

Examensarbete i ämnet biologi

The effect of a natural forest fire on beetle assemblages in the boreal forest of Sweden

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Abstract

The aim of this study was to determine the effect of a large natural forest fire on the beetle community. Large, naturally-occurring forest fires rarely occur in north Sweden and the ecological effect on these events has therefore seldom been studied. It is, however, important to gain knowledge of natural forest fires to be able to mimic these events in conservation measures and fire management. I studied the changes in abundance, species richness and assemblage composition of beetles following the fire in 2007. I performed detailed analyses on four different groups of beetles: saproxylic beetles, beetles favoured or strongly favoured by fire, and cambium consuming beetles. In the latter group, many of the typical pest species are found. 18 traps where placed in the fire field and the same number of traps where placed in a control area close to the fire field. Data from this survey where later analysed with Wilcoxons rank sum test and PERMANOVA. The total number of species found was 335 consisting of 8586 individuals. In the fire field I caught 239 species consisting of 5507 individuals while I in the control area caught 221 species consisting of 3079 individuals. I found a significant increase in the abundance and changes in the species composition of beetles following fire. In contrast, I did not find clear evidence that the species richness was higher in the burned areas. However, my study, in agreement with earlier studies, showed that the forest fires have a strong effect on the species composition of beetles. Further research on the effects of large scale forest fire in landscapes with a long history of fire suppression is important for the development of sustainable management plans for species favoured by and adapted to fire.

Keywords: Beetles, boreal forest, forest fire, decaying wood, saproxylic species, forest management

Introduction

Fire is an important disturbance in the boreal forest landscape (Niklasson and Granström, 2000). A forest fire releases nutrients and creates large amounts of dead wood from dead and dying trees which is important substrate for many fungi and beetle species. The long term effects of fire are increased structural diversity and changes in tree species assemblage which is important for the long term survival of species adapted to early succession stages and dead wood continuity on the landscape level (Wikars and Ås, 1999). Additionally about 40 species of insects and approximately 50 species of fungi are directly dependent on scorched wood or ground, while another 100 species are favoured by fire (Swedish Environmental Protection Agency, 2005).

Structural diversity in the boreal forest landscape created by forest fires has been dramatically reduced due to intensive forest management focused on high production (Schimmel and Granström, 1991). Fire suppression has effectively reduced the number of fires (Lundberg, 1984; Wikars, 1992; Esseen et al., 1997; Swedish Environmental Protection Agency, 2006) and forestry measures like clear cutting, soil scarification, planting and thinning has replaced diverse natural forests with dense, even aged stands (Linder and Östlund 1998). Historically, a general decrease in forest fires can be seen from around 1870 (Niklasson and Granström, 2000). However, global warming is expected to result in increased frequency of forest fires in the near future because of a drier climate (IPPC, 2007). Today less than 0.016% of Sweden's forest area is burned annually, while approximately 1% (~200 000 hectares) was burned annually 150 years ago (Wikars, 1992; Swedish Environmental Protection Agency, 2005). Nordic countries are now facing a situation where areas with a natural fire regime are scarce (Esseen et al., 1997). Many organisms dependent on burned forest have decreased in population size and some species were close to extinction in the early 1980's (Lundberg, 1984).

Although it appears intuitively that clear cutting and forest fires gives similar ecological effects there are clear differences, e.g. biomass in the form of trees are taken away after clear cutting while forest fires leave most of the trees standing or lying down hence creating large amounts of substrate for saproxylic (wood-living) organisms (Wikars and Ås, 1999). Furthermore, because not all trees die at once, fire creates substrate over a long time (Wikars, 1997). Another difference is that fires generally increases structural heterogeneity while clear cutting generally decreases structural diversity and the amount of available substrate (Wikars and Ås, 1999). To reduce the negative impact from clear cutting it is possible to mimic the tree species assemblage among the remaining trees after clear cutting (Wikars and Ås, 1999) and also to leave and create dead wood in the managed forest has shown to be useful (Toivanen and Kotiaho, 2007). However, many effects from forest fires, e.g. reduction of the field layer and production of large amounts of dead wood creating opportunities for less competitive species to invade before opportunist species recolonize the area, can not be completely mimicked by conservation oriented measures in clear cuts (Wikars and Ås, 1999). Another method for mimicking natural fires is prescribed burning where old growth forest as well as clear cuts can be used. During recent years an increased number of prescribed burnings has been performed to fulfil the certification demands (PEFC and FSC) in order to improve the situation for fire favoured and fire dependent species. Decreased forestry activity in naturally, fire damaged forest might also be a reason for this positive trend (Swedish Environmental Protection Agency, 2006).

To better understand the effects of fire on saproxylic insects assemblages it is important to study the assemblage composition and succession of insects after such a rare event as a large scale boreal forest fire. The large scale effect of naturally-occurring forest fires on beetles is, because of the scarcity of the event, poorly studied up until now. In this study I have had the opportunity to study the effects of such an event. Because it is well known that fire benefits various saproxylic and fire favoured species of beetles, I used the data to evaluate whether the abundance and species richness is in fact increased by natural fire. I also chose to analyse possible alterations in the assemblage composition of beetles. It is also well known that some pest species benefits and might mass reproduce during these occasions and I therefore analysed the same factors in this group of beetles.

I specifically asked the following questions:

- 1. What is the effect of fire on the abundance and species richness of saproxylic, fire favoured, strongly fire favoured and cambium consuming beetles?
- 2. What is the effect of fire on the assemblage composition of these four groups?

Material and methods

Study area and experimental design

The study area was situated in Bodträskfors, Boden municipality, in northern Sweden (approx. 66° 9' N, 20° 49' E) and constitutes a 1700 ha burned area which is a result of a natural forest fire in 2006. It is situated on and around Stora Klusåberget (Fig. 1). The fire was severe and most of the field layer was burned as well as higher parts with boulders and flat rocks. It has been decided that parts of the fire field owned by the Swedish forest company Sveaskog (approximately 250 hectares) will become a nature reserve.

