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Comments

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Near-Real-Time Measurement of Sea-Salt Aerosol during the SEAS Campaign: Comparison of Emission-Based Sodium Detection with an Aerosol Volatility Technique

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ABSTRACT

The first deployment of an emission-based aerosol sodium detector (ASD), designed to chemically characterize marine aerosols on a near-real-time basis, is reported. Deployment occurred as part of the Shoreline Environment Aerosol Study (SEAS) from 16 April to 1 May 2000 at Bellows Air Force Base on the east side of Oahu, where the University of Hawaii's Department of Oceanography maintains a tower for aerosol measurements. The instrument was operated in size-unsegregated mode and measurements were made that included two extended continuous sampling periods, each of which lasted for 24 h. During this time, the ASD was compared with measurements that used aerosol volatility coupled with optical particle counting to infer sea-salt size distributions. A reasonable agreement was obtained between the instruments when sampling in clean air, suggesting that under these conditions both approaches can provide reliable sea-salt distributions. The combination of these measurements suggested that sea salt was the dominant constituent of aerosol particles with diameters larger than 500 nm and that sulfate was the dominant constituent at smaller diameters during clean air sampling.

1. Introduction

It is clear that attempts to better understand and quantify both the direct and indirect forcing effects of atmospheric aerosols have changed our view of their role in climate change (Prospero 2002). While many early studies focused exclusively on sulfate aerosols (Charlson et al. 1992), it is now clear that sea salt, mineral dust, and organic aerosols play a significant effect in direct scattering of solar radiation (Piliinis et al. 1995; Satheesh et al. 1999). Initial studies of the indirect effects of forcing via their impact on cloud formation and cloud droplet size distribution again focused on sulfate aerosol based on the assumption that this was the dominant source of cloud condensation nuclei (CCN) in the marine boundary layer (Charlson et al. 1987). In recent years this view has been challenged, and the potential role of sea salt and organic compounds has been a subject of several investigations. An early study using electron microscopy identified submicron sea-salt particles in membrane filter samples based on its cubic structure (Meszaros and Vissy 1974). A number of subsequent

studies used more direct techniques and found the fraction of submicron sea salt to be extremely small. Radke and Hobbs (1969) used a flame emission instrument to simultaneously measure altitude profiles of CCN and sodium-containing particles in the Olympic Mountains in Washington State. They concluded that the Pacific Ocean was the main source of sodium-containing particles, but these constituted less than 1% of the active CCN. A similar study by Hobbs (1971) over the Pacific Ocean that sampled between sea level and 10 000 ft again concluded that less than 1% of the active CCN contained sodium.

The physical and chemical properties of aerosols are extremely variable, and much of the current, very limited database was obtained with indirect measurements such as aerosol volatility. This approach was used by Clarke (1991) and Clarke and Porter (1993) to distinguish between sea salt, sulfuric acid, and sulfate aerosols. In measurements in the equatorial Pacific they measured sea-salt aerosol concentrations of typically 10 cm^{-3} for the fraction of particles with diameters below $0.5 \mu\text{m}$. This was typically about 10% of the total particle concentration in this size fraction. In more recent work, O'Dowd et al. (1993, 1997, 1999) used aerosol volatility to examine submicron sea-salt production and its dependence on wind speed. They suggested that a

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high wind speeds sea-salt aerosol may provide the primary source of CCN, even under sulfate-rich conditions. One of the problems in assessing the relative roles of sea salt and sulfate aerosol is related to the difficulty of measuring in situ “chemically resolved” aerosol size distributions.

Direct measurements of size-segregated individual aerosol particles using laser ablation coupled with analysis by time-of-flight mass spectrometry provide contrasting results. Measurements at Cape Grim during the first Aerosol Characterization Experiment (ACE-1) reported that almost all aerosols larger than the 0.13- μm instrument detection limit contained some sea salt (Murphy et al. 1998). In contrast, measurements during the Indian Ocean Experiment (INDOEX) campaign employed a similar technique and found little sea-salt contribution to the submicron aerosol (Guazzotti et al. 2001). Both sampled clean air but with a significant difference in the wind speed and hence the sea-salt environment. Laser ablation coupled with mass spectrometry provides a powerful tool for analysis of single particles; however, it is a semi-quantitative technique. In this work we describe the field deployment of instrumentation that is designed to provide quantitative, rapid, near-real-time analysis of the sodium content of marine aerosols in an effort to increase our understanding of the factors that control submicron sea-salt distributions. The Shoreline Environment Aerosol Study (SEAS) deployment offered the opportunity to compare this method with measurements that used aerosol volatility techniques coupled with optical particle counting to infer sea-salt size distributions. The campaign took place from 16 April to 1 May 2000 at Bellows Air Force Base on the east side of Oahu, where the University of Hawaii (UH) Department of Oceanography maintains a tower for aerosol measurements (cf. Clarke and Kapustin 2003).

