A Continuous Measure of Gross Primary Production for the Conterminous U.S. Derived from MODIS and AmeriFlux Data

Jingfeng Xiao, Qianlai Zhuang, Beverly E. Law, Jiquan Chen, Dennis D. Baldocchi, David R. Cook, Ram Oren, Andrew D. Richardson, Sonia Wharton, Siyan Ma, Timothy A. Martin, Shashi B. Verma, Andrew E. Suyker, Russell L. Scott, Russell K. Monson, Marcy Litvak, David Y. Hollinger, Ge Sun, Kenneth J. Davis, Paul V. Bolstad, Sean P. Burns, Peter S. Curtis, Bert G. Drake, Matthias Falk, Marc L. Fischer, David R. Foster, Lianhong Gu, Julian L. Hadley, Gabriel G. Katul, Roser Matamala, Steve McNulty, Tilden P. Meyers, J. William Munger, Asko Noormets, Walter C. Oechel, Kyaw Tha Paw U, Hans Peter Schmid, Gregory Starr, Margaret S. Torn, Steven C. Wofsy

This work was supported by the Director, Office of Science, Office of Basic Energy Sciences, of the U.S. Department of Energy under Contract No. DE-AC02-05CH11231.

1

2

A Continuous Measure of Gross Primary Production for the Conterminous U.S. Derived from MODIS and AmeriFhux Data

3

-	
4	Jingfeng Xiao ^{1*} , Qianlai Zhuang ² , Beverly E. Law ³ , Jiquan Chen ⁴ , Dennis D. Baldocchi ⁵ ,
5	David R. Cook ⁶ , Ram Oren ⁷ , Andrew D. Richardson ⁸ , Sonia Wharton ⁹ , Siyan Ma ⁵ , Timothy A.
6	Martin ¹⁰ , Shashi B. Verma ¹¹ , Andrew E. Suyker ¹¹ , Russell L. Scott ¹² , Russell K. Monson ¹³ ,
7	Marcy Litvak ¹⁴ , David Y. Hollinger ¹⁵ , Ge Sun ¹⁶ , Kenneth J. Davis ¹⁷ , Paul V. Bolstad ¹⁸ , Sean P.
8	Burns ¹³ , Peter S. Curtis ¹⁹ , Bert G. Drake ²⁰ , Matthias Falk ⁹ , Marc L. Fischer ²¹ , David R.
9	Foster ²² , Lianhong Gu ²³ , Julian L. Hadley ²⁴ , Gabriel G. Katul ⁷ , Roser Matamala ²⁵ ,
10	Steve McNulty ¹⁶ , Tilden P. Meyers ²⁶ , J. William Munger ²⁷ , Asko Noormets ²⁸ , Walter C.
11	Oechel ²⁹ , Kyaw Tha Paw U ⁹ , Hans Peter Schmid ³⁰ , Gregory Start ³¹ , Margaret S. Torn ³² ,
12	Steven C. Wofsy ³³
13	Department of Earth & Atmospheric Sciences, Purdue Climate Change Research Center,
14	Purdue University, West Lafayette, IN 47907, USA
15	² Department of Earth & Atmospheric Sciences, Department of Agronomy, Purdue Climate
16	Change Research Center, Purdue University, West Lafayette, IN 47907, USA
17	³ College of Forestry, Oregon State University, Corvallis, OR 97331, USA
18	⁴ Department of Environmental Sciences, University of Toledo, Toledo, OH 43606, USA
19	⁵ Ecosystem Science Division, Department of Environmental Science, Policy and Management,
20	University of California, Berkeley, Berkeley, CA 94720, USA
21	⁶ Argonne National Laboratory, Environmental Science Division, Argonne, IL 60439, USA

22 ⁷Nicholas School of the Environment, Duke University, Durham, NC 27708, USA

- 23 ⁸Complex Systems Research Center Institute for the Study of Earth, Oceans and Space,
- 24 University of New Hampshire, Durham, NH 03824, USA
- ⁹Department of Land, Air and Water Resources, University of California, Davis, Davis, CA
 95616, USA
- 27 ¹⁰School of Forest Resources & Conservation, University of Florida, Gainesville, FL 32611,
- 28 USA
- 29 ¹¹School of Natural Resources, University of Nebraska-Lincoln, Lincoln, NE 68583, USA
- 30 ¹²USDA-ARS Southwest Watershed Research Center, Tucson, AZ 85719, USA
- 31 ¹³Department of Ecology and Evolutionary Biology, University of Colorado, Boulder, CO
- 32 80309, USA
- 33 ¹⁴Department of Biology, University of New Mexico, Albuquerque, NM 87131, USA
- 34 ¹⁵USDA Forest Service, Northeastern Research Station, Durham, NH 03824, USA
- 35 ¹⁶USDA Forest Service, Southern Research Station, Raleigh, NC 27606, USA
- 36 ¹⁷Department of Meteorology, Pennsylvania State University, University Park, PA 16802,
- 37 USA
- 38 ¹⁸Department of Forest Resources, University of Minnesota, St. Paul, MN 55108, USA
- 39 ¹⁹Department of Evolution, Ecology, and Organismal Biology, Ohio State University,
- 40 Columbus, OH 43210, USA
- 41 ²⁰Smithsonian Environmental Research Center, Edgewater, MD 21037, USA
- 42 ²¹Lawrence Berkeley National Laboratory, Environmental Energy Technologies Division,
- 43 Atmospheric Science Department, Berkeley, CA 94720, USA
- 44 ²²Harvard Forest and Department of Organismic and Evolutionary Biology, Harvard
- 45 University, Petersham, MA 01366, USA

46 ²³Oak Ridge National Laboratory Environmental Sciences Division, Oak Ridge, TN 37831,

47 USA

48 ²⁴Harvard Forest, Harvard University, Petersham, MA 01366, USA

- 49 ²⁵Argonne National Laboratory, Biosciences Division, Argonne, IL 60439, USA
- 50 ²⁶NOAA/ARL, Atmospheric Turbulence and Diffusion Division, Oak Ridge, TN 37831, USA
- 51 ²⁷Department of Earth and Planetary Sciences, Harvard University, Cambridge, MA 02138,

52 USA

- 53 ²⁸Department of Forestry and Environmental Resources and Southern Global Change Program,
- 54 North Carolina State University, Raleigh, NC 27695, USA
- 55 ²⁹Department of Biology, San Diego State University, San Diego, CA 92182, USA
- 56 ³⁰Department of Geography, Indiana University, Bloomington, IN 47405, USA; Atmospheric
- 57 Environmental Research, Institute of Meteorology and Climate Research, Research Center
- 58 Karlsruhe (FZK/IMK-IFU), Kreuzeckbahnstr, 19, 82467 Garmisch-Partenkirchen, Germany
- 59 ³¹Department of Biological Sciences, University of Alabama, Tuscaloosa, AL 35487, USA
- 60 ³²Lawrence Berkeley National Laboratory, Earth Science Division, Berkeley, CA 94720, USA
- 61 ³³Division of Engineering and Applied Science/Department of Earth and Planetary Science,
- 62 Harvard University, Cambridge, MA 02138, USA
- *Corresponding author: <u>jing@psu.edu</u>. Now at Department of Meteorology, Pennsylvania
 State University, University Park, PA 16802
- 65 Keywords: Gross Primary Productivity; MODIS; AmeriFlux; Eddy Covariance; Regression
- 66 Tree; US; Carbon Fluxes; Interannual Variability; Satellite Data; Biomes
- 67 Submitted to Remote Sensing of Environment on 01/28/2009
- 68 Revised: 10/08/2009

69 Abstract The quantification of carbon fluxes between the terrestrial biosphere and the 70 atmosphere is of scientific importance and also relevant to climate-policy making. Eddy 71 covariance flux towers provide continuous measurements of ecosystem-level exchange of 72 carbon dioxide spanning diurnal, synoptic, seasonal, and interannual time scales. However, 73 these measurements only represent the fluxes at the scale of the tower footprint. Here we used 74 remotely-sensed data from the Moderate Resolution Imaging Spectroradiometer (MODIS) to 75 upscale gross primary productivity (GPP) data from eddy covariance flux towers to the 76 continental scale. We first combined GPP and MODIS data for 42 AmeriFlux towers 77 encompassing a wide range of ecosystem and climate types to develop a predictive GPP model 78 using a regression tree approach. The predictive model was trained using observed GPP over 79 the period 2000-2004, and was validated using observed GPP over the period 2005-2006 and 80 leave-one-out cross-validation. Our model predicted GPP fairly well at the site level. We then 81 used the model to estimate GPP for each 1 km \times 1 km cell across the U.S. for each 8-day 82 interval over the period from February 2000 to December 2006 using MODIS data, Our GPP 83 estimates provide a spatially and temporally continuous measure of gross primary production for the U.S. that is a highly constrained by eddy covariance flux data. Our study demonstrated 84 85 that our empirical approach is effective for upscaling eddy flux GPP data to the continental 86 scale and producing continuous GPP estimates across multiple biomes. With these estimates, 87 we then examined the patterns, magnitude, and interannual variability of GPP. We estimated a gross carbon uptake between 6.91 and 7.33 Pg C yr⁴ for the conterminous U.S. Drought, fires, 88 89 and hurricanes reduced annual GPP at regional scales and could have a significant impact on 90 the U.S. net ecosystem carbon exchange. The sources of the interannual variability of U.S. 91 GPP were dominated by these extreme climate events and disturbances.

92 1. Introduction

93 The quantification of ecosystem carbon fluxes for regions, continents, or the globe can 94 improve our understanding of the feedbacks between the terrestrial biosphere and the 95 atmosphere in the context of global change and facilitate climate policy decisions (Law et al. 96 2006). Gross primary productivity (GPP) is the amount of carbon fixed by vegetation through 97 photosynthesis and a key component of ecosystem carbon fluxes and the carbon balance 98 between the biosphere and the atmosphere (Mäkelä et al. 2008). The accurate estimation of 99 GPP is essential for the quantification of net ecosystem carbon exchange (NEE) as the latter is 100 often a small difference of two large carbon fluxes – GPP and ecosystem respiration (R_{a}). The 101 estimation of GPP for regions, continents, or the globe, however, can only be made by using 102 ecosystem models (e.g., Prince and Goward 1995) and/or remotely sensed data (e.g., Running 103 et al. 2004).

104 Eddy covariance flux towers have been providing continuous measurements of 105 ecosystem-level exchange of carbon, water, and energy spanning diurnal, synoptic, seasonal, 106 and interannual time scales since the early 1990s (Wofsy et al. 1993; Baldocchi et al. 2001). At 107 present, over 500 eddy covariance flux towers are operating on a long-term and continuous 108 basis around the world (FLUXNET, http://daac.ornl.gov/FLUXNET). This global network 109 encompasses a large range of climate and biome types (Baldocchi et al. 2001), and provides 110 probably the best estimates of ecosystem-level carbon fluxes. The flux towers directly measure 111 NEE that can be separated into two major components: GPP and R_e (Reichstein et al. 2005; 112 Desai et al. 2008). However, these estimates only represent fluxes at the scale of the tower 113 footprint with longitudinal dimensions ranging between a hundred meters and several 114 kilometers depending on homogeneous vegetation and fetch (Schmid, 1994; Göckede et al.,

115 2008). To quantify the exchange of CO_2 between the terrestrial biosphere and the atmosphere, 116 significant efforts are needed to scale up flux tower measurements from the stand scale to 117 landscape, regional, continental, or global scales.

118 Satellite remote sensing is a potentially valuable tool for scaling-up efforts (Running et 119 al. 1999; Turner et al. 2003; Xiao et al. 2008). Several studies have integrated flux data with 120 remote sensing data to quantify GPP over large areas. Zhang et al. (2007) estimated GPP for 121 the Northern Great Plains grasslands using satellite and flux tower data. Yang et al. (2007) 122 linked satellite observations to flux tower GPP data for the estimation of GPP for two broad 123 vegetation types in the U.S. using a machine learning approach. Despite these efforts, to our 124 knowledge, no study has upscaled AmeriFlux GPP data to the continental scale to produce 125 spatially-explicit estimates of GPP across multiple biomes and to examine the patterns, 126 magnitude, and interannual variability of GPP over the conterminous U.S.

