map (right). The first two PCoA axes explained 99.85% of variance.

1970 *Hipparchia aristaeus* (Bonelli, 1826)

1971

1972 The taxonomic status of the species belonging to the *H. semele* group is highly debated 1973 mostly on Mediterranean islands (e.g. Cesaroni et al. 1994; Balletto et al. 2007; Dapporto 1974 2010; Kudrna et al. 2015; Vodă et al. 2016). For COI *Hipparchia aristaeus* represents a clearly 1975 distinct clade separated by more than 4% of divergence from *H. semele* and for this reason 1976 we analysed it separately.

- 1977
- 1978 **1900-2005**
- 1979 Gorgona: 1999-2002 (Dapporto, 2002).
- 1980 Capraia: 1993 (Terzani, 1994).
- 1981 Elba: 1908, 1920, 1921, 1963, 1968, 1980, 1987, 1988, 2002 (Biermann & Hesch, 1982,
  1982 Balletto et al. 2007), 2002 (Favilli et al. 2003), 2000 (authors).
- 1983 Giglio: 1908, 1979, 1989 (Biermann & Hesch, 1982, Balletto et al. 2007).
- 1984 **2006-2015**
- 1985 Elba: 2008-2010, 2012, 2013, 2015 (authors).
- 1986 Giglio: 2010, 2013, 2014 (authors).
- 1987
- 1988 This species showed a very low genetic variation over the study area (maximum p-distance
- 1989 0.6%) and almost no spatial structure (Dst 0.019, Gst 0.119). The status of island endemic did 1990 not allow the calculation of uniqueness



Figure S44. PCoA projection (left) of genetic distances among specimens (dots) in the bidimensional RGB colour space. Individual colours have been subsequently plotted on the map (right). The first two PCoA axes explained 87.53% of variance. 

1998 Hipparchia statilinus (Hufnagel, 1766)

1999

2000 **1900-2005** 

2001 Elba: 1908, 1916, 1917, 1920, 1963, 1980 (Biermann & Hesch, 1982, Balletto et al. 2007).

- 2002 **2006-2015**
- 2003 Elba: 2009, 2012, 2013 (authors).

2004

This species showed a low diversification among haplotypes (maximum p-distance 0.5%), and their distribution was slightly spatially structured resulting in low population diversification (Dst 0.053 and Gst 0.281). The haplotypes occurring on Elba have been also found on the mainland.

2009



2010

2011

Figure S45. PCoA projection (left) of genetic distances among specimens (dots) in the bidimensional RGB colour space. Individual colours have been subsequently plotted on the map (right). The first two PCoA axes explained 85.09% of variance.

2015

### 2018 Supplementary results





2020

2021

Figure S46. PCA scatterplot for wingspan (left). Data belong to four literature sources WingspanP (Pamperis, 2009), WingspanL (Lafranchis, 2000), WingspanHR (Higgins & Riley, 1970), WingspanfemalesT female data from Tshikolovets (2011), WingspanmalesT male data from Tshikolovets (2011). Tags for butterflies are obtained by using the first letter of the genus name and the first three letters of the species name. The eigenvalue plot for the five components (right) shows that only the first PC (red dot) had an eigenvalue higher than 1.





Figure S47. PCA scatterplot for phenology traits (left). Tags for butterflies are obtained by using the first letter of the genus name and the first three letters of the species name. The eigenvalue plot for the five components (right) shows that the first two PCs (red dots) had an eigenvalue higher than 1.





Figure S48. PCA scatterplot for physiological traits (left). Tags for butterflies are obtained by using the first letter of the genus name and the first three letters of the species name. The eigenvalue plot for the five components (right) shows that the first two PCs (red dots) had an eigenvalue higher than 1.

	Host_Plants				
MobilityPC1	-0.168	MobilityPC1			
PhenolPC1	0.277	0.119	PhenolPC1		
PhenolPC2	0.277	-0.200	0.000	PhenolPC2	
EcophyPC1	-0.351	0.048	-0.308	-0.114	EcophyPC1
EcophyPC2	0.090	0.109	-0.021	0.073	0.000

Table S1. Correlations among host plants and the six traits obtained after PrincipalComponent Analysis.