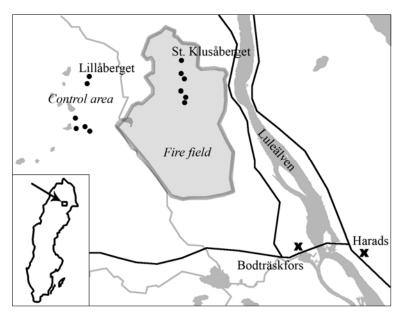


Fig. 1. Map showing study areas (black dots), location of the fire field and closest localities. Location of the section is presented in the inset (lover left corner).

The dominant tree species within the area was Scots pine although some stands contained considerable amounts of deciduous trees like goat willow, aspen and birch. I chose six areas within the fire field for sampling of insects. To the west of the fire field the same number of areas was chosen as controls. In each area I placed three window traps making up a total of 18 (6 x 3 traps) traps in each treatment. All stands included in the study were dominated by Scots pine with ages varying between 140-170 years. The controls were chosen to correspond to the circumstances in the burned area in terms of tree species composition, stand age, soil type, and direction of slope.

I collected insects between 31 May and 21 September 2007 by placing three IBL window traps in each of the twelve areas. IBL-traps were made out of a thin, semi transparent plastic sheet tightened between three thicker plastic parts, making a triangular, window-like flight intercept with a surface of approximately 0.35 m². The sides, made of thicker plastic fabric, constituted a channel for the trapped insects to slide down rather than falling to the ground. At the bottom a special, two level funnel collected the sliding insects. The funnel hade a special mechanism making rain water to go through a pipe on the side to prevent the jar from overflowing. The funnel was connected to a 725 cc collective jar by screw threads (Fig. 2).

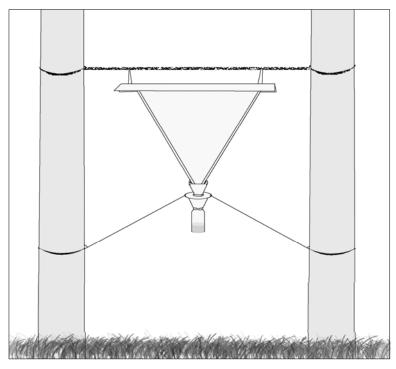


Fig. 2. The IBL-trap attached with threads between two trees. Funnels and collective jar is hanging under the triangular plastic sheet.

I tied the traps between appropriate trees with threads and placed them in a triangular pattern to cover different angels in the landscape. Collective jars were filled up to 1/3 with water, glycol and detergent to preserve the trapped insects.

Identification and classification of beetles

The entire material consisting of 8587 individuals was determined to species level during the winter 2007-2008 by S. Lundberg (Luleå, Sweden). Classification into functional groups was made using the wood decomposer database (Dahlberg and Stokland 2004), literature e.g. (Saalas 1923, Palm 1948) and gathered knowledge by R. Pettersson (Swedish University of Agricultural Sciences, Umeå, Sweden), J. Hilszczanski (Forest Research Institute, Warsaw, Poland) and S. Lundberg (Luleå, Sweden). I chose to include the following functional groups in the study: saproxylic beetles (species that are dependent or favoured by decaying wood meaning both obligate and facultative species), fire favoured (species favoured by fire through making suitable substrate available), strongly fire favoured (species strongly favoured by fire) and cambium consumers (phloem and cambium consumers and consumers of cortex on living trees). Overlap occurs between groups.

Statistical analysis

I analysed the effect of treatment (fire) on abundance and species richness in the different groups, saproxylic, fire favoured, fire dependent and cambium consuming beetles with Wilcoxon rank sum test in R version 2.6.0 (The R Foundation for Statistical Computing) because the data did not meet the assumptions of any parametric test method. Data from each area was pooled without transformation before analysis making 6 pooled samples for each treatment.

Species assemblage was analysed with Permutational Multivariate Analysis of Variance (PERMANOVA) (Anderson, 2003). In contrast to traditional MANOVA tests, permutation based tests requires no specific assumptions concerning correlation. Data might also contain a high number of zeroes due to the occurrence of rare species which is difficult to handle in traditional tests. In addition, the test requires no assumptions concerning correlation between the variables and the number of variables or their individual distributions (Anderson, 2001). In this case, the data did contain a large number of zeroes. I used 4999 permutations, the Bray-Curtis similarity index, and data was fourth-root transformed to reduce the impact of species with high abundance. Area was treated as a nested factor within the treatments. To present the species dissimilarities from the PERMANOVA analysis I plotted the data with non-metric multidimensional scaling (nMDS) (Clarke, 1993). To analyse which species had the greatest impact on the difference in species assemblages, I performed similarity percentages analysis (SIMPER). This is made by exploring how much the overall percentage contribution each species adds to the average dissimilarity between two groups. The species are then listed in decreasing order of importance (Clarke and Gorley, 2001) in distinguishing the two groups from the fire field and the control.

Results

Abundance and species richness

The total number of beetle species recorded was 335, consisting of 8586 individuals. The traps in the fire field caught 239 species consisting of 5507 individuals while the traps in the control caught 221 species consisting of 3079 individuals in total. The abundance was higher in the fire field than in the control area in all the examined groups (Wilcox, P = 0.002, 0.005, 0.026, 0.026 for the groups cambium consumers, strongly fire favoured, fire favoured and saproxylic beetles respectively, see Figure 3). The species mainly responsible for the difference in the abundance per group are presented in Figure 4.

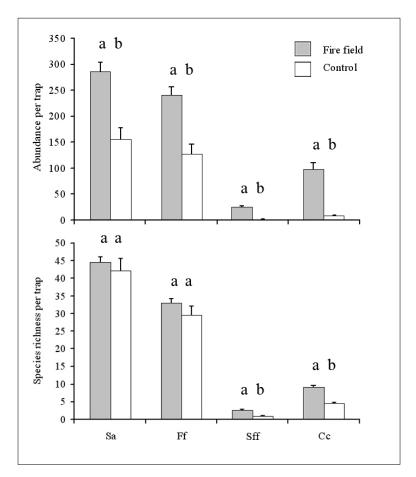


Fig. 3. The abundance and the number of species per trap in the four different categories saproxylic beetles (Sa), fire favoured (Ff), Strongly fire favoured (Sff) and Cambium consuming beetles (Cc) caught during the summer 2007. (mean \pm SE, N=18) Overlaps occur between different groups.