127 Here we used a regression tree approach and remotely-sensed data from the Moderate 128 Resolution Imaging Spectroradiometer (MODIS) to upscale flux tower GPP to the continental 129 scale and produced wall-to-wall GPP estimates for multiple biomes across the conterminous 130 U.S. First, we developed a predictive GPP model based on site-specific MODIS and flux tower 131 GPP data, and validated the model using eddy flux data in both temporal and spatial domains. 132 Second, we applied the model to estimate GPP for each 1 km \times 1 km cell across the 133 conterminous U.S. for each 8-day interval over the period 2000-2006 using wall-to-wall 134 MODIS data. Third, we examined the patterns, magnitude, and interannual variability of GPP 135 across the conterminous U.S.

136 2. Data and Methods

137 2.1. Regression tree approach

138 We used a modified regression tree approach implemented in the commercial software, 139 Cubist, to upscale flux tower GPP to the continental scale. Regression tree algorithms typically 140 predict class membership by recursively partitioning a dataset into more homogeneous subsets. 141 The partitioning process splits each parent node into two child nodes, and each child node is 142 treated as a potential parent node. Regression tree models can account for a aonlinear 143 relationship between predictive and target variables and allow both continuous and discrete 144 variables. Previous studies showed that regression tree methods are not only more effective 145 than simple techniques including multivariate linear regression, but also easier to understand than neural networks (e.g., Huang and Townshend 2003). 146

I47 Cubist constructs an unconventional type of regression tree, in which the terminal 148 nodes or leaves are linear regression models instead of discrete values (Minasny and 149 McBratney 2008). Cubist produces rule-based models containing one or more rules, each of 150 which is a set of conditions associated with a multivariate linear submodel. Cubist is a 151 powerful tool for generating rule-based predictive models. A Cubist model resembles a 152 piecewise linear model, except that the rules can overlap with one another (RuleQuest 2008). 153 Details on regression tree approaches and Cubist were described in Yang et al. (2003), Wylie 154 et al. (2007), and Xiao et al. (2008). In our previous study, we used Cubist to develop a 155 predictive NEE model and upscaled NEE estimates to the continental scale for the 156 conterminous U.S. (Xiao et al. 2008). In this study, we used Cubist to construct a predictive 157 GPP model based on MODIS and AmeriFlux GPP data. Cubist uses three statistical measures 158 to evaluate the quality of the constructed predictive model, including mean absolute error 159 (MAE), relative error (RE), and product-moment correlation coefficient (Yang et al. 2003; 160 Xiao et al. 2008). MAE is calculated as:

161
$$MAE = \frac{1}{N} \sum_{i=1}^{N} |y_i - \hat{y}_i|$$
(1)

where N is the number of samples used to establish the predictive model, and y_i and \hat{y}_i are the actual and predicted values of the response variable, respectively. RE is calculated as:

$$RE = \frac{MAE_T}{MAE_\mu}$$
(2)

165 where MAE_T is the MAE of the constructed model, and MAE_p is the MAE that would result 166 from always predicting the mean value. All three statistical measures were used to evaluate the 167 performance of the constructed model.

168 2.2. Explanatory variables

169 GPP is influenced by a variety of physical, physiological, atmospheric, hydrological, 170 and edaphic variables. At the leaf level, GPP is influenced by several factors, including 171 incoming solar radiation, air temperature, vapor pressure deficit, soil moisture, and nitrogen 172 availability (Ruimy et al. 1995; Clark et al. 1999, 2004). At the canopy or ecosystem level, 173 GPP is also influenced by leaf area index (LAI) (Ruimy et al. 1995) and canopy phenology 174 (Richardson et al. 2008a). At the stand or regional level, GPP is significantly affected by 175 disturbances such as fire and harvest (Law et al. 2004). Many of these factors can be 176 effectively assessed by satellite remote sensing. Surface reflectance depends on vegetation 177 type, biophysical properties (e.g., biomass, leaf area, and stand age), soil background, soil 178 moisture conditions, and sun-object-sensor geometry (Ranson et al. 1985; Penuelas et al. 1993; 179 Schmidt and Skidmore 2003). Vegetation indices including normalized difference vegetating 180 index (NDVI) and enhanced vegetation index (EVI) are closely correlated to the fraction of 181 photosynthetically active radiation (fPAR; Asrar et al. 1984), and are also related to vegetation 182 biomass and fractional vegetation cover (e.g., Tucker et al. 1985; Persson et al. 1993; Myneni

183 et al. 2001; Chen et al. 2004). Compared to NDVI, EVI is more responsive to canopy structural 184 variations, such as LAI, canopy type, plant physiognomy, and canopy architecture (Gao et al. 185 2000). The normalized difference water index (NDWI; Gao 1996) was shown to be strongly 186 correlated with leaf water content (Jackson et al. 2004) and soil moisture (Fensholt & Sandholt 187 2003) over time. LAI and fPAR characterize vegetation canopy functioning and energy 188 absorption capacity (Myneni et al. 2002) and are key parameters in most ecosystem 189 productivity and biogeochemical models (Sellers et al. 1997). We therefore selected surface 190 reflectance, EVI, LST, LAI, fPAR, and NDWI as explanatory variables. All of these variables 191 were derived from MODIS data, which also avoided the complications and difficulties to 192 merge disparate data sources (Xiao et al. 2008).

193 2.3. AmeriFlux data

194 We obtained the following three types of data: GPP from eddy covariance flux towers, 195 explanatory variables derived from MODIS, and a land cover map. The AmeriFlux network 196 coordinates regional analysis of observations from eddy covariance flux towers across North 197 America, Central America, and South America (Law 2006). We obtained the Level 4 data 198 product for 42 AmeriFlux sites over the period 2000-2006 from the AmeriFlux website 199 (http://public.ornl.gov/ameriflux) (Table 1). This product includes NEE data from most of the 200 active flux sites in the network. These sites are distributed across the conterminous U.S. and 201 cover a range of vegetation types: forests, shrublands, savannas, grasslands, and croplands. 202 Moreover, the distribution of these sites in the mean annual climate space indicates that the 203 sites we selected are fairly representative of typical U.S. climate types (Xiao et al. 2008). In 204 addition, some of the forested sites (e.g., Austin Cary, FL; Metolius new young pine, OR; 205 Metolius intermediate aged ponderosa pine, OR; Wisconsin intermediate hardwood, WI) are at different stages since stand replacing disturbance, which are located in disturbance clusters of sites. In addition, some of the sites have received treatment, including the Howland Forest West Tower (ME; nitrogen fertilizer) and the Mead cropland sites (NE; irrigation versus rainfed, continuous maize versus maize/soybean rotation). We therefore believe that these sites are fairly representative of typical U.S. ecosystem and climate types.

211 The Level 4 product consists of two types of GPP data, including standardized 212 (GPP_st) and original (GPP_or) GPP. GPP was calculated from NEE and ecosystem 213 respiration (R_e):

- $GPP_st = R_c NEE_st$ (3)
- 215 and

216

 $GPP_{or} = R_{o} - NEE_{or}$ (4)

217 where NEE_st and NEE_or are standardized and original NEE, respectively. NEE_st was 218 calculated using the storage obtained from the discrete approach (single point on the top of the 219 tower) with the same approach for all the sites, whereas NEE_or was calculated using the 220storage sent by the principal investigators that can be obtained with the discrete approach or 221 using a vertical CO₂ profile system. Both NEE_st and NEE_or were gap-filled using the 222 Marginal Distribution Sampling (MDS) method (Reichstein et al. 2005) and the Artificial 223 Neural Network (ANN) method (Papale and Valentini 2003). The ANN method was generally, 224 if only slightly, superior to the MDS method (Moffat et al. 2007). A number of methods are 225 available for estimating GPP. Although Stoy et al. (2006) showed that the non-rectangular 226 hyperbolic method (Gilmanov et al. 2003) produce estimates more consistent with independent 227data, we chose to use a method that relies on gap-filled nighttime data because it is more 228 frequently used and less computationally demanding. We used GPP calculated from NEE data that was gap-filled using the ANN method. For each site, if the percentage of the remaining missing values for GPP_st was lower than that for GPP_or, we selected GPP_or; otherwise, we used GPP_st. GPP_st was the first choice so that the processing procedure for GPP was the same for as many sites as possible. We used 8-day average GPP data ($g C m^{-2} day^{-1}$) to match the compositing intervals of MODIS data.

234 2.4. MODIS data

235 We used the following four MODIS data products (Collection 4), including surface 236 reflectance (MOD09A1; Vermote and Vermeulen 1999), daytime and nighttime LST 237 (MOD11A2; Wan et al. 2002), EVI (MOD13A1; Huete et al. 2002), and LAI/fPAR 238 (MOD15A2; Myneni et al., 2002). Surface reflectance and EVI are at a spatial resolution of 239 500m, while LST, LAI, and fPAR are at spatial resolution of 1 km. Surface reflectance, LST, 240 LAI, and fPAR are at a temporal resolution of 8 days, while EVI is at a temporal resolution of 241 16 days. Sims et al. (2005) showed that the midday values of gross CO_2 exchange during 242 satellite overpasses can be used to estimate 8-day mean gross CO₂ exchange, bridging the 243 connection between continuous measurements of flux tower data and 8-day MODIS data. We 244 used the 16-day EVI product instead of EVI calculated from 8-day surface reflectance despite 245 the lower temporal resolution of the 16-day EVI product. Each 16-day EVI composite was 246 composited from 16 daily observations (Huete et al. 2002). The VI algorithm applies a filter to 247 the data based on quality, cloud, and viewing geometry, and only the higher quality, cloud-248 free, filtered data are retained for compositing; the maximum value composite (MVC) method 249 employed selects the observation with the highest VI value to represent the composting period 250 (16 days) (Huete et al. 2002). MVC minimizes the contamination of clouds and aerosols and 251 the effects of sensor view angles on VI (Hoblen 1986). For the 8-day surface reflectance product, each pixel contains the best possible daily observation during an 8-day period as selected on the basis of high observation coverage, low view angle, the absence of clouds or cloud show, and aerosol loading (Vermote and Kotchenova, 2008), and the EVI calculated from the 8-day surface reflectance is less representative of the composting period than the MODIS EVI product.

257 For each AmeriFlux site, we obtained MODIS ASCII (American Standard Code for 258 Information Interchange) subsets (Collection 4) consisting of $7 \text{ km} \times 7 \text{ km}$ regions centered on 259 the flux tower, including surface reflectance, daytime and nighttime LST, EVI, LAI, and fPAR 260over the period 2000-2006 from the Oak Ridge National Laboratory's Distributed Active 261 Archive Center (ORNL DAAC 2006). We extracted average values for the central $3 \text{ km} \times 3$ 262 km area within the 7 km \times 7 km cutouts to better represent the flux tower footprint (Schmid 263 2002; Rahman et al. 2005; Xiao et al. 2008). For each variable, we determined the quality of 264 the value of each pixel within the area using the quality assurance (QA) flags included in the 265 product. At each time step, we averaged the values of each variable using the pixels with good 266 quality within the area to represent the values at the flux site. If none of the values within the 3 267 \times 3 km area were of good quality, we treated the period as missing. Each 16-day EVI value 268 was used for the two 8-day intervals corresponding with the compositing interval of other 269 MODIS data products. NDWI was calculated from band 2 and band 6 of the surface 270reflectance product.

To estimate GPP at the continental scale, we obtained wall-to-wall MODIS data including surface reflectance, daytime and nighttime LST, LAI, and EVI over the period from February 2000 to December 2006 from the Earth Observing System (EOS) Data Gateway. For each variable, we determined the quality of the value of each pixel using the QA flags and

replaced the bad-quality value using a linear interpolation approach (Zhao et al. 2005). The NDWI was calculated from band 2 (near-infrared, 841-876 nm) and band 6 (shortwave infrared, 1628-1652 nm) of the surface reflectance product (MOD09A1). Similarly, each 16day EVI composite was used for two 8-day intervals corresponding to the compositing interval of other MODIS products. NDWI was calculated from band 2 and band 6 of the surface reflectance product for each 8-day interval.

281 We also used the MODIS 8-day GPP product (MOD17A2; Running et al. 2004) to 282 evaluate the performance of our model at both the site level. The MODIS GPP product is at a 283 spatial resolution of 1 km, and a temporal resolution of 8 days. We obtained MODIS ASCI 284 subsets (Collection 4) for the MODIS 8-day GPP product over the period 2005-2006 from the 285 Oak Ridge National Laboratory's Distributed Active Archive Center (ORNL DAAC 2006). 286 We also obtained the MODIS annual GPP product (MOD17A3; Running et al. 2004) for 2005 287 from the Numerical Terradynamic Simulation Group, University of Montana 288 (http://www.ntsg.umt.edu) to evaluate the performance of our model at the continental scale.

