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Fall 2008

Ecology, distribution, quantification, and impact of introduced, Asian Porphyra yezoensis f yezoensis Ueda and Porphyra yezoensis f narawaensis A Miura in the Northwestern Atlantic

Jeremy C. Nettleton *University of New Hampshire, Durham*

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ECOLOGY, DISTRIBUTION, QUANTIFICATION, AND IMPACT OF INTRODUCED, ASIAN *PORPHYRA YEZOENSIS* F. *YEZOENSIS* UEDA AND PORPHYRA YEZOENSIS F. NARAWAENSIS A. MIURA IN THE NORTHWESTERN ATLANTIC

BY

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BS Biology, Iowa State University, 1998

THESIS

Submitted to the University of New Hampshire in Partial Fulfillment of the Requirements for the degree of

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31 July 2008

Date

DEDICATION

To Brita, my darling wife and best friend

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Jeremy Nettleton

University of New Hampshire, September, 2008

Invasive species pose a threat to the balance of intertidal ecosystems. Recently, two forms of the non-native species, *Porphyra yezoensis* Ueda, were found at multiple sites between New York and Downeast Maine. A 2007 New England survey confirmed the presence of P. *yezoensis* f. *yezoensis* at nine sites, including two beyond its reported distribution. *Porphyra yezoensis f narawaensis* A. Miura was found at four sites in Long Island Sound. To assess the ecological impact of f. *yezoensis* and f. *narawaensis* on Northwest Atlantic macroalgal communities, monthly density and biomass data were gathered in 2008 from seven southern New England sites along 20 m transect lines. *P. yezoensis* f. *yezoensis* was not detected at two historic sites. The f. narawaensis has expanded to Cape Cod. Fucoid algae epiphytized by *P. yezoensis* demonstrated no stature reduction. A *Porphyra* species of cryptic origins, *P.* spp. *'stamfordensis,'* may be competing with *P. yezoensis.*

INTRODUCTION

Human consumption of *Porphyra* (nori) as food has occurred in Asia for over 1,000 years (Xia and Abbott, 1987). *Porphyra* is valued for its cholesterol regulating agent, taurine (Tsujii et al., 1983) and its high protein content (29-35% dry weight) which is 1.7 times higher by weight than beef (Arasaki and Arasaki, 1983). *Porphyra* is also an excellent source of vitamin A, being 67 times higher than found in eggs, and vitamin C, having 1.5 times more than in oranges (Xia and Abbott, 1987).

Porphyra has been a staple of healthy diets in Asia for centuries (Mumford and Miura, 1989). In China, *Porphyra* is eaten in several ways including: sushi; lightly fried and flavored with soy sauce, sugar, or sesame oil; in soups; with pork in dumplings; or stir fried with other vegetables and meat (Xia and Abbott, 1987).

Porphyra also has important medical and scientific uses in that it contains the phycobilin red pigment r-phycoerythrin, which is utilized as a fluorescent tag for labeling antibodies, proteins, and nucleic acids. Phycobiliprotein dyes can be used in applications such as immunofluorescence microscopy, microarrays, and flow cytometry.

The widespread production of *Porphyra* as a food stuff and fluorescent tag was not possible until its complete life history was understood. For hundreds of years before, farmers recognized the diploid stage *of Porphyra* (Figure 1), they 'farmed' the blades by rock cleaning and, to increase production, bamboo 'planting'(Tseng, 1984). Based upon experience, early Chinese and Japanese nori farmers readied their rocks rods at times of year they expected the arrival *of Porphyra* 'seeds' (Tseng, 1984). Heavy reliance on the

abundance of nature without fully understanding the developmental processes of *Porphyra* kept Asian farmers from producing industrial levels of nori.

Kathleen M. Drew's 1955 discovery of *'Conchocelis rosea'* as a microscopic life history phase *of Porphyra umbilicalis* Kiitz removed the largest obstacle to the successful phycoculture of various *Porphyra* species. With the understanding that the highly resilient, shell boring, diploid sporophytic conchocelis stage was the source of 'seed' for the valuable haploid gametophytic blade phase of *Porphyra,* large scale production of nori began in earnest in Asia in the late 1950s (Tseng, 1984). By placing nets seeded with conchospores from the cultured conchocelis into coastal waters, Asian nori farmers were able to boost production to unprecedented annual values, which, by the 1990s, neared US \$1.5 billion (FAO, 1997; Hanisak, 1998).

Not only have Asian nori farmers learned how to maximize the production of their native seaweeds, but they have developed, since the 1960s, fast growing cultivars of their most desirable *Porphyra,* including *P. tenera* Kjellman and *P. yezoensis* (Patwary and van der Meer, 1992). *Porphyra yezoensis* f. *narawaensis* A. Miura, which is now found in the northwest Atlantic, was developed from a single strain in the late 1960s at a nori farm in Narawa in the Chiba Prefecture of Japan (Niwa and Aruga 2003). Cultivars derived from this strain are highly prized in Japan for rapid growth, lengthy vegetative period, blade size (up to 1 m in length). When used in the production of hoshi-nori, its texture and flavor are deemed superior (Miura 1984). The smooth, dark-green to black, rectangular sheets can be eaten alone or as a wrapper for sushi containing vinegared rice, thinly sliced vegetables, and fish. By the late 1980s, because of their growth characteristics and quality, forma narawaensis cultivars were nearly the only ones grown

in Japanese nori-culture (Miura and Aruga 1987). Also at this time, Miura and Aruga (1987) determined that nori farming along the Japanese coast was so extensive as to be nearly saturated.

Because the production of nori in Asia was highly lucrative, scientists and speculators in the United States and Canada became interested in bringing commercial nori-culture to North America. In the late 1970s, *Porphyra* cultivation began in western North America under the impetus of Thomas Mumford, Jr. and J.E. Merrill at the University of Washington and the Washington State Department of Natural Resources (Merrill, 1981; Mumford, 1990). Merrill, having studied for a year under Miura in Tokyo, pushed for *Porphyra* cultivation in Washington State for two reasons: the coastal waters of Washington were nearly ideal year-round, whereas in Japan only the winter months were suitable for blade development; the sushi industry had begun to flourish in the US, driving up imports (Mumford, 1990). After extensive consultation and assistance from Japanese nori-culture experts, Merrill determined that *Porphyra* cultivation, using established technology and techniques, was possible in coastal Washington (Merrill, 1981). A Washington Department of Natural Resources study followed, which determined that the US could 'enter and compete in the market for products of the red seaweed *Porphyra''* (Kramer et al., 1982).

Nori farming at several sites in Washington US began in the early 1980s, using some native North American species and several cultivars imported from Japan (Mumford, 1987). Researchers and businessmen opted to use species from Japan, including *Porphyra yezoensis* f. *narawaensis,* because there was an established, global market for this species. They believed the cultivars posed little risk of permanent

introduction to the region due to two factors: they believed that, in Washington, the combination of coastal water temperatures and day lengths would not allow the conchocelis to reproduce; and they believed the area had likely been inoculated with Japanese *Porphyra* conchocelis, through the shells imported for oyster culture, for fifty years without establishment of the species (Mumford and Hansen, 1987; Conway et al., 1975).

Utilizing the techniques of modern Asian nori-culture, several private companies grew *Porphyra yezoensis* with some success in the late 1980s, including New Channel Nori in the San Juan Islands that produced the equivalent of nearly 500,000 nori sheets processed by Canada West Nori (Mumford, 1990). Because of the successes of the few establish nori-farming firms in Washington, Mumford estimated that the industry could have been well developed in that area by the year 2000.

Although the cultivation *of Porphyra* in western North America showed promise, unanticipated political and ecological obstacles interfered with the establishment of the industry. Individual coastal land owners and special interest groups fought against the permitting of nori farms based on ecological concerns (the potential introduction of new species and resulting impacts). They also argued that the floats and nets used in noriculture had a negative impact on coastal views and, therefore, property values. The grass roots pressure swayed the legislative process and resulted in severe permitting difficulties (Mumford 1990). When permits were given, it was determined that floating debris was a greater problem than anticipated. The region's unique hydrogeographic characteristics caused drifting logs, branches, plastics, and seaweeds to become entangled in the

Porphyra floats, which led to production inefficiencies and, in some cases, required the construction of expensive barrier systems (Mumford 1990).

Despite the setbacks in the nori farming attempts in western North America, scientists and entrepreneurs in New England attempted to cultivate *Porphyra* for economic purposes in the 1990s. Initially, Coastal Plantations International attempted to grow *Porphyra yezoensis* f. *narawaensis* in Downeast Maine for its potential use as a food product and for phycobilin pigment production. It was hoped the industry would be a financial boon to the struggling economy of Washington County, Maine (Levine, 1998). Again, permits were approved for the culture and outplanting of this non-native species due to the belief that photoperiod and water temperatures would not allow sexual reproduction or permanent establishment of the species (Watson et al., 1998). Attempts to successfully farm the commercial Japanese cultivars failed due to nutrient limitations and a lack of understanding of the seasonality *of P. yezoensis* in New England. The gametophytic blade phase of this species only appears in the late winter and early spring months in New England, but Coastal Plantations International attempted to grow the blades during the summer months believing temperatures and light levels were superior at that time (Yarish, personal communication). Their initial failures did not end the attempts to grow nori in Maine.

