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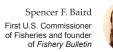
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NOAA National Marine Fisheries Service

Fishery Bulletin





Abstract-Fish species of the Middle Atlantic Bight (MAB) continental shelf are well known; however, species occupying hard-bottom habitats, particularly on the outer shelf, are poorly documented. Reef-like habitats are relatively uncommon on the MAB shelf; therefore, shipwrecks may represent a significant habitat resource. During fall 2012 and spring 2013, 9 sites (depths: 42-126 m) near Norfolk Canyon were surveyed by using remotely operated vehicles. One site consisted of sand bottom, one consisted of predominantly natural hard bottom, and 7 sites included 8 large shipwrecks. Of 38 fish taxa identified, 33 occurred on hard bottom and 25 occurred on soft substrata. Fourteen fish taxa occurred almost exclusively on hard bottom, and 6 species were observed only on soft bottom. The most abundant taxa, especially on reef habitat, were the chain dogfish (Scyliorhinus retifer), a scorpionfish (Scorpaena sp.), the yellowfin bass (Anthias nicholsi), the red barbier (Baldwinella vivanus), the black sea bass (Centropristis striata), unidentified anthiine serranids, and the deepbody boarfish (Antigonia capros). Depth, location, and season did not significantly influence fish assemblages. Fish assemblages on natural and artificial hard-bottom habitat were similar but significantly different from soft-bottom assemblages. Deep-reef fishes of the southern MAB may be constrained by zoogeography, depth, and inadequate habitat-limitations that could increase their vulnerability.

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Fish species associated with shipwreck and natural hard-bottom habitats from the middle to outer continental shelf of the Middle Atlantic Bight near Norfolk Canyon

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The fish fauna of the shelf and upper slope of the U.S. Middle Atlantic Bight (MAB) (from Cape Hatteras to Cape Cod) is considered cool temperate, although fish enter from colder and warmer regions to the north and south, respectively. The estuarine and shelf fishes are particularly well studied in this region (e.g., Grosslein and Azarovitz, 1982; Colvocoresses and Musick, 1984; Gabriel, 1992; Murdy et al., 1997; Able and Fahay, 1998), in large part, because of decades of standardized, fishery-independent trawl surveys. Although fish communities have been documented on the open shelf and upper slope, their presence on untrawlable habitats (i.e., canyon walls, rocky bottom, and shipwrecks) has not been well documented.

The shelf of the Middle Atlantic Bight has a lower percentage of exposed natural hard substrata than that of other areas in U.S. Atlantic

waters (Steimle and Zetlin, 2000; SEAMAP-SA, 2001). Therefore, habitat may be limiting for fauna in the MAB that require hard substrata, and therefore introduced shipwrecks or other reef-like habitats probably represent significant habitat resources. Even so, there has been little assessment of the fishes associated with either natural or artificial hard-bottom habitats in the MAB (Eklund, 1988; Adams, 1993; Steimle and Zetlin, 2000). Although direct observation techniques are preferred for assessment of the fauna of rugged hard substrata (e.g., Caillet et al., 1999; Quattrini and Ross, 2006; Ross and Quattrini, 2007), these methods have not been widely applied on the MAB shelf. Three studies that involved nearshore surveys in the MAB used direct observation to document fishes on various bottom types, including hard bottom, at depths ≤55 m (Auster et al., 1991;

Table 1

Details from dives of remotely operated vehicles (ROVs) on the continental shelf of the Middle Atlantic Bight during 2012 (ROV Kraken II) and 2013 (ROV Jason II), which were conducted from the NOAA ship Nancy Foster and the NOAA ship Ronald H. Brown. For site names, W=shipwreck sites; SS=shallow, soft-substrata site; NHB=natural hard-bottom site. Total time and depth range are for times when the ROV was on the bottom. Daytime (D)=0800-2000 h EDT; nighttime (N)=2000-0800 h EDT. n/a=not available.

Dive no.	Site name	Date	Time	Total time (min)	Start latitude (N)	Start longitude (W)	End latitude (N)	End longitude (W)	Depth range (m)
ROV-2012-NF-21	SS	20-Sep-12	D	304	37°10.90′	74°56.24′	37°10.85′	74°56.26′	42–43
ROV-2012-NF-22	W-1	22-Sep-12	D	622	37°09.40′	74°45.30′	n/a	n/a	81
ROV-2012-NF-23	W-2	23-Sep-12	D	612	37°09.40′	74°34.60′	37°09.20′	74°34.40′	113
ROV-2012-NF-24	W-3	24-Sep-12	D	519	37°13.90′	74°33.00′	37°14.00′	74°33.00′	124-126
ROV-2012-NF-26	W-4	26-Sep-12	D	223	37°11.50′	74°34.40′	37°11.50′	74°34.40′	100-106
ROV-2012-NF-27	W-5	26-Sep-12	D	363	37°16.90′	74°32.10′	37°17.20′	74°32.00′	118-119
ROV-2012-NF-28	NHB	27-Sep-12	D	291	37°01.06′	74°39.26′	37°00.92′	74°39.64′	98 - 117
ROV-2012-NF-29	W-6	27-Sep-12	D	251	36°54.80′	74°42.40′	36°54.80′	74°42.40′	84-85
ROV-2012-NF-30	W-7	28-Sep-12	D	174	37°11.90′	74°45.40′	37°11.90′	74°45.40′	68-69
ROV-2013-RB-692	W-4	19-May-13	N	295	37°11.50′	74°34.50′	37°11.50′	74°34.40′	91 - 105
ROV-2013-RB-693	W-2	20-May-13	D	894	37°09.40′	74°34.40′	37°09.40′	74°34.70′	90-116
ROV-2013-RB-694	W-3	21-May-13	D	861	37°13.90′	74°33.10′	37°14.00′	74°33.10′	101-126
ROV-2013-RB-695	W-5	22-May-13	D	504	37°16.80′	74°32.10′	37°17.00′	74°32.20′	106-121
ROV-2013-RB-696	W-2	23-May-13	D	197	37°09.40′	74°34.50′	37°09.40′	74°34.60′	90-114

