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Evaluation of 16S next-generation sequencing of hypervariable region 4 in wastewater samples: an unsuitable approach for bacterial enteric pathogen identification

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Abstract

Recycled wastewater can carry human-infectious microbial pathogens and therefore wastewater treatment strategies must effectively eliminate pathogens before recycled wastewater is used to supplement drinking and agricultural water supplies. This study characterised the bacterial composition of four wastewater treatment plants (WWTPs) (three waste stabilisation ponds and one oxidation ditch WWTP using activated sludge treatment) in Western Australia. The hypervariable region 4 (V4) of the bacterial 16S rRNA (16S) gene was sequenced using next-generation sequencing (NGS) on the Illumina MiSeq platform. Sequences were pre-processed in USEARCH v10.0 and denoised into zero-radius taxonomic units (ZOTUs) with UNOISE3. Taxonomy was assigned to the ZOTUs using QIIME 2 and the Greengenes database and cross-checked with the NCBI nr/nt database. Bacterial composition of all WWTPs and treatment stages (influent, intermediate and effluent) were dominated by Proteobacteria (29.0-87.4%), particularly Betaproteobacteria (9.0-53.5%) and Gammaproteobacteria (8.6-34.6%). Nitrifying bacteria (*Nitrospira* spp.) were found only in the intermediate and effluent of the oxidation ditch WWTP, and denitrifying and floc-forming bacteria were detected in all WWTPs, particularly from the families Comamonadaceae and Rhodocyclales. Twelve pathogens were assigned taxonomy by the Greengenes database, but comparison of sequences from genera and families known to contain pathogens to the NCBI nr/nt database showed that only three pathogens (*Arcobacter venerupis*, *Laribacter hongkongensis* and *Neisseria canis*) could be identified in the dataset at the V4 region. Importantly, Enterobacteriaceae genera could not be differentiated. Family level taxa assigned by Greengenes database agreed with NCBI nr/nt in most cases, however, BLAST analyses revealed erroneous taxa in Greengenes database. This study highlights the importance of validating taxonomy of NGS sequences with databases such as NCBI nr/nt, and recommends including the V3 region of 16S in future short amplicon NGS studies that

aim to identify bacterial enteric pathogens, as this will improve taxonomic resolution of most, but not all, Enterobacteriaceae species.

Keywords: Wastewater, next-generation sequencing, 16S rRNA, V4, Greengenes, Enterobacteriaceae.

1. Introduction

Water is becoming an increasingly scarce global resource, and as the overall demand for water grows, the quantity of wastewater produced and its overall pollution load are continuously increasing worldwide (Connor et al., 2017). Recycled wastewater is an essential resource in addressing this problem, as properly treated water can be safely released back into the environment, and used to supplement limited drinking water supplies. However, unless effectively treated, recycled wastewater has the potential to carry microbial pathogens (viruses, bacteria, protozoa and helminths), toxic chemicals and heavy metals. Therefore, treatment strategies must effectively eliminate these major public health risks (Rodriguez-Manzano et al., 2012).

Wastewater recycling in urban areas typically employs reverse osmosis membranes or advanced oxidation treatment after activated sludge wastewater treatment. This results in high purity recycled water, fit for potable reuse, but is technically challenging and expensive (Rajasulochana and Preethy, 2016; Garrido-Cardenas et al., 2017). In contrast, rural WWTPs typically use simple, non-mechanical waste stabilisation ponds (WSPs) or lagoons consisting of open basins that rely on natural microorganisms and algae to assist in the breakdown and settlement of degradable organic matter. Wastewater "influent" enters on one side of the WSP and exits on the other side as "effluent", after spending days or even months undergoing treatment processes in the pond, depending on plant capacity and flow rate. The treated

effluent is discharged generally for non-potable purposes, such as irrigation of public open spaces or agricultural/horticultural uses (Von Sperling, 2007; Anon, 2009). These WSPs are widely used across the world as passive wastewater treatment for domestic wastewaters as they can offer low cost, low maintenance and effective pathogen removal (Von Sperling, 2007; Ho et al., 2017; Eland et al., 2018).

Removal and inactivation of pathogens from WSPs is achieved via long retention times, increased temperature and pH, the presence of algal antibacterial compounds and sunlight penetration. Therefore shallow (<1 m) WSPs with low turbidity, high pH and maximal sunlight exposure will achieve the most efficient pathogen removal (Sharafi et al., 2012). While WSP systems can achieve high removal efficiencies (4-6 log₁₀), the efficiency of pathogen removal in full-scale systems is highly variable, and many WSP systems achieve only 2-3 log₁₀ removal (Verbyla et al., 2017).

In contrast to WSPs, many conventional WWTPs use an activated sludge process in which a biological sludge containing living microorganisms is mixed with wastewater and aerated in a reactor, forming a mixed liquor. Microbial populations within the activated sludge include a range of bacteria, yeast, fungi, protozoa and higher organisms such as rotifers that can digest organic matter in wastewater, and clump together (by flocculation), producing a treated wastewater that is relatively free from suspended solids and organic material. The removal mechanisms of pathogens in an activated sludge system are inactivation, hunting by ciliate protozoa, adsorption to solids and capsulation inside the sludge flocs (Sharafi et al., 2012).

Understanding the diversity of bacterial microorganisms in wastewater is essential for understanding the performance for biological wastewater treatment systems (Inaba et al., 2017). DNA-based approaches for identification of bacteria, such as polymerase chain reaction (PCR) and Sanger sequencing, can overcome the limitations of conventional

bacterial identification techniques (e.g. microscopy, culture-dependent assays and biochemical techniques) that are laborious and time-consuming, by allowing for the identification of microbes that are morphologically indistinguishable, uncultivable, fastidious, and obligate intracellular. Molecular bacterial identification approaches often target the 16S rRNA (16S) gene, which enables species differentiation based on genetic dissimilarity. However, the throughput of species identification with PCR and Sanger sequencing is limited by individual clone library preparation, and species-specific PCR approaches require *a priori* hypotheses regarding the taxa to be targeted. Wastewater can be comprised of hundreds of bacterial species (Berlec 2012; Kim et al. 2015), therefore assessments of bacterial diversity on this scale using PCR and/or Sanger sequencing is impractical. The rapid advances of next-generation sequencing (NGS) technologies have revolutionised the ability to identify large numbers of bacteria from various types of environmental and biological samples (Garrido-Cardenas et al., 2017). Primers targeting one or more of the nine hypervariable (V) regions of 16S can be used with NGS to identify bacteria. Other studies that have used 16S NGS to identify bacteria in WWTPs have targeted V3-4 (Lu et al., 2015), V4 (Zhang et al., 2012) and V5-6 (McLellan et al., 2010), and there is no consensus on the most suitable region to target for bacterial assessments in WWTPs. The V4 region of 16S is commonly targeted in microbiome studies with the widely used 515F/806R primers (Caporaso et al., 2011). These primers are recommended in the Earth Microbiome Project's Illumina NGS protocol (<http://www.earthmicrobiome.org/protocols-and-standards/16s/>) and have been modified by other studies to include additional degeneracies to allow amplification of additional taxa (Aprill et al., 2015; Parada et al., 2015). Therefore, the present study evaluated the ability of the V4 16S NGS to identify bacteria, particularly enteric pathogens, in WWTPs, and used this NGS approach to characterise bacterial compositions in different treatment stages (influent, intermediate and effluent) of three WSPs and one oxidation ditch WWTP, which is

a modified activated sludge WWTP, that utilises prolonged aeration to remove biodegradable organic compounds (Baars, 1962), in Western Australia (WA).

2. Methods

2.1 Study sites and sample collection

Wastewater samples ($n = 26$) were collected from three WSPs (WWTPs 1, 2 and 3) and an oxidation ditch (WWTP 4) in 2015 in WA (Table 1 and Figure 1). Samples were collected in February, July and September in 2015 and covered two seasons for each site. Samples were collected from WWTP 1, located in north-west WA and in a tropical climate, during the wet and dry seasons, while samples from WWTPs 2, 3 and 4, located in south-west WA and in a temperate climate, were collected during summer and winter (Table 1). Wastewater samples were also collected at different treatment stages (influent, intermediate and effluent) during summer and winter (or dry and wet seasons for WWTP 1 samples) (Table 1). The wastewater samples were collected in 1 L sterile containers that were treated with chlorine and rinsed with the sample before filling. Samples were kept cool in an ice box during transport back to the laboratory, and then stored at 4 °C and processed within 48 hours prior to DNA isolation.

2.2 DNA isolation

After 100 mL of each wastewater sample was filtered through sterile 0.2 µm Sterivex filters (Millipore, USA), genomic DNA (gDNA) was extracted from the filters using a PowerWater Sterivex DNA Isolation Kit (MO BIO Laboratories, California, USA). Extraction reagent blank controls (ExCs; $n = 6$) were included alongside each batch of gDNA extractions. Purified DNA was stored at -20 °C prior to molecular analysis.

2.3 Next-generation sequencing library preparation

The NGS library was prepared and sequenced following the 16S metagenomic sequencing library preparation protocol from Illumina (Part # 15044223 Rev. B; Illumina, USA), with minor modifications to the first stage PCR. V4 16S was amplified using modified 515F/806R primers [originally designed by Caporaso et al. (2011)]: 515FB 5'-GTGYCAGCMGCCGCGGTAA-3' (Parada et al., 2015) and 806RB 5'-GGACTACNVGGGTWTCTAAT-3' (Aprill et al., 2015). The 515FB/806RB primers were modified to include Illumina MiSeq adapter sequences (Part # 15044223 Rev. B; Illumina, USA), and conventional PCRs were carried out as described elsewhere (www.earthmicrobiome.org/protocols-and-standards/16s/; <https://doi.org/10.17504/protocols.io.nuudeww>). No-template controls (NTCs) were included alongside each PCR. The V4 16S library was sequenced on the Illumina MiSeq platform (San Diego, CA, USA) with v2 sequencing chemistry.

2.4 16S Bioinformatic analysis

Paired-end 16S reads were merged (minimum 50 bp overlap), trimmed of primers and distal bases, quality filtered (maximum expected error threshold of 1.0) and singletons were removed with USEARCH v10.0 (Edgar, 2010), resulting in reads that were 247 bp in length on average. Reads were denoised into zero-radius operational taxonomic units (ZOTUs) and chimeras were filtered with UNOISE3 (Edgar, 2016). Taxonomic assignment of ZOTUs was performed in QIIME 2 v2018.2 (Caporaso et al., 2010, <https://qiime2.org>) using the QIIME 2 feature classifier plugin (Bokulich et al, 2018) and the August 2013 release of the Greengenes sequence database (McDonald et. al., 2012). The sequences were also BLAST searched against the National Center for Biotechnology Information (NCBI) non-redundant nucleotide (nr/nt) database to cross-check Greengenes assigned taxonomy. ZOTUs that were in low

abundance (<0.05% sequence composition) across all samples may represent PCR or sequencing error, therefore, they were bioinformatically removed from the samples. To assess sequencing depth, alpha rarefaction plots were generated with the R package *vegan* (Oksanen et al., 2018) using R software (R Core Team, 2013).

2.5 Phylogenetic analysis

Enterobacteriaceae ZOTUs were aligned using the MAFFT program (Katoh et al., 2002) with closely related sequences retrieved from the NCBI nr/nt database in Geneious v10.2.2 (<http://www.geneious.com>, Kears e et al., 2012). Sequences in the alignment were trimmed to the same length, then were imported into the program PhyML (Guindon et al., 2010) and assessed for the most appropriate nucleotide substitution model (GTR+G+I) based on Akaike Information Criterion (AIC). Maximum likelihood trees were constructed using RAxML (Stamatakis, 2014). Genetic distance estimates were calculated with Kimura distance matrices (Kimura, 1980) in Geneious v10.2.2.

