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1        **Modelling occurrence and status of mat-forming lichens in boreal forests to**  
2 **assess the past and current quality of reindeer winter pastures**

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12

13      **ABSTRACT**

14      Lichens play an essential role in northern ecosystems as important contributors to the  
15 water, nutrient and carbon cycles, as well as the main winter food resource for reindeer  
16 (*Rangifer tarandus*, also called caribou in North America), the most abundant herbivores in  
17 arctic and subarctic regions. Today, climate change and several types of land use are rapidly  
18 transforming northern ecosystems and challenging lichen growth. Since lichens are important  
19 indicators of ecosystem health and habitat suitability for reindeer, large-scale assessments are  
20 needed to estimate their past, present and future status. In our study, we aimed to develop  
21 models and equations that can be used by stakeholders to identify the occurrence of lichen-  
22 dominated boreal forests and to determine lichen conditions in those forests. Data were  
23 collected in Sweden and most input data are publicly available. We focused on mat-forming  
24 lichens belonging to the genera *Cladonia* and *Cetraria*, which are dominant species in the  
25 reindeer and caribou winter diet. Our models described lichen-dominated forests as being  
26 dominated by Scots pine (*Pinus sylvestris*), having low basal area and thin canopy cover, and  
27 being located in south- and west-facing areas with low winter temperatures and on gentle  
28 slopes. Within those forests, lichen biomass was positively related to tree canopy cover and  
29 summer precipitation, while negatively and exponentially related to intensity of use of the  
30 area by reindeer. Forest, meteorological, topographic and soil data can be used as input in our  
31 models to determine lichen conditions without having to estimate lichen biomass through  
32 demanding and expensive fieldwork.

33

34      Keywords: ground lichen; lichen growth; lichen volume; reindeer forage; reindeer  
35 husbandry; terricolous lichen

36

37           **1. INTRODUCTION**

38           Climate change and rapid landscape transformation are challenging northern ecosystems  
39 around the world. Lichens play an essential role in those ecosystems. They are important  
40 contributors to the carbon, water, and nutrient cycles (Cornelissen et al., 2007). Moreover,  
41 mat-forming lichens are an essential food resource in winter for an economically and  
42 ecologically important herbivore, the reindeer (*Rangifer tarandus*) (Heggberget et al., 2002).  
43 Despite their importance, lichens have suffered rapid declines in several parts of the world.  
44 The increase and mechanization of forestry activities, coupled in some regions with intense  
45 reindeer grazing, have strongly altered the abundance of mat-forming lichens. Examples  
46 come from Sweden (Sandström et al., 2006; Sandström et al., 2016), Finland (Kumpula et al.,  
47 2000; Uotila et al., 2005; Virtanen et al., 2003), Norway (Evans, 1996; Nygaard and  
48 Ødegaard, 1999; Virtanen et al., 2003), Alaska (Collins et al., 2011; Joly et al., 2007a; Joly et  
49 al., 2007b), Russia (Rees et al., 2003), some parts of Northern Canada (Rickbeil et al., 2017),  
50 and to a lesser extent western Canada (Coxson and Marsh, 2001). On the contrary, forest  
51 management and fire have favored the expansion of lichen woodlands in eastern Canada, to  
52 the expense of the closed-crown boreal forest (Girard et al., 2008; Payette and Delwaide,  
53 2003). Air pollution was the cause of the declines of forest and mountain heath lichens  
54 registered between 1973 and 1999 at the border between Norway and Russia (Aamlid et al.,  
55 2000; Tømmervik et al., 2003). Mat-forming lichens are expected to be additionally  
56 challenged worldwide by the foreseen expansion of vascular plants into arctic and subarctic  
57 regions, as a consequence of climate warming and increased nutrient availability (Cornelissen  
58 et al., 2001; Joly et al., 2009; Olthof and Pouliot, 2010).

59           Lichens are a symbiotic association between a fungus (the mycobiont) and an alga and/or  
60 cyanobacterium (the photobiont). *Cladonia arbuscula*, *C. mitis*, *C. rangiferina*, *C. stygia*, *C.*  
61 *stellaris*, and *Cetraria islandica* are the mat-forming lichen species preferred by reindeer in

62 winter (Andreyev, 1954) and the most abundant in northern ecosystems. All six species have  
63 circumpolar arctic and boreal distribution and low growth rates (Sandström et al., 2006;  
64 Thomson, 1984). *Cladonia* spp. are characterized by a branched, fruticose growth form and  
65 are common on nutrient-poor soils in bogs, tundra, and boreal forests, while *Cetraria* spp.  
66 have a leaf-like shape and grow in dry or wet tundra and in old spruce forests (Thomson,  
67 1984). Light exposure, humidity, and air temperature are the key factors determining lichen  
68 presence, abundance, and growth (Gaio-Oliveira et al., 2006; Jonsson Čabrajič et al., 2010).  
69 Indeed, lichens are poikilohydric organisms that can survive in a metabolically inactive state  
70 throughout long dry periods and regain their metabolic and photosynthetic activity only when  
71 enough humidity is present. The amount of light that reaches them during this wet period  
72 determines their growth rate. In Swedish forests, mat-forming lichens grow primarily in Scots  
73 pine (*Pinus sylvestris*) heaths on dry oligotrophic soils (Ahti, 1961). In general, lichen cover  
74 decreases in old pine forests on dry sites, probably due to reduced light availability and to  
75 increased nutrient availability that promotes the expansion of mosses and shrubs which  
76 outcompete lichens (reviewed in Berg et al., 2008). *C. stellaris* and *C. islandica* reach growth  
77 peaks at intermediate light exposure and their growth rate is mainly determined by total  
78 irradiance they receive when wet, chlorophyll concentration, site openness, and is negatively  
79 correlated to air temperature (Čabrajič Jonsson et al., 2010). Čabrajič Jonsson et al. (2010)  
80 found that tree basal area ( $\text{m}^2 \text{ha}^{-1}$ ) can be used as a proxy for light exposure to determine  
81 potential lichen growth. Reindeer grazing can also limit lichen growth (den Herder et al.,  
82 2003; Moen and Danell, 2003), keeping mat-forming lichens at a height of few centimeters  
83 (Roturier and Roué, 2009). Similarly, reindeer trampling may damage lichens, especially  
84 when reoccurring frequently (reviewed in Crittenden, 2000). On the contrary, in some  
85 occasions trampling and grazing by reindeer can thin the lichen mats and thus promote  
86 recovery of the remaining lichen fragments (Gaio-Oliveira et al., 2006).

87       Despite the essential role that lichens play in boreal forests, large-scale tools to monitor  
88 their status are rare. Some national inventories collect information on lichen horizontal  
89 extent, usually quantified in terms of lichen cover. One example is the Swedish National  
90 Forest Inventory (NFI, Anonymous, 2015). However, the thickness of the lichen mats, which  
91 is strictly correlated to lichen biomass (Moen et al., 2007; Olofsson et al., 2011), is rarely  
92 monitored on a large scale. Such monitoring is essential to quantify total lichen biomass and  
93 to predict how climate change and human disturbances will affect lichens, ecosystem  
94 functioning, and reindeer survival in the future. Reindeer herders, practitioners and  
95 conservationists would greatly benefit from tools to estimate the past conditions of mat-  
96 forming lichens and to detect current lichen hotspots. The purpose of this study was therefore  
97 to develop regression models that can be translated into equations which allow the  
98 assessment of lichen conditions when forest, meteorological, topographic and soil  
99 characteristics of a certain area are known. We first developed a model describing the  
100 occurrence of forests dominated by mat-forming lichens. Secondly, we developed models  
101 describing lichen biomass, height (i.e., lichen vertical growth), and cover (i.e., lichen  
102 horizontal extension) in those forests in which the ground layer is dominated by mat-forming  
103 lichens (fig. 1). We hypothesized those forests to be dominated by Scots pine and  
104 characterized by dry soils (Ahti, 1961). We also hypothesized that lichen biomass would be  
105 favored by low basal area and thin canopy cover (Berg et al., 2008; Gaio-Oliveira et al.,  
106 2006; Jonsson Čabrajč et al., 2010). Lastly, we hypothesized reindeer grazing to negatively  
107 affect lichen height (den Herder et al., 2003; Holt et al., 2008; Moen and Danell, 2003), while  
108 positively affecting lichen cover (Gaio-Oliveira et al. 2006).

109

## 110 2. METHODS

### 111 2.1 Predicting the occurrence of lichen-dominated forests

#### 112 2.1.1 Input open data

113 Since the 1920s, each year the NFI has been recording data on the Swedish forests in  
114 circular temporary plots (<http://www.slu.se/nfi>). Since 1953 the plots, with a 10 m radius,  
115 have been organized in clusters, distributed over a grid covering the whole country. Each  
116 cluster has a squared shape and three to four plots per edge, the length of which can vary  
117 between 1 and 2 km (Fridman et al., 2014). The distance between clusters varies between  
118 northern and southern Sweden, with clusters in the south being closer to each other than in  
119 the north. We selected all forest plots (n = 48267) which were sampled by the NFI between  
120 1983 and 2014, and were located within the reindeer herding husbandry area of northern  
121 Sweden, i.e. in the counties of Jämtland, Västerbotten, and Norrbotten. We assigned a unique  
122 code to each annual cluster of plots, hereafter referred to as *Cluster*. The NFI classifies each  
123 forest plot based on the vegetation group dominating the ground layer, differentiating among  
124 dry mosses, wet mosses, and mat-forming lichens. Based on the NFI classification, we  
125 divided the plots into two categories: moss-dominated and lichen-dominated. We defined as  
126 lichen-dominated those plots classified by the NFI as either “lichen dominant” (>50% lichen  
127 cover), “lichen moderate/Sphagnum type” (25-50% lichen cover), or “lichen moderate” (25-  
128 50% lichen cover) (Anonymous, 2015). We defined all other plots as moss-dominated. The  
129 NFI also records several forest characteristics at each plot, e.g. basal area, tree canopy cover,  
130 forest type, forest age, and tree height.

131 We obtained data on monthly average air temperature and monthly total precipitation from  
132 the Swedish Meteorological and Hydrological Institute (SMHI). Data were provided as  
133 monthly maps covering the whole country and divided by year (2005-2014). We averaged the  
134 monthly temperature data and summed monthly precipitation data by season (winter:

135 December-February; spring: March-May; summer: June-August; fall: September-November).  
136 The temperature map for June 2009 was missing, so we did not develop a temperature map  
137 for summer 2009. Similarly, we did not develop temperature and precipitation maps for  
138 winter 2005 because maps for December 2004 were not available. A preliminary analysis  
139 revealed that meteorological data averaged over a 5-year period (2010-2014) were highly  
140 correlated to data averaged over a 10-year period (2005-2014). Therefore, we assumed that  
141 data averaged over the 10-year period could confidently represent the spatial variability in  
142 climatic conditions among plots in our study area. Similar patterns were suggested by  
143 Jonsson Čabrajič et al. (2010). This assumption allowed us to test the importance of  
144 meteorological conditions in determining lichen dominance even for those years for which  
145 meteorological data were not available in map format (i.e. 1983-2004).