2. Instrumentation

The basic design and calibration of the aerosol sodium detector (ASD) is described in Clark et al. (2001), which contains a schematic of the instrument. The basic principle of operation of the ASD is the volatilization of aerosol particles in a high-temperature flame, atomization of the sodium salts to give sodium atoms, and detection of the emission at 589.0 (D_2 line) and 589.6 nm (D_1 line) from thermally excited sodium atoms. The ASD consists of an aerosol sampling and injection system to introduce aerosol particles into the flame, a pre-mixed laminar hydrogen/air flame for volatilization of the aerosol and atomization of sodium salts, and photomultiplier tubes (PMTs) that detect the emission and associated electronics for data acquisition. It is possible to monitor size-resolved aerosols by sampling through a differential mobility analyzer (DMA), however, as noted below, size segregation was not used for work reported here. The field instrument deployed in the SEAS

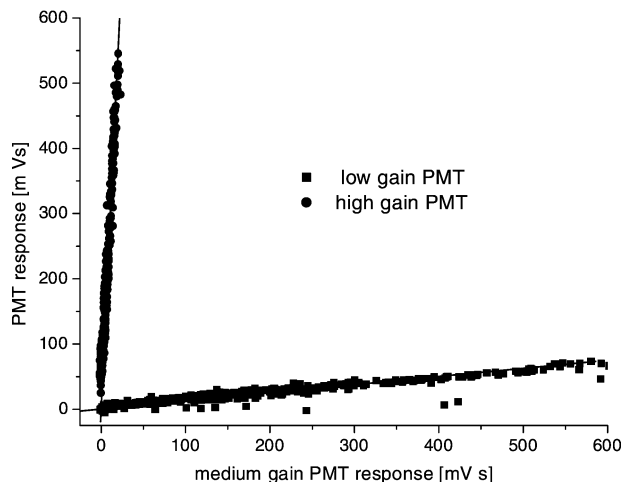


FIG. 1. Correlations of the signal response of the three detectors for one typical SEAS sample.

campaign was a more rugged, improved version of the ASD described previously, modified to employ a 50% larger sampling volume. In addition, a smaller burner head allowed for faster flame velocities, attaining higher flame temperatures while reducing total gas consumption. Entrainment of sodium aerosols from the surrounding air into the ASD flame was particularly troublesome in the trailer in which the ASD was located at the SEAS site. In order to discriminate against coincidental or spurious aerosols from around the burner, a makeshift aluminum chimney was constructed on site. This shielded the flame quite effectively, but $\sim 1\%$ of the sample could be identified as entrained laboratory particles. This problem precluded the use of size segregation since a much lower count rate was obtained after passage through the DMA. Under these conditions spurious counts from entrained particles dominated the signal. Because the instrument's probed quantity is mass, a large dynamic range is needed to cover the aerosol size range of interest. By combining three PMTs with different gain settings, we achieved a dynamic range of about 2000 (180–2300-nm equivalent dry-NaCl diameter). The detection volume in this configuration is better defined, improving discrimination against spurious aerosol events. Figure 1 shows the correlation between the three PMTs for a typical emission dataset taken during SEAS. For all measurements during SEAS the differential mobility analyzer was not used to select a particular size aerosol; thus the ASD sampled the aerosol size distribution transmitted by the inlet system. Hence we measure a distribution of emission signals that are proportional to the sodium content of the aerosol particle. However, we have no information on the actual mass of the particle that produced the emission signal. The emission signals are converted to an absolute sodium mass using the calibration procedure described below, and size distributions are obtained by converting

sodium mass to an equivalent volume of “dry sodium chloride” based on an assumed density.

a. Calibration

The ASD is calibrated with monodisperse aerosols of known sizes, produced by a vibrating orifice aerosol generator (VOAG; TSI Model 3050; Berglund and Liu 1973). The VOAG produces droplets with a known sodium content that are then dried to give sea-salt aerosol particles of a known mass. Particles are produced with diameters between 0.2 and 3 μm , providing an absolute calibration standard for the emission signal. In a variety of tests described previously (Clark et al. 2001) we found that ASD emission signals were linear, increasing with increasing sodium concentration as expected. A linear response of the ASD to variation in droplet sodium concentration demonstrates that we are able to reproducibly vary the initial droplet sodium concentration, that volatilization of the aerosols is complete and independent of sodium concentration, and that the flame remains optically thin. In addition, we demonstrated the absence of any chemical interference effects.

After the completion of the SEAS campaign the instrument underwent an extensive recalibration that addressed issues relating to absolute calibration, transmission, and data acquisition. Generation of well-defined monodisperse aerosol particles of variable sodium concentration is critical for ASD calibration. The initial size of the droplet as a geometric diameter produced by the VOAG can be calculated from (Westenberger et al. 1990)

$$D_d = \sqrt[3]{\frac{6Q}{\pi f}}, \quad (1)$$

where f is the frequency of orifice vibration (s^{-1}), Q is the liquid flow rate ($\text{cm}^3 \text{ s}^{-1}$), and D_d is the droplet diameter (cm). The absolute amount of sodium is calculated from the droplet volume and solution concentration. Hence the absolute concentration calibration depends critically on an accurate knowledge of the initial droplet diameter. An error in the droplet size calibration will produce a linear response in emission versus calculated sodium mass but with an error in the absolute sodium mass. To increase reproducibility of the VOAG aerosol output, an He–Ne laser beam was focused on the liquid jet, about 8 mm downstream from the vibrating orifice, and the absorption pattern was monitored with a photodiode, allowing the droplet breakup process to be monitored in real time. Under correct operation a uniform absorption trace was observed with a frequency that matched the driving frequency of the VOAG crystal. All calibration experiments were performed while simultaneously monitoring the droplet size. The VOAG drying tube was modified to run at very high dilution flows and hence low relative humidity (RH). It was possible to reach RH as low as 20% under these con-