Because *Porphyra* species have been determined to be highly efficient in the uptake of nutrients commonly found in eutrophic waters, it was proposed that they could be used as bioremediators in areas of established fish farms (Chopin and Yarish, 1998). Uptake and growth observations using native *Porphyra* species *(P. purpurea* (Roth) C. Agardh and *P. umbilicalis)* and non-native *P. yezoensis* were made in natural habitats, on

nets, and in integrated aquacultural systems. Comparisons were made between ambient and tissue P and N levels. *Porphyra* species reduced P and N to non-harmful levels, and it was estimated that between 22 and 27 nori nets would be needed per ton of finfish produced per year to offset the P and N (Chopin and Yarish, 1999), with *P. yezoensis* and *P. purpurea* deemed the best bioremediators tested. Further assessment of the bioremediation of other native species, including *P. leucosticta, P. amplissima, P. linearis,* was proposed along with a cultivar improvement program (Yarish et al., 1999).

While, to date, commercial scale nori farming has been largely unsuccessful in North America, *Porphyra yezoensis* has become established in regions of the continent's coast. Extensive field collections, herbarium specimens, and molecular evaluations have confirmed the occurrence and the distribution of two distinct *P. yezoensis* genotypes in the northwestern Atlantic (Bray, 2006; Mathieson et al., 2008; Neefus et al., 2008). One of the two genotypes has an ITS-1 sequence identical to a GenBank sequence from a specimen of P. *yezoensis* f. *yezoensis* that was collected from the wild near Nanaehama, Hakodate, Hokkaido Japan (Neefus et al., 2008). The distribution of this forma extends from Maine to New York. The ITS-1 sequence of the second forma is identical to more than a dozen recently developed commercial cultivars of *P. yezoensis* f. *narawaensis;* Its distribution is more limited and extends from Hammonassett State Park near Madison, Connecticut, in the west, to Westport, Massachusetts, in the east (Figure 2). While f. *narawaensis* occurred within the range off. *yezoensis,* Bray (2006) reported that at sites where f. narawaensis occurred, f. yezoensis was absent. Such patterns suggested that competition favored the commercial cultivar (Bray, 2006, Neefus et al. 2008).

The modes of introduction of alien species have been studied extensively due to the sometimes devastating ecological and economic effects recognized since the early 1900s (Ostenfeld, 1908; Elton 1958; Carlton 1999). It has been determined that no region of the world is without established alien marine species (Carlton, 1979), including 260 alien marine macroalgal species (Hewitt et al., 2007). Because introduced macroalgal species are not easily eradicated or controlled once established, much effort has been directed at identifying vectors of transport and release. The modes of transport and inoculation of alien marine species into new regions have been detailed by many and include wooden-hull boring; fouling of and subsequent transport of fishing nets, relocation of oil rigs, and untreated metal ship hulls, recreational boat hulls; dry ballast (intertidal rocks and sand); ballast water uptake and release; attachment to sea chests or propellers, intentional transfer of maricultural organisms (including shellfish, finfish, and seaweeds); accidental transfer of organisms associated with maricultural organisms; improper disposal of live, frozen, or dried seafood; accidental release from aquaculture; improper release of aquarium stock; and the improper disposal of seaweeds used as packing material for live bait (Elton 1958, Carlton 1996, Weigle et al., 2005).

Because *Porphyra yezoensis* is a resilient organism with a complex life history that includes sexual and multiple forms of asexual reproduction, it could be transported from its point of origin to new regions by most of the above mentioned modes. Hewitt et al. (2007) delineated the likelihood of encounter and the survival constraints associated with the common modes of alien transport, which included ease of uptake in ballast water, association with a target species (oysters) or habitat (subtidal conchocelis), ability to survive the shear stresses of transport on the exterior of a vessel, survival of

desiccation, darkness, crushing stress, and exposure to climate change. The conchocelis stage of P. *yezoensis* is likely to encounter and survive the uptake transport modes of most dispersal vectors.

Though one might point to nori-culture in America as the source for the establishment of *Porphyra yezoensis* f. *narawaensis* in the northwestern Atlantic, evidence suggests that it is not the vector to blame. Although f. *narawaensis* cultivars were imported and out-planted into the waters of Cobscook Bay in Downeast Maine, this genotype has not subsequently been discovered north of Long Island Sound (Bray, 2008). Because the water temperature and light regimes are markedly different north and south of the Cape Cod, and because hydrogeographic mixing and boat traffic north to south across this barrier are minimal, Coastal Plantation's nori farms in Maine are not the likely source off. *narawaensis* populations in New England (Neefus et al. 2008). Also, because Coastal Plantations International did not attempt to cultivate f. *yezoensis* and because there are herbarium specimens of f. yezoensis from New England that pre-date the CPI's operation, they cannot be its source of introduction in the region (Bray, 2006).

Although *Porphyra yezoensis* has the potential to be distributed to new regions by many of the known transport vectors. Both forms of P. *yezoensis* were likely transported to New England as shell boring conchocelis associated with organisms imported for use in mariculture (Bray 2006). Clokie and Boney (1980) found a close association between conchocelis infected shells in the subtidal zone and high density of *Porphyra* blades in the intertidal zone in the Firth of Clyde, Scotland. The earliest voucher specimens of P. yezoensis f. yezoensis were collected in the region in the 1960s (Bray, 2006), around the time another Asian algal species *Codium fragile* ssp *fragile* (Suringar) Hariot was

introduced with oysters brought from Peconic Bay, Long Island, NY (Galstoff, 1962). It is believed that f. *narawaensis* was introduced to New England, in the 1980s, at about the time the cultivars were developed for widespread use in Japan (Neefus et al. 2008).

The establishment of the two forms of *Porphyra yezoensis* in the northwestern Atlantic is significant in that only a small percentage of macroalgae are ever found beyond their points of origin (260 of thousands). In Williamson and Fitter's (1996) treatise on invaders, they proposed that only one in ten species are ever introduced to new regions via anthropomorphic transport vectors. Of these introduced species, only one in ten survive the transportation and the new environment for long enough to become established in the new region. Once established, one in ten of these introduced aliens becomes invasive (destructive environmentally and/or economically) in the new region. As *P. yezoensis* f. *yezoensis* and *P. yezoensis* f. *narawaensis* have been successfully transported and established in New England, according to Williamson and Fitter's estimation, the species has a ten percent probability of becoming invasive in this new region.

To predict future invading organisms Nyberg and Wallentinus (2005) delineated the traits common to successful invaders and produced of a list of thirteen characteristics indicative of invasive potential. The traits were based on the summaries given by Boudeouresque and Verlaque (2002), Ribera Siguan (2002), and Wallentinus (2002). and included: current geographical distribution (organisms found in roughly half of the world's regions were more likely to be invaders than those found in few or nearly all regions); probability of being transported; survival time out of water; salinity survival range; temperature survival range; tolerance to pollutants; reproductive flexibility;

growth strategy (stress tolerant, competitive, ruderal) including surface area to volume; defense mechanisms against grazing and infestation; thallus size (larger organisms being more likely to negatively impact new environments); morphology (crust and mat forming increases negative impact); and life span.

Using these criteria, Nyberg and Wallentinus (2005) evaluated 113 algal species introduced to Europe and an equal number of equivalent native European taxa. Species were awarded scores between 0 and 1 for each criterion, with the overall score being the average across the thirteen traits. The authors deemed the results of the evaluation reliable in that fifteen of the twenty-six invasive species were listed in the twenty highest ranked taxa. In this study, *Porphyra yezoensis* ranked 16th among the 77 red algal species evaluated. Although this ranking was high compared to other red algal species, *P. yezoensis* was not considered to have the potential of being highly invasive.

While hypothetical species-trait risk assessments can be useful for determining an organism's overall invasiveness potential, it has been common to find that a species invasive to one region is not invasive in another. An example of this phenomenon is *Codium fragile* ssp. *fragile* which has had a significant negative impact on the western Atlantic coast, while at the same time has had a minimal effect on the east Atlantic Ocean (Chapman, 1999; Schaffelke and Hewitt, 2007), though both regions are abiotically similar. Disturbance in the receiving community (through nutrient, substrata, or water temperature disruption, macroalgal removal through grazing or disease, and ecosystem "meltdown" caused by high levels of other invaders) has been the key to nearly all the successful macroalgal invasions in which the inoculation mechanism is known (Valentine et al., 2007), with a notable exception being the invasion of the Mediterranean by

Caulerpa taxifolia (M. Vahl) C. Agardh. *Sargassum muticum* (Yendo) Fensholt and C. *fragile* ssp. *fragile* both require a disturbance of native canopy-forming algae in order to become established (Johnson, 2007). The same was found to be true for *Undaria pinnatifida* (Harvey) Suringar in Tasmania (Johnson, 2007). These introduced species have also been agents of habitat modification in disturbed areas, whereas they have remained background species at other undisturbed sites. Resistance to invasion has been highest in regions with extensive seagrass or macroalgal cover (Cecchereli and Cinelli, 1999). Therefore alien macroalgal species do not typically outcompete native species unless the growth of native assemblages is limited by disturbance.