ZOTU sequences generated from this study have been submitted to GenBank under the accession numbers MH892609 to MH892828. Raw sequence files were deposited in the NCBI Sequence Read Archive under the accession number PRJNA526519 (refer to Table 1 for sample names and metadata).

3. Results

3.1 Next-generation sequencing library summary

Approximately 1.4 million paired-end V4 16S sequences were obtained for all samples and controls ($n = 34$) (Table 2). After the reads were pre-processed (merged, quality filtered with singletons and chimeras removed), there was a total of ~800,000 sequences for all samples (~24,000 average). The processed 16S sequences (total of ~700,000) excluded

sequences that were not classified as bacteria and low abundance (<0.05%) ZOTUs, and on average, there were ~27,000 processed bacterial 16S sequences for the WWTP samples ($n = 26$). Few sequences were detected in the ExCs and NTCs, which had an average of 8 sequences (Table 2).

3.2 Bacterial sequence composition in WWTPs

A total of 3,598 ZOTUs (Supplementary File B.1) were obtained for the pre-processed sequences, and a total of 1,644 ZOTUs remained for the processed sequences. For the processed sequences, sequencing depth was adequate for all samples at ~5,000 sequences (Supplementary Figures A.3 and A.4), but the alpha rarefaction plots did not reach a plateau for the pre-processed sequences (Supplementary Figures A.1 and A.2). The archaeal sequence compositions were low and two archaeal phyla were detected: Euryarchaeota was found in the influent of WWTP 4 (<0.1%) and effluent of WWTP 2 (0.1%), and Parvarchaeota was found in the effluent of WWTP 4 (0.1%). Two different types of Euryarchaeota were detected, *Methanobrevibacter* sp. from the class Methanobacteria in WWTP 4 influent and *Methanosaeta* sp. from the class Methanomicrobia in WWTP 2 effluent. The taxonomy for Parvarchaeota was assigned as Parvarchaea for class, and WCHD3-30 and YLA114 for Parvarchaea orders, with no further taxonomic classifications assigned by Greengenes (Supplementary File B.2).

Bacteria were classified into 28 phyla (Supplementary File B.2); the most dominant phylum was Proteobacteria across all WWTPs and treatment stages (influent, intermediate and effluent), with sequence compositions ranging from 29.0% in the effluent of WWTP 2 to 87.4% in the intermediate stage of WWTP 3 (Figure 2). Other abundant phyla (>10% composition in WWTP samples) were Bacteroidetes (ranging from 4.1% in WWTP 1 influent to 31.5% WWTP 3 effluent), Cyanobacteria (0% (not detected) in WWTP 1 and 3 influent to

47.2% WWTP 2 effluent), Firmicutes (0.1% in WWTP 3 effluent to 22.1% in WWTP 1 influent) and Actinobacteria (1.1% in WWTP 4 influent to 10.3% in WWTP 2 influent) (Figure 2).

Six classes of Proteobacteria were identified: Alphaproteobacteria, Betaproteobacteria, Deltaproteobacteria, Epsilonproteobacteria, Gammaproteobacteria and "TA18". Betaproteobacteria and Gammaproteobacteria sequences were abundant ($\geq 8.6\%$) and prevalent across all WWTPs and treatment stages. There were also six classes for Bacteroidetes: WWTP 1 and 4 exhibited a similar pattern in sequence composition of Bacteroidetes, with classes Bacteroidia and Flavobacteriia detected in the influent, and in addition to these two classes, three other classes (Saprospirae and Sphingobacteriia) were also detected in the intermediate and effluent stages (Figure 2). Like WWTP 1 and 4, Bacteroidia and Flavobacteriia were detected in all stages of WWTP 3, and the same classes that were found in WWTP 1 and 4 were also found in WWTP 3, but in the intermediate stage of WWTP 3, only Bacteroidia, Flavobacteriia and Saprospirae sequences were obtained. WWTP 2 had similar Bacteroidetes in the influent and effluent; all aforementioned Bacteroidetes classes were found in the influent and effluent, and an additional class, Rhodothermi, was also found in the effluent (Figure 2 and Supplementary File B.2).

Cyanobacteria were not found in the influent of WWTP 1 and 3, but sequences were detected in the intermediate and effluent stages of these plants, and were detected in all stages of WWTP 2 and 4. Oscillatoriophyceae was dominant in the intermediate and effluent of WWTP 1 and 2 (11.2% and 14.3%, respectively) and was also detected in the effluent of WWTP 2 and 3. Other classes of Cyanobacteria included Synechococcophyceae in the intermediate of WWTP 1 and effluent of WWTP 1 and 4, Nostocophyceae in WWTP 2 effluent, and a class designated as 4C0d-2 by the Greengenes database was found in WWTP 3 intermediate and WWTP 4 intermediate and effluent. Among the Firmicutes, three classes

were detected: Bacilli (Bacillales, Lactobacillales and Turicibacterales), Clostridia (Clostridiales) and Erysipelotrichi (Erysipelotrichales). Bacilli and Clostridia sequences were the most abundant classes of Firmicutes and were found in the influent of WWTPs 1, 3 and 4, ranging from 0.8-22.1%, and sequences were in low abundance ($\leq 2.0\%$) or not detected in the intermediate and effluent stages (Figure 2 and Supplementary File B.2).

Five Actinobacteria classes were identified: Acidimicrobiia, Actinobacteria, Coriobacteriia, Nitrospirae and Thermoleophilia. The sequence composition of the class Actinobacteria was higher than other classes of the phylum Actinobacteria (which were all $\leq 2.0\%$ in various treatment plants and stages), particularly in the intermediate and effluent of WWTPs 1, 3 and 4 (1.7-8.0%), and the influent and effluent of WWTP 2 (9.1% and 4.8%, respectively) (Figure 2). Acidimicrobiia and Thermoleophilia were detected in the intermediate and effluent of WWTP 1, influent and effluent of WWTP 2 and intermediate of WWTP 4. A low sequence composition of Actinobacteria and Coriobacteriia were found in the influent of WWTPs 1, 3 and 4. Nitrospirae was only found in the effluent of WWTP 2 (Supplementary File B.2).

3.3 Bacterial pathogen identification

Based on Greengenes taxonomic assignments, seven ZOTUs were assigned to the family Enterobacteriaceae (Gammaproteobacteria: Enterobacteriales). Four of these ZOTUs were not assigned further taxonomy by Greengenes, but the remaining ZOTUs were designated as *Citrobacter* sp., *Escherichia coli* and *Trabulsiella* sp. with high confidence (0.95-1). However, comparison of the Enterobacteriaceae sp. ZOTUs to the NCBI nr/nt database using BLAST revealed that all these ZOTUs were 100% similar to multiple Enterobacteriaceae sp. genera (Table 3). The phylogenetic tree constructed with Enterobacteriaceae sp. ZOTUs and Enterobacteriaceae sequences from the NCBI nr/nt

database showed that different genera grouped closely with short branch lengths, and most bootstrap values were low (Figure 3; refer to Supplementary file B.3 for pairwise genetic distances, which range from 94.3-100%). This supports the BLAST results from Table 3 that suggest that the Enterobacteriaceae sp. ZOTUs can only be confidently assigned to the family level, and suggests that the V4 region of 16S cannot distinguish between many Enterobacteriaceae species and genera.

Other ZOTUs from the class Gammaproteobacteria that were assigned to pathogenic species, or to taxa that contain pathogens, based on Greengenes taxonomy included *Acinetobacter* (*Acinetobacter johnsonii*, *Acinetobacter lwoffii* and unassigned species), Aeromonadaceae (unassigned genera and *Tolumonas*), Coxiellaceae (genus unassigned), Legionellaceae (genus unassigned), *Enterococcus* spp. (Enterococcaceae), Pseudomonadaceae (*Pseudomonas alcaligenes*, *Pseudomonas fragi*, *Pseudomonas nitroreducens*, *Pseudomonas stutzeri*, *Pseudomonas veronii*, *Pseudomonas viridiflava* and unassigned species), Piscirickettsiaceae (genus unassigned) and Pseudoalteromonadaceae (genus unassigned). Greengenes taxonomy that conflicted with BLAST analysis was identified for *Tolumonas* (Aeromonadaceae; ZOTU 483), which was most similar to *Pseudoaeromonas sharmana* (100%; GenBank® accession no. MF280154), *Aeromonas sharmana* (99.2%; JF496528) and *Tolumonas* sp. (98.8%; MG801837). *Acinetobacter* ZOTUs assigned to the species level (*Acinetobacter johnsonii* and *Acinetobacter lwoffii*) with high confidence naïve Bayes confidence scores (0.94-1) were also 100% similar to several other *Acinetobacter* species. Similarly, many *Pseudomonas* species had sequence similarities of 100%, therefore had incorrect species level taxonomy assigned with Greengenes. Greengenes taxonomy was more conservative for ZOTU 589 (Pseudoalteromonadaceae sp.) as BLAST results showed that this sequence could be assigned to the genus *Vibrio*, but like

Acinetobacter and *Pseudomonas*, many *Vibrio* species were also 100% similar at the V4 region (Supplementary File B.4).

The BLAST results agreed with Greengenes taxonomic assignments for most Betaproteobacteria (Alcaligenaceae spp., Neisseriaceae spp. and *Vitreoscilla* spp.), except for ZOTU 55 that was classed as *Microvirgula* sp., but was 100% similar to *Laribacter hongkongensis* sequences (NR025167). Five Campylobacteraceae (class Epsilonproteobacteria) ZOTUs that were assigned to the genus *Arcobacter* or *Arcobacter cryaerophilus* were also 100% similar to *Campylobacter* sequences and therefore could only be confidently assigned to the Campylobacteraceae family. *Corynebacterium* spp. and *Mycobacterium* spp. (phylum Actinobacteria) taxonomy agreed for both Greengenes and BLAST results, but there were discrepancies for Streptococcaceae spp. (phylum Firmicutes) that were assigned as *Streptococcus* spp., *Streptococcus luteciae* and *Streptococcus minor* by Greengenes, but could not be assigned to the species or genus level by BLAST in most cases. The Greengenes taxa *Candidatus Rhabdochlamydia* sp. (ZOTU 1597; phylum Chlamydiae), *Clostridium* spp., *Proteiniclasticum* sp. (phylum Firmicutes) or *Treponema* spp. (Spirochaetes) could also not be confidently assigned to the genus level based on BLAST results (Supplementary File B.4).

The sequence compositions for pathogenic and potentially pathogenic taxa that were given final taxonomic assignments based on Greengenes and NCBI nr/nt sequence database comparisons are summarised in Table 4. Briefly, sequence compositions for these taxa were generally higher in the influent for WWTP 1, 3 and 4 and lower in the intermediate and effluent, with the exception of *Acinetobacter* spp. in the intermediate stage of WWTP 3 (17.1%) and *Aeromonas* sp. in the effluent of WWTP 1 (8.6%). Potentially pathogenic sequence compositions were relatively low in the influent and effluent of WWTP 2, with the highest composition of 2.0% in the effluent for Alcaligenaceae sp. (Table 4).