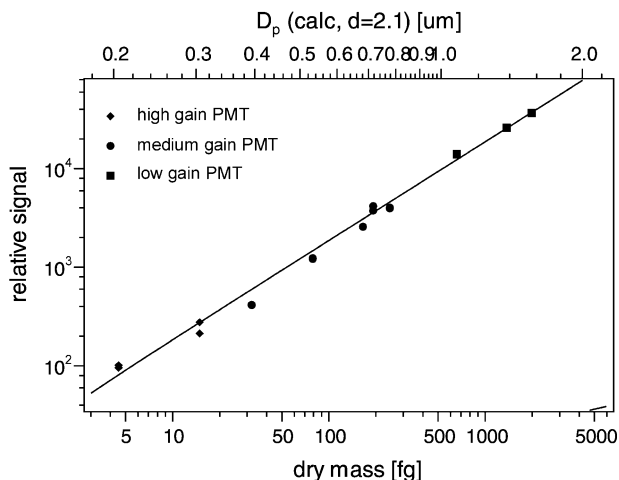


FIG. 2. An 11-point calibration of the ASD over the dynamic range of the three photomultipliers, normalized to the response of the highest-gain detector. Both dry NaCl mass and calculated diameter (for a dry density of 2.1) of the synthetic aerosols are plotted.

ditions. In order to achieve good counting statistics, actual calibrations were taken at an RH of 45%–50%. Provided that the absorption trace is clean, the output of the VOAG is always monodisperse. Under these conditions we monitored the sodium emission signal as a function of sodium concentration. Figure 2 shows the relative emission signal for all three PMTs, normalized to the sensitivity of the highest-gain PMT and plotted against the calculated dry mass of sodium. The linearity over almost three orders of magnitude is quite remarkable. Standard errors were in the range of 3% but are mainly caused by the width of the “monodisperse” VOAG distribution. As noted above, this demonstrates the linearity of the detection but does not validate the absolute mass calibration since errors in the calculation of droplet volume would produce a linear response. To ensure that the initial droplet volume calculation was correct, we attempted to completely dry the aerosols and compare the calculated and measured particle sizes. Figure 3 shows the correlation of the calculated VOAG aerosol sizes with the actual sizes measured with an aerosol particle sizer (APS; TSI Model 3310). The aerodynamic particle sizes measured by the APS were corrected to give geometric diameters assuming spherical particles, and theoretical sizes were calculated assuming the density of the aerosol to be that of pure sea salt, 2.1 g cm^{-3} (Tang 1997). The measured particle sizes agree well with those calculated assuming dry sea salt. As a result of these measurements we discovered that our original SEAS calibration, and hence a preliminary report of a SEAS distribution (Clark et al. 2001), was in error by a factor of 2 in mass. Since a cubic factor relates mass and diameter, the reported distribution, given as an equivalent dry NaCl diameter, has calculated diameters that are a factor of 1.3 too low. As noted in the

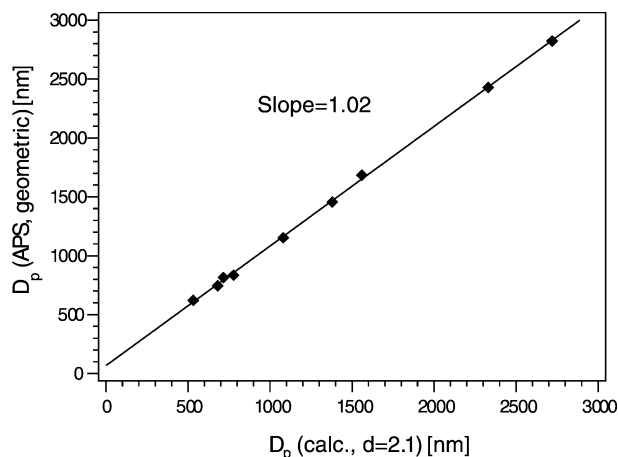


FIG. 3. Correlation between calculated dry sizes for VOAG-generated aerosols and the sizes measured with the APS, assuming dry aerosols.

paper, this distribution was not corrected for transmission efficiency through the ASD.

The strong correlation of calculated and measured diameters independent of size and sizing instrument gives us confidence that our current absolute calibration is accurate. Thus we can extrapolate a theoretical detection limit of the ASD of around 150 nm in equivalent dry NaCl diameter. The current operational detection limit is set by the need to reliably trigger the data acquisition sequence off an emission event. At SEAS it was necessary to raise the triggering threshold for the acquisition electronics in order to discriminate against small room aerosols as well as false triggers, and the operational detection limit was around 200 nm in equivalent dry NaCl diameter.

b. Transmission efficiency

The transmission efficiency of the ASD was determined as a function of particle diameter by sampling monodisperse VOAG particles of various diameters and comparing the ASD count rate per cycle with the measured particle number density. Small particles were passed through a DMA (TSI Model 3081) set at the maximum of the distribution, and measurements were taken at the output with both the ASD and a condensation particle counter. Most VOAG particles are multiply charged, so transmission through the DMA was low and counting statistics were poor. Particles above 900 nm were measured with the APS and ASD in similar plumbing configurations. Figure 4 summarizes the measured transmission ratios.