Cambium consumers and strongly fire favoured beetles showed higher species richness in the fire field than in the control area (Wilcox, P = 0.009, 0.004 respectively) while the two remaining groups fire favoured and saproxylic beetles showed no significant difference in species richness (Wilcox, P = 0.688, 0.810 respectively). The total species richness in the group saproxylic beetles was actually slightly higher in the control area than in the fire field (see table 1) while the species richness per trap (mean value of all traps per treatment) was slightly higher in the fire field. This difference in the outcome of the test in comparison to the observed value is probably due to the highly elevated abundance in the fire field which causes the likelihood of trapping each species in each trap to be much higher in the fire field.

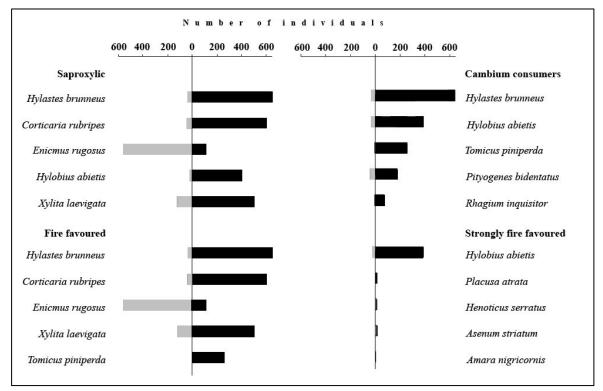


Fig. 4. The abundance per treatment for the five species with the biggest difference in abundance between the fire field and the control per group. Grey represents the control area and black the fire field.

Group	Fire field		Control a	area	Total	
	Species	Abundance	Species	Abundance	Species	Abundance
Saproxylic	158	5148	162	2801	221	7949
Fire favoured	109	4343	109	2283	148	6626
Strongly fire favoured	9	450	5	22	11	472
Cambium consumers	24	1747	19	147	29	1894

Table. 1. Total species richness and abundance per treatment and group. To the right is shown the total species richness and abundance per group.

Species assemblages

The assemblage composition of beetles differed significantly between the fire field and the control area in all the examined groups. No significant effect of area was recorded in the two group's cambium consumers and strongly fire favoured beetles indicating that the variation among the areas was low (table 2, Figure 5). The nMDS figure shows good grouping of all four groups and separated the fire field from the control. Fire favoured and saproxylic beetles showed very similar patterns which might have been an outcome of the strong species overlap. SIMPER analysis showed that the species mainly responsible for the difference in species assemblage between the fire field and the control was for the group cambium consumers *Tomicus piniperda* 14.38%, strongly fire favoured *Hylobius abietis*

	df	MS	F	Р
Saproxylic				
Treatment	1	17671.78	12.56	0.002
Area	5	1406.48	1.52	0.000
Fire favoured				
Treatment	1	18727.76	15.43	0.002
Area	5	1213.68	1.40	0.004
Cambium consumers				
Treatment	1	23250.21	14.72	0.002
Area	5	1579.01	1.36	0.089
Strongly fire favoured				
Treatment	1	17659.83	6.76	0.002
Area	5	2613.80	1.61	0.068

Tab. 2. PERMANOVA shoving the effect of treatment (fire and control) on assemblage of saproxylic, fire favoured, fire dependent and cambium consuming beetles. Error degrees of freedom is 35 for all groups.

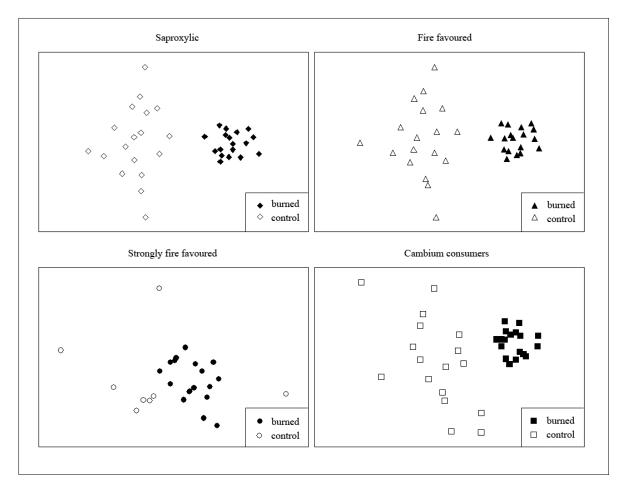


Fig. 5. Result from nMDS ordination. Stress values 0.15, 0.19, 0.09 and 0.15 for saproxylic, fire favoured, strongly fire favoured and cambium consuming beetles respectively

42.80%, fire favoured *T. piniperda* 3.59% and saproxylic beetles *T. piniperda* 2.65% (see table 3 below). Note that the fire favoured species *Scolytus ratzeburgi* was more abundant in the control area than in the fire field.

	Contribution in %	Comparisons
Saproxylic		
Tomicus piniperda	2.65	B > C
Hylobius abietis	2.20	B > C
Atomaria pulchra	2.13	B > C
Corticaria rubripes	2.08	B > C
Hylastes brunneus	1.92	B > C
Fire favoured		
Tomicus piniperda	3.59	B > C
Atomaria pulchra	2.89	B > C
Corticaria rubripes	2.82	B > C
Hylastes brunneus	2.59	B > C
Phloeonomus lapponicus	2.59	B > C
Strongly fire favoured		
Hylobius abietis	42.80	B > C
Henoticus serratus	16.62	B > C
Placusa atrata	12.52	B > C
Asenum striatum	11.77	B > C
Scolytus ratzeburgi	3.98	B < C
Cambium consumers		
Tomicus piniperda	14.38	B > C
Hylobius abietis	11.65	B > C
Rhagium inquisitor	9.86	B > C
Hylastes brunneus	9.48	B > C
Pityogenes bidentatus	8.99	B > C

Tab. 3. The five most important species from the similarity percentages analysis (SIMPER) listed in decreasing order importance.