3.4 Nitrifying, denitrifying and floc-forming bacteria

Other bacteria of interest in WWTPs, such as nitrifying, denitrifying and floc-forming bacteria, also had Greengenes taxonomy validated with BLAST results from the NCBI nr/nt database. Compared to pathogenic bacteria, the nitrifying, denitrifying and floc-forming bacterial ZOTUs had more taxonomic assignments that agreed with both databases. All were assigned to the appropriate family, but some ZOTUs had conflicting genera. For example, ZOTU 387 was assigned as *Dechloromonas* sp. by Greengenes, but was also 100% similar to *Azonexus hydrophilus* (LN650477), and ZOTU 766 Greengenes taxonomy was *Comamonas* sp., but this ZOTU was 100% similar to *Comamonas* spp. (MH174324) and *Delftia* spp. (MF156914). Results of taxonomy database comparisons for nitrifying, denitrifying and floc-forming bacteria are provided in Supplementary File B.5, and the sequence compositions of validated taxa are presented in Table 5. Nitrifying bacteria, *Nitrospira* spp. (Nitrospirales: Nitrospiraceae), were only detected in the intermediate and effluent of WWTP 4, with sequence compositions of 1.2% and 1.5%, respectively. In WWTP 1, denitrifying bacteria with the highest compositions were found in the influent for *Comamonas* sp. (Comamonadaceae; 6.9%) and *Thauera* spp. (Rhodocyclaceae; 3.4%). The comamonads *Hydrogenophaga* spp. and *Aquabacterium* sp. had highest compositions in the effluent (2.4%) and influent (1.3%) of WWTP 2, respectively, and the floc-forming bacteria *Flavobacterium* spp. were higher in the effluent (4.7%) than in the influent (2.8%) of WWTP 2. WWTP 3 had a greater diversity of denitrifying and floc-forming bacteria in the influent and intermediate stages than the effluent; the highest sequence compositions in the influent was 4.4% for *Comamonas* sp., 6.0% for *Thauera* spp. in the intermediate stage and 7.6% for *Flavobacterium* spp. in the effluent. The abundance of *Comamonas* sp. sequences was also high in the influent of WWTP 4, and the denitrifying bacteria *Uliginosibacterium* spp. were

highest in the intermediate stage, and *Flavobacterium* spp. was highest in the effluent of WWTP 4 (Table 5). Pseudomonadaceae (*Pseudomonas*) are also denitrifying bacteria, and are summarised in Table 4.

4. Discussion

Evaluation of BLAST results from the NCBI nr/nt database of V4 16S sequences that were assigned taxonomy by the 16S Greengenes taxonomy database to pathogenic species (or to bacterial groups that contain pathogenic species) showed that the V4 region of 16S resolves poorly at the species level, and genus level identification was also impeded in many instances. Comparison of the ZOTU sequences to the NCBI nr/nt database revealed that only three ZOTUs were 100% identical to the following pathogenic species: *Laribacter hongkongensis*, which causes gastroenteritis and diarrhoea (Beilfuss et al., 2015); *Neisseria canis*, which usually infects cats and dogs, but can also infect humans (Safton et al., 1999); and *Arcobacter venerupis*. There are 15 species of *Arcobacter*, and three (*Arcobacter butzleri*, *Arcobacter cryaerophilus* and *Arcobacter skirrowii*) have been associated with gastrointestinal infections (Kayman et al., 2012). *Arcobacter venerupis* has previously only been isolated from shellfish (Levican et al., 2012), and these sequences were in low abundance (0.7%) and only found in the influent of WWTP 1. The sequences from *L. hongkongensis* and *N. canis* were found in the influent of WWTPs 1, 3 and 4, and were in low abundance ($\leq 0.1\%$) or not detected in the intermediate and effluent stages of these plants. Genera known to contain pathogenic species that were validated by BLAST analyses of the ZOTUs against the NCBI nr/nt database included *Aeromonas* sp., *Acinetobacter* spp., *Arcobacter* spp., *Candidatus Rhabdochlamydia* sp., *Corynebacterium* sp., *Enterococcus* spp., *Legionella* sp., *Mycobacterium* spp., *Neisseria* sp., *Pseudomonas* spp., *Streptococcus* spp., *Turneriella* sp. and *Vibrio* sp. A previous 16S NGS study on WWTPs in Australia that also

used the Illumina MiSeq platform identified 25 potentially pathogenic genera (Ahmed et al., 2017), while another study of municipal activated sludge plants across four countries (China, USA, Canada and Singapore) identified 16 pathogenic genera using pyrosequencing (Ye and Zhang, 2011). The abundance of pathogenic genera may vary among studies due to DNA extraction kits, different sequencing technologies, inherent amplification biases during PCR and the 16S hypervariable region(s) targeted (Haft and Tovchigrechko 2012).

Other pathogenic genera that can infect people via contaminated drinking water, *Campylobacter* spp. and *Leptospira* spp., were not identified in the present study, but we cannot exclude the possibility of their presence, as several Campylobacteraceae spp. and Leptospiraceae spp. ZOTUs could not be resolved to the genus level (Table 4 and Supplementary File B.4). Most of the potentially pathogenic genera identified had higher sequence compositions in the influent, and had low compositions or were not detected in the effluent (Table 4). However, *Aeromonas* sp. had relatively high sequence compositions in WWTP 1 intermediate (2.2%) and effluent (8.6%) samples, and a similar trend was observed for *Aeromonas* sp. in WWTP 3, with compositions of 4.7% in the influent, 1.1% in intermediate samples and 4.8% in the effluent. *Acinetobacter* spp. also had a high sequence composition in the intermediate samples of WWTP 3 (17.1%), but was not detected in the effluent. Other studies have also found that *Acinetobacter* spp. sequence compositions were not significantly lower in treated wastewater samples compared to the influent (Ahmed et al., 2017). *Mycobacterium* spp. and *Pseudomonas* spp., which had lower compositions or were not detected in the influent, had higher compositions in the intermediate and effluent (Table 4). The absence or lower abundance of bacteria associated with human waste in the influent compared to the intermediate and effluent may be partly explained by the lack of sample replicates, as only 100 mL grab samples were collected per season at each site and treatment stage. However, the number of 16S sequences obtained by NGS does not represent the

number of bacterial organisms present. A number of factors affect sequence composition, including PCR amplification bias (Hong et al., 2009), sequencing depth and copy number variation in the 16S gene (Kembel et al., 2012).

Many enteric bacteria (Enterobacteriaceae) can be transmitted to humans by faecal-oral transmission and can cause gastrointestinal illnesses with symptoms of abdominal pain, diarrhoea, fever, nausea and vomiting. Human enteric pathogens include *Citrobacter freundii*, *Escherichia coli*, *Klebsiella aerogene*, *Salmonella bongori*, *Salmonella enterica* and *Shigella* spp. (Cabral, 2010). Other pathogens from the family Enterobacteriaceae that can cause gastrointestinal illnesses are *Yersinia enterocolitica*, which is a food-borne pathogen associated with pork products (Bhaduri et al., 2005), and *Raoultella ornithinolytica* (formerly *Klebsiella ornithinolytica*), which has been found in aquatic environments and hospitals, with one report of its isolation from human digestive organs (Seng et al., 2016). Enterobacteriaceae pathogens that cause urinary tract infections and other illnesses in humans include *Proteus mirabilis*, *Proteus penneri*, *Proteus vulgaris* and *Serratia marcescens* (Guentzel et al., 1996). Unfortunately, the V4 region of 16S lacked sufficient variability to distinguish between Enterobacteriaceae genera (Table 3 and Figure 3). The same issue was likely encountered in the V4 16S NGS study by Zhang et al. (2012), that reported the detection of sequences from the order Enterobacteriales. Similarly, a 16S NGS study on WWTPs that targeted the V6 region could not resolve one OTU that was 100% similar to several Enterobacteriaceae genera (primarily *Klebsiella* and *Shigella*) (McLellan et al., 2010). Other 16S wastewater studies that have targeted 16S regions that span two hypervariable regions appear to have been able to resolve Enterobacteriaceae genera. For example, Ahmed et al. (2017) sequenced regions V5-6 (300 bp), and reported the detection of *Escherichia/Shigella* (unclear if these could be differentiated), *Salmonella* and *Yersinia*, but species level assignments were not made. Lu et al. (2015) targeted the V3-4 region (460 bp)

and reported the presence of *Klebsiella pneumoniae* and *Serratia* spp., but performed shotgun sequencing to identify pathogens to the species level, which included *E. coli*, *S. enterica*, *Shigella sonnei* and *Yersinia pestis*. According to a study by Chakravorty et al. (2007), V3 is a more suitable region for the differentiation of Enterobacteriaceae genera, and these authors recommended targeting V2, V3 and V6 to identify the bacterial genera assessed in their study, including *Acinetobacter*, *Bacillus*, *Clostridium*, *Corynebacterium*, *Chlamydia*, *Enterococcus*, *Listeria*, *Mycobacterium*, *Neisseria*, *Pseudomonas*, *Streptococcus*, *Staphylococcus*, *Treponema* and *Vibrio*. Using these three regions means that most of the 110 species examined in their study could be identified to the species level (Chakravorty et al., 2007). Using multiple regions does have some challenges, however. For example, the V2 region of *E. coli* starts at nucleotide (nt) position 137 and V6 ends at nt position 1,043 (Brosius et al., 1978), therefore V2-6 spans 906 bp of 16S. This amplicon is too long for current amplicon NGS sequencers; the maximum length is 600 bp on the Illumina MiSeq with v3 chemistry (<http://www.illumina.com/>). Regions V2-3 and V6 could be targeted separately, or full length 16S could be sequenced on long-read sequencing platforms such as PacBio for improved taxonomic resolution of a greater variety of taxa (Ibal et al. 2019). It is important for Enterobacteriaceae species such as *Escherichia coli* for serotypes to be differentiated at the strain level, as some strains are harmless gut bacteria whereas others are pathogenic, e.g. enterohemorrhagic *Escherichia coli* O157:H7. While some studies state that 16S sequencing is unsuitable for differentiating *E. coli* and *Shigella* spp. serotypes as the sequence similarity is high (97.9-99.9%) (Devanga Ragupathi et al. 2017), Ibal et al. (2019) were able to classify *E. coli* strains based on full length 16S sequences (Ibal et al. 2019). Other housekeeping genes that are conserved among bacteria, such as *gyrB*, *rpoB* and *mdh* have greater genetic variability for distinguishing *E. coli* and *Shigella* spp. strains than 16S (Devanga Ragupathi et al. 2017; Fukushima et al. 2002). These genes could also be targeted

using amplicon NGS approaches for improved taxonomic resolution of bacterial strains, however the use of universal primers is more limited than 16S. Alternatively, shotgun sequencing could be performed, which can provide greater taxonomic and functional information (e.g. pathogenicity islands and toxin-producing genes) than amplicon NGS of several target genes (Sanapareddy et al., 2009; Lu et al., 2015). Shotgun sequencing of metagenomes has been considerably more expensive than amplicon NGS (Goodwin et al., 2016), however costs are reducing, particularly with new approaches such as "shallow shotgun sequencing", which can produce more accurate species level taxonomic and functional profiles of the human microbiome than 16S sequencing (Hillmann et al. 2018).

A large portion of the V4 16S sequences (68%) collected in this current study were not assigned to the genus level with the Greengenes database. Other 16S NGS studies on wastewater have used RDP Classifier (Zhang et al., 2012; Ahmed et al., 2017) and SILVA (McLellan et al., 2010; Lu et al., 2015) databases for taxonomic assignment. According to a recent study that compared the major taxonomy databases (Greengenes, RDP classifier, SILVA, NCBI and OTT), there were few conflicts when SILVA, RDP and Greengenes were mapped into NCBI and OTT (Balvočiūtė et al., 2017). However, we found many genus level conflicts, when potentially pathogenic and denitrifying bacteria were compared to the NCBI nr/nt database (Supplementary Files B.4 and B.5). Furthermore, we found erroneous taxonomy in the Greengenes database that causes 16S sequences deriving from chloroplasts in algae and plants to be classified to the bacterial phylum Cyanobacteria and the class "Chloroplast", which is not a valid taxon. For 44 ZOTUs in our dataset that were classified to the class "Chloroplast", the orders provided by the Greengenes database were Chlorophyta (phylum of green algae), Euglenozoa (phylum of flagellate excavates) and Stramenopiles (infrakingdom of algae and oomycetes). While chloroplast sequences in the Greengenes database can be useful to identify such sequences in an NGS dataset, researchers that aim to

only analyse bacterial 16S sequences at higher levels of classification (kingdom and phylum) may be unaware that the chloroplast sequences are classified in the database at the class level. Classifying the chloroplast sequences as “Chloroplast” at the kingdom level, rather than as “Bacteria” may help researchers to detect these sequences at an earlier stage of the data analysis. We have provided a modified version of the Greengenes 99 OTU taxonomy file for all chloroplast sequences with the kingdom “Bacteria” renamed as “Chloroplast” in Supplementary File B.6. A custom curated sequence database for waterborne pathogens, with quality-checked sequences and taxonomy validated by phylogenetic analyses, may also reduce the errors in bacterial taxonomic assignment experienced with other 16S sequence databases.