The fact that alien macroalgal species require environmental disturbance to become invasive may explain why some species may become invasive in a particular location and not at another. Initial short term studies of the effect of introduced *Caulerpa taxifolia* on the density of the seagrass *Cymodocea nodosa* (Ucria) Ascherson in the Mediterranean pointed to a reduction of *Cymodocea* shoots, whereas long term studies demonstrated that the two organisms coexisted without future shifts in the competitive balance (Ceccherelli and Cinelli, 1997). Harris and Tyrell's (2001) twenty five year study of the northwestern Atlantic demonstrated a shift in abundance from kelp to a *C. fragile* and red algal dominated assemblage. The same ecosystems have reverted in recent years with kelp abundances increasing and *Codium* levels decreasing to the point where it may no longer be damaging particular communities (Harris, personal communication). Although few studies have examined sites prior to, or in the early stages of, invasion (Schaffelke and Hewitt, 2007) it has been observed that aliens often

remain background species with little impact for some time before expanding to the point of becoming invasive (Stockwell et al., 2003).

Although there are no published reports of *Porphyra yezoensis* f. *yezoensis* or *P. yezoensis* f. *narawaensis* becoming invasive following introduction into new regions, evidence from their home range suggests that the commercial cultivars of f. *narawaensis* have the potential to cause ecological damage. The cultivars were developed to grow rapidly, efficiently absorb nutrients, and to proliferate through the production of neutral spores (Miura 1984). While these qualities have been highly beneficial to the nori industry, they have had some negative consequences on the Japanese coast. In areas of heavy nori-culture, f. *narawaensis* has migrated from the coastal bay nets, on which it was seeded, to the open coast where it has become firmly established. The cultivar has subsequently displaced and even caused the extinction of other native Japanese macroalgal species (Miura and Aruga 1987).

The impact *of Porphyra yezoensis* f. *yezoensis* and *P. yezoensis* f. *narawaensis* in the northwestern Atlantic has been unclear, but their presence had been noted with concern. The Connecticut Aquatic Nuisance Species Working Group (2005) included *P. yezoensis* as a Management Class 4 species, which means 'it is established in the waters of Connecticut and may have the potential to cause impacts, but current knowledge is insufficient to determine if control actions are warranted.' The management actions for such organisms include the prevention of further introduction, the interruption of the export pathways from Connecticut, further research to evaluate invasive potential and ecosystem impact, and continued monitoring of existing populations to determine rates of spread.

In accordance with these management recommendations, the current study set out to look for changes in the distribution of *Porphyra yezoensis* f. *yezoensis* and *P. yezoensis* f. *narawaensis* throughout New England and to monitor existing southern New England populations of f. *yezoensis* and f. *narawaensis* during the growing season, through monthly measurements of density and biomass at sites where the organism was previously collected (Bray 2006). To determine the possible ecological impacts of f. *yezoensis* and f *narawaensis,* density and biomass measurements were also taken for all macroalgal taxa growing in close proximity to either form. Because both forms of P. *yezoensis* often grow epiphytically on long-lived fucoid algae (Miura 1988; Bray 2006), this study also attempted to determine the impact of P. *yezoensis* on host organisms. Therefore, stature measurements of host organisms were compared to those of non-host organisms of the same species found in the same locations.

While previous studies had done much to define the range, seasonality, and population locations *of Porphyra yezoensis* f. *yezoensis* and *P. yezoensis* f. *narawaensis* (Bray, 2006; Mathieson et al., 2008) in the northwest Atlantic, little effort had been given to ecological quantification. It is hoped that the current study will provide valuable baseline data for further comparisons and long term monitoring of this introduced species.

Through the course of this study a *Porphyra* species of cryptic origins, *Porphyra* spp. *'stamfordensis'* (Bray 2006), was also detected at several sites. Because *P.* spp. *'stamfordensis'* may have been introduced this decade, and it was discovered in high density and biomass at several sites, special attention has been given to its collection data.

Material and Methods

Rapid Assessment Survey

In the winter of 2006-2007, sites from Lubec, ME to western Connecticut were surveyed for *Porphyra yezoensis f yezoensis* and *P. yezoensis* f *narawaensis* using rapid assessment techniques. Likely population locations (channels, breachways, narrow tidal rapids, boat ramps, marinas, etc.) were visited briefly and visually scanned for *Porphyra* species. Possible *P. yezoensis* blades were collected for molecular analysis. The sites thus examined were Rocky Neck State Park, Niantic, CT; Niantic Boat Valet, Niantic, CT; Black Point, Narragansett, RI; Village Inn Beach, Narragansett, RI; Mackerel Cove, RI, the Westport Boat Ramp, Westport, MA; Pope's Island Marina, Fairhaven, MA; Buzzard's Bait Bridge, Wareham, MA; Victory Rd Park, S. Boston, MA; Morrissey Boat Ramp, S. Boston, MA; Carson Beach, S. Boston, MA; Lead Hazard Bridge, Marblehead, MA; Marblehead Neck, MA; Salem Willows, MA; Goose Cove, Gloucester, MA; Dover Point, Newington, NH; Seapoint, Kittery,ME; Leeman Hwy, Brunswick, ME; Great Island, Harpswell, ME; Orr Island, Harpswell, ME; Cundy's Harbor, ME; Machiasport, ME; Cutler, ME; Pikeland, Lubec, ME; Lubec Town Dock, Lubec, ME (Figure 10). For comparative purposes, it was decided that only sites in and surrounding the known distribution of f. *narawaensis* would be further examined in this study.

Field Procedures for Quantification

During the winter-spring growing season (December 2007 through May 2008), *Porphyra yezoensis* sites, documented by Bray (2006), were monitored monthly (Figure 2, Table 1). Once *P. yezoensis* blades appeared (initial month of appearance varied by

location), a twenty meter transect line was established and twenty sample quadrats (0.5 m x 0.5 m) were established along this line in the low intertidal zone at each site. The quadrats were used to determine occurrence and density of all macroalgal taxa. An attempt was made to establish an equal number of quadrats in areas containing *P. yezoensis* and areas devoid of P. *yezoensis,* but similar in substrata, wave exposure, temperature, salinity, slope, currents, and nutrients. However, after close inspection of collected materials in the laboratory, it was determined that most "non-P. *yezoensis'"* quadrats contained some small epiphytic *P. yezoensis* blades that were undetectable in the field. Quadrats were digitally photographed for percent cover calculations, but due to the small size of most *P. yezoensis* blades, this technique was ineffective.

A destructive macroalgal sample $(0.1 \text{ m} \times 0.1 \text{ m})$ was collected from a random location from within each of the larger (0.5 m x 0.5 m) quadrats during each month. To randomly select the destructive sampling area, the large quadrats were divided into 25 sectors (10 cm x 10cm) being five sectors across by five sectors down. Prior to sampling, a ten-sided die was rolled twice to determine the coordinates of the sample. The first roll determined the across value, and the second determined the down value. With rolls of 6 or above (the zero reading equaling ten), the proper sectors were determined by subtracting 5 from the rolled value. Therefore, a roll of 8 was actually a coordinate of 3. Once the coordinates were determined, a paint scraper was used to remove the algae from the substrata. Each destructive sample was placed in its own labeled plastic bag and transported, untreated, to a processing lab in the Spaulding Life Science building at the University of New Hampshire.

Collection processing

In the lab, the contents of each bag from destructive sampling were examined for *Porphyra* specimens. A subsample of *Porphyra* was removed, floated in seawater, and pressed on labeled herbarium sheets. Using a razor, 2 cm x 2 cm sections were cut from a selection of blades to use for molecular identifications. Each piece removed was placed in its own labeled 1.7 ml tube, along with silica beads. The remainder of the destructive sample in each bag was frozen at -20°C freezer for between 1 to 4 weeks before further processing. Upon removal from the freezer each destructive sample was placed in an aquarium net and rinsed in warm water to thaw and remove sediments. The samples were then floated in tap-water in a 28 x 43 cm pan. Species were sorted and counted on dry trays. Identifications of macroalgae were made based upon macroscopic and microscopic characters using keys to the marine algae of the northwestern Atlantic (Sears, 2002; Bohnsack-Villalard, 1995). Once counted, the individuals of each taxon where clumped together, squeezed until damp dry, and fresh weight (FW) was determined to the nearest hundredth of a gram (Mettler Toledo PR503 Delta Range). Biomass of each taxon (g FW/m²) and density counts (individuals/ m²) were estimated by multiplying the measured values by 100. Voucher specimens of each taxon from each collection were pressed and will be deposited in the Albion R. Hodgdon Herbarium (NHA) at the University of New Hampshire.

Prior to weighing, fucoid algae were segregated into those with and without epiphytic loads of *Porphyra yezoensis.* The individual lengths of all intact fucoid algae from both groups were measured from the holdfast to the tip. Likewise, the lengths of a

representative subset of *Porphyra* were measured from each destructive sample during the months of greatest luxuriance (April and May).