Adams, 1993; Diaz et al., 2003). Similar assessments in deeper waters of the middle to outer shelf are lacking, aside from those obtained from submersible surveys directed toward tilefish (Lopholatilus chamaeleonticeps) at depths of 117–268 m (Grimes et al., 1986). In these studies, the physical structure of habitat was observed to be correlated with fish distribution patterns. Higher profile, more complex habitats generally supported greater fish species richness and higher abundance for some species. Bioengineering by tilefish and associated species in and near canyon heads also created complex habitats for other outer shelf fauna (Grimes et al., 1986).

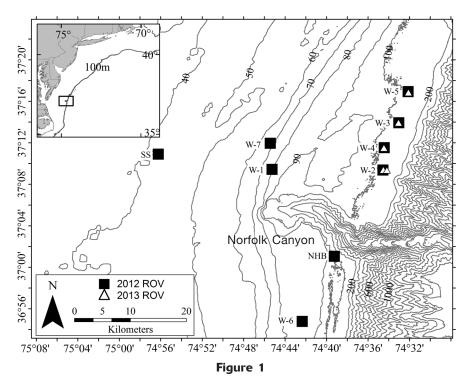
Non-natural hard substrata (e.g., shipwrecks) aggregate fish and invertebrates. The effects of artificial reefs composed of shipwrecks and other structures (e.g., drilling platforms and fish attracting devices) are well known but their use as fish habitat is still being debated (Stephan and Lindquist, 1989; Grossman et al., 1997; Perkol-Finkel et al., 2006). It is unclear whether artificial reef structures actually increase populations of fish as opposed to simply concentrating them, and understanding the role of artificial reefs is increasingly important considering the decline of natural reefs worldwide (Perkil-Finkel et al., 2006). However, Arena et al. (2007) reported that vessel-reefs off southeastern Florida supported significantly higher fish species richness and abundance than at natural reefs and that different community structures and trophic patterns were observed for the 2 habitat types, and they suggested that vessel-reefs enhanced local fish populations. The extent to which artificial reefs mimic natural reef functions requires further study, and artificial reefs may only approach the functions of natural reefs if their physical structures are similar (Perkol-Finkel et al., 2006).

As part of a larger survey of submarine canyons and nearby features in the MAB, historically important shipwrecks, naturally occurring hard bottom, and sandy bottom areas on the outer continental shelf near Norfolk Canyon were surveyed with remote operated vehicles (ROVs) in 2012 and 2013. In this article, we document 1) species of overall fish communities on shelf-depth artificial (shipwrecks) substrata and natural hard substrata and nearby soft-bottom habitats, 2) relative abundance of fish species in those communities, and 3) behaviors and distributions of fishes on shipwreck and nonshipwreck open bottom for 2 seasons (fall in 2012 and spring in 2013). The degree to which fishes were associated with hard bottom and the degree to which such habitats supported unique communities were investigated.

Materials and methods

Study area

On the middle to outer continental shelf (depths of 42–126 m) of the southern MAB, 9 locations in the vicinity of Norfolk Canyon were surveyed with ROVs (Table 1, Fig. 1). These study sites had been mapped with multibeam sonar in 2011. The shallowest location (site SS) was the only one entirely composed of flat, soft sediment



Locations of the 9 sites, 7 shipwreck sites (W-1 through W-7), 1 mostly natural hard-bottom (NHB) site, and 1 soft-bottom (SS) site, where remotely operated vehicles (ROVs) were used to collect data concerning fish assemblages on 20–28 September 2012 and 19–23 May 2013. Depth contours are given in meters. The inset illustrates the Middle Atlantic Bight; the rectangle indicates the study area.

bottom. All other sites comprised predominately hard, complex substrata, but they were also surrounded by soft substrata (see "Habitat definitions" section). The natural hard-bottom (NHB) location was dominated by hard, rough bottom, including boulders, rubble fields, and walls of consolidated mud. The dominant habitat in 7 of the study sites (e.g., W-1) was composed of 8 historically important shipwrecks, all sunk during the early 1920s (i.e., 6 of the ships were part of the "Billy Mitchell fleet" [Lee, 1949]). These shipwrecks had the following lengths and maximum heights off bottom: W-1 (45×6 m), W-2 (167×18 m), W-3 (141×7 m), W-4 (301×3 m), W-5 (2 shipwrecks about 685 m apart; 64×3 m and 53×2 m), W-6 (171×14 m), W-7 (72×3 m). The shipwrecks were surrounded by soft substrata (sand or gravel). All shipwrecks were covered to varying degrees with lost fishing gear (trawls, Fig. 2, A and C).