Overall, transmission is fairly constant between 500 and 2000 nm, perhaps increasing slightly below 500 nm. The falloff between 2000 and 3000 nm is to be expected from the deposition velocities for such aerosols in our injection volume (84 cm long, 7.1-mm inner diameter, 16-s residence time). Transmission was measured

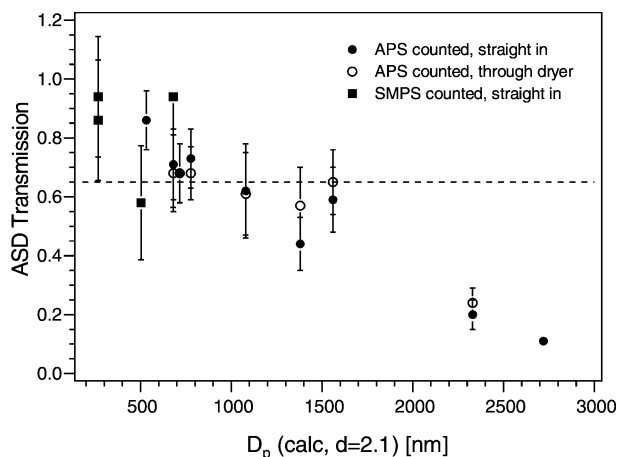


FIG. 4. Transmission efficiencies of the ASD for different sizes, categorized by counting method. Also, for some sizes, the efficiency of aerosols dried through the nafion dryer is shown. The black dotted line shows the average value used for analysis of the SEAS data.

through a nafion dryer with similar results. For the analysis of the SEAS data, in particular the comparison with aerosol volatility/optical particle counter (OPC) measurements, an average value of 65% was taken. Since we did not measure a distribution of aerosol particle diameters directly we felt a more detailed correction for transmission was not appropriate.

c. Sampling configuration at SEAS

Aerosols were sampled at the tower at Bellows at a height of 12 m above sea level. The inlet used in this study was a 3/8-in. inner diameter black-carbon-rich silicone tube pointing into the wind, drawing 11 L min^{-1} , of which 0.4 L were dried through a 60-cm-long 1/4-in. steel nafion dryer (Permapure Inc., Model MD-24) and drawn into the ASD. The actual sampling velocity at the top of the inlet was 2.5 m s^{-1} . At typical wind speeds of 6–8 m s^{-1} , such an inlet should slightly oversample big particles. Because the ASD is not designed for detection of particles larger than 3 μm aerodynamic diameter, oversampling should not be a concern as long as there are no other losses.

Some tests were run in order to validate the inlet transmission. On two different days, a scanning mobility particle sizer (SMPS) and an APS were set up to sample the air going into the ASD injection loop. These measurements were compared to aerosol measurements using an APS on the sampling tower (Clarke et al. 2003). The inlet for the aerosol instruments was 2 m below the inlet for the ASD. Figure 5 compares the data from the APS at the ASD injection loop with APS measurements taken on the tower. The tower APS measurements are the average of a 120-min sampling period compared with a 5-min sampling period at the ASD inlet that was taken 1 h earlier. In the range above 1- μm diameter, where the APS performs most reliably, the agreement

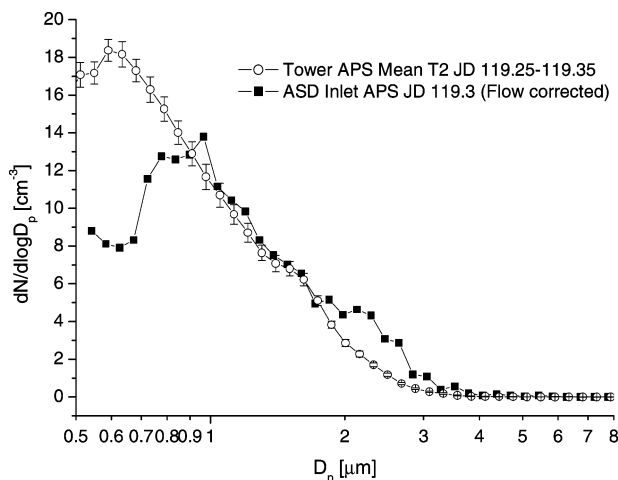


FIG. 5. Comparison of data taken with an APS at the ASD inlet injection loop (sampling from tower) with an APS located on the sampling tower, 2 m below the ASD sampling inlet. Aerodynamic diameters are shown.

between the tower APS and the injection loop APS is excellent up to $2 \mu\text{m}$. Beyond $2 \mu\text{m}$, the injection loop APS measurements were higher than the tower APS measurements, but this might be caused by poor counting statistics since the injection loop APS sampling period was 6 times shorter than that of the tower APS. This may also be the source of the discrepancy between 0.5 and $0.8 \mu\text{m}$, since, as we show below, agreement was reasonable below $0.5 \mu\text{m}$.

Figure 6 compares SMPS (TSI Model 3936) measurements at the injection loop with tower measurements. The agreement is satisfactory, so oversampling and diffusional losses were probably not an issue in this region. We conclude that our inlet provided a representative sample of outside air for particles between 100 and $2\text{-}\mu\text{m}$ diameter.

3. Results

a. Time frame

The ASD operated continuously for two 24-h periods, one starting at 1100 LST 26 April 2000 and the other at 1100 LST 28 April 2000 [Julian days (JDs) 117.875–118.875 and 119.875–120.875, respectively]. Samples were taken on a 2-min cycle time that consisted of 1 min for sampling and 1 min for analysis. Each sample gave 50–100 sodium signals. To improve counting statistics, 10 injection cycles were combined, giving an effective time resolution of 20 min. During JD 118 a front came over Bellows Beach (Clarke and Kapustin 2003), and during the second part of that day the origin of the air mass changed, although local conditions such as wind and humidity remained unchanged. Accordingly we measured both undisturbed marine as well as post-frontal aerosol.