Molecular Methods

The *Porphyra* samples that were dried for molecular analysis were ground in labeled 1.7 ml microcentifuge tubes using disposable plastic pestles, a few grains of molecular grade sand, and 300 ml of Gentra Puregene® Cell Lysis Solution (D-5002). The DNA was extracted with a Gentra Puregene ® Isolation Kit as per the manufacturer's instructions. Samples were incubated in a 65°C heatblock for one hour inverting 10 times at 30 minutes and cooled to room temperature before 100 ul of Protein Precipitation Solution (Gentra D-5003) was added. Samples were inverted 150 times and chilled at -20°C for 45 minutes before they were centrifuged for 15 minutes at 13,000 rpm. The supernatant was then poured into at new 1.7 ml microcentifuge tube containing 300 ul of 100% isopropanol and inverted 50 times before centrifugation for 10 minutes at 13,000 rpm. The alcohol was decanted and replaced with 300 ul of 70% ethanol before inversion and 5 minutes of centrifugation at 13,000 rpm. The alcohol was decanted, and the sample was air dried for 60 minutes before 50 μ l of DNA Hydration Solution (Gentra D-5004) was added. After briefly mixing, the samples were incubated in a 65°C heatblock for one hour and centrifuged for 5 minutes.

Polymerase chain reactions were carried out in 50 µl volumes containing 4 µl extracted DNA, 10 µl Taq buffer (Promega GoTaq® Flexi Green), (0.2 mM) Mg^{2+} , 1 µl dNTPs, 1 ul each (20 mM) primer, and 0.25 ul Taq polymerase (GoTaq® Flexi). The segment of DNA amplified was 1481 bp in length extending from position 67 of *rbcL*

through the *rbcL-spcS* intergenic spacer to the beginning of the small subunit. The evaluation was done using the F67 and rbc-spc primers (Teasdale et al., 2000).

The PCR products were separated by electrophoresis on a Cyber-Safe® treated low-melt agarose gel (0.8%) in nTBE Buffer (0.5x). On a UV lightbox, the desired DNA bands were excised using microscope slide covers and transferred to 1.7 ml tubes, incubated at in a 65°C heatblock for five minutes, and then transferred to 37°C heatblock. To each tube, 1.5 *\xl* of agarase (Sigma A6303, 50 units/ml) were added, and the mixture was incubated overnight.

Concentrations of DNA were quantified using an Invitrogen™ Quant-iT™ dsDNA BR Assay Kit (Q32851) and an Invitrogen™ Qubit™ fluorometer (Q32857) as per the manufacturer's instructions, and appropriate volumes of DNA and primers were sent to Hubbard Genomic Center (UNH) for clean-up and sequencing reactions using Applied Biosystems BigDye Terminator Cycle Sequencing Kits (vl.l and v3.1). The DNA samples were resolved by capillary electrophoresis on an ABI3130 DNA Analyzer.

Resulting sequences were trimmed in Chromas (version 2.2, Technelysium, Pty. Ltd., Tewantin, Queensland, Australia). Sequence assembly, alignments were made and proofed using Seq Man II (version 7.1 for Windows, DNAStar, Inc., Madison, Wisconsin). Comparative alignments and GenBank searches were performed using MegAlign (version 7.1 for Windows, DNAStar, Inc., Madison, Wisconsin).

Site Descriptions

Seven study sites visited monthly from December 2007 through May 2008:(1) Lighthouse Point, New Haven, CT; (2) Guilford Marina, Guilford, CT; (3) Rocky Neck

State Park, Old Lyme, CT; (4) Charlestown Breachway, Charlestown, RI, (5) Black Point, Narragansett; (6) Westport, MA; and (7) Falmouth Heights, MA (Table 1, Figures 2-9). Transect/Quadrat sampling was conducted monthly beginning at each site with the initial appearance of *Porphyra yezoensis* blades.

Light House Point, New Haven, CT (Figure 3), also known as Morris Point from colonial times and Five Mile Point (due to the fact that it is located five miles from the center of New Haven), marks the eastern end of New Haven Harbor. Its tidal amplitudes range from lows of-1 ft to highs of 7.6 ft above Mean Low Water (MLW). Because the location is moderately exposed, it experiences low to moderate wave action. Its granitic boulder and sandy substrata support the growth of fucoid algae, *Chondrus crispus* Stackhouse, *Ulva* spp., and multiple *Porphyra* species. The rocky point lies beside a wide sand beach designated for public swimming and sunbathing ([http://www.cityofnewhaven.com/Parks/ParksInformation/lighthousepoint.asp\).](http://www.cityofnewhaven.com/Parks/ParksInformation/lighthousepoint.asp) As an indication of the level of pollution in the harbor, the New Haven Board of Health frequently monitors the area waters in the summer for unhealthy levels of bacteria, and resulting beach closures are not uncommon (East Shore Ranger, Terry McCool, personal communication).

The Guilford Marina, Guilford, CT (Figure 4), site is located in shallow Guilford Harbor sheltered by Faulkner's Island. Its tidal amplitudes range from lows of-.8 ft to highs of 5.6 ft. The Marina was designated a Connecticut Clean Marina in 2007 by the Department of Environmental protection for its efforts to control pollutants from fuels and litter along with efforts to properly clean boat hulls

[\(http://www.ct.gov/Dep/cwp/view.asp?A=2712&Q=329898\).](http://www.ct.gov/Dep/cwp/view.asp?A=2712&Q=329898) The Marina is home to

slips and moorings for upwards of thirty residential and recreational boats. The study site is located on a wide, manmade retaining wall comprised of granitic boulders located at the mouth of the marina. The boulders predominantly support the growth of fucoid algae and associated epiphytes. Swans frequent the study site.

The Rocky Neck State Park, Niantic, CT site (Figure 5) is located on an exposed point on the western edge of the park. Tidal amplitudes range from lows of-0.5 ft to highs of 3.6 ft. The granitic and basaltic bedrock substrata support the growth of barnacles, fucoid and ulvoid algae, *Chondrus crispus,* and multiple *Porphyra* species. Wave action at this site can be heavy with an apparently strong current running away from the point. For example, a sample bag accidentally dropped into the water, was immediately carried straight away from shore and was out of sight in minutes. Ducks and geese frequent the study site.

The Charlestown Breachway, site in Charlestown, RI (Figure 6) is located along the inside of a manmade jetty channel that was constructed in the middle of a miles-wide stretch of sand beach on the southwestern coast of Rhode Island. The breachway was originally a natural feature of the coastline that connected the Atlantic Ocean to the Pawaget, Ninigret, and Charlestown Ponds. Because nature's breachway was sandy and tended to fill in with sand and other sediments, the people of the Charlestown region, during the late 1800s and early 1900s, pushed for the construction of a permanent breachway and jetties composed of 400 pound field stones stacked as retaining walls. The labor required to build the breachway was extensive and used horses, railways, and rail carts. The construction was done in hopes of preserving the common practice of cultivating and harvesting oysters in the ponds

[\(http://www.riparks.com/charlestownhistory.htm\)](http://www.riparks.com/charlestownhistory.htm). The large fieldstone walls currently support the growth of fucoid and ulvoid algae, along with multiple species of *Porphyra*. The breachway is heavily used for saltwater fishing in spring and summer, and it serves as the point of ocean access for the Ocean House Marina that has slips and dry dock storage space for more than fifty recreational boats. Tidal amplitudes at this site range from lows of-.05 ft to highs of 3.7 ft.

The Black Point, Narragansett, RI site (Figure 7) is highly exposed and wave action is extreme due to an abrupt granitic bedrock ledge. Due to the pounding of the waves, fucoid algae are nearly absent from this site, and ulvoid algae and *Porphyra* species are found growing attached to blue mussels *(Mytilus edulis* L.), barnacles *(Semibalonus balanoides* L.), and, in low areas, *Chondrus crispus. Scytocyphon lomentaria* (Lyngbye) J. Agardh is also abundant at this site. The tidal amplitudes at Black Point range from lows of-0.5 ft to highs of 4.6 ft. While boat traffic close to this site is unlikely, the Block Island ferry terminal lies within miles.

The Westport, MA site (Figure 8) lies in a completely sheltered estuarine environment at the western edge of Buzzards Bay. The site is near the confluence of the eastern and western branches of the Westport River. Collections were made along a transect line placed on a short manmade jetty comprised of large field stones that support the growth *ofFucus vesiculosis* L., *Ascophyllum nodosum* (L.) Le Jolis, and ulvoid algae along with a few *Porphyra* species. The jetty lies within twenty yards of a seasonally operated seafood restaurant on the west and an oft used public boat landing on the east. Multiple marinas and marine businesses lie within a mile of this location. Tidal

amplitudes at this site range from -0.3 ft lows to 4.3 ft highs. Currents along the tip of the jetty can be dangerously strong at points in the tidal cycle.

The Falmouth Heights, MA study site on Cape Cod (Figure 9) lies in a short (50 m), narrow (15 m) manmade fieldstone-walled channel that drains from Little Pond into Vineyard Sound at low tide. The boulders and sandy substrata support the growth of fucoid and ulvoid algae, *Chondrus crispus,* and several *Porphyra* species. Currents in the channel are moderate. Wave action is minimal. East and west of the channel lie miles of sandy beach with heavy public use in summer months. Tidal amplitudes range from lows of-0.2 ft to highs of 1.7 ft.