Remotely operated vehicle

Dives of the ROV were conducted with the University of Connecticut ROV *Kraken II* deployed from the NOAA Ship *Nancy Foster* on 20–28 September 2012 and, during a second research cruise, with the Woods Hole Oceanographic Institution ROV *Jason II* deployed from the NOAA Ship *Ronald H. Brown* on 19–23 May 2013. The site SS, the site NHB, and sites W-1, W-6, and W-7 each were sampled with 1 ROV dive, and each

of the other shipwreck sites were sampled either 2 (W-3, W-4, and W-5) or 3 (W-2) times for a total of 14 ROV dives (Table 1). The position of the ROV was recorded continuously by using an ultrashort baseline tracking system, and navigation data were time synchronized with all imagery and samples. An SBE 911plus1 conductivity, temperature, and depth (CTD) instrument (Sea-Bird Electronics Inc., Bellevue, WA) was attached to the ROVs to record conductivity (in microSiemens per centimeter), temperature (degrees Celsius), salinity, density ($\sigma\theta$, in kilograms per cubic meter), dissolved oxygen (DO, in milliliters per liter), depth, and pH at a frequency of once per second during each dive. Only temperature, salinity, and DO data recorded during dives while the ROVs were on or near bottom are presented.

Digital video was collected as the ROVs moved along transects at slow speeds, <~25 cm/s (<0.5 kt), across all habitat types and with the vehicles as near to bottom as possible. At each shipwreck, video and still image photo-mosaics were recorded over a series of parallel transects that covered the entire shipwreck. This digital imagery was used for vessel identification, examination of present ambient conditions and documentation

¹ Mention of trade names or commercial companies is for identification purposes only and does not imply endorsement by the National Marine Fisheries Service, NOAA.

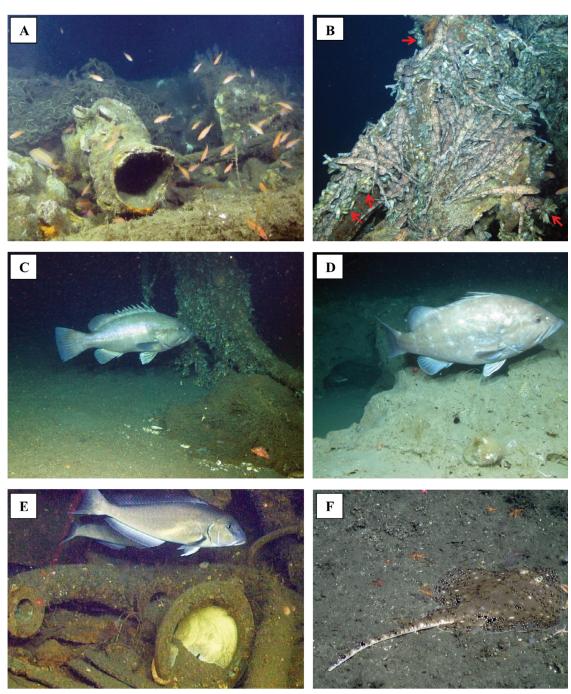


Figure 2

Photographs of fishes and habitats surveyed with remotely operated vehicles in 2012 and 2013 near Norfolk Canyon in the Middle Atlantic Bight: (A) school of unidentified anthiine serranids and at least one yellowtail bass (Anthias nicholsi, mid-left) on shipwreck site W-1, 81 m, 22 September 2012; (B) dense aggregations of chain dogfish (Scyliorhinus retifer) lying on shipwreck structure (site W-5, ~115 m, 26 September 2012), 4 yellowfin bass (2 upper right, 2 lower right), and red arrows indicate clusters of egg cases of chain dogfish; (C) warsaw grouper (Hyporthodus nigritus), and a scorpionfish (Scorpaena sp., lower right lying on trawl net) on shipwreck site W-5, 118 m, 26 September 2012; (D) snowy grouper (Hyporthodus niveatus) on the natural hard-bottom (NHB) site, ~110 m, 27 September 2013; the scaling laser dots near the anal fin indicate that this fish is at least 150 cm long in total length; (E) 2 blueline tilefish (Caulolatilus microps) on shipwreck site W-2, ~100 m, 20 May 2013; and (F) rosette skate (Leucoraja garmaini) on sandy habitat near shipwreck site W-2, ~100 m, 23 May 2013.

of biota and habitats. The color video cameras attached to the ROVs had scaling lasers (10-cm spacing) used to estimate total length (TL) of fish. During transect surveys, the cameras were positioned to record directly in front of the ROV and were set on wide angle (or near wide angle). The video cameras recorded continuously throughout the ROV dives (whether the ROV was moving over transects or was stationary on bottom), and digital still images were taken frequently to augment video collection.

Habitat definitions

A main objective was to determine to what degree fishes were associated with general habitats on a large scale; therefore, habitat definition was reduced to 2 broad, relatively simple types: 1) soft substrata (SS) of sand or mud—relatively flat substrata and with few structuring features aside from gravel, burrows, depressions, and animal tracks; and 2) artificial (shipwreck) substrata and natural hard bottom (AS/NHB), which included World War I-era shipwrecks with substantial vertical profile and one site with natural hard bottom (consolidated mud, ledges, and boulders). Additional habitat metrics included bottom depth and environmental data recorded by the CTD instruments mounted on the ROVs.

Video analysis: community and habitat association

A preferred method for documenting fauna in complex habitats, visual observations (here based on ROV-collected video) were used to describe the fish communities and associated habitats at the 9 study sites. Tracks of ROV dives were processed initially to conservatively remove erroneous tracking data (location points) as described by Quattrini et al. (2012). To determine community structure and habitat associations of fishes at sites, much as described in Ross and Quattrini (2007), videos from each dive were viewed multiple times for habitat classifications (see "Habitat definitions" section) and for identification (to the lowest possible taxa) and enumeration of fishes by time of observation. Video segments were designated when the ROV stopped or started movement, when the video quality changed, or when the habitat changed. Depth was recorded by the ROV-mounted SBE 911plus for every time segment. Unusable video (out of focus, too far off bottom, because of malfunction, sediment clouds) was removed from the data set.

