1 Opinion

2	Temporal instability of evidence base: a threat to policy making?
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16 Abstract

18	A shift towards evidence-based conservation and environmental management over the last
19	two decades has resulted in an increased use of systematic reviews and meta-analyses as
20	tools to combine the existing scientific evidence. However, to guide policy making decisions
21	in conservation and management the conclusions of meta-analyses need to remain stable
22	for at least some years. Alarmingly, numerous recent studies indicate that the magnitude,
23	statistical significance and even the sign of the effects reported in the literature might
24	change over relatively short time periods. We argue that such rapid temporal changes in
25	cumulative evidence represent a real threat to policy making in conservation and
26	environmental management and call for systematic monitoring of temporal changes in
27	evidence and exploration of their causes.
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evidence at a more or less arbitrary point in time whereas scientific evidence is not static
and tends to change over time as more research on the topic accumulates [7]. New studies
may either strengthen or challenge the conclusions of previous reports. If the above
changes in cumulative evidence over time are rapid and of considerable magnitude, the
conclusions of meta-analysis will strongly depend on when the review was conducted and
the policy-relevant recommendations derived from these reviews will quickly go out of date.

47 Worryingly, a growing number of studies demonstrates that substantial changes in the magnitude, statistical significance or even sign of the reported effects over time are 48 49 common in ecology and evolutionary biology [8-13] as well as other disciplines [14-17]. In most cases decreases in the magnitude of the estimated effect are reported over time, a 50 phenomenon which has been dubbed 'a **decline effect**' in some fields [18]. As a result, the 51 52 conclusions of systematic reviews and meta-analyses may go out of date very rapidly as 53 well. For instance, a survey of 100 meta-analyses in medicine showed that clinically 54 important evidence that alters review conclusions about the effectiveness and harms of 55 treatments can accumulate within relatively short time frames, i.e. 2-5 years [19]. While no similar surveys have been conducted in ecology and evolution, meta-analyses in these fields 56 57 are often performed on topics where results of studies are contradictory, sample sizes are 58 low, and the expected magnitudes of the effects are relatively small [20]. This makes 59 temporal changes in cumulative evidence more likely. The failure of later studies to reproduce the results of the earlier studies exemplifies a broader concern about the 60 61 reproducibility in science [21].

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63 Despite its obvious scientific and practical importance, temporal changes in evidence base for conservation and environmental management have received little attention so far [7]. In 64 this Opinion piece we review possible causes of such temporal trends, draw attention 65 66 towards their potential implications for policy making and evidence-based conservation, 67 and discuss the methods of detection of temporal changes. We argue that rapid temporal changes in cumulative evidence represent a real threat to policy making in conservation and 68 69 environmental management and call for systematic exploration of their extent and causes in 70 applied ecology.

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72 Causes of temporal instability of the evidence base

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74 Temporal changes in reported effects may occur for three main reasons. First, temporal 75 trends may reflect true changes in the magnitude or direction of a biological effect, e.g. due 76 to shifts in the strength and relative importance of the drivers of biodiversity loss [22-24] 77 and to rapid adaptive evolution [25]. A well-known example in medicine is the development of antibiotic resistance which might decrease treatment efficacy over time [26]. Similar 78 79 adaptive responses may occur in ecological and evolutionary studies as a result of selection 80 pressure imposed by humans directly or indirectly. Examples of such changes include 81 reductions in body size in animals as a result of warming temperatures [27-29] and shifting 82 song frequencies in birds in response to anthropogenic noise [30]. As the above selection pressures increase over time, it is likely that studies published few decades ago would 83 84 report smaller effects compared to the more recent studies.

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86 Second, temporal trends in estimated **effect sizes** may occur even when the true effect size remains the same, but the proportion of studies with particular characteristics which 87 influence the magnitude and direction of the effect (known as **moderators** in meta-analysis) 88 89 changes over time. An example of such evidence reversal is discussed in Box 1. If there is 90 significant heterogeneity in effect sizes (i.e. not all studies share the same effect) and 91 effects are smaller or larger under particular conditions, any changes in frequency of studies 92 on the above condition over time relative to other conditions may result in corresponding 93 temporal changes in the magnitude of the overall estimated effect (Box 1, [11, 31]). Changes 94 in prevalence of particular research or statistical methods over time may also result in 95 similar effects if such methods differ in the magnitude of the estimated effects that they produce [32, 33]. It is therefore crucial to examine the amount of heterogeneity and its 96 causes in a meta-analysis, particularly as high heterogeneity should be expected in 97 98 ecological and evolutionary studies [34].

