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Miniature baited remote underwater video (mini-BRUV) reveals the response of cryptic fishes to seagrass cover

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Abstract

Seagrass habitats worldwide are degrading and becoming fragmented, threatening the important ecosystem services they provide. Fauna associated with seagrasses, particularly cryptic species, are expected to respond to these changes, but are difficult to detect at ecologically meaningful scales using non-extractive techniques. We used a small, wide-angle camera (GoPro) and a small quantity of bait positioned within the canopy of Posidonia australis meadows in Jervis Bay, New South Wales to assess the response of fishes to seagrass cover. We saw a clear positive relationship with the condition of P. australis; a high cover of this seagrass had positive effects on the diversity and abundance of cryptic fauna. Our findings highlight ecosystem shifts associated with the loss and fragmentation of biogenic habitat. These changes are of particular relevance for P. australis meadows given their current status as an endangered ecological community in several locations in NSW and their slow rate of recovery from disturbance.

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23	Running Head: Cryptic fishes respond to seagrass cover

26 Abstract

27 Seagrass habitats worldwide are degrading and becoming fragmented, threatening the 28 important ecosystem services they provide. Fauna associated with seagrasses, 29 particularly cryptic species, are expected to respond to these changes, but are difficult 30 to detect at ecologically meaningful scales using non-extractive techniques. We used a 31 small, wide-angle camera (GoPro) and a small quantity of bait positioned within the 32 canopy of *Posidonia australis* meadows in Jervis Bay, New South Wales to assess the 33 response of fishes to seagrass cover. We saw a clear positive relationship with the 34 condition of P. australis; a high cover of this seagrass had positive effects on the 35 diversity and abundance of cryptic fauna. Our findings highlight ecosystem shifts 36 associated with the loss and fragmentation of biogenic habitat. These changes are of 37 particular relevance for P. australis meadows given their current status as an 38 endangered ecological community in several locations in NSW and their slow rate of 39 recovery from disturbance.

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41 Keywords: fish assemblages; seagrass; *Posidonia australis;* sampling method; cryptic
42 species; GoPro camera

44 Introduction

45 Seagrass meadows are valuable ecosystems, playing important roles in nutrient 46 recycling, carbon sequestration, trophic transfers, and providing nurseries for a large 47 range of fish and other taxa including key economic species (Duarte 2002; Fourqurean 48 et al. 2012). Despite the value of this habitat, seagrasses are suffering loss and 49 degradation at an alarming rate worldwide and the rate of loss appears to be accelerating 50 (Waycott et al. 2009). In NSW, approximately 85% of estuarine seagrass beds have been 51 lost (RJ West pers. comm.) and six populations of Posidonia australis are listed as 52 endangered (DPI 2012). With increasing declines in seagrass worldwide, it is essential 53 to understand how these losses will affect the abundance and diversity of seagrass 54 associated fauna.

55 Seagrass meadows are affected by an array of human activities. Declines in 56 seagrass cover have been linked to moorings (Demers et al. 2013), eutrophication 57 (Cardoso et al. 2004), anchoring (Okudan et al. 2011) and dredging (Badalamenti et al. 58 2011). Anthropogenic activities can impact upon a variety of seagrass characteristics 59 that may have important implications for fish assemblages. These characteristics may 60 be classed as either 'landscape' or 'structural' and will have different effects depending 61 on the species and trophic group in question (Hovel et al. 2002). Landscape 62 characteristics include large-scale attributes such as size, number, shape and perimeter-63 area ratio (i.e. proportion of edge) of seagrass beds in meadows (Connolly and Hindell 64 2006), whilst structural characteristics include small-scale local attributes such as 65 seagrass cover, shoot density, epiphyte load and canopy height (Jelbart et al. 2007).