146 We derived topographic data from DEM maps with 50 m resolution downloaded from the  
147 Lantmateriet website (accessed on April 28, 2016: <http://www.lantmateriet.se/sv/Kartor-och-geografisk-information/Hojddata/>). For those areas where a 50 m resolution map was not  
148 available, we used maps with 2 m resolution. In ArcGIS 10.2.1 (ESRI, 2014), we derived  
149 slope and aspect maps from the DEMs. We obtained soil data, i.e. a map describing the  
150 percentage of sand content and a map of Available Water Capacity (AWC) in the topsoil,  
151 from the European Soil Data Centre, <http://eusoils.jrc.ec.europa.eu/content/topsoil-physical-properties-europe-based-lucas-topsoil-data>  
152 (Ballabio et al., 2016). Lastly, we extracted  
153 information from the meteorological, topographic and soil maps for each plot.  
154

155

### 156 *2.1.2 Model development*

157 We developed a quasibinomial mixed-effect regression model in which lichen dominance  
158 was the response variable, taking the value 1 for lichen-dominated plots and the value 0 for  
159 moss-dominated plots. A quasibinomial model was necessary because the corresponding



160 binomial model suffered of overdispersion. The candidate predictor variables were basal area,  
161 tree canopy cover, forest type, forest age, spring, summer and winter precipitation, summer  
162 and winter temperature, slope, aspect, sand percentage in the soil (*sand*) and AWC. We did  
163 not include spring and fall temperature as candidate predictor variables because they were  
164 highly correlated with winter temperature (Pearson's correlation coefficient  $r = 0.80$  and  $0.88$ ,  
165 respectively). Similarly, we excluded fall precipitation from the analysis because it was  
166 correlated with winter precipitation ( $r = 0.88$ ). We did not include elevation in the model due  
167 to its correlation with summer precipitation and temperature ( $r = 0.66$  and  $-0.73$ ,  
168 respectively). We added *Cluster* as a random term in order to take into account the clustered  
169 sampling design used by the NFI. We plotted a semivariogram for the within-group residuals,  
170 using the *Variogram* function in the nlme package for R (Pinheiro et al., 2018), which  
171 suggested that the model residuals were not spatially autocorrelated (Appendix A, fig. A.1,  
172 panel A). In the full model, some of the candidate predictor variables were not statistically  
173 significant ( $p\text{-value} > 0.05$ ). Therefore, we used the *Anova* function in the car package for R  
174 (Fox and Weisberg, 2011) to detect which candidate predictor variables could be removed  
175 from the full model ( $p\text{-value}$  in the likelihood ratio test  $> 0.05$ ). Models were developed in R  
176 3.3.0 (R Development Core Team, 2017).

177

## 178 **2.2 Predicting lichen biomass, height and cover in lichen-dominated forests**

### 179 *2.2.1 Study area*

180 In July and September 2015, we visited 98 sample forest plots distributed in the boreal  
181 forest zone within the Swedish reindeer herding husbandry area. Sample plots had been  
182 previously inventoried and classified by the NFI as lichen-dominated, but we restricted the  
183 selection to plots visited between 2010 and 2014 in order to take advantage of the detailed  
184 description of forest characteristics compiled by the NFI. We located all sample plots using

185 the spatial coordinates provided by the NFI, and in most cases the original location was  
186 confirmed by a wooden stick left by the NFI to mark the center of the plot. The plots had a  
187 10 m radius, the same as the original NFI plots. We visited areas that are used both by forest  
188 reindeer herding districts, which have both winter and summer pastures within the boreal  
189 forest, and mountain herding districts, which use the boreal forest only during winter. Two  
190 plots were in recent clear-cuts, five were dominated by lodgepole pine (*Pinus contorta*), one  
191 by Norway spruce (*Picea abies*), 76 by Scots pine (*Pinus sylvestris*), two were in mixed  
192 coniferous forests, and 12 were in mixed forests containing both conifers and deciduous trees,  
193 predominantly birches (*Betula* spp.).

194

#### 195 2.2.2 *Input open data*

196 For each plot visited in 2015, we obtained data on forest type, age, canopy cover and basal  
197 area from the NFI dataset. We updated data on forest age to the year of study (i.e. 2015). For  
198 recent clear-cuts, we set age, canopy cover and basal area to 0. Since boreal forests have very  
199 slow growth rates (Archibold, 1995), data on all other forest characteristics were recent  
200 enough to be included in our models as provided by the NFI.

201 Because mat-forming lichens have very slow growth rates (den Herder et al., 2003; Pegau,  
202 1968; Scotter, 1963; Thomson, 1984), we hypothesized that the meteorological conditions of  
203 several previous years would affect current lichen conditions. Since in our study area  
204 meteorological data were correlated over a 5- and a 10-year period (see subsection 2.1.1), we  
205 decided to consider the average meteorological conditions of each plot over the 5 years  
206 preceding the field measurements (i.e., 2010-2014), keeping the data divided by season as  
207 described in subsection 2.1.1. For each plot, we extracted information about topography and  
208 soil from the same maps described in subsection 2.1.1.

209

210           2.2.3 *Field measurements*

211           We measured lichen height in all sample plots as the average height of all mat-forming  
212 lichen species described in the Introduction, in 20cm-radius circles (hereafter, hits) regularly  
213 spaced one meter apart in the direction of the cardinal and half-cardinal points starting from  
214 the center of the sample plot, following Uotila et al. (2005) (Appendix A, fig. A.2 – panel A).  
215 We used a graduated rod with a plate that rests on the lichen thalli to take the measurements  
216 (Olofsson et al., 2011). During the measurement, the rod was held perpendicular to the soil  
217 without penetrating into the litter and humus layer. Lichen height was measured with a  
218 precision of 0.5 cm. This technique provided 81 measurements of lichen height for each plot.  
219 If lichens were not present, we noted lichen height = 0 cm. For each hit, we also recorded  
220 which lichen species were present.

221           We estimated the intensity of use of the area by reindeer by counting reindeer pellet  
222 groups in five subplots within each sample plot using the fecal standing crop technique  
223 (Appendix A, fig. A.2 – panel B) (McClanahan, 1986). We only counted pellet groups that  
224 included at least 50 pellets and which laid for at least half of their extent in the plots  
225 (following Skarin, 2007). In mountain herding districts the boreal forest is only used in  
226 winter, while in forest herding districts lichen-dominated forests can be used or at least  
227 travelled on also during the snow-free season. Therefore, we only counted winter pellets.  
228 During winter, reindeer pellets are dryer and appear as separate drops. Summer pellets are  
229 wetter and the individual pellets are clumped together, making them easy to distinguish from  
230 winter ones.

231

232           2.2.4 *Model development*

233           Based on the field data collected in 2015, we developed three separate regression models  
234 with three proxies of lichen conditions as response variables: lichen biomass (LB), lichen

235 height (LH), and lichen cover (LC). For each sampling plot, we estimated LB by averaging  
236 all 81 measurements of lichen height taken in a plot, i.e. including the hits where lichen  
237 height = 0 cm. LB is therefore expressed in centimeters. LB is a comprehensive measurement  
238 that takes into account both lichen height and cover, thus being a good approximation for  
239 food availability for reindeer (Moen et al., 2007). LB is also strictly correlated with lichen  
240 volume (Appendix A, fig. A.3). We estimated LH by averaging lichen height over all those  
241 hits in which lichens were present (i.e., lichen height > 0 cm). Lastly, we estimated LC as the  
242 proportion of hits where lichens were present in each plot.

243 We started by running a Gaussian mixed-effect linear regression model (GLMM) with LB  
244 as response variable, *Cluster* as random term, and all the variables described in Table 1 as  
245 candidate predictor variables, plus interaction terms between summer temperature and  
246 precipitation and between winter temperature and precipitation, with the purpose of taking  
247 into account the effect that extreme meteorological conditions may have on lichen growth  
248 (Skuncke, 1969: 29). By visual inspection we determined that the relationship between pellet  
249 group counts (*pellets*) and LB followed a decreasing exponential curve, so we included  
250 *pellets* in the form of  $\exp(-\textit{pellets})$ . The GLMM had a lower Akaike Information Criterion  
251 (AIC, Burnham and Anderson, 2002) compared to an analogous fixed-term regression model,  
252 so we retained the random term. A semivariogram for the within-group residuals, drawn  
253 using the *Variogram* function in the nlme package for R (Pinheiro et al., 2018), suggested  
254 that the model residuals were not spatially autocorrelated (Appendix A, fig. A.1, panel B).  
255 Subsequently, we used the *stepAIC* function in the MASS package for R 3.3.0 (Venables and  
256 Ripley, 2002) to run an automatic bidirectional elimination procedure in order to detect the  
257 set of predictor variables that provided the best-fit model based on AIC. In addition to the  
258 best-fit model, we also developed a reduced model by removing those variables for which the  
259 p-value in the likelihood ratio test provided by the Analysis of Variance table produced by

260 the *anova* function in R 3.3.0 (R Development Core Team, 2017) was  $> 0.05$ . We developed  
261 the reduced model because the purpose of our study was to create relatively simple equations  
262 for stakeholders' use. Thus, we believe that a model that performs slightly worse than the  
263 best-fit model but contains less predictor variables is more valuable to stakeholders.

264 Similarly we ran a GLMM with LH as a response variable, and the same random and fixed  
265 terms as for the LB model as predictors, with the exception of winter precipitation  
266 (*precip\_1014w*) which we included as a second-order polynomial because of its parabolic  
267 relationship with LH. The semivariogram for the within-group residuals suggested that the  
268 model residuals were not spatially autocorrelated (Appendix A, fig. A.1, panel C).  
269 Comparing the GLMM with an analogous fixed-term regression model as above, we  
270 determined that the random term (*Cluster*) was not needed (Standard Deviation: 0.37), so we  
271 proceeded with a fixed-effect linear regression model. Finally, we developed a best-fit and a  
272 reduced model following the same procedure as for LB.

273 Subsequently, we ran a mixed-effect quasibinomial model, i.e. a GLMM with logit  
274 function, to link LC to the candidate predictor variables described in Table 1, with the  
275 exception of *pellets*, which was included in the form of  $\ln(\textit{pellets}+1)$  because we determined  
276 by visual inspection that its relationship with the logit of LC followed a logarithmic curve.  
277 The +1 allows the calculation of the logarithm of values = 0. *Cluster* was the random term. A  
278 quasibinomial model was necessary because the corresponding binomial model suffered of  
279 overdispersion. The semivariogram for the within-group residuals suggested that the model  
280 residuals were not spatially autocorrelated (Appendix A, fig. A.1, panel D). Since AIC cannot  
281 be calculated for quasibinomial models, we used the *Anova* function in the *car* package for R  
282 3.3.0 (Fox and Weisberg, 2011) to detect which predictor variables could be removed from  
283 the full model, based on a likelihood-ratio test.

284        Lastly, we repeated the three procedures above but starting with models which did not  
285        contain reindeer pellet counts (*pellets*) as predictor variable, with the purpose of creating  
286        equations that could describe past lichen conditions, i.e. when pellet counts are not available.  
287

288 **Table 1.** List of all forest, meteorological, biotic, and topographic characteristics included as predictor variables in our models aimed to  
 289 predict the occurrence of lichen-dominated forests, as well as lichen biomass, height and cover in those forests. All continuous variables are  
 290 highlighted in italic. See the Methods section for a description of the data sources. The descriptive statistics refer to the two datasets used to  
 291 model lichen occurrence and lichen conditions (i.e. lichen biomass, height and cover) respectively, and are reported as mean (standard deviation)  
 292 [minimum; maximum].  
 293

<b>Variable</b>	<b>Description</b>	<b>Descriptive statistics (lichen occurrence)</b>	<b>Descriptive statistics (lichen conditions)</b>
<i>basal area</i>	Expressed in m <sup>2</sup> /ha. For details, see Anonymous (2015).	14.63 (13.08) [0.00; 493.22]	9.18 (8.67) [0.00; 41.89]
<i>age</i>	Average age (in years), estimated as the average age of at least two trees representative for the whole plot. For details, see Anonymous (2015).	66.85 (50.03) [0; 345]	54.92 (47.45) [0; 232]
<i>canopy cover</i>	Tree canopy cover, estimated visually and expressed as a percentage. For details, see Anonymous (2015).	55.97 (20.17) [0; 99]	38.10 (18.42) [0; 72]
<i>pellets</i>	Number of reindeer pellet groups (see section 2.2.3 for details).		1.87 (3.16) [0; 16]