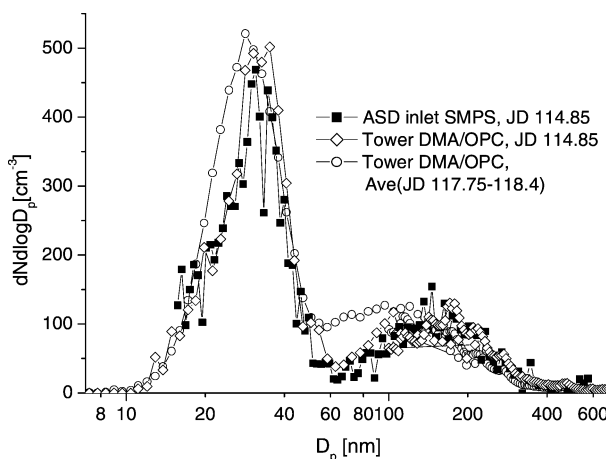


FIG. 6. As in Fig. 5, but comparing data taken with our SMPS through our inlet with UH data. Also shown is the average of the UH data during our first 24-h measurement period.

b. Analysis

Instrumental problems precluded analysis of 3 h of data from the first day of measurements. All other datasets were analyzed using the corrected calibrations taken on JDs 119 and 121. As stated, the gain ratios calculated from the photomultiplier correlations in each dataset were stable within 10% over the course of the 2 days, as shown in Fig. 1. There was a small long-term drift on the order of 5% in the sodium mass calibration, probably due to small instabilities in the flame flow or warming of the photodetectors, and the calculated equivalent NaCl diameters were corrected for this.

Each emission signal was inspected prior to integration, in order to discriminate against coincidental or spurious aerosols from around the burner. As noted above, $\sim 1\%$ of the sample could be identified as contaminant particles based on the shape of the emission traces and the correlation between the PMT signals. Very small emission signals with doubtful traces were discarded, which might have resulted in our undercounting at or near the detection limit.

The equivalent NaCl aerosol diameters, calculated using a density of 2.1 , were converted into a 25-“channel” aerosol number distribution with a step width, $d\log D_p = 0.05$. This number of bins seemed to be the best compromise between resolution and optimizing counting statistics. Figure 7 shows such a calculated distribution for a typical sodium mass dataset. It also shows the impact of uncertainties in the sodium mass calibration for a worst-case scenario. The statistical calibration error is around $\pm 6\%$, and the 20% mass calibration uncertainty shown in the figure reflects the variation of calibration results over a year of measurements, including two complete realignments. It translates, on average, to a 10%–20% error for $dN/d\log D_p$, except at the edges of the distribution, as discussed below. Overall, most distributions were fairly similar to the one shown, with

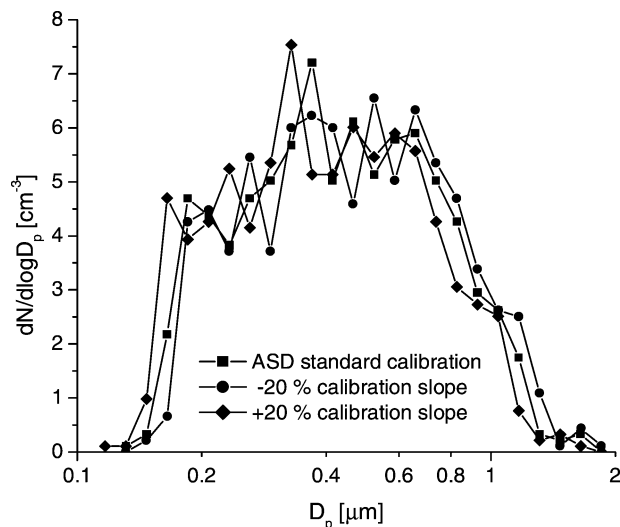


FIG. 7. Typical sodium size distribution for the second day of measurements (JD 120) showing the effect of the 20% mass calibration uncertainty on size.

statistical errors on the order of 10%–20%. In order to discern trends, we examined the temporal variability of a three-bin distribution (<500, 500–1000, >1000 nm) for all 123 available datasets, as shown in Fig. 8. The concentrations of aerosols bigger than 500 nm remained nearly constant over the course of the 2 days. This was expected, as most of those particles came from sea spray and wind speeds were fairly constant. We saw no tidal dependency in particle concentration. Aerosol particle concentrations dropped twice in response to rain events, on JDs 118.6 and 118.82, in all sizes ranges. Total particle concentrations below 500 nm increased starting around JD 118.3 and remained at higher levels until the last third of the second day of measurements. This most likely reflects the reported postfrontal change in air mass during JD 118.

Based on the preliminary trends in the sub-500-nm range, the data was subdivided into four time periods: JD 118.07–118.25 (prefront), JD 118.3–118.75 (new air mass, rain events were stripped), JD 119.875–120.25 (same aerosol count, 1 day later), and JD 120.4–120.875 (smaller aerosol load). For each time period, all sodium masses were aggregated into one dataset and one size distribution generated, as described. Errors resulted from the statistical counting error and the calibration error ($\pm 15\%$). Area and volume distributions were also calculated from the number distribution using average diameters. Figure 9 shows the resulting distributions.