289 2.5. Land cover

290 To construct a predictive GPP model, we obtained the land cover type for each 291 AmeriFlux site based on the site descriptions (Table 1) and categorized each site into a class of 292 the UMD (University of Maryland) land-cover classification system. Although the 42 293 AmeriFlux sites used in this study cover a variety of vegetation classes of this classification 294 system, some classes (e.g., deciduous needleleaf forests, open shrublands) were not covered by 295 any site. We therefore reclassified all vegetation classes of the UMD classification system to 296 seven broader classes, following Xiao et al. (2008). Specifically, evergreen needleleaf forests 297 and every green broadleaf forests were merged to every even forests, deciduous needleleaf forests

and deciduous broadleaf forests to deciduous forests, closed shrublands and open shrublands toshrublands, and woody savannas and savannas to savannas.

300 To estimate GPP for each $1 \text{ km} \times 1 \text{ km}$ pixel across the conterminous U.S., we 301 obtained the land-cover type for each pixel from the MODIS land-over map with the UMD 302 classification system (Friedl et al. 2002). Similarly, we reclassified the vegetation classes of the 303 MODIS land-cover map to the seven broader classes. We then used the reclassified land cover 304 map to specify the land cover of each 1 km \times 1 km cell across the conterminous U.S.

305 2.6. Model development

306 We developed a predictive GPP model using Cubist based on the site-specific MODIS 307 and AmeriFlux GPP data. Our explanatory variables included land cover, surface reflectance 308 (bands 1-7), daytime and nighttime LST, EVI, NDWI, fPAR, and LAI, and our response variable was GPP (g C m⁻² day⁻¹). Land cover was included as a categorical variable in the 309 310 model. We split the site-level data set of AmeriFlux and MODIS data into a training set (2000-31 F 2004) and a test set (2005-2006). If a site only had GPP observations for the period 2000-2004, 312 the site was only included in the training set; if a site only had GPP observations for the period 313 2005-2006, the site was only included in the test set; otherwise, the site was included in both 314 training and test sets. The training and test sets included 40 and 34 AmeriFlux sites, 315 respectively. We had a total of 4529 and 2240 data samples for the training and test sets, 316 respectively. In addition to the full model that includes all of the 14 independent variables, we 317 also developed a series of models by dropping one or more variables at a time using Cubist. To 318 select the best model, we evaluated the performance of each model based on MAE, RE, and 319 correlation coefficient. We chose the model with the minimal MAE and RE and maximum 320 correlation coefficient as the best model. We also evaluated the model performance using the Root Mean Squared Error (RMSE), scatterplots of predicted GPP versus observed GPP, and
 seasonal variations between the predicted and observed GPP.

We also evaluated the performance of our model in the spatial domain using leave-oneout cross-validation. In this approach, the data from a single site was used for validation, and the data from the remaining sites were used for training. The training and validation data were from different sites and were therefore independent from each other as these sites are generally hundreds of kilometers away from each other and the spatial autocorrelation between these sites was negligible. The leave-one-out cross-validation was conducted for each site, separately.

330 2.7. Continental-scale estimation of GPP

331 As mentioned earlier, the AmeriFlux sites used in this study are fairly representative of 332 typical U.S. ecosystem and climate types. We believe that the predictive GPP model 333 constructed from the 42 sites can be extrapolated to the conterminous U.S. We used the model 334 to estimate GPP for each 1 km \times 1 km cell across the conterminous U.S. for each 8-day 335 interval over the period 2000-2006 using wall-to-wall MODIS data. GPP was not estimated for 336 non-vegetated cells (e.g., urban, barren), and water bodies. We compared our estimate with the 337 MODIS GPP product (MOD17A3; Running et al. 2004). With our 8-day GPP estimates, we 338 examined the patterns, magnitude, and interannual variability of GPP.

- 339 3. Results and Discussion
- 340 3.1. Model development
- 341 3.1.1 Predictive GPP model

We chose the model containing five explanatory variables - land cover, EVI, daytime LST, LAI, and NDWI as the best model to predict GPP at the continental scale (RE = 0.38,

344	MAE = 1.22 g C m ⁻² day ⁻¹ , R ² = 0.74). The performance of the model was comparable to that
345	of the full model (RE = 0.37, MAE = 1.19 g C m ⁻² day ⁻¹ , $R^2 = 0.74$). Having only five
346	explanatory variables could substantially reduce the computational complexity for continental-
347	scale predictions compared to the full model. The model consisted of five committee models,
348	each of which was made of a number of rule-based submodels. For instance, the first
349	committee model was made of the following 23 rule-based submodels:
350	Rule 1: if land cover in {Deciduous forests, Mixed forests}, EVI <= 0.37, and LAI <=
351	2.64, then
352	$GPP = 0.28 + 0.012 LST_{day}$
353	Rule 2: if $LST_{day} \le 3.27$, $EVI > 0.22$, then
354	GPP = -0.64 + 4.8EVI + 0.6NDWI + 0.06LAJ
355	Rule 3: if land cover in {Evergreen forests, Shrublands, Savannas, Grasslands,
356	Croplands}, EVI <= 0.22, LAI <= 2.64, then
357	$GPP = 0.15 + 1.4EVJ + 0.01 LST_{day}$
358	Rule 4: if land cover in {Deciduous forests, Mixed forests, Savannas, Croplands},
359	$LST_{day} > 3.26$, EVI <= 0.37, LAI <= 1.78, then
360	GPP = -1.10 + 0.75LAI + 4.6EVI
361	
362	Rule 22: if NDWI <= -0.28 , LST _{day} > 17.65, EVI > 0.58, then
363	$GPP = 10.49 + 1.9EVI + 0.02LST_{day} - 0.7NDWI$
364	Rule 23: if land cover = Croplands, $LST_{day} > 29.92$, EVI > 0.37, then
365	$GPP = -24.85 + 5.47LAI + 58.9EVI + 0.124LST_{day} - 0.5NDWI$

where LST_{day} is the daytime LST. As mentioned earlier, the rules of the model could overlap with one another. For instance, rule 1 overlapped with rule 4 as land cover could be deciduous forests in both cases; rules 22 and 23 overlapped with each other as EVI could be greater than 0.58 in both cases.

370 3.1.2. Model evaluation

371 The analysis of model residuals indicated that the residuals were not randomly 372 distributed. Low GPP values were generally associated with low prediction errors, whereas 373 high GPP values were associated with high prediction errors. The uncertainties of carbon flux 374 measurements are directly proportional to the magnitudes of the fluxes (Richardson et al. 375 2008b). The residuals also exhibited a systematic component. For example, large GPP tended 376 to have consistently negative residuals. The residuals also had a random component that arose 377 partially from errors/uncertainties in the measured fluxes as well as MODIS data. Random 378 errors in AmeriFlux GPP data are significant (Hagen et al. 2006) and these errors may 379 ultimately limit the agreement between observed and predicted GPP values. In addition, the 380 explanatory variables included in the model could not completely explain the variance of GPP. 381 For example, the independent variables used in the model could not account for nitrogen 382 availability, and may affect the accuracy of the model.

We compared our GPP estimates with observed GPP for each AmeriFlux site over the period 2005-2006 (Fig. 2). Our estimates captured most features of observed GPP including seasonality and year-to-year variations over the period 2005-2006. GPP was under- and overpredicted for some sites. The model could not capture exceptionally high GPP values for some sites, such as Audubon Research Ranch (AZ), Santa Rita Mesquite (AZ), and Fort Peck (MT). We averaged observed and predicted 8-day GPP for each site, and plotted mean predicted GPP

389 against observed GPP (Fig. 3). The model estimated GPP fairly well at the site level (y = 0.95x) + 0.21, $R^2 = 0.84$, p < 0.0001; RMSE = 0.77 g C m⁻² day⁻¹). Overall, the model slightly under-390 and overestimated GPP for values greater or lower than $\sim 4 \text{ g C m}^2$ day¹, respectively. The **39**1 392 model performance also varied with site. Large underestimation occurred at some sites such as 393 the North Carolina Loblolly Pine (NCP, NC), Freeman Ranch Mesquite Juniper (FRM, TX), 394 and Walnut Gulch Kendall Grasslands (WGK, AZ), whereas large overestimation occurred at some sites such as Lost Creek (LC, WI) and Mead Rainfed (MR, NE). The model oredicted 395 GPP remarkably well at the biome level (y = 0.99x - 0.13, $R^2 = 0.91$, p < 0.00001; RMSE = 396 0.42 g C m^{-2} day⁻¹). The model slightly overestimated GPP for deciduous forests and croplands 397 398 and slightly underestimated GPP for all other biomes.

399 The disagreement between predicted and observed GPP values is likely due to the 400 following reasons. First, the MODIS and tower footprints do not match with each other and the 401 vegetation structure at the flux tower could be significantly different from that within the 402 MODIS footprint (Xiao et al. 2008). For example, the Tonzi Ranch site (CA) is dominated by 403 deciduous blue oaks (Ouercus douglasii) and the understory and open grassland are mainly 404 cool-season C_3 annual species (Ma et al. 2007). The MODIS footprint consists of a larger 405 fraction of grassland than the tower footprint. Blue oaks and grasses have distinct phenologies 406 (Ma et al. 2007) and therefore had differential contributions to the carbon fluxes integrated 407 over the MODIS footprint over time (Xiao et al. 2008). Second, MODIS data is less sensitive 408 to changes in understory vegetation and damage to canopies that do not increase canopy gaps, 409 leading to overestimation of carbon assimilation rates. Third, the independent variables 410 included in the model could not account for other factors such as nitrogen availability (Clark et 411 al. 1999, 2004) and stand age (Ryan et al. 2004), all of which may influence GPP. Finally, we

412 estimated GPP for each 8-day interval, and therefore our estimates may not capture the 413 variability of GPP within that period. The 8- or 16-day LST and EVI values do not always 414 represent average environmental conditions and average fluxes over the 8- or 16-day period 415 (Xiao et al. 2008), and the exclusion of days with high and low values could lead to 416 underestimation and overestimation of GPP values, respectively. In addition, during drought or 417 days with high vapor pressure deficits, the midday GPP may not be representative because of 418 the skewed diurnal variation in GPP (Anthoni et al. 1999).

419 3.1.3. Model validation

420 We validated our model in both temporal and spatial domains. We first validated the 421 model in the temporal domain using the test set over the period 2005-2006 (Fig. 1a). Our model estimated GPP fairly well ($R^2 = 0.74$, p < 0.0001; RMSE = 1.99 g C m⁻² day⁻¹), although 422 it slightly under- and overestimated GPP values greater and less than 3 g C m⁻² day⁻¹, 423 424 respectively. By contrast, the MODIS GPP product estimated GPP for the AmeriFlux sites with a RMSE of 2.43 g C m⁻² day⁻¹ (y = 0.50x + 1.0], $R^2 = 0.58$, p < 0.0001; Fig. 1b). The R^2 425 426 and RMSE of our model were 28% higher and 18% lower than those of the MODIS GPP 427 product, respectively,

We then validated the model in the spatial domain using leave-one-out cross validation. The cross-validation also showed that our model estimated GPP fairly well (Fig. 4). The model performance varied with site and biome type. Our model had a higher performance for forest ecosystems and croplands than shrublands, savannas, and grasslands. We averaged observed and predicted 8-day GPP for each site, and then plotted mean predicted GPP against observed GPP (Fig. 5). The cross-validation showed that our model estimated GPP fairly well at the site level (y = 0.73x + 1.07, R² = 0.69, p < 0.0001; RMSE = 1.19 g C m⁻² day⁻¹) although the slope and R^2 were lower and the RMSE was higher than those of the validation in the temporal domain, respectively. The cross-validation also showed that our model estimated GPP remarkably well at the biome level (y = 0.86x + 0.48, R^2 = 0.91, p < 0.0001; RMSE = 0.33 g C m⁻² day⁻¹).

The validation of the model in both temporal and spatial domains showed that the performance of our model is encouraging, given the diversity in ecosystem types, age structures, fire and insect disturbances, and management practices. Our approach extensively used eddy covariance flux tower data involving typical U.S. ecosystem and climate types. Our study demonstrated that our empirical approach has great potential for upscaling flux tower GPP data to continental scales across a variety of biomes.