Variable	Description	Descriptive statistics	Descriptive statistics
		(lichen occurrence)	(lichen conditions)
<i>precip_sp</i>	Total spring precipitation (mm) averaged for either the period 2005-2014 or 2010-2014	77.62 (10.76) [0; 186]	72.13 (6.43) [59; 89]
<i>precip_su</i>	Total summer precipitation (mm) averaged for either the period 2005-2014 or 2010-2014	189.41 (27.22) [0; 298]	195.07 (19.79) [126; 245]
<i>precip_w</i>	Total winter precipitation (mm) averaged for either the period 2005-2014 or 2010-2014	101.63 (21.15) [0; 235]	94.91 (20.38) [65; 145]
<i>temp_su</i>	Average summer temperature (°C), averaged for either the period 2005-2014 or 2010-2014	12.22 (0.97) [0; 14]	12.25 (0.74) [10; 14]
<i>temp_w</i>	Average winter temperature (°C), averaged for either the period 2005-2014 or 2010-2014	-9.91 (1.87) [-15; 0]	-10.35 (1.72) [-14; -7]
<i>slope</i>	Expressed in degrees and derived from a 50 m Digital Elevation Model (DEM), except for a few plots in Jämtland for which we used a 2 m DEM	4.04 (3.54) [0.00; 37.84]	3.55 (2.92) [0.02; 14.63]
<i>sand</i>	Percentage of sand content in the topsoil	68.88 (10.83) [0.00; 98.81]	70.87 (10.14) [50.81; 94.81]



Variable	Description	Descriptive statistics	Descriptive statistics
		(lichen occurrence)	(lichen conditions)
<i>AWC</i>	Available Water Capacity in the topsoil	0.07 (0.01) [0.00; 0.12]	0.07 (0.01) [0.05; 0.08]
<i>HD</i>	Herding district type: forest or mountain		
	Determined starting from the NFI classification referring to the proportion of each tree species. We assigned a plot to a specific forest type based on the dominant tree species (covering $\geq 70$ % of the plot). In some cases, we corrected the NFI classification based on our field observations. We defined forests $\leq 5$ years old as clear-cuts. If there was not any dominant tree species (i.e. no species constituted $> 70$ % of all trees), we defined forest type as “mixed” (including both deciduous trees and conifers) or “mixed conifer” (only including conifers).		
<i>aspect</i>	Derived from the DEM. Then, converted to a categorical		

Variable	Description	Descriptive statistics (lichen occurrence)	Descriptive statistics (lichen conditions)
	variable with 10 categories, divided as follows:		
	Flat: -1		
	North: 0-22.5		
	Northeast: 22.5-67.5		
	East: 67.5-112.5		
	Southeast: 112.5-157.5		
	South: 157.5-202.5		
	Southwest: 202.5-247.5		
	West: 247.5-292.5		
	Northwest: 292.5-337.5		
	North: 337.5-360		

295 **3. RESULTS**

296 **3.1 Predicting the occurrence of lichen-dominated forests**

297 Based on data collected in the forests of northern Sweden from 1983 to 2014, we assessed  
298 that the odds of a plot being lichen-dominated are higher in Scots pine forests compared to  
299 any other forest type, while they are lower on north facing slopes than in any other aspect  
300 category (Table 2 and Appendix A, Table A.1). Moreover, the odds of a forest being  
301 dominated by lichens are higher if the forest is older and characterized by lower basal area  
302 and thinner canopy cover (fig. 2). Finally, areas on gentle slopes with higher summer  
303 precipitation and lower winter precipitation and temperature favor lichen occurrence.

304  
305 **Table 2.** Quasibinomial mixed-effect regression model predicting the occurrence of  
306 lichen-dominated forests. The model was developed based on data from the Swedish National  
307 Forest Inventory describing boreal forests. The response variable was a dummy variable  
308 distinguishing between lichen-dominated (= 1) and moss-dominated (= 0) forests. For a list of  
309 the candidate predictor variables, see subsection 2.1. Random term standard deviation = 1.80.  
310 The categories of the “forest type” variable are marked with an asterisk. Scots pine (*Pinus*  
311 *sylvestris*) was the reference category. Lodgepole pine = *Pinus contorta*. Norway spruce =  
312 *Picea abies*. The categories of the “aspect” variable are marked with a °. North was the  
313 reference category. All continuous variables are highlighted in italic.  $\beta$  = regression  
314 coefficient mean estimate, which in a quasibinomial model is the log odd ratio; SE = standard  
315 error of the coefficient estimate.

	$\beta$	SE	p-value
intercept	-2.86	0.82	
<i>basal area</i>	-0.0441	0.0055	< 0.0001

	$\beta$	SE	p-value
<i>canopy cover</i>	-0.0345	0.0027	< 0.0001
clear-cut *	-4.64	0.69	< 0.0001
lodgepole pine *	-1.14	0.27	< 0.0001
mixed *	-2.45	0.16	< 0.0001
mixed conifer *	-1.51	0.25	< 0.0001
Norway spruce *	-2.93	0.20	< 0.0001
<i>age</i>	0.0024	0.0009	0.0107
<i>precip_su</i>	0.0119	0.0024	< 0.0001
<i>precip_w</i>	-0.0115	0.0036	0.0012
<i>temp_w</i>	-0.10	0.04	0.0110
<i>slope</i>	-0.06	0.01	< 0.0001
east °	0.44	0.17	0.0083
northeast °	0.45	0.16	0.0051
northwest °	0.73	0.19	0.0001
south °	0.94	0.16	< 0.0001
southeast °	0.70	0.18	0.0001
southwest °	1.00	0.16	< 0.0001
west °	0.92	0.17	< 0.0001

316

### 317 **3.2 Predicting lichen biomass, height and cover in lichen-dominated forests**

318 Lichen biomass (LB) was on average 3.98 ( $\pm$  2.15) cm (Appendix A, Table A.2) and was  
319 positively related to tree canopy cover (fig. 3, panel A) and summer precipitation (fig. 3,  
320 panel B), and higher in mountain reindeer herding districts compared to forest herding  
321 districts (fig. 3, panel C). The intensity of use of the area by reindeer negatively affected LB,

322 but in an exponential manner. These results are based on the reduced model detailed in Table  
323 3 and in Appendix A, Table A.3, while the best-fit model predicting LB is detailed in  
324 Appendix A, Table A.4. For a model without reindeer pellet counts, we refer the reader to  
325 Appendix B, Tables B.1 and B.2.

326 Lichen height (LH) was on average  $4.89 (\pm 2.29)$  cm (Appendix A, Table A.2) and was  
327 higher in forests with denser canopy cover and greater summer precipitation. LH decreased  
328 exponentially with an increasing use of the area by reindeer (fig. 4, panel A). Lastly, LH was  
329 higher in mountain herding districts compared to forest herding districts, and lower on south-  
330 and west-facing slopes compared to north facing slopes. These results are based on a reduced  
331 model which is detailed in Appendix A, Tables A.5 and A.6, while the best-fit model is  
332 detailed in Appendix A, Table A.7. Those models explained 65 % and 70 % of the variability  
333 in LH, respectively. For a model without reindeer pellet counts, we refer the reader to  
334 Appendix B, Tables B.3 and B.4.

335 Lichen cover (LC), estimated as a proportion, was on average  $0.82 (\pm 0.19)$ , Appendix A,  
336 Table A.2) and was positively related to use of the area by reindeer (fig. 4, panel B),  
337 negatively affected by the sand content in the soil, and highest in Scots pine forests compared  
338 to any other forest type, except lodgepole pine (Appendix A, Tables A.8 and A.9). For a  
339 model without reindeer pellet counts, we refer the reader to Appendix B, Tables B.5 and B.6.  
340

341 **Table 3.** Equations predicting lichen biomass (LB) in boreal forests dominated by mat-forming lichens. The equations were obtained from the  
 342 reduced regression model described in Appendix A, Table A.3, where the uncertainty in the coefficient estimates is also provided. Predictor  
 343 variables are described in Table 1 and in subsection 2.2. The regression model included one categorical variable (HD = herding district) and here  
 344 we report different equations for each category of that variable.

345

Categorical variable	Equation
HD = forest	$LB = -3.92 + 0.02 \textit{canopy cover} + 0.47 \exp(-\textit{pellets}) + 0.03 \textit{precip}_{su}$
HD = mountain	$LB = -2.81 + 0.02 \textit{canopy cover} + 0.47 \exp(-\textit{pellets}) + 0.03 \textit{precip}_{su}$

346

347

348 **4. DISCUSSION**

349 Mat-forming lichens thrive in Scots pine forests, with low basal area and thin canopy  
350 cover (Ahti, 1961; Table 2 and fig. 2, this study). The negative effect of a dense canopy cover  
351 on lichen growth has been previously demonstrated not only for boreal Scots pine forests in  
352 Scandinavia (Bråkenhielm and Persson, 1980; Jonsson Čabrajič et al., 2010; Uotila et al.,  
353 2005), but also for pine and spruce forests in North America (Boudreault et al., 2013; Coxson  
354 and Marsh, 2001; Foster, 1985). In dense forests, mat-forming lichens do not receive enough  
355 light for optimal growth, and the moisture and nutrient levels in the soil are more  
356 advantageous for mosses than for lichens (Sulyma and Coxson, 2001). This is the case in old  
357 forests which have not been thinned (Bråkenhielm and Persson, 1980) and in young forests,  
358 which nowadays in Scandinavia grow much faster and denser than in the past due to  
359 silviculture (Axelsson and Östlund, 2001). The agreement between previous studies and our  
360 results suggests that our model is robust and describes accurately lichen occurrence in boreal  
361 forests.

362 Once the forest ground layer is dominated by lichens, canopy cover seems to be the only  
363 forest characteristic influencing LB, which is higher in forests with denser canopy cover (fig.  
364 3, panel A). This result may seem contradictory with our model describing the occurrence of  
365 lichen-dominated forests (Table 2), which suggests that lichens occur in forests with thinner  
366 canopy cover (fig. 2, panel B). Čabrajič Jonsson et al. (2010) determined that the dry mass  
367 gain of mat-forming lichens peaks at sites with intermediate light exposure levels  
368 (corresponding to approximately 40 % canopy openness). A closer look at fig. 3, panel A  
369 suggests that LB increases up to 40 % canopy cover. At canopy covers denser than 40 %,  
370 variability in LB increases drastically. In forests where LB is high despite canopy cover being  
371 dense, lichens are probably tall and sparse, but may be locally abundant. Mat-forming lichens  
372 do not usually receive enough light for optimal growth in forests with dense canopy cover

373 (Boudreault et al., 2013; Bråkenhielm and Persson, 1980; Coxson and Marsh, 2001; Foster,  
374 1985), but the ones that manage to grow in those forests grow taller because they extend  
375 vertically in search for light inside the thick moss layer (pers. obs.). Our estimations of LB  
376 may be higher in areas with abundant summer precipitation for the same reason (fig. 3, panel  
377 B). We therefore advice the end users of the equations produced in this study to keep in mind  
378 that high LB values predicted by our equations for forests with dense canopy cover and  
379 greater summer precipitation may indicate that the lichen mat is patchy, but could be locally  
380 thick.

381 Reindeer use of the forests negatively affected LH (fig. 4, panel A), but was positively  
382 related to LC in winter grazing areas (fig. 4, panel B). Such effects were evident already at  
383 low intensity of use of the forests. In winter, lichens constitute the main component of  
384 reindeer diet (Heggberget et al., 2002). Thus, reindeer grazing is expected to shorten the  
385 lichen mat (den Herder et al., 2003; Holt et al., 2008; Moen and Danell, 2003). However,  
386 reindeer feed on lichens by opening craters in the snow in a patchy manner, so their grazing  
387 and trampling activities do not affect the lichen mat evenly and by breaking the lichen thalli,  
388 reindeer can promote lichen dispersion (Gaio-Oliveira et al. 2006). Moreover, the effects of  
389 reindeer grazing are not the same among lichen species. *Cetraria islandica* and *Cladonia*  
390 *stellaris* are the most sensitive to reindeer grazing (Andreyev, 1954; Väre et al., 1996; Väre et  
391 al., 1995), while grazing benefits *C. rangiferina* and *C. arbuscula* (Väre et al., 1996). During  
392 our 2015 fieldwork, we indeed observed that *C. rangiferina* and *C. arbuscula/mitis* dominate  
393 the boreal forests of the Swedish reindeer husbandry area, at the expenses of *C. stellaris* and  
394 *Cetraria islandica* (Appendix A, fig. A.4). However, the succession dynamics of different  
395 lichen species may also be involved in explaining the different abundance of the four species.  
396 *C. stellaris* is a late successional species within the lichen community, and if forest  
397 disturbance (due to harvesting, scarification, or fire) is frequent enough, late successional



398 lichen communities may have too little time to develop. According to Ahti (1977), *C.*  
399 *arbuscula* and *rangiferina* may be dominant 30-100 years after fire (i.e., they are primary  
400 succession species), while *C. stellaris* may not be dominant until 80-120 years after fire.