- - -



2070	
2071 2072	Figure S49. The topological constraints at family and subfamily levels applied to the COI phylogenetic tree based following the butterfly phylogeny published in Heikkila et al. (2012).
2073	
2074	
2075	
2076	



Figure S50. Supports for nodes in the phylogenetic tree used in phylogenetic regression. Family and subfamily nodes costrained according to Heikkila et al. (2012) are by default characterized by a maxium support.



Figure S51 Gst values for species mapped over the phylogenetic tree based on cytochrome *c* oxidase subunit 1 (COI) tree. Host plants represent the number of plant genera on which the larva of a given species has been reported: one leaf - one genus; two leaves two to four genera, three leaves more than four genera. The sizes of the butterfly silhouettes are directly correlated with the species size (wingspan). Ecophy 1 represents the first PC of physiology traits mostly represent mean temperature and is reported with colours

2094 representing quartiles of values (red, preference for warm temperatures; blue, preference

for cold temperatures). Ecophy 2 represents the second PC of physiology traits mostly

2096 correlated with precipitation tolerance and represented in quartile from small (high

tolerance to drought) to large drops (low tolerance to drought). Phenol 1 represent quartiles
the first PC of phenology correlated with the length of flying period from shortest (one black

- 2099 sector) to longest (four black sectors). Phenol 2, mostly linked to the period of emergence,
- represent spring species appearing early in the year (black sector right-top) to species with a
- 2101 later summer-autumn appearance (black sector left-top).
- 2102



- Figure S52 Uniqueness values for species mapped over the phylogenetic tree based on
- cytochrome *c* oxidase subunit 1 (COI) tree. Symbols for traits as in the previous figure.



Figure S53. Plots for observed vs fitted values in phylogenetic multiple regression models forGst, Dst and uniqueness.



Figure S54. Plot for observed vs fitted values in phylogenetic generalized linear model for extinction in Capraia.

In the main paper, we recognised as separate units most species accepted by the Fauna Europaea Project (Karsholt & Van Nieukerken, 2013; <u>www.faunaeur.org</u>) and considered as single unit taxa displaying a minimum COI p-distance lower than 3%. We re-performed all the analyses without setting any threshold. In this case we used the Fauna Europea taxonomy to identify units. The results turned out to be very similar and they are reported below:

## 2133

	Gst	
Dst	0.790***	Dst
Uni	0.103 n.s.	0.385**

2134

# 2135 Table S2. Spearman correlation between Gst, Dst and Uni

2136

Index	Pagel's lambda	р
Gst	0.136	0.519
Dst	0.342	0.053
Uni	0.050	0.709

2137

2138 Table S3. Phylogenetic signal for Gst, Dst and Uni.

2139

2140



2143 Figure S55. Overall genetic patterns obtained after comparisons among islands based on Gst. 2144 (left) The colours obtained in the RGB projection of the Principal Coordinate Analyses have 2145 been included in (right) the map of the studied region to show similarity and contrast among 2146 areas. The locality codes are: Arg, Argentario; Cap, Capraia; Cor, Corsica; Elb, Elba; Gia, 2147 Giannutri; Gig, Giglio; Gor, Gorgona; Mon, Montecristo; Pia, Pianosa; Sar, Sardinia; Tus, 2148 Tuscany; T\_C central Tuscany coast; T\_N, northern Tuscany coast; T\_S, southern Tuscany 2149 coast. Protest between PCoA configuration and spatial location, Procrustes correlation 2150 0.641, P=0.006.
		Estimate	StdErr	t	р	
	Gst					
	Host_Plants	-0.076	0.040	-1.880	0.068	
	MobilityPC1	-0.126	0.042	-2.903	0.006	
	EcophysiolPC2	0.066	0.038	1.769	0.085	
	Dst					
	Host_Plants	-0.080	0.088	-2.520	0.016	
	MobilityPC1	-0.098	0.048	-2.040	0.048	
	EcophysiolPC2	0.122	0.032	3.860	<0.001	
	Uniqueness					
	Host_Plants	-0.048	0.026	-1.843	0.073	
	EcophysiolPC2	0.093	0.025	3.680	<0.001	
2153						
2154						
2155						
2156 2157	Table S4 Trait va Uniqueness (Uni).	ariables enter	red in the A	IC phylogenet	ic regressions for Gst, Dst a	nd
2158						
		Estimate	StdErr	z.value	e p.value	
	Elba					