Seagrass structural characteristics are less often used to assess patterns of seagrass
fish assemblages due to the difficulty of sampling mobile fauna at small spatial scales
(Connolly and Hindell 2006). Instead, most studies assess patterns of infaunal and
epifaunal assemblages such as shrimp, gastropods and crabs due to their small size and

low mobility (e.g. Edgar 1992; Johnson and Heck 2006). Nevertheless, significant
relationships have been found between fish assemblages and the structural
characteristics of seagrass such as shoot density (Gullstrom et al. 2008), seagrass cover
(Hovel et al. 2002) and epiphyte load (Jackson et al. 2006).

Here, we use a small BRUV unit, deemed 'mini-BRUV' and based on GoPro technology, to sample fish assemblages within the seagrass canopy in areas ranging from degraded meadows to some of the most pristine *Posidonia australis* meadows in southeastern Australia. As cryptic species often rely on this habitat for protection from predators (Heck and Orth 1983) we reasoned that the loss of seagrass cover would affect their diversity and abundance.

80

81 Materials and Methods

82 Study locations and sampling

83 We sampled in Jervis Bay, a large, marine-dominated embayment located in 84 southeast Australia (35°08' S, 150°45' E) (Fig. 1). We selected five sample locations 85 (Hare Bay, Green Point, Long Beach, Hole in the Wall and Callala Bay) based on the 86 presence of Posidonia australis, an endemic Australian seagrass found in large 87 expanses throughout Jervis Bay. Some of these areas have suffered seagrass loss and 88 degradation from anchoring and mooring activities (Demers et al. 2013, author's pers. 89 obs.) All sampling occurred in shallow seagrass beds between 1.0-4.0 m depths. 90 Sampling was done over a 6-week period during October-November 2012.

We developed a miniaturized baited remote underwater video (BRUV) system
and placed it within the seagrass canopy. The mini-BRUV consisted of a GoPro (Hero
2) camera affixed to a standard clay brick (23 x 11 x 7 cm) and attached to a steel grid
(45 x 30 cm), which served to anchor the unit and flatten the seagrass (Fig. 1). Attached
to the grid and opposite the brick was a bait bag (18 x 8 x 6 cm) made of plastic mesh

96 containing a single pilchard (*Sardinops sagax*) weighing approximately 45 g and cut 97 into 4 pieces. Pilchards were replaced before each deployment. A rope and buoy were 98 attached to the unit and used to relocate the equipment. Each GoPro camera was set to 99 the highest possible resolution (Resolution/FPS: 1080-30p) and all LED lights were 100 deactivated. A GoPro flat lens was used on cameras to reduce distortion in water.

101 At each location, mini-BRUVs were deployed by snorkelers at 3 sites separated 102 by a minimum of 100 m. At each site, 4 mini-BRUV units were simultaneously 103 deployed and placed 20-30 m apart to minimize any overlap of bait plumes. Units were 104 placed at a minimum of 20 m from the seagrass-sand interface and 50 m from nearby 105 reef habitats to prevent sampling fishes associated with bare sand or reef environments. 106 As mini-BRUV units had a restricted field of view and minimal bait was used we were 107 not certain of the number of fish each mini-BRUV would attract. Hence, the abundance 108 and species richness recorded on the 4 mini BRUVs at each site were considered a 109 single deployment and were pooled to form a single replicate (but see section 'Video 110 and statistical analyses'). Mini-BRUVs were left to record for 35 minutes to allow for 111 30 minutes sampling time. All video sampling was carried out between 0800-1600 hrs 112 to avoid reduced visibility outside of these hours.

We estimated the cover of seagrass using point counts. Two 10m line transects were haphazardly placed within 5 m of each mini-BRUV unit. We recorded the presence or absence of seagrass every 0.5 m giving a total of 160 points for each replicate (8 transects per mini-BRUV replicate). Seagrass presence included any live material of seagrass including shoot, flower or rhizome. Seagrass cover was only quantified after mini-BRUVs had completed filming to prevent any effects of snorkeler presence on fish. Counts were then converted to percent cover.