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100 Third, changes in magnitude and significance of the effect size estimates over time may be 101 due to biases. Here, again, the true magnitude of the effect size might not change with time, but the estimate of the effect does. For instance, time lag in the publication of studies with 102 non-significant results may lead to decrease in the cumulative effect over time as the 103 104 number of studies with weak and non-significant effects increases. Jennions and Møller [9] 105 suggested that such time-lag bias against non-significant results is the most probable cause of the observed decrease in estimated effect sizes with time in ecological and evolutionary 106 meta-analyses. However, no studies so far have explored the relative importance of 107 108 different causes of temporal trends in reported effect sizes in ecology and evolution. On the 109 other hand, **publication bias** may also lead to overestimation of the overall effect. Nuijten et

110 al. [35] showed that if both the original study and its conceptual replication are subject to publication bias, combining the two studies to obtain an overall effect size will result in an 111 112 overestimation of the population effect size. Biases may also prevent the cumulative effects 113 from reaching statistical significance. For instance, the attractiveness of contradictory findings to researchers and editors may lead to publication of the succession of extreme 114 115 positive and negative effects, hence hindering the stabilization of the cumulative effect size 116 over time [36, 37]. Heleno [37] argued that the consequences of the "editorial love of 117 controversy" may be particularly severe in conservation-led decisions and might contribute 118 to an underestimation of the impacts of human pressure on the environment. Other biases 119 which may lead to temporal changes in cumulative evidence include bias in choice of study 120 organisms [12] and paradigm shifts [38].

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122 It is important to distinguish between the above causes of temporal changes in reported 123 effects because they determine whether the current conservation or management policy needs to be modified. If true biological effects are changing over time, then actions might 124 125 need to be taken to re-evaluate conservation status and conservation strategy for the given species or environmental management options might need to be reconsidered. On the 126 127 other hand, if temporal changes in estimated effect sizes are due to heterogeneity among 128 studies, the sources of this heterogeneity have to be identified to find out under what 129 conditions the proposed management and conservation strategies are effective. Examination of temporal trends in effect sizes is thus a good diagnostic tool for detection of 130 131 sources of heterogeneity. Finally, testing for presence of biases in a meta-analysis is absolutely essential, although it might be sometimes difficult to distinguish them from true 132 133 heterogeneity [39].

Potential implications of temporal changes in estimated effect sizes

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The magnitude and direction of the mean effect size and the breadth of its confidence 137 interval largely determine the conclusions drawn from a meta-analysis [4]. If the magnitude, 138 139 statistical significance or the direction of the estimated effect changes over time, any policy 140 recommendations derived from a meta-analysis are likely to change as well. In Box 1 we 141 show how two meta-analyses on the same topic conducted several years apart reached opposite conclusions on effectiveness of the same conservation measure. Such reversals in 142 conclusions of meta-analyses represent an example of evidence reversal, a phenomenon 143 that has only recently became a topic of formal exploration [40]. Reversals of evidence can 144 have significant impacts on evidence-based conservation and environmental management 145 146 and might necessitate revision of already implemented policies based on recommendations 147 from the previous meta-analysis. 148 149 Moreover, evidence reversals may affect not only the effectiveness of the currently implemented policies and measures, but also the society's and researcher's faith in the 150 approach to assessment of scientific evidence base. For instance, differences in the 151 152 conclusions between several meta-analyses on the same topic have sometimes led to 153 questioning whether meta-analyses constitute repeatable science [41]. While the results of two meta-analyses can differ for many other reasons (e.g. different inclusion criteria, 154 different statistical models and moderators tested), at present we do not know what 155 proportion of ecological meta-analyses on the same topic arrived to different conclusions 156

157 because of temporal changes in the estimated effect sizes.