Species composition and relative abundances (fish counts) were determined from the wide-angle video and were compared within each of and between the 2 habitat types. To compare abundances of all species within a habitat type, relative abundances were calculated in percentages as the number of individuals per taxa per habitat type divided by the total number of individuals observed per habitat type and then multiplied by 100. For comparisons between habitat type, analysis was restricted to benthic fishes identified to at least fam-

ily level and with overall abundances ≥2. Occurrence of at least 2 individuals allowed for the possibility of a taxa occurring in both habitat types. Relative abundances by habitat type were calculated for each taxon by dividing the number of individuals in a particular habitat type by the total number of individuals of the same species from both habitat types and multiplying the result by 100.

Multivariate analyses were conducted in PRIMER 6 and PERMANOVA+ (PRIMER-E Ltd., Ivybridge, U.K.) (Clarke and Warwick, 2001; Clarke and Gorley, 2006; Anderson et al., 2008) to determine differences in benthic fish assemblages between habitat types. Sample units were the numbers of each species per habitat type (SS or AS/NHB) per ROV dive; samples with no species present were removed from the data set. Because transect times were variable, abundances of species were standardized per sample by dividing the number of individuals per species by the total number of fishes per sample. Standardized abundances were fourth-root transformed to down-weight the abundant species in relation to rare species. The Bray-Curtis similarity coefficient was used to calculate similarities between samples, and on the basis of the resulting similarity matrix a nonmetric multidimensional scaling ordination (MDS) plot and a dendrogram with group average linking were created. One-way analysis of similarities (ANOSIM) and post-hoc multiple comparison tests were used to determine whether there were significant differences between fish assemblages in the 2 different habitat types. Similarity percentage (SIMPER) analysis was used to determine which species contributed to the dissimilarities among habitat types.

Results

On the 9 study sites (depths of 42–126 m), 14 ROV dives were completed, 9 dives in September 2012 and 5 dives in May 2013 (Table 1, Fig. 1), resulting in 84.4 h of usable video data on hard-bottom (AS/NHB) habitat and 16.5 h of video data on soft-bottom (SS) habitat. Soft-bottom habitat was observed exclusively with video collected during the dive at the shallowest site (Table 1, Fig. 1); however, because only 3 specimens of unidentified skates were observed during this dive, it made little contribution to our study. Although shipwrecks and natural hard bottom were the focus of the remaining dives, soft-bottom habitat surrounding those hard-bottom habitats was also surveyed during these dives.

In September 2012, mean bottom temperatures varied about 2.5°C across the study sites; the coldest temperatures (means: 11.9–13.0°C), lowest salinities (means: 33.1–34.8), and highest DO (means: 4.0–4.5 mL/L) occurred at the shallower sites (depths of 42–81 m) (Table 2). At each of the 5 deeper sites (depths of 84–126 m), bottom temperatures (means: 14.2–14.5°C), salinities (35.6–35.8), and DO (3.7–4.0 mL/L) were similar to each other. In May 2013, little variation was

Table 2

Environmental data (means, ranges, and standard errors of the means in parentheses) recorded at bottom by the SBE 911plus system (one exception indicated at the footnote) attached to the remotely operated vehicles Kraken II (2012) and Jason II (2013) during surveys of shipwrecks and sandy bottoms on the continental shelf near Norfolk Canyon in the Middle Atlantic Bight. Site locations are appended to the dive year in each dive no. (see Table 1). NA=not available. DO=dissolved oxygen.

Dive no. Temperature (°C)		Salinity	DO (mL/L)	
2012-NF-21-SS	12.16, 11.95–12.29 (0.0007)	33.14, 33.16–33.17 (0.0001)	4.51, 4.45–4.55 (0.0002)	
2012-NF-22-W1	11.94, 10.73 - 14.61 (0.0029)	34.16, 32.92–34.94 (0.0015)	4.22, 4.08–5.54 (0.0005)	
2012-NF-23-W2	NA	NA	NA	
2012-NF-24-W3	14.31, 14.10–14.47 (0.0003)	35.65, 35.32–35.80 (0.0005)	3.88, 3.67-4.14 (0.0003	
2012-NF-26-W4	14.47, 14.40–14.52 (0.0001)	35.78, 35.69-35.80 (0.0001)	3.98, 3.95-4.11 (0.0001)	
2012-NF-27-W5	14.15, 14.00–14.46 (0.0008)	35.75, 35.74–35.80 (0.0000)	3.69, 3.51-4.84 (0.0006)	
2012-NF-28-NHB	14.33, 14.22–14.39 (0.0002)	35.71, 35.50-35.77 (0.0004)	3.93, 3.83-4.06 (0.0004)	
2012-NF-29-W6	14.21, 13.86–14.30 (0.0009)	35.63, 35.12–35.72 (0.0013)	3.84, 3.67–4.85 (0.0004)	
2012-NF-30-W7	13.00, 12.53-13.41 (0.0013)	34.76, 34.45-35.09 (0.0008)	3.99, 3.87-4.78 (0.0004)	
2013-RB-692-W4	13.16, 13.09–13.32 (0.0005)	34.83, 32.20–35.72 (0.0085)	2.99, 1.81–4.22 (0.0042)	
2013-RB-693-W2	13.27, 13.18–13.50 (0.0003)	32.81, 29.66–33.14 (0.0033)	2.44, 1.38-4.31 (0.0018)	
2013-RB-694-W3	13.42, 12.90–13.49 (0.0004)	35.48, 32.64-35.94 (0.0027)	3.08, 2.09–4.46 (0.0021)	
2013-RB-695-W5	13.19, 13.10–13.45 (0.0004)	29.42, 26.38–35.18 (0.0010)	NA^{1}	
2013-RB-696-W2	$13.32, 13.00 - 13.60 \ (0.0012)$	35.69, 35.63–35.77 (0.0002)	$4.71,4.574.76\;(0.0003)$	

 $^{^{}I}$ Data were taken from the Jason~II conductivity, temperature, and depth (CTD) system, SBE 911plus, which was not operating.

again observed among the 5 deeper sites, but temperatures (means: 13.2–13.4°C) were on average a degree colder than they were in 2012. At these sites, more variations in salinity (means: 29.4–35.7) and DO (means: 2.4–4.7 mL/L) were recorded that those recorded in 2012. It seems unlikely that the small environmental variations were biologically significant to these temperate, wide-ranging fishes, particularly at the deeper sites, but monitoring over longer periods is required to determine the scale of environmental variation.

