

Replacement substances for the brominated flame retardants PBDE, HBCDD, and TBBPA

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| <p>Rapporttitel Replacement substances for the brominated flame retardants PBDE, HBCDD, and TBBPA</p> | <p>Beställare Naturvårdsverket 106 48 Stockholm Finansiering Nationell MÖ</p> |
| <p>Nyckelord för ämne Flamskyddsmedel</p> | |
| <p>Sammanfattning En litteratur- och databasstudie genomfördes med syfte att identifiera nya flamskyddsmedel, dvs. ämnen som används som ersättningskemikalier för polybromerade difenyletrar (PBDEs), hexabromocyklododekan (HBCDD) samt tetrabromobisfenol-A (TBBPA). Först studerades utvalda patent från den amerikanska patentdatabasen, där ett antal nya flamskyddsmedel kunde identifieras, bl.a. pentaerytritol, melamin och bis-(t-butylfenyl)fenylfosfat. Därefter granskades den öppna litteraturen (inklusive internationellt publicerade vetenskapliga artiklar och rapporter från olika miljömyndigheter) för att finna tidigare rapporterade koncentrationer av nya flamskyddsmedel i inomhusdamm, inom- och utomhusluft, vatten, sediment, slam, jord, atmosfärisk deposition, växter, djur samt människor. Genom denna granskning identifierades 66 nya flamskyddsmedel som detekterats i minst en av de studerade matriserna. Vidare identifierades sex listor över prioriterade ämnen i den genomsökta litteraturen. Dessa listor innehöll ca 50 nya flamskyddsmedel som anses ha hög miljömässig relevans. Information om förbrukningsmängder av olika flamskyddsmedel i Sverige och Europeiska unionen (EU) hämtades från två olika databaser (<i>Registered substances</i> from the European Chemicals Agency (ECHA) and the Swedish product register from the Swedish Chemicals Agency (KemI)). I Sverige är pentaerytritol det flamskyddsmedel som används i störst mängd, följt av bl.a. kortkedjade klorerade paraffiner (SCCPs), 2-etylhexyldifenylfosfat (EHDPP), 1,2-bis(2,4,6-tribromofenoxy)-etan (BTBPE) och tetrabromobisfenol-A-bis(2,3-dibromopropyl)eter (TBBPA-BDBPE). I EU används pentaerytritol samt melamin i högst kvantiteter, följt av bl.a. kort- och mediumkedjade klorerade paraffiner, 1,2-bis(2,3,4,5,6-pentabromofenyl)etan (DBDPE) och trietylfosfat (TEP). Från den svenska databasen erhöles också exponeringsindex som ger en uppskattning av risken för exponering (för ytvatten, luft, jord, avloppsreningsverk, konsument, samt vid yrkesmässig hantering) för de olika flamskyddsmedlen. De flamskyddsmedel som generellt utgör den högsta exponeringsrisken konstaterades vara pentaerytritol, tributylfosfat (TNBP), trifenylfosfat (TPHP), SCCPs och tritolylfosfat (TMPP). Från den svenska databasen var det dessutom möjligt att få fram information om tidstrender i risken för exponering. Ökande risk identifierades för TBBPA-BDBPE, tris(tribromoneopentyl)fosfat (TTBNPP), DBDPE, resorcinolbis(difenylfosfat) (PBDPP), TMPP och cresyldifenylfosfat (CDP). Slutligen, för att kunna prioritera mellan de identifierade flamskyddsmedlen, utvecklades en multikriteriomodell baserad på (i) användningsdata, (ii) tidstrender i risken för exponering, (iii) detekterbarhet i miljön och (iv) publicerade prioriteringslistor. De tio högst rankade flamskyddsmedlen från denna modell var TBBPA-BDBPE, DBDPE, BTBPE, TTBNPP, bis(2-etyl-1-hexyl)tetrabromoftalat (BEH-TEBP), etylenbis-tetrabromoftalimid (EBTEBPI), PBDPP, para-TMPP, TPHP, and tri(1-kloro-2-propyl)fosfat (TCIPP). Dessa flamskyddsmedel föreslås bli prioriterade i framtida miljöövervakningar.</p> | |

Replacement substances for the brominated flame retardants PBDE, HBCDD, and TBBPA

Ersättningsämnen för de bromerade flamskyddsmedlen PBDE, HBCDD och TBBPA

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Rapport till Naturvårdsverket

Överenskommelse: NV-08295-16

Uppsala, 2017-06-21

Preface

As an assignment from the Swedish Environmental Protection Agency (Naturvårdsverket), a literature- and database-review focusing on emerging organic flame retardants (FRs) was conducted. The study aimed at identifying new alternative flame retardants used as replacement chemicals for the legacy FRs PBDEs and HBCDD, and also for TBBPA. To gather relevant information, (i) selected patents of the US patent database were explored, (ii) usage data from the EU and Sweden were extracted from databases, and (iii) environmental concentrations from published research articles and reports were collected. To finally prioritize between the identified FRs, a multicriteria model was developed based on the collected data. As the study focuses on emerging FRs, information about legacy FRs such as PBDEs and HBCDD is not part of this report. Furthermore, the review focuses entirely on organic FRs, and thus information on inorganic FRs is not included.

Förord

På uppdrag av Naturvårdsverket genomfördes en litteratur- och databasstudie med fokus på nya organiska flamskyddsmedel. Syftet med studien var att identifiera flamskyddsmedel som används som ersättningsämnen för traditionellt använda flamskyddsmedel såsom PBDEer och HBCDD och även TBBPA. Relevant information inhämtades genom att (i) utforska utvalda patent ur den amerikanska patentdatabasen, (ii) sammanställa användardata från två olika databaser (EU och Sverige), och (iii) sammanställa rapporterade koncentrationer från publicerade forskningsartiklar och rapporter. För att kunna prioritera mellan de identifierade flamskyddsmedlen utvecklades en multikriteriomodell baserad på insamlade fakta. Eftersom studien fokuserar på nya flamskyddsmedel så inkluderas inte information om PBDEer och HBCDD. Vidare fokuserar studien enbart på organiska flamskyddsmedel, vilket innebär att information om oorganiska flamskyddsmedel inte heller inkluderas i rapporten.

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Summary

A literature and database review was conducted with the aim of identifying new alternative flame retardants (FRs) used as replacement chemicals for the traditionally used polybrominated diphenylethers (PBDEs) and hexabromocyclododecane (HBCDD), and also for tetrabromobisphenol-A (TBBPA). Firstly, selected patents from the US patent database were studied and a number of alternative FRs could be identified, including, e.g., pentaerythritol, melamine, and bis-(*t*-butylphenyl) phenyl phosphate. Secondly, two databases, containing quantity information on usage from Sweden and the EU, were searched to obtain usage data. In Sweden, the FR that is used in the highest quantities is pentaerythritol, followed by e.g., short-chained chlorinated paraffins (SCCPs), 2-ethylhexyl diphenyl phosphate (EHDPP), 1,2-bis(2,4,6-tribromophenoxy) ethane (BTBPE), and tetrabromobisphenol-A-bis(2,3-dibromopropyl) ether (TBBPA-BDBPE). In the EU, pentaerythritol and melamine are used in the highest quantities, followed by e.g., SCCPs, MCCPs, 1,2-bis(2,3,4,5,6-pentabromophenyl)ethane (DBDPE), and triethyl phosphate (TEP). From the Swedish database, exposure indices were obtained, indicating the potential of exposure for different environmental compartments to different FRs. The highest average potential of exposure was found for pentaerythritol, tributyl phosphate (TNBP), triphenyl phosphate (TPHP), SCCPs, and tritoyl phosphate (TMPP). In addition, time trends in the potential of exposure were obtained from the database and showed increasing trends for TBBPA-BDBPE, tris(tribromoneopentyl) phosphate (TTBNPP), DBDPE, Resorcinol bis(diphenyl phosphate) (PBDPP), TMPP, and cresyl diphenyl phosphate (CDP). Thirdly, the open literature (including international peer-reviewed articles and reports from environmental authorities), was reviewed in search for previously reported environmental concentrations of emerging FRs in indoor dust, indoor and outdoor air, water, sediment, sludge, soil, atmospheric deposition, plants and animals including humans. In total, 66 different FRs were detected in at least one of the studied matrices. In addition, six prioritization lists were identified, which included about 50 different FRs that were suggested to be of high environmental relevance. Finally, to be able to prioritize between the identified FRs for future screenings, a multicriteria model was developed based on (i) usage, (ii) time trends in the potential of exposure, (iii) environmental detection, and (iv) previous publication lists. From this multicriteria model, the top ten FRs were: TBBPA-BDBPE, DBDPE, BTBPE, TTBNPP, bis(2-ethyl-1-hexyl)tetrabromophthalate (BEH-TEBP), ethylene bis-tetrabromo phthalimide (EBTEBPI), PBDPP, *para*-TMPP, TPHP, and tri(1-chloro-2-propyl) phosphate (TCIPP). These FRs are suggested to be prioritized in future screenings.

Sammanfattning

En litteratur- och databasstudie genomfördes med syfte att identifiera nya flamskyddsmedel, dvs. ämnen som används som ersättningskemikalier för polybromerade difenyletrar (PBDEs), hexabromocyclododekan (HBCDD) samt tetrabromobisfenol-A (TBBPA). Först studerades utvalda patent från den amerikanska patentdatabasen, där ett antal nya flamskyddsmedel kunde identifieras, bl.a. pentaerytritol, melamin och bis-(t-butylfenyl)fenylfosfat. Därefter granskades den öppna litteraturen (inklusive internationellt publicerade vetenskapliga artiklar och rapporter från olika miljömyndigheter) för att finna tidigare rapporterade koncentrationer av nya flamskyddsmedel i inomhusdamm, inom- och utomhusluft, vatten, sediment, slam, jord, atmosfärisk deposition, växter, djur samt människor. Genom denna granskning identifierades 66 nya flamskyddsmedel som detekterats i minst en av de studerade matriserna. Vi identifierade sex listor över prioriterade ämnen i den genomsökta litteraturen. Dessa listor innehöll ca 50 nya flamskyddsmedel som anses ha hög miljömässig relevans. Information om förbrukningsmängder av olika flamskyddsmedel i Sverige och Europeiska unionen (EU) hämtades från två olika databaser (Registered substances from the European Chemicals Agency (ECHA) and the Swedish product register from the Swedish Chemicals Agency (KemI)). I Sverige är pentaerytritol det flamskyddsmedel som används i störst mängd, följt av bl.a. kortkedjade klorerade paraffiner (SCCPs), 2-etylhexyldifenylfosfat (EHDPP), 1,2-bis(2,4,6-tribromofenoxy)-etan (BTBPE) och tetrabromobisfenol-A-bis(2,3-dibromopropyl)eter (TBBPA-BDBPE). I EU används pentaerytritol samt melamin i högst kvantiteter, följt av bl.a. kort- och mediumkedjade klorerade paraffiner, 1,2-bis(2,3,4,5,6-pentabromofenyl)etan (DBDPE) och trietylfosfat (TEP). Från den svenska databasen erhöles också exponeringsindex som ger en indikation av risken för exponering (för ytvatten, luft, jord, avloppsreningsverk, konsument, samt vid yrkesmässig hantering) för de olika flamskyddsmedlen. De flamskyddsmedel som generellt utgör den högsta exponeringsrisken konstaterades vara pentaerytritol, tributylfosfat (TNBP), trifenylfosfat (TPHP), SCCPs och tritolylfosfat (TMPP). Från den svenska databasen var det dessutom möjligt att få fram information om tidstrender i risken för exponering. Ökande risk identifierades för TBBPA-BDBPE, tris(tribromoneopentyl)fosfat (TTBNPP), DBDPE, resorcinolbis-(difenylfosfat) (PBDPP), TMPP och cresyldifenylfosfat (CDP). Slutligen, för att kunna prioritera mellan de identifierade flamskyddsmedlen, utvecklades en multikriteriomodell baserad på (i) användningsdata, (ii) tidstrender i risken för exponering, (iii) detekterbarhet i miljön och (iv) publicerade prioriteringslistor. De tio högst rankade flamskyddsmedlen från denna modell var TBBPA-BDBPE, DBDPE, BTBPE, TTBNPP, bis(2-etyl-1-hexyl)tetrabromoftalat (BEH-TEBP), etylenbis-tetrabromoftalimid (EBTEBPI), PBDPP, para-TMPP, TPHP, and tri(1-kloro-2-propyl)fosfat (TCIPP). Dessa flamskyddsmedel föreslås bli prioriterade i framtida miljöövervakningar.

1. Introduction

Flame retardants (FRs) are substances used in different materials to provide fire protection. The FRs are designed to interrupt chemical reactions of combustion through different mechanisms (e.g., by halogens reacting with H and OH radicals formed in the flame), and thereby slowing down the fire development or ultimately quench the fire. FRs are widely used in many different materials, including e.g., textiles and plastics, which are part of everyday-life products such as furniture, electronics and building insulation [1, 2].

During the 1970's, the production and usage of plastics and synthetic fibers increased and partly replaced more traditional materials like wood and metals [3]. As a result of plastics being more flammable than traditional materials, the incorporation of FRs into these materials was desired, which led to an increased use of FRs. Following this, many nations introduced legislation towards high fire safety standards by requiring producers of e.g., furniture and electronics to add FRs into their products [4]. FRs are emitted during production, usage and disposal of the products, and as a result, many FRs are nowadays ubiquitously spread in the environment. The traditionally heavily used polybrominated diphenyl ethers (PBDEs) have e.g., been detected in numerous abiotic (e.g. soil, freshwater and sediment) and biotic (e.g. seabirds and mammals) matrices in the Arctic [4]. Also several alternative FRs (e.g., BTBPE and DBE-DBCH) have been detected in remote sites such as the Arctic [5]. Detection of FRs at far distances from emission sources demonstrate that these substances are persistent and undergo long range transport (LRT) without being transformed.

As a consequence of persistency, bioaccumulation potential, and toxicity (PBT), tetra- through hepta-PBDEs have been included in the Stockholm Convention. The use of two out of three technical PBDE products (pentaBDE and octaBDE) is forbidden in new materials in the European Union (EU) since 2009, while the third technical PBDE product (DecaBDE) has been suggested to be listed in the Stockholm Convention [6]. DecaBDE is, however, already banned from use in electrical and electronic appliances within the EU [7]. Two more FRs are listed in the Stockholm convention, namely hexabromocyclododecane (HBCDD) and hexabromobiphenyl [6]. Hexabromobiphenyl was one of the main components in the polybrominated biphenyl (PBB) mixture that was accidentally mixed into cattle feed in Michigan in 1973 [8], causing the widespread contamination of animal feed, animals and human food products.

The restriction of the previously heavily used PBDEs has increased the need of alternative FRs to be developed and used. For example, the use of TDCIPP in American couches has increased since the ban of PentaBDE [9]. The recent development and use of alternative FRs has created a need of an up-to-date overview of the current situation. The aim of this literature study was to identify (i) what FRs are used as replacement chemicals for PBDEs, HBCDD, and also TBBPA, (ii) what FRs that have been detected in indoor air and dust, and (iii) what FRs that have been detected in the environment, and finally (iv) what FRs that should be included in future environmental screening studies.

2. Exploration of available databases

A number of available databases were utilized in the search for relevant information on alternative FRs of interest. The used databases were: i) the US patent database [10], ii) *Registered substances* from the European Chemicals Agency (ECHA) [11], and iii) the Swedish product register from the Swedish Chemicals Agency (KemI) [12]. Before a product is being introduced on a market (such as the EU or the US), it is often registered in a patent database. Thus, patent databases can be useful information sources when trying to predict future use of chemicals such as FRs. A search of US patents (US patent database, www.uspto.gov) was done to identify alternative FRs using classification numbers given within the Cooperative Patent Classification (CPC)-system for polymers and beds, two types of products that often are treated to provide fire safety. Note that the results from the patent database search presented here are not comprehensive, but should rather be considered as an attempt of exploring the possibility of identifying trends in FRs usage through patent searches.

The database *Registered substances* from ECHA provides information on amounts of chemicals that is manufactured in and/or imported into the EU on an annual basis [11]. This database was used to obtain annual production/import data for the emerging FRs identified in this literature review. Furthermore, data on the use of emerging FRs in Sweden was obtained from the Swedish Chemicals Agency (Kemikalieinspektionen, KemI). EC/EINECS-numbers were used for the search if available and otherwise CAS-numbers (Table 1). Note that only non-confidential quantities are shown in these databases. Thus, the reported quantities may not necessarily reflect the real production/import, and additionally, intermediate substances used to produce other chemicals are not included in the databases.

All FRs ($n = 125$) identified in this literature review are listed with abbreviations, name, CAS- and EC-numbers (if available) in Table 1. Throughout this report, the FR abbreviations suggested by Bergman et al. (2012) [13] are used if available. Information on selected FRs are given in the Appendix, Table S1, and modelled physicochemical properties of selected FRs from a previous peer-reviewed publication [14] are given in the Appendix, Tables S16-S17.