Overall, of the 2 archaeal and 28 bacterial phyla detected, Proteobacteria, Bacteroidetes, Cyanobacteria, Firmicutes and Actinobacteria had high sequence compositions (>10%) in WWTP samples (Figure 2). The two most dominant phyla in all treatment stages for WWTPs 1-4 were Proteobacteria and Bacteroidetes, which has also been observed by a previous 16S NGS study that examined bacteria in activated sludge WWTPs across Australia, including Perth (Ahmed et al., 2017). The study by McLellan et al. (2010) that compared V6 16S NGS bacterial profiles in WWTP influent, surface water and human faecal samples, also found that the most dominant bacterial phylum in the WWTPs was Proteobacteria (overall 59% sequence composition), and like our study, Gammaproteobacteria and Betaproteobacteria were the most abundant classes. McLellan et al. (2010) also found that Actinobacteria, Bacteroidetes and Firmicutes were dominant taxa in the WWTP influent, and sewage samples had high compositions of Firmicutes, particularly Clostridia (the human faecal samples were comprised mostly (98%) of Clostridia) and Bacilli (McLellan et al., 2010). In the present study, Firmicutes had the highest compositions in the influent of WWTPs 1, 3 and 4, ranging from 16.8-20.5%; Bacilli ranged from 5.6-11.9% and Clostridia

ranged from 7.9-12.4% (Figure 2). *Bacteroides* is another faecal indicator bacterium (Kreader, 1995), and the sequence compositions for *Bacteroides* spp. in the present study ranged from 0.4% in the influent of WWTP 1 and 2 to 2.4% in the influent of WWTP 3, and sequence compositions were low ($\leq 0.8\%$) or undetectable in the intermediate and effluent (Supplementary file B.2). *Faecalibacterium* is also associated with faeces (Zheng et al., 2009), and was detected in the influent of WWTP 1 (1.0%), 3 (1.7%) and 4 (1.3%), and had low sequence compositions ($\leq 0.1\%$) or were not detected in the intermediate and effluent stages.

Nitrification is a fundamental process in the biological removal of nitrogen in WWTPs, and this two-step process is carried out by ammonia-oxidising bacteria (AOB) that convert ammonia to nitrite, then nitrite-oxidising bacteria (NOB) convert nitrite to nitrate (Bellucci and Curtis, 2011). *Nitrosomonas* and *Nitrospira* are two important genera of AOB in WWTPs, while *Nitrobacter* is a major NOB (Siripong et al., 2007). In the present study, *Nitrosomonas* and *Nitrobacter* were not detected, and *Nitrospira* spp. were only detected in the intermediate and effluent of WWTP 4 (sequence compositions 1.2% and 1.5%, respectively) (Table 5). Rhodocyclales are a widespread and abundant order of bacteria in WWTPs responsible for anaerobic nitrogen removal by denitrification (Yang et al., 2011). In the present study, 12 Rhodocyclales genera were identified: *Azoarcus* spp., *Azonexus* spp., *Azospira* spp., *Dechloromonas* spp., *Methyloversatilis* sp., *Propionivibrio* spp., *Rhodocyclaceae* spp., *Sterolibacterium* spp., *Sulfuritalea* sp., *Thauera* spp., *Uliginosibacterium* spp. and *Zoogloea* spp. (Table 5). In WWTP 1, Rhodocyclales were highest in the influent (*Thauera* spp. had the highest composition; 3.4%) and rare ($\leq 0.1\%$) or not detected in the intermediate and effluent samples. Rhodocyclales compositions were low in the influent (0.8%) and effluent (1.0%) of WWTP 2. WWTP 3 had higher Rhodocyclales compositions in the intermediate (13.2%; most abundant was *Thauera* spp. at 6.0%)

compared to the influent (4.3%), and no Rhodocyclales sequences were detected in WWTP 3 effluent. In addition to denitrification, certain *Thauera* and *Dechloromonas* strains can degrade oil derivatives such as toluene (Shinoda et al., 2004; Chakraborty et al., 2005) and therefore may be important in reducing the ecological burden of these aromatic compounds, but we were unable to identify the species and strains of these genera based on V4 16S amplicons. Unlike the WSPs, the oxidation ditch plant WWTP 4 had high Rhodocyclales compositions in both the intermediate and effluent (7.7% and 7.1%, respectively). Members of the family Comamonadaceae are also denitrifiers and are responsible for aromatic degrading processes (Xu et al., 2018). The nine Comamonadaceae genera identified were *Aquabacterium* sp., *Brachymonas* (*Brachymonas denitrificans*), *Comamonas* sp., *Delftia* sp., *Flavobacterium* spp., *Hydrogenophaga* spp., *Polaromonas* spp., *Rhodoferax* spp. and *Rubrivivax* spp. Comamonadaceae compositions in WWTP 1 were similar to those observed for Rhodocyclales in this treatment plant, with the highest composition observed in the influent (7.4%; *Comamonas* sp. had the highest composition of 6.9%) and compositions were low in the intermediate and effluent (1.6% and 1.7%, respectively). Comamonadaceae compositions were much higher than Rhodocyclales in WWTP 2, which had 4.7% in the influent and 7.7% in the effluent, and *Flavobacterium* spp. had the greatest sequence compositions in both influent (2.8%) and effluent (4.7%). For WWTP 3, the Comamonadaceae compositions were similar to Rhodocyclales in the influent and intermediate, but unlike Rhodocyclales, were detected (mostly *Flavobacterium* spp. 7.6%) in the effluent. For WWTP 4, the Comamonadaceae were mostly comprised of *Comamonas* sp. in the influent (4.4%) and *Flavobacterium* spp. (6.5%) in the effluent, and the composition of Comamonadaceae was low in the intermediate (1.2%). Comamonadaceae, Rhodocyclaceae, Flavobacteriaceae and Pseudomonadaceae also play important roles in flocculation in

activated sludge plants, and Comamonadaceae and Flavobacteriaceae are important for bulking and foaming (Shchegolkova et al., 2016).

5. Conclusions

In the present study, a total of 36 pathogenic or potentially pathogenic species were detected, but most could not be identified to species level. Of these, sequences belonging to 14 medically important genera that could possibly be from pathogens were identified primarily in the influent of WWTPs 1-4. In almost all cases, these bacteria were present in lower abundance in the effluent with the exception of *Aeromonas* sp. in the effluent of WWTP 1 (8.6%). The use of V4 16S NGS for bacterial pathogen identification has significant limitations for species level identification including the inability to differentiate Enterobacteriaceae genera that contain many important enteric pathogens of humans. Amplicon NGS is a useful tool for broad taxonomic surveys of bacteria, while tools such as quantitative PCR and droplet digital PCR could be used in follow-up studies to identify bacteria that could not be differentiated at the species or strain level. This would also allow quantification of pathogens before and after the wastewater treatment process. Future studies that aim for improved taxonomic resolution of bacterial pathogens in wastewater should consider sequencing full length 16S and more variable housekeeping genes such as *gyrB*, *rpoB* or *mdh* for differentiation of *E. coli* and *Shigella* strains. Shallow shotgun sequencing can also be used for pathogen identification and for gaining functional information that is important for public health.

Nitrifying, denitrifying and floc-forming bacteria could mostly be identified to the genus level. Only the activated sludge oxidation ditch plant showed the presence of an AOB, *Nitrospira* spp., for bacterial nitrification. However, both the lower technology WSPs and the activated sludge oxidation ditch plant showed the presence of Rhodocyclales, Comamonadaceae, Flavobacteriaceae and Pseudomonadaceae bacteria, which are responsible

for anaerobic nitrogen removal by denitrification (i.e. conversion of nitrate to nitrogen gas). These bacteria are also important for WWTP performance since they assist floc formation. Our current work is examining the presence, diversity and relative abundances of bacterial communities responsible for the nitrification and denitrification cycle in WSPs (e.g. *Nitrobacter*, *Nitrosomonas*, *Nitrospira*, *Nitrosococcus* and *Nitrosomonas*) using functional genes that encode key enzymes (*amoA*, *njfH*, *nirK*, *nosZ*, *norB*, *nxrB*, *narG*, *napA* and *nrfA*). This will help us to better understand the correlations between the concentrations of selected nitrogenous species present in wastewater and their contribution to the nitrogen cycle in WSPs.

Other limitations include the misidentification of 16S sequences from chloroplasts as Cyanobacteria by the Greengenes database. Due to the discrepancies between taxonomic assignments with Greengenes and the NCBI nr/nt database, we recommend that future studies use the Greengenes database for 16S NGS taxonomic assignment with caution and compare OTU or ZOTU sequences with the NCBI nr/nt database to validate taxonomic assignments.

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References

- Ahmed W., Staley C., Sidhu J., Sadowsky M., Toze S., 2017. Amplicon-based profiling of bacteria in raw and secondary treated wastewater from treatment plants across Australia. *Appl Microbiol Biotechnol.* 101(3), 1253-1266.
- Anonymous, 2009. Ponds for stabilising organic matter. www.water.wa.gov.au/__data/assets/pdf_file/0005/4100/84601.pdf
- Apprill, A., McNally, S., Parsons, R., Weber, L., 2015. Minor revision to V4 region SSU rRNA 806R gene primer greatly increases detection of SAR11 bacterioplankton. *Aquat. Microb. Ecol.* 75, 129-137.
- Baars, J.K., 1962. The use of oxidation ditches for treatment of sewage for small communities. *Bull. World Health Organ.* 26, 465-474.
- Balvočiūtė, M., Huson, D.H., 2017. SILVA, RDP, Greengenes, NCBI and OTT—how do these taxonomies compare? *BMC Genomics.* 18, 114.
- Beilfuss, H.A., Quig, D., Block, M.A., Schreckenberger, P.C., 2015. Definitive identification of *Laribacter hongkongensis* acquired in the United States. *J. Clin. Microbiol.* 53, 2385-2388.
- Bellucci, M., Curtis, T.P., 2011. Ammonia-oxidizing bacteria in wastewater. *Methods Enzymol.* 496, 269-286.
- Berlec, A., 2012. Novel techniques and findings in the study of plant microbiota: Search for plant probiotics. *Plant Sci.* 193-194,96-102.
- Bhaduri, S., Wesley, I.V., Bush, E.J., 2005. Prevalence of pathogenic *Yersinia enterocolitica* strains in pigs in the United States. *Appl Environ Microbiol.* 71, 7117-7121.