From analysis of the video from ROV dives, 38 unique fish taxa, representing at least 25 families, were identified (Urophycis sp., Hyporthodus sp., Anthiinae (unidentified), Caulolatilus sp., Labridae (unidentified), and unidentified fish not included in total counts; Table 3). Of those 38 taxa, 33 occurred on the AS/NHB habitat type (14 species were observed only on hard bottom), and 25 taxa occurred on the SS habitat type (6 occurred only on soft bottom) (Table 3). The lower number of species observed in the SS habitat type was at least partly due to lower dive effort there (Table 3). Three taxa, a requiem shark (Carcharhinus sp.), the greater amberjack (Seriola dumerili), and the ocean sunfish (Mola mola), that occurred over or near either habitat type are considered pelagic fish that are less constrained to benthic habitats.

Fish assemblages on each habitat type were numerically dominated by relatively few species. On the AS/NHB substrata, 96.5% of the community was composed of 7 taxa (in decreasing order of abundance): unidentified anthine serranids, the chain dogfish (*Scyliorhi*-

nus retifer), the yellowfin bass (Anthias nicholsi), the deepbody boarfish (Antigonia capros), the red barbier (Baldwinella vivanus), a scorpionfish (Scorpaena sp.), and the black sea bass (Centropristis striata). Anthiine serranids (all combined, including yellowfin bass, red barbier, and unidentified members of this subfamily) and the chain dogfish (Fig. 2, A and B) were each an order of magnitude (2 orders of magnitude compared with most species) more abundant than any other taxa in either habitat type. Most of the Anthiinae fishes that were observed were probably red barbier, but small, rapidly moving anthiines can be difficult to identify in situ; some of these fishes could have been the streamer bass (B. aureorubens), longtail bass (Hemanthias leptus), or threadnose bass (Choranthias tenuis). The smaller (~60-180 mm TL) fishes of the subfamily Anthiinae occurred as dense aggregations whose members swam rapidly around hard-bottom structures (Fig. 2 A), occasionally straying over nearby sandy bottom. Larger (usually ~130-200 mm TL) yellowtail bass were more solitary and often associated with the anthiine schools (Fig. 2, A and B).

Six taxa (in decreasing order of abundance: the chain dogfish, the deepbody boarfish, the black sea bass, *Scorpaena* sp., the yellowfin bass, and anthiine serranids), accounted for 93.6% of the fauna on the SS habitat type but usually exhibited a lower percent contribution to the SS habitat type than to the AS/NHB habitat type (Table 3). Species that were unique to either habitat type occurred in low abundance (<1% of total abundance within habitat).

Table 3

Relative abundance (%) of fishes observed during dives of remotely operated vehicles in 2012 and 2013 on 2 habitats types: 1) artificial (shipwreck) substrata and natural hard bottom (AS/NHB) and 2) soft substrata (SS) near Norfolk Canyon, Middle Atlantic Bight. Number of hours of observation (usable video) and depth ranges are provided under each habitat type.

Taxa	AS/NHB 84.42 h, 63–126 m	SS 16.48 h, 40–126 m
0. 11. 11.	•	•
Scyliorhinidae	00.150	F.O. 100
Scyliorhinus retifer, chain dogfish	30.158	56.489
Carcharhinidae	0.000	
Carcharhinus sp., requiem shark	0.002	
Rajidae		0.000
Leucoraja garmani, rosette skate		0.339
Rajidae (unidentified)	0.002	0.594
Ophichthidae		0.450
Ophichthus cruentifer, margined snake eel		0.170
Congridae		0.450
Conger oceanicus, conger eel	0.596	0.170
Gadiformes (unidentified), cods	0.002	0.085
Moridae		
Physiculus fulvus, metallic codling	0.094	
Phycidae		
Phycis chesteri, longfin hake	0.002	
Urophycis chuss, red hake	0.002	
Urophycis regia, spotted hake	0.006	
Urophycis sp.	0.083	0.085
Lophiidae		
Lophius americanus, goosefish		0.085
Trachichthyidae		
Gephyroberyx darwinii, big roughy	0.600	
Macroramphosidae		
Macroramphosus scolopax, longspined snipefish	0.557	0.254
Scorpaenidae		
Scorpaena sp., scorpionfish	1.249	6.107
Triglidae		
Prionotus sp., searobin	0.004	0.085
Polyprionidae		
Polyprion americanus, wreckfish	0.004	
Serranidae		
Anthias nicholsi, yellowfin bass	10.667	1.442
Anthias sp.	0.002	
Baldwinella vivanus, red barbier	3.904	0.085
Centropristis striata, black sea bass	1.180	6.531
Hyporthodus nigritus, warsaw grouper	0.015	
Hyporthodus niveatus, snowy grouper	0.058	0.085
Hyporthodus sp.	0.004	
Pronotogrammus martinicensis, roughtongue ba	ss 0.006	
Anthiinae (unidentified)	44.606	1.442
Malacanthidae		
Caulolatilus microps, blueline tilefish	0.369	0.594
Caulolatilus sp.		0.339
Pomatomidae		
Pomatomus saltatrix, bluefish	0.137	1.781
Carangidae		
Seriola dumerili, greater amberjack	0.062	0.085
Sparidae		
Stenotomus chrysops, scup	0.009	0.509