Table 1 Identified FRs with abbreviation, name, CAS-, and EC-number (alphabetic order).

| Abbreviation | Name | CAS# | EC# |
|-----------------------|---|------------|-----------|
| 2,4-DBP | 2,4-Dibromophenol | 615-58-7 | 210-436-5 |
| 2,6-DBP | 2,6-Dibromophenol | 608-33-3 | 210-161-0 |
| 2-BP | 2-Bromophenol | 95-56-7 | 251-200-1 |
| 3-BP | 3-Bromophenol | 591-20-8 | 209-706-5 |
| 4'-PeBPOBDE208 | Pentabromophenoxy-nonabromodiphenyl ether | 58965-66-5 | 261-526-6 |
| 4-BP | 4-Bromophenol | 106-41-2 | 203-394-4 |
| TBP-AE/ATE | Allyl 2,4,6-tribromophenyl ether | 221-913-2 | 221-913-2 |
| BADP | Bisphenol A bis(diphenyl phosphate) | 5945-33-5 | 425-220-8 |

| | | | |
|----------------------------|---|-------------------------|-----------|
| BATE | 2-Bromoallyl 2,4,6-tribromophenyl ether | - | - |
| bbDBP | Bis(2,3-Dibromopropyl) phosphite | 5412-25-9 | 226-493-4 |
| BCMP-BCEP/V6 | Tetrakis(2-Chloroethyl)dichloroisopentyl diphosphate | 38051-10-4 | 253-760-2 |
| BdPhP | Butyldiphenyl phosphate | 2752-95-6 | 220-398-1 |
| BEH-TEBP | Bis(2-ethyl-1-hexyl)tetrabromophthalate | 26040-51-7 | 247-426-5 |
| BTBPE | 1,2-Bis(2,4,6-tribromophenoxy) ethane | 37853-59-1 | 253-692-3 |
| CDP | Cresyl diphenyl phosphate | 26444-49-5 | 247-693-8 |
| Chlordene Plus | Chlordene Plus | - | - |
| CLP1 | Tris(2-chloroethyl) phosphite | 140-08-9 | 205-397-6 |
| DBDPE | 1,2-Bis(2,3,4,5,6-pentabromophenyl)ethane | 84852-53-9 | 284-366-9 |
| DBE-DBCH/TBECH | 1,2-Dibromo-4-(1,2-dibromoethyl)cyclohexane | 3322-93-8 | 222-036-8 |
| DBHCTD | Hexachlorocyclopentadienyl-dibromocyclooctane | 51936-55-1 | 257-526-0 |
| DBNPG | Dibromoneopentyl alcohol | 3296-90-0 | 221-967-7 |
| DBPhP | Dibutyl phenyl phosphate | 2528-36-1 | 219-772-7 |
| DBS | Dibromostyrene | 31780-26-4 | - |
| DDC-Ant/Dec-603 | Dechlorane 603 | 13560-92-4 | - |
| DDC-CO/DP | Dechlorane Plus | 13560-89-9 | 236-948-9 |
| DDC-DBF/Dec-602 | Dechlorane 602 | 31107-44-5 | 250-472-9 |
| Dec-604A/HCTBPH | Dechlorane 604 component A | 34571-16-9 | - |
| Dec-604B | Dechlorane 604 component B | 71245-27-7 ^a | - |
| Dibutyl chlorendate | Dibutyl 1,4,5,6,7,7-hexachlorobicyclo[2.2.1]-hept-5-ene-2,3-dicarboxylate | 1770-80-5 | 217-192-9 |
| DMP | Dimethyl phosphate | 813-78-5 | 212-389-6 |
| DOPP | Diocetyl phenyl phosphate | 6161-81-5 | 228-190-2 |
| DPhBP | Diphenyl butyl phosphate | 2752-95-6 | 220-398-1 |
| EBTEBPI | Ethylene bis-tetrabromo phtalimide | 32588-76-4 | 251-118-6 |
| EHDPP | 2-Ethylhexyl diphenyl phosphate | 1241-94-7 | 214-987-2 |
| EH-TBB | 2-Ethylhexyl 2,3,4,5-tetrabromobenzoate | 183658-27-7 | - |
| HBB | Hexabromobenzene | 87-82-1 | 201-773-9 |
| HBCYD | Hexabromocyclodecane | 25495-98-1 | - |
| HEEHP-TEBP | 2-(2-hydroxyethoxy)ethyl 2-hydroxypropyl 3,4,5,6-tetrabromophthalate | 20566-35-2 | 243-885-0 |
| IDP | Isodecyl diphenyl phosphate | 29761-21-5 | 249-828-6 |
| MCCP | Chlorinated paraffins (medium-chained) | 85535-84-8 | - |
| mDEP/dDEP | Diethyl phosphate (mono/di) | 598-02-7 | 209-912-5 |
| OBTMPI | 4,5,6,7-Tetrabromo-1,1,3-trimethyl-3-(2,3,4,5-tetrabromophenyl)-indane | 1084889-51-9 | - |
| PBB | Pentabromobenzene | 608-90-2 | - |
| PBB-Acr | Pentabromobenzyl acrylate | 59447-55-1 | 261-767-7 |
| PBBBr | Pentabromobenzyl bromide | 38521-51-6 | 253-985-6 |
| PBBC | Pentabromobenzyl chloride | 58495-09-3 | - |
| PBCH | Pentabromochlorocyclohexane | 87-84-3 | 201-776-5 |
| PBDPP/RDP | Resorcinol bis(diphenyl phosphate) | 57583-54-7 | 260-830-6 |
| PBDMPP | Tetrakis(2,6-dimethylphenyl) 1,3-phenylene bis(phosphate) | 139189-30-3 | 432-770-2 |
| PBEB | Pentabromoethylbenzene | 85-22-3 | 201-593-0 |
| PBP | Pentabromophenol | 608-71-9 | 210-167-3 |

| | | | |
|--------------------|---|-------------|-----------|
| PBPAE | Pentabromophenyl allyl ether | 3555-11-01 | - |
| PBT | Pentabromotoluene | 87-83-2 | 201-774-4 |
| SCCP | Chlorinated paraffins (short-chained) | 85535-85-9 | - |
| T2CPP | Tris(2-Chloropropyl)phosphate | 6145-73-9 | 228-150-4 |
| T3CPP | Tris(3-Chloropropyl)phosphate | 26248-87-3 | - |
| TBBBS | Tetrabromobisphenol S | 39635-79-5 | 254-551-9 |
| TBBPA | Tetrabromobisphenol A | 79-94-7 | 201-236-9 |
| TBBPA-BAE | Tetrabromobisphenol A bis(allyl ether) | 25327-89-3 | 246-850-8 |
| TBBPA-BDBPE | Tetrabromobisphenol A bis(2,3-dibromopropyl ether) | 21850-44-2 | 244-617-5 |
| TBBPA-BME | Tetrabromobisphenol A bismethyl ether | 108608-62-4 | 253-693-9 |
| TBBPA-DHEE | Tetrabromobisphenol A dihydroxyethyl ether | 4162-45-2 | 224-005-4 |
| TBBPS-DBPE | Tetrabromo-bisphenol-S-bis(2,3-dibromopropyl) ether | 42757-55-1 | 255-929-6 |
| TBCO | (1R,2R,5S,6S)-1,2,5,6-Tetrabromocyclooctane | 3194-57-8 | - |
| TBOEP | Tri(2-butoxyethyl) phosphate | 78-51-3 | 201-122-9 |
| TBP | 2,4,6-Tribromophenol | 118-79-6 | 204-278-6 |
| TBP-DBPE | 2,3-Dibromopropyl 2,4,6-tribromophenyl ether | 35109-60-5 | 252-372-0 |
| TBPP | Tris(4-tert-butylphenyl) phosphate | 78-33-1 | 201-106-1 |
| TBX | 2,3,5,6-tetrabromo-p-xylene | 23488-38-2 | 245-688-5 |
| TCBPA | Tetrachlorobisphenol A | 27360-90-3 | 201-237-4 |
| TCEP | Tris(2-chloroethyl) phosphate | 115-96-8 | 204-118-5 |
| TCIPP | Tri(1-chloro-2-propyl) phosphate | 13674-84-5 | 237-158-7 |
| TDBPP | Tris(2,3-dibromopropyl) phosphate | 126-72-7 | 204-799-9 |
| TDBP-TAZTO | 1,3,5-tris(2,3-dibromopropyl)-1,3,5-triazine-2,4,6(1H,3H,5H)-trione | 52434-90-9 | 257-913-4 |
| TDCIPP | Tris(1,3-dichloroisopropyl) phosphate | 13674-87-8 | 237-159-2 |
| TDCPP | Trisdichloropropyl phosphate | 26604-51-3 | 247-843-2 |
| TEBP-Anh | 3,4,5,6-Tetrabromophthalic anhydride | 632-79-1 | 211-185-4 |
| TEEdP | Tetraethyl(ethylene)diphosphonate | 995-32-4 | 213-625-0 |
| TEHP | Tris(2-ethylhexyl) phosphate | 78-42-2 | 201-116-6 |
| TEP | Triethyl phosphate | 78-40-0 | 201-114-5 |
| THP | Trihexyl phosphate | 2528-39-4 | 219-774-8 |
| TIBP | Triisobutyl phosphate | 126-71-6 | 204-798-3 |
| TiPP | Triisopropyl phosphate | 513-02-0 | 208-150-0 |
| TiPPP | Tris(2-isopropyl) phosphate | 64532-95-2 | 248-147-1 |
| TMP | Trimethyl phosphate | 512-56-1 | 208-144-8 |
| TMPP (m-) | Tritolyl phosphate | 563-04-2 | 209-241-8 |
| TMPP (o-) | Tritolyl phosphate | 78-30-8 | 201-103-5 |
| TMPP (p-) | Tritolyl phosphate | 1330-78-5 | 215-548-8 |
| TNBP | Tributyl phosphate | 126-73-8 | 204-800-2 |
| TPeP | Tripentyl phosphate | 2528-38-3 | 219-773-2 |
| TPHP | Triphenyl phosphate | 115-86-6 | 204-112-2 |
| TPP | Tripropyl phosphate | 513-08-6 | 208-151-6 |
| TTBNPP | Tris(tribromoneopentyl) phosphate | 19186-97-1 | 606-254-4 |
| TTBP-TAZ | 2,4,6-tris(2,4,6-tribromophenoxy)-1,3,5-triazine | 25713-60-4 | - |
| TXP | Trixylenyl phosphate | 25155-23-1 | 246-677-8 |

| | | | |
|---|---|------------------------|-----------|
| - | 1,3-hexylene dimelamine | - | - |
| - | 1,3-phenylene-bis(dixylenyl phosphate) | - | - |
| - | Acetoguanamine | 542-02-9 | 208-796-3 |
| - | Ammeline/Cyanurodiamide | 645-92-1 | 211-455-1 |
| - | Benzoguanamine | 91-76-9 | 202-095-6 |
| - | Bis-(isopropylphenyl) phenyl phosphate | 101299-37-0 | 248-849-8 |
| - | Bis-(t-butylphenyl) phenyl phosphate | 65652-41-7 | 265-859-8 |
| - | Brominated paraffins | - | - |
| - | Butylene diguanamine | - | - |
| - | Chlordene Plus | - | - |
| - | Dibutyl chlorendate | 1770-80-5 | 217-192-9 |
| - | Ethylene dimelamine | - | - |
| - | Hexamethylene dimelamine | - | - |
| - | Isopropylphenyl diphenyl phosphate | 28108-99-8 | 248-848-2 |
| - | Melamine/Cyanurotriamide | 108-78-1 | 203-615-4 |
| - | Melamine (poly)phosphate | 163183-93-5 | 243-601-5 |
| - | Melamine cyanurate | 37640-57-6 | 253-575-7 |
| - | Melamine pyrophosphate | 15541-60-3 | 239-590-1 |
| - | Methylene diguanamine | - | - |
| - | Norbornene diguanamine | - | - |
| - | Octyl diphenyl phosphate | 115-88-8 | 204-113-8 |
| - | Pentaerythritol/Tetra(hydroxymethyl)methane | 115-77-5 | 204-104-9 |
| - | Phthalodiguanamine | 5118-79-6 ^b | - |
| - | Piperazine (poly)phosphate | 1951-97-9 | 217-775-8 |
| - | Piperazine pyrophosphate | 66034-17-1 | 457-330-7 |
| - | t-Butylphenyl diphenyl phosphate | 83242-23-3 | 260-391-0 |
| - | Tetramethylene dimelamine | - | - |
| - | Trimethylene dimelamine | - | - |
| - | Tris-(isopropylphenyl) phosphate | 26967-76-0 | 248-147-1 |
| - | Xylenyl diphenyl phosphate | 25155-24-2 | - |

^aCAS-number for Dechlorane 604; ^buncertain CAS-number.

2.1 Patent database search

Within the CPC-system, flame retarded polymers and beds are given the classifications, C08L2201/02 and Y10S5/954, respectively. These CPC-codes were used to conduct the database search. When using the CPC-code for fireproof beds, two patents were retrieved, *Flame resistant filler cloth and mattresses incorporating same* (US patent 9,006,118) published in April, 2015, and *Fire resistant flange for removable top panels for use in mattress assemblies* (US patent 8,893,337) published in November, 2014. FRs mentioned in the patents can be utilized in accordance with embodiments of the invention, but the patents are not restricted to those FRs. In the patent published in 2015, several inorganic FRs (e.g., mono- and diammonium phosphate, boric acid, and ammonium sulfomate) are mentioned together with several organic FRs (including organic

phosphate esters in general, BDE-209, chlorinated and brominated paraffin, and chlorinated and brominated binders), indicating a current or future use of those FRs in these type of products. In the second patent (published in 2014), in addition to inorganic FRs, several organic FRs are mentioned, including organic phosphorus compounds in general, BDE-209, and chlorinated paraffin, again, indicating a current or future use of those FRs in these type of products.

The CPC-code C08L 2201/02 refers to polymers with FR properties. Conducting a search using this CPC-code resulted in 100 found patents, of which the two newest were selected for further evaluation. The first patent refers to an insulated electrical wire for automobile (US patent 9,583,234) and was published in February, 2017. This patent allows the usage of two inorganic FRs and one organic FR, which is sold under the tradename SAYTEX 8010, produced by Albemarle corporation and contains the BFR DBDPE [15]. The second patent (published in February 2017) refers to a cellulose ester-based resin composition (US patent 9,580,580). In this patent, a large number of FRs are mentioned. The organic FRs include triazine ring-containing FRs (i.e. melamine, ammeline, benzoguanamine, acetoguanamine, phthalodiguanamine, melamine cyanurate, melamine pyrophosphate, butylene diguanamine, norbornene diguanamine, methylene diguanamine, ethylene dimelamine, trimethylene dimelamine, tetramethylene dimelamine, hexamethylene dimelamine and 1,3-hexylene dimelamine), organophosphorus compounds (i.e. TMP, TEP, TNBP, TBOEP, TCEP, TDCPP, TPHP, TMPP, CDP, trixylenyl phosphate (TXP), octyl diphenyl phosphate, xylenyl diphenyl phosphate, TiPPP, EHDPP, t-butylphenyl diphenyl phosphate, bis-(t-butylphenyl) phenyl phosphate, TBPP, isopropylphenyl diphenyl phosphate, bis-(isopropylphenyl) phenyl phosphate, and tris-(isopropylphenyl) phosphate, PBDPP, 1,3-phenylene-bis(dixylenyl phosphate), and BADP, and the non-halogenated FR pentaerythritol. Furthermore, a number of organic polyphosphate-based FRs are mentioned, including melamine polyphosphate, piperazine polyphosphate, and piperazine pyrophosphate. Surprisingly, the forbidden legacy FR BDE209 was mentioned in two of the patents. Even though only four different patents were investigated, it was still possible to identify a large number of relatively unknown FRs, including brominated paraffins, melamine, ammeline, benzoguanamine, acetoguanamine, phthalodiguanamine, melamine cyanurate, melamine pyrophosphate, butylene diguanamine, norbornene diguanamine, methylene diguanamine, ethylene dimelamine, trimethylene dimelamine, tetramethylene dimelamine, hexamethylene dimelamine, 1,3-hexylene dimelamine, TXP, octyl diphenyl phosphate, xylenyl diphenyl phosphate, t-butylphenyl diphenyl phosphate, bis-(t-butylphenyl) phenyl phosphate, isopropylphenyl diphenyl phosphate, bis-(isopropylphenyl) phenyl phosphate, and tris-(isopropylphenyl) phosphate, 1,3-phenylene-bis(dixylenyl phosphate), pentaerythritol, melamine polyphosphate, piperazine polyphosphate, and piperazine pyrophosphate. To determine the environmental relevance of these chemicals (and thus the need for screening), production and use data would be needed in combination with risk assessment. Other more known FRs were also identified within the patents, of which the majority are OPFRs, indicating a future interest for this class of FRs. Based only on this type of data it is not possible to

determine which FRs to include in future screening as the environmental relevance is difficult to estimate due to lacking information for many compounds.

2.2 Use in the European Union

According to the ECHA database *Registered substances*, 10 000 to 100 000 tons are used in the EU annually of chlorinated paraffins, DBDPE, and TEP (Table 2). Other FRs, including EHDPP, PBDPP, TBBPA, TBBPA-BDBPE, TBOEP, TDCIPP, TEHP, TIBP, TNBP, and TPHP, are used in lower amounts (1000 to 10 000 tons per year), while BADP usage is reported as >1000 tons per year. BEH-TEBP, DBNPG, EBTEBPI, HEEHP-TEBP, IDP, PBB-Acr, TTBNPP, and BCMP-BCEP are all used between 100 to 1000 tons annually. TEBP-Anh and PBDMPP are used between 10 to 100 tons per year, while 3-BP is only used as an intermediate and therefore no quantities are available. TCEP, TCIPP, and TBP have not been registered within the REACH-regulation (Registration, Evaluation, Authorisation and restriction of Chemicals) of the EU. However, as discussed below, these three substances have frequently been detected in e.g., the Nordic countries. For 17 FRs (e.g., BATE, Chlordene plus, and Dec 604B), neither CAS- nor EC-number were available, and as a result no search could be made for these compounds. Finally, for the remaining FRs, no data was available in the database, suggesting that they are either not used within the EU or that the use information is confidential. However, due to limitations in the REACH legislation, finished products, such as electronics, that are imported into the EU may still contain FRs not registered within REACH.

For many of the FRs identified through the patent search, CAS- or EC-numbers were not found and thus no search in the ECHA database could be done for these chemicals. Melamine and pentaerythritol were found to be used in very high quantities, 100 000-1000 000 tons per year, while benzoguanamine and melamine cyanurate are used at 10 000-100 000 tons per year, trixylenyl phosphate at 100-1000 tons per year, and piperazine (poly)phosphate at 10-100 tonnes per year. However, several of these chemicals are likely to have other commercial use than as FRs, e.g., as intermediates in the production of other chemicals. One example is melamine, which is used in cooking utensils, paper, and as a fertilizer [16]. Another example is TBBPA of which >25% of the annual amount is transformed into other substances during use (e.g., through synthesis or burning of fuels) according to the Swedish product register (Table S2 in Appendix) [12]. Acetoguanamine is not registered in REACH and can thus be assumed not to be used during production within the EU but can still be present in imported goods. No data was available for ammeline, melamine pyrophosphate, octyl diphenyl phosphate, xylenyl diphenyl phosphate, t-butylphenyl diphenyl phosphate, bis-(t-butylphenyl) phenyl phosphate, isopropylphenyl diphenyl phosphate, bis-(isopropylphenyl) phenyl phosphate, tris-(isopropylphenyl) phosphate, and melamine (poly)phosphate, indicating that they are either not used during manufacture within the EU or that the information is confidential.