120 Video and statistical analyses

Video analysis commenced 3 min after Bottom Time (BT), which was the time when the BRUV unit settled on the sea floor. This allowed fish to recover from the presence of snorkelers. Videos were analyzed for exactly 30 min from BT and all fish and crab individuals observed during this time were recorded.

At the time of sampling, this was the first study in which mini-BRUVs had been used to assess cryptic fish in seagrass environments and combined with a restricted field of view and minimal bait, we were uncertain of the number of fish (particularly cryptic fish) each mini-BRUV would attract. Hence, the species richness recorded on the 4 mini-BRUVs at each site were considered a single deployment and were pooled to form a single replicate.

Fish abundance was estimated with *max* N, the maximum number of individuals of a species observed in a single frame throughout the 30 min video. This prevented recounting the same individual and is therefore a conservative measure of relative fish abundance. As mini-BRUV units comprising a replicate were relatively close to each other, we recorded abundance as the highest *max* N observed by a single camera across the 4 units comprising each mini-BRUV replicate.

We then examined the relationship between the cover of seagrass for the 15 replicate mini-BRUV deployments and the diversity and abundance of cryptic and noncryptic fishes with least squares regression (JMP V9). We confirmed that model residuals did not show evidence of significant spatial autocorrelation (Moran's I) using the 'spdep' package in R, before proceeding with these regressions.

Fish were categorized as cryptic and non-cryptic based on information from a variety of sources (Hutchins and Swainston 1986, Cappo et al. 2004, Colton and Swearer 2010) as well as our own observations of species behaviour. Cryptic species were typically small (<10cm TL), shy, slow swimming, sedentary and were usually solitary. Many cryptic species were benthic, traveling amongst seagrass blades or close to the substratum. In contrast, non-cryptic fish (also known as pelagic/transient fish)
typically school, were medium- to large-sized, highly mobile and often swam above the
seagrass canopy or in the water column.

150 **Results**

Our mini-BRUV deployments revealed 252 individuals from 28 species (21
families) in 30 hrs of recordings. We considered 9 of these species to be cryptic (Table
1). In addition, a large benthic crab, *Nectocarcinus integrifrons*, was observed on 73%
of our recordings.

155 The cover of seagrass showed a strong positive effect on and explained a 156 significant amount of the variation in the abundance ($r^2=0.49$, P<0.01) and diversity $(r^2=0.59, P<0.001)$ of cryptic fishes within the *P. australis* meadows of Jervis Bay (Fig. 157 2a and b). In contrast, seagrass cover explained little of the variation in the abundance 158 159 $(r^2=0.03, P>0.05)$ and diversity $(r^2=0.02, P>0.05)$ of non-cryptic fish species (Fig. 2c 160 & d). A large school of juvenile Gerres subfasciatus represented an outlier in the 161 abundance of non-cryptic fishes and it was removed from figure 2. Its exclusion did 162 not change the nature of the relationship or our interpretation.

163

164 **Discussion**

Our use of a miniaturized BRUV unit ('mini-BRUV') placed within the seagrass canopy of *Posidonia australis* proved highly effective at sampling seagrass fishes, particularly small, cryptic fish species not easily detected by methods such as UVC and conventional BRUV. Importantly, mini-BRUVs produced negligible damage to the environment, which is of particular importance in MPAs and sensitive habitats such as seagrass where trawls would be considered inappropriate. The compact size of these mini-BRUVs combined with the use of a small amount of bait, made this an ideal method to examine the relationship between fish assemblages and small-scalestructural attributes such as seagrass cover.