159	Conversely, a lack of temporal changes in the estimated effect sizes may also convey
160	important policy information when changes in effectiveness over time are expected. For
161	instance, agri-environment schemes (AES) in Europe have been used for ca 25 years and are
162	the biggest conservation expenditure in Europe [42]. National AES programs are revised
163	every 7 years allowing countries to use novel scientific insights and modify their agri-
164	environmental programs to increase their efficiency. However, a meta-analysis by Batáry et
165	al. [42] showed that effectiveness of AES has not changed as a result of the revision of the
166	EU's agri-environmental programmes in 2007. The authors point out that this lack of
167	increase in effectiveness over time is worrying in view of forthcoming reductions in AES
168	budget as it is unlikely that increased effectiveness of the scheme will compensate for the
169	future budget cuts.
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172	Testing for temporal trends and updating the results of systematic reviews and meta-
173	analyses
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175	Several relatively simple and straightforward statistical approaches which allow testing for
176	temporal trends in estimated effect sizes are available (reviewed in [7, 43, 44], and Box 2),
177	but are unfortunately seldom used by ecologists. For instance, only 5% of 322 meta-analyses
178	in plant ecology published between 1996 and 2013 have tested for temporal changes in
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	estimated effects [45]. We argue that such tests have to become a routine part of ecological
180	estimated effects [45]. We argue that such tests have to become a routine part of ecological meta-analyses and one of the important criteria for review quality control evaluation [46].

publication year as a moderator into meta-regression [13, 16] (Figure IIA). A cumulative
meta-analysis (CMA) in which studies are entered into the analysis in chronological order
provides another useful tool for detection of changes in cumulative evidence over time [47].
As all visual tools, CMA plots might be subject to misinterpretation and should be
supplemented by formal statistical methods which should take into account multiple testing
inherent in CMA [44]. Therefore, we recommend the use of cumulative meta-analysis in
combination with control plots [44](Box 2), which can be plotted using R package qcc [48].

Another class of methods has been developed for sequential clinical trials in medicine where 190 191 the accumulated evidence is periodically reviewed as the trial progresses with a view of stopping the trial early if required. Applications of these techniques to meta-analysis exist 192 [49-52], but we do not recommend their use (see critique of these approaches in [53, 54]). 193 194 Furthermore, some ecological meta-analyses assess temporal changes in effect sizes by 195 subdividing studies into groups based on the publication year (e.g. by decades or published 196 before and after year X) and comparing mean effect sizes between the groups [42, 55]. This 197 relatively crude approach ignores likely gradual character of temporal changes and their possible occurrence within as well as between the studied groups, therefore we do not 198 199 recommend it.

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Use of tests for temporal changes in estimated effect sizes within individual meta-analyses may prove particularly effective if such changes occur mainly early on. For instance, Fanelli et al. [56] have recently shown that declines in magnitude of the effect sizes with publication year in meta-analyses are not linear and there is a strong "first-year" effect, in which the earliest studies are more likely to overestimate the overall effect than all later

ones. This effect might occur if early studies are statistically underpowered [57]. As a result,
 first meta-analyses on the topic based on the first few early primary studies available are
 particularly likely to overestimate the effect and results of such meta-analyses need to be
 treated with caution.

210

211 In addition to testing for temporal trends within meta-analyses, updating existing meta-212 analyses can also be an effective tool in early detection of evidence reversal. Useful 213 guidelines on when and how to update systematic reviews have been recently published by the Cochrane panel [58]. In order to enable such an update, the transparency of methods 214 215 used in the published ecological meta-analyses needs to improve. For instance, the 216 database on which previous meta-analysis has been based need to be available as well as the detailed literature search strategy. Unfortunately, the majority of published ecological 217 218 meta-analyses do not fulfil these criteria [45]. Another problem is that publication of meta-219 analyses and any subsequent updates can take many months, which means that by the time 220 of publication these reviews are already out of date. Shojania et al. [19] proposed that when 221 the process of submission and rejection from other journals has resulted in the passage of more than one year from the date of the previous search, authors should update the search 222 223 before resubmission. Another approach to narrowing the time gap between evidence and 224 practice and to reducing the evidence reversal impact is to conduct living systematic 225 reviews, online summaries updated as new research becomes available [59]. This approach, however, similarly to cumulative meta-analysis, might inflate the rate of false-positive 226 findings due to repeated testing. Therefore, previously discussed methods or the Bayesian 227 228 approach discussed in Elliott et al. [52] should be be used for monitoring accumulating 229 evidence while reducing the probability of false positives.