Table 2 Amounts of FRs produced/imported into the EU annually (tons) [11].

| Compound | Annual amount (tons) |
|--|------------------------------|
| Melamine | 100 000-1 000 000 |
| Pentaerythritol | 100 000-1 000 000 |
| Benzoguanamine | 10 000-100 000 |
| SCCP/MCCP | 10 000-100 000 |
| DBDPE | 10 000-100 000 |
| Melamine cyanurate | 10 000-100 000 |
| TEP | 10 000-100 000 |
| EHDPP | 1000-10 000 |
| PBDPP | 1000-10 000 |
| TBBPA | 1000-10 000 |
| TBBPA-BDBPE | 1000-10 000 |
| TBOEP | 1000-10 000 |
| TDCIPP | 1000-10 000 |
| TEHP | 1000-10 000 |
| TIBP | 1000-10 000 |
| TNBP | 1000-10 000 |
| TPHP | 1000-10 000 |
| BADP | 1000+ |
| BEH-TEBP | 100-1 000 |
| DBNPG | 100-1 000 |
| EBTEBPI | 100-1 000 |
| HEEHP-TEBP | 100-1 000 |
| IDP | 100-1 000 |
| PBB-Acr | 100-1 000 |
| Trixylenyl phosphate (TXP) | 100-1 000 |
| TTBNPP | 100-1 000 |
| BCMP-BCEP | 100-1 000 |
| PBDMPP | 10-100 |
| TEBP-Anh | 10-100 |
| Piperazine (poly)phosphate | 10-100 |
| Piperazine pyrophosphate | 10-100 |
| 1,3-hexylene dimelamine | - ^a |
| 1,3-phenylene-bis(dixylenyl phosphate) | - ^a |
| TBP | Not registered in REACH |
| 2,4-DBP | No data available |
| 2,6-DBP | No data available |
| 2-BP | No data available |
| 3-BP | Used only as an intermediate |
| 4'-PeBPOBDE208 | No data available |
| 4-BP | No data available |
| Acetoguanamine | Not registered in REACH |
| Cyanurodiamide | No data available |
| ATE | No data available |
| BATE | - ^a |

| | |
|---|-------------------|
| bBDBP | No data available |
| Bis-(isopropylphenyl) phenyl phosphate | No data available |
| Bis-(t-butylphenyl) phenyl phosphate | No data available |
| Brominated paraffins | - ^a |
| BTBPE | No data available |
| Butylene diguanamine | - ^a |
| CDP | No data available |
| Chlordene Plus | - ^a |
| CLP1 | No data available |
| DBE-DBCH | No data available |
| DBHCTD | No data available |
| DBPhP | No data available |
| DDC-DBF | No data available |
| DDC-Ant | No data available |
| Dec-604A/HCTBPH | No data available |
| Dec-604B | No data available |
| Dibutyl chlorendate | No data available |
| DMP | No data available |
| DOPP | No data available |
| DDC-CO | No data available |
| DPhBP | No data available |
| EH-TBB | No data available |
| Ethylene dimelamine | - ^a |
| HBB | No data available |
| HBCYD | No data available |
| Hexamethylene dimelamine | - ^a |
| Isopropylphenyl diphenyl phosphate | No data available |
| mDEP/dDEP | No data available |
| Melamine (poly)phosphate | No data available |
| Melamine pyrophosphate | No data available |
| Methylene diguanamine | - ^a |
| Norbornene diguanamine | - ^a |
| OBTMPI | No data available |
| Octyl diphenyl phosphate | No data available |
| PBB | No data available |
| PBBC | No data available |
| PBCH | No data available |
| PBEB | No data available |
| PBP | No data available |
| PBT | No data available |
| Phthalodiguanamine | No data available |
| T2CPP | No data available |
| T3CPP | No data available |
| TBBBS | No data available |
| TBBPA-BAE | No data available |
| TBBPA-BME | No data available |

| | |
|---|-------------------------|
| TBBPA-DHEE | No data available |
| TBBPS-DBPE | No data available |
| TBCO | No data available |
| TBP-DBPE | No data available |
| TBPP | No data available |
| t-Butylphenyl diphenyl phosphate | No data available |
| TBX | No data available |
| TCBPA | No data available |
| TCEP | Not registered in REACH |
| TCIPP | Not registered in REACH |
| TMPP (<i>m</i>-) | No data available |
| TMPP (<i>o</i>-) | No data available |
| TMPP (<i>p</i>-) | No data available |
| TDBP-TAZTO | No data available |
| TDCPP | No data available |
| TEEdP | No data available |
| Tetramethylene dimelamine | - ^a |
| THP | No data available |
| TiPP | No data available |
| TMP | No data available |
| TPeP | No data available |
| TPP | No data available |
| Trimethylene dimelamine | - ^a |
| Tris-(isopropylphenyl) phosphate | No data available |
| TTBP-TAZ | No data available |
| Xylenyl diphenyl phosphate | No data available |

^aDatabase search not possible due to lacking CAS- and EC-numbers.

2.3 Flame retardant use in Sweden

Quantitative FR data archived at KemI are mostly confidential and thus not publically available. To circumvent the confidentiality, the data was transformed into quantity indices (QI) ranging from 1 to 7, where 7 represents a high usage and 1 represents a low usage. Table 3 shows quantity indices (for FRs with available data, $n = 46$) obtained from KemI, while the whole dataset is given in the Appendix, Table S2. For most FRs, the compiled data represent the early 1990's to 2015.

Based on the data obtained from KemI, no FR was indexed into quantity Group 7. Pentaerythritol was indexed 6, showing an extensive use of this specific compound (Table 3). Furthermore, 18 FRs were indexed into quantity Group 5, meaning that they are used in fairly high volumes in Sweden. These FRs include BADP, CDP, SCCP, DBDPE, EHDPP, IDP, melamine, melamine cyanurate, PBDPP, TBBPA-BDBPE, TBOEP, TCIPP, *p*-TMPP, TEHP, TEPP, TIBP, TNBP, and TPHP. Quantity Group 4 includes BEH-TEBP, BTBPE, DBPhP, DPhBP, EBTEBPI, HEEHP-TEBP, melamine pyrophosphate, T2CPP, TCEP, and

TXP. Smaller volumes are used of bis-(*t*-butylphenyl) phenyl phosphate, TBPP, *o*-TMPP, and TTBNPP, which are all indexed 3. No FR was indexed 2 but four compounds, including ammeline, DBNPG, DDC-CO, and TBBPA were indexed 1, indicating only a small usage in Sweden. Nine FRs (4'-PeBPOBDE208, Acetoguanamine, CLP1, MCCP, DMP, TDCIPP, TMP, and BCMP-BCEP) were indexed 0 meaning that they are not used in Sweden or that the information is confidential. For the remaining FRs listed in Table 2 but not in Table 3, no Swedish quantity data is available.

Table 3 Quantity Index (QI) of FRs in Sweden [12]. Indices were calculated from the Swedish Product Register based on registered use patterns in year 2015.

| Abbreviation | Quantity index ^a |
|---|-----------------------------|
| Pentaerythritol | 6 |
| BADP | 5 |
| CDP | 5 |
| SCCP | 5 |
| DBDPE | 5 |
| EHDPP | 5 |
| IDP | 5 |
| Melamine | 5 |
| Melamine cyanurate | 5 |
| PBDPP | 5 |
| TBBPA-BDBPE | 5 |
| TBOEP | 5 |
| TCIPP | 5 |
| TMPP (<i>p</i> -) | 5 |
| TEHP | 5 |
| TEP | 5 |
| TIBP | 5 |
| TNBP | 5 |
| TPHP | 5 |
| BEH-TEBP | 4 |
| BTBPE | 4 |
| DBPhP | 4 |
| DPhBP | 4 |
| EBTEBPI | 4 |
| HEEHP-TEBP | 4 |
| Melamine pyrophosphate | 4 |
| T2CPP | 4 |
| TCEP | 4 |
| TXP | 4 |
| bis-(<i>t</i> -butylphenyl) phenyl phosphate | 3 |
| TBPP | 3 |
| TMPP (<i>o</i> -) | 3 |
| TTBNPP | 3 |
| Ammeline | 1 |
| DBNPG | 1 |

| | |
|-----------------------|---|
| DDC-CO | 1 |
| TBBPA | 1 |
| 4'-PeBPOBDE208 | 0 |
| Acetoguanamine | 0 |
| Benzoguanamine | 0 |
| CLP1 | 0 |
| MCCP | 0 |
| DMP | 0 |
| TDCIPP | 0 |
| TMP | 0 |
| BCMP-BCEP | 0 |

³Indices range from 1-7 where 7 represents the highest quantity.

3. Flame Retardants in the environment

This chapter is based on the literature review and summarizes efforts made to detect FRs in various environmental matrices such as indoor dust, indoor and outdoor air, water, sediment, sludge, soil, atmospheric deposition, plants and animals including humans. It includes a broad range of FRs; however, legacy FRs PBDEs and HBCDD are excluded from the compilation. In total, about 60 references were reviewed, and in those, 66 different FRs were screened for in one or more matrices. The cited literature encompasses international peer-reviewed literature and reports from environmental authorities within the Nordic countries (e.g., the Swedish Environmental Protection Agency and the Norwegian Pollution Control Authority). A special emphasis was paid to studies conducted within the Nordic countries.

3.1 Indoor air and dust

In total, 50 different alternative FRs were analysed in indoor air in the cited literature [17-25]. In public areas, 25 FRs were detected, in homes 22 FRs and in offices 20 FRs (Appendix, Table S3). Six brominated FRs (DBDPE, DBE-DBCH, HBB, PBEB, PBT and TBX), ten non-halogenated OPFRs (DOPP, EHDPP, TBOEP, TMPP, TEP, TEHP, TIBP, TNBP, TPHP, and TPP), and three halogenated OPFRs (TCEP, TCIPP, and TDCIPP) were detected above 1 ng m^{-3} . Cequier et al. (2014) [18] and Schlabach et al. (2011) [20] detected DBDPE in air of Norwegian living rooms, school classrooms, and offices at concentrations up to 1 ng m^{-3} , while it was not detected by Møskeland et al. (2009) [23] in an electronics store. DBE-DBCH have been detected up to 4.1 ng m^{-3} in living rooms and school classrooms in Norway, with a detection frequency of 100% [18]. In the same study, TBX was detected at concentrations up to 2.8 ng m^{-3} with a detection frequency of 38% and 17% in living rooms and classrooms, respectively. No other study targeted TBX in indoor air. Remberger et al. (2014) analysed indoor air in a recycling hall for electronics in Sweden and found concentrations of $220\text{-}530 \text{ ng m}^{-3}$, $6.7\text{-}8.3 \text{ ng m}^{-3}$, $1\ 400\text{-}1\ 600 \text{ ng m}^{-3}$, not detected (n.d.)- 12 ng m^{-3} , and $14\text{-}25 \text{ ng m}^{-3}$ for DBDPE, DBE-DBCH, HBB, PBEB, and PBT, respectively [25].

Non-halogenated OPFRs have frequently been analysed and detected in indoor air. Two exceptions are DOPP (analysed in one study, concentrations up to 4.8 ng m^{-3} [21]) and TPP (the same study, concentrations up to 8.4 ng m^{-3}). EHDPP has been detected in a number of studies (e.g., [17-19]) at concentrations up to 14 ng m^{-3} . Also, TBOEP has been frequently detected in indoor air (e.g., [17, 18, 24]) at concentration up to $1\ 300 \text{ ng m}^{-3}$. Interestingly, both Cequier et al. (2009) [18] and Green et al. (2007) [22] reported detection frequencies of 100% in air samples from Norwegian homes and public places. On the other hand, Luongo et al. (2015) detected TBOEP in only 5% of their sampled Swedish homes [19]. When interpreting the results, it should be kept in mind that detection frequencies are dependent on the detection limits. TEP and TIBP have been detected at concentrations up to 300 ng m^{-3} and 66 ng m^{-3} , respectively, with detection frequencies up to 100% in several different indoor air environments in Sweden and Norway [17, 19]. For TNBP, three different studies in Sweden and Norway reported detection frequencies of 100% with concentrations up to

320 ng m⁻³ [17-19, 22]. The by far highest reported concentration of any FR in the reviewed literature is 47 000 ng m⁻³, which was determined for TPHP in a Norwegian shopping center [22]. Concentrations of TPHP were generally high in this study of public places in Norway, ranging between 2 300 ng m⁻³ and 47 000 ng m⁻³, with 100% detection frequency. In Sweden, the highest reported concentration of TPHP is 25 ng m⁻³ (in a home) [19]. The detection frequency in this study was 15%. The three halogenated OPFRs with levels above 1 ng m⁻³ have all been frequently detected in both Sweden and Norway at rather high concentrations and detection frequencies. The highest reported concentrations are 730 ng m⁻³, 1 200 ng m⁻³, and 150 ng m⁻³ of TCEP, TCIPP, and TDCIPP, respectively [19, 21]. FRs that have been detected at lower concentrations (<1 ng m⁻³) in indoor air in Sweden and Norway but with high detection frequencies (≥50%) in at least one study include TBP-AE, HBB, PBB, PBT, TBP-DBPE, and TMPP. BEH-TEBP, BTBPE, CLP1, syn-DDC-CO, anti-DDC-CO, and EH-TBB have also been detected but at lower detection frequencies.

In dust, 44 different alternative FRs have been targeted for in the cited literature (Table S4) [17-19, 24-33]. A high number of FRs were detected in homes ($n = 30$) and public places ($n = 39$), but a variety of FRs were also detected in offices ($n = 12$) and special point sources, such as inside cars and recycling halls for electronics ($n = 22$). OPFRs and BFRs are frequently found in dust. Detection frequencies of >50% have been reported in at least one study for the BFRs BEH-TEBP, BTBPE, DBDPE, DBE-DBCH, syn-DDC-CO, anti-DDC-CO, EH-TBB, HBB, PBB, PBT, and TBP-DBE with concentrations up to 0.81, 0.23, 23, 0.17, 0.59, 0.31, 0.25, 8.2, 0.00068, 0.064, and 0.021 µg g⁻¹, respectively. For OPFRs, detection frequencies >50% were reported in at least one study for EHDPP, TBOEP, TCEP, TCIPP, TMPP, TDCIPP, TEP, TiBP, TNBP, and TPHP with maximum concentrations of 540, 11 000, 1 800, 370, 36, 860, 4.7, 47, 160, and 390 µg g⁻¹, respectively. In general, OPFRs showed higher concentrations than the BFRs. Other less frequently detected (<50%, but still detected in at least one study) FRs in dust include CLP1, DOPP, PBEB, TBX, and THP. Also TEEedP have been detected in indoor dust, but detection frequencies were not reported.

3.2 Outdoor air

In total, 38 FRs were analysed in outdoor air samples, out of which 34 were detected in at least one study (Appendix, Table S5) [20, 22, 23, 25, 33-38]. TPHP is the FR that has been detected at the highest concentration (12 000 pg m⁻³) [34]. Surprisingly, this sample, which was collected at a background location in northern Finland, showed about twelve times higher concentration of TPHP than samples collected in urban areas in Norway (which were up to 1 000 pg m⁻³) [22]. The detection of TPHP in a remote area indicates a potential of long-range transport by this FR. However, TPHP was not detected at background locations in Norway [22]. Other FRs detected in the same area as the high levels of TPHP include TCEP, TCIPP, TDCIPP, TMP (only tentatively identified), and TNBP, found at concentrations of 1.6, 810, 20, 24, and 280 pg m⁻³ [34], respectively. These concentrations are considerably lower than concentrations measured in Norwegian urban areas, which have been found to be up to 3700, 3700, 72, and 3700 pg m⁻³ of TCEP, TCIPP, TDCIPP and TNBP, respectively [22]. In addition, several other FRs have been detected at

relatively high concentration in urban areas in Norway, including EHDPP, TBOEP, TIBP, and BCMP-BCEP at concentrations up to 1 100, 340, 4 400, and 5 200 pg m^{-3} , respectively. EHDPP, TBOEP and TIBP have also been detected in remote areas, but at lower concentrations, namely up to 260, 150 and 230 pg m^{-3} , respectively [22]. Not surprisingly, the detection frequencies were generally higher in the urban areas, e.g., 100% for TCEP, TCIPP, and TIBP, compared to in the remote areas (14% for TCEP and TCIPP, and 86% for TIBP) [22]. Regarding BFRs, some FRs have been found at relatively high concentrations in outdoor air in China. BTBPE, DBDPE, and TBBPA-BDBPE have been detected at concentrations of 3.8-67 pg m^{-3} , 402-3578 pg m^{-3} , and 130-1 200 pg m^{-3} , respectively, in the Pearl River delta [33]. In Sweden, Norway and Finland, outdoor air concentrations found for these FRs are considerably lower, with DBDPE ranging from n.d. to 44 pg m^{-3} [20, 23, 25, 38], while concentrations of BTBPE have been detected from n.d. to 2.2 pg m^{-3} [20]. TBBPA-BDBPE has, to our knowledge, never been detected in outdoor air in the Nordic countries. However, other BFRs have been detected at similar or higher concentrations in Sweden, Norway and Denmark, including anti-DDC-CO, detected up to 120 pg m^{-3} , syn-DDC-CO (up to 42 pg m^{-3} , and TBBPA, up to 280 pg m^{-3} [20]. Also TBP, 2,4-DP, and DBE-DBCH (individual isomers) have been detected at concentrations up to 27 pg m^{-3} , 21 pg m^{-3} and 18 pg m^{-3} , respectively [20]. However, bromophenols, such as TBP and 2,4-DBP, also occur naturally [23]. Other BFRs that have been detected in Nordic outdoor air (at concentrations $\leq 10 \text{ pg m}^{-3}$) include 2/3-BP, 2,6-DBP, 4-BP, TBP-AE, BATE, BEH-TEBP, EH-TBB, HBB, PBEB, PBP, PBT, and TBP-DBPE [20, 38]. In a study by Haglund et al. (2015), with samples collected at background sites in Sweden and Finland, TBP, BTBPE, DBE-DBCH, anti-DDC-CO, syn-DDC-CO, EH-TBB, HBB, PBEB, and PBT were detected in all samples, while BEH-TEBP and DBDPE were detected in 92% and 83% of the samples, respectively [38]. Even though concentrations were all $<1 \text{ pg m}^{-3}$, the high detection frequency in background areas shows the wide spread of these type of FRs.