401 In this study, we have used long-term and large-scale datasets to describe the optimal  
402 habitat for the occurrence and growth of mat-forming lichens. To our knowledge, our study is  
403 the first to propose a description of the environmental characteristics that benefit the  
404 occurrence of mat-forming lichens based exclusively on publicly available data. Moreover,  
405 our LB models are based on a novel method to estimate biomass of mat-forming lichens  
406 which can be applied in future studies. Using traditional techniques, one needs to collect  
407 lichen samples in the field, take them to a laboratory, dry them and finally weigh them (see  
408 e.g. den Herder et al., 2003), which is a cumbersome procedure. With our technique, lichen  
409 biomass can instead be quantified directly in the field from measurements of lichen height  
410 and cover, or be estimate it directly from forest, meteorological, topographic and soil data  
411 using the equations proposed in this study. For a more detailed model predicting lichen  
412 growth, we refer the reader to Jonsson Čabradič et al. (2010).

413 Due to the recent strong decline in the extent of lichen-dominated forests in northern  
414 Sweden (Sandström et al., 2016), we suggest that the equations reported in this study can be  
415 useful to a variety of stakeholders, e.g. to detect areas that should receive targeted  
416 conservation or management efforts. To calculate the probability of occurrence of lichen-  
417 dominated forests, LB, LH, or LC, one has to obtain data on the variables included in the  
418 right end side of the equations and make the calculation according to the formula. The  
419 equations can be used retrospectively to estimate past conditions of mat-forming lichens in  
420 the boreal forest, as well as to map their current distribution or to foresee their future status  
421 under different climatic and environmental scenarios.

422

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429

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433

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576  
577



578 **Figure 1.** Conceptual representation of the aim and application of the study.

579

580 **Figure 2.** Variability in basal area (y-axis, panel A) and canopy cover (y-axis, panel B) in  
581 moss- and lichen-dominated forests (x-axes). For a forest to be classified as being dominated  
582 by a certain vegetation group, that group must comprise at least 25 % of the forest ground  
583 layer. Data were collected by the Swedish National Forest Inventory in 48267 forest plots,  
584 visited from 1983 to 2014. All plots were located in the Swedish reindeer husbandry area. In  
585 each boxplot, the median of the data is represented by the bold horizontal bar, the  
586 interquartile range is denoted by the horizontal edges of the box, and the dashed vertical lines  
587 extend to the range of data. Outliers were removed in order to improve the visibility of the  
588 main box. The median and interquartile range of basal area and canopy cover are slightly  
589 lower for lichen- compared to moss-dominated forests, which suggests that lichen-dominated  
590 forests have usually lower tree density and less dense tree canopy cover compared to moss-  
591 dominated forests.

592

593 **Figure 3.** Relationship between lichen biomass (LB) and tree canopy cover (panel A),  
594 summer precipitation (*precip\_su*, panel B), and type of herding district (HD, panel C) in  
595 lichen-dominated forests. A description of how lichen biomass was estimated is available in  
596 subsection 2.2. For details on the predictor variables (x-axes) see Table 1.

597

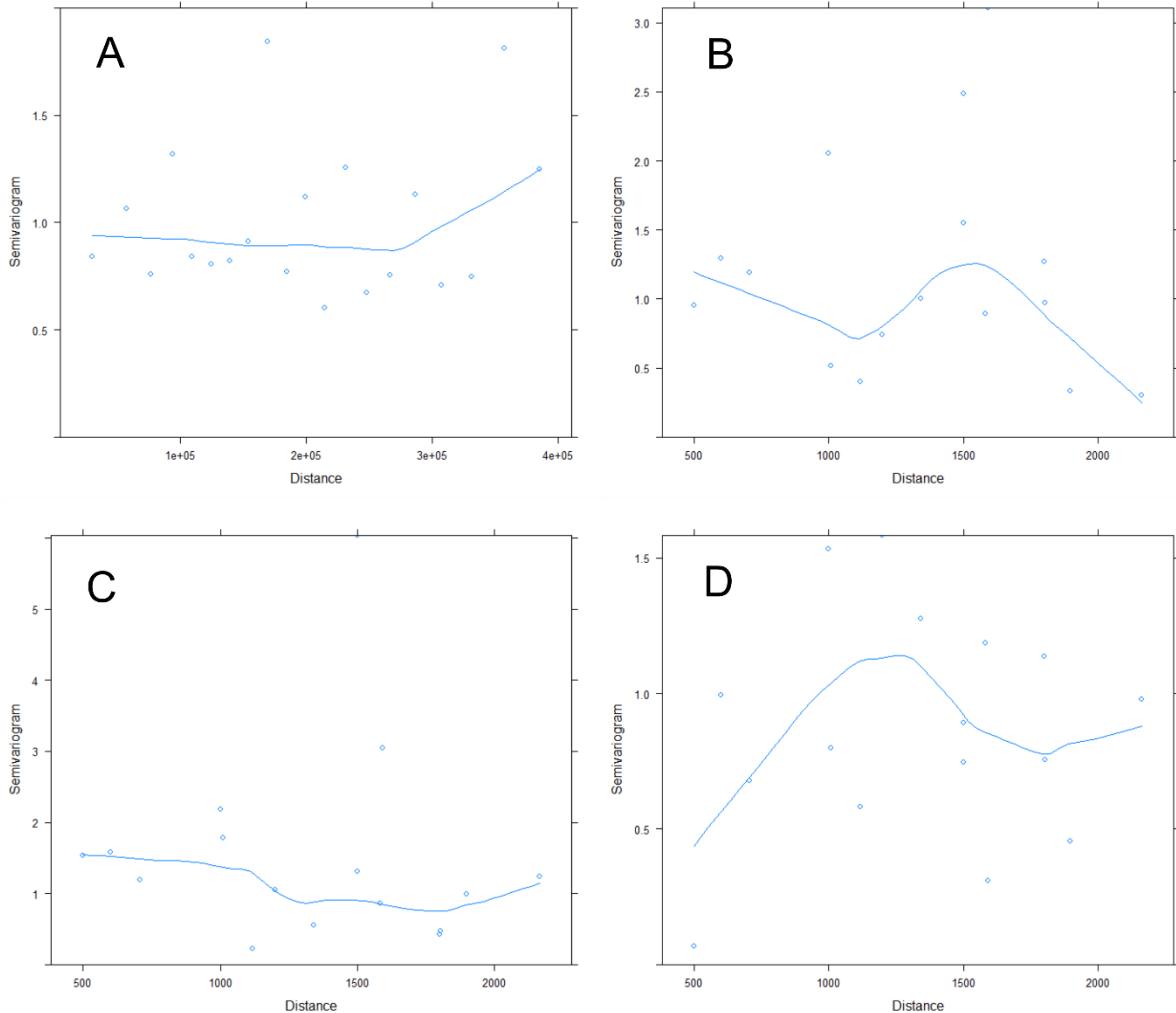
598 **Figure 4.** Relationship between lichen height (LH, y-axis in panel A) and lichen cover  
599 (LC, y-axis in panel B) and intensity of use by reindeer of boreal forests dominated by  
600 lichens (x-axes), which was estimated based on reindeer pellet counts (*pellets*). A description  
601 of how LH, LC, and *pellets* were obtained is available in subsection 2.2.

# Appendix A

## 1 Modelling occurrence and status of mat-forming lichens in boreal forests to 2 assess the past and current quality of reindeer winter pastures

3 Alessia Uboni, Alexander Blochel, Danijela Kodnik, and Jon Moen

4  
5

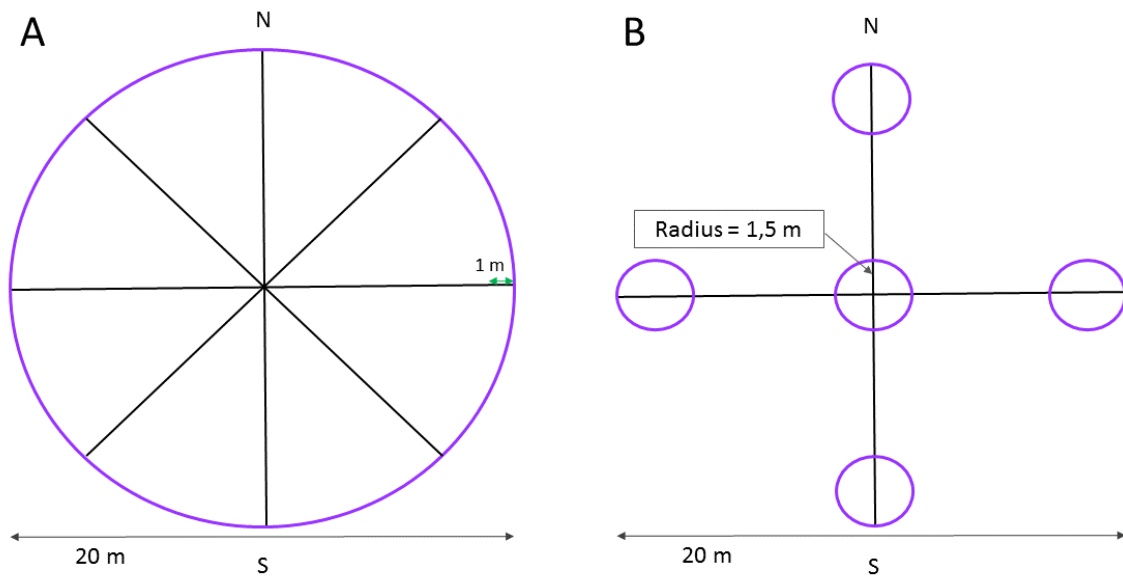


6  
7 **Figure A.1.** Semivariogram for the within-group residuals from regression models relating  
8 the occurrence of lichen-dominated forests (panel A, quasibinomial mixed-effect model),  
9 lichen biomass (panel B, linear mixed-effect model), lichen height (panel C, linear mixed-  
10 effect model), and lichen cover (panel D, quasibinomial mixed-effect model) with forest,  
11 meteorological, topographic and soil characteristics. For details on the model development  
12 see sections 2.1.2 and 2.2.4 in the main manuscript. The x-axes have been limited to distances  
13 up to 400000 m.