4. Discussion

The distributions from Fig. 9 all show a steep decline below 210 nm that is due to aerosol undercounting around the triggering threshold and a conservative discrimination against small potentially spurious emission

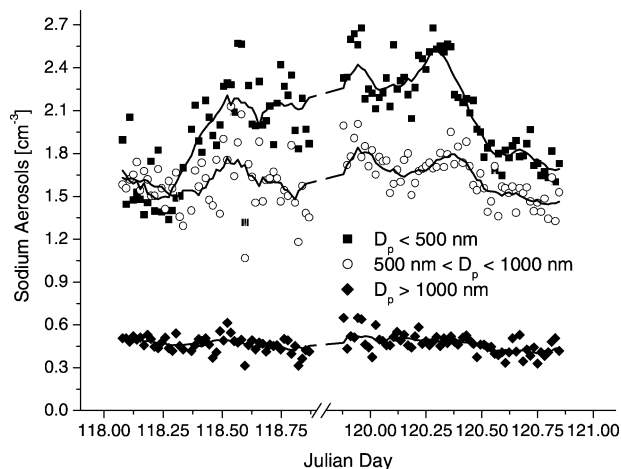


FIG. 8. Number of sodium-containing aerosols measured per 20-min sampling cycle over the 48-h measurement period during SEAS, broken down by equivalent size.

signals during raw data analysis. Above 1.1 μm , on the other hand, both the volume and area distributions seem to fall off short of the expected mode maximum for sea salt around 3 μm (Clarke et al. 2003). This is somewhat lower than the 2.0–2.5- μm cutoff that the transmission measurements with VOAG aerosols suggest, but, as size was not monitored separately, this might well be due to issues such as the lower density of sea salt versus pure NaCl dried aerosols, longer drying times, or the presence of other salts. This would lead in all cases to a bigger particle diameter at the same equivalent NaCl diameter and, hence, lower transmission.

Because the ASD operated in an unsized mode, that is, without a DMA for size segregation, the distributions shown in Fig. 9 are plotted as equivalent dry sodium diameters. Hence the sodium signal could, in principle, arise from an internally mixed aerosol of larger diameter with a fractional sea-salt composition. However, as we discuss below, Fig. 10 shows that the ASD distributions are in good agreement with the refractory (i.e., sea salt) distributions obtained by the OPC in clean air, particularly above 700 nm. This correlation of OPC refractory particle counts with the ASD distributions suggests that the amount of internally chemically mixed aerosols was low and that the ASD distributions shown in Fig. 9 represent particles that consist largely of sea salt. The number distributions change with time, increasing during the second half of the first measurement period, corresponding to an increase in wind speed. We find the number distributions peaking at 500 nm. In all four cases we measure submicrometer sea salt down to our current detection limit of 200 nm.

Comparison with OPC data

The SEAS experiments offered the opportunity to compare the ASD measurements with heated OPC mea-