Discussion

The fire had an enormous effect on the abundance and assemblage composition of beetles in all the examined groups. The total abundance was 79% higher in the fire field than in the control area and the species richness was 9% higher. I will first discuss the impact from the fire on abundance and species richness and later the impact on the assemblage composition of these different groups of beetles.

116 of the total 221 species of saproxylic beetles were more abundant in the unburned control area than in the fire field (see appendix) indicating that there are some saproxylic beetles that react negatively to fire. One reason for this might be that species inhabiting the area before the fire still have not re-colonized, another possible explanation is the fact that some saproxylic species who are also fungivores react negatively to forest fire. Species living on polypores and mushrooms are known to decline after forest fires and rapidly recover when their fungal host re-establish at the site (Muona and Rutanen, 1994). The two group's saproxylic beetles and fire favoured are as previously mentioned overlapping and similar results were expected. The great overlap might be due to a generous definition of fire favoured species where many species generally favoured by disturbance have fallen into the group fire favoured. Hyvärinen et al. (2005) studied the effects of controlled burning on beetle assemblages by using controlled burning. They used different levels of tree retention levels on burned or un-burned plots. In their study they noticed similar patterns as in this study i.e., slightly decreased species richness among saproxylic beetles in their un-harvested, burned plots.

My results are consistent with previous studies (Toivanen and Kotiaho, 2007) where the effects on beetle diversity were studied by using controlled burning and harvesting in combination with the creation of dead wood. Plots were also left un-harvested to mimic natural fire. They concluded was that the use of fire as a tool for conservation increases the overall abundance and species richness of not only pyrophilous (i.e., species strongly attracted to burning or newly burned areas, and species that have their main occurrence in burned forest 0-5 years after the fire (Wikars, 1992)) beetles but also beetle species in general. In the group strongly fire favoured beetles I found two red listed species, Denticollis borealis (NT) which thrives at later succession fire fields (Ehnström et al. 1995) and Tropideres dorsalis (NT), both species connected to birch (Lundberg, 1984). Although the examined area in the fire field was dominated by Scots pine, damp areas dominated by birch might constitute good habitat for these two species. In total six red listed species exclusive for the fire field was found. D. borealis and T. dorsalis were two among those. Both species richness and abundance of strongly fire favoured species was higher in fire field. After excluding the most abundant species, *H. abietis* making 89% of the total group abundance, the abundance was still approximately 9 times higher in the fire field. This confirms previous studies (Muona and Rutanen, 1994) showing that many groups including fire specialist's benefits from forest fire.

The results show that cambium consumers are strongly favoured by forest fire. The high numbers of individuals in the burned area was due to high abundance of the bark beetle *Hylastes brunneus*. However, after exclusion of this species the total abundance of cambium consumers was still approximately 5 times higher in the burned area compared with the control and only 6 species out of the total 29 was more abundant in the control area. Both species richness and the abundance were highly elevated in the fire field. Other studies (Johansson et al., 2007; Wikars, 2002) where logs have been treated with fire and later examined has shown that some fire favoured cambium consumers react negatively to burned wood because the cambium dries out during the fire, making it less appropriate as

food for early succession bark beetles (Saint-Germain et al., 2004). The examined part of the fire field in Bodträskfors contained a high number of weakened, standing trees that constituted good substrate for these species, which might be one reason for the higher abundance of cambium consumers in the burned area compared with the control area. This indicates that mimicking or examining real fire fields is superior to direct treatment of substrate where much of the important, structural diversity created by natural fires is lost (Vanha-Majamaa et al., 2007).

As seen in Figure 4, a few species dominated the material in each group. These species were mainly *H. brunneus, Corticaria rubripes, H. abietis, T. piniperda* and *Enicmus rugosus* which all but *E. rugosus* had a highly elevated abundance in the fire field. In contrast, *E. rugosus* was more abundant in the control area. Similar results have previously been recorded by Toivanen and Kotiaho (2007). It is previously noted that *E. rugosus* is a species which appears later in the succession after fires (Wikars 1997) and might not have reached its peak in abundance in this study because I only sampled insects the first year after fire. 649 individuals of *H. abietis* were found in the traps at the fire field while 39 were found in the control area. *H. abietis* is considered to be one of the worst pest species in Swedish forests (Ehnström and Axelsson, 2002), a fact well worth thinking about while planning prescribed forest fires in boreal forests. However, the impact of pest bark beetle species may be reduced if the fire occurs after swarming (late June) and it has also been said that the problems with some pest species can be avoided or reduced by increasing the intensity of the fire (Wikars, 1992), something that can be achieved during prescribed fires. However these methods are not yet well tested.

The assemblage of beetle species was dramatically altered by the fire. All four groups experienced large changes in the composition of species. As seen in the nMDS ordination plots, there is a clear separation in the grouping of the beetle assemblages between the fire field and the control area (see Figure 5). The species mostly responsible for the difference in assemblage species composition in this study is also among the most abundant. All species tabulated (Table 2) but one, *S. ratzeburgi*, were more abundant in the fire field which might be surprising when at least parts of the fire field constituted good habitat for this species with high density of living or dying birch. However, only two 2 individuals were found in the control area which makes it difficult to draw any firm conclusions concerning their habitat preference. The occurrence in control areas might be due to random events.