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# Appendix A

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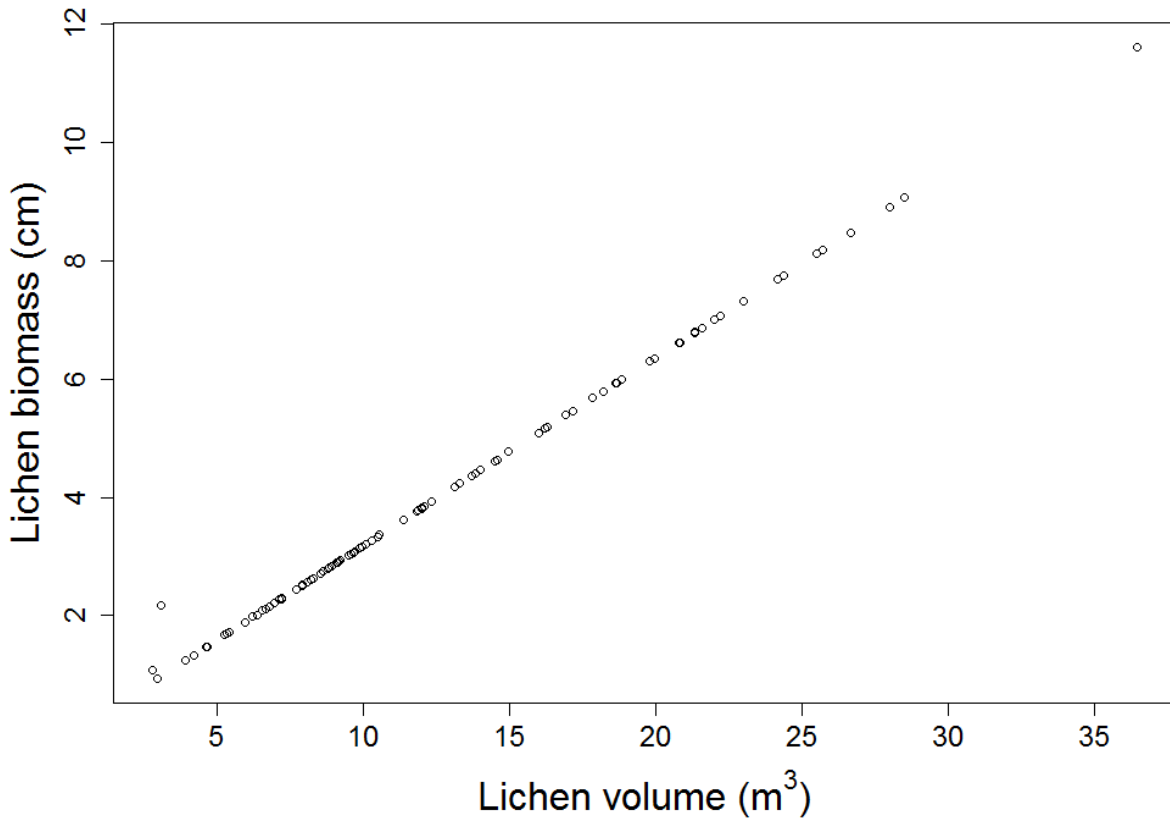
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**Figure A.2.** Data collection design. Panel A represents the protocol followed in 2015 to measure lichen height in Swedish boreal forests dominated by mat-forming lichens. Lichen height was measured in 20cm-radius circles regularly spaced one meter apart along all transects depicted in the figure. Panel B represents the protocol followed to count reindeer pellet groups. Reindeer pellet groups were counted in the 5 subplots depicted in purple in the figure.

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**Figure A.3.** Correlation between lichen biomass (y-axis) and lichen volume (x-axis) in 98 forest plots distributed within the reindeer herding husbandry area of northern Sweden. Lichen biomass was estimated as the average height of the lichen mat, including all measurements taken in a plot, i.e. also those in which lichen height = 0. Lichen volume was calculated by multiplying average lichen height by lichen cover. In the main manuscript lichen height is referred to as LH and is measured in centimeters, but for this calculation it was converted to meters. Here, lichen cover was estimated by multiplying LC (see main manuscript) by the area of the plot. Thus, here lichen cover is the proportion of the area of the plot, in m<sup>2</sup>, covered by mat-forming lichens. Each sample plot had an area of 314.16 m<sup>2</sup>. For details about the methods used to collect data on LH and LC, see subsections 2.2.3 and 2.2.4 in the Methods section of the main manuscript. The Pearson correlation coefficient between lichen biomass and lichen volume was 0.99.

## Appendix A

**Table A.1.** Equations predicting the probability (p) of a forest plot being lichen-dominated. The equations were obtained from the regression model described in Table 2. Predictor variables are described in Table 1 and in subsection 2.1 of the main manuscript. The regression model included two categorical variables, aspect (A) and forest type (FT), and here we report different equations for each category of those variables. Scots pine = *Pinus sylvestris*; Lodgepole pine = *Pinus contorta*; Norway spruce = *Picea abies*.

Categorical variable	Equation
A = north; FT = Scots pine	$p = \frac{\exp(-2.86 - 0.04 \textit{ basal area} - 0.03 \textit{ canopy cover} + 0.0024 \textit{ age} + 0.01 \textit{ precip\_su} - 0.01 \textit{ precip\_w} - 0.10 \textit{ temp\_w} - 0.06 \textit{ slope})}{\exp(-2.86 - 0.04 \textit{ basal area} - 0.03 \textit{ canopy cover} + 0.0024 \textit{ age} + 0.01 \textit{ precip\_su} - 0.01 \textit{ precip\_w} - 0.10 \textit{ temp\_w} - 0.06 \textit{ slope}) + 1}$
A = north; FT = clear-cut	$p = \frac{\exp(-7.50 - 0.04 \textit{ basal area} - 0.03 \textit{ canopy cover} + 0.0024 \textit{ age} + 0.01 \textit{ precip\_su} - 0.01 \textit{ precip\_w} - 0.10 \textit{ temp\_w} - 0.06 \textit{ slope})}{\exp(-7.50 - 0.04 \textit{ basal area} - 0.03 \textit{ canopy cover} + 0.0024 \textit{ age} + 0.01 \textit{ precip\_su} - 0.01 \textit{ precip\_w} - 0.10 \textit{ temp\_w} - 0.06 \textit{ slope}) + 1}$
A = north; FT = Lodgepole pine	$p = \frac{\exp(-4.00 - 0.04 \textit{ basal area} - 0.03 \textit{ canopy cover} + 0.0024 \textit{ age} + 0.01 \textit{ precip\_su} - 0.01 \textit{ precip\_w} - 0.10 \textit{ temp\_w} - 0.06 \textit{ slope})}{\exp(-4.00 - 0.04 \textit{ basal area} - 0.03 \textit{ canopy cover} + 0.0024 \textit{ age} + 0.01 \textit{ precip\_su} - 0.01 \textit{ precip\_w} - 0.10 \textit{ temp\_w} - 0.06 \textit{ slope}) + 1}$
A = north; FT = mixed	$p = \frac{\exp(-5.30 - 0.04 \textit{ basal area} - 0.03 \textit{ canopy cover} + 0.0024 \textit{ age} + 0.01 \textit{ precip\_su} - 0.01 \textit{ precip\_w} - 0.10 \textit{ temp\_w} - 0.06 \textit{ slope})}{\exp(-5.30 - 0.04 \textit{ basal area} - 0.03 \textit{ canopy cover} + 0.0024 \textit{ age} + 0.01 \textit{ precip\_su} - 0.01 \textit{ precip\_w} - 0.10 \textit{ temp\_w} - 0.06 \textit{ slope}) + 1}$
A = north; FT = mixed conifer	$p = \frac{\exp(-4.37 - 0.04 \textit{ basal area} - 0.03 \textit{ canopy cover} + 0.0024 \textit{ age} + 0.01 \textit{ precip\_su} - 0.01 \textit{ precip\_w} - 0.10 \textit{ temp\_w} - 0.06 \textit{ slope})}{\exp(-4.37 - 0.04 \textit{ basal area} - 0.03 \textit{ canopy cover} + 0.0024 \textit{ age} + 0.01 \textit{ precip\_su} - 0.01 \textit{ precip\_w} - 0.10 \textit{ temp\_w} - 0.06 \textit{ slope}) + 1}$
A = north; FT = Norway spruce	$p = \frac{\exp(-5.79 - 0.04 \textit{ basal area} - 0.03 \textit{ canopy cover} + 0.0024 \textit{ age} + 0.01 \textit{ precip\_su} - 0.01 \textit{ precip\_w} - 0.10 \textit{ temp\_w} - 0.06 \textit{ slope})}{\exp(-5.79 - 0.04 \textit{ basal area} - 0.03 \textit{ canopy cover} + 0.0024 \textit{ age} + 0.01 \textit{ precip\_su} - 0.01 \textit{ precip\_w} - 0.10 \textit{ temp\_w} - 0.06 \textit{ slope}) + 1}$
A = east; FT = Scots pine	$p = \frac{\exp(-2.41 - 0.04 \textit{ basal area} - 0.03 \textit{ canopy cover} + 0.0024 \textit{ age} + 0.01 \textit{ precip\_su} - 0.01 \textit{ precip\_w} - 0.10 \textit{ temp\_w} - 0.06 \textit{ slope})}{\exp(-2.41 - 0.04 \textit{ basal area} - 0.03 \textit{ canopy cover} + 0.0024 \textit{ age} + 0.01 \textit{ precip\_su} - 0.01 \textit{ precip\_w} - 0.10 \textit{ temp\_w} - 0.06 \textit{ slope}) + 1}$
A = east; FT = clear-cut	$p = \frac{\exp(-7.06 - 0.04 \textit{ basal area} - 0.03 \textit{ canopy cover} + 0.0024 \textit{ age} + 0.01 \textit{ precip\_su} - 0.01 \textit{ precip\_w} - 0.10 \textit{ temp\_w} - 0.06 \textit{ slope})}{\exp(-7.06 - 0.04 \textit{ basal area} - 0.03 \textit{ canopy cover} + 0.0024 \textit{ age} + 0.01 \textit{ precip\_su} - 0.01 \textit{ precip\_w} - 0.10 \textit{ temp\_w} - 0.06 \textit{ slope}) + 1}$
A = east; FT = Lodgepole pine	$p = \frac{\exp(-3.55 - 0.04 \textit{ basal area} - 0.03 \textit{ canopy cover} + 0.0024 \textit{ age} + 0.01 \textit{ precip\_su} - 0.01 \textit{ precip\_w} - 0.10 \textit{ temp\_w} - 0.06 \textit{ slope})}{\exp(-3.55 - 0.04 \textit{ basal area} - 0.03 \textit{ canopy cover} + 0.0024 \textit{ age} + 0.01 \textit{ precip\_su} - 0.01 \textit{ precip\_w} - 0.10 \textit{ temp\_w} - 0.06 \textit{ slope}) + 1}$
A = east; FT = mixed	$p = \frac{\exp(-4.86 - 0.04 \textit{ basal area} - 0.03 \textit{ canopy cover} + 0.0024 \textit{ age} + 0.01 \textit{ precip\_su} - 0.01 \textit{ precip\_w} - 0.10 \textit{ temp\_w} - 0.06 \textit{ slope})}{\exp(-4.86 - 0.04 \textit{ basal area} - 0.03 \textit{ canopy cover} + 0.0024 \textit{ age} + 0.01 \textit{ precip\_su} - 0.01 \textit{ precip\_w} - 0.10 \textit{ temp\_w} - 0.06 \textit{ slope}) + 1}$









# Appendix A

## Categorical variable

## Equation

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A = west; FT = Lodgepole pine	$p = \frac{\exp(-3.08 - 0.04 \textit{ basal area} - 0.03 \textit{ canopy cover} + 0.0024 \textit{ age} + 0.01 \textit{ precip\_su} - 0.01 \textit{ precip\_w} - 0.10 \textit{ temp\_w} - 0.06 \textit{ slope})}{\exp(-3.08 - 0.04 \textit{ basal area} - 0.03 \textit{ canopy cover} + 0.0024 \textit{ age} + 0.01 \textit{ precip\_su} - 0.01 \textit{ precip\_w} - 0.10 \textit{ temp\_w} - 0.06 \textit{ slope}) + 1}$
A = west; FT = mixed	$p = \frac{\exp(-4.39 - 0.04 \textit{ basal area} - 0.03 \textit{ canopy cover} + 0.0024 \textit{ age} + 0.01 \textit{ precip\_su} - 0.01 \textit{ precip\_w} - 0.10 \textit{ temp\_w} - 0.06 \textit{ slope})}{\exp(-4.39 - 0.04 \textit{ basal area} - 0.03 \textit{ canopy cover} + 0.0024 \textit{ age} + 0.01 \textit{ precip\_su} - 0.01 \textit{ precip\_w} - 0.10 \textit{ temp\_w} - 0.06 \textit{ slope}) + 1}$
A = west; FT = mixed conifer	$p = \frac{\exp(-3.45 - 0.04 \textit{ basal area} - 0.03 \textit{ canopy cover} + 0.0024 \textit{ age} + 0.01 \textit{ precip\_su} - 0.01 \textit{ precip\_w} - 0.10 \textit{ temp\_w} - 0.06 \textit{ slope})}{\exp(-3.45 - 0.04 \textit{ basal area} - 0.03 \textit{ canopy cover} + 0.0024 \textit{ age} + 0.01 \textit{ precip\_su} - 0.01 \textit{ precip\_w} - 0.10 \textit{ temp\_w} - 0.06 \textit{ slope}) + 1}$
A = west; FT = Norway spruce	$p = \frac{\exp(-4.88 - 0.04 \textit{ basal area} - 0.03 \textit{ canopy cover} + 0.0024 \textit{ age} + 0.01 \textit{ precip\_su} - 0.01 \textit{ precip\_w} - 0.10 \textit{ temp\_w} - 0.06 \textit{ slope})}{\exp(-4.88 - 0.04 \textit{ basal area} - 0.03 \textit{ canopy cover} + 0.0024 \textit{ age} + 0.01 \textit{ precip\_su} - 0.01 \textit{ precip\_w} - 0.10 \textit{ temp\_w} - 0.06 \textit{ slope}) + 1}$