- 445 3.2. Gross Primary Production
- 446 3.2.1. Seasonal patterns

447 Our 8-day GPP estimates were highly constrained by eddy flux data, and provided a 448 spatially and temporally continuous measure of GPP with high spatial and temporal resolution 449 for the conterminous U.S., which made it possible to examine the patterns, magnitude, and 450 interannual variability of GPP across the U.S. Our estimates showed that GPP exhibited large 451 spatial variability and strong seasonal fluctuations (Fig. 6). The seasonal patterns of GPP and 452 its spatial variability reflected the controlling effects of climate conditions. In the spring 453 months, the Southeast and the Gulf Coast significantly assimilated carbon with GPP values reaching $-100-250 \text{ g C m}^{-2} \text{ mo}^{-1}$ as the growing season started in early to mid-spring in these 454 455 regions. The Pacific Coast is dominated by evergreen forests, and these ecosystems also 456 assimilated carbon due to mild temperatures and moist conditions during the spring (Anthoni et 457 al. 2002). The Mediterranean regions in California also assimilated carbon in the spring

458 because of a surplus of precipitation and relatively warm temperatures (Xu & Baldocchi 2004; 459 Ma et al. 2007). By contrast, the Upper Great Lakes region and the northern Great Plains are 460 dominated by croplands with most crops planted between April and June (Shroyer et al. 1996), 461 whereas the New England region and the northern portion of the Upper Great Lakes region are 462 dominated by temperate-boreal transitional forests. The relatively late greenup in these regions 463 led to low GPP.

In the summer months, the eastern U.S., the Coastal Pacific Northwest, and some regions in California exhibited high GPP values (~250-450 g C m⁻² mo⁻¹) owing to favorable temperature and soil moisture conditions, while the vast majority of western landscapes, including the Great Basin, the Colorado Plateau, and the western Great Plains exhibited much lower GPP values due to spare vegetation and precipitation deficits.

In the fall months (September-November), the GPP values of the Southeast and the Gulf Coast substantially decreased relative to those in the summer because vegetation began to senesce and days became shorter in these regions. The spatial patterns and magnitude of GPP were similar to those of the spring. The Upper Great Lakes region and the Great Plains had very low GPP values due to the harvesting of crops.

In the winter months (December-February), the majority of the U.S. had little or no photosynthesis as the canopies of most ecosystems were dormant. Some regions in the Pacific Coast, California, the Gulf Coast, and the Southeast slightly assimilated carbon, because of the dominance of evergreen forests and mild temperatures (Waring & Franklin 1979; Clark et al. 1999; Anthoni et al. 2002; McGarvey et al. 2004). For example, Douglas-fir, a major species in the Pacific Northwest and California, is known to be highly plastic and able to photosynthesize in winter when temperatures are above the freezing point (Xiao et al. 2008).

481 Fig. 7 showed the trajectories of the spatially averaged and integrated 8-day GPP for 482 each biome from February 2000 to December 2006. Deciduous forests and croplands had the 483 largest intra-annual variability in spatial averaged GPP, followed by mixed forests; evergreen 484 forests and savannas had intermediate intra-annual variability; grasslands and shrublands had 485 the least variability (Fig. 7a). The temporal variability of spatially integrated GPP (or spatial 486 total) also showed clear dependence on biome (Fig. 7). Collectively, the terrestrial ecosystems substantially assimilated carbon and had a peak spatial total of 50-55 Tg C day⁻¹. Taken 487 488 separately, croptands had the highest intra-annual variability in spatially integrated GPP, with a peak spatial total of 25-30 Tg C day⁻¹, followed by deciduous forests; evergreen forests, mixed 489 490 forests, savannas, and grasslands had intermediate intra-annual variability in spatially 491 integrated GPP; shrublands had the least variability. Both spatially averaged and integrated 492 GPP showed interannual variability for each biome.

493 3.2.2. Annual fluxes

494 We calculated annual GPP for each year over the period 2001-2006 from our 8-day 495 GPP estimates, and then calculated the average annual GPP over the 6-year period (Fig. 8). 496 Annual GPP varied considerably over space, and exhibited a large spatial gradient from the 497 east to the west. The Gulf Coast, the Southeast, the coastal Pacific Northwest, and a part of the Pacific Southwest had high annual GPP (~1500-2000 g C m² yr⁻¹); the Midwest and the 498 Northeast had intermediate values (-1200-1500 g C m⁻² yr⁻¹), and the majority of the western 499 500 half of the country including the Southwest, the western Great Plains, and the Rocky Mountain region had GPP values generally lower than 500 g C m⁻² yr⁻¹. 501

502 We estimated a total gross carbon uptake of 7.06 Pg C yr⁻¹ for the conterminous U.S. 503 over the period 2001-2006. A quantitative breakdown of the 6-yr average annual GPP map by regions (Table 2) showed that the North Central and South Central regions had the highest GPP, followed by the Southeast, the Rocky Mountain region, and the Northeast; the Pacific Northwest and the Pacific Southwest had the lowest GPP. The spatially averaged annual GPP of the U.S. was \sim 1100 g C m⁻² yr⁻¹ (Table 2). Regionally, the Southeast had the highest spatially averaged annual GPP, followed by the Northeast and the South Central regions; the North Central region, the Pacific Northwest, and the Pacific Southwest had intermediate values; the Rocky Mountain region had the lowest spatially averaged annual GPP.

511 A quantitative breakdown of the 6-yr average annual GPP by biomes (Table 3) showed 512 spatially integrated annual GPP varied with biome. Croplands had the highest spatially 513 integrated annual GPP; shrublands had lowest annual GPP; other biomes including evergreen 514 forests, deciduous forests, mixed forests, savannas, and grasslands had intermediate annual 515 GPP. Spatially averaged annual GPP also varied with biome (Table 3). Deciduous forests had 516 the highest spatially averaged annual GPP, followed by evergreen forests, mixed forests, 517 croplands, and savannas; grasslands had intermediate values; shrublands had the lowest values. 518 We compared our annual GPP estimate with annual GPP estimate from the MODIS 519 GPP product (MOD17A3; Running et al. 2004) for 2005 (Fig. 9). Both estimates showed a 520 large spatial gradient from the east to the west: the Southeast had the highest annual GPP; the 521 Midwest had intermediate annual GPP, while the Rocky Mountain region had the lowest GPP. Both estimates showed that annual GPP is $-300 \text{ g C m}^{-2} \text{ yr}^{-1}$ in the Rocky Mountain region. 522 523 Annual GPP is also very similar in the New England region and the Upper Peninsula of 524 Michigan. However, large discrepancies were observed between our estimate and the MODIS 525 GPP product. Our estimate exhibited larger spatial variability than the estimate from the 526 MODIS GPP product. Moreover, compared to our estimate, the MODIS GPP product substantially underestimated GPP in some regions, particularly the Midwest and the coastal Pacific Northwest where ecosystems are highly productive. For example, our annual GPP for the croplands in the Midwest is ~1200-1500 g C m⁻² yr⁻¹, while the MODIS annual GPP is only ~700 g C m⁻² yr⁻¹. In many areas in the Southeast, our annual GPP estimate is ~1500-2000 g C m⁻² yr⁻¹, while the MODIS annual GPP is only ~1000-1500 g C m⁻² yr⁻¹.

532 The large discrepancies in annual GPP between our estimate and the MODIS GPP 533 product can be attributed to the following reasons. First, the MODIS GPP product was 534 developed using an algorithm optimized for global applications and meteorological fields with 535 coarse resolution (1 by 1.25 degree) and large uncertainties (Zhao et al. 2006), and thereby 536 likely contributed to the smaller spatial variability of MODIS GPP and affected the accuracy of 537 the MODIS GPP estimates. Second, the maximum light use efficiency (ϵ_{max}) is an essential 538 parameter of the LUE model used to develop the MODIS GPP product. The smax is only 0.68 g C MJ⁻¹ for croplands in the MOD17 algorithm (Heinsch et al. 2003), which is likely too low 539 540 for croplands and results in substantial underestimation of cropland GPP. Zhang et al. (2008) 541 showed that MODIS annual GPP for an irrigated cropland in China was only about 20-30% of 542 annual GPP derived from eddy covariance flux measurements, and attributed the substantial underestimation of cropland GPP to the underestimation of ε_{max} in the MOD17 algorithm. 543 544 Third, our estimate was highly constrained eddy flux data, while some geographical regions 545 and biomes are underrepresented by the AmeriFlux network, which could affect the accuracy 546 of our estimates. For example, we merged savannas (tree cover 10-30%) and woody savannas 547 (tree cover 30-60%) together in the development of the model because no sites representing 548 typical savannas with tree cover below 30% were available in the model, which could lead to 549 overestimation of GPP for areas that were classified as savannas (tree cover 10-30%).

Global annual GPP has been estimated to be 120 Pg C yr⁻¹ using ¹⁸O measurements of 550 atmospheric CO₂ (Ciais et al. 1997) and 110 Pg C yr⁻¹ from 2001 to 2003 using the MODIS 551 552 GPP product (Zhao et al. 2005). Our estimates suggested that the terrestrial ecosystems in the 553 conterminous U.S. accounted for 5.9-6.5% of the global annual GPP, while its land area 554 accounts for ~5.4% of the global land area. Our estimate of U.S. annual GPP was higher than 555 other estimates. For example, the average annual GPP over the period 2001-2006 derived from MODIS GPP product (Running et al. 2004) was ~6.2 Pg C yr⁻¹. Our estimate was about 14% 556 557 higher than the MODIS GPP estimate. Potter et al. (2007) estimated annual NPP between 2.67 and 2.79 Pg C yr⁻¹ over the period 2001-2004 using MODIS data and the NASA-CASA model. 558 559 Our estimate was also higher than that estimated by Potter et al. (2007) assuming that NPP is 560 about half of GPP (Lloyd & Farguhar 1996; Waring et al. 1998).

561 Our predictive model has advantages over empirical or process-based ecosystem 562 models. Most ecosystem models are dependent on site-level parameterizations that are used as 563 default parameters for a much broader spectrum of vegetation types and conditions, which may 564 limit the accuracy of model simulations over large areas (e.g., Prince and Goward 1995; 565 Running et al. 2004; Xiao et al. 2009). By contrast, our model was highly constrained by eddy 566 flux data from a number of towers encompassing a range of ecosystem and climate types, and 567 may lead to model parameters that are more representative of the full spectrum of vegetation 568 and climate types and thereby more accurate estimates of carbon fluxes at regional scales. 569 Moreover, our model consisted of rule-based, multivariate linear regression models, and is 570 easier to understand and implement. Our model could substantially reduce computational 571 complexity compared to many ecosystem models. On the other hand, however, our model also 572 has disadvantages over empirical or process-based ecosystem models. Our model is an

573 empirical approach, and does not involve ecosystem processes such as photosynthesis and 574 nitrogen cycling. Moreover, our model did not explicitly consider some factors influencing 575 GPP such as nitrogen availability, stand age, and disturbance history that may be explicitly 576 simulated in process-based ecosystem models (e.g., Aber et al. 1997). All these model 577 differences contributed to the discrepancies in annual GPP estimates between our empirical 578 approach and ecosystem models.

579 3.2.3. Interannual variability

The U.S. annual GPP varied between 6.91 and 7.33 Pg C yr⁻¹ over the period 2001-580 581 2006. The years 2002, 2004, and 2006 had lower GPP values relative to 2001, 2003, and 2005. 582 Annual GPP exhibited positive and negative anomalies for each year, and the spatial patterns 583 of these anomalies also varied from year to year (Fig. 10). The annual GPP anomalies were 584 likely due to climate variability, disturbances, and management practices. Most notably, large 585 negative GPP anomalies occurred in the Great Plains in 2002 and 2006. Moderate to severe 586 drought affected over 50% of the country in both years, including the Southwest, the Great 587 Plains, the Gulf Coast, and the coastal Southeast, particularly Texas and Oklahoma (U.S. 588 Drought Monitor, http://www.drought.unl.edu). The annual precipitation of these two years 589 was 467 and 458 mm for the U.S., respectively - lower than the 30-year mean annual 590 precipitation (480 mm) taken over the PRISM climate database. Our GPP anomaly maps 591 showed large negative GPP anomalies in many of the drought-affected regions. Notably, large 592 negative GPP anomalies occurred in the Great Plains in 2002 and 2006. Our results further 593 demonstrated severe drought could substantially affect ecosystem carbon fluxes (Xiao et al. 594 2009).