- Bokulich, N.A., Kaehler, B.D., Rideout, J.R., Dillon, M., Bolyen, E., Knight, R., Huttley, G.A., Caporaso, J.G., 2018. Optimizing taxonomic classification of marker-gene amplicon sequences with QIIME 2's q2-feature-classifier plugin. *Microbiome*. 6, 90.
- Brosius, J., Palmer, M. L., Kennedy, P. J., Noller, H. F., 1978. Complete nucleotide sequence of a 16S ribosomal RNA gene from *Escherichia coli*. *Proc Natl Acad Sci U S A*. 75, 4801-4805.
- Cabral, J.P., 2010. Water microbiology. Bacterial pathogens and water. *Int J Environ Res Public Health*. 7, 3657-3703.
- Caporaso, J.G., Kuczynski, J., Stombaugh, J., Bittinger, K., Bushman, F.D., Costello, E.K., Fierer, N., Peña, A.G., Goodrich, J.K., Gordon, J.I., Huttley, G.A., Kelley, S.T., Knights, D., Koenig, J.E., Ley, R.E., Lozupone, C.A., McDonald, D., Muegge, B.D., Pirrung, M., Reeder, J., Sevinsky, J.R., Turnbaugh, P.J., Walters, W.A., Widmann, J., Yatsunenko, T., Zaneveld, J., Knight, R., 2010. QIIME allows analysis of high-throughput community sequencing data. *Nat Methods*. 7, 335-336.
- Caporaso, J. G., Lauber, C. L., Walters, W. A., Berg-Lyons, D., Lozupone, C. A., Turnbaugh, P. J., Noah Fierer, N., Knight, R., 2011. Global patterns of 16S rRNA diversity at a depth of millions of sequences per sample. *Proc. Natl. Acad. Sci. USA*. 108, 4516-4522.
- Chakraborty, R., O'Connor, S.M., Chan, E., Coates, J.D., 2005. Anaerobic degradation of benzene, toluene, ethylbenzene, and xylene compounds by *Dechloromonas* strain RCB. *Appl. Environ. Microbiol.* 71, 8649-8655.
- Chakravorty, S., Helb, D., Burday, M., Connell, N., Alland, D., 2007. A detailed analysis of 16S ribosomal RNA gene segments for the diagnosis of pathogenic bacteria. *J Microbiol Methods*. 69, 330-339.

- Connor, R., Renata, A., Ortigara, C., Koncagül, E., Uhlenbrook, S., Lamizana-Diallo, B.M., Zadeh, S.M., Qadir, M., Kjellén, M., Sjödin, J., Hendry, S., 2017. The United Nations world water development report. Wastewater: the untapped resource. Paris, UNESCO. <https://reliefweb.int/sites/reliefweb.int/files/resources/247153e.pdf>.
- Cydzik-Kwiatkowska, A., Zielińska, M., 2016. Bacterial communities in full-scale wastewater treatment systems. *World J Microbiol Biotechnol.* 32, 66.
- Devanga Ragupathi, N.K., Muthuirulandi Sethuvel, D.P., Inbanathan, F.Y., Veeraraghavan, B., 2017. Accurate differentiation of *Escherichia coli* and *Shigella serogroups*: challenges and strategies. *New Microbes New Infect.* 21, 58-62.
- Edgar, R.C., 2010. Search and clustering orders of magnitude faster than BLAST. *Bioinformatics.* 26, 2460-2461.
- Edgar, R.C., 2016. UNOISE2: improved error-correction for Illumina 16S and ITS amplicon sequencing. *bioRxiv.* 081257.
- Eland, L.E., Davenport, R.J., Santos, A.B., Mota Filho, C.R., 2018. Molecular evaluation of microalgal communities in full-scale waste stabilisation ponds. *Environ. Technol.* In press. doi: 10.1080/09593330.2018.1435730.
- Fukushima, M., Kakinuma, K., Kawaguchi, R., 2002. Phylogenetic analysis of *Salmonella*, *Shigella*, and *Escherichia coli* strains on the basis of the *gyrB* gene sequence. *J Clin Microbiol.* 40, 2779-2785.
- Garrido-Cardenas, J.A., Polo-López, M.I., Oller-Alberola, I., 2017. Advanced microbial analysis for wastewater quality monitoring: metagenomics trend. *Appl. Microbiol. Biotechnol.* 101, 7445-7458.

- Goodwin, S., McPherson, J.D., McCombie, W.R., 2106. Coming of age: ten years of next-generation sequencing technologies. *Nat Rev Genet.* 17, 333-351.
- Guentzel, M.N., 1996. *Escherichia, Klebsiella, Enterobacter, Serratia, Citrobacter, and Proteus.*
In: Baron S, editor. *Medical Microbiology.* 4th edition. Galveston (TX): University of Texas Medical Branch at Galveston.
- Guindon, S., Dufayard, J.F., Lefort, V., Anisimova, M., Hordijk, W., Gascuel, O., 2010. New algorithms and methods to estimate maximum-likelihood phylogenies: assessing the performance of PhyML 3.0. *Syst. Biol.* 59, 307-321.
- Haft, D.H., Tovchigrechko, A., 2012. High-speed microbial community profiling. *Nat Methods.* 9, 793-794.
- Hillmann, B., Al-Ghalith, G. A., Shields-Cutler, R.R., Zhu, Q., Gohl, D.M., Beckman, K.B., Knight, R., Knights, D., 2018. Evaluating the information content of shallow shotgun metagenomics. *mSystems.* 3, e00069-00018.
- Ho, L.T., Van Echelpoel, W., Goethals, P.L.M., 2017. Design of waste stabilization pond systems: a review. *Water Res.* 123, 236-248.
- Ibal, J.C., Pham, H.Q., Park, C.E., Shin, J-H., 2019. Information about variations in multiple copies of bacterial 16S rRNA genes may aid in species identification. *PLoS One.* 14, e0212090.
- Inaba, T., Hori, T., Aizawa, H., Ogata, A., Habe, H., 2017. Architecture, component, and microbiome of biofilm involved in the fouling of membrane bioreactors. *NPJ Biofilms Microbiomes.* 3, 5.

- Katoh, K., Misawa, K., Kuma, K.-i., Miyata, T., 2002. MAFFT: a novel method for rapid multiple sequence alignment based on fast Fourier transform. *Nucleic Acids Res.* 30, 3059-3066.
- Kayman, T., Abay, S., Hizlisoy, H., Atabay, H.İ., Diker, K.S., Aydin, F., 2012. Emerging pathogen *Arcobacter* spp. in acute gastroenteritis: molecular identification, antibiotic susceptibilities and genotyping of the isolated arcobacters. *J Med Microbiol.* 61, 1439-1444.
- Kearse, M., Moir, R., Wilson, A., Stones-Havas, S., Cheung, M., Sturrock, S., Buxton, S., Cooper, A., Markowitz, S., Duran, C., Thierer, T., Ashton, B., Meintjes, P., Drummond, A., 2012. Geneious Basic: an integrated and extendable desktop software platform for the organization and analysis of sequence data. *Bioinformatics.* 28, 1647-1649.
- Kembel, S.W., Wu, M., Eisen, J.A. and Green, J.L., 2012. Incorporating 16S gene copy number information improves estimates of microbial diversity and abundance. *PLoS computational biology*, 8(10), p.e1002743.
- Kim, Y., Koh, I., Rho, M., 2015. Deciphering the human microbiome using next-generation sequencing data and bioinformatics approaches. *Methods.* 79–80, 52-59.
- Kimura, M., 1980. A simple method for estimating evolutionary rate of base substitutions through comparative studies of nucleotide sequences. *J. Mol. Evol.* 16, 111-120.
- Levican, A., Collado, L., Aguilar, C., Yustes, C., Diéguez, A.L., Romalde, J.L. Figueras, M.J., 2012. *Arcobacter bivalviorum* sp. nov. and *Arcobacter venerupis* sp. nov., new species isolated from shellfish. *Syst Appl Microbiol.* 35, 133-138.
- Lu, X., Zhang, X.-X., Wang, Z., Huang, K., Wang, Y., Liang, W., Tan, Y., Liu, B., Tang, J., 2015. Bacterial pathogens and community composition in advanced sewage treatment

- systems revealed by metagenomics analysis based on high-throughput sequencing. *PLoS One*. 10, e0125549.
- McDonald, D., Price, M.N., Goodrich, J., Nawrocki, E.P., DeSantis, T.Z., Probst, A., Andersen, G.L., Knight, R., Hugenholtz, P., 2012. An improved Greengenes taxonomy with explicit ranks for ecological and evolutionary analyses of bacteria and archaea. *ISME J.* 6, 610.
- McLellan, S., Huse, S., Mueller Spitz, S., Andreishcheva, E., Sogin, M., 2010. Diversity and population structure of sewage derived microorganisms in wastewater treatment plant influent. *Environ. Microbiol.* 12, 378-392.
- Oksanen, J., Blanchet, F.G., Friendly, M., Kindt, R., Legendre, P., McGlinn, D., Minchin, P.R., O'Hara, R. B., Simpson, G.L., Solymos, P., Henry, M., Stevens, H., Szoecs, E., Wagner, H., 2018. *Vegan: Community ecology package*. R package version 2.5-2. <https://CRAN.R-project.org/package=vegan>.
- Parada, A.E., Needham, D.M., Fuhrman, J.A., 2015. Every base matters: assessing small subunit rRNA primers for marine microbiomes with mock communities, time series and global field samples. *Environ. Microbiol.* 18, 1403-1414.
- R Core Team, 2013. *R: A language and environment for statistical computing*. R Foundation for Statistical Computing, Vienna, Austria. URL <http://www.R-project.org/>.
- Rajasulochana, R., Preethy, V., 2016. Comparison on efficiency of various techniques in treatment of waste and sewage water – a comprehensive review. *Resource-Efficient Technologies.* 2, 175-184.
- Rodriguez-Manzano, J., Alonso, J.L., Ferrus, M.A., Moreno, Y., Amoros, I., Calgua, B., Hundesa, A., Guerrero-Latorre, L., Carratala, A., Rusinol, M., Girones, R., 2012. Standard and new faecal indicators and pathogens in sewage treatment plants,

- microbiological parameters for improving the control of reclaimed water. *Water Sci. Technol.* 66, 2517-2523.
- Safton, S., Cooper, G., Harrison, M., Wright, L. and Walsh, P., 1999. *Neisseria canis* infection: a case report. *Commun Dis Intell.* 23, 221.
- Sanapareddy, N., Hamp, T.J., Gonzalez, L.C., Hilger, H.A., Fodor, A.A., Clinton, S.M., 2009. Molecular diversity of a North Carolina wastewater treatment plant as revealed by pyrosequencing. *Appl Environ Microbiol.* 75, 1688-1696.
- Seng, P., Boushab, B.M., Romain, F., Gouriet, F., Bruder, N., Martin, C., Paganelli, F., Bernit, E., Le Treut, Y.P., Thomas, P., Papazian, L., 2016. Emerging role of *Raoultella ornithinolytica* in human infections: a series of cases and review of the literature. *Int J Infect Dis.* 45, 65-71.
- Sharafi, K., Davil, M.F., Heidari, M., Almasi, A., Taheri, H. 2012. Comparison of conventional activated sludge system and stabilization pond in removal of chemical and biological parameters. *Int. J. Environ. Health Eng.* 1, 1-5.
- Shchegolkova, N.M., Krasnov, G.S., Belova, A.A., Dmitriev, A.A., Kharitonov, S.L., Klimina, K.M., Melnikova, N.V., Kudryavtseva, A.V., 2016. Microbial community structure of activated sludge in treatment plants with different wastewater compositions *Front Microbiol.* 7, 90.
- Shinoda, Y., Sakai, Y., Uenishi, H., Uchihashi, Y., Hiraishi, A., Yukawa, H., Yurimoto, H., Kato, N., 2004. Aerobic and anaerobic toluene degradation by a newly isolated denitrifying bacterium, *Thauera* sp. strain DNT-1 *Appl. Environ. Microbiol.* 70, 1385-1392.