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## Appendix A

47 **Table A.2.** Lichen biomass (LB), lichen height (LH), and lichen cover (LC) values measured  
 48 in 2015 in 98 plots distributed across the Swedish reindeer husbandry area. For details on  
 49 how the measurements were performed, see subsections 2.2.3 and 2.2.4 in the Methods  
 50 section of the main manuscript. SD = standard deviation  
 51

Plot ID	LB		LH		LC
	Mean	SD	Mean	SD	
1	5.68	3.59	7.30	2.14	0.78
2	5.46	3.28	6.15	2.80	0.89
3	2.61	2.09	3.85	1.27	0.68
4	1.70	2.82	5.98	1.48	0.28
5	2.02	2.14	3.57	1.59	0.57
6	0.94	1.86	4.22	1.30	0.22
7	2.75	3.48	6.74	1.59	0.41
8	3.27	1.54	3.35	1.47	0.98
9	2.21	2.69	4.48	2.11	0.49
10	2.51	1.28	2.79	1.02	0.90
11	2.79	1.06	2.79	1.06	1.00
12	7.69	3.47	8.42	2.64	0.91
13	6.79	2.03	6.79	2.03	1.00
14	11.60	2.60	11.75	2.26	0.99
15	8.19	3.42	8.61	2.93	0.95
16	6.35	2.49	6.51	2.30	0.98
17	8.48	2.21	8.59	2.01	0.99
18	4.40	3.65	5.93	2.96	0.74
19	1.67	2.22	4.22	1.29	0.40
20	1.08	1.98	3.40	1.37	0.26

## Appendix A

Plot ID	LB		LH		LC
	Mean	SD	Mean	SD	
21	1.48	1.96	2.49	2.00	0.59
22	2.28	2.52	4.40	1.69	0.52
23	3.20	1.55	3.55	1.19	0.90
24	2.80	2.07	3.91	1.26	0.72
25	2.29	1.24	2.32	1.22	0.99
26	2.18	0.93	0.98	0.16	1.00
27	2.63	1.60	3.09	1.25	0.85
28	2.44	1.23	2.71	0.98	0.90
29	3.04	1.17	3.04	1.17	1.00
30	2.80	1.52	3.06	1.31	0.91
31	1.72	1.11	2.02	0.92	0.85
32	1.24	1.53	2.51	1.23	0.49
33	3.17	2.09	4.07	1.38	0.78
34	6.30	2.75	6.63	2.40	0.95
35	2.27	2.11	2.55	2.07	0.89
36	4.23	2.32	4.69	1.94	0.90
37	5.19	2.22	5.19	2.22	1.00
38	5.09	3.69	6.34	3.00	0.80
39	2.93	1.98	3.29	1.78	0.89
40	3.84	3.25	5.65	2.28	0.68
41	2.60	3.97	7.81	2.48	0.33
42	1.89	1.90	3.12	1.44	0.60
43	2.52	2.73	4.98	1.55	0.51

## Appendix A

Plot ID	LB		LH		LC
	Mean	SD	Mean	SD	
44	5.16	2.00	5.57	1.42	0.93
45	2.89	1.45	3.16	1.20	0.91
46	2.72	2.70	5.00	1.38	0.54
47	3.07	1.74	3.36	1.53	0.91
48	4.46	1.65	4.46	1.65	1.00
49	1.47	0.84	1.47	0.84	1.00
50	1.48	0.73	1.48	0.73	1.00
51	5.93	2.53	6.58	1.66	0.90
52	5.93	1.44	5.93	1.44	1.00
53	3.09	2.53	4.72	1.43	0.65
54	4.36	1.91	4.77	1.42	0.91
55	2.09	1.52	2.09	1.52	1.00
56	2.16	1.47	2.43	1.33	0.89
57	3.07	1.41	3.36	1.09	0.91
58	4.62	2.34	4.99	2.01	0.93
59	4.77	2.04	4.95	1.84	0.96
60	5.39	2.96	6.42	1.94	0.84
61	2.92	2.61	4.38	1.94	0.67
62	5.79	4.00	7.82	2.37	0.74
63	3.93	1.97	3.93	1.97	1.00
64	3.81	1.88	3.81	1.88	1.00
65	3.38	2.65	4.80	1.76	0.70
66	9.07	3.25	9.42	2.77	0.96

## Appendix A

Plot ID	LB		LH		LC
	Mean	SD	Mean	SD	
67	7.00	4.34	8.34	3.35	0.84
68	7.75	4.33	9.52	2.45	0.81
69	8.91	3.02	9.14	2.69	0.98
70	8.11	3.62	8.76	2.89	0.93
71	6.78	4.03	8.20	2.82	0.83
72	7.31	3.96	8.35	3.04	0.88
73	7.07	4.79	9.10	3.30	0.78
74	6.62	3.75	7.25	3.29	0.91
75	6.61	5.52	10.30	3.01	0.64
76	2.12	1.50	2.49	1.32	0.85
77	6.86	3.34	7.83	2.25	0.88
78	2.02	1.26	2.34	1.05	0.86
79	2.81	1.16	2.84	1.12	0.99
80	2.57	1.25	2.57	1.25	1.00
81	1.33	0.56	1.33	0.56	1.00
82	1.98	1.12	2.00	1.11	0.99
83	3.77	1.49	3.81	1.44	0.99
84	2.21	1.12	2.21	1.12	1.00
85	2.90	2.02	3.61	1.58	0.80
86	3.33	1.19	3.38	1.14	0.99
87	4.64	2.41	4.88	2.22	0.95
88	4.18	2.65	5.21	1.83	0.80
89	2.83	2.05	3.53	1.66	0.80

## Appendix A

Plot ID	LB		LH		LC
	Mean	SD	Mean	SD	
90	3.21	2.34	4.41	1.47	0.73
91	3.02	1.48	3.02	1.48	1.00
92	3.14	2.97	5.53	1.48	0.57
93	2.12	2.71	5.06	1.59	0.42
94	3.62	2.85	4.81	2.24	0.75
95	3.93	2.61	4.42	2.34	0.89
96	5.99	3.43	6.94	2.64	0.86
97	3.82	2.89	4.62	2.52	0.83
98	3.78	2.82	5.10	1.97	0.74
Mean	3.98		4.89		0.82
SD	2.15		2.29		0.19

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## Appendix A

54 **Table A.3.** Reduced mixed-effect regression model predicting lichen biomass in boreal  
55 forests dominated by mat-forming lichens. This model was derived from the best-fit model  
56 detailed in Table A.4 by removing the non-significant variables as detailed in the Methods  
57 section. Mountain is one of the two categories of the variable HD (= herding district), where  
58 forest is the reference category. Random term standard deviation = 0.92.  $\beta$  = regression  
59 coefficient mean estimate; SE = standard error of the coefficient estimate. This model had an  
60 Akaike Information Criterion (AIC, Burnham and Anderson, 2002) = 348.87, which was  
61 slightly higher than the AIC of the best-fit model reported in Table A.4 (AIC = 345.96).  
62

<b>Predictor</b>	<b><math>\beta</math></b>	<b>SE</b>	<b>p-value</b>
intercept	-3.92	1.31	
<i>canopy cover</i>	0.02	0.01	0.0087
<i>exp(-pellets)</i>	0.47	0.36	0.1937
mountain	1.11	0.38	0.0049
<i>precip_su</i>	0.03	0.01	< 0.0001

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## Appendix A

64 **Table A.4.** Best-fit mixed-effect regression model predicting lichen biomass in boreal forests  
 65 dominated by mat-forming lichens. This model was developed from a full model including all  
 66 predictor variables described in Table 1 and in subsection 2.2 of the main manuscript, with  
 67 *Cluster* as random term (standard deviation = 0.91). Meteorological variables refer to  
 68 averages calculated over the 5 years preceding the collection of lichen biomass data (i.e.  
 69 2010-2014). Mountain is one of the two categories of the variable HD (= herding district),  
 70 where forest is the reference category. The categories of the “aspect” variable are marked  
 71 with a °, and north was the reference category. All continuous variables are highlighted in  
 72 italic. A colon mark (:) indicates an interaction term.  $\beta$  = regression coefficient mean  
 73 estimate; SE = standard error of the coefficient estimate.

74

Predictor	$\beta$	SE	p-value
intercept	-36.43	24.24	
<i>age</i>	-0.01	0.00	0.1047
<i>canopy cover</i>	0.03	0.01	0.0135
<i>exp(-pellets)</i>	0.63	0.40	0.1229
mountain	0.99	0.38	0.0120
<i>precip_su</i>	0.21	0.12	0.0960
<i>temp_su</i>	2.85	1.91	0.1464
east °	0.52	0.60	0.3904
northeast °	0.64	0.57	0.2729
northwest °	1.38	0.86	0.1177
south °	-0.56	0.72	0.4396
southeast °	-0.66	0.61	0.2905
southwest °	-0.23	0.58	0.7008
west °	-0.68	0.76	0.3764
<i>sand</i>	-0.03	0.02	0.1343
<i>precip_su:temp_su</i>	-0.02	0.01	0.1353

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## Appendix A

77 **Table A.5.** Reduced linear regression model predicting lichen height in boreal forests  
 78 dominated by mat-forming lichens. This model was derived from the best-fit model detailed  
 79 in Table A.7 by removing the non-significant variables as detailed in the Methods section.  
 80 Mountain is one of the two categories of the variable HD (= herding district), where forest is  
 81 the reference category. The categories of the “aspect” variable are marked with a °, and north  
 82 is the reference category.  $\beta$  = regression coefficient estimate; SE = standard error of the  
 83 coefficient estimate. This model had an adjusted  $R^2 = 0.65$  and an Akaike Information  
 84 Criterion (AIC, Burnham and Anderson, 2002) = 350.91, which was higher than the AIC of  
 85 the best-fit model reported in Table A.6 (AIC = 341.61).