Several other sites where *Porphyra yezoensis* had been previously reported by Bray (2006) were visited monthly for collecting, but were not used for macroalgal community quantification in most cases, because significant *P. yezoensis* populations never appeared. These sites were from west to east: Hammonasset State Park, Madison, CT; Niantic Boat Valet, Niantic, CT; and Fort Rodman, New Bedford, MA. These sites were not examined as thoroughly as the main study sites for a few reasons. Due to profound lack of *Porphyra* of any kind, the New Bedford, MA site was omitted. Although the Fort Taber site (also known as Fort Rodman) in New Bedford, MA was historically reported to support populations of P. *yezoensis,* no such populations were found in this study. Only five blades, of other *Porphyra* species, were found in rapid surveys from January through March. The Niantic, CT boat valet site was not selected for this quantification study because of the high densities of the morphologically similar *Porphyra* spp. *'stamfordensis'* and seemingly low densities of P. *yezoensis* f *yezoensis.*

The site was left out because visually separating the two species was difficult if not impossible without reproductive markings, which neither species displayed regularly.

The Hammonassett State Park site was not used for quantification because the *Porphyra yezoensis* populations were located in a precarious position far out on a jetty surrounded by deep water.

Results

Rapid Assessment Survey

Table 2 summarizes all of the *Porphyra* species collected through the winter 2006-2007 rapid assessment survey. *Porphyra yezoensis* f. *yezoensis* was confirmed at more sites, 9 of 25, than was *P. yezoensis* f. *narawaensis,* 4 of 25 (Figure 10). *Porphyra* ssp. *'stamfordensis'* was not confirmed at any of the 25 survey sites.

During the survey, *Porphyra yezoensis* f. *yezoensis* was collected at two sites outside of its previously published distribution. In May of 2007, voucher specimens of f. *yezoensis* were collected at the town dock in Lubec, ME, more than 60 miles north of the distributional limits reported by Bray (2006). In April of 2007, voucher specimens off. *yezoensis* were collected at the Niantic Boat Valet, Niantic, CT. This is the only known population of P. *yezoensis* f. *yezoensis* that exists within the distributional range of P. *yezoensis* f. *narawaensis* in the Northwest Atlantic.

Quantification Study

Table 3 summarizes all of the macroalgal species obtained through destructive quadrat sampling along the line transects of each of the seven study sites. The greatest number of taxa (fourteen) was recorded for the New Haven, Rocky Neck, Charlestown, and Black Point sites. Each of these locations had seasonal populations of *Porphyra yezoensis* throughout the study period (February-May), but none supported populations of both *P. yezoensis* f. *yezoensis* and *P. yezoensis* f. *narawaensis.*

Two sites with fewer number of taxa, Guilford (8 species), and Falmouth Heights (10 species), had seasonal populations of *Porphyra* spp. *'stamfordensis'* throughout the

study period. No *P. yezoensis* specimens were collected from Guilford, and very few from Falmouth Heights (only in March and April). The collections of P. *yezoensis* f. *narawaensis* from Falmouth Heights represent the first records east of Westport, MA for this genotype.

Of the macroalgal species listed for each site, those commonly found growing epiphytically on *Fucus* included: *P. yezoensis* f. *narawaensis, P. yezoensis* f. *yezoensis, P. leucosticta* Thuret, *P. olivii* Orfanidis, Neefus & Bray, *P.* spp. *'stamfordensis', Ulva intestinalis* L., *Polysiphonia stricta* (Dillwyn) Greville, *Elachista fucicola* (Velley) J. E. Areschoug, *Ulothrix flacca* (Dillwyn) Thuret and *Pylaiella littoralis* (L.) Kjellman.

The monthly density values (individuals/ $m²$) for each species collected on the twenty quadrats from each site are summarized in Appendix A. Counts of minute epiphytic species *(Elachista fucicola, Pylaiella littoralis, Blidingia minima* (Nagelli ex Kiitzing) Kylin, *Bangia fuscopurpurea* (Dillwyn) Lyngbye, *Ulothrix flacca)* were not included in the species density enumerations, as it was difficult to accurately document in a reasonable amount of time. *Porphyra* species and *Ulva intestinalis,* growing epiphytically or epibiotically (on *Mytilus edulis* and *Semibalanus balanoides* at Black Point) in dense clusters, typically had the highest counts per month at each site.

The monthly biomass data (g $FW/m²$) for each species on the twenty quadrats from each site were also summarized (Appendix B). Unlike the density data, all species were included in the biomass recordings. For each site, excluding Black Point, biomass yields were highest for fucoid algal species *[Ascophyllum nodosum* (Westport), *Fucus spiralis* (Falmouth Heights), *Fucus vesiculosis* (all other sites)]. Because only one frond

of *Fucus vesiculosis* was collected at Black Point during the study, *P. yezoensis* f. *narawaensis* supplied the bulk of that site's biomass.

Seasonality and abundance in both *Porphyra yezoensis* forms were estimated using monthly means from each site (Figures 11 and 12). *Porphyra yezoensis* f. *yezoensis* exhibited an earlier peak density period (February) at New Haven than the f. *narawaensis* populations at Rocky Neck, Charlestown Breachway, Black Point, and Falmouth Heights (March to April). The mean population density of P. *yezoensis* at peak periods was more than twice as high for the f. *yezoensis* at New Haven (10,150 blades per m^2 + 1796.6 SE) than for f. *narawaensis* at Black Point (5003 blades per m^2 + 1119.6 SE). The Falmouth Heights site contained a population of f. *narawaensis* that had a low density (15 blades per $m^2 \pm 10.9$ and 15 SE) during April and May.

Figures 13 and 14 summarize mean monthly biomass at each site during February to May. *Porphyra yezoensis* f. *yezoensis* biomass peaked earlier (February) at New Haven compared with *P. yezoensis* f. *narawaensis* at the other sites (March at Rocky Neck and Charlestown Breachway, April at Black Point and Falmouth Heights). The peak biomass was more than twice as great for the Charlestown Breachway *P. yezoensis* f. *narawaensis* populations (511.3 g/ m^2 + 441.7 SE) compared with *P. yezoensis* f. *yezoensis* populations from New Haven (237.3 g/ $m^2 \pm 40.2$ SE). *Porphyra yezoensis* f. *narawaensis* populations at Falmouth Heights had a mean biomass of only 7.15 g/ m^2 + 5.4 SE at peak in April.

Because significant populations of *Porphyra* spp. *'stamfordensis'* occurred at several study sites, its seasonality and abundance was also estimated (Figure 15) and mean monthly values were enumerated for each site. The taxon was collected on
transects from February through May at four sites. Peak population biomass yields occurred in February at Charlestown (5.15 $g/m^2 \pm 2.57$ SE), while maximum values occurred in March at Guilford (147.3 g/ $m^2 \pm 59.4$ SE), Falmouth Heights (108 g/ $m^2 \pm$ 20.26 SE) and Wesport (262.8 g/m² \pm 61.5 SE).

The maximum biomass contribution of the dominant *Porphyra* species to the total macroalgal community biomass is summarized in Table 4. Due to the absence of large fucoid algae at Black Point, *Porphyra yezoensis* f. *narawaensis* biomass contribution was substantially higher (81%) than at all other sites: New Haven—1%, Rocky Neck—2%, Charlestown—2%. *Porphyra* spp. *'stamfordensis'* was a small contributor of biomass to all of the communities it occupied (Guilford—1%, Westport—1%, Falmouth Heights— 2%).

To evaluate the impact of epiphytic *Porphyra* loads on long lived fucoid algae, the percentage of epiphytized plants were calculated during March and April (Figure 16). Both the highest and lowest values were recorded for *Porphyra* spp. *'stamfordensis'* populations at Falmouth Heights (67%) and Guilford (16%), respectively. The values for *P. yezoensis* f. *yezoensis* at New Haven was greater (48%) than those found at both sites occupied by *P. yezoensis* f. *narawaensis*—i.e. Rocky Neck (36%) and Charlestown (34%).

The mean frond lengths for *Fucus* with and without epiphytic *Porphyra yezoensis* were enumerated during March and April (Figure 17). ANOVA revealed that epiphytized *Fucus* plants were longer than those without *Porphyra* loads (P value <0.01). The difference was most clearly demonstrated at New Haven, Charlestown, and Westport (P values each ≤ 0.01).

Table 5 summarizes mean frond length of *Porphyra yezoensis* f. *yezoensis* and *P. yezoensis* f. *narawaensis* during March and April (i.e. peak values). The mean blade lengths were \leq 5 cm for each site. The longest mean blade length (4.38 cm \pm 1.5 cm) was recorded for the *P. yezoensis* f. *narawaensis* at Black Point, and the shortest mean blade length (2.40 cm + 0.91 cm) was recorded for the f. *yezoensis* at New Haven. The range of individual blade lengths for *P. yezoensis* f. *yezoensis* varied from < 0.5 cm to 9 cm, while *P. yezoensis* f. *narawaensis* ranged from < 0.5 cm to 10 cm.

The mean frond lengths for all other *Porphyra* species at each site were also determined during peak periods of March and April (Table 6). Again the mean blade lengths for each species at different sites were all less than 6 cm. The longest mean blade length (5.33 cm \pm 4.72 cm) was recorded for *Porphyra leucosticta* from the Charlestown Breachway, while the shortest values $(2.54 \text{ cm} + 1.63 \text{ cm})$ were recorded for *P. olivii* at New Haven. Mean blade lengths for *Porphyra* spp. *'stamfordensis'* ranged from 3.51 cm \pm 3.00 cm at Guilford to 4.42 cm \pm 3.10 cm at Westport.