595 At landscape to regional scales, annual GPP also exhibited large anomalies. For 596 example, our results showed large negative GPP anomalies in the region experienced the 597 Biscuit fire in Oregon (Fig. 11). The 2002 Biscuit Fire was among the largest forest fires in 598 modern U.S. history, encompassing > 2000 km^2 primarily within the Rogue-Siskiyou National 599 Forest (RSNF) in southwest Oregon (Campbell et al., 2007; Thompson et al. 2007). Large fires 600 such as the Biscuit Fire damaged both overstory and/or understory vegetation, leading to a 601 reduction in GPP and large negative GPP anomalies in the region in 2003, which may result in 602 net carbon release into the atmosphere. Numerous wildfires occurred over the western half of the country due to dry weather and high winds, burning $1.5-4.0 \times 10^4$ km² of forests from 2000 603 604 to 2006 (U.S. Fire Administration, http://www.usfa.dhs.gov). The drought along with wildfires 605 likely led to the negative GPP anomalies in the western half of the U.S.

606 Our results also showed hurricanes could reduce GPP and lead to large negative GPP anomalies (Fig. 12). For example, Hurricane Katrina occurred in late August 2005 affected 607 over 2×10^4 km² of forest across Mississippi, Louisiana and Alabama, with damage ranging 608 609 from downed trees, snapped trunks and broken limbs to stripped leaves (USDA Forest Service, 610 http://www.srs.fs.usda.gov). Forest inventories indicated that the potential timber losses from 611 Hurricane Katrina amounted to roughly 1.2×10^8 m³ (USDA Forest Service, 612 http://www.srs.fs.usda.gov). Our results showed large negative GPP anomalies in 2006 in the 613 Gulf Coast region severely affected by Hurricane Katrina (Fig. 12). The reduction in GPP and 614 increased Re resulting from increased litter could lead to substantial carbon release into the 615 atmosphere (Chambers et al. 2007).

616 4. Conclusions

617 We used a regression tree approach and remotely-sensed data from MODIS to upscale 618 AmeriFlux GPP data to the continental scale and to produce a GPP dataset with 8-day temporal 619 resolution and 1 km spatial resolution for the conterminous U.S. over the period 2000-2006. 620 Our results demonstrated that our empirical approach has great potential for upscale eddy flux. 621 GPP data to large areas across multiple biomes. Our GPP estimates provided a spatially and 622 temporally continuous measure of gross primary production for the conterminous U.S. Our 623 estimates also provided an alternative, independent dataset from the MODIS GPP product and 624 simulations with biogeochemical models. Our GPP estimates were highly constrained by flux 625 tower data from towers encompassing a large range of ecosystem and climate types as well as 626 disturbance history. Our approach can be applied to the entire North America, other geographic 627 regions including Europe, Southeast Asia, and South America, or to the global scale, and to 628 produce continuous GPP estimates over continents or the globe. This approach can also be 629 used to upscale other fluxes including evapotranspiration to large areas.

630 Our GPP estimates exhibited large spatial variability and strong seasonal variations, 631 which reflected the controlling effects of climate conditions and vegetation distributions. We estimated a total gross carbon uptake of 7.06 Pg C yr⁴ for the conterminous U.S. over the 632 633 period 2001-2006. Annual GPP varied substantially with geographical region and biome type. 634 Our annual GPP estimate exhibited large spatial variability than the MODIS GPP product 635 (MOD17A3; Running et al. 2004). Our results also showed that the U.S. annual GPP varied between 6.91 and 7.33 Pg C yr⁻¹ over the 6-year period. Extreme climate events (e.g., drought) 636 637 and disturbances (e.g., fires and hurricanes) reduced annual GPP at regional scales and could 638 have a significant impact on the U.S. net ecosystem carbon exchange. The interannual

639 variability of GPP was mainly caused by these extreme climate events (e.g., drought) and640 disturbances (e.g., fire, hurricane).

641 Acknowledgements:

642 This research was partly supported by the National Science Foundation (NSF) Carbon 643 and Water Program (EAR-0630319). We thank the principal investigators of the MODIS data 644 products including Drs. A.R. Huete, R.B. Myneni, S.W. Running, E.F. Vermote, and Z. Wan. 645 We also thank D.A. Sims, T.A. Boden, S.K.S. Vannan, M. Zhao, Z. Wan, and L. Delp for 646 helpful discussion. The MODIS data products were obtained from the Oak Ridge National 647 Laboratory (ORNL) Distributed Active Archive Center (DAAC) and the Earth Observing 648 System (EOS) Data Gateway for making these products available. The PRISM climate 649 database University provided PRISM Group. Oregon State was the bγ 650 (http://www.prismclimate.org).

651

652 References:

Aber, J.D., Ollinger, S.V., Driscoll, C.T. (1997). Modeling nitrogen staturation in forest
 ecosystems in response to land use and atmospheric deposition. *Ecological Modelling*, 101,
 61-78.

656 Anthoni, P.M., Unsworth, M.H., Law, B.E., Irvine, J., Baldocchi, D.D., Tuyl, S.V., Moore, D.

657 (2002). Seasonal differences in carbon and water vapor exchange in young and old-growth
 658 ponderosa pine ecosystems. *Agricultural Forest Meteorology*, 111, 203-222.

Asrar, G., Fuchs, M., Kanemasu, E.T., Hatfield, J.L. (1984). Estimating of absorbed
 photosynthesis radiation and leaf area index from spectral reflectance in wheat. Agron. J.,
 76, 300-306.

- Baldocchi, D., Falge, E., Gu, L. et al. (2001). FLUXNET: A new tool to study the temporal
 and spatial variability of ecosystem-scale carbon dioxide, water vapor, and energy flux
 densities. Bulletin of the American Meteorological Society, 82, 2415-2434.
- 665 Campbell, J., Donato, D., Azuma, D., & Law, B. (2007). Pyrogenic carbon emission from a
 666 large wildfire in Oregon, United States. J. Geophys. Res., 112, G04014.
- 667 doi:10.1029/2007JG000451.
- Chambers, J.Q., Fisher, J.I., Zeng, H., Chapman, E.L., Baker, D.B., Hurtt, G.C. (2007).
 Hurricane Katrina's carbon footprint on U.S. Gulf Coast forests. Science, 318, 1107.
- 670 Chen, X., Vierling, L., Powell, E., DeFelice T. (2004). Using lidar and effective LAI data to
- 671 evaluate IKONOS and Landsat 7 ETM+ vegetation cover estimates in ponderosa pine
 672 forest. Remote Sensing of Environment, 91, 14-26.
- 673 Ciais, P. et al. (1997). A thee-dimensional synthesis study of y180 in atmospheric CO2. 1.
 674 surface Fluxes. Journal of Geophysical Research, D102, 5857-5872.
- 675 Clark, K.L., Ghoiz, H.L., Moncrieff, J.B., Cropley, F., Loescher, H.W. (1999). Environmental
- 676 controls over net carbon dioxide from contrasting Florida ecosystems. Ecological
 677 Applications, 9, 936-948.
- 678 Clark, K.L., Gholz, H.L., Castro, M.S. (2004). Carbon dynamics along a chronosequence of
 679 slash pine plantations in N. Florida. *Ecological Applications*, 4, 1154-1171.
- 680 Cook, B.D., Davis, K.J., Wang, W. et al. (2004). Carbon exchange and venting anomalies in an
- upland deciduous forest in northern Wisconsin, USA. Agricultural Forest Meteorology,
 126, 271-295.
- 683 DeFries, R.S., Houghton, R.A., Hansen, M.C., Field, C.B., Skole, D., & Townshend, J. (2002).
- 684 Carbon emissions from tropical deforestation and regrowth based on satellite observations

- 685 for the 1980s and 1990s, Proceedings of the National Academy of Sciences of the United 686 States of America, 99, 14256-14261.
- 687 Desai, A.R., Bolstad, P.V., Cook, B.D., Davis, J., & Carey, E.V. (2005), Comparing net 688 ecosystem exchange of carbon dioxide between an old-growth and mature forest in the
- 689 upper Midwest, USA. Agricultural Forest Meteorology, 128, 33-55.
- 690 Desai, A.R., Richardson, A.D., Moffat, A.M. et al. (2008). Cross site evaluation of eddy **69**1 covariance GPP and RE decomposition techniques. Agricultural and Forest Meteorology, 692 in press.
- 693 Dore, S., Hymus, G.J., Johnson, D.P., Hinkle, C.R., Valentini, R., & Drake, B.G. (2003). Cross
- 694 validation of open-top chamber and eddy covariance measurements of ecosystem CO2 695 exchange in a Florida scrub-oak ecosystem. Global Change Biology, 9, 84-95.
- 696 Falk, M., Wharton, S., Schroeder, M., Ustin, S., Paw U, K.T. (2008). Flux partitioning in an 697 old-growth forest: seasonal and interannual dynamics. Tree Physiology, 28, 509-520.
- 698 Fensholt, R., & Sandholt, I. (2003). Derivation of a shortwave infrared water stress index from
- 699 MODIS near- and shortwave infrared data in a semiarid environment. Remote Sensing
- 700 Environment, 87, 111-121.
- 701 Friedl, M.A., McIver, D.K., Hodges, J.C.F. et al. (2002). Global land cover mapping from
- 702 MODIS: algorithms and early results. Remote Sensing Environment, 83, 287-302.
- 703 Gao, B.C. (1996). NDWI - A normalized difference water index for remote sensing of 704 vegetation liquid water from space. Remote Sensing Environment, 58, 257-266.
- 705
- 706 vegetation spectra without background contamination. Remote Sensing of Environment, 74,

Gao, X., Huete, A. R., Ni, W., Miura, T. (2000). Optical-biophysical relationships of

707 609-620.

- Gilmanov, T.G., Verma, S.B., Sims, P.L., Meyers, T.P., Bradford, J.A., Burba, G.G., Suyker,
 A.E. (2003). Gross primary production and light response parameters of four Southern
 Plains ecosystems estimated using long-term CO₂-flux tower measurements. *Global Biogeochem. Cycles*, 17, 1071, doi:10.1029/2002GB002023.
- 712 Göckede, M., Foken, T., Aubinet, M. et al. (2008). Quality control of CarboEurope flux data -
- Part 1: Coupling footprint analyses with flux data quality assessment to evaluate sites in
 forest ecosystems. *Biogeosciences*, 5, 433-450.
- Gough, C.M., Vogel, C.S., Schmid, H.P., Su, H.-B., Curtis, P.S. (2008). Multi-year
 convergence of biometric and meteorological estimates of forest carbon
 storage. Agricultural and Forest Meteorology, 148, 158-170.
- Gu, L., Meyers, T., Pallardy, S.G. et al. (2007). Influences of biomass heat and biochemical
 energy storages on the land surface fluxes and radiative temperature. *Journal of Geophysical Research*, D02107, doi:10.1029/2006JD007425.
- Gu, L., Meyers, T., Pallardy, S.G. et al. (2006). Direct and indirect effects of atmospheric
 conditions and soil moisture on surface energy partitioning revealed by a prolonged
 drought at a temperate forest site. *Journal of Geophysical Research*, *D16102*,
 doi:10.1029/2006JD007161.
- Hagen, S.C., Braswell, B.H., Linder, E., Frolking, S., Richardson, A.D., Hollinger, D.Y.
 (2006). Statistical uncertainty of eddy-flux based estimates of gross ecosystem carbon
 exchange at Howland Forest, Maine. Journal of Geophysical Research, 111, D08S03,
 doi:10.1029/2005JD006154.

- Heinsch, F.A., Zhao, M., Running, S.W. (2006). Evaluation of remote sensing based terrestrial
 productivity from MODIS using regional tower eddy flux network observations. *IEEE Transactions on Geosciences and, Remote Sensing*, 44, 1908-1925.
- Hoblen, B.N. (1986). Characteristics of maximum value composite images from temporal
 VHRR data. International Journal of Remote Sensing, 7, 1417–1434.
- 734 Hollinger, D.Y., J.T., К., Goltz. S.M., Davidson. E.A., Lee, Tu, 735 Valentine, H.T. (1999). Seasonal patterns and environmental control of carbon 736 dioxide and water vapor exchange in an ecotonal boreal forest. Global 737 Change Biology, 5, 891-902
- Hollinger, D.Y., Aber, J., Dail, B. et al. (2004). Spatial and temporal variability in
 forest-atmosphere CO₂ exchange. *Global Change Biology*, 10, 1689-1706.
- Hollinger, S.E., Bernacchi, C.J., Meyers, T.P. (2005). Carbon budget of mature no-till
 ecosystem in North Central Region of the United States. Agricultural and Forest
 Meteorology, 130, 59-69, 2005.
- 743 Huang, C., Townshend, J.R.G. (2003). A stepwise regression tree for nonlinear approximation:
- applications to estimating subpixel land cover. International Journal of Remote Sensing,
 24, 75-90.
- 746 Huete, A., Didan, K., Miura, T., Rodríguez, E.P., Gao, X., & Ferreira, L.G. (2002). Overview
- 747 of the radiometric and biophysical performance of the MODIS vegetation indices. Remote
- 748 Sensing of Environment, 83, 195-213.
- 749 Irvine, J., Law, B.E., Hibbard, K.A. (2007). Postfire carbon pools and fluxes in semiarid
- ponderosa pine in Central Oregon. Global Change Biology, 13, 1748-1760.