	AS/NHB	SS
Taxa	84.42 h, 63–126 m	16.48 h, 40–126 m
Labridae		
Tautoga onitis, tautog	0.008	
Tautogolabrus adspersus, cunner	0.729	0.254
Labridae (unidentified)	0.002	
Caproidae		
Antigonia capros, deepbody boarfish	4.776	21.628
Paralichthyidae		
Paralichthys dentatus, summer flounder	0.002	0.085
Paralichthys oblongus, fourspot flounder	0.004	0.085
Bothidae (unidentified), lefteye flounders		0.254
Pleuronectidae		
Glyptocephalus cynoglossus, witch flounder	0.002	
$Hippoglossoides\ platessoides, American\ plaice$	0.002	
Cynoglossidae		
Symphurus stigmosus, blotchfin tonguefish		0.085
Molidae		
Mola mola, ocean sunfish	0.002	
Unidentified fishes	0.096	0.339

Chain dogfish were less abundant on the natural hard bottom (site NHB) than on the shipwreck (AS) or SS habitats. They occurred in massive numbers at the shipwreck sites, where they were often so densely packed that they lay on top of each other in layers that were several individuals thick (Fig. 2B). Individuals and aggregations of individuals were observed on all areas of the shipwreck sites, including on and within the trawl nets that covered sections of the shipwrecks. Although many chain dogfish were observed lying on soft bottom, they did so generally within tens of meters of the shipwrecks. The aggregations of chain dogfish probably reflect activity related to spawning because thousands of their egg cases were attached to the shipwreck structures and the nets that covered them (Fig. 2B).

Because shelf communities are subjected to seasonal environmental variability and may exhibit seasonal distribution patterns, multivariate analysis was used to examine seasonal differences (fall 2012 versus spring 2013) in fish distributions at the 4 study sites that were sampled during both seasons; 17 video samples (8 for fall, 9 for spring) and 29 species were examined in this analysis. Season did not have a significant impact on fish assemblages (ANOSIM, coefficient of multiple correlation [R] = -0.024, P = 0.55). Likewise, there were no differences in assemblage structure over the limited depth range examined (8 sites, depths of 68–126 m, dive NF-21 at site SS excluded, R=0.026, P=0.400). The greatest distances between sites were no more than 50 km, and fish assemblages (excluding observations from dive NF-21) at the 8 sites were not significantly different (R=0.130, P=0.090) in regard to distance from one another or distance from Norfolk Canyon. Therefore, all data were combined for analysis of habitat influence on fish assemblages.

Multivariate analysis of 26 video samples (excluding those from the shallow dive NF-21) and 41 taxa, indicated a significant difference (R=0.499, P=0.001) in fish assemblage structure between the soft bottom (SS) and hard bottom (AS/NHB) habitat types (Fig. 3). The video samples from the AS/NHB habitat type were 60% dissimilar from the 2 sample groups affiliated with the SS habitat type; the video sample associated with natural hard-bottom habitat (at site NHB) grouped with the shipwreck hard-bottom (AS) samples (dive number 28, Fig. 3). The fishes most influencing the group of samples from the AS/NHB habitat type (on the basis of SIMPER analysis) were the chain dogfish, members of Anthiinae, the yellowfin bass, the deepbody boarfish, the conger eel (Conger oceanicus), Scorpaena sp., the red barbier, the cunner (Tautogolabrus adspersus), and the blueline tilefish (Caulolatilus microps). Fishes most influencing the groups of samples from the SS habitat type were the chain dogfish, Scorpaena sp., the deepbody boarfish, and the black sea bass.

Data indicates at least some difference in fish communities along isobaths. Within the hard bottom group, video samples (n=9) from the 4 deeper shipwreck habitats (AS) north of Norfolk Canyon (depths of 91-126 m; Fig. 1) grouped closely together (Fig. 3), although data were collected in 2 different years and seasons. The 3 shipwreck hard-bottom (AS) video samples from the shallower middle shelf (depths of 68-85 m; Fig. 1) were offset together in the overall group of samples from the AS/NHB habitat type (dive numbers 22, 29, 30; Fig. 3). The 3 samples from sandy bottom on the middle shelf (SS habitat type, dive numbers 22, 29, 30; Fig. 3) also

were set apart from most other SS samples (Fig. 3). Fishes that occurred in deeper waters (>90 m) that were missing from those 3 middle shelf sites, regardless of habitat type, were the metallic codling (*Physiculus fulvus*), the big roughy (*Gephyroberyx darwinii*), the longspine snipefish (*Macroramphosus scolopax*), groupers (*Hyporthodus* spp.), and the deepbody boarfish. Two species common on the middle shelf sites, the scup (*Stenotomus chrysops*) and the tautog (*Tautoga onitis*), were not observed on the deeper sites.

Habitat preference also was indicated by relative abundance patterns of most benthic fishes. The hard bottom (AS/NHB) habitat type contained ≥89% of the abundance of each of 22 fish taxa (i.e., the first 22 species in Fig. 4), and far more individuals were observed overall in that habitat type than in the SS habitat type. Several species, including a searobin (*Prionotus* sp.), the fourspot flounder (Paralichthys oblongus), the summer flounder (P. dentatus), and the scup, used both soft- and hard-bottom habitats frequently. A few taxa, including lefteye flounders (Bothidae), the margined snake eel (Ophichthus cruentifer), and the rosette skate (Leucoraja garmani), were observed only on soft bottom (Figs. 2F and 4).