Discussion

As *Porphyra yezoensis* is an introduced species in the northwestern Atlantic, coastal managers have been wary of its potential negative impact on native macroalgal communities (Anonymous, 2005). To assess the extent its introduction, a coastal survey was conducted by Bray (2006) to determine the distribution both f. yezoensis and f. *narawaensis.*

The survey methods employed were designed to rapidly assess presence and absence of f. *yezoensis* and f. *narawaensis* in channels, breachways, narrow tidal rapids, boat ramps, marinas, etc. along the New England Coast. These methods revealed the presence of dozens of P. *yezoensis* populations from New York to Downeast Maine (Bray 2006).

During the winter/spring of 2007,1 conducted another rapid assessment survey of *Porphyra* species along the New England Coast (Table 2; Figure 10). This study revealed the presence of P. *yezoensis* f. *yezoensis* at 9 of 25 sites, including two collections (Lubec, ME and Niantic, CT) off. *yezoensis* outside of the distribution reported by Bray (2006).

The voucher specimens of f. *yezoensis* from Lubec, ME are the first collections of this form north of Bar Harbor, ME. This marks a range expansion of 60 miles. Although this area was not surveyed by Bray (2006), it was extensively monitored for escapes in the winter and spring of 1998 and 1999 following the region's nori culture attempts (Watson et al., 2000), and no form of P. *yezoensis* was found. Thus, it is likely that P. *yezoensis* f. *yezoensis* has expanded to this region within the last decade.

The voucher collections off. *yezoensis* from Niantic, CT mark the first discovery of a population off. *yezoensis* within the distribution off. *narawaensis* in New England. Because the distribution of f. *yezoensis* in the Northwest Atlantic is interrupted by f. *narawaensis,* which arrived later, it has been proposed that the distribution of f. *yezoensis* was once continuous in the region (Bray, 2006; Neefus, personal communication). The f. *yezoensis* in Niantic, CT is either a holdover population that has been long established in the region, or it has recently arrived. Because there are no collection records from the Niantic Boat Valet site prior to 2007, one can only speculate as to the history of the population at this location. Because of the sheltered nature of this site, I suspect this is a holdover population still residing in this location. Because competitive exclusion favoring f. *narawaensis* has been suggested (Bray, 2006), it is likely that the Niantic, CT site has never been successfully inoculated with f. *narawaensis.*

To quantify the level of establishment of *Porphyra yezoensis* f. *yezoensis, P. yezoensis* f. *narawaensis,* and *P.* spp. *'stamfordensis'* in New England, and to assess the macroalgal communities they occupy, I conducted monthly biomass and density assessments of all macroalgal taxa, growing within transects, at the established study sites from New Haven, CT to Falmouth Heights, MA, during the season of maximum blade growth in New England (February-May).

The study intended to measure f. *yezoensis* populations at four sites (New Haven, CT; Guilford, CT; New Bedford, MA; and Falmouth Heights) and f. *narawaensis* populations at four others (Rocky Neck State Park, Niantic, CT; Charlestown, RI; Black Point, Narragansett, RI; and Westport, MA). The lack of detection of f. *yezoensis* or f. *narawaensis* at Guilford, CT is interesting in that the site lies at the distributional

convergence of both forms. It is also important to note that P. spp. *'stamfordensis',* which has likely been introduced recently, was the dominant *Porphyra* species at this site. Because there have been no previous collection records of any kind from this location, we do not know if either form of P. *yezoensis* was ever established in this site. But with the proximity of the marina and heavy recreational boat traffic, it is unlikely that this site has only been inoculated with P. spp. *'stamfordensis'.* If competition with either form of P. *yezoensis* has occurred at this site, it appears to have favored *P.* spp. *'stamfordensis'.*

That *P. yezoensis* f. *narawaensis* was not collected at Westport, MA in 2008 is curious in that it had been collected at the site, along with *P.* spp. *'stamfordensis',* by Bray (2006). Again, competition, at least in blade recruitment, has favored *P.* spp. *'stamfordensis.'* at this site.

The absence off. *yezoensis* and the presence of both f. *narawaensis Porphyra* spp. *'stamfordensis',* at the Falmouth Heights site is of great interest. Both f. *yezoensis* and *P.* spp. *'stamfordensis'* had been collected at this site previously (f. *yezoensis* in April 2004 and *P.* spp. *'stamfordensis'* in January of 2005) by Bray (2006), and f. *narawaensis* had not. *Porphyra* spp. *'stamfordensis'* was the dominant *Porphyra* species at this site throughout the 2008 study period, with f. *narawaensis* being first detected in April, at low density (300 total blades across two quadrats). The appearance of f. *narawaensis* was months behind its emergence at all other f. *narawaensis* sites in this study. That successful gametophytic blade recruitment of f. *narawaensis* followed the peak density period of P. spp. *'stamfordensis'* at this site suggests a competitive advantage for P. spp. *'stamfordensis'* during its months of peak production. It is likely that the recently

introduced f. *narawaensis* has lower conchocelis density at this stage of its introduction than does the established *P.* spp. *'stamfordensis'* and is therefore releasing fewer conchospores than its competitor.

That f. *yezoensis* was not detected at Falmouth Heights in 2008, following the arrival of f. *narawaensis,* is further evidence of competitive exclusion favoring the cultivar. Further investigation of the site's short, narrow, shallow channel, which connects Little Pond to the Atlantic Ocean, could reveal much about the nature of competition between *P. yezoensis* f. *yezoensis, P. yezoensis* f. *narawaensis,* and *P.* spp. *'stamfordensis',* especially if one could locate and observe the conchocelis phase of each throughout the year, or if one conducted laboratory culture experiments.

The lack of detection of f. *yezoensis* populations at New Bedford, MA and Falmouth Heights, MA, and the lack of detection of f. *narawaensis* at Westport, MA, is puzzling. If their absence was not the result of sampling error, it is possible that the forms have been completely eradicated from these locations. Another possibility is that the forms continue to exist at these locations in the perennating conchocelis phase, and no gametophytes successfully recruited into the intertidal zone this year due to: spore release during and ebb versus flood tide; rain or ice event that interfered with spore attachment; or the conchocelis many not release spores every year. The total absence of *P. yezoensis* at these locations is doubtful in that the conchocelis stage of *Porphyra* species is quite resilient and can remain viable for years under refrigeration without the addition of nutrients or exposure to sunlight (C. Yarish, personal communication).

But some evidence suggests that the conchocelis of the *P. yezoensis* forms may no longer reside at these sites. Given the right conditions, a very small amount of

conchocelis can give rise to an incredible number of progeny in a limited time period. For example, in a study of free-living *P. leucosticta* conchocelis, He and Yarish (2006) found that 1 g dry weight of conchosporangia could release over 20 million conchospores at peak production. With that level of fecundity it seems that if there were conchocelis reproducing in these locations, as has happened in the past, some of the millions of conchospores would have successfully recruited.

Although thorough collections at some previously identified *Porphyra yezoensis* sites did not detect the expected populations, the current studies were useful in measuring the presence, biomass, and density of entire macroalgal communities growing in close association f. *yezoensis,* f. *narawaensis,* and *P.* spp. *'stamfordensis'.* In doing so, baseline data was assembled for future comparative studies, which may be able to detect further changes in these macroalgal communities across time. Such comparisons are of great importance in assessing the effect of an introduced species on its host community.

The biomass and density data is of critical importance at the present time in documenting the autecology of different *Porphyra* populations (Figures 10 and 11). That is, peak blade production in *P. yezoensis* f *yezoensis* occurred earlier (February) than in *P.* spp. *'stamfordensis* '(March) and *P. yezoensis £ narawaensis* (March to April). The differential timing of production in the forms of P. *yezoensis* may reflect the genetic difference between the two. It is also possible that the trends seen in peak production time are not based on genetic differences between the forms, but rather are the result of biotic or abiotic differences between the various sites. Comparative examination of blade development in the two *P. yezoensis* forms, under controlled laboratory conditions, could better resolve this issue. If there is a genetic basis for the differential production timing

in the two forms of P. *yezoensis,* and blade recruitment space was a limiting factor, f. *yezoensis* would hold an advantage as a preemptor of space in recruitment competition with both P. spp. *'stamfordensis'* and *P. yezoensis* f. *narawaensis.*

The mean blade lengths of both forms of P. *yezoensis* were determined during the months of greatest luxuriance (March and April). The mean blade length of either form of *Porphyra yezoensis* was less than 5 cm at each site (Table 5). Considering that cultured nori blade lengths routinely exceed 60 cm and can reach lengths of up to 1.0 m (Miura and Aruga, 1987), the size of the *P. yezoensis* blades growing along Long Island Sound are very short. That the longest individual *P. yezoensis* blade recorded in the study was a mere 10 cm, only 10% of the maximum expected length, indicates that conditions for growth at these sites varies greatly from those in Asian nori-culture in which the blades are grown subtidally in protected bays and are thus protected from the stresses of wave action and exposure at low tide. The reduced stature of P. *yezoensis* in New England may result from exposure to the above stressors, lower nutrients, or a shorter growing season. In addition, the reduced stature observed could have been caused by grazing. Although little has been published on ingestion of P. *yezoensis* by grazing organisms (Noda et al., 2003), ducks, geese, or swans were observed in each of the study sites. However, consumption of P. *yezoensis* blades by these animals was never observed.