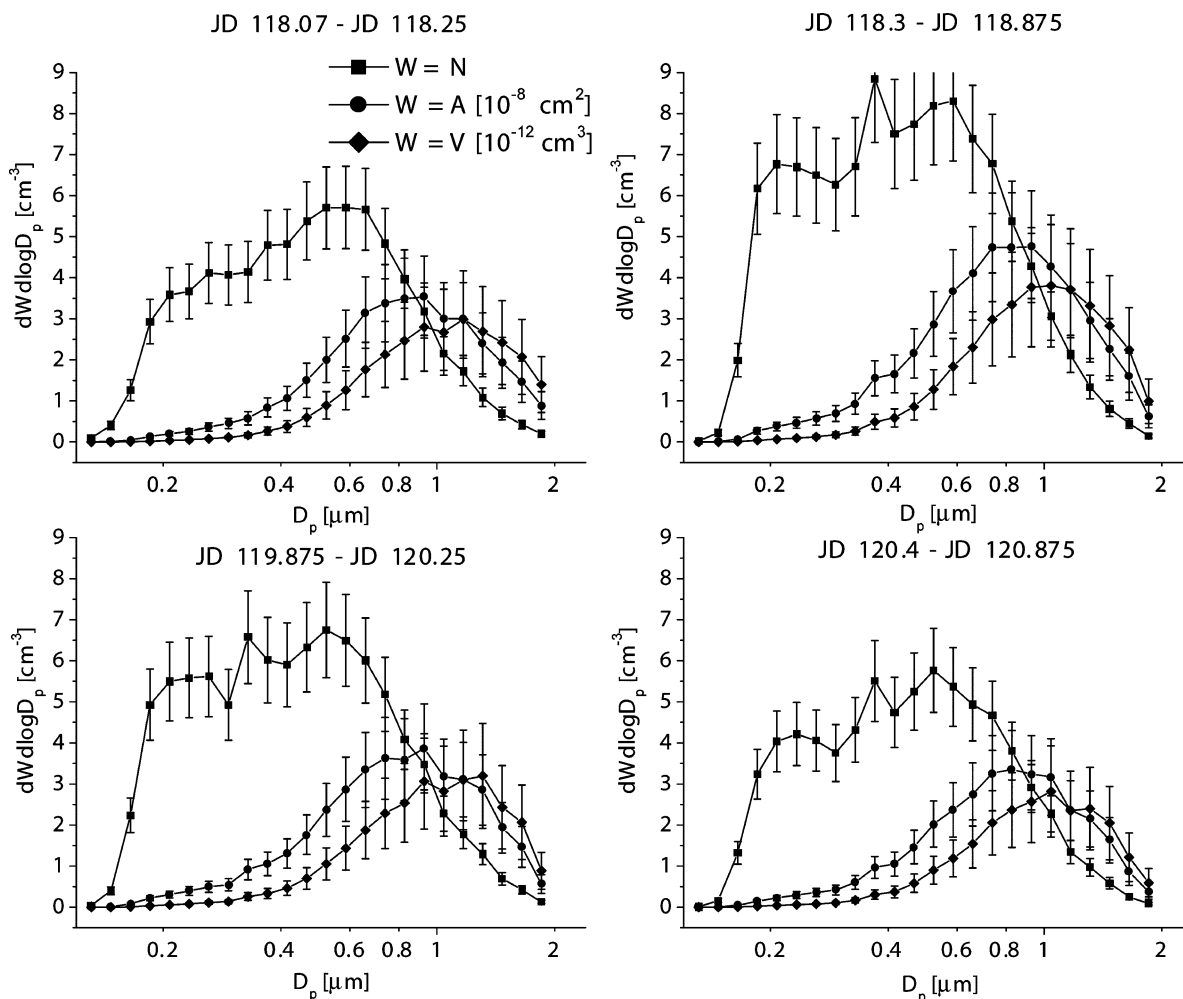


FIG. 9. Averages for $dN/d\log D_p$, $dA/d\log D_p$, and $dV/d\log D_p$ for sodium-containing particles during the four time periods shown in Fig. 8. Particle diameters are equivalent dry sodium diameters.

measurements. In the OPC system ambient particle counts are compared with those obtained after heating the particles to 300°C. It is assumed that all volatile components such as water and sulfate are removed and a size distribution of the nonvolatile components such as sea salt and dust can be obtained. Since the ASD operated in a non-size-resolved mode, comparison with size-resolved distributions is essential for the interpretation of our data.

The OPC data were averaged over the same time periods used to analyze the ASD data. No measurements are available for most of JD 119, so no comparison could be drawn for this time period. Figure 10 compares the averages of the OPC distributions heated to 50°C (dry ambient particles) and 300°C with our data for the first time periods.

There is fair agreement between the heated OPC distribution and the ASD data between 500 and 2000 nm. The fine structure at 750 nm in the OPC distribution is not real aerosol structure but is caused by Mie oscil-

lations. The ASD appears to undercount particles with diameters larger than 1100 nm, but counting statistics are poor. However, both datasets agree within error limits; an agreement that, considering the assumptions involved in the calculation of the ASD distributions, is quite remarkable. As noted above, most particles in this range seem to be 100% sea salt.

Between 250 and 500 nm, the agreement is again reasonable for this period when we sampled clean air. It suggests that the heated OPC measurements reliably reflect the sea-salt distribution in the absence of significant concentrations of mineral dust aerosols. It further indicates that sulfate aerosol particles are the dominant constituent of the total aerosol distribution in this diameter range. Figure 11 shows the comparison for two periods after the passage of the front. During this period the ASD counts are lower than the refractory OPC counts. The agreement is particularly poor below 500 nm, when the OPC counted up to 5 times more aerosols at 250 nm. Interpretation of this data is clearly com-

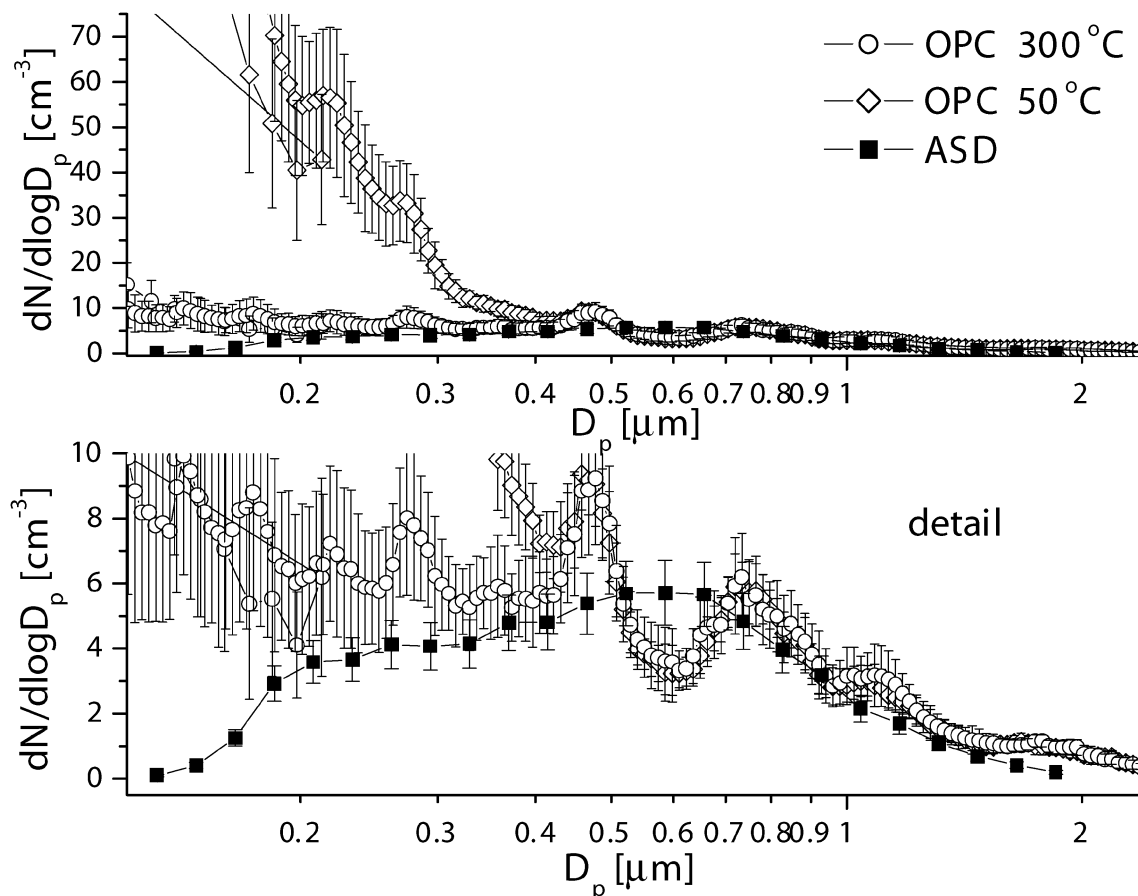


FIG. 10. Comparison of the average ASD size distributions (i.e., equivalent geometric diameter of dry sea salt) with the average of the OPC distributions for the unpolluted time period (JD 118.07–118.25), taken 2 m below the ASD inlet, after heating the aerosols to 50° and 300°C, respectively (cf. Clarke et al. 2003).