174 The positioning of mini-BRUV units within the seagrass allowed timid cryptic 175 fishes to approach the bait without the risks associated with open water. The use of a 176 small quantity bait may have been advantageous, as the large volumes of bait used in 177 conventional BRUVs often attract schools of fish including sharks (Bernard and Götz 178 2012), this may intimidate smaller species (Klages et al. 2014). Although we observed 179 predatory species around mini-BRUV units, their abundance (max N) remained low and many individuals were observed simply swimming past, rather than being attracted 180 181 to the bait. In addition, we did not observe intimidation of cryptic fishes by the large 182 benthic crab, Nectocarcinus integrifrons observed in our recordings; the diversity and 183 abundance of cryptic fishes was not significantly correlated with crab abundance (r=0.4184 and r=0.01 respectively, df=15).

185 The use of small quantities of bait in mini-BRUVs may also see useful 186 applications in seascape ecology. The bait plume of conventional BRUVs has the 187 potential to spread over hundreds of metres depending on current speed, direction and 188 soak time (i.e. length of time BRUVs are in water) (Taylor et al. 2013). Thus schooling 189 species drawn from a large area perhaps encompassing several habitats may concentrate 190 around a single BRUV unit, thereby misrepresenting the fish assemblage associated 191 with a particular habitat. The small volume of bait used with mini-BRUVs has the 192 potential to resolve this issue although further research is required to determine bait 193 plume size and optimum bait levels associated with mini-BRUVs over small spatial 194 scales.

195 This study underscores the importance of seagrass cover to cryptic fishes. We 196 observed a clear positive relationship in cryptic fish abundance and species richness 197 associated with seagrass cover. As many cryptic taxa are highly specialized and likely 198 occupy small ecological niches, they are expected to be vulnerable to habitat loss. Our 199 findings have important implications for seagrass-dependent species as seagrass 200 meadows continue to degrade worldwide. This calls for increased efforts to protect 201 seagrass meadows to ensure the maintenance of the biodiversity that these key habitats 202 support.

203

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210 Group, University of Wollongong

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212 Compliance with ethical standards

213 Conflict of Interest: The authors declare that they have no conflict of interest.

214 This work was completed with ethics approval from the University of Wollongong

215 (AE12/07r15).

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Table 1: List of fish and crab⁺ species observed across all deployments of mini BRUV
units in *Posidonia australis* meadows in Jervis Bay, Australia. Total abundance (*max*N) was categorised into the following; *=1-5; **=6-10; ***=11-15; and ****>16.
Species underlined represent cryptic organisms (Hutchins and Swainston 1986; Cappo
et al. 2004; Colton and Swearer 2010).

2	n	1
Ζ	9	Τ

Species	Fish Abundance (Mini BRU⊉92	
Acanthopagrus australis	**	
Aldrichetta forsteri	*	
Arripis trutta	*	293
Atypichthys strigatus	*	
Brachaluteres jacksonianus	*	
Dasyatis brevicaudata	*	294
Dinolestes lewini	**	
Enoplosus armatus	*	295
Gerres subfasciatus	****	293
Girella tricuspidata	****	
Haletta semifasciata	***	
Heteroclinus tristis	*	
Meuschenia freycineti	**	
Mugil cephalus	*	
Myliobatis australis	*	
Nelusetta ayraudi	***	
Neoodax balteatus	***	
Pagrus auratus	*	
Pelates sexlineatus	****	
Pseudocaranx dentex	****	
<u>Scobinichthys granulatus</u>	**	
Sillago ciliata	*	
Sphyraena novaehollandiae	**	
Torquigener pleurogramma	*	
Trachurus novaezelandiae	*	
Trygonorrhina fasciata	**	
Upeneichthys lineatus	*	
<u>Nectocarcinus integrifrons</u> ⁺	****	

296 Figures Captions

297	Fig. 1: Mini BRUV unit in the seagrass <i>Posidonia australis</i> at Jervis Bay, Australia.
298	The unit consists of; 1) GoPro Hero 2 camera 2) brick (weight) 3) grid and 4) bait bag.
299	
300	Fig. 2: Regression plots for total abundance (max N) and species richness of a) and b)
301	cryptic fish and; c) and d) non-cryptic fish against the cover of seagrass P. australis
302	within Jervis Bay, Australia.