86

Predictor	$\beta$	SE	p-value
intercept	-2.78	1.08	
<i>canopy cover</i>	0.02	0.01	0.0122
<i>exp(-pellets)</i>	2.08	0.37	<0.0001
mountain	1.52	0.31	<0.0001
<i>precip_su</i>	0.03	0.00	<0.0001
east °	-0.39	0.62	0.5316
northeast °	-0.21	0.63	0.7373
northwest °	-0.97	0.89	0.2793
south °	-1.87	0.72	0.0107
southeast °	-1.35	0.68	0.0495
southwest °	-0.98	0.60	0.1051
west °	-1.70	0.77	0.0298

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## Appendix A

89 **Table A.6.** Equations predicting lichen height (LH) in boreal forests dominated by mat-forming lichens. The equations were obtained from the  
 90 reduced regression model described in Table A.5. Predictor variables are described in Table 1 and in the subsection 2.2 of the main manuscript.  
 91 The regression model included two categorical variables and here we report different equations for each combination of their categories.  
 92

Categorical variables	Model
HD = forest; aspect = north	$LH = -2.78 + 0.02 \text{ canopy cover} + 2.08 \exp(-\text{pellets}) + 0.03 \text{ precip}_{su}$
HD = forest; aspect = east	$LH = -3.17 + 0.02 \text{ canopy cover} + 2.08 \exp(-\text{pellets}) + 0.03 \text{ precip}_{su}$
HD = forest; aspect = northeast	$LH = -2.99 + 0.02 \text{ canopy cover} + 2.08 \exp(-\text{pellets}) + 0.03 \text{ precip}_{su}$
HD = forest; aspect = northwest	$LH = -3.75 + 0.02 \text{ canopy cover} + 2.08 \exp(-\text{pellets}) + 0.03 \text{ precip}_{su}$
HD = forest; aspect = south	$LH = -4.65 + 0.02 \text{ canopy cover} + 2.08 \exp(-\text{pellets}) + 0.03 \text{ precip}_{su}$
HD = forest; aspect = southeast	$LH = -4.13 + 0.02 \text{ canopy cover} + 2.08 \exp(-\text{pellets}) + 0.03 \text{ precip}_{su}$
HD = forest; aspect = southwest	$LH = -3.76 + 0.02 \text{ canopy cover} + 2.08 \exp(-\text{pellets}) + 0.03 \text{ precip}_{su}$
HD = forest; aspect = west	$LH = -4.48 + 0.02 \text{ canopy cover} + 2.08 \exp(-\text{pellets}) + 0.03 \text{ precip}_{su}$
HD = mountain; aspect = north	$LH = -1.26 + 0.02 \text{ canopy cover} + 2.08 \exp(-\text{pellets}) + 0.03 \text{ precip}_{su}$
HD = mountain; aspect = east	$LH = -1.65 + 0.02 \text{ canopy cover} + 2.08 \exp(-\text{pellets}) + 0.03 \text{ precip}_{su}$
HD = mountain; aspect = northeast	$LH = -1.47 + 0.02 \text{ canopy cover} + 2.08 \exp(-\text{pellets}) + 0.03 \text{ precip}_{su}$
HD = mountain; aspect = northwest	$LH = -2.23 + 0.02 \text{ canopy cover} + 2.08 \exp(-\text{pellets}) + 0.03 \text{ precip}_{su}$
HD = mountain; aspect = south	$LH = -3.13 + 0.02 \text{ canopy cover} + 2.08 \exp(-\text{pellets}) + 0.03 \text{ precip}_{su}$

## Appendix A

<b>Categorical variables</b>	<b>Model</b>
HD = mountain; aspect = southeast	$LH = -2.60 + 0.02 \text{ canopy cover} + 2.08 \exp(-\text{pellets}) + 0.03 \text{ precip}_{su}$
HD = mountain; aspect = southwest	$LH = -2.24 + 0.02 \text{ canopy cover} + 2.08 \exp(-\text{pellets}) + 0.03 \text{ precip}_{su}$
HD = mountain; aspect = west	$LH = -2.96 + 0.02 \text{ canopy cover} + 2.08 \exp(-\text{pellets}) + 0.03 \text{ precip}_{su}$

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94

## Appendix A

95 **Table A.7.** Best-fit linear regression model predicting lichen height in boreal forests  
 96 dominated by mat-forming lichens. This model was developed from a full model including all  
 97 predictor variables described in Table 1 and in subsection 2.2 of the main manuscript.  
 98 Meteorological variables refer to averages calculated over the 5 years preceding the  
 99 collection of lichen height data (i.e. 2010-2014). Mountain is one of the two categories of the  
 100 variable HD (= herding district), where forest is the reference category. North was the  
 101 reference category for the “aspect” variable and the other categories are marked with a °. All  
 102 continuous variables are highlighted in italic. A colon mark (:) refers to an interaction term.  $\beta$   
 103 = regression coefficient estimate; SE = standard error of the coefficient estimate.  
 104

<b>Predictor</b>	<b><math>\beta</math></b>	<b>SE</b>	<b>p-value</b>
intercept	-43.89	21.37	
<i>canopy cover</i>	0.01	0.01	0.0939
<i>exp(-pellets)</i>	1.56	0.38	0.0001
mountain	1.58	0.40	0.0002
<i>precip_su</i>	0.23	0.10	0.0277
<i>precip_w</i>	-42.30	17.71	0.0193
<i>precip_w<sup>2</sup></i>	7.29	1.99	0.0004
<i>temp_su</i>	3.00	1.64	0.0705
<i>temp_w</i>	1.62	0.70	0.0230
east °	0.04	0.63	0.9479
northeast °	-0.11	0.60	0.8552
northwest °	-0.80	0.87	0.3606
south °	-1.22	0.71	0.0882
southeast °	-0.97	0.66	0.1444
southwest °	-0.51	0.59	0.3898
west °	-1.79	0.78	0.0241
<i>AWC</i>	61.39	28.50	0.0343
<i>temp_su:precip_su</i>	-0.02	0.01	0.0446
<i>temp_w:precip_w</i>	-0.02	0.01	0.0151

## Appendix A

105 **Table A.8.** Quasibinomial mixed-effect regression model predicting lichen cover in boreal  
 106 forests dominated by mat-forming lichens. This model was developed from a full model  
 107 including all predictor variables described in Table 1 and in subsection 2.2 of the main  
 108 manuscript. *Cluster* was the random term (standard deviation: 0.74). The categories of the  
 109 “forest type” variable are marked with an asterisk. Scots pine (*Pinus sylvestris*) was the  
 110 reference category. Lodgepole pine = *Pinus contorta*. Norway spruce = *Picea abies*. All  
 111 continuous variables are highlighted in italic.  $\beta$  = regression coefficient mean estimate, which  
 112 in a quasibinomial model is the log odd ratio; SE = standard error of the coefficient estimate.  
 113

<b>Predictor</b>	<b><math>\beta</math></b>	<b>SE</b>	<b>p-value</b>
intercept	4.82	1.02	
clear-cut *	-1.32	0.62	0.0414
lodgepole pine *	-0.15	0.44	0.7345
mixed *	-0.85	0.28	0.0041
mixed conifer *	-1.93	0.73	0.0108
Norway spruce *	-2.03	1.00	0.0468
<i>age</i>	-0.0045	0.0023	0.0520
<i>log(pellets + 1)</i>	0.69	0.17	0.0003
<i>sand</i>	-0.04	0.01	0.0023

114

115

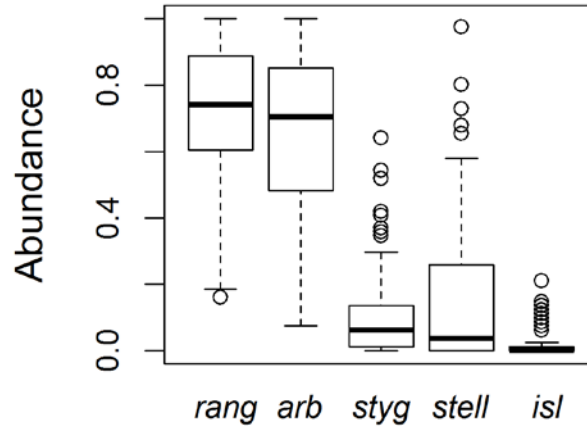
## Appendix A

116 **Table A.9.** Equations predicting lichen cover (LC) in boreal forests dominated by mat-forming lichens. The equations were obtained from the  
 117 reduced regression model described in Table A.8. Predictor variables are described in Table 1 and in subsection 2.2 of the main manuscript. The  
 118 regression model included one categorical variable, forest type (FT) and here we report different equations for each category of that variable.  
 119 Scots pine = *Pinus sylvestris*; Lodgepole pine = *Pinus contorta*; Norway spruce = *Picea abies*.  
 120

Categorical variable	Equation
FT = Scots pine	$LC = \frac{\exp(4.82 + 0.69 \log(\text{pellets} + 1) - 0.04 \text{ sand})}{\exp(4.82 + 0.69 \log(\text{pellets} + 1) - 0.04 \text{ sand}) + 1}$
FT = clear-cut	$LC = \frac{\exp(3.50 + 0.69 \log(\text{pellets} + 1) - 0.04 \text{ sand})}{\exp(3.50 + 0.69 \log(\text{pellets} + 1) - 0.04 \text{ sand}) + 1}$
FT = Lodgepole pine	$LC = \frac{\exp(3.35 + 0.69 \log(\text{pellets} + 1) - 0.04 \text{ sand})}{\exp(3.35 + 0.69 \log(\text{pellets} + 1) - 0.04 \text{ sand}) + 1}$
FT = mixed	$LC = \frac{\exp(2.50 + 0.69 \log(\text{pellets} + 1) - 0.04 \text{ sand})}{\exp(2.50 + 0.69 \log(\text{pellets} + 1) - 0.04 \text{ sand}) + 1}$
FT = mixed conifer	$LC = \frac{\exp(0.57 + 0.69 \log(\text{pellets} + 1) - 0.04 \text{ sand})}{\exp(0.57 + 0.69 \log(\text{pellets} + 1) - 0.04 \text{ sand}) + 1}$
FT = Norway spruce	$LC = \frac{\exp(-1.46 + 0.69 \log(\text{pellets} + 1) - 0.04 \text{ sand})}{\exp(-1.46 + 0.69 \log(\text{pellets} + 1) - 0.04 \text{ sand}) + 1}$

121

## Appendix A



122  
123

124 **Figure A.4.** Distribution of the abundance (y axis) of each of five mat-forming lichen species  
125 (x axis) in lichen-dominated boreal forests of northern Sweden. Data were collected in 2015  
126 in 98 forest plots located in the counties of Jämtland, Västerbotten and Norrbotten, Sweden.  
127 Forest plots are the sample unit. Species abundance is reported as the proportion of hits in a  
128 plot containing a certain species (see section 2.2 in the main manuscript for details on data  
129 collection). In each boxplot the median is represented by a bold horizontal bar, the  
130 interquartile range corresponds to the horizontal edges of the box, the dashed vertical lines  
131 extend to the range of data, and the circles outside the box indicate outliers. rang = *Cladonia*  
132 *rangiferina*; arb = *Cladonia arbuscula/mitis*; styg = *Cladonia stygia*; stell = *Cladonia*  
133 *stellaris*; isl = *Cetraria islandica*. *C. arbuscula* and *C. mitis* were pooled because they are  
134 impossible to distinguish visually. The median and interquartile range of the abundance of *C.*  
135 *rangiferina* and *C. arbuscula* are much higher than for the other species, which suggests that  
136 those two species are the most abundant in the lichen-dominated boreal forests of northern  
137 Sweden.

138

## Appendix A

### 139 **References**

140 Burnham, K.P., Anderson, D.R., 2002. Model selection and multi-model inference: a  
141 practical information-theoretic approach. Springer-Verlag, New York, USA.

142



## Appendix B

### 1 **Modelling occurrence and status of mat-forming lichens in boreal forests to** 2 **assess the past and current quality of reindeer winter pastures**

3 Alessia Uboni, Alexander Blochel, Danijela Kodnik, and Jon Moen  
4  
5  
6  
7

8 **Table B.1.** Reduced mixed-effect regression model predicting lichen biomass in boreal  
9 forests dominated by mat-forming lichens. Compared to the model detailed in Appendix A,  
10 Table A.3, this model did not include reindeer pellet count as a predictor variable. Mountain  
11 is one of the two categories of the variable HD (= herding district), where forest is the  
12 reference category. Random term standard deviation = 0.97.  $\beta$  = regression coefficient mean  
13 estimate; SE = standard error of the coefficient estimate.  
14

<b>Predictor</b>	<b><math>\beta</math></b>	<b>SE</b>	<b>p-value</b>
intercept	-3.90	1.34	
<i>canopy cover</i>	0.02	0.01	0.0054
mountain	1.24	0.37	0.0015
<i>precip_su</i>	0.03	0.01	< 0.0001

15

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## Appendix B

17 **Table B.2.** Equations predicting lichen biomass in boreal forests dominated by mat-forming lichens. The equations were obtained from the  
18 reduced regression model described in Table B.1. Predictor variables are described in Table 1 and in subsection 2.2 of the main manuscript. The  
19 regression model included one categorical variable (HD = herding district) and here we report different equations for each category of that  
20 variable.  
21

<b>Categorical variable</b>	<b>Equation</b>
HD = forest	$LB = -3.90 + 0.02 \text{ canopy cover} + 0.03 \text{ precip}_{su}$
HD = mountain	$LB = -2.66 + 0.02 \text{ canopy cover} + 0.03 \text{ precip}_{su}$

22

## Appendix B

23 **Table B.3.** Reduced linear regression model predicting lichen height in boreal forests  
 24 dominated by mat-forming lichens. Compared to the model detailed in Appendix A, Table  
 25 A.5, this model did not include reindeer pellet count as a predictor variable. Mountain is one  
 26 of the two categories of the variable HD (= herding district), where forest is the reference  
 27 category. The categories of the “aspect” variable are marked with a °, and north is the  
 28 reference category. A colon mark (:) refers to an interaction term.  $\beta$  = regression coefficient  
 29 estimate; SE = standard error of the coefficient estimate. This model had an adjusted  $R^2$  =  
 30 0.61.