231 Concluding Remarks and Future Perspectives

233	We believe that more widespread application of methods for monitoring of temporal
234	changes in reported effects (Box 2) and for updating meta-analyses will facilitate
235	conclusions on sufficiency of evidence for policy making and timely detection of evidence
236	reversal. Moreover, analysis of causes of temporal changes in cumulative evidence will
237	reveal whether these changes require adjustment in previously accepted management
238	policies. Ultimately this will allow saving of time and resources in the development of
239	management strategies thus making conservation action more effective.

241 Box 1. An example of evidence reversal in conservation biology

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Two meta-analyses on effects of predator removal on bird population provide a good 243 example of how heterogeneity in effect size can lead to evidence reversal and change the 244 conclusions and practical recommendations. The first meta-analysis by Coté and Sutherland 245 246 [60] showed that predator removal significantly increases postbreeding population sizes (i.e. 247 autumn densities) of the target bird species, but does not significantly affect breeding 248 population sizes (Fig. I). Coté and Sutherland concluded therefore that predator removal 249 fulfils the goal of game management (enhancing harvestable postbreeding populations) but 250 is of less use for conservation management (increasing bird breeding population sizes). 251 However, a more recent meta-analysis on the same topic by Smith et al. [61] arrived at the 252 opposite conclusion, showing that the predator removal effect on breeding population 253 numbers is statistically significant, but the effect of predator removal on postbreeding 254 populations is no longer significant (Fig. I). Smith et al. concluded therefore that predator 255 removal is an effective strategy for the conservation of bird populations, but not for game 256 management. Hence, two meta-analyses on the same topic conducted 13 years apart reached opposite conclusions on the effectiveness of the assessed conservation measures. 257 In this particular case the difference in the results of the two meta-analyses was not due to 258 259 changes in true biological effects but due to heterogeneity. Smith et al. have revealed that 260 predator removal was effective in increasing postbreeding bird populations on mainland, but not on islands. Since the proportion of studies conducted on islands increased with time 261 and was higher in meta-analysis by Smith et al. than in the earlier meta-analysis on the same 262 topic by Coté and Sutherland, the magnitude of the overall effect estimate of predator 263 264 removal on postbreeding populations was much smaller in the former meta-analysis. This

- 265 example shows the importance of updating the results of previous meta-analyses as new
- studies on the topic are published as well as the importance of examining the sources of
- 267 variation in effect sizes and drawing inference from studies conducted under similar
- 268 ecological conditions.
- 269

270 Box 2. Methods of detection of temporal changes in reported effects

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The simplest way to visualize a potential temporal trend in a meta-analytic dataset is by 272 273 plotting effect sizes from individual primary studies against their publication years (Fig. IIA). 274 In order to statistically test the above relationship, publication year can be used as a 275 moderator in a meta-regression model [13, 16, 62]. Alternatively, cumulative meta-analysis 276 (CMA) where studies are added to the analysis in chronological order and meta-analytic 277 means are cumulatively calculated over the years can be used to visually detect temporal trends (Fig. IIB, [47]). Finally, methods of statistical quality control such as Xbar charts and 278 279 **CUSUM charts** can be used to detect possible outliers and trends over time in meta-analysis 280 [44, 63]. Xbar charts are based on detecting outlying observations under normality. The control limits on Xbar charts are usually plotted at 3 standard deviations, corresponding to a 281 282 significance level of α = 0.0027. The CUSUM charts plot the cumulative sums of the 283 deviations of the sample values from a target value. The chart is restricted from falling 284 below zero, and often two one-sided CUSUM charts (for positive and negative deviations) 285 are plotted simultaneously.

286

We demonstrate the application of four different methods for detection of temporal trends in effect sizes on Figure II using a subset from the meta-analysis by Batáry et al. [64] on effects of agri-environment schemes on biodiversity as an example. A bubble plot (Fig. IIA) shows decrease in effect sizes with publication year, particularly between 1995 and 2005. The cumulative meta-analysis plot (Fig. IIB) demonstrates similar trend with initial increase of the effect until the fourth study was added to the analysis and the subsequent decrease in the magnitude of the effect. The cumulative effect size becomes significantly different

294	from 0 at study 6, and even more so at study 7, but then the effect declines as more studies
295	are added to the analysis. In this example, the effect size reached at study 7 (d= 1.165) is
296	monitored over time. The Xbar chart (Fig. IIC) shows one high outlier (study 4), two low
297	outliers (studies 11 and 14) and one significant run rule violation (a series of more than 7
298	negative deviations from the target value), suggesting a shift in the process mean. CUSUM
299	chart (Fig. IID) shows that while the cumulative effects were significantly above 1.165 at
300	studies 4 and 5, the cumulative results are significantly below this value for the last 4
301	studies, indicating a decrease in the mean effect size.