Management and conservation of boreal forest

Sweden is committed under the Convention on Biological Diversity (CBD) to protect biodiversity and to use natural resources sustainably, but production is seemingly more important to the Swedish government than nature conservation. The government of Sweden is now preliminarily planning to increase production in the forest (Greenpeace, 2008), introducing new measures with unknown effects on biodiversity. It is thus more important than ever to develop methods to prevent further biodiversity loss in the forest landscape. One important measure is the creation of dead wood through prescribed burnings. Johansson et al. (2007) concluded that burning of logs was insufficient to create the variation needed to sustain the survival of pyrophilious species. I recommend that fires occurring on government property should be left burning under surveillance to produce larger areas with continuously burned forest. It is of course important to keep the threat from pest species in mind during these actions, but, as previously mentioned, forest fires after swarm time may reduce the risk for mass-outbreaks (Wikars, 1992). It is also an aid to preserve as much as possible of the forest damaged by natural occurring forest fires in the future by the creation of reserves. Prescribed fires as a conservation tool should be planned on large clear cuts where substantial amounts of retention trees are left to create sufficient volumes of substrate and a continuity of burned ground and substrate in the landscape. Although prescribed forest fires, should, as far as possible, be planned in reserves to reduce conflicting interests with forestry, it might be an aid to locate some fires in managed forest. Not all reserves may be suitable for prescribed forest fires, and it is also my conclusion that many reserves is located in the northern part of north Sweden and covering a small part of the forest land in Sweden which might create a geographic bias in the distribution of firedamaged forest. Because the best response from fire favoured insects have been observed in northern parts of Sweden (Swedish Environmental Protection Agency, 2006) it might be wise to concentrate these actions to this area. To create new dead-wood and to preserve continuity of adequate substrate in the forest landscape, forest fires play a key role and it is therefore needed to further use and investigate the effects of naturally-occurring fires as well as prescribed fires. This study has given support to suggestions of the importance of natural occurring fires by showing a clear enhancement in the abundance and species richness of species strongly favoured by fire as well as a strong alteration in the assemblage composition of several functional groups of beetles. Further research on the effects of large scale forest fires in a landscape with a long history of fire suppression is important to better be able to develop sustainable management plans for species favoured by and adapted to fire.

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References

- Anderson, M.J. 2001. A new method for non-parametric multivariate analysis of variance. - Austral Ecololgy 26: 32-46.
- Anderson, M.J. 2003. NPMANOVA: A FORTRAN computer program for non-parametric analysis of variance (for any two-factor ANOVA design) using permutation tests. Department of Statistics, University of Auckland, New Zealand.
- Clarke, K.R. and Gorley, R.N. 2001. Primer v5: user manual/tutorial. Primer-E Ltd, Plymoth.
- Dahlberg, A. and Stokland, J. 2004. Vedlevande arters krav på substrat. Rapport 7. Skogsstyrelsen, Jönköping, Sweden.
- Ehnström, B. 2002. Insektsgnag i Bark och Ved. Artdatabanken, SLU.
- Ehnström, B. et al. 1995. Insects in burned forests-forest protection and faunal conservation (preliminary results). Entomologica Fennica 6: 109-117.
- Engelmark, O. et al. 1993. Fire and age structure of Scots pine and Norway spruce in northern Sweden during the past 700 years. New Phytologist 126: 163-168.
- Esseen, P.-A. et al. 1997. Boreal forest. Ecological Bulletins 46: 16-47.

- Flannigan, M.D. et al. 1998. Future wildfire in circum-boreal forests in relation to global warming. Journal of Vegetation Science 9: 469-476.
- Greenpeace. 2008. The Nordic myth State of forest biodiversity in Sweden and Finland. http://www.greenpeace.org
- Hyvärinen, E. et al. 2005. Short-term effects of controlled burning and green-tree retention on beetle (Coleoptera) assemblages in managed boreal forests. – Forest Ecology and Management 212: 315-332.
- IPCC, 2007. Impacts, adaptation and vulnerability. Cambridge University Press.
- Johansson, T. et al. 2007. Variable response of different functional groups of saproxylic beetles to substrate and forest management: Implications for conservation strategies. Forest Ecology and Management. 242: 496-510.
- Linder, P. and Östlund, L. 1998. Structural changes in three mid-boreal Swedish forest landscapes, 1885-1996. –Biological Conservation 85: 9-19.
- Lundberg, S. 1984. Den brända skogens skalbaggsfauna i Sverige. Entomologisk tidskrift. 105: 129-141.
- Muona, J. and Rutanen, I. 1994. The short-term impact of fire on the beetle fauna in boreal coniferous forest. Annales zoologici Fennici 31: 109-121.
- Niklasson, M. and Granström, A. 2000. Numbers and sizes of fires: Long-term spatially explicit fire history in a Swedish boreal landscape. Ecology 81: 1484-1499.
- Palm, T. 1948. Svensk insektfauna (in Swedish). Entomologiska Föreningen i Stockholm, Stockholm.
- Saint-Germain, M. et al. 2004. Xylophagous insect species composition patterns of substratum use on fire killed black spruce in central Quebec. Canadian Journal of Forest Research 34: 677-685.
- Schimmel, J. and Granström, A. 1991. Skogsbränderna och vegetationen. Skog och forskning 4: 39-46.
- Saalas, U. 1923. Die Fichtenkäfer Finlands. II. Spezieller teil und larvenbestimmungstabelle, vol 22, Annales Academiae Scientiarum Fennicae. Series A, Helsinki, Finland.
- Swedish Environmental Protection Agency. 2005. Naturvårdsbränning: Vägledning för brand och bränning i skyddad skog. Rapport 5438. Swedish Environmental Agency.
- Swedish Environmental Protection Agency. 2006. Åtgärdsprogram för bevarande av brandinsekter i boreal skog. Rapport 5610. Swedish Environmental Agency.
- The R Foundation for Statistical Compting. 2007.
- Toivanen, T. and Kotiaho, J.S. 2007. Mimicking natural disturbances of boreal forests: The effects of controlled burning and creating dead wood on beetle diversity. Biodiversity and Conservation 16: 3193-3211.
- Vanha-Majamaa, I. et al. 2007. Rehabilitating boreal forest structure and species composition in Finland trough logging, dead wood creation and fire: The EVO experiment. Forest Ecology and Management 205: 77-88.
- Wikars, L.-O. 1992. Skogsbränder och insekter. Entomologisk tidskrift 113: 1-12.
- Wikars, L.-O. 1997. Brandinsekter I Orsa Finnmark: Biologi, Utbredning och artbevarande. – Entomologisk tidskrift 118: 155-169.
- Wikars, L.-O. 2002. Dependence on fire in wood-living insects: An experiment with burned and unburned spruce and birch logs. Journal of Insect Conservation 6: 1-12.
- Wikars, L.-O. and Ås, S. 1999. Skalbaggarna som följer på branden. Skog och forskning 2: 53-58.
- Zackrisson, O. 1977. Influence of forest fires on the north Swedish boreal forest. Oikos 29: 22-32.