- Jackson, T.J., Chen, D., Cosh, M., Li, F., Anderson, M., Walthall, C., Doriaswamy, P., Hunt,
 E.R. (2004). Vegetation water content mapping using Landsat data derived normalized
 difference water index fro corn and soybeans. *Remote Sensing Environment*, 92, 475-482.
- 754 Jenkins, J.P., Richardson, A.D., Braswell, B.H., Ollinger, S.V., Hollinger, D.Y., Smith, M.-L.
- 755 (2007). Refining light-use efficiency calculations for a deciduous forest canopy using
- rsf
 rsf
- Law, B.E. (2006). Carbon dynamics in response to climate and disturbance: recent progress
 from multiscale measurements and modeling in AmeriFlux. In S. Yatnamoto, ed. Plant
 Responses to Air Pollution and Global Change. Springer, Tokyo, Japan.
- Law, B.E., Sun, O.J., Campbell, J., Tuyl, S.V., & Thornton, P.E. (2003). Changes in carbon
 storage and fluxes in a chronosequence of ponderosa pine. *Global Change Biology*, 9, 510 524.
- 764 Law, B.E., Turner, D., Campbell, J., Sun, O.J., Van Tuyl, S., Ritts, W.D., & Cohen, W.B.
- 765 (2004). Disturbance and climate effects on carbon stocks and fluxes across Western
 766 Oregon USA. Global Change Biology, 10, 1429-1444.
- 767 Law, B.E., Turner, D., Lefsky, M., Campbell, J., Guzy, M., Sun, O., Van Tuyl, S., Cohen, W.
- 768 (2006). Carbon fluxes across regions: Observational constraints at multiple scales. In J.
- 769 Wu, B. Jones, H. Li, O. Loucks, eds. Scaling and Uncertainty Analysis in Ecology:
- 770 Methods and Applications. Springer, USA. Pages 167-190.
- 771 Lipson, D.A., Wilson, R.F., Oechel, W.C. (2005). Effects of Elevated Atmospheric CO₂ on
- 772 Soil Microbial Biomass, Activity, and Diversity in a Chaparral Ecosystem. Applied and
- 773 Environmental Microbiology, 71, 8573-8580.

774	Lloyd, J., Farquhar, G.D. (1996). The CO ₂ dependence of photosynthesis, plant growth
775	responses to elevated atmospheric CO_2 concentrations, and their interaction with soil
776	mutrient status: I. General principles and forest ecosystems. Functional Ecology, 10 , $4 - 32$.
777	Lutes, D.C., Keane, J.F., Caratti, C.H., Key, C.H., Benson, N.C., & Gangi, L.J. (2004).
778	FIREMON: Fire Effects Monitoring and Inventory System (US Department of Agriculture
779	Forest Service, Rocky Mountain Research Station, Ogden, UT), Vol RMRS-GTR-164-CD,
780	p 400.
781	Ma, S., Baldocchi, D.D., Xu, L., & Hehn, T. (2007). Inter-annual variability in carbon dioxide
782	exchange of an oak/grass savanna and open grassland in California. Agricultural Forest

- 783 Meteorology, 147, 157-171.
- Mäkelä, A., Pulkkinen, M.A., Kolari, P. et al. (2008). Developing an empirical model of stand
 GPP with the LUE approach: analysis of eddy covariance data at five contrasting confier
 sites in Europe. *Global Change Biology*, 14, 92-108.
- McGarvey, R.C., Martin, T.A., & White, T.L. (2004). Integrating within-crown variation in net
 photosynthesis in lobiolly and slash pine families. *Tree Physiology*, 24, 1209-1220.
- 789 Minasny, B., McBratney, A.B. (2008), Regression rules as a tool for predicting soil properties
- from infrared reflectance spectroscopy. Chemometrics and Intelligent Laboratory Systems,94, 72-79.
- 792 Monson, R.K., Turnipseed, A.A., Sparks, J.P., Harley, P.C., Scott-Denton, L.E., Sparks, K., &
- 793 Huxman, T.E. (2002). Carbon sequestration in a high-elevation, subalpine forest. Global
- 794 Change Biology, 8, 459-478.

- Moffat, A.M., Papale, D., Reichstein, M. et al. (2007). Comprehensive comparison of gapfilling techniques for eddy covariance net carbon fluxes. Agricultural Forest Meteorology,
 147, 209-232.
- 798 Myneni, R.B., Dong, J., Tucker, C.J., Kaufmann, R.K., Kauppi, P.E., Liski, J., Zhou, L.,
- 799 Alexeyev, V., Hughes, M.K. (2001). A large carbon sink in the woody biomass of northern
- 800 forests Proc. Natl. Acad. Sci. U.S.A., 98, 14784-14789.
- 801 Myneni, R.B., Hoffman, S., Knyazikhin, Y. et al. (2002). Global products of vegetation leaf
- 802 area and fraction absorbed PAR from year one of MODIS data. Remote Sensing of
 803 Environment, 83, 214-231.
- 804 Noormets, A., Chen, J., Crow, T.R., 2007. Age-dependent changes in ecosystem carbon fluxes
- 805 in managed forests in northern Wisconsin, USA. *Ecosystems* 10, 187-203.
- 806 Noormets, A., Desai, A.R., Cook, B.D., Euskirchen, E.S., Ricciuto, D.M., Davis, K.J., Bolstad,
- 807 P.V., Schmid, H.P., Vogel, C.V., Carey, E.V., Su, H.B., Chen, J., 2008a. Moisture
- 808 sensitivity of ecosystem respiration: Comparison of 14 forest ecosystems in the Upper
- 809 Great Lakes Region, USA. Agricultural and Forest Meteorology 148, 216-230.
- 810 Noormets, A., McNulty, S.G., DeForest, J.L., Sun, G., Li, Q., Chen, J., 2008b. Drought during
- 811 canopy development has lasting effect on annual carbon balance in a deciduous temperate
- 812 forest. New Phytologist: doi: 10.1111/j.1469-8137.2008.02501.x.
- 813 Noormets, A., McNulty, S.G., Gavazzi, M.J., Sun, G., Domec, J.C., King, J., Chen, J., 200X.
- 814 High drought tolerance of carbon fluxes in a coastal plain loblolly pine forest. Global
- 815 Change Biology, submitted.
- 816 Oak Ridge National Laboratory Distributed Active Archive Center (ORNL DAAC) (2006).
- 817 MODIS subsetted land products, Collection 4. Available on-line

- 818 (http://www.daac.ornl.gov/MODIS/modis.html) from ORNL DAAC, Oak Ridge,
- 819 Tennessee, U.S.A. Accessed Month 06, 2007.
- 820 Oren, R., Ewers, B.E., Todd, P., Phillips, N., & Katul, G. (1998). Water balance delineates the
 821 soil layer in which moisture affects canopy conductance. *Ecological Applications*, 8, 990-
- 822 1002.
- Oren, R., Hsieh, C.-I., Stoy, P., Albertson, J., McCarthy, H.R., Harrell, P., & Katul, G.G.
 (2006). Estimating the uncertainty in annual net ecosystem carbon exchange: spatial
 variation in turbulent fluxes and sampling errors in eddy-covariance measurements. *Global Change Biology*, 12, 883-896.
- Papale, D., & Valentini, A. (2003). A new assessment of European forests carbon exchange by
 eddy fluxes and artificial neural network spatialization. *Global Change Biology*, 9, 525535.
- Pataki, D.E., & Oren, R. (2003). Species difference in stomatal control of water loss at the
 canopy scale in a bottomland deciduous forest. Advances in Water Research, 26, 12671278.
- Penuelas, J., Gamon, J.A., Griffin, K.L., & Field, C.B. (1993). Assessing community type,
 plant biomass, pigment composition, and photosynthetic efficiency of aquatic vegetation
 from spectral reflectance. *Remote Sensing of Environment*, 46, 110-118.
- Persson, P., Hall-Könyves, K. Sjöström, G., Pinzke, S. (1993). NOAA/AVHRR data for crop
 productivity estimation in Sweden. Adv. Space Res., 13, 111-116.
- 838 Peters, W., Jacobson, A.R., Sweeney, C. et al. (2007). An atmospheric perspective on North
- 839 American carbon dioxide exchange: carbontracker. Proc Natl Acad Sci USA, 104, 18925-
- 840 18930 (2007).

- 841 Potter, C., Klooster, S., Huete, A., & Genovese, V. (2007). Terrestrial carbon sinks for the
 842 United States predicted from MODIS satellite data and ecosystem modeling. *Earth*
- 843 Interactions, 11, 1-21.
- 844 Poweil, T.L., Gholz, H.L., Clark, K.L., Starr, G., Cropper, W.P., Martin, T.A. (2008). Carbon
- 845 exchange of a mature, naturally-regenerated pine forest in north Florida. Global Change
- 846 Biology, 14, 2523-2538, doi: 10.1111/j.1365-2486.2008.01675.x.
- 847 Prince, S.D., Goward, S.N. (1995). Global primary production: a remote sensing approach.
 848 Journal of Biogeography, 22, 815-835.
- 849 PRISM Climate Group, Oregon State University, http://www.prismclimate.org, created 4 Feb
 850 2004.
- 851 Rahman, A.F., Sims, D.A., Cordova, V.D., El-Masri, B.Z. (2005). Potential of MODIS EVI
- and surface temperature for directly estimating per-pixel ecosystem C fluxes. *Geophysical Research Letters*, 32, L19404, doi:10.1029/2005GL024127.
- 854 Ranson, K.J., Daughtry, C.S.T., Biehl, L.L., Bauer, M.E. (1985). Sun-view angle effects on
- 855 reflectance factors of corn canopies. Remote Sensing of Environment, 18, 147-161.
- 856 Reichstein, M., Falge, E., Baldocchi, D. et al. (2005). On the separation of net ecosystem
- 857 exchange into assimilation and ecosystem respiration: review and improved algorithm.
- 858 Global Change Biology, 11, 1424-1439.
- 859 Richardson, A.D., Hollinger, D.Y., Dail, D.B., Lee, J.T., Munger, J.W., O'Keefe, J. (2008a).
- 860 Influence of spring phenology on seasonal and annual carbon balance in two contrasting
- 861 New England forests. *Global Change Biology*, under review.

- 862 Richardson, A.D., Mahecha, M.D., Falge, E. et al. (2008b). Statistical properties of random
- 863 CO₂ flux measurement uncertainty inferred from model residuals. Agricultural and Forest
 864 Meteorology, 148, 38-50.
- 865 Ruimy, A., Jarvis, P.G., Baldocchi, D.D., & Saugier, B. (1995). CO2 fluxes over plant
- 866 canopies and solar radiation: a review. Advances in Ecological Research, 26, 1-68.
- 867 RuleQuest, 2008. http://www.rulequest.com. Visited on 10/18/2007.
- 868 Running, S.W., Baldocchi, D.D., Turner, D.P., Gower, S.T., Bakwin, P.S., & Hibbard, K.A.
- 869 (1999). A global terrestrial monitoring network integrating tower fluxes, flask sampling,
- 870 ecosystem modeling and EOS satellite data. Remote Sensing of Environment, 70, 108-127.
- 871 Running, S.W., Nemani, R.R., Heinsch, F.A., Zhao, M., Reeves, M., & Hashimoto, H. (2004).
- 872 A continuous satellite-derived measure of global terrestrial primary production.
 873 BioScience, 54, 547-560.
- 874 Ryan, M.G., Binkley, D., Fownes, J.H., Giardina, C.P., Senock, R.S. (2004). An experimental
- 875 test of the causes of forest growth decline with stand age. *Ecological Monographs*, 74, 393876 414.
- 877 Schmid, H.P. (1994). Source areas for scalars and scalar fluxes. *Boundary Layer Meteorology*,
 878 67, 293-318.
- 879 Schmid, H.P. (2002). Footprint modeling for vegetation atmospheric exchange studies.: A
 880 review and perspective. Agricultural and Forest Meteorology, 113, 159-183.
- 881 Schmid, H.P., Grimmond, C.S.B., Cropley, F., Offerle, B., & Su, H.B. (2000). Measurements
- 682 of CO₂ and energy fluxes over a mixed hardwood forest in the mid-western United States.
- 883 Agricultural and Forest Meteorology, 103, 357-374.