Although the average blade length *of Porphyra yezoensis* was short compared to Bray's (unpublished data, 2006) descriptions (f. *yezoensis* mean length of 8.9 cm + 0.89 SE and f. *narawaensis* mean length of 15.6 cm \pm 3.5 SE) and the lengths described for this species in nori-culture, they were similar to average blade length calculations for all

other *Porphyra* species recorded in this study (Table 6). Two species *(P. olivii, P.* spp. *'stamfordensis*') had mean blades lengths of less than 5 cm, and the mean blade length of *P. leucosticta* was less than 6 cm. The mean lengths were longer for two of these species in Bray's (unpublished data, 2006) study (P. spp. 'stamfordensis' = 7.6 cm + 0.98 SE and *P. leucosticta =* 7.3 cm + 1.2 SE). Discrepancies in length descriptions between the current study and that conducted by Bray are likely the result of sampling differences due to collection purpose and technique, with Bray's study selecting conspicuous, therefore larger, blades.

As a measure of impact, data collected in this study were used to determine the population sizes of both forms of *Porphyra yezoensis* in comparison to other macroalgal species in their host communities. The biomass of all macroalgal taxa was measured, and the mean percent contribution of P. *yezoensis* was calculated for each site. The percent contribution of both forms of P. *yezoensis* biomass was minimal (1% to 4%) for each site with significant *Fucus* populations. Even when *P. yezoensis* density was highest (greater than 10,000 blades per m^2 for f. *yezoensis* at New Haven), the biomass of the short, thin blades contributed little to the total community. It has been argued that only large, canopy forming, or turf forming macroalgae can become damaging, and therefore invasive, upon introduction to a host community. With their large thalli, these organisms modify the habitats in which they grow through space preemption and light blocking.

Although *Porphyra yezoensis* blades are not long enough to dwarf most macroalgae, they can block sunlight penetration to the organisms on which they grow epiphytically. As *P. yezoensis* commonly grows attached to long lived fucoid algae, which are keystone species in many macroalgal communities, this study sought to

determine if such epiphytic growth was harmful to these host plants. Therefore, the percent of *Fucus* with epiphytic *Porphyra* was determined for each site (Figure 16). During the peak months of growth (March and April), nearly half (48%) of all *Fucus* plants from New Haven bore loads of P. *yezoensis fyezoensis,* and roughly one-third (34% and 36%) of the *Fucus* plants from Charlestown and Rocky Neck bore loads *P. yezoensis* f *narawaensis.* Although these loads were substantial, it is also clear that *P. yezoensis* has yet to saturate its preferred substratum at these sites. With 52% to 66% of all *Fucus* fronds completely uncolonized, *P. yezoensis* populations have not likely reached maximal levels.

While the epiphytic *Porphyra yezoensis* loads were substantial, negative impact could not be assumed. Lengths *of Fucus* plants bearing *P. yezoensis* were compared to those free of epiphytic *Porphyra.* Because it was reasoned that plants bearing loads of epiphytic *P. yezoensis* would get less light, and possibly less nutrients than those without, it was hypothesized that load bearing plants would be shorter than non-load bearing plants. However, my findings were the opposite (Figure 17).

The observed greater mean lengths for *Fucus* plants bearing loads of P. *yezoensis* are not likely caused by the presence of these epiphytes. Rather their blades are likely found more often attached to longer *Fucus* plants because the longer plants have a greater surface area on which the blades can recruit, and/or longer *Fucus* plants are older and more worn, which may enhance their susceptibility to epiphytes.

Despite the statistical significance of the relationship between epiphytic *Porphyra yezoensis* and *Fucus* length, this study was not totally comprehensive as other epiphytic organisms *{Elachista fuciola, Ulva intestinalis, Ulothrix flacca,* and *Pylaiella littoralis)*

grew along with, or in the absence of, *P. yezoensis* on the measured *Fucus* fronds. Thus, the true effect, if any, of epiphytic *Porphyra* growth would be difficult to determine.

Another measure commonly used to assess the impact of an introduced organism on its host community is to compare the species richness of affected communities to the richness of unaffected but otherwise ecologically similar communities. Although the some study sites that possessed substantial populations *of P. yezoensis* varied markedly on some environmental parameters (salinity, waved action, exposure to tidal currents, substrata) from sites that did not contain populations of P. *yezoensis,* Table 3 shows that the *P. yezoensis* sites were home to more macroalgal species (14 species at New Haven, Rocky Neck, Charlestown, and Black Point) compared to the *P.* spp. *'stamfordensis'* dominated sites (11 at Westport and 8 at Guilford). Perhaps these higher species counts are an indication that both forms *P. yezoensis* exist as a background species at their sites and have not yet acted to exclude other macroalgal species. However, conclusions about the impact off. *yezoensis,* f. *narawaensis,* and *P.* spp. *'stamfordensis'* on the diversity of the study sites are only speculative due to the absence of pre-invasion data.

In summary, the current study was effective at establishing baseline structure data for seven macroalgal communities from New Haven, CT to Falmouth Heights, MA. A single season snapshot of density and biomass data was recorded for populations of the introduced Asian red algal species *Porphyra yezoensis* at these seven sites. The present data set will be useful to coastal managers conducting future comparative assessments of the macroalgal assemblages at these locations. The density and biomass data were useful for determining peak production times for both *P. yezoensis* forms and *P.* spp. *'stamfordensis'* across the study sites, as well as for evaluating the contribution of this

species to these communities. The epiphytic load *of P. yezoensis* on *Fucus* was determined for each site, and the effect of said epiphytes on *Fucus* stature was examined. Species counts for sites with and without current blade phase populations *of P. yezoensis* were compared. None of these evaluations revealed a clear negative impact of P. *yezoensis* on its host macroalgal communities.

The main difficulty with conducting impact assessments of introduced species is that most studies are conducted post invasion (Schaffelke and Hewitt, 2007). The present study is certainly an example of this phenomenon. The best way to conduct an impact assessment study is to thoroughly examine and catalog the algal community at a particular location both *before* and after an introduction. Failing this, many researchers have attempted to study concurrent and seemingly similar sites to compare the structure of communities with and without invaders. Because it is unclear if the uncolonized sites are uncolonized because they are abiotically or biotically different from colonized sites, some studies have suggested that the lack of pre-invasion data significantly limits the ability of a researcher to make inferences about the impact of the introduced species (Taylor 2002). Many researchers have also attempted to make post-invasion impact assessments through manipulation of the invaded environment. In the bulk of these studies, the introduced species is removed and the site is treated as uncolonized. However, some researchers have proposed that such techniques are flawed in that removal of the introduced organism from a site may reset the assemblage to an earlier successional stage rather than to pre-invasion condition (Edgar et al., 2004). Therefore, reliable impact inferences are limited.

The effectiveness of impact studies is also impaired due to limited time, with most studies lasting from weeks to, at most, a few years (Schaffelke and Hewitt, 2007). While this time scale is practical from the standpoint of research effort, it does not allow for the lag time between introduction and full blown invasion. Many invasive marine species initially exist at low levels for a period before increasing in number and expanding into new territory (Stockwell et al., 2003). The lag time may be caused by adaptation to environmental controls such as competition and herbivory. Also, density dependent survivorship thresholds might need to be reached before expansion can occur.

Studies of brief duration may also overestimate the impact of introduced species. Because grazers may initially avoid the ingestion of an introduced species, populations may expand rapidly. Over time, however, herbivore preferences have been seen to shift (Stimson et al., 2001). Such a shift reduces the competitive advantage and negative impact of the introduced organism on its host community.

Regardless of the duration of the impact assessment studies, some have argued that the invasiveness risk is minimal for most marine macroalgal organisms. Of the 260 introduced species worldwide, only 17 have been considered at all with only 4 of these *(Caulerpa taxifolia, Undaria pinnatifida, Codium fragile* ssp. *fragile,* and *Sargassum muticum)* being highly studied (Johnson, 2007). Hence the impact of most introduced seaweeds is minimal or their impacts are often unclear. Because alien species that actively modify habitats have a much higher negative impact on new environments than organisms that do not (Wallentinus and Nyberg 2007), the impact of introduced marine animals are undoubtedly greater than the impact of algal species.

While the impact of both forms of *Porphyra yezoensis* on their host communities in New England is currently unclear, the data gathered from this study will provide a baseline for further monitoring of their impacts on macroalgal assemblages. Although this study revealed that *P. yezoensis* had seemingly disappeared from three sites where it had previously been observed (f. *narawaensis* from Westport, MA, and f. *yezoensis* from both New Bedford, MA and Falmouth Heights, MA), the range of f *narawaensis* had expanded east nearly 100 miles to Falmouth Heights, on Cape Cod (Figure 18). Range expansion of introduced species is a concern, as it indicates the ability of this species to continue to spread throughout New England, inhabiting new communities, with potential negative effects.