3.3 Atmospheric deposition

A few studies have analysed emerging FRs in atmospheric deposition. In total, 19 FRs have been targeted, and 14 were detected in at least one study (Appendix, Table S6) [25, 34, 39, 40]. Newton et al. (2013) detected DBE-DBCH and DDC-CO in wet and dry deposition at $3.1 \pm 3.6 \text{ ng m}^{-2} \text{ month}^{-1}$ and $22 \pm 21 \text{ ng m}^{-2} \text{ month}^{-1}$, respectively, in a boreal catchment in Sweden (Krycklan Catchment Study area) and $3.5 \pm 2.8 \text{ ng m}^{-2} \text{ month}^{-1}$ and $1.1 \pm 0.52 \text{ ng m}^{-2} \text{ month}^{-1}$, respectively, in Abisko [40]. Both locations are rather remote, thus indicating potential of long-range air transport of these FRs. In another study, DBDPE, DBE-DBCH, HBB, PBEB, and PBT were analysed in deposition samples from the Swedish west coast, but no FRs were detected [25]. Several OPFRs have been detected in wet and dry deposition in background areas in Finland. TCEP, TCIPP, TMP (tentatively identified), and TNBP were found at levels of 16 500, 15 300, 33, and 6 900 $\text{ng m}^{-2} \text{ month}^{-1}$, respectively, while TBOEP, TMPP, TDCIPP, TEHP, TPHP, and TPP were not detected [34]. TBOEP, TCEP, TCIPP, p-TMPP, TDCIPP, TEHP, TMP (tentatively quantified), TNBP, and TPHP have been found at higher concentrations in snow close to a road (concentration ranges in ng L^{-1} : 4-12, 7-12,

110-170, n.d., 8-230, n.d.-130, n.d.-10, 11-20, and 4-68, respectively) and an airport (7-94, 29-39, 100-210, 260-9 900, 4-15, 1-95, 11-28, 2 100-23 000, and 120-830 ng L⁻¹, respectively) compared to a background location in Sweden (2, 7, 68, n.d., 29, n.d., n.d., 19, and 4 ng L⁻¹, respectively), showing that traffic may act as a point source of these FRs [34]. Similarly, urban rain was found to contain similar or higher levels of some OPFRs (i.e. TBOEP, TCEP, TCIPP, TIBP, and TNBP) than background rain and snow in Germany [39].

3.4 Water

In total, 35 out of 39 targeted emerging FRs have been detected in water, including freshwaters, marine waters and effluents from point sources such as WWTPs (Appendix, Table S7) [22, 23, 25, 35-37, 41-52]. A number of studies have investigated FR concentrations in water from point sources. The highest reported concentration was from a Swedish WWTP, where levels up to 52 000 ng L⁻¹ were reported for TNBP [42]. Also other OPFRs showed high concentrations in this study, DOPP showed concentrations of n.d.-2 000 ng L⁻¹, TBOEP 3 100-35 000 ng L⁻¹, TCEP 90-450 ng L⁻¹, TCIPP 1 100-24 000 ng L⁻¹, TDCIPP 130-450 ng L⁻¹, TEHP n.d.-130 ng L⁻¹, TMP n.d.-584 ng L⁻¹, and TPHP 41-290 ng L⁻¹. Generally, slightly lower concentrations have been found for TBOEP (500-9 200 ng L⁻¹), TCIPP (400-2 900 ng L⁻¹), TDCIPP (86-820 ng L⁻¹), and TNBP (160-2 800 ng L⁻¹) in WWTPs in Germany and Norway, while reported concentrations of TCEP (130-2 500 ng L⁻¹) and TPHP (1 700-14 000 ng L⁻¹) were higher [22, 41]. In Norway, concentrations of EHDPP in WWTPs ranged from 250-710 ng L⁻¹ [22]. For BFRs, reported concentrations in water from WWTPs are generally lower than for OPFRs. The highest concentration of a BFR has been reported for DBDPE (250-1 500 ng L⁻¹) and PBEB (up to 91 ng L⁻¹) in stormwater in Gothenburg and of DBDPE (330-1800 ng L⁻¹) and HBB (11-1 200 ng L⁻¹) in extinguishing water from a fire [25]. DBDPE has also, together with BTBPE and TBBPA-BDBPE, been found at relatively high concentrations close to point sources in Norway (up to 185.7, 107.0, and 159.6 ng L⁻¹, respectively) [23] while DBDPE concentration in a WWTP in Sweden ranged up to 420 ng L⁻¹ [25]. In Norwegian WWTPs, TBBPA-BDBPE has showed concentrations up to 18 ng L⁻¹ [47]. In conclusion, WWTP do act as point sources for the emission of OPFRs and BFRs.

Several studies have investigated the presence of emerging FRs in European rivers [41, 43, 46, 48-52]. Again, OPFRs dominate the FR content with e.g., EHDPP concentrations up to 46 ng L⁻¹ [51], TBOEP up to 4600 ng L⁻¹ [51], TCEP up to 330 ng L⁻¹ [51], TCIPP up to 26000 ng L⁻¹ [52], TDCIPP up to 200 ng L⁻¹ [51], and TIBP up to 1200 ng L⁻¹ [51]. The four BFRs detected at the highest concentration in any European river are TEBP-Anh, TBBPA, TCBPA, and EHTBB with maximum concentrations of 67, 62, 56, and 24 ng L⁻¹, respectively [49]. All the highest BFR concentrations were detected in the same study [49]. Few other studies have, to our knowledge, included these FRs. Other FRs detected in the Swedish rivers in the same study include TBP (n.d.-20 ng L⁻¹), BTBPE (n.d.-4.7 ng L⁻¹), syn-DDC-CO (n.d.-12 ng L⁻¹), EHDPP (n.d.-9.2 ng L⁻¹), HBB (n.d.-0.13 ng L⁻¹), PBB-Acr (n.d.-2.6 ng L⁻¹), PBT (n.d.-2.5 ng L⁻¹), TBX (n.d.-0.022 ng L⁻¹), TCEP (n.d.-14 ng L⁻¹), TCIPP (n.d.-30 ng L⁻¹), *o*-TMPP (n.d.-1.4 ng L⁻¹), *p*-TMPP (n.d.-11 ng L⁻¹),

Σ TDCIPP+TEHP (n.d.-48 ng L⁻¹), TNBP (n.d.-24 ng L⁻¹), TPHP (n.d.-66 ng L⁻¹), and TTBNPP (n.d.-3.6 ng L⁻¹). Finally, TEP and TPP have, to our knowledge, only been analysed in one study each and were found at concentrations of 13-51 ng L⁻¹ and 40 ng L⁻¹, respectively [41, 43].

Five OPFRs have been detected in rural/remote surface waters, which indicate potential of long-range transport [44]. These OPFRs include TBOEP found at mean concentrations up to 31 ng L⁻¹, TCEP up to 33 ng L⁻¹, TCIPP up to 312 ng L⁻¹, TIBP up to 11 ng L⁻¹, and TNBP up to 7 ng L⁻¹ [44]. Four BFRs (i.e. anti-DDC-CO, syn-DDC-CO, HBB, and TBP-DBPE) have been detected in marine water at very low concentrations (<0.02 ng L⁻¹) [36, 37].

3.5 Sediment and sludge

In total, 44 FRs have been targeted in sediment and sewage sludge [20, 22, 23, 25, 33, 35, 42, 43, 47, 48, 51, 53-60]. Out of these, 34 and 32 compounds have been detected in sediment and sludge, respectively (Appendix, Table S8 and S9). Concentrations are generally higher in sewage sludge than in sediment. Marklund et al. (2005) analysed OPFRs in Swedish sewage sludge and found comparably high concentrations [42]. EHDPP was found at concentrations ranging from 320-4 600 ng g⁻¹ dw, TBOEP from nd-1900 ng g⁻¹ dw, TCEP from 6.6-110 ng g⁻¹ dw, TCIPP from 61-1 900 ng g⁻¹ dw, TDCIPP from 3.3-260 ng g⁻¹ dw, TIBP from 27-2 700 ng g⁻¹ dw, TNBP from 39-850 ng g⁻¹ dw, and TPHP from 52-320 ng g⁻¹ dw. Other studies have reported concentrations of BFRs in sludge from the Nordic countries, with concentrations of TBP up to 100 ng g⁻¹ dw, 2,4-DBP up to 40 ng g⁻¹ dw, TBP-AE up to 27 ng g⁻¹ dw, BATE up to 4.1 ng g⁻¹ dw, BEH-TEBP up to 42 ng g⁻¹ dw, BTBPE up to 3.9 ng g⁻¹ dw, DBDPE up to 190 ng g⁻¹ dw, α -DBE-DBCH up to 4.7 ng g⁻¹ dw, β -DBE-DBCH up to 2.6 ng g⁻¹ dw, $\Sigma\gamma+\delta$ -DBE-DBCH up to 1.7 ng g⁻¹ dw, anti-DDC-CO up to 25 ng g⁻¹ dw, syn-DDC-CO up to 14 ng g⁻¹ dw, EH-TBB up to 2.6 ng g⁻¹ dw, HBB up to 1.6 ng g⁻¹ dw, PBP up to 3.5 ng g⁻¹ dw, PBT up to 5.2 ng g⁻¹ dw, TBBPA up to 59 ng g⁻¹ dw, and TBP-DBPE up to 120 ng g⁻¹ dw [20, 23, 25, 35, 47]. Similar concentrations of DBDPE, anti-DDC-CO, and syn-DDC-CO have been reported from Spain [59]. In the Spanish study, low concentrations (< 1 ng g⁻¹ dw) of DDC-DBF and DDC-Ant were also reported. Furthermore, low concentrations have been reported for TCBPA in sludge from Canada, while TBBPA concentrations were higher (2.1-28 ng g⁻¹ dw) [60]. Elevated concentrations of some FRs have been detected in the Pearl River delta, China, with high concentrations of DBDPE and TBBPA-BDBPE up to 2 000 and 9 000 ng g⁻¹ dw, respectively, while BTBPE concentration (up to 1.66 ng g⁻¹ dw) were more similar to concentrations detected on other locations [33].

Also in sediment from Pearl River delta, high concentrations of some FRs have been found, including BTBPE (0.05-22 ng g⁻¹ dw), DBDPE (39-360 ng g⁻¹ dw), and TBBPA-BDBPE (n.d.-2 300 ng g⁻¹ dw) [33]. Even higher levels of DBDPE (up to 440 ng g⁻¹ dw) have been detected in sediment from Spain [51]. However, in two other Spanish studies, concentrations of DBDPE were lower (up to 32 ng g⁻¹ dw and 24 ng g⁻¹ dw) [55, 59]. In the Nordic countries, reported DBDPE concentrations are lower, up to 2.4 ng g⁻¹ dw

[20]. BTBPE has been detected up to $4.5 \text{ ng g}^{-1} \text{ dw}$ in sediment in Norway, while TBBPA-BDBPE has not been detected [23, 47]. Other BFRs detected in sediment from the Nordic countries include TBP (up to $7.8 \text{ ng g}^{-1} \text{ dw}$), 2,4-DBP (up to $2.9 \text{ ng g}^{-1} \text{ dw}$), BEH-TEBP (up to $3.3 \text{ ng g}^{-1} \text{ dw}$), DBE-DBCH ($< 1 \text{ ng g}^{-1} \text{ dw}$), anti-DDC-CO (up to $2.5 \text{ ng g}^{-1} \text{ dw}$), syn-DDC-CO (up to $0.99 \text{ ng g}^{-1} \text{ dw}$), EH-TBB (up to $0.21 \text{ ng g}^{-1} \text{ dw}$), HBB (up to $1.8 \text{ ng g}^{-1} \text{ dw}$), PBEB (up to $0.1 \text{ ng g}^{-1} \text{ dw}$), PBT (up to $2.7 \text{ ng g}^{-1} \text{ dw}$), TBBPA (up to $16 \text{ ng g}^{-1} \text{ dw}$), and TBBPA-BAE (up to $2.4 \text{ ng g}^{-1} \text{ dw}$) [20, 23]. In North America, BTBPE concentrations up to $1.6 \text{ ng g}^{-1} \text{ dw}$, anti-DDC-CO concentrations up to $120 \text{ ng g}^{-1} \text{ dw}$, syn-DDC-CO concentrations up to $34 \text{ ng g}^{-1} \text{ dw}$, and PBEB concentration up to $0.1 \text{ ng g}^{-1} \text{ dw}$ have been reported [54, 57]. Four studies identified in this literature review have analysed OPFRs in sediments from Austria, Norway, and Spain. The highest concentration of any OPFR was found of TCIPP in Austrian sediments, with concentrations up to $1300 \text{ ng g}^{-1} \text{ dw}$ [43]. In Norway and Spain, concentrations of TCIPP were lower, up to 54 and 370 $\text{ng g}^{-1} \text{ dw}$, respectively [51, 56]. Other detected OPFRs in the four studies include EHDPP (up to $63 \text{ ng g}^{-1} \text{ dw}$), TBOEP (up to $130 \text{ ng g}^{-1} \text{ dw}$), TCEP ($160 \text{ ng g}^{-1} \text{ dw}$), *o*-TMPP (up to $1.5 \text{ ng g}^{-1} \text{ dw}$), *p*-TMPP (up to $290 \text{ ng g}^{-1} \text{ dw}$), TDCIPP (up to $8.7 \text{ ng g}^{-1} \text{ dw}$), TEHP (up to $290 \text{ ng g}^{-1} \text{ dw}$), TEP (up to $81 \text{ ng g}^{-1} \text{ dw}$), TIBP (up to $8.4 \text{ ng g}^{-1} \text{ dw}$), TNBP (up to $50 \text{ ng g}^{-1} \text{ dw}$), and TPHP (up to $160 \text{ ng g}^{-1} \text{ dw}$) [43, 51, 56]. Out of the detected OPFRs, high detection frequencies ($>50\%$) were reported for EHDPP, TBOEP, TCEP, TCIPP, *p*-TMPP, TEHP, TIBP, TNBP, and TPHP, indicating wide spread in sediments [56]. In addition to the previously mentioned studies, Green et al. (2008) reported concentrations of several OPFRs in sediment and sludge from Norway. Reported concentrations are in the unit of $\mu\text{g kg}^{-1}$ loss-of-ignition weight [22]. Detected OPFRs in this study include EHDPP, TBOEP, TCEP, TCIPP, TDCIPP, TIBP, TNBP, TPHP, and BCMP-BCEP (for concentrations, see Appendix, Table S9).

3.6 Soil and plants

Studies of emerging FRs in soil are scarce. However, three BFRs have been detected in farmland soil in China, including soils from an e-waste processing area (Appendix, Table S10) [33]. BTBPE were detected up to $6.2 \text{ ng g}^{-1} \text{ dw}$, DBDPE up to $36 \text{ ng g}^{-1} \text{ dw}$, and TBBPA-BDBPE up to $60 \text{ ng g}^{-1} \text{ dw}$ [33]. In Norway, BEH-TEBP has been detected at $1.0 \text{ ng g}^{-1} \text{ dw}$, with 100% detection frequency [58].

Three studies report concentrations of emerging FRs in mosses and tree needles collected nearby potential point sources in the Nordic countries (Appendix, Table S11) [20, 23, 35]. In total, 23 different FRs were targeted, out of which 14 were detected in at least one study. In needles, most of the targeted FRs were not detected. Nevertheless, a few compounds including DBDPE (n.d.- $0.1 \text{ ng g}^{-1} \text{ ww}$), HBB (n.d.- $0.05 \text{ ng g}^{-1} \text{ ww}$), and TBBPA-BDBPE (n.d.- $0.16 \text{ ng g}^{-1} \text{ ww}$) were detected. [23, 35]. DBDPE and HBB have also been detected in mosses at concentrations up to $0.1 \text{ ng g}^{-1} \text{ dw}$ (note the different unit compared to needles) [23, 35]. On Faroe Islands TBP, 2,4-DBP, BEH-TEBP, BTBPE, DBDPE, DBE-DBCH, anti-DDC-CO, syn-DDC-CO, HBB, PBEB, PBT, and TBP-DBPE have been detected in mosses close to an incineration plant (concentrations shown in Table S12) [20].