- Schmidt, K.S., & Skidmore, A.K. (2003). Spectral discrimination of vegetation types in a
 coastal wetland. *Remote Sensing of Environment*, 85, 92-108.
- 886 Sellers, P.J., Randall, D.A., Betts, A.K. et al. (1997). Modeling the exchanges of energy, water,
- and carbon between continents and the atmosphere. *Science*, 275, 502-509.
- 888 Shroyer, J.P., Thompson, C., Brown, R., Ohlenbach, P.D., Fjell, D.L., Staggenborg, S.,
- 889 Duncan, S., & Kilgore, G.L. (1996). Kansas crop planting guide. Publication, Vol. L-818
- 890 (pp. 2). Manhattan, KS: Kansas State University.
- 891 Sims, D.A., Rahman, A.F., Cordova, V.D. et al. (2005). Midday values of gross CO₂ flux and
- 892 light use efficiency during satellite overpasses can be used to directly estimate eight-day
- 893 mean flux. Agricultural and Forest Meteorology, 131, 1-12.
- 894 Stoy, P.C., Katul, G.G., Siqueira, M.B.S., Juang, J.Y., Novick, K.A., Uebetherr, J.M., Oren, R.
- (2006). An evaluation of models for partitioning eddy covariance-measured net ecosystem
 exchange into photosynthesis and respiration. Agricultural and Forest Meteorology, 141,
 2-18.
- 898 Thompson, J.R., Spies, T.A., Ganio, L.M. (2007). Reburn severity in managed and unmanaged
- vegetation in a large wildfire. Proceedings of the National Academy of Sciences of the
 United States of America, 104, 10743-10748.
- 901 Tucker, C.J., Vanpraet, C.L., Sharman, M.J., Van Ittersum, G. (1985). Satellite remote sensing
 902 of total herbaceous biomass production in the Senegalese Sahel: 1980-1984. Remote Sens.
 903 Environ., 17, 233-249.
- 904 Urbanski, S., Barford, C., Wofsy, S. et al. (2007). Factors controlling CO2 exchange on
- 905 timescales from hourly to decadal at Harvard Forest. Journal of Geophysical Research,
- 906 *112*, G02020, doi:10.1029/2006JG000293.

- 907 Verma, S.B., Dobermann, A., & Cassman, K.G. (2005). Annual carbon dioxide exchange in
 908 irrigated and rainfed maize-based agroecosystems. Agricultural and Forest Meteorology,
 909 131, 77-96.
- 910 Vermote, E.F., Vermeulen, A. (1999). MODIS Algorithm Technical Background Document --
- 911 Atmospheric Correction Algorithm: Speciral Reflectances (MOD09), Version 4.0.
- 912 http://modis.gsfc.nasa.gov/data/atbd/atbd_mod08.pdf.
- 913 Vermote, E.F., Kotchenova, S.Y. (2008). MOD09 (Surface Reflectance) User's Guide, Version
 914 1.1, March, 2008, http://modis-sr.ltdri.org.
- 915 Wan, Z., Zhang, Y., Zhang, Q., & Li, Z.-L. (2002). Validation of the land-surface temperature
- 916 products retrieved from Terra Moderate Resolution Imaging Spectroradiometer data.
- 917 Remote Sensing of Environment, 83, 163-180.
- Waring, R.H., & Franklin, J.F. (1979). Evergreen coniferous forests of the Pacific Northwest.
 Science, 204, 1380-1386.
- 920 Waring, R.H., Landsberg, J.J., & Williams, M. (1998), Net primary production of forests: A
- 921 constant fraction of gross primary production? *Tree Physiology*, 18, 129–134.
- 922 Watts, C.J., Scott, R.L., Garatuza-Payan, J., Rodriguez, J.C., Prueger, J., Kustas, W., Douglas,
- 923 M. (2007). Changes in vegetation condition and surface fluxes during NAME 2004.
- 924 Journal of Climate, 20, 1810-1820.
- 925 Wofsy, S.C., Goulden, M.L., Munger, J.W., Fan, S.-M., Bakwin, P.S., Daube, B.C., Bassow,
- 926 S.L., & Bazzaz, F.A. (1993). Net exchange of CO₂ in a mid-latitude forest. Science, 260,
- 927 1314-1317.

- Wylie, B.K., Fosnight, E.A., Gilmanov, T.G. et al. (2007). Adaptive data-driven models for
 estimating carbon fluxes in the Northern Great Plains. *Remote Sensing of Environment*,
 106, 399-413.
- 931 Xiao, J., Zhuang, Q., Baldocchi, D.D. et al. (2008). Estimation of net ecosystem carbon
- 932 exchange for the conterminous United States by combining MODIS and AmeriFlux data.

933 Agricultural and Forest Meteorology, 148, 1827-1847.

- 934 Xiao, J., Zhuang, Q., Liang, E., McGuire, A.D., Moody, A., Kicklighter, D.W., Shao, X.,
- 935 Melillo, J.M. (2009). Twentieth-century droughts and their impacts on terrestrial carbon
- 936 cycling in China. Earth Interactions, 13, Paper No. 10, 1-31, DOI: 10.1175/2009EI275.1
- 937 Xu, L., & Baldocchi, D.D. (2004). Seasonal variation in carbon dioxide exchange over a
- 938 Mediterranean annual grassland in California. Agricultural and Forest Meteorology, 123,
 939 79-96.
- 940 Yang, L., Huang, C., Homer, C., Wylie, B.K., Coan, M.J. (2003). An approach for mapping
- 941 large-area impervious surfaces: synergistic use of Landsat-7 ETM+ and high spatial
 942 resolution imagery. *Canadian Journal of Remote Sensing*, 29, 230-240.
- 943 Yang, F., Ichii, K., White, M.A., Hashimoto, H., Michaelis, A.R., Votava, P., Zhu, A.-X.,
- 944 Huete, A., Running, S.W., & Nemani, R.R. (2007). Developing a continental-scale measure
- 945 of gross primary production by combining MODIS and AmeriFlux data through support
- 946 machine approach. *Remote Sensing of Environment*, 110, 109-122.
- 947 Zhang, L., Wylie, B., Loveland, T., Fosnight, E., Tieszen, L.L., Ji, L., & Gilmanov, T. (2007).
- 948 Evaluation and comparison of gross primary production estimates for the Northern Great
- 949 Plains grasslands. Remote Sensing of Environment, 106, 173-189.

95 0	Zhang, Y., Yu, Q., Jiang, J., Tang, Y. (2008). Calibration of Terra/MODIS gross primary								
951	production over an irrigated cropland on the North China Plain and an alpine meadow on								
952	the Tibetan Plateau. Global Change Biology, 14, 757-767.								
953	Zhao, M., Heinsch, F.A., Nemani, R.R., & Running, S.W. (2005). Improvements of the								
954	MODIS terrestrial gross and net primary production global data set. Remote Sensing of								
955	Environment, 95, 164-175.								
956	Zhao, M., Running, S.W., & Nemani, R.R. (2006). Sensitivity of moderate resolution imaging								
957	spectroradiometer (MODIS) terrestrial primary production to the accuracy of								
958	meteorological reanalyses. Journal of Geophysical Research, 111, G01002,								
959	doi:10.1029/2004JG000004.								
960									
961									
962									
963									
964									
965									
966									
967									
968									
969									
970									
971									
972									

973 Table 1. Site descriptions including name, latitude, longitude, vegetation type, years of data

974 available, and references for each flux site.

Site	State	Lat	Loa	Vegetation type	Year	References
Audubon Research	AZ	31.59	-110.51	Grasslands	2002-2006	
Ranch (ARR)						
Santa Rita Mesquite (SRM)	AZ	31.82	-110.87	Savannas	2004-2006	Watts et al. 2007
Walnut Gulch	AZ	31.74	-109.94	Grassiands	2004-2006	
Kendall Grasslands (WGK)						
Sky Oaks Old Stand (SOO)	CA	33.37	-116.62	Shrublands	2004-2006	Lipson et al. 2005
Sky Oaks Young stand (SOY)	CA	33.38	-116.62	Shrublands	2001-2006	Lipson et al. 2005
Tonzi Ranch (TR)	CA	38.43	-120.97	Savannas	2001-2006	Ma et al. 2007
Vaira Ranch (VR)	CA	38.41	-120.95	Grasslands	2001-2006	Xu and Baldocchi 2004
Niwot Ridge Forest (NRF)	co	40.03	-105.55	Evergreen forests	2000-2003	Monson et al. 2002
Kennedy Space Center -Scrub Oak (KSC)	FL.	28.61	-80.67	Shrublands	2000-2006	Dore et al. 2003
Austin Cary - Slash Pine (AC)	瓧	29.74	-82.22	Evergreen forests	2001-2005	Powell et al. 2008
Bondville (Bon)	IL	40.01	-88.29	Croplands	2001-2006	Hollinger et al. 2005
FNAL agricultural site (FAg)	IL.	41.86	-88.22	Croplands	2005-2006	
FNAL Prairie site (FPr)	IL.	41.84	-88.24	Grasslands	2004-2006	
Morgan Monroe State Forest (MMS)	IN	39.32	-86.41	Deciduous forests	2000-2005	Schmid et al. 2000
Harvard Forest EMS Tower (HFE)	MA	42.54	-72.17	Deciduous forests	2000-2004	Urbanski et al. 2007
Harvard Forest Hemlock Site (HFH)	MA	42.54	-72.18	Evergreen forests	2004	
Little Prospect Hill (LPH)	МА	42.54	-72.18	Deciduous forests	2002-2005	
Howland forest (HF)	ME	45.20	-68.74	Evergreen forests	2000-2004	Hollinger et al., 1999, 2004
Howland forest (west tower) (HFW)	MĒ	45.21	-68.75	Deciduous forests	2000-2004	Hollinger et al., 1999, 2004
Sylvania Wilderness Area (SWA)	м	46.24	-89.35	Mixed forests	2002-2006	Desai et al., 2005
Univ. of Mich. Biological Station (UMB)	МІ	45.56	-84.71	Mixed forests	2000-2003	Gough et al., 2008
Missouri Ozark (MQ)	мо	38.74	-92.20	Deciduous forests	2004-2006	Gu et al. 2006, 2007
Goodwin Creek (GC)	MS	34.25	-89.97	Grasslands	2002-2006	
Fort Peck (FPe)	MŤ	48.31	-105.10	Grassiands	2000-2006	
Duke Forest loblolly pine (DFP)	NC	35.98	-79.09	Evergreen forests	2001-2005	Oren et al. 1998, 2006

Duke Forest hardwoods (DFH)	NC	35.97	-79.10	Deciduous forests	2003-2005	Pataki and Oren, 2003
North Carolina toblolly pipe (NCP)	NC	35.80	-76.67	Evergreen forests	2005-2006	Noormets et al. submitted
Mead irrigated continuous maize site (MIC)	NE	41.17	-96.48	Croplands	2001-2005	Verma et al. 2005
Mead inigated	NE	41.16	-9 6.47	Croplands	2001-2005	Verma et al. 2005
Mead rainfed (MR)	NE	41.18	-96.44	Croplands	2001-2005	Verma et al. 2005
Bartlett Experimental Forest (BEF)	NH	44.06	-71.29	Deciduous forests	2004-2005	Jenkins et al., 2007
Toledo Oak Openings (TOO)	QН	41.55	-83.84	Savannas	2004-2005	Noormets et al. 2008b
ARM Oklahoma (ARM)	ОК	36.61	-97.49	Croplands	2003-2006	
Metolius intermediate aged ponderosa pine (MI)	OR	44,45	-121.56	Evergreen forests	2003-2005	Law et al. 2003; Irvine et al. 2007
Metolius new young pine (MN)	OR	44.32	-121.61	Evergreen forests	2004-2005	Law et al. 2003; Irvine et al. 2007
Brookings (Bro)	SD	44.35	-96.84	Grasslands	2004-2006	
Freeman Ranch Mosquite Juniper (FRM)	ΤX	29.95	-98.00	Savannas	2004-2006	
Wind River Crane Site (WRC)	WA	45.82	-121.95	Evergreen forests	2000-2004	Falk et al., 2008
Lost Creek (LC)	WI	46.08	-89.98	Deciduous forests	2000-2005	
Willow Creek (WC)	WI	45.81	-90.08	Deciduous forests	2000-2006	Cook et al., 2004
Wisconsin intermediate	WI	46.73	-91.23	Deciduous forests	2003	Noormets et al. 2008a
Wisconsin mature red pine (MRP)	WI	46 .74	-91.17	Evergreen forests	2002-2005	Noormets et al. 2007

- - -

-

Table 2. Spatially averaged (spatial mean) and integrated (spatial total) annual GPP over the
period 2001-2006 for each region: Northeast (NE), North Central (NC), Southeast (SE), South
Central (SC), Rocky Mountain (RM), Pacific Northwest (PNW), and Pacific Southwest
(PSW).