Host_Plants	-0.383	0.467	-0.819	0.413	
MobilityPC1	0.410	0.481	0.851	0.395	
PhenologyPC1	-0.401	0.555	-0.723	0.469	
PhenologyPC2	-0.521	0.438	-1.190	0.234	
EcophysiolPC1	-0.843	0.622	-1.355	0.176	
EcophysiolPC2	0.790	0.568	1.392	0.164	
Capraia					
Host_Plants	-0.969	0.974	-0.995	0.320	
MobilityPC1	1.628	1.216	1.339	0.181	
PhenologyPC1	-6.470	3.055	-2.118	0.034	
PhenologyPC2					
	2.906	1.905	1.524	0.127	
EcophysiolPC1	2.906 -1.877	1.905 1.124	1.524 -1.508	0.127 0.132	

Table S5. Species traits associated to disappearance on Elba and Capraia based on phylogenetic GLM with logistic model.

2164 **References** 

2165

2166 Boillat H (2002) *Coenonympha lyllus* Esper, 1805, spec. rev. Une nouvelle approche 2167 taxinomique du complexe pamphilus. *Alexanor*, 22, 243-309.

2168

2169 Bryk F (1932) Neue Thais-Rassen. *Parnassiana*, 2, 66.

2170

Dapporto L, Cecchi B, Lo Cascio P, Sforzi A (1999) Contributi alla conoscenza
 dell'artropodofauna dell'Isola di Pianosa (Arcipelago Toscano). II. Prima nota sui
 Macrolepidotteri. (Insecta, Lepidoptera). Bollettino della Società Entomologica Italiana,
 131, 245–252.

2175

Dapporto L (2002) I Lepidotteri raccolti all'Isola di Gorgona con trappola Malaise nel periodo
 1999-2002 (Insecta, Lepidoptera). *Frustula Entomologica*, 25, 49-56.

2178

Dapporto L (2007) Cuticular lipid diversification in *Lasiommata megera* and *Lasiommata paramegaera*: the influence of species, sex, and population (Lepidoptera:
 Nymphalidae). *Biological Journal of the Linnean Society*, 91, 703-710.

2182

Dapporto L, Cini A (2007) Faunal patterns in Tuscan archipelago butterflies: the dominant
 influence is recent geography not paleogeography. *European Journal of Entomology*, 104, 497-503.

2186

Dapporto L (2008) Geometric morphometrics reveal male genitalia differences in the
 *Lasionmata megera/paramegaera* complex (Lepidoptera, Nymphalidae) and the lack of
 a predicted hybridization area in the Tuscan Archipelago. Journal of Zoological
 Systematics and Evolutionary Research, 46, 224-230.

Dapporto L, Strumia F (2008) The thorny subject of insular endemic taxonomy:
 morphometrics reveal no evidence of speciation between *Coenonympha corinna* and
 *Coenonympha elbana* butterflies (Lepidoptera: Nymphalidae). *Zootaxa*, 1755, 47-56.

2195

Dapporto L, Bruschini C, Baracchi D, Cini A, Gayubo SF, Gonzalez JA, Dennis RLH (2009)
Phylogeography and counter-intuitive inferences in island biogeography: evidence from
morphometric markers in the mobile butterfly *Maniola jurtina* (Linnaeus)(Lepidoptera,
Nymphalidae). *Biological Journal of the Linnean Society*, 98, 677-692.

2200

Dapporto L (2010) Satyrinae butterflies from Sardinia and Corsica show a kaleidoscopic
 intraspecific biogeography (Lepidoptera, Nymphlidae). *Biological Journal of the Linnean Society*, 100, 195-212.

2204

Dapporto L, Vodă R, Dincă V, Vila R (2014) Comparing population patterns for genetic and
 morphological markers with uneven sample sizes. An example for the butterfly *Maniola jurtina*. *Methods in Ecology and Evolution*, 5, 834-843.