Appendix 1.
All
species
caught :
В.
the fin
e field
1 and
the
species caught in the fire field and the control area.
area.

DD: data deficient. Species written in bold letters where more common in the fire field. pyrophilous, Ca: cambium consumer, D: detritivore, F: Fungivore, P: predator, W: wood borer, NT: near threatened, VU: vulnerable, SxO: obligate saproxylic, SxF: facultative saproxylic, NS: non-saproxylic, FF: fire favoured, SFF: strongly fire favoured, PF:

Species name	Saproxylic group	Fire cathegory	Functional group	Redlist cathegory	Burned	Control
Abdera affinis	SXO	FF	F	Contraction of the local division of the loc		ω
Abdera triguttata	SxO	Ŧ	п		N	
Absidia schoenherri	SxO		70		ω	6
Acanthocinus aedilis		ΡF			4	
Acrotona pusilla	Unknown	Unknown	Unknown	Unknown	-	
Acrotrichis insularis			п			9
Acrotrichis intermedia			T			
Acrulia inflata	SxF		?F			
Agathidium arcticum	SXF		TI			4
Agathidium mandibulare	SxF		Ţ	TN	4	N
Agathidium pallidum	SxF		т	TN		
Agathidium pisanum	SxO	Ŧ	т			8
Agathidium rotundatum	SXF	FF	TI		ω	7
Agathidium seminulum	SxF		п		78	45
Aleochara fumata			dč			N
Aleochara moerens			ζþ			ω
Aleochara sparsa	Unknown	Unknown	Unknown	Unknown	N	
Aleochara stichai	SxF		dč		<u>د</u>	
Aloconota planifrons			γp			
Altica longicollis			т		8	
Amara familiaris	SN				-	
Amara nigricomis	1 100	SFF	H,P		N	
Amischa analis			γp		ω	
Ampedus balteatus	SXO	FF	P,W		13	22
Ampedus nigrinus	SxO	FF	PW		10	48
Ampedus tristis	SxO	FF	P,W		68	υ
Anaspis bohemica	SXO		70			
Anaspis rufilabris	SXO		q			4
Anisotoma axillaris	SxO	FF	П		ω	ω
Anisotoma castanea	SXO	ΡF	T		<u>ن</u> ت	ω
Anisotoma glabra	SxO	ΗĘ	T		18	12
Anisotoma humeralis	SXO	FF	TI			2
Anomognathus cuspidatus	SxO	ŦF	D,P			ω
Anoplodera reyi	SXO		W		ف	
Anotylus clavatus			dλ			-
Aphodius ater			O		4	
Aphodius depressus			0			4

Cartodere constricta SXF Catops alpinus			Cardiophorus ruficollis SxO	Cantharis paludosa	Calitys scabra	Caenoscelis terruginea SxF		Bryoporus cernuus SxF	sicus	Boreophila eremita Unknown	Bolitophagus reticulatus SxO	Bolitochara pulchra SxF	S	Bibloporus bicolor SxO	Bembidion lampros	Bembidion grapii	Batrisodes hubenthali	S	sde	Atrecus affinis SxO	Atomaria umbrina SxO			ula	đ.		Atomaria nitidula Unknown		Atomaria fuscata SxF		Atomaria atrata		scus	Atheta procera Unknown		CA.	x	si	Atheta aquatica Unknown		Arpidiphorus orbiculatus SxF	Aphodius rufipes	Aphodius nemoralis	mm	Species name Saproxylic group
					FF	TT				Unknown	ΡF		Η	Ħ			FF	ŦF			ΗF		ŦF	FF			Unknown			HH.	Ŧ			Unknown				Unknown	Unknown	SFF	FF				oup Fire cathegory
D		ΤI	D,W	dć	ł	т	T	dč	dć	Unknown	Ъ	γF	P	ŋ	ס	P		jþ	ζþ	P	F	н	Ŧ	H	п	Ŧ	Unknown	T	Ŧ	T	т	TI	ס	Unknown				Unknown	Unknown	C,W	T	0	D	D	Functional group
					VU					Unknown							2										Unknown					NT		Unknown				Unknown	Unknown						Redlist cathegory
		21	-	4	-	13	4		4	<u>ند</u>	N	N	-1	-	4	-	-	2	-				187	-	-			-	-	16	9	0	U		-	თ	4		-	10	2	13	20	10	Burned
	2						-	<u> </u>					ω	13					თ	11	2	-	9			ω	-			32	1	12	ω	2				cn		-	6	30	10	12	Control