3.7 Wildlife

A wide variety of different wildlife animal species have been analysed for emerging FRs, including e.g., arctic fox, polar bear, moose, seal, cod, salmon, mussels, crab, common eider and herring gull [22, 23, 25, 33, 35, 47, 53, 56, 57, 61-67]. Starting with the mammals, no OPFRs have to our knowledge been targeted in mammals (except humans), while BFRs have been analysed in a number of studies (e.g., [57, 61]). The highest concentration of an emerging FR was detected for TBP in harbour seal ($160 \pm 84 \text{ ng g}^{-1} \text{ ww}$) with a detection frequency of 100% [58] (Appendix, Table S12). However, TBP is not only a FR but also a naturally occurring compound [23]. TBP has also been detected in moose ($81 \pm 45 \text{ ng g}^{-1} \text{ ww}$), mouse ($54 \pm 43 \text{ ng g}^{-1} \text{ ww}$), and shrew liver ($27.1 \pm 6.9 \text{ ng g}^{-1} \text{ ww}$) in Norway, all with detection frequencies $\geq 88\%$ [58]. Sagerup et al. (2010) detected TBP in ringed seal liver ($0.050 \pm 0.023 \text{ ng g}^{-1} \text{ ww}$) in the Norwegian Arctic but not in arctic fox liver or polar bear plasma [63]. On the other, Harju et al. (2013) detected TBP in polar bear plasma at $26 \pm 15 \text{ ng mL}^{-1}$ (note the unit) with 100% detection frequency [61]. In the same study, BEH-TEBP was detected in polar bear plasma at $0.15 \pm 0.16 \text{ ng mL}^{-1}$ with 95% detection frequency. The same compound has also been detected in ringed seal liver ($0.57 \pm 0.20 \text{ ng g}^{-1} \text{ ww}$) from the Arctic with 60% detection frequency [63], showing the ubiquitous spread. Another ubiquitously spread FR is DBDPE. It has, for example, been detected in moose, mouse, shrew and Arctic harbor seal livers at concentrations up to $26 \pm 9 \text{ ng g}^{-1} \text{ ww}$, with 100% detection frequency [61]. It has also been detected in polar bear adipose in Canada [62] and in polar bear and ringed seal plasma in the Norwegian arctic [61]. Chlorinated paraffins (CCPs) have to our knowledge only recently been addressed as FRs that may pollute the environment. Both SCCPs and MCCPs have been detected in polar bear and ringed seal plasma in the Arctic [61]. Concentrations of SCCPs were determined to $4\text{-}5 \text{ ng mL}^{-1}$ in both polar bear and seal plasma, while the MCCPs showed slightly lower concentrations of $2\text{-}3 \text{ ng mL}^{-1}$. Detection frequencies were high ($\geq 90\%$) in both species. Two other other chlorinated FRs, anti-DDC-CO and syn-DDC-CO, have been detected at similar concentrations (up to approx. $3.5 \text{ ng g}^{-1} \text{ lw}$, note the unit) in harbor seal blubber in the US [57]. Detection frequencies of both DDC-COs in adult seals were 100%, and concentrations were higher in adult seals than in cubs. Finally, another three FRs have been detected in mammal tissues; EH-TBB has been detected in arctic fox liver, polar bear plasma, and ringed seal liver at concentrations up to $3.5 \pm 2.5 \text{ ng g}^{-1} \text{ ww}$ with detection frequencies $\geq 90\%$ [63]. HBB has been detected in polar bear plasma up to $3.4 \text{ ng g}^{-1} \text{ lw}$ [62], and PBEB has been detected in polar bear plasma up to $1.7 \text{ ng g}^{-1} \text{ lw}$ [62] and in harbor seal blubber up to $0.5 \text{ ng g}^{-1} \text{ lw}$ [57]. In adult seals, the detection frequency was 100%, while it was slightly lower (83%) in the younger seals.

A number of Nordic studies have analysed emerging FRs in different types of water living organisms, such as e.g., fish and mussels (Appendix, Table S13) (e.g., [35, 53, 65]). In total, 37 different FRs have been detected in this type of biota. As previously mentioned, TBP is a naturally occurring compound but also produced and used as a FR [23]. It has been detected in aquatic biota in several studies up to $130 \text{ ng g}^{-1} \text{ ww}$

[23] and up to $86 \mu\text{g g}^{-1} \text{lw}$ [20], and often with high detection frequency [61]. A rarely analysed FR is DPhBP, which has been detected in freshwater carp/perch in Sweden close to a potential point source at concentrations up to $2000 \text{ ng g}^{-1} \text{lw}$ [64]. Concentrations in marine mussel were, however, much lower (up to $0.5 \text{ ng g}^{-1} \text{lw}$), and it was not detected in a Norwegian study [56]. BEH-TEBP has been detected at $0.72 \pm 0.29 \text{ ng g}^{-1} \text{ww}$ in whole capelin in the Norwegian Arctic with 90% detection frequency [63]. However, both lower concentrations [58] and detection frequencies have been observed in other studies [58, 65]. BTBPE was not detected in US [57] and Canada [65], but has been detected in fish both in the Nordic countries and China at concentrations up to $0.20 \text{ ng g}^{-1} \text{ww}$ and $0.15 \text{ ng g}^{-1} \text{lw}$, respectively [20, 33]. Regarding chlorinated compounds, Chlordene plus (up to $8.7 \pm 7.3 \text{ ng g}^{-1} \text{lw}$), DDC-DBF (up to $24 \pm 20 \text{ ng g}^{-1} \text{lw}$), Dec 604B (up to $140 \pm 130 \text{ ng g}^{-1} \text{lw}$), anti-DDC-CO (up to $2.8 \text{ ng g}^{-1} \text{lw}$), and syn-DDC-CO (up to $9.1 \text{ ng g}^{-1} \text{lw}$) have all been detected in northern pike and muskellunge in Canada at detection frequencies ranging from 36% up to 91% [65]. DDC-CO (anti- and syn-) has also been detected in fish and mussel in the Nordic countries at concentrations up to approx. $25 \text{ ng g}^{-1} \text{ww}$ [20] while anti-DDC-CO also has been detected in fish from the US (up to $3.7 \text{ ng g}^{-1} \text{lw}$) with detection frequencies up to 83% [57]. Chlorinated paraffins (short- and medium-chained) have been detected in Atlantic and polar cod liver at concentrations up to $10 \pm 11 \text{ ng g}^{-1} \text{ww}$ with detection frequencies up to 100% [61]. Another FR with high detection frequencies is DBDPE. It has, for example, been detected in brown trout liver and mussels in the Nordic countries at concentrations up to $11 \pm 9 \text{ ng g}^{-1} \text{ww}$ and also in cod in the Arctic with 100 % detection frequency [58]. DBDPE was, however, only rarely detected in fish from Canada [65] and not at all in other studies from Norway [23], Sweden [25] and China [33]. In a study by Sundkvist et al. (2010), several OPFRs were detected in aquatic biota [64]. DBPhP was found at concentrations up to $3\ 300 \text{ ng g}^{-1} \text{lw}$ in freshwater perch/carp close to point sources, and EHDPP was detected up to $14\ 000 \text{ ng g}^{-1} \text{lw}$ in marine eelpout (in other fish species it was $<190 \text{ ng g}^{-1} \text{ww}$). EHDPP has also been detected in fish from Svalbard (up to $52 \text{ ng g}^{-1} \text{ww}$, note the different unit) [53] and in fish from Norway (up to $1.1 \text{ ng g}^{-1} \text{ww}$) with a detection frequency up to 33% [56]. Sundkvist et al. (2010) [64] also detected TBOEP at concentrations ranging between 240 and $1\ 000 \text{ ng g}^{-1} \text{lw}$ in perch/carp caught close to point sources in Sweden but not in any other sampled fish. Nevertheless, TBOEP has been detected in Burbon liver (up to $410 \text{ ng g}^{-1} \text{ww}$, note the unit) in Norwegian [56] and Canadian fish (up to $9.8 \text{ ng g}^{-1} \text{ww}$) [66]. Sundkvist et al. (2010) [64] also detected TCEP (up to $160 \text{ ng g}^{-1} \text{lw}$), TCIPP (up to $1300 \text{ ng g}^{-1} \text{lw}$), *o*-TMPP (up to $2.5 \text{ ng g}^{-1} \text{lw}$), *p*-TMPP (up to $140 \text{ ng g}^{-1} \text{lw}$), TDCIPP (up to $140 \text{ ng g}^{-1} \text{lw}$), TNBP (up to $4\ 900 \text{ ng g}^{-1} \text{lw}$), and TPHP (up to $180 \text{ ng g}^{-1} \text{lw}$). TCEP, TCIPP, TDCIPP, TNBP, and TPHP have also been detected in other studies at concentrations up to 26, 17, <0.88 , 17, and $44 \text{ ng g}^{-1} \text{ww}$, respectively [22, 53, 56, 66]. Other detected FRs in aquatic biota include 2,4-DBP (up to $6.4 \text{ ng g}^{-1} \text{ww}$) [20], BATE (up to $0.00072 \text{ ng g}^{-1} \text{ww}$) [20], DBE-DBCH (e.g., α + β -isomers, up to $0.14 \text{ ng g}^{-1} \text{ww}$) [25], HBB (up to $0.047 \text{ ng g}^{-1} \text{ww}$ [20], $4.6 \text{ ng g}^{-1} \text{ww}$ [25]), PBEB (up to $0.044 \text{ ng g}^{-1} \text{ww}$ [20], $3.9 \text{ ng g}^{-1} \text{ww}$ [25]), PBT (up to $0.021 \text{ ng g}^{-1} \text{ww}$) [20], T2CPP (up to $8.9 \text{ ng g}^{-1} \text{ww}$)

[53], TBP-DBPE (up to 0.049 ng g⁻¹ ww) [20], TDCIPP (up to 8.1 ng g⁻¹ ww) [53], TEHP (up to 4.6 ng g⁻¹ ww)[53, 56], and TIBP (up to 4.9 ng g⁻¹ ww) [53].

Seven studies within this literature review have reported on concentrations of emerging FRs in birds and/or bird eggs [20, 33, 53, 56, 57, 61, 63]. In total, 29 different FRs have been detected in birds/eggs (Appendix, Table S14). TBP has been detected in common eider liver (detection frequency 90%) and Kittiwake egg (83%) from the Arctic [63], and in common eider (90%) and herring gull eggs (80%) from the Nordic countries [61], at concentrations up to 66 ng g⁻¹ ww. It has also been detected in glaucous gull plasma from the Arctic (31 ± 9 ng mL⁻¹) with 100% detection frequency [61] and in eggs from Sweden and Faroe islands (up to 1.4 ng g⁻¹ ww) [20]. Also BEH-TEBP has been detected in a number of different bird species and eggs from both the Arctic and the Nordic countries. Reported concentrations of BEH-TEBP range between n.d. and 2.0 ± 2.6 ng g⁻¹ ww with detection frequencies of 17-100% [61, 63]. This FR has also been detected up to 0.021 ng g⁻¹ ww in eggs from different wild bird species from Sweden and Faroe Islands and at 0.026 ± 0.001 ng mL⁻¹ in glaucous gull plasma in the Arctic [20, 61]. BTBPE has been detected in birds from an e-waste area in China at concentrations up to 2.4 ng g⁻¹ lw, which is much lower than the concentrations found of DBDPE (up to 120 ng g⁻¹ lw) in the same study [33]. In eggs from Faroe Islands and Sweden, BTBPE has been detected up to 0.042 ng g⁻¹ ww [20], while it was not detected in cormorant eggs from the US [57]. DBDPE has been frequently detected, even in Arctic birds. Concentrations ranged up to 1.0 ± 1.6 ng g⁻¹ ww in eggs, and 6.4 ± 2.62 ng mL⁻¹ in gull plasma, most often with 100% detection frequency [61, 63]. Other BFRs that has been detected in birds include DBE-DBCH (e.g., up to 0.33 ng g⁻¹ ww of the α -isomer) [20], EHDPP (up to 1.2 ± 0.98 ng g⁻¹ ww [63], and 0.18 ng g⁻¹ ww [20]), HBB (up to 0.03 ng g⁻¹ ww), PBEB (up to 0.0014 ng g⁻¹ ww), PBP (up to 0.43 ng g⁻¹ ww), and PBT (up to 0.0063 ng g⁻¹ ww [20]. Regarding chlorinated FRs, Chlordene Plus has been detected in guillemot eggs in the Arctic (0.66 ± 0.37 ng g⁻¹ ww, df 40%) but was not detected in common eider liver or Kittiwake liver [63]. SCCPs and MCCPs have been detected at high frequency (>67%) in eggs from the Arctic at concentrations up to 7.8 ± 8.3 and 4.9 ± 4.9 ng g⁻¹ ww, respectively [61]. In the same study, S/MCCPs were also detected in glaucous gull plasma. Both anti-DDC-CO and syn-DDC-CO have been detected in eggs from different wild bird species from Sweden and Faroe Islands at concentrations up to 0.057 and 0.026 ng g⁻¹ ww [20], while syn-DDC-CO also has been detected in cormorant eggs from the US at concentrations up to 1.1 ng g⁻¹ lw with 100% detection frequency [57]. Only a few studies have analysed OPFRs in birds and bird's eggs, and in general the detection frequencies are lower (usually <31%) than for many brominated and chlorinated FRs. Evenset et al. (2009) detected DBPhP (up to 0.33 ng g⁻¹ dw), EHDPP (up to 28 ng g⁻¹ dw), T2CPP (up to 2.6 ng g⁻¹ dw), TCEP (up to 4.7 ng g⁻¹ dw), TEHP (up to 4.6 ng g⁻¹ dw), TIBP (up to 2.6 ng g⁻¹ dw), TNBP (up to 6.8 ng g⁻¹ dw), and TPHP (up to 3.3 ng g⁻¹ dw) in seabirds from Svalbard [53]. In another study, Leonards et al. (2010) analysed blood and eggs from Norwegian birds [56] and detected DBPhP (up to 5.6 ng g⁻¹ ww), EHDPP (up

to 3.1 ng g⁻¹ ww), TBOEP (up to 57 ng g⁻¹ ww), TCEP (up to 6.1 ng g⁻¹ ww), TCIPP (up to 10 ng g⁻¹ ww), TDCIPP (up to 1.9 ng g⁻¹ ww), TEHP (up to 8.7 ng g⁻¹ ww), and TPHP (up to 14 ng g⁻¹ ww).

3.8 Humans

In Sweden, FRs have been analysed in human blood serum and breast milk [25, 38, 64]. Eight BFRs including 2/3-BP (up to 0.19 ng g⁻¹), TBP (up to 0.27 ng g⁻¹), 2,4-DBP (up to 0.076 ng g⁻¹), 2,6-DBP (up to 0.047 ng g⁻¹), 4-BP (up to 0.16 ng g⁻¹), BTBPE (up to 0.78 ng g⁻¹), EH-TBB (up to 0.0055 ng g⁻¹), and PBEB (up to 0.072 ng g⁻¹) have been detected in blood serum [25, 38] (Appendix, Table S15). Both TBP and BTBPE were detected in all samples. The detected OPFRs in human milk were EHDPP (up to 13 µg g⁻¹ lw), TBOEP (up to 63 µg g⁻¹ lw), TCEP (up to 8.2 µg g⁻¹ lw), TCIPP (up to 82 µg g⁻¹ lw), *p*-TMPP (up to 3.7 µg g⁻¹ lw), TDCIPP (up to 5.3 µg g⁻¹ lw), TNBP (up to 57 µg g⁻¹ lw), and TPHP (up to 11 µg g⁻¹ lw) [64].

4. Estimation of potential exposure in Sweden

In addition to Swedish quantity data, also data concerning the potential exposure to FRs for different environmental compartments was obtained from KemI [12]. Again, the data was transformed into indices to circumvent confidentiality. The exposure index (EI) was calculated based on registered use patterns from the Swedish Product Register and gives an estimate of the potential “worst case” exposure to FRs for different environmental compartments, such as e.g., surface water, air, and waste water treatment plant (WWTP) water. Indices 1 to 7 are used, where 7 represents a high potential of exposure and 1 represent a low potential of exposure. Since no physicochemical properties are included in the calculations, the exposure is only valid close to the release source. Table 4 shows the obtained exposure indices together with the average exposure index for each FR for which data is available. Darker cell color indicates a higher potential of exposure. In addition, Table 4 contains indices for trends in exposure for humans and the environment. These indices range between -2 and 2, where a negative value indicates a decreasing time trend in the potential exposure while a positive value indicates an increasing trend.

Pentaerythritol (EI = 6.7), TNBP (6.5), TPHP (6.3), SCCPs (6.2), and TMPP (6.0) all had average EIs ≥ 6 , indicating a high potential of exposure for many environmental compartments. In fact, all compartments (except air for a few FRs) had an EI of 6 or 7 for these compounds. Seven FRs (i.e. CDP (5.7), TBOEP (5.7), TIBP (5.7), DBPhP (5.5), TXP (5.5), TEHP (5.3), DPhBP (5.2), and melamine (5.0)) had average EIs between 5 and 6, also indicating a generally high potential of exposure to these FRs. Both bis-(*t*-butylphenyl) phenyl phosphate and TCIPP had average EIs of 4.8, while TEP, EHDPP, IDP, melamine cyanurate, and BEH-TEBP had EIs of 3.7, 3.3, 3.3, 3.3, and 3.0, respectively. The remaining FRs all had average EIs < 3 with generally higher EIs for WWTP and occupational than for the other categories.

Regarding time trends in exposure, two FRs were indexed with a 2 (TBBPA-BDBPE in humans and environment, and TTNPP in humans), indicating a strong increasing trend in the potential of exposure. Furthermore, in humans, an increasing potential (index 1) can be observed for DBDPE, PBDPP, and *o*-TMPP. In the environment, an increasing potential of exposure (1) can be associated with TMPP, CDP, DBDPE, PBDPP, and TTNPP. One FR, SCCP, showed a decreasing potential of exposure (-1) for the environment.