Discussion

Fishes that occupied natural and artificial hard-bottom habitats on the middle to outer shelf of the MAB exhibited an assemblage structure in our study that was different from that of the well-documented (e.g., Murawski et al., 1983; Mahon et al., 1998) ichthyofauna of MAB soft-bottom habitats. Although the most abundant reef (i.e., hard-bottom) species also were observed and counted on soft-bottom habitat, in most cases they were never far from reef structures. The hard-bottom habitats surveyed in our study were dominated by cooltemperate and warm-temperate species that are generally considered to be reef associates, although some (e.g., the chain dogfish, black sea bass, and conger eel) have broad depth and latitudinal distributions and large-scale habitat use. Other species common to the hard-bottom habitats (e.g., most Serranidae, the tautog, the blueline tilefish, and the deepbody boarfish) exhibited more restricted distributions and tighter association with reefs. Species of Serranidae (excluding the black sea bass) in particular seemed constrained to a relatively narrow depth range (from ~70 m to at least 150 m) in the MAB, most likely because of the generally warmer (>10°C) and less variable bottom water temperatures along the outer shelf of the southern MAB (Colvocoresses and Musick, 1984).

In contrast to most of the fishes associated with

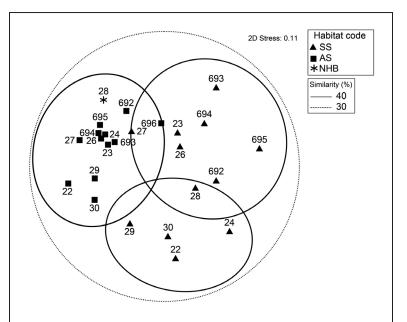
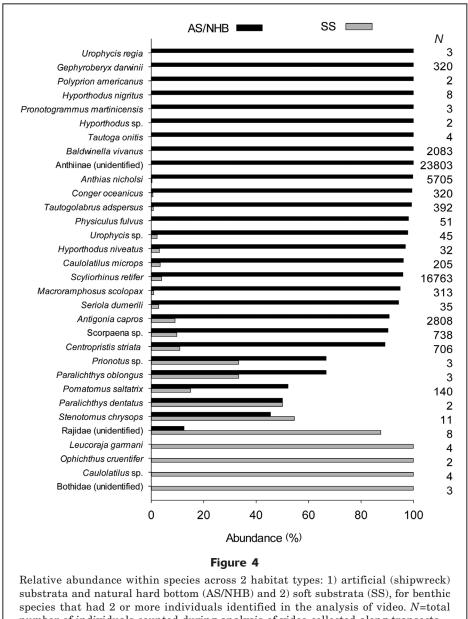


Figure 3

Multidimensional scaling ordination of 26 video samples from habitats of artificial (shipwreck) substrata (AS), natural hard bottom (NHB), and soft substrata (SS), based on the Bray-Curtis similarity matrix calculated from standardized, fourth-root transformed fish abundances (41 taxa). Numbers by symbols are the dive numbers given to each dive of the remotely operated vehicles (see Table 1).

soft-bottom habitats, several of the abundant hardbottom species (Fig. 2, A-E; e.g., the yellowfin bass, the red barbier, groupers, and the blueline tilefish) were further constrained by being at or near the northern limits of their adult ranges (Moore et al., 2003; Anderson and Heemstra, 2012). Because many of the common hard-bottom species (e.g., most of the Serranidae and the chain dogfish and deepbody boarfish) likely have an obligate association with reef-like habitats (Able and Flescher, 1991; Craig et al., 2011; Anderson and Heemstra, 2012), the relatively limited extent of hard bottom in the MAB (Steimle and Zetlin, 2000) would also affect their distribution. Therefore, an abundant component of the hard-bottom fish community in the southern MAB is restricted by habitat availability, depth, and zoogeography, of which the latter 2 constraints probably are related to bottom temperature. Although this reef community of the outer shelf appears to flourish, these limitations likely make it vulnerable to overfishing, habitat damage, and large-scale environmental variations.

Results presented here differ substantially from those of other surveys of the MAB. Our ROV study sites overlapped with some of the areas of fish groupings that were based on decades of bottom trawl surveys (Colvocoresses and Musick, 1984; Mahon et al., 1998); however, the majority of species common to this study and these 2 trawl-based studies were species that are reported to be most common on sandy bottoms



number of individuals counted during analysis of video collected along transects.

(e.g., goosefish [Lophius americanus], spotted hake [*Urophycis regia*], fourspot flounder, summer flounder). The most abundant taxa observed in our study (i.e., the chain dogfish, Anthiinae, and the deepbody boarfish), as well as others known to be reef associates (e.g., groupers, the blueline tilefish, and wrasses [Labridae]), were not abundant or were not reported in those earlier trawl-based studies (see also Grosslein and Azarovitz, 1982). The differences in species composition, largely resulting from sampling constraints imposed by trawls, emphasize the high degree of separation between fish communities on soft-bottom and those on reef-like habitats in the MAB.