Species name	Saproxylic group	Fire cathegory	Functional group	Redlist cathegory	Burned	Control
Cerylon deplanatum	A R R R R R R R R R R R R R R R R R R R	ΗF	-1		10111	2
Cerylon ferrugineum	SXO	17	1 71			14
Cerylon histeroides	OXS	Ŧ	Ŧ		289	138
Chilocorus renipustulatus	NS NS	1			<u>د</u> د	
Cis bideniatus	axo	TT	1 7		10	3 2
	SXO	R	n 7		n 40	47
Cis dentatus	SVO	H :	Π-	TN	c	- 2
Cis hispidus	SXO		. IL		7	30
Cis jacquemartii	SXO		п			-
Cis lineatocribratus	SxO		וד		ω	14
Cis punctulatus	SxO	FF	H		2	49
Coccinella trifasciata			σ		8	-
Coccinula quadriguttatus	SN				ω	
Coccinula quattuordecimpustulata			P		ω	
Corticaria abietorum	SxF	FF	Ŧ		د.	
Corticaria fennica	SXO		TI	DD	ľ	4
Corticaria ferruginea	SxF	뀨	п		56	-
Corticaria interstitialis	SxF		TI			6
Corticaria lapponica	SXO	Η	п		-	1
Corticaria orbicollis	SXO	FF	т		U	6
Corticaria rubripes	SxF	뀨	н		607	40
Corticaria saginata			F		-	
Corticarina fuscula	SXF	Η	П		11	
Cortinicara gibbosa	SxF	FF	T		-	
Cryptophagus confertus	د.		н		-	
Cryptophagus confusus	SxO	FF	TI			ω
Cryptophagus dentatus	SxF		F		12	14
Cryptophagus dorsalis	SXF	Ŧ	TI		2	@
Cryptophagus lapponicus	SxF	뀨	'n		15	96
Cryptophagus lysholmi	SxO	ΡF	TI	VU	ω	
Cryptophagus saginatus	Unknown	Unknown	Unknown	Unknown	فد	
Cryptophagus subdepressus	Unknown	Unknown	Unknown	Unknown	0	
Crypturgus cinereus	SXO	Ŧ	C		N	
Crypturgus pusillus	SxO	FF	С			4
Curtimorda maculosa	SxO	FF	Ŧ		<u>ب</u>	
Cyphon coarctatus			Ξ		N	
Cyphon padi			D,H		7	<u>ن</u> ــ
Cyphon pubescens			D,H		35	4
Cytilus sericeus			н		-	1
Dacne bipustulata	SxO	FF	T		353	311
Dasytes obscurus	SXO		dί		Ì	2
Dendrophagus crenatus	SxO	FF	T			ω
Dendrophilus pygmaeus	SxF		ס		2	
Denticollis borealis	SXO	SFF	PW	TN	-	

Species name	Saproxylic group	Fire cathegory	Functional group	Redlist cathegory	Burned	COLLION
Dictyoptera aurora	SXO	нн	J		2	7
Dinaraea aequata	SXO	Ŧ	ŻĿ		2	-
Dinaraea linearis	SxO	Ŧ	γF			4
Dorcatoma robusta	SxO	ŦF	п			2
Dromius agilis	SxF	FF	P			7
Dryocoetes autographus	SXO		C		20	ω
Dryocoetes hectographus	SxO		C		32	17
Eanus costalis	3		Ś		23	91
Enicmus fungicola	SxO	FF	Ŧ		ω	4
Enicmus planipennis	SXO		п			ω
Enicmus rugosus	SXO	FF	T		113	560
Ennearthron cornutum	0.00	FF				2
Ennearthron laricinum	SXO	FF	Ŧ	NT		2
Enochrus affinis	Unknown	Unknown	Unknown	Unknown		N
Episernus angulicollis	SXO	FF	W		2	6
Epuraea aestiva	SXF		п			-
Epuraea angustula	SxO	ΗF	F,P			4
Epuraea biguttata	SXO	귀	т		2	
Epuraea boreella	SxO	FF	ם,ח		Ch	ω
Epuraea contractula	SxO	Η	т		œ	
Epuraea marseuli	SXO	FF	D,F		13	-
Epuraea pygmaea	SXO	ŦĦ	D,F		თ	4
Epuraea rufomarginata	SxF	FF	T		б	
Epuraea silacea	SXO		п		4	31
Epuraea variegata	SxO	FF	T		-	
Euaesthetus ruficapillus	SN					
Euconnus maeklinii	SxF					-
Eudectus giraudi	SXO	FF	?F			ω
Euplectus piceus	SxF		P		2	-
Euplectus punctatus	SXO		ס		12	40
Euryusa castanoptera	SxO	FF	γF		4	4
Sabrius expectatus	SxF	FF	q		8	44
Glischrochilus quadripunctatus	SxO	FF	Ę,P		46	-
Globicornis emarginata	SXO	FF	D			-
Gyrophaena affinis	SXF		T			
Gyrophaena boleti	SXO		FI			1
Hadrobregmus confusus		FF		VU		N
Hadrobregmus pertinax	SXO	ŦĦ	W		G	19
Hallomenus binotatus	SXO		Ŧ			2
Hapalaraea clavigera	SXO	FF	άć	NT	ω	4
Hapalaraea linearis	SxO	ŦF	di		-	
Haplogiossa marginalis	SXF		?F		د	
Haploglossa nidicola	Unknown	Unknown	Unknown	Unknown		-
Haploglossa villosula	SxF		?F		32	60
Henoticus serratus	SxF	SFF	TI		12	