Table 4 Exposure Indices (EI) of FRs in Sweden [12]. Indices were calculated based on registered use patterns in year 2015 from the Swedish Product Register, considering only diffuse emissions, mainly during end product use. The FRs are sorted based on the average EI, from high to low.

| Abbreviation/ name | EI Surface water | EI Air | EI Soil | EI WWTP | EI Consumer | EI Occupational | Average EI | EI Human trend | EI Environmental trend |
|---|------------------------|-----------|------------|------------|----------------|--------------------|---------------|----------------------|------------------------------|
| Pentaerythritol | 7 | 6 | 7 | 6 | 7 | 7 | 6.7 | 0 | 0 |
| TNBP | 6 | 6 | 7 | 6 | 7 | 7 | 6.5 | 0 | 0 |
| TPHP | 6 | 5 | 7 | 6 | 7 | 7 | 6.3 | 0 | 0 |
| SCCP | 6 | 3 | 7 | 7 | 7 | 7 | 6.2 | 0 | -1 |
| TMP (o-, m-, p-) | 7 | 2 | 7 | 6 | 7 | 7 | 6.0 | 0 | 1 |
| TMPP (p-) | 7 | 2 | 7 | 6 | 7 | 7 | 6.0 | 0 | 1 |
| CDP | 6 | 3 | 6 | 6 | 6 | 7 | 5.7 | 0 | 1 |
| TBOEP | 5 | 3 | 7 | 6 | 6 | 7 | 5.7 | 0 | 0 |
| TIBP | 6 | 2 | 7 | 6 | 6 | 7 | 5.7 | 0 | 0 |
| DBPhP | 5 | 2 | 7 | 5 | 7 | 7 | 5.5 | 0 | 0 |
| TXP | 5 | 2 | 7 | 5 | 7 | 7 | 5.5 | 0 | 0 |
| TEHP | 5 | 5 | 5 | 5 | 5 | 7 | 5.3 | 0 | 0 |
| DPhBP | 5 | 1 | 7 | 5 | 6 | 7 | 5.2 | 0 | 0 |
| Melamine | 3 | 3 | 4 | 6 | 7 | 7 | 5.0 | 0 | 0 |
| Bis-(t-butylphenyl) phenyl phosphate | 6 | 1 | 6 | 3 | 6 | 7 | 4.8 | 0 | 0 |
| TCIPP | 4 | 3 | 4 | 6 | 5 | 7 | 4.8 | 0 | 0 |
| TEP | 2 | 2 | 2 | 4 | 5 | 7 | 3.7 | 0 | 0 |
| EHDPP | 1 | 1 | 3 | 5 | 3 | 7 | 3.3 | 0 | 0 |
| IDP | 1 | 1 | 1 | 4 | 6 | 7 | 3.3 | 0 | 0 |
| Melamine cyanurate | 2 | 1 | 2 | 6 | 2 | 7 | 3.3 | 0 | 0 |
| BEH-TEBP | 2 | 2 | 2 | 1 | 4 | 7 | 3.0 | 0 | 0 |
| BADP | 1 | 1 | 1 | 5 | 1 | 7 | 2.7 | 0 | 0 |
| DBDPE | 1 | 1 | 1 | 5 | 1 | 7 | 2.7 | 1 | 1 |
| TCEP | 2 | 1 | 1 | 4 | 1 | 7 | 2.7 | 0 | 0 |
| PBDPP | 1 | 1 | 1 | 4 | 1 | 7 | 2.5 | 1 | 1 |
| TBBPA-BDBPE | 1 | 1 | 1 | 4 | 1 | 7 | 2.5 | 2 | 2 |
| BTBPE | 1 | 1 | 1 | 3 | 1 | 7 | 2.3 | 0 | 0 |
| EBTEBPI | 1 | 1 | 1 | 3 | 1 | 7 | 2.3 | 0 | 0 |
| HEEHP-TEBP | 1 | 1 | 1 | 3 | 1 | 7 | 2.3 | 0 | 0 |
| TBPP | 1 | 1 | 4 | 1 | 2 | 4 | 2.2 | 0 | 0 |
| TTBNPP | 1 | 1 | 1 | 2 | 1 | 7 | 2.2 | 2 | 1 |
| Melamine pyrophosphate | 1 | 1 | 1 | 2 | 1 | 6 | 2.0 | 0 | 0 |
| TMPP (o-) | 1 | 1 | 1 | 1 | 1 | 6 | 1.8 | 1 | 1 |
| TBBPA | 1 | 1 | 1 | 1 | 1 | 5 | 1.7 | 0 | 0 |
| Ammeline | 1 | 1 | 1 | 1 | 1 | 1 | 1.0 | 0 | 0 |
| DBNPG | 1 | 1 | 1 | 1 | 1 | 1 | 1.0 | 0 | 0 |
| DDC-CO | 1 | 1 | 1 | 1 | 1 | 1 | 1.0 | 0 | 0 |
| T2CPP | 1 | 1 | 1 | 1 | 1 | 1 | 1.0 | 0 | 0 |
| 4'-PeBPO-BDE208 | 0 | 0 | 0 | 0 | 0 | 0 | 0.0 | 0 | 0 |
| Acetoguanamine | 0 | 0 | 0 | 0 | 0 | 0 | 0.0 | 0 | 0 |
| Benzoguanamine | 0 | 0 | 0 | 0 | 0 | 0 | 0.0 | 0 | 0 |

| | | | | | | | | | |
|-----------|---|---|---|---|---|---|-----|---|---|
| CLP1 | 0 | 0 | 0 | 0 | 0 | 0 | 0.0 | 0 | 0 |
| MCCP | 0 | 0 | 0 | 0 | 0 | 0 | 0.0 | 0 | 0 |
| DMP | 0 | 0 | 0 | 0 | 0 | 0 | 0.0 | 0 | 0 |
| HBCDD | 0 | 0 | 0 | 0 | 0 | 0 | 0.0 | 0 | 0 |
| TDCIPP | 0 | 0 | 0 | 0 | 0 | 0 | 0.0 | 0 | 0 |
| TMP | 0 | 0 | 0 | 0 | 0 | 0 | 0.0 | 0 | 0 |
| BCMP-BCEP | 0 | 0 | 0 | 0 | 0 | 0 | 0.0 | 0 | 0 |

5. Prioritization lists in the open literature

Several prioritization lists for FRs have been published [58, 68-72]. Harju et al. (2009) recommended a number of BFRs to be included in future monitoring in the Norwegian environment [58]. The list is based upon production volumes, product usage, potential of long range atmospheric transport (LRAT), potential of bioaccumulation, persistence, levels in the environment and environmental transport processes. In total, 14 BFRs were recommended for future studies, including TBP, TBP-AE, BEH-TEBP, BTBPE, DBDPE, EBTEBPI, EH-TBB, HBB, PBEB, PBT, TBBPA-BAE, TBBPA-BDBPE, TBP-DBPE, and TEBP-Anh. Furthermore, a step-by-step approach for prioritizing chemicals to include in long-term air monitoring programs have been suggested and applied to 138 organic chemicals, including several FRs [68]. This prioritization generated a list of 15 high-priority substances, out of which three are FRs, specifically short-chain paraffins, HBCDD, and PBP. In another study, Howard and Muir (2010) identified 610 organic chemicals with persistent and bioaccumulative properties that had not been included in measurement programs before. They determined top ten priority compounds within brominated, chlorinated, fluorinated, siloxanes and other chemical classes. Among the top ten priority chemicals in the different classes, TBBPA, PBCH, HBCDD, TBECH, BEH-TEBP, EBTEBPI, BTBPE, DBDPE, OBTMPI, Dibutyl chlorendate, and DDC-CO were included [69]. Covaci et al. (2011) have reported TBBPA-BDBPE, HCDBCO, TBP, DBDPE, BTBPE, HBB, EH-TBB, BEH-TEBP, PBEB, DBE-DBCH, TBP-AE, PBT, TTBP-TAZ, EBTEBPI, and TBBPA-BAE as the most important alternative BFRs [70].

The European Food Safety Authority (EFSA, 2012) have evaluated 27 emerging FRs in terms of potential persistence in the environment, potential of bioaccumulation and toxicity [71]. A modeling exercise was conducted based on chemical properties. Fifteen (15) out of the 27 compounds was determined to have a high overall persistence, while a high potential of bioaccumulation was determined for 14 compounds. Both high potential for bioaccumulation and a high overall persistence was found for 10 FRs, including BTBPE, DBHCTD, EBTEBPI, HBB, HBCYD, HCTBPH, PBB-Acr, PBEB, PBT, and TBX. As a result, these FRs were recommended for further studies. In fact, experimental data have shown that HBB and BTBPE have high bioaccumulation factors in several aquatic species. Furthermore, experimental studies have shown that both DBNPG and TDBPP are genotoxic and carcinogenic [71], which can motivate inclusion in future studies. In another modeling study, 37 BFRs and 24 OPFRs were evaluated in terms of persistence and ability of LRT by comparing physicochemical properties to previously known persistent organic pollutants [72]. Out of the tested BFRs, eight or 13 (depending on the chosen parameters) were predicted to be of low concern, typically having low molecular weight, low degree of bromination, and containing OH-Groups, while the remaining compounds (i.e. TBBPA, HBCDD, DBDPE, BTBPE, BEH-TEBP, EH-TBB, TBBPA-BDBPE, PBT, PBEB, 2,4-DBP, TBX, PBP, pentabromobenzyl chloride (PBBC), PBPAE, TBP-DBPE,

HBB, PBBBr, TBBBS (Tetrabromobisphenol-S), TBBPA-DHEE, Tetrabromobisphenol A bismethyl ether (TBBPA-BME), TBBPA-BAE, DBDPE, TDBP-TAZTO, PBB-Acr, DBHCTD, HCTBPH, EBTEBPI, HBCYD, and TTBP-TAZ) were suggested to be as persistent as the legacy PBDEs. Three of those (PBT, HBB, and TBX) may also travel extremely long distances in the atmosphere and were therefore suggested to be of extra concern. For OPFRs, the applied model suggested potentially lower hazard for the environment in terms of persistence and LRT compared to for BFRs and PBDEs. All OPFRs except PBDMPP, PBDPP, BADP, TIPPP, and TTBNPP are likely compounds of lower concern. However, authors stress that the atmospheric half-lives of particle-bound FRs can be much higher than for FRs in the gas-phase, which may increase their persistence and LRT compared to the model predictions. Furthermore, neither toxicity nor potential of bioaccumulation were considered in the model, which might be important to consider. TCEP, TCIPP, TPHP, CDP, TMPP, DBNPG, DBS, and TBECH have toxic or bioaccumulative properties making them less suitable as replacement chemicals for the legacy FRs.

6. Discussion on prioritization among the identified flame retardants

The following discussion on prioritization among the identified flame retardants is based on usage, exposure time trend, environmental detection, and previous prioritization lists. It would have also been relevant to include environmental risk, e.g. as the ratio between measured environmental concentrations (MECs) and predicted no effect concentrations (PNEC). However, it was out of the scope of the current study to also include toxicity assessments in the ranking among FRs. Thus, the toxicity aspect has not been given priority. Neither in the previously published prioritization lists [58, 68-72] was toxicity considered, explained by e.g., the need of vast toxicity data for the estimation of toxicity thresholds which is often lacking for newer substances [68].

6.1 Usage and time trends in exposure

Combining the results from the databases of the EU and Sweden, in total 45 different chemicals have been identified as being used as FRs. Pentaerythritol and melamine appear to be used in the highest quantities with indices of 0.92 and 0.83, respectively (calculated as summed indices of Swedish and EU usage, normalized against the highest possible score (12)). However, it is likely that these chemicals also have other usages than as FRs. Other highly used FRs include SCCPs (0.75), DBDPE (0.75), TEP (0.75), melamine cyanurate (0.75), TBBPA-BDBPE (0.67), TPHP (0.67), EHDPP (0.67), PBDPP (0.67), TIBP (0.67), TEHP (0.67), BADP (0.67), IDP (0.58), BEH-TEBP (0.50), EBTEBPI (0.50), HEEHP-TEBP (0.50), and TXP (0.50). Interestingly, there are twelve FRs that are reported to be used in Sweden, but for which there is no information available in the ECHA database. Similarly, two FRs (TCEP and TCIPP) are among the most extensively used FRs in Sweden, but according to the EU database, they have not been registered within REACH. This mismatch may be explained by differences in the reporting limits of the two databases where the ECHA database has a higher threshold for when a chemical needs to be registered. If prioritization of FRs for environmental monitoring was entirely based on use data, the list could include up to 45 FRs depending on the quantity threshold.

Increasing time trends in exposure indices were observed for TMPP (environment, human), CDP (environment), DBDPE (environment, human), PBDPP (environment, human), TBBPA-BDBPE (environment and human), and TTNPP (environment, human). A decreasing trend in the potential of exposure was observed for SCCPs (environment). Based on the increasing trend in exposure to some FRs, it can be argued that these FRs should be included in future studies. Depending on the type of samples (i.e. water, soil, air etc.) collected, some FRs may be more important to include than other.

6.2 Environmental detection

In indoor air, 19 different FRs have been detected at concentrations higher than 1 ng m^{-3} , including DBDPE, DBE-DBCH, HBB, PBEB, PBT, TBX, DOPP, EHDPP, TBOEP, TMPP, TEP, TEHP, TIBP, TNBP, TPHP, TPP, TCEP, TCIPP, and TDCIPP (Appendix, Table S3). Detection frequencies up to 100% have been

reported in at least one study for DBE-DBCH, TBOEP, TEP, TIBP, TNBP, and TPHP [18, 19, 22]. Other FRs (HBB, PBB, PBT, TBP-AE, TBP-DBPE, and TMPP) have been detected in Sweden and Norway at lower concentrations but still with high detection frequencies ($\geq 50\%$) [18]. Other detected FRs in indoor air, but with lower detection frequencies, are BEH-TEBP, BTBPE, CLP1, syn-DDC-CO, anti-DDC-CO, and EH-TBB [18, 20, 21]. In dust, detection frequencies $>50\%$ (often $>90\%$) have been reported in at least one study for BEH-TEBP, BTBPE, DBDPE, DBE-DBCH, syn-DDC-CO, anti-DDC-CO, EH-TBB, HBB, PBB, PBT, TBP-DBPE, EHDPP, TBOEP, TCEP, TCIPP, TMPP, TDCIPP, TEP, TiBP, TNBP, and TPHP [18, 19, 22, 26, 27, 29] (Appendix, Table S4). Other FRs detected in dust include CLP1, DOPP, PBEB, TBX, THP, and TEEedP [18-21, 25, 32]. In general, concentrations in dust of OPFRs were higher than for BFRs, which may indicate a more frequent use of OPFRs than of BFRs or possibly a higher ability of OPFRs of leaking from the materials. As a result, in a prioritization process, it could be argued that including the OPFRs in future indoor screenings is more important than the BFRs; however, some BFRs such as e.g., DBE-DBCH, DBDPE, and HBB may also be important to include.

Many of the FRs that have been detected in indoor environments have also been detected in outdoor air, sometimes even in remote regions (Appendix, Table S5). TPHP, TCEP, TCIPP, TDCIPP, TMP (tentatively identified), TNBP, EHDPP, TBOEP, TIBP, TBP, BTBPE, DBE-DBCH, anti-DDC-CO, syn-DDC-CO, EHTBB, HBB, PBEB, PBT, BEH-TEBP, and DBDPE have all been detected in remote areas within the Nordic countries [22, 25, 34, 38], showing that they all are likely to undergo LRT. With respect to this, these chemicals may be important to include in future screenings. It is, however, important to remember that bromophenols such as e.g., TBP are also naturally produced. Other FRs detected in outdoor air of urbanized areas in the Nordic countries include BCMP-BCEP, TBBPA, 2,4-DBP, 2/3-BP, 2,6-DBP, 4-BP, TBP-AE, BATE, PBP, and TBP-DBPE [20, 22, 37, 38]. In China, also TBBPA-BDBPE has been detected [33].

In Sweden and Finland, DBE-DBCH, DDC-CO, TCEP, TCIPP, TMP (tentatively identified), and TNBP have been detected in atmospheric deposition in remote areas [34, 39, 40] (Appendix, Table S6). Again, this is showing LRT potential. A number of OPFRs (TBOEP, TCEP, TCIPP, *p*-TMPP, TDCIPP, TEHP, TMP (tentatively quantified), TNBP, and TPHP) have been detected in snow close to a road and an airport, indicating these as being emitted from point sources [34]. TBOEP, TCEP, TCIPP, TIBP, and TNBP have also been detected in urban rain in Germany [39]. Possibly, all FRs mentioned in this section may be important to include in future screenings, and especially the ones found in remote areas.

Five OPFRs have been detected in rural/remote surface waters, including TBOEP, TCEP, TCIPP, TIBP, and TNBP [44], which makes them important to include in future screenings (Appendix, Table S7). In general, concentrations of OPFRs appear to be higher than for BFRs in European rivers. Detected OPFRs in European rivers include EHDPP, TBOEP, TCEP, TDCIPP, TIBP and TCIPP [41, 43, 46, 51, 52]. In the screening study of FRs in Swedish rivers, TEBP-Anh, TBBPA, TCBPA, and EHTBB were detected at the

highest concentrations, but also TBP, BTBPE, syn-DDC-CO, EHDPP, HBB, PBB-Acr, PBT, TBX, TCEP, TCIPP, *o*-TMPP, *p*-TMPP, ΣTDCIPP+TEHP, TNBP, TPHP, and TTBNPP [49]. Even though concentrations were relatively low (<67 ng L⁻¹), it could be argued that these FRs are important to include in future studies since they are in fact polluting the Swedish environment. Finally, also TEP and TPP have been detected in European rivers [41, 43]. In addition to many of the FRs detected in surface waters, several other FRs have been detected in European WWTP, including DOPP, TEHP, TMP, DBDPE, PBEB, and TBBPA-BDBPE [23, 35, 42, 47]. As these are likely to also reach surface waters, including them in future screenings may be important.

The OPFRs EHDPP, TBOEP, TCEP, TCIPP, TDCIPP, TIBP, TNBP, and TPHP have all been detected in sewage sludge from Sweden [42], while the BFRs TBP, 2,4-DBP, TBP-AE, BATE, BEH-TEBP, BTBPE, DBDPE, DBE-DBCH, anti-DDC-CO, syn-DDC-CO, EH-TBB, HBB, PBEB, PBP, PBT, TBBPA, and TBP-DBPE have been detected in sewage sludge from the Nordic countries [20] (Appendix, Table S9). Other detected FRs in sewage sludge include DDC-DBF and DDC-Ant in Spain [59], TCBPA in Canada [60], and TBBPA-BDBPE in China [33]. Also in sediment, many FRs have been detected (Appendix, Table S8). In the Nordic countries, detected FRs in sediment include TBP, 2,4-DBP, BEH-TEBP, DBE-DBCH, anti-DDC-CO, syn-DDC-CO, EH-TBB, HBB, PBEB, PBT, TBBPA, and TBBPA-BAE [20, 23, 61]. Within Europe, the OPFRs EHDPP, TBOEP, TCEP, *o*-TMPP, *p*-TMPP, TDCIPP, TEHP, TEP, TIBP, TNBP, TPHP and BCMP-BCEP have been detected in sediment, some with high detection frequencies [22, 43, 51, 56]. In North America, China and Norway, BTBPE has been detected, while TBBPA-BDBPE has, to our knowledge, only been detected in China [20, 23, 33, 47, 54, 57]. Based on this, it could be argued that all FRs that have been detected in sludge and/or sediment in the Nordic countries should be included in future studies. Perhaps is also BTBPE a compound of interest, as it is frequently detected in other environmental compartments. Furthermore, FRs such as DDC-DBF, DDC-Ant, TCBPA, and TBBPA-BDBPE may also be interesting to include, even though they have not (yet) been detected in Nordic sediment or sludge.