Also of note, *Porphyra* spp. *'stamfordensis',* a species of unknown origin, was found to inhabit five of the seven study sites, and it was the dominant species at three sites. Given that this species was first detected at Hammonassett State Park by Neefus in 2004 (Bray, 2006), and nothing more is known about its introduction to the region, continued monitoring and impact assessment of this species should be coupled with these same efforts for *P. yezoensis.*

While *Porphyra yezoensis* f. *yezoensis, P. yezoensis* f. *narawaensis,* and *P.* spp. *'stamfordensis'* are established and possibly invasive in New England, attempts to eradicate them would be difficult, if not impossible, at this time. Physical removal of all gametophytic blades would be implausible due to their abundance and small size. Chemical treatment of the infected shores would likely have little effect on the subtidaldwelling, and blade-producing conchocelis stages of these organisms and would likely be devastating to native species.

Efforts to reduce the further spread of these species would be difficult and costly. The curtailing of recreational boat traffic, fishing, and shipping in affected regions is unreasonable, at this time, in light of the fact that negative impacts of these species have yet to be observed in New England.

TABLES

Table 1 GPS coordinates of study sites

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P.
umbilicalus linearis yezoensis narawaens/s 'stamfordensis' 'collinsii' leucosticta tsengii purpurea spatulata amplissima umbilicalus linearis yezoensis f yezoensis f P. spp. P. spp. P. P. P. P. P. P. P. $\overline{\mathsf{x}}$ \times \times $\times \times \times \times$ $\pmb{\times}$ P.
purpurea spatulata amplissima \times \times \times \times \times \times \times \times X \times \times P.
tsengii \times P.
Ieucosticta x \times X X P spp.
'collinsii' \times \times P. spp.
'stamfordensis' $\hat{\mathbf{x}}$ P.
yezoensis f
narawaensis $\tilde{\mathbf{x}}$ *p. p.* \times \times X P.
yezoensis f
yezoensis **X?** X \times X X X X X X Lead Hazard Bridge, Marblehead, MA Lead Hazard Bridge, Marblehead, MA Morrissey Boat Ramp, S. Boston, MA Buzzard's Bait Bridge, Wareham, MA Morrissey Boat Ramp, S. Boston, MA Rocky Neck State Park, Niantic, CT Westport Boat Ramp, Westport, MA Pope's Island Marina, Fairhaven, MA Pope's Island Marina, Fairhaven, MA Buzzard's Bait Bridge, Wareham, MA Westport Boat Ramp, Westport, MA Rocky Neck State Park, Niantic, CT Village Inn Beach, Narragansett, RI Village Inn Beach, Narragansett, Rl Victory Rd Park, S. Boston, MA Victory Rd Park, S. Boston, MA Niantic Boat Valet, Niantic, CT Niantic Boat Valet, Niantic, CT Carson Beach, S. Boston, MA Carson Beach, S. Boston, MA Goose Cove, Gloucester, MA Leeman Hwy, Brunswick, ME Black Point, Narragansett, RI Black Point, Narragansett, Rl Goose Cove, Gloucester, MA Dover Point, Newington, NH Leeman Hwy, Brunswick, ME Great Island, Harpswell, ME Dover Point, Newington, NH Great Island, Harpswell, ME Orr Island, Harpswell, ME Orr Island, Harpswell, ME Marblehead Neck, MA Marblehead Neck, MA Seapoint, Kittery, ME Pikeland, Lubec, ME Seapoint, Kittery.ME Cundy's Harbor, ME Cundy's Harbor, ME Salem Willows, MA Salem Willows, MA Mackerel Cove, RI Mackerel Cove, Rl Machiasport, ME Machiasport, ME Cutler, ME

Table 2 Winter 2006-2007 rapid assessment survey results - *Porphyra* species by site Table 2 Winter 2006-2007 rapid assessment survey results - Porphyra species by site

The'?' symbol means species identitification was based on morphology alone. The "? symbol means species identification was based on morphology alone.

X

X X

Pikeland, Lubec, ME Lubec Town Dock, Lubec, ME

Lubec Town Dock, Lubec, ME

Site	Contribution	Porphyra species
New Haven	4%	P. yezoensis f yezoensis
Guilford	1%	P. spp. 'stamfordensis'
Rocky Neck	2%	P. yezoensis f narawaensis
Charlestown	2%	P. yezoensis f narawaensis
Black Point	81%	P. yezoensis f narawaensis
Westport	1%	P. spp. 'stamfordensis'
Falmouth Heights	2%	P. spp. 'stamfordensis'

Table 4 Maximum percent biomass contribution for dominant *Porphyra* species at different sites.

 $\hat{\mathcal{A}}$

Porphyra yezoensis Mean Blade Length (cm)						
	March	April	Overall Average	S.D.	N	
New Haven	2.05	2.85	2.40	0.91	427	
Rocky Neck	4.13	4.45	4.24	194	386	
Charlestown	3.79	4.08	3.88	1.27	349	
Black Point	4.51	4.21	4.38	150	501	

Table 5 Mean blade length of *Porphyra yezoensis* fronds at different sites.

 $\bar{\gamma}$

Table 6 Mean blade length of different *Porphyra* taxa at different sites.

FIGURES

Figure 1 Life history phases of *Porphyra yezoensis.* Terminology from Holmes and Brodie (2005).

o Porphyra yezoensis f narawaensis

 \triangle Porphyra yezoensis f yezoensis

Figure 2 Study sites with previously confirmed *Porphyra yezoensis* populations. Sites from west to east are New Haven (NH), Guilford (G), Hammonassett State Park (H), Rocky Neck State Park (RN), Niantic Boat Valet (NI), Charlestown Breachway (CB), Black Point (BP), Westport Boat Ramp (W), New Bedford (NB), and Falmouth Heights (FH).

Figure 3 Wide and close aerial views of New Haven Light study site. Image courtesy of Google Earth [™] mapping service.

Figure 4 Wide and close aerial views of the Guilford Marina study site. Image courtesy of Google Earth ™ mapping service.

Figure 5 Wide and close aerial views of the Rocky Neck State Park study site. Image courtesy of Google Earth ™ mapping service.

Figure 6 Wide and close aerial views of the Charlestown Breachway study site. Image courtesy of Google Earth ™ mapping service.

Figure 7 Wide and close aerial views of the Black Point study site. Image courtesy of Google Earth ™ mapping service.

Figure 8 Wide and close aerial views of the Westport Boat Ramp study site. Image courtesy of Google Earth ™ mapping service.

Figure 9 Wide and close aerial views of the Google Earth ™ mapping service. Falmouth Heights study site. Image courtesy of

Figure 10 Winter 2007 rapid assessment survey results- *Porphyra yezoensis* f. *yezoensis* and P. *yezoensis* f. *narawaensis* presence/absence by site

Figure 11 Mean monthly *Porphyra yezoensis* density by site.

Figure 12 Monthly mean density of *Porphyra yezoensis* f. *yezoensis* and *P. yezoensis* f. *narawaensis.* Asterisks indicate significant differences between the two forms ($p < 0.05$).

Figure 13 Mean monthly *Porphyra yezoensis* biomass by site.

Figure 15 Mean monthly *Porphyra* spp. *'stamfordensis'* biomass by site.

 α **C** α Figure 16 Percentage of Fucus plants beari were made using all *Fucus* fronds collected in March and April. Abbreviations repr Porphyra yezoensis wild type, P. yezoensis f narawaensis, and P. spp. *'stamfordensis*

Figure 17 Mean *Fucus* length with and without epiphytic *Porphyra* loads. The checkered bars represent mean lengths *of Fucus* plants without epiphytic loads. *Porphyra yezoensis f. yezoensis* is represented by striped bars, *P.* spp. *'stamfordensis'* by dotted bars, and *P. yezoensis fnarawaensis* by cross hatched bars. An asterisk between two bars represents a significant length difference as calculated by Tukey's test.

Figure 18 *Porphyra* species collected by site in the current study. Sites from west to east are New Haven (NH), Guilford (G), Hammonassett State Park (H), Rocky Neck State Park (RN), Niantic Boat Valet (NI), Charlestown Breachway (CB), Black Point (BP), Westport (W), and Falmouth Heights (FH).

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 $\hat{\mathcal{A}}$

New Haven Light, New Haven, CT **New Haven Light, New Haven, CT**

Destructive Sampling Count per m A 2 February

Scytosiphon lomentaria

 $\frac{1}{2}$

Destructive Sampling Count per m

A 2 March

200 200 400

Falmouth Heights, MA **Falmouth Heights, MA**

Appendix B: Monthly biomass by site **Appendix B: Monthly biomass by site**

New Haven Light, New Haven, CT **New Haven Light, New Haven, CT**

Destructive Sampling Damp/Dry weight (g/m Destructive Sampling Damp/Dry weight (g/m^2) February 2) February

Pylaiella littoralis

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J

Rocky Neck State Park, Niantic, CT

Rocky Neck State Park, Niantic, CT

 $\hat{\mathcal{A}}$

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Charlestown Breachway, Charlestown, RI **Charlestown Breachway, Charlestown, Rl**

 $\hat{\mathcal{L}}$