species name	Sabroxylic group	File Califeguiy	I unicuonal group	Neulist cattleguly	Dattien	COLLING
Nephus bisignatus	ALL	The second s	- D		2	1000
Vevraphes coronatus	SxF		סי נ			
Vouopnius aquaucus			το		2 X	
Notothecta flavipes	Unknown	Unknown	Unknown	Unknown		-
Nudobius lentus	SXO	뀨	ס		Ν	
Omalium caesum	SxF		Чć	~	cs ا	
Orchesia fasciata	SXO		H	NT	1	
Orchesia micans	SXO	Ŧ	TI		12	15
Orchesia minor	SXO		п	NT		-
Orithales serraticomis			Ş		თ	
Orthocis alni	SXO	Ŧ	H			7
Orthocis linearis	SxO	FF	ц		-	
Orthomicus suturalis	SXO				2	
Orthoperus atomus	SxF	FF	T		i	1
Ostoma ferruginea	SXO	Ŧ	п		2	ω
Oxymirus cursor	SXO	FF	W			ω
Oxypoda elongatula		1	?F		د.	
Oxypoda skalitzkyi	SXF		, Έ		N	4
Oxypoda spectabilis	SxF		żΕ			4
Oxypoda umbrata	Unknown	Unknown	Unknown	Unknown		
Philonthus sordidus	Unknown	Unknown	Unknown	Unknown	-	
Phloeonomus lapponicus	SxO	FF	P		78	
Phloeonomus planus	SXO	FF	γņ		49	-
Phloeonomus pusillus	SXO	TT I	ס		N	2
^o hloeonomus sjoebergi	SXO	뭐	dč			-
Inloeopora concolor		Ŧ			قت .	
Phloeopora corticalis	SxO	17	9; 9;		د .	4
-inverpora testacea	ovo	LL.	, i		×	1
Phyliodrena melanocenhala	S×⊑ OXC	R	¢		ł	л –
Dissodes av/lenhalii	S.O.		ר		3	c
Pissodes pini	SXO	H H	o 0		ז רס	
Pityogenes bidentatus	SXO	Ŧ	o		198	20
Pityogenes chalcographus	SXO	FF	C		ω	
Pityogenes quadridens	SXO	Ŧ	C		10	
Pityophagus ferrugineus	SXO	FF	P		51	
Piacusa atrata	SXO	SFF	d;		17	<u>_</u>
Placusa depressa	SXO	FF	?P		-	
Platycerus caprea	SXO	ĘĘ	W		2	4
Platydracus fulvipes	Unknown	Unknown	Unknown	Unknown		1
Plegaderus vulneratus	SXO	H	q			د.
Podabrus alpinus			۶Þ			
Pogonocherus decoratus	SXO		C,W		-	
Pogonocherus fasciculatus	SXO	FF	C,W			-

species name	Saproxylic group	Fire cathegory	Functional group	Redlist cathegory	Burned	Control
Polygraphus punctifrons	SXO	FF	C			2
Polygraphus subopacus	SxO	ŦF	0		2	1
Porrhodites fenestralis	Unknown	Unknown	Unknown	Unknown	33	
Potosia cuprea spp metallica	SXF					2
Prosternon tessellatum	Unknown	Unknown	Unknown	Unknown	N	б
Pseudomedon obscurellus	Unknown	Unknown	Unknown	Unknown	-	
Pterostichus adstrictus		FF	0		-	
Pterostichus oblongopunctatus		Η	ס			
Pteryx suturalis	SxO	ΕF	п			4
Pytho depressus	SXO	뭐	ი		ω	
Quedius plagiatus	SXO	FF	ט		-	0
Quedius tenellus	SXF		dč			2
Rhagium inquisitor	SxO	FF	C,W		76	ω
Rhagonycha atra	<i>د</i> .		ζþ		2	4
Rhagonycha elongata	ć		?P		4	4
Rhizophagus bipustulatus		ŦF				ω
Rhizophagus dispar	SxF	FF	σ			2
Rhizophagus ferrugineus	SxO	귀	ס		35	15
Rhizophagus nitidulus	SxO	FF	σ			G
Rhizophagus parvulus	SxO	Ŧ	J		193	32
Rhyncolus sculpturatus	SxO	FF	W		4	10
Ropalodontus strandi	(tom)					4
Salpingus ruficollis	SxO	ΗF	σ			4
Scaphisoma agaricinum	SxF	FF	.9E		71	113
Sciodrepoides fumatus	SN					2
Sciodrepoides watsoni		0	dč		-	
Scolytus ratzeburgi	SXO	SEE	C			2
Scymnus limbatus	SN				-	
Selatosomus aeneus			Ś		-	-
Selatosomus impressus	SxF		σ		υ'n	18
Selatosomus melancholicus			Ś		16	
Sepedophilus testaceus	SxF		T			10
Sericoda quadripunctata		PF			-	
Sericus brunneus	Sector Sec.		<i>د</i> .		23	00
Sogda suturalis	Unknown	Unknown	Unknown	Unknown		
Soronia grisea	SXO		D			N
Sphindus dubius	SXF	FF	т		N	4
Stenichnus bicolor	SXF	FF	ס			4
Stenotrachelus aeneus	SxO	SFF	C,W		N	
Stenus clavicornis			P		4	
Stenus europaeus	Unknown	Unknown	Unknown	Unknown	-	
Stenus strandi	Unknown	Unknown	Unknown	Unknown	-	
Stephanopachys linearis		PF		VU	4	
Stephostethus pandellei	SxF	Ŧ	п			N
Stephostethus rugicollis	SXF		п		-	4

Species name	Saproxylic group	Fire cathegory	Functional group	Redlist cathegory	Burned	Control
Strophosoma capitatum	State of the state	Η	H	10 10 10 10 10 10 10 10 10 10 10 10 10 1	9	2
Sulcacis affinis	SXO	FF	т		N	-
Synchita humeralis		SFF				4
Syneta betulae			Т			-
Tachinus atripes			?F,?P			-
Tachinus basalis				B		
Tachinus elongates	SN				_	4
Tachinus pallipes			?F,?P			2
Tachinus proximus			?F,?P		-	4
Tachyta nana	SxO	SFF	σ		-	
Tanasimus formicarius	SxO				2	
Tetratoma ancora	SXO	Ŧ	T		1	4
Thalycra fervida	Unknown	Unknown	Unknown	Unknown	U	2
Thanasimus formicarius	SXO	Ŧ	σ		23	
Thyasophila angulata			di		-	
Thyasophila canaliculata	Unknown	Unknown	Unknown	Unknown		-
Tomicus piniperda	SxO	FF	0		258	2
Triplax aenea	SXO	FF	п		-	6
Triplax russica	SXO	FF	TI		12	49
Triplax scutellaris	SXO	Ŧ	LI.		б	40
Tropideres dorsalis		SEE		TN	نہ	
Trypodendron lineatum	SxO	뀨	Ŧ		37	13
Trypodendron proximum	SxO				44	35
Tyrus mucronatus	SxF	ŦF	,Έ		00	
Xenota myrmecobia	Unknown	Unknown	Unknown	Unknown	10	თ
Xylechinus pilosus	SXO		0			30
Xylita laevigata	SxO	FF	т		506	122
Zilora ferruginea	SxO	FF	н	NT	E.	
Total abundance					5507	3079
Total species richness					239	221