Although Grimes et al. (1986) and Ross et al. (2015)

conducted visual surveys that covered extensive complex habitats in the region, they reported only 2 (25% overlap) and 10 (12% overlap) fish species, respectively, in common with those observed in our study. In those 2 studies, the faunal differences can be attributed to deeper waters or a sampling areas much farther north than those surveyed in our study. Although also completed farther north (~ 41°N), visual surveys (Auster et al., 1995) conducted over flat, primarily sand and shell bottom (sites at depths of 55, 240, 712 m) yielded 37.5% fish species in common with our study, and most of those species exhibited broad habitat affinity or affinity for soft bottom. In contrast, the MAB hard-bottom habitats surveyed in our study shared 43% of the fish fauna with a deep (depths of 237–253 m) shipwreck off Cape Fear, North Carolina, that was assessed during one earlier ROV dive (Quattrini and Ross, 2006). Many (~30%) of the reef-associated fishes reported here are common on outer shelf hard grounds throughout the southeastern United States (Grimes et al., 1982; Quattrini and Ross, 2006), also indicating a warm-temperate affinity for reef fishes of the southern MAB.

As with the reefs of the middle to outer shelf, hardbottom habitats of the shallower inner shelf of the MAB were dominated by relatively few, but different, fish species (i.e., the black sea bass, tautog, cunner, and scup). On the deeper hard-bottom habitats surveyed in this study, the black sea bass, tautog, and scup ranked below the top 6 species in abundance. Previous shelf studies were conducted in much shallower water (depths <35 m) (Briggs, 1975; Feigenbaum et al., 1985; Eklund and Targett, 1991; Adams, 1993), and not many small species were caught in 2 studies that relied heavily on data from fish traps. However, taxa like the Anthiinae, most common on deeper reefs (Anderson and Heemstra, 2012), are unlikely to occur on inshore reefs. Although regionally limited in general, hard-bottom habitats and associated data that can be recorded there are even more rare along the outer shelf (i.e., ~100-m depth zone; Steimle and Zetlin, 2000). Despite the restricted scope of these deep shelf reefs, they support economically important fishes (e.g., groupers, tilefishes, and black sea bass) and exhibit a high species richness of fishes, as do reefs at similar depths south of Cape Hatteras (Parker and Ross, 1986; Quattrini and Ross, 2006).

A degree of faunal stability along the outer shelf of the southern MAB is indicated by similarities between years or seasons for the 4 study sites sampled in both seasons and both years. Although many fish species shift distributions by season in the MAB (Murawski et al., 1983), such movements may be less pronounced in deeper waters. A relatively small (~2°C) variation in bottom temperatures along the outer shelf (depth ~100 m) was correlated with consistent groupings of softbottom species across seasons and years (Colvocoresses and Musick, 1984). Grimes et al. (1986) noted that the region from southern New England to the MAB was often occupied by a warmer (9-14°C) bottom water mass from depths of about 100-300 m. In contrast, there were distinct seasonal differences in fish communities correlated with temperature, which varied over a range of 16.7°C, on an artificial reef in much shallower water, at a depth of 21 m, off Virginia (Adams, 1993). For the deeper shipwreck sites that we sampled during 2 seasons, only a mean bottom temperature difference <1.5°C was observed between the 2 survey periods. Although more continuous and long-term environmental data are needed to capture more accurate means and especially variability, our results agree with the larger data set from Colvocoresses and Musick (1984).

Colton (1972) noted a series of warming and cooling trends on the shelf in the Gulf of Maine, but there was also little apparent change in the distributions of

4 groundfish species correlated with these temperature shifts. Because obligate reef fishes usually exhibit strong site or area fidelity, as long as bottom temperatures remain within tolerances, much of the reef fish community (e.g., Anthiinae and *Hyporthodus* spp.) on the outer shelf of the MAB should continue to occupy these hard-bottom sites. However, episodic intrusions of cold water from the north or from the deep sea can jeopardize some species of the reef fish community of the MAB outer shelf and could cause mass mortalities as documented for tilefish (Marsh et al., 1999).

As previously suggested (Murawski et al., 1983; Nye et al., 2009; Møller et al., 2010), it is tempting to propose that hard-bottom habitats of the southern MAB are increasingly invaded by more warm-temperate species, possibly in response to rising ocean temperatures. North Carolina is the closest southern source where many of the species noted here are abundant on extensive outer shelf hard-bottom habitats (Grimes et al., 1982; Parker and Mays, 1998; Quattrini and Ross, 2006). Although Cape Henry, Virginia, was listed with question as the northern limit of blueline tilefish (Dooley, 1978), our observations confirm its presence in the MAB (Fig. 2E) and extend its range north of Norfolk Canyon. That species, and the yellowfin bass, had been reported from this region from the early 20th century (Firth, 1933, 1937). Snowy grouper (Hyporthodus niveatus) and warsaw grouper (H. nigritus) (Fig. 2, C and D) were reported in New England waters as early as the late 19th century, but in most of these cases the fish were juveniles collected inshore and assumed to be strays (Smith, 1971). Large adults (documented in world tackle records) of snowy grouper recently occurred in the MAB recreational hook-and-line fishery (as did blueline tilefish), but data presented here are the first descriptions of their relative abundance and adult habitat along the outer shelf of the MAB. Recent collections of red barbier near Wilmington Canyon represent the first records of that species for the MAB (Moore et al., 2003), but this small, deep-reef-specific fish could have easily escaped detection. Similarly, our observations of 3 individuals of roughtongue bass (Pronotogrammus martinicensis) at depths of 92 m on the natural hard bottom represent a new northern range limit (from North Carolina, Anderson and Heemstra, 2012) for this species, but this observation does not necessarily mean that this species is newly arrived to the MAB. Although historical data have been inadequate (because of a lack of appropriate sampling on deep reefs) to allow an evaluation of long-term changes in the patterns of hard-bottom species composition on the MAB outer shelf, this study, the first to examine outer shelf reef fishes of this region, should provide a baseline for future assessments.

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