31

Predictor	$\beta$	SE	p-value
intercept	-3.12	1.96	
<i>canopy cover</i>	0.03	0.01	0.0017
mountain	1.55	0.44	0.0007
<i>precip_su</i>	0.03	0.01	<0.0001
<i>precip_w</i>	-36.68	18.44	0.0500
<i>precip_w</i> <sup>2</sup>	9.07	2.17	0.0001
<i>temp_w</i>	1.58	0.71	0.0281
east °	0.27	0.67	0.6919
northeast °	0.24	0.67	0.7202
northwest °	-0.26	0.97	0.7925
south °	-0.60	0.74	0.4201
southeast °	-0.88	0.72	0.2293
southwest °	-0.23	0.63	0.7179
west °	-2.55	0.85	0.0036
<i>precip_w : temp_w</i>	-0.02	0.01	0.0365

32

33

## Appendix B

34 **Table B.4.** Equations predicting lichen height (LH) in boreal forests dominated by mat-forming lichens. The equations were obtained from the  
 35 reduced regression model described in Table B.3. Predictor variables are described in Table 1 and in the subsection 2.2 of the main manuscript.  
 36 The regression model included two categorical variables and here we report different equations for each combination of their categories. A colon  
 37 mark (:) refers to an interaction term (i.e. one must multiply the two variables).  
 38

Categorical variables	Model
HD = forest; aspect = north	$LH = -3.12 + 0.03 \textit{canopy cover} + 0.03 \textit{precip}_{su} - 36.68 \textit{precip}_w + 9.07 \textit{precip}_w^2 + 1.58 \textit{temp}_w - 0.02 \textit{precip}_w:\textit{temp}_w$
HD = forest; aspect = east	$LH = -2.85 + 0.03 \textit{canopy cover} + 0.03 \textit{precip}_{su} - 36.68 \textit{precip}_w + 9.07 \textit{precip}_w^2 + 1.58 \textit{temp}_w - 0.02 \textit{precip}_w:\textit{temp}_w$
HD = forest; aspect = northeast	$LH = -2.88 + 0.03 \textit{canopy cover} + 0.03 \textit{precip}_{su} - 36.68 \textit{precip}_w + 9.07 \textit{precip}_w^2 + 1.58 \textit{temp}_w - 0.02 \textit{precip}_w:\textit{temp}_w$
HD = forest; aspect = northwest	$LH = -3.38 + 0.03 \textit{canopy cover} + 0.03 \textit{precip}_{su} - 36.68 \textit{precip}_w + 9.07 \textit{precip}_w^2 + 1.58 \textit{temp}_w - 0.02 \textit{precip}_w:\textit{temp}_w$
HD = forest; aspect = south	$LH = -3.72 + 0.03 \textit{canopy cover} + 0.03 \textit{precip}_{su} - 36.68 \textit{precip}_w + 9.07 \textit{precip}_w^2 + 1.58 \textit{temp}_w - 0.02 \textit{precip}_w:\textit{temp}_w$
HD = forest; aspect = southeast	$LH = -4.00 + 0.03 \textit{canopy cover} + 0.03 \textit{precip}_{su} - 36.68 \textit{precip}_w + 9.07 \textit{precip}_w^2 + 1.58 \textit{temp}_w - 0.02 \textit{precip}_w:\textit{temp}_w$
HD = forest; aspect = southwest	$LH = -3.35 + 0.03 \textit{canopy cover} + 0.03 \textit{precip}_{su} - 36.68 \textit{precip}_w + 9.07 \textit{precip}_w^2 + 1.58 \textit{temp}_w - 0.02 \textit{precip}_w:\textit{temp}_w$
HD = forest; aspect = west	$LH = -5.67 + 0.03 \textit{canopy cover} + 0.03 \textit{precip}_{su} - 36.68 \textit{precip}_w + 9.07 \textit{precip}_w^2 + 1.58 \textit{temp}_w - 0.02 \textit{precip}_w:\textit{temp}_w$
HD = mountain; aspect = north	$LH = -1.57 + 0.03 \textit{canopy cover} + 0.03 \textit{precip}_{su} - 36.68 \textit{precip}_w + 9.07 \textit{precip}_w^2 + 1.58 \textit{temp}_w - 0.02 \textit{precip}_w:\textit{temp}_w$
HD = mountain; aspect = east	$LH = -1.30 + 0.03 \textit{canopy cover} + 0.03 \textit{precip}_{su} - 36.68 \textit{precip}_w + 9.07 \textit{precip}_w^2 + 1.58 \textit{temp}_w - 0.02 \textit{precip}_w:\textit{temp}_w$
HD = mountain; aspect = northeast	$LH = -1.33 + 0.03 \textit{canopy cover} + 0.03 \textit{precip}_{su} - 36.68 \textit{precip}_w + 9.07 \textit{precip}_w^2 + 1.58 \textit{temp}_w - 0.02 \textit{precip}_w:\textit{temp}_w$
HD = mountain; aspect = northwest	$LH = -1.83 + 0.03 \textit{canopy cover} + 0.03 \textit{precip}_{su} - 36.68 \textit{precip}_w + 9.07 \textit{precip}_w^2 + 1.58 \textit{temp}_w - 0.02 \textit{precip}_w:\textit{temp}_w$
HD = mountain; aspect = south	$LH = -2.17 + 0.03 \textit{canopy cover} + 0.03 \textit{precip}_{su} - 36.68 \textit{precip}_w + 9.07 \textit{precip}_w^2 + 1.58 \textit{temp}_w - 0.02 \textit{precip}_w:\textit{temp}_w$

## Appendix B

### Categorical variables

### Model

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HD = mountain; aspect = southeast	$LH = -2.45 + 0.03 \text{ canopy cover} + 0.03 \text{ precip\_su} - 36.68 \text{ precip\_w} + 9.07 \text{ precip\_w}^2 + 1.58 \text{ temp\_w} - 0.02 \text{ precip\_w:temp\_w}$
HD = mountain; aspect = southwest	$LH = -1.80 + 0.03 \text{ canopy cover} + 0.03 \text{ precip\_su} - 36.68 \text{ precip\_w} + 9.07 \text{ precip\_w}^2 + 1.58 \text{ temp\_w} - 0.02 \text{ precip\_w:temp\_w}$
HD = mountain; aspect = west	$LH = -4.12 + 0.03 \text{ canopy cover} + 0.03 \text{ precip\_su} - 36.68 \text{ precip\_w} + 9.07 \text{ precip\_w}^2 + 1.58 \text{ temp\_w} - 0.02 \text{ precip\_w:temp\_w}$

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39

40

## Appendix B

41 **Table B.5.** Quasibinomial mixed-effect regression model predicting lichen cover in boreal  
 42 forests dominated by mat-forming lichens. Compared to the model detailed in Appendix A,  
 43 Table A.8, this model did not include reindeer pellet count as a predictor variable. *Cluster*  
 44 was the random term (standard deviation: 0.78). Mountain is one of the two categories of the  
 45 variable HD (= herding district), where forest is the reference category. The categories of the  
 46 “forest type” variable are marked with an asterisk. Scots pine (*Pinus sylvestris*) was the  
 47 reference category. Lodgepole pine = *Pinus contorta*. Norway spruce = *Picea abies*. All  
 48 continuous variables are highlighted in italic.  $\beta$  = regression coefficient mean estimate, which  
 49 in a quasibinomial model is the log odd ratio; SE = standard error of the coefficient estimate.  
 50

<b>Predictor</b>	<b><math>\beta</math></b>	<b>SE</b>	<b>p-value</b>
intercept	6.53	1.02	
clear-cut *	-1.39	0.65	0.0395
lodgepole pine *	-0.50	0.45	0.2734
mixed *	-1.01	0.29	0.0011
mixed conifer *	-1.62	0.78	0.0414
Norway spruce *	-1.23	1.04	0.2456
<i>age</i>	-0.0046	0.0023	0.0489
mountain	-0.89	0.31	0.0066
<i>sand</i>	-0.05	0.01	0.0003

51  
52

## Appendix B

53 **Table B.6.** Equations predicting lichen cover (LC) in boreal forests dominated by mat-forming lichens. The equations were obtained from the  
 54 reduced regression model described in Table B.5. Predictor variables are described in Table 1 and in subsection 2.2 of the main manuscript. The  
 55 regression model included two categorical variables, herding district type (HD) and forest type (FT). Here, we report different equations for each  
 56 category of those variables. Scots pine = *Pinus sylvestris*; Lodgepole pine = *Pinus contorta*; Norway spruce = *Picea abies*.  
 57

Categorical variable	Equation
HD = forest; FT = Scots pine	$LC = \frac{\exp(6.53 - 0.0046 \textit{age} - 0.05 \textit{sand})}{\exp(6.53 - 0.0046 \textit{age} - 0.05 \textit{sand}) + 1}$
HD = forest; FT = clear-cut	$LC = \frac{\exp(5.14 - 0.0046 \textit{age} - 0.05 \textit{sand})}{\exp(5.14 - 0.0046 \textit{age} - 0.05 \textit{sand}) + 1}$
HD = forest; FT = lodgepole pine	$LC = \frac{\exp(6.04 - 0.0046 \textit{age} - 0.05 \textit{sand})}{\exp(6.04 - 0.0046 \textit{age} - 0.05 \textit{sand}) + 1}$
HD = forest; FT = mixed	$LC = \frac{\exp(5.53 - 0.0046 \textit{age} - 0.05 \textit{sand})}{\exp(5.53 - 0.0046 \textit{age} - 0.05 \textit{sand}) + 1}$
HD = forest; FT = mixed conifer	$LC = \frac{\exp(4.91 - 0.0046 \textit{age} - 0.05 \textit{sand})}{\exp(4.91 - 0.0046 \textit{age} - 0.05 \textit{sand}) + 1}$
HD = forest; FT = Norway spruce	$LC = \frac{\exp(5.31 - 0.0046 \textit{age} - 0.05 \textit{sand})}{\exp(5.31 - 0.0046 \textit{age} - 0.05 \textit{sand}) + 1}$
HD = mountain; FT = Scots pine	$LC = \frac{\exp(5.65 - 0.0046 \textit{age} - 0.05 \textit{sand})}{\exp(5.65 - 0.0046 \textit{age} - 0.05 \textit{sand}) + 1}$

## Appendix B

Categorical variable	Equation
HD = mountain; FT = clear-cut	$LC = \frac{\exp(4.26 - 0.0046 \textit{ age} - 0.05 \textit{ sand})}{\exp(4.26 - 0.0046 \textit{ age} - 0.05 \textit{ sand}) + 1}$
HD = mountain; FT = lodgepole pine	$LC = \frac{\exp(5.15 - 0.0046 \textit{ age} - 0.05 \textit{ sand})}{\exp(5.15 - 0.0046 \textit{ age} - 0.05 \textit{ sand}) + 1}$
HD = mountain; FT = mixed	$LC = \frac{\exp(4.64 - 0.0046 \textit{ age} - 0.05 \textit{ sand})}{\exp(4.64 - 0.0046 \textit{ age} - 0.05 \textit{ sand}) + 1}$
HD = mountain; FT = mixed conifer	$LC = \frac{\exp(4.03 - 0.0046 \textit{ age} - 0.05 \textit{ sand})}{\exp(4.03 - 0.0046 \textit{ age} - 0.05 \textit{ sand}) + 1}$
HD = mountain; FT = Norway spruce	$LC = \frac{\exp(4.42 - 0.0046 \textit{ age} - 0.05 \textit{ sand})}{\exp(4.42 - 0.0046 \textit{ age} - 0.05 \textit{ sand}) + 1}$