- Siripong, S. and Rittmann, B.E., 2007. Diversity study of nitrifying bacteria in full-scale municipal wastewater treatment plants. *Water Res.* 41, 1110-1120.
- Stamatakis, A., 2014. RAxML version 8: a tool for phylogenetic analysis and post-analysis of large phylogenies. *Bioinformatics.* 30, 1312-1313.
- Verbyla, M., von Sperling, M., Maiga, Y., 2017. Waste Stabilization Ponds. In: J.B. Rose and B. Jiménez-Cisneros, (eds) *Global Water Pathogens Project*. <http://www.waterpathogens.org> (C. Haas, J.R. Mihelcic and M.E. Verbyla) (eds) Part 4 Management of risk from Excreta and Wastewater) www.waterpathogens.org/book/waste-stabilization-ponds Michigan State University, E. Lansing, MI, UNESCO.
- Von Sperling, M., 2007. *Waste stabilisation ponds*. IWA Publishing. www.iwapublishing.com/sites/default/files/ebooks/9781780402109.pdf.
- Xu, S., Yao, J., Ainiwaer, M., Hong, Y., Zhang, Y., 2018. Analysis of bacterial community structure of activated sludge from wastewater treatment plants in winter. *Biomed Res Int.* 2018, 8278970.
- Yang, C., Zhang, W, Liu, R, Li, Q, Li, B, Wang, S, Song, C, Qiao, C, Mulchandani, A., 2011. Phylogenetic diversity and metabolic potential of activated sludge microbial communities in full-scale wastewater treatment plants. *Environ Sci Technol.* 45, 7408-7415.
- Ye, L., Zhang, T., 2011. Pathogenic bacteria in sewage treatment plants as revealed by 454 pyrosequencing. *Environ Sci Technol* 45, 7173-7179.
- Zhang, T., Shao, M.-F., Ye, L., 2012. 454 pyrosequencing reveals bacterial diversity of activated sludge from 14 sewage treatment plants. *ISME J.* 6, 1137-1147.

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Figure Legends

Figure 1. WWTP localities and different treatment stages sampled.

Figure 2. 16S NGS sequence percent composition plot of phyla (P) and classes (C) detected in different treatment stages of wastewater sampled from WWTPs 1-4. Treatment stages include influent (I), intermediate (INT) and effluent (E). Phyla with $\leq 10\%$ overall sequence composition are grouped as “other”.

Figure 3. Maximum likelihood tree of a 247 bp alignment (gaps excluded) of genomic 16S Enterobacteriaceae sequences trimmed to the V4 region. The seven Enterobacteriaceae ZOTU sequences derived from this study are in bold typeface. Values at nodes indicate Bootstrap values from 1,000 replicates. Outgroup of tree *Vibrio cholerae* (2614873) not shown.

Appendices

Appendix A. Supplementary Figures

Figure A.1. Alpha rarefaction plot of 16S sequencing depth and ZOTUs detected in WWTP samples prior to low read abundance ($<0.05\%$) filtering.

Figure A.2. Alpha rarefaction plots of 16S sequencing depth and ZOTUs detected prior to low read abundance ($<0.05\%$) filtering for WWTPs 1-4 and treatment stages.

Figure A.3. Alpha rarefaction plot of 16S sequencing depth and ZOTUs detected in WWTP samples after low read abundance ($<0.05\%$) filtering.

Figure A.4. Alpha rarefaction plots of 16S sequencing depth and ZOTUs detected after low read abundance ($<0.05\%$) filtering for WWTPs 1-4 and treatment stages.

Appendix B. Supplementary Data

Supplementary File B.1. List of 3,598 16S V4 region ZOTU sequences generated by this study.

Supplementary File B.2. Sequence totals and compositions.

Supplementary File B.3. Pairwise genetic distance matrix of the 247 bp alignment (gaps excluded) of genomic 16S Enterobacteriaceae sequences trimmed to the V4 region that was used to construct the phylogenetic tree in Figure 3.

Supplementary File B.4. Comparison of Greengenes and NCBI nr/nt database taxa to ZOTUs potentially from pathogenic bacteria.

Supplementary File B.5. Comparison of Greengenes and NCBI nr/nt database taxa to ZOTUs from nitrifying, denitrifying and floc-forming bacteria.

Supplementary File B.6. Chloroplast sequences in the Greengenes 99 OTU taxonomy file renamed with the kingdom “Chloroplast”.

Table 1. Rural wastewater treatment plant samples analysed in the present study.

WWTP	Treatment technology	Location	Climate	Sample ID	Wastewater treatment stage	Sample collection date; season
WWTP 1	Stabilisation pond: Combined anaerobic and aerobic pond system, followed by two maturation ponds	Northwest Western Australia	Tropical climate. Wet and dry seasons.	WWTP 1-1	Influent	19-Feb-2015; Wet
				WWTP 1-2	Effluent (pre-chlorination)	
				WWTP 1-3	Effluent (post-chlorination)	
				WWTP 1-4	Influent	7-Sep-2015; Dry
				WWTP 1-5	Intermediate (post maturation pond 1)	
				WWTP 1-6	Intermediate (post maturation pond 2)	
				WWTP 1-7	Effluent (pre-chlorination)	
				WWTP 1-8	Effluent (post-chlorination)	
WWTP 2	Stabilisation	Wheatbelt,	Hot dry	WWTP 2-1	Influent	12-Feb-2015;

	pond: One facultative pond	Western Australia	summers and mild winters. Four distinct seasons.			Summer
				WWTP 2-2	Effluent (final pond)	
				WWTP 2-3	Effluent (storage basin)	
				WWTP 2-4	Influent	13-Jul-2015; Winter
				WWTP 2-5	Effluent (final pond)	
				WWTP 2-6	Effluent (storage basin)	
WWTP 3	Stabilisation pond: Two primary facultative ponds, and one secondary pond	Southwest Western Australia	Temperate climate. Four distinct seasons.	WWTP 3-1	Influent	23-Feb-2015; Summer
				WWTP 3-2	Intermediate (post-pond)	
				WWTP 3-3	Effluent	
				WWTP 3-4	Influent	14-July-2015; Winter
				WWTP 3-5	Intermediate (post-pond)	
				WWTP 3-6	Effluent	
WWTP 4	Activated sludge: Oxidation ditches	Southwest Western Australia	Temperate climate. Four distinct seasons.	WWTP 4-1	Influent	23-Feb-2015; Summer
				WWTP 4-2	Intermediate (oxidation	

	followed by sedimentation tanks				ditch)	
				WWTP 4-3	Effluent	
				WWTP 4-4	Influent	14-July-2015; Winter
				WWTP 4-5	Intermediate (oxidation ditch)	
				WWTP 4-6	Effluent	

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Table 2. V4 16S NGS sequence statistics.

Statistics	Raw (unprocessed)	Pre-processed ^a	Processed 16S sequences ^b			
	Grand total (<i>n</i> = 34)		Samples (<i>n</i> = 26)	Extraction controls (<i>n</i> = 6)	NTCs (<i>n</i> = 2)	Grand total (<i>n</i> = 34)
Average	27,965	23,805	26,746	8	8	20,454
Standard deviation	27,254	24,239	20,608	7	2	21,314
Min	2,646	2	4,681	2	6	2
Max	182,113	95,135	85,305	21	9	85,305
Total	1,426,191	809,368	695,400	48	19	695,463

^aMerged, quality filtered sequences with singletons and chimeras removed

^bMerged, quality filtered sequences with singletons, chimeras, unassigned sequences and low abundance sequences (<0.05%) removed

Table 3. Enterobacteriaceae (Gammaproteobacteria: Enterobacteriales) ZOTUs Greengenes assigned taxonomy cross-checked against the NCBI nr/nt database.

ZOTU no.	Accession no.	Final taxonomy	Greengenes results		NCBI nr/nt results			Correct Greengenes taxonomy?		
			Assigned taxonomy	Confidence scores ^a	GenBank ® accession no.	Species	Percent identity	Family	Genus	Species
28	MH89261 5	Enterobac teriaceae sp.	Enterobac teriaceae sp.	0.95	MH38442 6	<i>Enterobac ter xiangfang ensis</i>	100	✓	*	*
					MH19022 0	<i>Erwinia aphidicol a</i>	100	✓	*	*
					MH41122 0	<i>Klebsiella pneumoni ae</i>	100	✓	*	*
54	MH89262 2	Enterobac teriaceae sp.	<i>Escherich ia coli</i>	0.96	MH39673 7	<i>Escherich ia coli</i>	100	✓	✗	✗
					MH35216 4	<i>Salmonell a enterica subsp. enterica</i>	100	✓	✗	✗
					MH37132 7	<i>Shigella flexneri</i>	100	✓	✗	✗
170	MH89263 7	Enterobac teriaceae sp.	<i>Citrobact er sp.</i>	0.95	NR15605 2	<i>Citrobact er europaeus</i>	100	✓	✗	*
					MH37132 2	<i>Citrobact er freundii</i>	100	✓	✗	*
					MH35220 5	<i>Salmonell a enterica subsp. enterica</i>	100	✓	✗	*

183	MH89263 8	Enterobac teriaceae sp.	Enterobac teriaceae sp.	0.94	CP020089	<i>Enterobac ter cloacae</i>	100	✓	*	*
					MF36001	<i>Klebsiella michigane nsis</i>	100	✓	*	*
					MH19634	<i>Klebsiella oxytoca</i>	100	✓	*	*
417	MH89265 6	Enterobac teriaceae sp.	Enterobac teriaceae sp.	0.92	MG89020	<i>Leclercia adecarbox ylata</i>	100	✓	*	*
					MG02265	<i>Raoultella electrica</i>	100	✓	*	*
					MG51611	<i>Raoultella ornithinol ytica</i>	100	✓	*	*
546	MH89266 3	Enterobac teriaceae sp.	<i>Trabulsiel la</i> sp.	1	MH08545	<i>Citrobact er amalonati cus</i>	100	✓	✗	*
					MF18660	<i>Citrobact er farmeri</i>	100	✓	✗	*
					MH16920	<i>Kosakoni a oryzendop hytica</i>	100	✓	✗	*
619	MH89267 2	Enterobac teriaceae sp.	Enterobac teriaceae sp.	1	MH14147	<i>Cronobac ter sakazakii</i>	100	✓	*	*
					MH16920	<i>Kluyvera georgiana</i>	100	✓	*	*
					MG89020	<i>Pseudocit robacter faecalis</i>	100	✓	*	*

*Taxon was unassigned, which was the correct choice based on BLAST results.

^aConfidence scores are probabilities generated by the naïve Bayes algorithm implemented by QIIME 2 feature classifier (<https://docs.qiime2.org/2018.6/tutorials/feature-classifier/>).

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Table 4. Sequence composition (%) of pathogens and possible pathogens in WWTPs 1-4 influent (I), intermediate (INT) and effluent (E) with taxonomy confirmed with Greengenes and NCBI nr/nt sequence databases.

						WWTP 1			W WTP 2		WWTP 3			WWTP 4		
Class	Order	Family	Taxonomic assignment ^a	ZOT U no.	Accession no.	I	INT	E	I	E	I	INT	E	I	INT	E
Actinobacteria																
Actinobacteria	Actinobacteria	Actinobacteria	<i>Actinobacteria</i>	603	H892	0.1	-	-	-	-	-	-	-	-	-	-
				704	H892	-	2.5	2.2	0.1	-	-	0.1	0.3	-	0.2	0.1
				6;	H892											
				332;	617;											
				588;	MH8											
				1469;	9265											
				1651;	1;											
				1756;	MH8											
				1801	9266											
				8;												
					MH8											
				9269												
				2;												
					MH8											
				9269												
				6;												
					MH8											

					0											
Firmicutes																
acilli	actobacillales	treptococcales	actobacillales	1; 134; 443	H892 616; MH8 9263 2; MH8 9265 7	1.5	-	-	-	-	2.3	-	-	1.9	0.2	0.2
			treptococcales	10	H892 678	0.1	-	-	-	-	-	-	-	-	-	-
			treptococcus	; 559	H892 611; MH8 9266 5	12.1	-	0.2	-	-	2.6	-	-	3.2	0.4	0.2
lostridiales	lostridiales	lostridiales	lostridiales	57; 1161	H892 652; MH8 9268 4	-	-	0.2	0.1	-	-	-	-	-	-	-
			uminococcales	387	H892 688	<0.1	-	-	-	-	-	-	-	-	-	-
						0.1	0.4	-	-	-	0.1	-	-	0.1	-	-

			lostri diale s spp.	59; 460; 1119; 1967	H892 653; MH8 9265 9; MH8 9268 3; MH8 9270 1											
Bacilli	Lactobacillales	nterococcaeae	<i>nterococcus</i> spp.	86; 288; 1246	H892 646; MH8 9264 8; MH8 9268 6	0.2	-	0.1	-	-	0.2	-	-	0.3	-	-
roteobacteria																
etaproteobacteria	urkholderiales	lcaligenaceae	lcaligenaceae sp.	9	H892 619	-	-	<0.1	0.2	2.0	-	-	-	-	-	-
	eisseriales	eisseriaceae	<i>aribaacter hongkongensis</i>	5	H892 623	0.8	-	-	-	-	0.9	0.1	-	0.7	0.1	0.1