Not many alternative FRs have been analysed or detected in soil (Appendix, Table S10). However, BTBPE, DBDPE, and TBBPA-BDBPE have been detected in soil in China and BEH-TEBP in Norway [33, 61]. In plants, DBDPE, HBB, TBBPA-BDBPE, TBP, 2,4-DBP, BEH-TEBP, BTBPE, DBDPE, DBE-DBCH, anti-DDC-CO, syn-DDC-CO, HBB, PBEB, PBT, and TBP-DBPE have been detected in mosses/tree needles close to point sources in the Nordic countries [20, 23, 35] (Appendix, Table S11).

A wide variety of different animal species have been analysed for emerging FRs, including e.g., arctic fox, polar bear, moose, seal, cod, salmon, mussels, crab, common eider and herring gull (Appendix, Tables S12 and S13). Several BFRs, including BEH-TEBP, DBDPE, CCPs, anti- and syn-DDC-CO, EH-TBB, HBB, PBEB, and TBP have been detected in Arctic species such as e.g., polar bear [61-63], indicating potential of LRT and possibly also bioaccumulation. Thus, these FRs may be important to include in upcoming studies.

However, as previously mentioned TBP is also a naturally occurring compound. In this literature review, no studies concerning OPFRs in mammals (except in humans) were found. The number of detected FRs in aquatic animals (37) is higher compared to detected FRs in mammals (8). BEH-TEBP, S/MCCPs, DBDPE, and EHDPP have all been detected in fish from the Arctic [53, 61, 63], indicating potential of LRT and possibly also bioaccumulation. These may therefore be important to include in future studies. In the Nordic countries, the following FRs have been detected in fishes and mussels etc.: TBP, BEH-TEBP, BTBPE, anti-DDC-CO, syn-DDC-CO, DBDPE, DBPhP, EHDPP, TBOEP, TCEP, TCIPP, o-TMPP, p-TMPP, TDCIPP, TNBP, TPHP, 2,4-DBP, BATE, DBE-DBCH, HBB, PBEB, PBT, T2CPP, TBP-DBPE, TEHP, and TIBP [20, 22, 23, 25, 35, 56, 61, 64]. Also these FRs may be important to include in future studies, as they are polluting the Nordic environment with possible bioaccumulation. In Canada, the chlorinated FRs Chlordene plus, DDC-DBF, Dec 604B, anti-, and syn-DDC-CO have been detected in fish [65]. To our knowledge, chlordene Plus, DDC-DBF and Dec-604 component B have never been analysed in any of the Nordic countries before (except in the Norwegian Arctic), which could motivate to include also them in future studies. In birds and/or bird eggs, a total of 29 FRs have been detected (Appendix, Table S14). FRs that has been detected in Arctic birds/eggs include TBP, BEH-TEBP, DBDPE, Chlordene Plus, S/MCCPs, DBPhP, EHDPP, T2CPP, TCEP, TEHP, TIBP, TNBP, and TPHP [53, 61, 63]. Other FRs that have been detected in birds/eggs from other areas include BTBPE, DBE-DBCH, EHDPP, HBB, PBEB, PBP, PBT, anti-DDC-CO and syn-DDC-CO, TBOEP, TCIPP, and TDCIPP [20, 33, 56, 57, 61]. In general, detection frequencies of OPFRs were lower than for BFRs and CFRs.

Finally, eight BFRs have been detected in human blood serum from Sweden, including 2/3-BP, TBP, 2,4-DBP, 2,6-DBP, 4-BP, BTBPE, EH-TBB, and PBEB [25, 38] (Appendix, Table S15). Both TBP and BTBPE were detected in 100% of the analysed samples, showing the wide-spread contamination of these FRs. In human breast milk, also from Sweden, EHDPP, TBOEP, TCEP, TCIPP, *p*-TMPP, TDCIPP, TNBP, and TPHP were detected [64].

6.3 Prioritization lists

In total, 51 different FRs have been suggested as priority substances in literature. These include TBP, TBP-AE, BEH-TEBP, BTBPE, DBDPE, EBTEBPI, EH-TBB, HBB, PBEB, PBT, TBBPA-BAE, TBBPA-BDBPE, TBP-DBPE, TEBP-Anh, S/MCCPs, PBP, TBBPA, PBCH, OBTMPI, Dibutyl chlorendate, DDC-CO, DBE-DBCH, TTBP-TAZ, DBHCTD, HBB, HBCYD, HCTBPH, PBB-Acr, TBX, DBNPG, TDBPP, 2,4-DBP, PBBC, PBPAE, PBBBr, TBBBS, TBBPA-DHEE, TBBPA-BME, TDBP-TAZTO, HBCYD, PBDMPP, PBDPP, BADP, TiPPP, TTBNPP, TCEP, TCIPP, TPHP, CDP, TMPP, and DBS. As all FRs addressed in published prioritization lists have for different reasons, such as e.g., potential of bioaccumulation, persistence and toxicity, been thought of as posing a special hazard to the environment and humans, all of them may be important to include in future screenings.

6.4 Summary

In total, approximately 120 alternative FRs have been identified within this literature review through the search in patents, databases of chemical use, literature from previous studies and available prioritization lists. Additionally, five FRs (bBDBP, mDEP/dDEP, T3CPP, TBBPS-DBPE, and TPeP) are being sold as FR reference standards by two chemical suppliers (Accustandard and BOC sciences), despite no other information about these compounds seems to be available.

7. Multicriteria model for prioritization among flame retardants

A simple (but novel) multicriteria model was created to aid in prioritization of FRs to be included in future screening studies. The model takes four different types of data into account, namely (i) usage, (ii) time trend data of exposure for humans and the environment, (iii) environmental detection (for those that have been screened for in the environment), and (iv) previously published prioritization lists. Each FR was ranked individually between 0 (low importance) and 1 (high importance) in each data category, resulting in a maximum possible score of 3 or 4 (3 for those that were never screen for). It should be noted that the model is not compartment specific. If a specific environmental media should be screened (e.g., surface water), the model needs to be refined to also include exposure media specific indicies. Toxicity was not included since such toxicity data often is lacking, especially for newer compounds.

For the category of usage, Swedish usage data was used if available [12], otherwise the EU usage data was used [11]. If no usage data was available a score of 0 was given indicating little or no use. The Swedish usage data was indexed from 0 to 7, and the data was normalized to 0 to 1 by dividing each index by 7, which was the highest possible score. The EU usage data was transformed into indices by assigning the FRs with the highest usage amount (i.e. 100 000 to 1 000 000 tonnes per year) with an index of 5, FRs with the second highest usage amount (i.e. 10 000 to 100 000 tonnes per year) with an index of 4 and so on. If no data was available, the index was put to 0. Then the the data were normalized to 0 to 1 by dividing each index by 5 which was the highest possible score.

For environmental detection, 11 different environmental compartments (including indoor air, indoor dust, outdoor air, atmospheric deposition, water, sewage sludge, sediment, soil, plants, animals and humans) were reviewed within this literature study (see Section 3). To create a score based on environmental detection of FRs, one point was given per detected compartment, followed by normalization against the highest possible score, which was 11. As an example, if a FR has been detected in 5 out of the 11 environmental compartments, this resulted in a score of $5/11 = 0.45$. As rarely monitored FRs (≤ 1 study) would be discriminated in this category, these FRs ($n = 40$) were considered to have a maximum score of 3, and they were consequently normalized to 3. The rarely monitored FRs include TTBNPP, EBTEBPI, PBDPP, CDP, BADP, bBDBP, pentaerythritol, IDP, melamine, melamine cyanurate, melamine pyrophosphate, DPhBP, HEEHP-TEBP, TXP, DBNPG, TBPP, PBDMPP, HCTBPH, HBCYD, mDEP/dDEP, TBBPS-DBPE, TPeP, TTBP-TAZ, piperazine polyphosphate, piperazine pyrophosphate, DBS, dibutyl chlorendate, PBBBr, PBBC, PBCH, PBPAE, T3CPP, TBBPA-BME, TBBPA-DHEE, TBBPS, TDBPP, TDBP-TAZTO, TiPPP, and cyanurodiamide.

For the prioritization lists, six prioritization lists were identified [58, 68-72] within this literature review. For each list where a FR was included, the FR was given one point. By dividing the obtained score with the highest possible score in this category (6), scores between 0 and 1 were obtained.

The last category, time trends in exposure calculated from the Swedish product database, was already indexed with values between -2 and 2 [12]. To normalize this category, the assigned indices in the two exposure categories (human and ecosystem) were summed and divided by the highest possible score. As each exposure category could have an index up to 2, the highest possible score was 4. The actual score for each FR was thus normalized by dividing by 4.

Finally, the score from each data category were summed together into one final total score for each FR. This total score was converted into total scores (0 to 100) to obtain an easily understandable and interpretable value (Table 5). The calculation of the total scores was performed by dividing the total score by the highest possible score of 3 or 4 and multiplying with 100.

Based on the score, ranging between 0 and 100, the FRs can be ranked and prioritized for future studies (Table 5). The highest total score was obtained for TBBPA-BDBPE (69), DBDPE (65), BTBPE (58), TTBNPP (48) BEH-TEBP (47), EBTEBPI (47), PBDPP (46) *p*-TMPP (44), TPHP (42), and TCIPP (40). When interpreting the results, it should be kept in mind that there are some limitations of this simplistic model. For example, the levels of the FRs are not considered, only the occurrence. As a result, the score in the category of environmental detection is highly dependent on the detection limits in each specific study. Furthermore, although ~60 peer-reviewed papers were included for the category of environmental detection, the literature review may not be completely comprehensive. Moreover, the total scores is influenced by whether there is data available or not (e.g., for usage and time trend); if no data are available (or if EC/CAS numbers are lacking and no database search is possible), the score will be low. Another limitation is that the model is not compartment specific. If a specific environmental media should be screened (e.g., surface water), the model needs to be refined to also include exposure media specific indices. FRs that are used but only have been targeted in maximum one study are marked with an asterisk (*) in Table 5. These FRs may be interesting to include in future studies despite a comparably low ranking. Nevertheless, the developed multicriteria model provide a simple and effective way of prioritizing FRs, and can readily be used to obtain priorities of FRs that are important to include in future monitoring studies.

Table 5 Multicriteria model results for prioritization of alternative FRs.

| | Usage | Environmental detection | Prioritization lists | Trends | Σ score | Total score |
|--------------------------------|-------|-------------------------|----------------------|--------|----------------|-----------------------|
| TBBPA-BDBPE | 0.71 | 0.55 | 0.50 | 1.00 | 2.76 | 69^a |
| DBDPE | 0.71 | 0.73 | 0.67 | 0.50 | 2.61 | 65^a |
| BTBPE | 0.57 | 0.91 | 0.83 | 0 | 2.31 | 58^a |
| TTBNPP* | 0.43 | 0.09 | 0.17 | 0.75 | 1.44 | 48^b |
| BEH-TEBP | 0.57 | 0.64 | 0.67 | 0 | 1.87 | 47^a |
| EBTEBPI* | 0.57 | 0 | 0.83 | 0 | 1.40 | 47^b |
| PBDPP* | 0.71 | 0 | 0.17 | 0.50 | 1.38 | 46^b |
| TMPP (<i>p</i>-) | 0.71 | 0.64 | 0.17 | 0.25 | 1.77 | 44^a |
| TPHP | 0.71 | 0.82 | 0.17 | 0 | 1.70 | 42^a |
| TCIPP | 0.71 | 0.73 | 0.17 | 0 | 1.61 | 40^a |
| TCEP | 0.57 | 0.82 | 0.17 | 0 | 1.56 | 39^a |
| TMPP (<i>o</i>-) | 0.43 | 0.45 | 0.17 | 0.50 | 1.55 | 39^a |
| TBOEP | 0.71 | 0.82 | 0 | 0 | 1.53 | 38^a |
| TNBP | 0.71 | 0.82 | 0 | 0 | 1.53 | 38^a |
| CDP* | 0.71 | 0 | 0.17 | 0.25 | 1.13 | 38^b |
| EHDPP | 0.71 | 0.73 | 0 | 0 | 1.44 | 36^a |
| HBB | na | 0.73 | 0.67 | 0 | 1.40 | 35^a |
| PBEB | na | 0.73 | 0.67 | 0 | 1.40 | 35^a |
| PBT | na | 0.73 | 0.67 | 0 | 1.40 | 35^a |
| TIBP | 0.71 | 0.64 | 0 | 0 | 1.35 | 34^a |
| DBE-DBCH | na | 0.73 | 0.50 | 0 | 1.23 | 31^a |
| BADP* | 0.71 | 0 | 0.17 | 0 | 0.88 | 29^b |
| TEHP | 0.71 | 0.45 | 0 | 0 | 1.17 | 29^a |
| Pentaerythritol* | 0.86 | 0 | 0 | 0 | 0.86 | 29^b |
| EH-TBB | na | 0.64 | 0.50 | 0 | 1.14 | 28^a |
| DDC-CO | 0.14 | 0.82 | 0.17 | 0 | 1.13 | 28^a |
| TEP | 0.71 | 0.36 | 0 | 0 | 1.08 | 27^a |
| MCCP | 0.80 | 0.09 | 0.17 | 0 | 1.06 | 26^a |
| SCCP | 0.80 | 0.09 | 0.17 | 0 | 1.06 | 26^a |
| TBP | 0.00 | 0.64 | 0.33 | 0 | 0.97 | 24^a |
| IDP* | 0.71 | 0 | 0 | 0 | 0.71 | 24^b |
| Melamine* | 0.71 | 0 | 0 | 0 | 0.71 | 24^b |
| Melamine cyanurate* | 0.71 | 0 | 0 | 0 | 0.71 | 24^b |
| PBB-Acr | 0.40 | 0.18 | 0.33 | 0 | 0.92 | 23^a |
| TBP-DBPE | na | 0.55 | 0.33 | 0 | 0.88 | 22^a |
| TBBPA | 0.14 | 0.36 | 0.33 | 0 | 0.84 | 21^a |
| TDCIPP | 0 | 0.82 | 0 | 0 | 0.82 | 20^a |
| DPhBP* | 0.57 | 0 | 0 | 0 | 0.57 | 19^b |
| HEEHP-TEBP* | 0.57 | 0 | 0 | 0 | 0.57 | 19^b |
| Melamine pyrophosphate* | 0.57 | 0 | 0 | 0 | 0.57 | 19^b |
| TXP* | 0.57 | 0 | 0 | 0 | 0.57 | 19^b |
| 2,4-DBP | na | 0.55 | 0.17 | 0 | 0.71 | 18^a |
| DBPhP | 0.57 | 0.09 | 0 | 0 | 0.66 | 17^a |

| | | | | | | |
|--|------|------|------|---|------|-----------------------|
| T2CPP | 0.57 | 0.09 | 0 | 0 | 0.66 | 17^a |
| DBNPG* | 0.14 | 0 | 0.33 | 0 | 0.48 | 16^b |
| ATE | na | 0.27 | 0.33 | 0 | 0.61 | 15^a |
| PBP | na | 0.27 | 0.33 | 0 | 0.61 | 15^a |
| TBX | na | 0.27 | 0.33 | 0 | 0.61 | 15^a |
| TBBPA-BAE | na | 0.09 | 0.50 | 0 | 0.59 | 15^a |
| bis-(t-butylphenyl) phenyl phosphate* | 0.43 | 0 | 0 | 0 | 0.43 | 14^a |
| TBPP* | 0.43 | 0 | 0 | 0 | 0.43 | 14^b |
| DBHCTD | na | 0 | 0.50 | 0 | 0.50 | 13^a |
| PBDMPP* | 0.2 | 0 | 0.17 | 0 | 0.37 | 12^b |
| TEBP-Anh | 0.2 | 0.09 | 0.17 | 0 | 0.46 | 11^a |
| Dec-604A/ HCTBPH | na | 0 | 0.33 | 0 | 0.33 | 11^b |
| HBCYD | na | 0 | 0.33 | 0 | 0.33 | 11^b |
| TTBP-TAZ | na | 0 | 0.33 | 0 | 0.33 | 11^b |
| TMPP (<i>m</i>-) | 0 | 0.27 | 0.17 | 0 | 0.44 | 11^a |
| BATE | na | 0.27 | 0 | 0 | 0.27 | 7^a |
| DOPP | na | 0.27 | 0 | 0 | 0.27 | 7^a |
| TMP | 0 | 0.27 | 0 | 0 | 0.27 | 7^a |
| TPP | na | 0.27 | 0 | 0 | 0.27 | 7^a |
| Piperazine (poly)phosphate* | 0.20 | 0.00 | 0 | 0 | 0.20 | 7^b |
| DBS | na | 0 | 0.17 | 0 | 0.17 | 6^b |
| Dibutyl chlorendate | na | 0 | 0.17 | 0 | 0.17 | 6^b |
| PBBBr | na | 0 | 0.17 | 0 | 0.17 | 6^b |
| PBBC | na | 0 | 0.17 | 0 | 0.17 | 6^b |
| PBCH | na | 0 | 0.17 | 0 | 0.17 | 6^b |
| PBPAE | na | 0 | 0.17 | 0 | 0.17 | 6^b |
| TBBPA-BME | na | 0 | 0.17 | 0 | 0.17 | 6^b |
| TBBPA-DHEE | na | 0 | 0.17 | 0 | 0.17 | 6^b |
| TBBPS | na | 0 | 0.17 | 0 | 0.17 | 6^b |
| TDBPP | na | 0 | 0.17 | 0 | 0.17 | 6^b |
| TDBP-TAZTO | na | 0 | 0.17 | 0 | 0.17 | 6^b |
| TiPPP | na | 0 | 0.17 | 0 | 0.17 | 6^b |
| Ammeline* | 0.14 | 0 | 0 | 0 | 0.14 | 5^b |
| Piperazine pyrophosphate | 0.14 | 0 | 0 | 0 | 0 | 5^b |
| 2,6-DBP | na | 0.18 | 0 | 0 | 0.18 | 5^a |
| 2-BP | na | 0.18 | 0 | 0 | 0.18 | 5^a |
| 3-BP | na | 0.18 | 0 | 0 | 0.18 | 5^a |
| 4-BP | na | 0.18 | 0 | 0 | 0.18 | 5^a |
| CLP1 | na | 0.18 | 0 | 0 | 0.18 | 5^a |
| DDC-DBF | na | 0.18 | 0 | 0 | 0.18 | 5^a |
| PBB | na | 0.18 | 0 | 0 | 0.18 | 5^a |
| TCBPA | na | 0.18 | 0 | 0 | 0.18 | 5^a |
| BCMP-BCEP | 0 | 0.18 | 0 | 0 | 0.18 | 5^a |
| OBTMPI | na | 0 | 0.17 | 0 | 0.17 | 4^a |
| BdPhP | na | 0.09 | 0 | 0 | 0.09 | 2^a |