	GPP	NE	NC	SE	SC	RM	PNW	PSW	US
	Spatial Mean	1604.20	1212.55	2033.91	1457.97	438.20	1007.35	927.77	1103.92
	(g C m ⁻² yr ⁻¹)								
	Spatial Total	0.67	2.00	1.00	1.93	0.81	0.36	0.29	7.06
	(Pg C yr' ¹)								
988			<u>_</u>						
989									
9 90									
991									
992									
993									
994									
995									
996									
997									
998									
999									

1000 Table 3. Spatially averaged (spatial mean) and integrated (spatial total) annual GPP over the

1001 period 2001-2006 for each vegetation type: evergreen forests (EF), deciduous forests (DF),

1002	mixed forests (MF), shrublands (Sh), savannas (S	Sa), grasslands (Gr), and croplands (Cr).
------	--------------------------------------------------	-------------------------------------------

GPP	EF	DF	MF	Sh	Sa	Gr	Cr	All
Spatial Mean	1431.26	1774.74	1447.31	303.78	1500.96	589.24	1500.38	1103.92
(g C m ⁻² yr ⁻¹)								
Spatial Total	0.85	0.83	0.67	0.33	0.63	0.83	2.92	7.06
$(Pg C yr^{-1})$								

Figure captions:

Fig. 1. Scatterplots of observed 8-day GPP versus predicted 8-day GPP. (a) Our estimate (y = 0.85x + 0.37, $R^2 = 0.74$, p < 0.0001). (b) MODIS GPP product (Running et al. 2004) (y = 0.50x + 1.01, $R^2 = 0.58$, p < 0.0001). For each plot, the solid line is the 1:1 line, and the dashed line is the regression line.

Fig. 2. Examples of time series plots of observed (open circles) and predicted (solid circles) 8day GPP (g C m⁻² day⁻¹) for each AmeriFlux site over the period 2005-2006: (1) evergreen forests – AC (FL) and MRP (WI); (2) deciduous forests – DFH (NC); (3) mixed forests – SWA (MI); (4) shrublands – KSC (FL); (5) savannas – SRM (AZ), TR (CA), and VR (CA); (6) grasslands – ARR (AZ); (7) croplands – MIC (NE). For x-axis, the starting dates (month/day) of every two 8-day intervals are provided in parentheses under interval numbers. Dashed lines are used to separate 2005 from 2006. Site abbreviations are used here, and their full names are given in Table 1.

Fig. 3. Scatterplot of observed mean 8-day GPP versus predicted mean 8-day GPP across the AmeriFlux sites. Error bars are standard errors (defined as the standard deviation divided by the square root of the number of observations) of the observed and predicted 8-day mean GPP. The solid line indicates the 1:1 line, and the dashed line indicates the regression line (y = 0.95 * x + 0.21, $R^2 = 0.84$, p < 0.0001). Site abbreviations are used here, and their full names are given in Table 1.

Fig. 4. Examples of leave-one-out cross-validation scatterplots with observed 8-day GPP versus predicted 8-day GPP: (1) evergreen forests – HF (y = 0.85x+1.00, $R^2 = 0.91$, p < 0.0001) and NRF (y = 0.67x + 0.63, $R^2 = 0.63$, p < 0.0001); (2) deciduous forests – HFE (y = 0.97x + 0.24, $R^2 = 0.87$, p < 0.0001) and WC (y = 0.73x + 0.62, $R^2 = 0.86$, p < 0.0001); (3)

mixed forests – SWA (y = 0.89x + 0.03, $R^2 = 0.89$, p < 0.0001) and UMB (y = 0.92x + 0.66, $R^2 = 0.93$, p < 0.0001); (4) shrublands – KSC (y = 0.28x + 2.35, $R^2 = 0.14$, p < 0.0001); (5) savannas – FRM (y = 0.83x + 1.94, $R^2 = 0.49$, p < 0.001); (6) grasslands – GC (y = 0.76x + 2.03, $R^2 = 0.55$, p < 0.0001); (7) croplands – MR (y = 1.04x + 0.81, $R^2 = 0.86$, p < 0.0001). The solid line indicates the 1:1 line, and the dashed line indicates the regression line. Site abbreviations are used here, and their full names are given in Table 1.

Fig. 5. Leave-one-out cross-validation scatterplot of observed mean 8-day GPP versus predicted mean 8-day GPP across the AmeriFlux sites. Error bars are standard errors of the observed and predicted 8-day GPP. The solid line indicates the 1:1 line, and the dashed line indicates the regression line (y = 0.73 * x + 1.07, $R^2 = 0.69$, p < 0.0001). Site abbreviations are used here, and their full names are given in Table 1.

Fig. 6. Monthly GPP (g C m^{-2} mo⁻¹) for the conterminous U.S. from January through December in 2005.

Fig. 7. Spatially averaged and integrated 8-day GPP for each biome across the conterminous U.S. over the period 2001-2006. (a) Spatially averaged 8-day GPP (g C m⁻² day⁻¹). (b) Spatially integrated 8-day GPP (Tg C day⁻¹).

Fig. 8. Average annual GPP (g C m⁻² yr⁻¹) of the conterminous U.S. over the period 2001-2006. The gray lines indicate state boundaries. The black lines indicate boundaries of geographical regions: Northeast (NE), Southeast (SE), North Central (NC), South Central (SC), Rocky Mountain (RM), Pacific Northwest (PNW), and Pacific Southwest (PSW).

Fig. 9. Annual GPP (g C m^2 yr⁻¹) for the conterminous U.S. for 2005. (a) Our estimate. (b) The MODIS GPP product (MOD17A3; Running et al. 2004).

Fig. 10. Annual GPP anomalies (g C m⁻² yr⁻¹) and annual precipitation anomalies (mm) for the conterminous U.S. for 2002 and 2006. The anomalies of annual GPP were relative to the 6-year period from 2003 to 2006, and the anomalies of annual precipitation were relative to the 30-year period from 1971 to 2000 taken from the PRISM climate database (PRISM Climate Group 2004).

Fig. 11. Impact of the Biscuit Fire (> 2000 km²) in Oregon on annual GPP in 2004. (a) Burned area. (b) The dots represent fire detections from Terra MODIS and Aqua MODIS MODIS (USDA Forest Service MODIS Active Fire Mapping Program, http://activefiremaps.fs.fed.us). (c) Burn severity based on the difference normalized burn ratio (dNBR; Lutes et al. 2004) from Landsat Thematic Mapper (TM) data acquired before and immediately after the fire. (d) Annual GPP in 2003 (g C m⁻² yr⁻¹).

Fig. 12. Impact of Hurricane Katrina on annual GPP in 2006. The units are g C m^{-2} yr⁻¹. The white lines indicate the isotachs, including tropical storm, hurricane category 1, and hurricane category 2.



Fig. 1. Scatterplots of observed 8-day GPP versus predicted 8-day GPP. (a) Our estimate (y = 0.85x + 0.37, $R^2 = 0.74$, p < 0.0001). (b) MODIS GPP product (Running et al. 2004) (y = 0.50x + 1.01, $R^2 = 0.58$, p < 0.0001). For each plot, the solid line is the 1:1 line, and the dashed line is the regression line.



Fig. 2. Examples of time series plots of observed (open circles) and predicted (solid circles) 8day GPP (g C m² day⁻¹) for each AmeriFlux site over the period 2005-2006: (1) evergreen forests – AC (FL) and MRP (WI); (2) deciduous forests – DFH (NC); (3) mixed forests – SWA (MI); (4) shrublands – KSC (FL); (5) savannas – SRM (AZ), TR (CA), and VR (CA); (6)

grasslands – ARR (AZ); (7) croplands – MIC (NE). For x-axis, the starting dates (month/day) of every two 8-day intervals are provided in parentheses under interval numbers. Dashed lines are used to separate 2005 from 2006. Site abbreviations are used here, and their full names are given in Table 1.



Fig. 3. Scatterplot of observed mean 8-day GPP versus predicted mean 8-day GPP across the AmeriFlux sites. Error bars are standard errors (defined as the standard deviation divided by the square root of the number of observations) of the observed and predicted 8-day mean GPP. The solid line indicates the 1:1 line, and the dashed line indicates the regression line (y = 0.95 * x + 0.21, $R^2 = 0.84$, p < 0.0001). Site abbreviations are used here, and their full names are given in Table 1.



Fig. 4. Examples of leave-one-out cross-validation scatterplots with observed 8-day GPP versus predicted 8-day GPP: (1) everygen forests – HF (y = 0.85x+1.00, $R^2 = 0.9$), p < 0.9

0.0001) and NRF (y = 0.67x + 0.63, $R^2 = 0.63$, p < 0.0001); (2) deciduous forests – HFE (y = 0.97x + 0.24, $R^2 = 0.87$, p < 0.0001) and WC (y = 0.73x + 0.62, $R^2 = 0.86$, p < 0.0001); (3) mixed forests – SWA (y = 0.89x + 0.03, $R^2 = 0.89$, p < 0.0001) and UMB (y = 0.92x + 0.66, $R^2 = 0.93$, p < 0.0001); (4) shrublands – KSC (y = 0.28x + 2.35, $R^2 = 0.14$, p < 0.0001); (5) savannas – FRM (y = 0.83x + 1.94, $R^2 = 0.49$, p < 0.001); (6) grasslands – GC (y = 0.76x + 2.03, $R^2 = 0.55$, p < 0.0001); (7) croplands – MR (y = 1.04x + 0.81, $R^2 = 0.86$, p < 0.0001). The solid line indicates the 1:1 line, and the dashed line indicates the regression line. Site abbreviations are used here, and their full names are given in Table 1.



Fig. 5. Leave-one-out cross-validation scatterplot of observed mean 8-day GPP versus predicted mean 8-day GPP across the AmeriFlux sites. Error bars are standard errors of the observed and predicted 8-day GPP. The solid line indicates the 1:1 line, and the dashed line indicates the regression line (y = 0.73 * x + 1.07, $R^2 = 0.69$, p < 0.0001). Site abbreviations are used here, and their full names are given in Table 1.



Fig. 6. Monthly GPP (g C m^{-2} mo⁻¹) for the conterminous U.S. from January through December in 2005.



Fig. 7. Spatially averaged and integrated 8-day GPP for each biome across the conterminous U.S. over the period 2001-2006. (a) Spatially averaged 8-day GPP (g C m⁻² day⁻¹). (b) Spatially integrated 8-day GPP (Tg C day⁻¹).



Fig. 10. Annual GPP anomalies (g C m^{-2} yr⁻¹) and annual precipitation anomalies (nm) for the conterminous U.S. for 2002 and 2006. The anomalies of annual GPP were relative to the 6-year period from 2001 to 2006, and the anomalies of annual precipitation were relative to the 30-year period from 1971 to 2000 taken from the PRISM climate database (PRISM Climate Group 2004).



Fig. 11. Impact of the Biscuit Fire (> 2000 km²) in Oregon on annual GPP in 2004. (a) Burned area. (b) The dots represent fire detections from Terra MODIS and Aqua MODIS MODIS (USDA Forest Service MODIS Active Fire Mapping Program, http://activefiremaps.fs.fed.us). (c) Burn severity based on the difference normalized burn ratio (dNBR; Lutes et al. 2004) from Landsat Thematic Mapper (TM) data acquired before and immediately after the fire. (d) Annual GPP in 2003 ($g C m^2 yr^{-1}$).



Fig. 12. Impact of Hurricane Katrina on annual GPP in 2006. The units are g C m⁻² yr⁻¹. The white lines indicate the isotachs, including tropical storm, hurricane category 1, and hurricane category 2.