licated by the fact that the ASD operated in an unresolved sizing mode. However, we believe these differences are real and that the two instruments were measuring aerosol particles of different chemical composition. The most obvious explanation is that after the frontal passage the air mass contained a substantial fraction of refractory small particles that were not composed primarily of sea salt. There is some evidence for this in light absorption measurements (Masonis et al. 2003) and in refractory CN number for this period (A. Clarke 2002, personal communication). Without measurements of the chemical composition of these aerosols it is not possible to definitively determine the source of this discrepancy.

5. Conclusions

The measurements reported here represent the first field deployment of the ASD, which was demonstrated to be capable of providing an extensive dataset of sea-salt distributions that are quantitative and have high temporal resolution. The opportunity to deploy the instrument as part of a field campaign focusing exclusively on aerosol particle measurements was invaluable, par-

ticularly the comparison with a heated OPC instrument. We obtained reasonable agreement between the instruments when sampling in clean air, suggesting that under these conditions both approaches can provide reliable sea-salt distributions. The combination of these measurements indicates that sea salt was the dominant constituent of aerosol particles with diameters larger than 500 nm and that, based on volatility, sulfate was the dominant constituent at smaller diameters.

There were significant differences in measurements by the two instruments after the passage of a front that brought a polluted air mass into the site, suggesting that the instruments were measuring different things. One possible explanation is the presence of refractory aerosol particles of anthropogenic origin; however, size-resolved measurements from the ASD would do much to illuminate the origin of this discrepancy.

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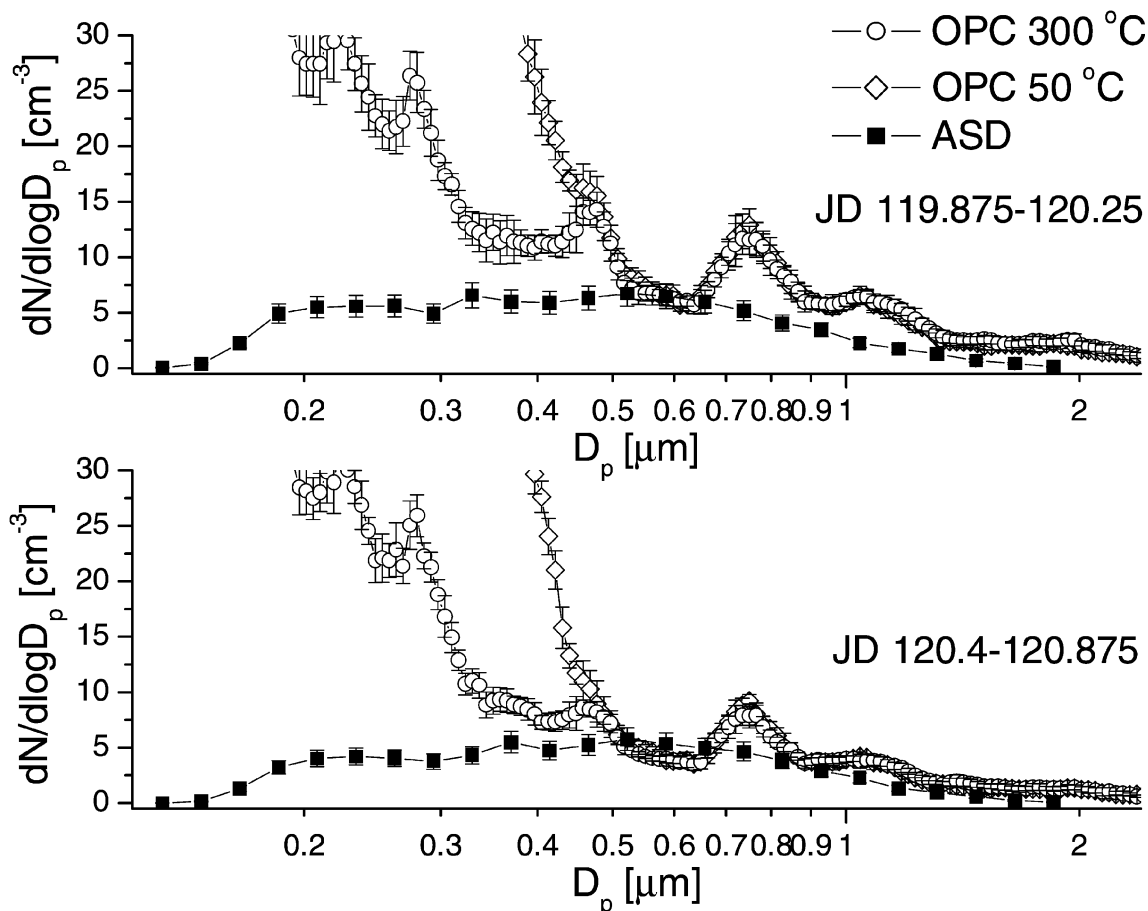


FIG. 11. As in Fig. 10, but for the second day of measurements.

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