2208

Dincă V, Montagud S, Talavera G, Hernández-Roldán J, Munguira ML, García-Barros E, Hebert
 PD, Vila R (2015) DNA barcode reference library for Iberian butterflies enables a
 continental-scale preview of potential cryptic diversity. *Scientific reports*, *5*.

2212

Favilli L, Piazzini S, Manganelli G (2003) Nuovi dati sulla distribuzione in Toscana meridionale
 ed insulare di alcuni lepidotteri diurni (Hesperoidea, Papilionoidea) rari o poco noti. *Atti della Società toscana di scienze naturali residente in Pisa: memorie. Serie B*, **110**, p.25.

2216

Hernández-Roldán JL, Dapporto L, Dincă V, Vicente JC, Hornett EA, Šíchová J, Lukhtanov VA,
 Talavera G, Vila R (2016) Integrative analyses unveil speciation linked to host plant shift
 in Spialia butterflies. Molecular Ecology, 25, 4267-4284.

Higgins LG, Riley ND (1970) A field guide to the butterflies of Britain and Europe. Collins,London.

2223

2224 John E, Wiemers M, Makris C, Russell P (2013) The Pontia daplidice (Linnaeus, 1758)/Pontia 2225 edusa (Fabricius, 1777) complex (Lepidoptera: Pieridae): confirmation of the presence of 2226 Pontia daplidice in Cyprus, and of Cleome iberica DC. as a new host-plant for this species 2227 in the Levant. Entomologist's Gazette, 64, 69-78. 2228 2229 Kodandaramaiah U, Wahlberg N (2009) Phylogeny and biogeography of Coenonympha 2230 butterflies (Nymphalidae: Satyrinae)-patterns of colonization in the Holarctic. Systematic Entomology, 34, 315-323. 2231 2232 2233 Kudrna O (1977) On the Status of *Lasiommata paramegaera* (Hübner) (Lep., Nymphalidae). 2234 Atalanta, 8, 290-293. 2235 Lafranchis T (2000) Les Papillons de jour de France, Belgique et Luxembourg et leurs 2236 2237 chenilles. Biotope Éditions, France. 2238 2239 Meirmans, P.G., Hedrick, P.W. (2011) Assessing population structure: FST and related 2240 measures. Molecular Ecololgy Resources, 11, 5-18. 2241 Nei, M. (1987) Molecular Evolutionary Genetics, Columbia University Press. 2242 2243 Pamperis LN (2009) The Butterflies of Greece. Second edition. Editions Pamperis, Greece. 2244 2245 2246 Sañudo-Restrepo CP, Dincă V, Talavera G, Vila R (2013) Biogeography and systematics of Aricia butterflies (Lepidoptera, Lycaenidae). Molecular Phylogenetics and Evolution, 66, 2247 2248 369-379. 2249

Schmitt T, Röber S, Seitz A (2005) Is the last glaciation the only relevant event for the present
 genetic population structure of the Meadow Brown butterfly Maniola jurtina
 (Lepidoptera: Nymphalidae)? *Biological Journal of the Linnean Society*, 85, 419-431.

- 2254Thomson G (1987) Enzyme variation at morphological boundaries in Maniola and related2255genera (Lepidoptera: Nymphalidae: Satyrinae). DPhil Thesis, University of Stirling.
- Tshikolovets VV (2011) Butterflies of Europe and the Mediterranean area. TshikolovetsPublications, Czech Republic.

Vandewoestijne S, Baguette M, Brakefield PM, Saccheri IJ (2004) Phylogeography of *Aglais urticae* (Lepidoptera) based on DNA sequences of the mitochondrial COI gene and
 control region. *Molecular Phylogenetics and Evolution*, 31, 630-646.

Vodă R, Dapporto L, Dincă V, Shreeve TG, Khaldi M, Barech G, Rebbas K, Sammut P, Scalercio
 S, Hebert PD, Vila R (2016) Historical and contemporary factors generate unique
 butterfly communities on islands. *Scientific Reports*, 6.

Weingartner E, Wahlberg N, Nylin S (2006) Speciation in *Pararge* (Satyrinae: Nymphalidae)
 butterflies–North Africa is the source of ancestral populations of all Pararge
 species. *Systematic Entomology*, 31, 621-632.