			<i>eisseriacanis</i>	4	H892 626	1.1	-	-	-	-	0.6	-	-	0.3	<0.1	0.1
			<i>eisseria</i> sp.	251	H892 687	<0.1	-	-	-	-	-	-	-	-	-	-
			<i>eisseriaceae</i> spp.	9; 197; 287; 960	H892 614; MH8 9263 9; MH8 9264 7; MH8 9267 9	0.9	-	-	0.1	-	4.2	0.3	-	4.0	0.2	0.1
			<i>itreoscilla</i> spp.	7; 169; 466	H892 629; MH8 9263 6; MH8 9266 0	0.2	-	-	-	-	1.0	0.1	-	1.0	-	-
psiloproteobacteria	ampylobacteriales	ampylobacteraceae	<i>rcobacter</i> spp.	18; 289; 598; 1174	H892 642; MH8 9264 9;	0.3	-	-	-	-	0.6	-	-	0.1	-	-

					MH8 9267 0; MH8 9268 5											
			<i>rcob acter vener upis</i>	14	H892 641	0.7	-	-	-	-	-	-	-	-	-	
			<i>ampy lobac terac eae spp.</i>	; 37; 59; 158; 229	H892 609; MH8 9261 8; MH8 9262 4; MH8 9263 5; MH8 9264 3	13.4	0.2	1.6	0.9	0.1	14.9	1.9	-	20.7	3.8	0.7
<i>amm aprot eoba cteria</i>	<i>erom onad ales</i>	<i>erom onad aceae</i>	<i>erom onad aceae spp.</i>	83; 639; 1476	H892 661; MH8 9267 3; MH8 9269 3	0.2	-	-	-	-	0.1	-	-	-	-	-

			<i>erom</i>			6.0	2.2	8.6	0.4	0.2	4.7	1.1	4.8	2.8	0.5	0.8
			<i>onas</i> sp.		H892 610											
	ntero bacte riales	ntero bacte riace ae	ntero bacte riace ae spp.	8; 54; 170; 183; 417; 546; 619	H892 615; MH8 9262 2; MH8 9263 7; MH8 9263 8; MH8 9265 6; MH8 9266 3; MH8 9267 2	3.7	0.2	0.6	-	-	2.6	-	0.6	2.4	0.1	0.3
	egion ellale s	oxiell aceae	oxiell aceae sp.	92	H892 677	-	-	-	-	-	-	-	-	-	-	0.4
			egion ellale s sp.	079	H892 677	-	-	-	-	<0.1	-	-	-	-	-	-
		egion	<i>egion</i>	554	H892	-	-	-	-	-	-	-	-	-	-	0.1

		ellac eae	<i>ella</i> sp.		703											
	seud omon adale s	oraxe llace ae	<i>cinet</i> <i>obact</i> <i>er</i> spp.	0; 16; 41; 45; 75; 101; 283; 317; 564; 584; 886; 991; 992 7; 9263 0; MH8 9264 5; MH8 9265 0; MH8 9266 6; MH8 9266 7; MH8 9267 6;	H892 612; MH8 9261 3; MH8 9262 0; MH8 9262 1; MH8 9262 MH8 9263 0; MH8 9264 5; MH8 9265 0; MH8 9266 6; MH8 9266 7; MH8 9267 6;	1.0	-	0.1	-	-	8.5	17.1	-	12.8	0.7	0.7

					MH8 9268 1; MH8 9268 2											
		seud omon adace ae	<i>seud</i> <i>omon</i> <i>as</i> spp.	5; 77; 109; 147; 151; 210; 233; 361; 402; 515; 556; 607; 678; 722; 1402	H892 625; MH8 9262 8; MH8 9263 1; MH8 9263 3; MH8 9263 4; MH8 9264 0; MH8 9264 4; MH8 9265 4; MH8 9265 5; MH8 9266	1.1	-	1.3	0.1	1.3	0.3	3.1	1.4	0.6	-	.1

					2; MH8 9266 4; MH8 9267 1; MH8 9267 4; MH8 9267 5; MH8 9268 9											
			amm aprot eoba cteria spp.	440; 1709	H892 690; MH8 9269 7	-	-	-	-	-	-	-	-	-	0.2	0.1
	ibrio nales	seud oalter omon adace ae	<i>ibrio</i> sp.	89	H892 669	-	-	-	-	0.1	-	-	-	-	-	-
Spirochaetes																
eptos pirae	eptos pirale s	eptos pirac eae	eptos pirac eae sp.	146	H892 702	-	-	-	-	-	-	-	-	-	0.1	-
						-	-	-	-	-	-	-	-	-	<0.1	-

			piroc haete s sp.	988	H892 706											
			<i>urner</i> <i>iella</i> sp.	946	H892 700	-	-	-	-	-	-	-	-	-	0.1	0.1
piroc haeti a	piroc haeta les	piroc haeta ceae	piroc haeta ceae spp.	45; 965; 1564	H892 658; MH8 9268 0; MH8 9269 4	-	0.1	0.1	0.1	-	-	-	-	-	-	-

^aMost specific level of taxonomy designated after comparing ZOTUs to Greengenes and NCBI nr/nt databases.

Table 5. Sequence composition (%) of nitrifying, denitrifying and floc-forming bacteria in WWTPs 1-4 influent (I), intermediate (INT) and effluent (E) with taxonomy confirmed with Greengenes and NCBI nr/nt sequence databases.

						W			W			W				
						WTP 1			WTP 2		WTP 3		WTP 4			
Class	Order	Family	OT	Accessio n no.	Species	I	N	E	I	E	I	IN	E	I	IN	E
							T				T			T		
Bacteroidetes																
Flavobacteri ia	Flavobacte riales	Flavobacter iaceae	35;	MH892	<i>Flavobacteri</i>	-	0.	0.	2.	4.	-	0.	7.	0.	0.	6.
			44;	717;	<i>um spp.</i>	8	4	8	7	2	6	1	4	5		
			57;	MH892												
			103	718;												
			;	MH892												
			153	720;												
			;	MH892												
			155	728;												
			;	MH892												
			172	733;												
			;	MH892												
			178	734;												
			;	MH892												
			263	737;												
			;	MH892												
			365	738;												
;	MH892															
474	745;															
;	MH892															
660	752;															
;	MH892															
766	763;															
;	MH892															
970	771;															

				; MH892														
			114	779;														
			2;	MH892														
			124	792;														
			1;	MH892														
			188	798;														
			3;	MH892														
			191	800;														
			2;	MH892														
			204	811;														
			2;	MH892														
			224	813;														
			2;	MH892														
			234	815;														
			9;	MH892														
			237	816;														
			4;	MH892														
			237	817;														
			5;	MH892														
			249	818;														
			3;	MH892														
			264	819;														
			9;	MH892														
			290	820;														
			5;	MH892														
			323	821;														
			1;	MH892														
			337	823;														
			1	MH892														
				825;														
				MH892														
				826														
Nitrospirae																		
Nitrospira	Nitrospirales	Nitrospira	404	MH892	<i>Nitrospira</i>	-	-	-	-	-	-	-	-	-	-	-	1.	1.
			;	758;	spp.												2	5

			614	MH892												
			;	767;												
			157	MH892												
			4;	808;												
			269	MH892												
			0	822												
Proteobacteria																
Betaproteobacteria	Burkholderiales	Comamonadaceae	27	MH892 715	<i>Aquabacterium</i> sp.	-	0.	0.	1.	0.	0.	3.	0.	0.	0.	0.
			647	MH892 769	<i>Brachymonas denitrificans</i>	0.	-	-	-	-	-	-	-	-	-	-
			94;	MH892	Comamonadaceae spp.	0.	0.	0.	-	0.	0.	0.	-	0.	0.	1.
			677	726;		3	1	1		3	6	6		4	6	1
			;	MH892												
			356	772;												
			;	MH892												
			492	751;												
			;	MH892												
			776	765;												
			;	MH892												
			580	780;												
			;	MH892												
			926	766;												
			;	MH892												
			161	790;												
			9;	MH892												
			310	809;												
			1;	MH892												
			855	824;												
				MH892												
				785												
			11	MH892	Comamonas sp.	6.	-	0.	0.	-	4.	1.	-	4.	0.	0.
				713		9		1	1		4	4		4	2	2
			133	MH892	Delftia sp.	-	-	-	-	-	-	-	-	-	-	-

			6	802												
			73;	MH892	<i>Hydrogenop</i>	-	0.	0.	0.	2.	-	2.	0.	<0	-	<0
			85;	723;	<i>haga</i> spp.		2	6	4	4		8	3	.1		.1
			100	MH892												
			;	724;												
			184	MH892												
			;	727;												
			225	MH892												
			;	741;												
			275	MH892												
			;	744;												
			319	MH892												
			;	747;												
			426	MH892												
			;	749;												
			449	MH892												
			;	759;												
			454	MH892												
			;	761;												
			142	MH892												
			0	762;												
				MH892												
				805												
			110	MH892	<i>Polaromona</i>	-	-	-	0.	<0	-	<0	-	-	-	0.
			9;	796;	<i>s</i> spp.				1	.1		.1				2
			140	MH892												
			3	804												
			716	MH892	<i>Rhodoferax</i>	-	0.	-	-	-	-	-	-	-	-	0.
			;	774;	spp.		2									5
			904	MH892												
				788												
			762	MH892	<i>Rubrivivax</i>	-	-	<0	-	0.	-	-	-	-	-	-
			;	778;	spp.			.1		1						
			349	MH892												
			8	828												

	Rhodocycl ales	Rhodocycla ceae	171 ; 823 ; 189 8	MH892 736; MH892 782; MH892 812	<i>Azoarcus</i> spp.	-	-	-	-	0. 6	-	-	-	-	-	-
			387 ; 980	MH892 755; MH892 794	<i>Azonexus</i> spp.	<0 .1	-	-	-	0. 2	-	-	-	-	-	-
			193 ; 863	MH892 743; MH892 786	<i>Azospira</i> spp.	0. 1	<0 .1	-	-	-	-	-	-	-	2. 3	1. 5
			11; 30; 71; 302 ; 490	MH892 713; MH892 716; MH892 722; MH892 748; MH892 764	<i>Dechloromo nas</i> spp.	1. 8	-	0. 1	0. 1	-	0. 4	5. 6	-	0. 3	1. 0	1. 2
			146 2	MH892 806	<i>Methylovers atilis</i> sp.	-	-	-	-	<0 .1	-	-	-	-	-	-
			49; 70; 104 ; 389 ; 717 ; 847	MH892 719; MH892 721; MH892 729; MH892 756; MH892	<i>Propionivibr io</i> spp.	2. 2	-	-	0. 2	-	2. 1	0. 9	-	1. 8	0. 3	0. 1

			; 775; 901 MH892 784; MH892 787												
			623 MH892 ; 768; 113 MH892 9; 797; 130 MH892 5; 801; 136 MH892 7; 803; 154 MH892 6; 807; 148 MH892 ; 732; 269 MH892 ; 746; 373 MH892 ; 753; 754 MH892 ; 777; 115 MH892 1 799	Rhodocyclac eae spp.	0. 6	<0 .1	- 3	0. 1	0. 4	0. 1	- 6	0. 3	0. 3	0. 4	
			787 MH892 ; 781; 385 MH892 ; 754; 977 MH892 793	<i>Sterolibacter</i> <i>ium</i> spp.	-	-	-	-	-	-	-	-	1. 3	1. 2	
			179 MH892 6 810	<i>Sulfuritalea</i> sp.	-	<0 .1	-	-	-	-	-	-	-	-	
			23; MH892 91; 714;	<i>Thauera</i> spp.	3. 4	- .1	<0 -	- 2	0. 5	0. 0	6. 4	- 0.	- 4	0. 2	

			126 ; 924	MH892 725; MH892 731; MH892 789												
			191 ; 197 4	MH892 742; MH892 814	<i>Uliginosibac terium</i> spp.	-	-	<0 .1	-	-	-	-	-	-	2. 4	1. 4
			120 ; 180 ; 181 ; 335	MH892 730; MH892 739; MH892 740; MH892 750	<i>Zoogloea</i> spp.	0. 7	-	<0 .1	0. 1	-	0. 8	0. 6	-	0. 5	0. 1	1. 0
	Unclassifi ed	Unclassifie d	165 ; 394 ; 441 ; 655 ; 693 ; 728 ; 825 ; 938 ; 101	MH892 735; MH892 757; MH892 760; MH892 770; MH892 773; MH892 776; MH892 783; MH892 791; MH892	<i>Candidatus</i> <i>Accumuliba</i> <i>cter</i> spp. ^b	-	-	-	-	-	-	-	-	-	5. 0	5. 9

			5;	795;															
			340	MH892															
			4	827															

^aMost specific level of taxonomy designated after comparing ZOTUs to Greengenes and NCBI nr/nt sequences.