303 Figures:



Fig. I. Differences in estimates of the effects of predator removal on postbreeding and

306 breeding population size of birds (data from meta-analyses by Côté and Sutherland [60]

and Smith et al. [61]). Error bars represent 95% confidence intervals; mean effects are not

308 significantly different from 0 if confidence intervals include 0. Number of studies included in

the analysis: 13 and 51 for breeding population size estimates and 10 and 19 for

- 310 postbreeding population size estimates in Côté and Sutherland and Smith et al.,
- 311 respectively.

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321 landscapes within croplands and published before 2006. Effect sizes are standardized mean 322 differences (Hedges' d) between biodiversity measures in extensively and intensively 323 managed fields. A: a bubble plot showing the results of meta-regression with publication year as a moderator. Effect sizes are weighted by their precision; larger bubbles indicate 324 325 more precise estimates and smaller bubbles less precise. B: cumulative meta-analysis showing changes in cumulative mean effect size and the 95% confidence interval as more 326 recent studies are added in the analysis. C. Xbar chart. Horizontal central line on Xbar chart 327 328 corresponds to the combined effect size of the first seven studies (d= 1.165). D. CUSUM chart. Control limits (dashed lines) are at ±3SD, out-of-control values are in red, run test 329 violations (a series of consecutive deviations from the expected value which are of the same 330 331 sign) are in orange.

332

334	Glossary
335	
336	Cumulative meta-analysis: a type of meta-analysis in which effect sizes from individual
337	studies are entered into the analysis sequentially, one study at the time, based on some
338	predetermined order (most commonly chronological); the mean effect size and confidence
339	intervals are recalculated at each step.
340	
341	CUSUM chart: a cumulative sum (CUSUM) chart is a type of control chart used to monitor
342	changes in the process mean. It plots the cumulative sum of deviations of the sample values
343	from a target value.
344	
345	Decline effect: decrease in support for scientific claims over time as original studies are
346	repeated.
347	
348	Effect size: a quantitative measure of the magnitude of study outcome that puts all
349	responses across studies in a meta-analysis on the same scale. It provides a "common
350	currency" for comparisons of the results across studies. Metrics of effect size most
351	commonly used in ecology include standardized mean differences, response ratios and
352	correlation coefficients.
353	
354	Evidence-based conservation: conservation management actions and policy making based
355	on systematic assessment (e.g. systematic review and meta-analysis) of existing scientific
356	evidence of current effectiveness of different management interventions.
357	

358	Evidence reversal: occurs when an existing claim is tested and the original evidence is
359	contradicted by new evidence.
360	
361	Heterogeneity: the variation in the effect size estimates among studies.
362	
363	Meta-analysis: a set of statistical methods for combining magnitudes of the effects across
364	different data sets addressing the same research question.
365	
366	Meta-regression: an extension of basic meta-analysis model in which moderators are used
367	to explain between-study variation in effect sizes (heterogeneity).
368	
369	Moderator: a variable (continuous or categorical) which is used in meta-regression to
370	explain between-study variation in effect sizes.
371	
372	Publication bias: influence of magnitude, direction, and/or statistical significance of
373	research findings on the probability of a study to be published.
374	
375	Systematic review: the type of research synthesis on a precisely defined topic using
376	systematic and explicit methods to identify, select, critically appraise, and analyse relevant
377	research. Systematic review may or may not include <i>meta-analysis</i> of the data.
378	
379	Time-lag bias: influence of study results on the time it takes to complete and publish a
380	study; often refers to delayed publication of non-significant results.
381	

382 **Xbar** (\overline{X}) chart: a type of control chart that is used to monitor the means of successive 383 samples based on detecting outlying observations under normality. The control limits on 384 Xbar charts are usually plotted at 3 standard deviations, corresponding to a significance 385 level of $\alpha = 0.0027$.

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