^b*Candidatus* Accumulibacter spp. was assigned by Greengenes to the family Rhodocyclaceae, but is a recently discovered bacterium that has not yet been classified to an order or family.

ACCEPTED MANUSCRIPT

Highlights

- Waste water samples were screened with bacterial 16S next-generation sequencing
- The V4 region of 16S could not differentiate Enterobacteriaceae
- Only three pathogens could be identified to the species level
- Erroneous taxa in the 16S Greengenes database were identified
- NCBI nr/nt database comparisons provided more accurate taxonomic assignments

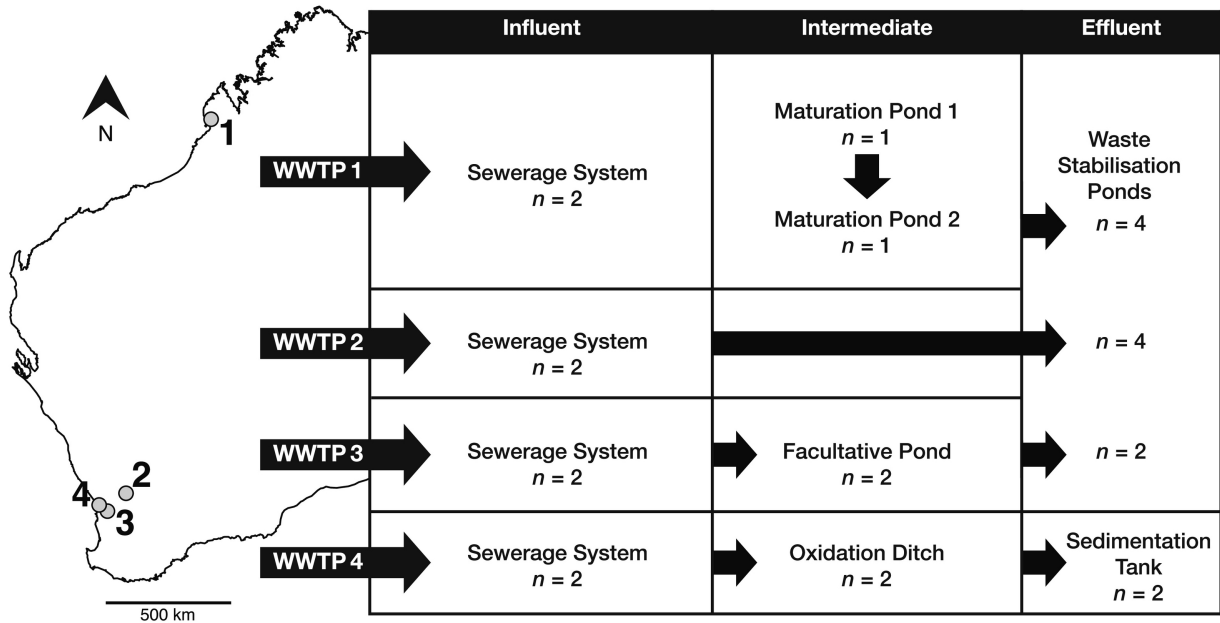


Figure 1

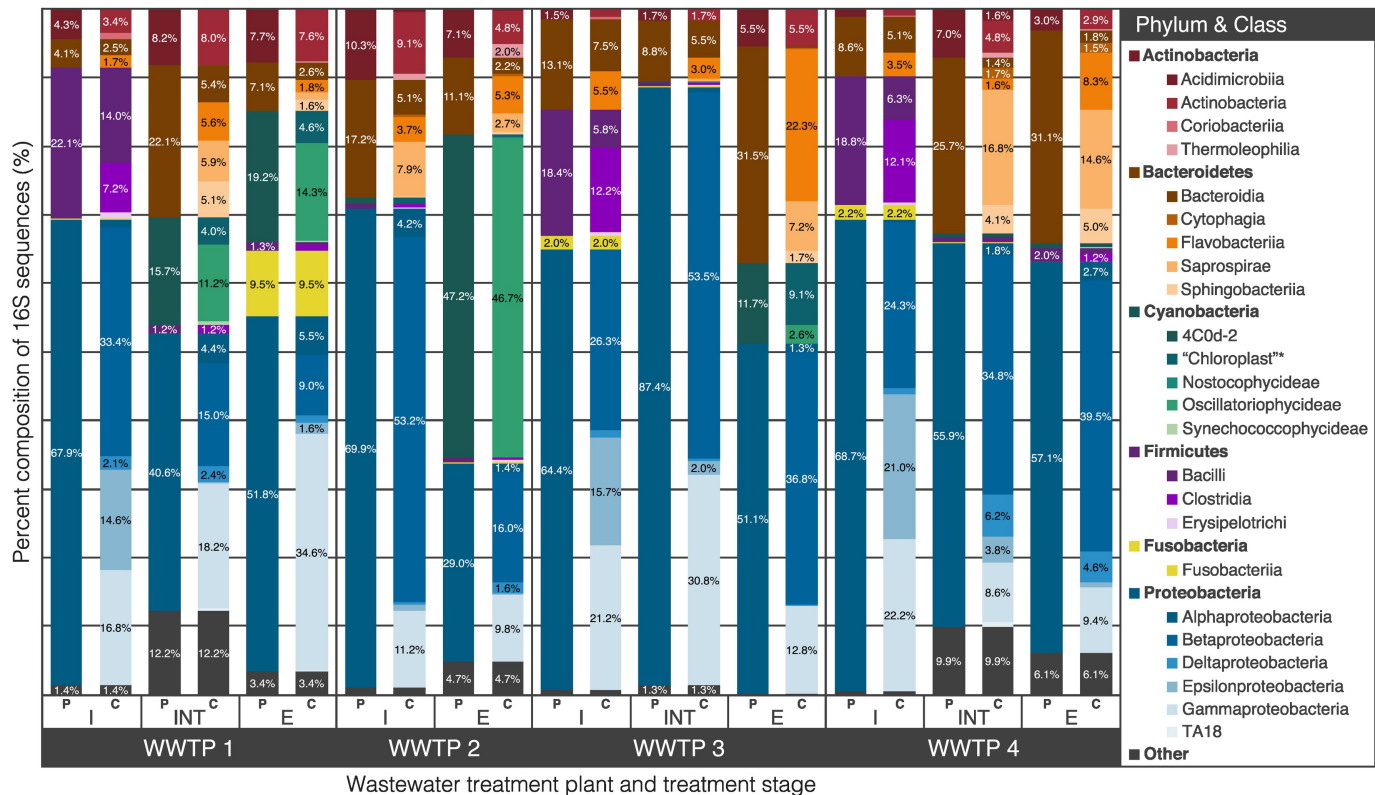
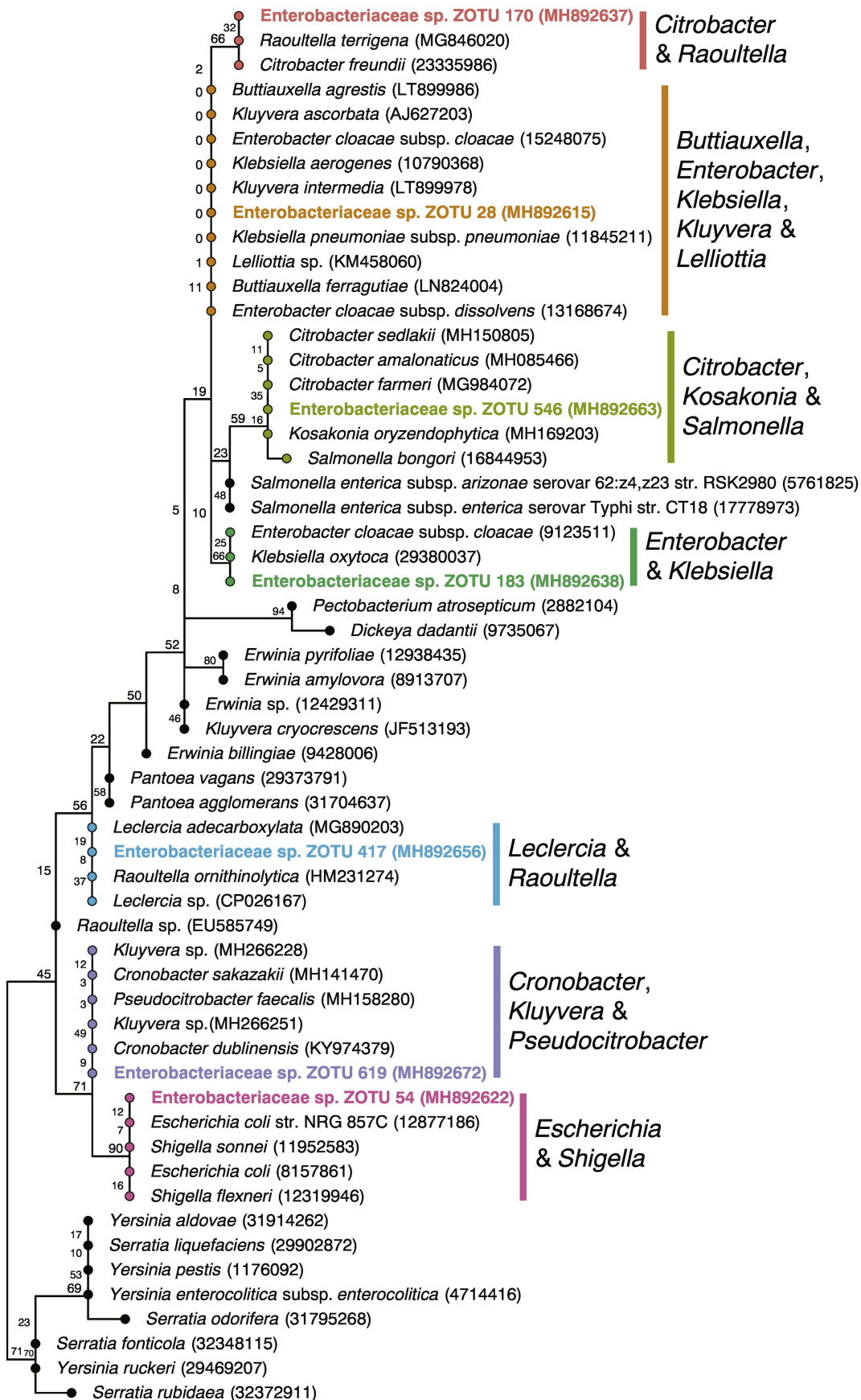


Figure 2



0.02

Figure 3