| | | | | | | |
|---|-----------------|------|---|---|------|----------------------|
| Chlordene Plus | na | 0.09 | 0 | 0 | 0.09 | 2^a |
| Dec 604B | na | 0.09 | 0 | 0 | 0.09 | 2^a |
| DDC-Ant | na | 0.09 | 0 | 0 | 0.09 | 2^a |
| TEEdP | na | 0.09 | 0 | 0 | 0.09 | 2^a |
| THP | na | 0.09 | 0 | 0 | 0.09 | 2^a |
| TiPP | na | 0.09 | 0 | 0 | 0.09 | 2^a |
| 1,3-hexylene dimelamine | na | 0 | 0 | 0 | 0 | 0^b |
| 1,3-phenylene-bis(dixylenyl phosphate) | na | 0 | 0 | 0 | 0 | 0^b |
| 4'-PeBPOBDE208 | 0 | 0 | 0 | 0 | 0 | 0^b |
| Acetoguanamine | 0 | 0 | 0 | 0 | 0 | 0^b |
| bBDBP | na ^c | 0 | 0 | 0 | 0 | 0^b |
| Benzoguanamine* | 0 | 0 | 0 | 0 | 0 | 0^b |
| bis-(isopropylphenyl) phenyl phosphate | na | 0 | 0 | 0 | 0 | 0^b |
| Brominated paraffins | na | 0 | 0 | 0 | 0 | 0^b |
| Butylene diguanamine | na | 0 | 0 | 0 | 0 | 0^b |
| DMP | 0 | 0 | 0 | 0 | 0 | 0^b |
| Ethylene dimelamine | na | 0 | 0 | 0 | 0 | 0^b |
| Hexamethylene dimelamine | na | 0 | 0 | 0 | 0 | 0^b |
| Isopropylphenyl diphenyl phosphate | na | 0 | 0 | 0 | 0 | 0^b |
| mDEP/dDEP | na ^c | 0 | 0 | 0 | 0 | 0^b |
| Melamine (poly)phosphate | na | 0 | 0 | 0 | 0 | 0^b |
| Methylene diguanamine | na | 0 | 0 | 0 | 0 | 0^b |
| Norbornene diguanamine | na | 0 | 0 | 0 | 0 | 0^b |
| Octyl diphenyl phosphate | na | 0 | 0 | 0 | 0 | 0^b |
| Phthalodiguanamine | na | 0 | 0 | 0 | 0 | 0^b |
| T3CPP | na ^c | 0 | 0 | 0 | 0 | 0^b |
| TBBPS-DBPE | na ^c | 0 | 0 | 0 | 0 | 0^b |
| TBCO | na | 0 | 0 | 0 | 0 | 0^b |
| t-butylphenyl diphenyl phosphate | na | 0 | 0 | 0 | 0 | 0^b |
| TDCPP | na | 0 | 0 | 0 | 0 | 0^b |
| Tetramethylene dimelamine | na | 0 | 0 | 0 | 0 | 0^b |
| TPeP | na ^c | 0 | 0 | 0 | 0 | 0^b |
| Trimethylene dimelamine | na | 0 | 0 | 0 | 0 | 0^b |
| Tris-(isopropylphenyl) phosphate | na | 0 | 0 | 0 | 0 | 0^b |
| Xylenyl diphenyl phosphate | na | 0 | 0 | 0 | 0 | 0^b |

*Used but rarely monitored; ^aA highest possible score of 4 was used in the calculation of total scores; ^bA highest possible score of 3 used in the calculation of total scores; ^cNo search was done in the Swedish database as the FR was identified after the database search was finished.

8. Conclusions

In total, ~125 chemical substances used as alternative FRs were studied within this literature review. The majority of these have previously been targeted in environmental monitoring; however, through the search of US patents, some alternative FRs were identified, although only a limited number of patents were investigated. This indicates a high potential of using patents in the search for new chemicals on developing markets. Searching new chemicals through patent documentation studies is, however, not straight-forward, as generally no CAS- or EC-numbers are provided in the patents. Several FRs were reported in the Swedish database as being used in Sweden, although there was no usage information in the European database (ECHA). Furthermore, two FRs (i.e. TCEP and TCIPP) are extensively used in Sweden, but according to the ECHA database, they have not been registered within REACH. This mismatch may be explained by differences in the reporting limits of the two databases where the ECHA database has a higher threshold for when a chemical needs to be registered. Thus, the Swedish database may contain information on chemicals produced in lower amounts which are excluded by the ECHA database. However, the mismatch could also indicate that the two databases may not be completely up-to-date or that they simply contain erroneous information. Another explanation might be that information could be excluded due to confidentiality. Through the application of a multicriteria model, the FRs were ranked for future studies based on their (i) usage, (ii) time trends in exposure, (iii) environmental detection, and (iv) published prioritization lists. Out of the top 20 ranked FRs, eight were BFRs and twelve were OPFRs. This signals the importance of including both BFRs and OPFRs in future screening studies.

Acknowledgement

This study was financed by the Swedish Environmental Protection Agency (Naturvårdsverket), reference number: NV-08295-16.

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Appendix

Table S1 Brief information about FRs identified in a previous internal literature study.^a

| Brominated flame retardants (BFRs) | | | |
|---|-------------------------------|---|----------|
| DBNPG | Dibromoneopentyl glucol | DBNPG is a reactive FR incorporated in rigid polyurethane foams and unsaturated polyester resins for molded products. It is genotoxic and carcinogenic. | [71] |
| TBNPA | Tribromoneopentyl alcohol | TBNPA is considered potentially persistent and may cause aquatic environmental damage. | [71] |
| 4-BP | 4-bromophenol | No information found. | |
| 2,4-DBP | 2,4-dibromophenol | Produced as an industrial BFR but also naturally occurring in the marine environment. | [20] |
| 2,6-DBP | 2,6-dibromophenol | No information found. | |
| 2,4,6-tribromophenol | TBP | TBP is used as a reactive FR, but also as an antifungal agent and a chemical intermediate. It is considered to be a high production volume (HPV) chemical in the EU. TBP is not only anthropogenically produced; it is also produced naturally by for example marine algae. Furthermore, it can also be formed as a biotransformation product of PBDEs and is often found as a by-product in commercial BTBPE products. | [70] |
| PBP | Pentabromophenol | No information found. | |
| HBB | Hexabromobenzene | HBB can be used as an additive FR in products like paper, woods, textiles and electronics. It is not registered as a compound that is produced within EU, but HBB is produced in China. It may undergo debromination under environmental conditions. | [70, 71] |
| PBT | Pentabromotoluene | PBT is listed as a low production volume (LPV) chemical in the EU. It is used in for example plastics such as polystyrene, polyethylene and polypropylene. PBT is thought to be a stable compound that might undergo debromination in the environment. | [70, 71] |
| PBBBr | Pentabromobenzyl-bromide | No information found. | |
| DBS | Dibromostyrene | DBS is not completely defined but is likely to consist of a mixture of isomers. It is used as a FR in styrenic polymers, probably incorporated in the polymeric chain (reactive FR). | [71] |
| TBCO | 1,2,5,6-tetrabromocyclooctane | TBCO consists of two diastereomers which are easily transformed into one another upon heating. TBCO has the potential of being an aquatic hazardous, very persistent and | [70] |

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| | | bioaccumulative substance. | |
| DBE-DBCH | 4-(1,2-dibromo-ethyl)-1,2-dibromocyclohexane | DBE-DBCH consists of four isomers (α , β , δ and γ) of which α and β are the main compounds in technical mixtures. However, at high temperature α - and β -isomers rearrange to δ - and γ -isomers. These high temperatures can be expected during production of the materials. DBE-DBCH is used as an additive FR in polystyrene and polyurethane materials. DBE-DBCH has been predicted to have a high bioconcentration potential, slow biodegradation and high potential of long-range atmospheric transport (LRAT). | [70, 71] |
| TBX | 2,3,5,6-tetrabromo-p-xylene | TBX is considered to be a stable compound in the environment but with possibility to undergo debromination. | [71] |
| PBEB | Pentabromoethylbenzene | PBEB is no longer in use. Earlier, it was incorporated as an additive in polyester materials like circuit boards and textiles. PBEB is considered to be a persistent and toxic compound that is likely to bioaccumulate. | [58, 70, 71] |
| TEBP-Anh | 3,4,5,6-Tetrabromophthalic anhydride | TEBP-Anh is used both as an additive and reactive FR in materials such as unsaturated polyesters, rigid polyurethane foams and paper. It is proposed as a LPV chemical in the EU. It has also been proposed as a compound with similar properties as those undergoing LRAT, which means that it might undergo LRAT. On the other hand, TEBP-Anh is also considered to be easily hydrolyzed and thus not persistent or bioaccumulative to any extent at all. | [70, 71] |
| TBP-AE | 2,4,6-tribromophenyl allyl ether | ATE is used as an additive FR in expandable styrene and polystyrene foam, but can also be used as a reactive FR. TBP-AE is listed as a LPV chemical in the EU but when TBP-DBPE is degraded or biotransformed, TBP-AE can be formed. TBP-AE might be able to undergo LRAT since its structure and partitioning properties are similar to many of those compounds found in the Arctic. | [70] |
| BATE | 2-Bromoallyl 2,4,6-tribromophenyl ether | No information found. | |
| PBPAE | Pentabromophenyl allyl ether | No information found. | |
| TBP-DBPE | 2,3-Dibromopropyl-2,4,6-tribromophenyl ether | TBP-DBPE is not a commercial product anymore. Can degrade into TBP-AE. | [70] |
| HBCYD | Hexabromocyclodecane | No information found. | |
| PBB-Acr | Pentabromobenzylacrylate | PBB-Acr is a reactive FR used in polybutylene terephthalate and polyethylene | [70, 71] |

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| | | terephthalate. It is classified as a LPV chemical in the EU. Very little is known about this FR. | |
| HBCDD | Hexabromocyclo-dodecane | During 1999-2004, the use of HBCDD in Sweden fluctuated between 0 and 70 tonnes/year. No data is available after 2004. | [73] |
| TDBP-TAZTO | Tris(2,3-dibromopropyl)-isocyanurate | TDBP-TAZTO is an additive FR used in many different materials, e.g. polyurethane, polyolefin, polyvinyl chloride (PVC) and synthetic rubber. It is classified as a LPV chemical in the EU and is used in a mixture with DBP-TAZTO and BDBP-TAZTO. According to the Danish Environmental Protection Agency (2000), TDBP-TAZTO can be classified as a toxic, persistent and very bioaccumulative compound. | [70, 71] |
| DBP-TAZTO | 1-(2,3-dibromopropyl)-3,5-diallyl-1,3,5-triazine-2,4,6(1H,3H,5H)-trione | DBP-TAZTO has been found in plastic consumer products in Switzerland. It is used in a mixture with TDBP-TAZTO and BDBP-TAZTO. | [71] |
| BDBP-TAZTO | 1,3-bis(2,3-dibromopropyl-5-(2-propen-1-yl)-1,3,5-triazine-2,4,6(1H,3H,5H)-trione | BDBP-TAZTO has been found in plastic consumer products in Switzerland. It is used in a mixture with TDBP-TAZTO and DBP-TAZTO. | [71] |
| DBDPE | Decabromo-diphenylethane | DBDPE was introduced to the market in the mid-1980's as a replacement for BDE-209. It is used in materials like plastics, rubber and polymers used in electronic equipment. DBDPE is listed as a LPV chemical in the European chemical substances Information System (ESIS). In Sweden, 39 tonnes of DBDPE were used in 2008 but only about 7 tonnes in 2012. However, in China, DBDPE is the second most used BFR. The hatching success of some fish has been found to be suppressed by DBDPE. | [71], [73] |
| BTBPE | 1,2-Bis(2,4,6-tribromophenoxy)-ethane | BTBPE has been produced since the 1970's and is used as an additive FR in e.g. thermoplastics and polycarbonate. In the EU, it is listed as a LPV chemical. BTBPE has shown potential biomagnification in aquatic food webs, and it is considered to be highly persistent in the environment even though it might be debrominated. | [70, 71] |
| BEH-TEBP | Bis(2-ethyl-1-hexyl)tetrabromophthalate | BEH-TEBP is used in e.g., PVC, neoprene, wire insulation and wall coverings. It is often used in a mixture with EH-TBB in polyurethane foam. Both compounds undergo debromination in photo-degradation experiments. For BEH-TEBP, this can result | [70] |

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| EH-TBB | 2-Ethylhexyl- 2,3,4,5-tetrabromo-benzoate | in formation of the plasticizer bis(2-ethylhexyl)-phtalate which is restricted or banned in many countries. | |
| HEEHP-TEBP | 2-(2-hydroxyethoxy)-ethyl 2-hydroxy-propyl- 3,4,5,6-tetrabromo-phthalate | HEEHP-TEBP is classified as a LPV chemical in the EU. It consists of a pair of enantiomers. It has been reported to be used during production of foam for insulation. | [71] |
| OBTMPI | Octobromo-1,3,3-trimethyl-1-phenylindane | Very little is known about OBTMPI. It consists of one enantiomeric pair and can be expected to undergo debromination in the environment. | [71] |
| EBTEBPI | Ethylene bis-tetrabromo-phthalimide | EBTEBPI is used in polyethylene, polypropylene, thermoplastics, rubber, textiles and other plastic materials. EBTEBPI has been reported as bioaccumulative, but conflicting results have been reported in literature. | [70] |
| 4'-PeBPOBDE208 | Pentabromo-phenoxy-nonabromo-diphenyl ether | 4'-PeBPOBDE-208 is classified as a LPV in the EU. It is used in high performance polyamide and linear polyester engineering resins and alloys. Its bioavailability is likely to be restricted by its high molecular mass, and it is expected to undergo photolysis in the environment. | [71] |
| TBBPA | Tetrabromo-bisphenol-A | TBBPA is the most used BFR worldwide. It can be used as both an additive and reactive component and is used in printed circuit boards, plastics and resins. The content of TBBPA in printed circuit boards can be as high as 34 % (by weight). TBBPA has been shown to be a potential disruptor of the thyroid hormone system. In Sweden, the usage of TBBPA peaked at 438 tonnes/year in 2000. After that, the usage has decreased and was 24 tonnes/year in 2010. | [20], [60], [73] |
| TBBPA-DHEE | Tetrabromo-bisphenol-A dihydroxyethyl ether | TBBPA-DHEE is an additive FR used in engineering polymers and coatings. It is listed as a LPV chemical in the EU and is likely to be stable at normal environmental pH. | [70] |
| TBBPA-BDBPE | Tetrabromo-bisphenol A bis(2,3-dibromopropyl ether) | TBBPA-BDBPE is an additive FR found in pipes, water barriers, kitchen hoods and electronics. In the EU it is listed as a LPV chemical. The compound can be hydrolyzed to form TBBPA bis-(bromopropenyl ether) which might be more prevalent in sediment than the original compound. | [70] |
| TBBPA-BAE | Tetrabromo-bisphenol-A bis(allyl ether) | TBBPA-BAE is used both as an additive and reactive FR. It is used in polystyrene foam and expanded polystyrene. TBBPA-BAE is considered to be a LPV chemical in the EU. Laboratory experiments show that it might be resistant to degradation in the environment. | [70] |
| TBBPS-DBPE | Tetrabromo- | No information found. | |

| | | | |
|--|--|--|--------------|
| | bisphenol-S-bis(2,3-dibromopropyl)ether | | |
| TTBP-TAZ | 2,4,6-tris(2,4,6-tribromophenoxy)-1,3,5-triazine | TTBP-TAZ is used in high-impact polystyrene polymers and in acrylonitrile butadiene styrene polymers. | [71] |
| Chlorinated and brominated flame retardants (CBFRs) | | | |
| DDC-CO | Dechlorane Plus | DDC-CO is used in e.g., coatings for electrical wires and in plastic roofing materials. DDC-CO can be assumed to no longer exist on the Nordic chemical product market but can still be present in imported goods. | [20] |
| PBCH | Pentabromo-chlorocyclohexane | No information found. | |
| HCTBPH | 1,2,3,4,7,7-hexachloro-5-(2,3,4,5-tetrabromophenyl)-bicyclo[2.2.1]hept-2-ene | HCTBPH belongs to the group of chemicals called Dechloranes. It has been used since the mid1960's as an additive FR in plastics, rubber, paint and electrical equipment. It is expected to undergo dehalogenation in the environment. | [71] |
| DBHCTD | Hexachloro-cyclopentadienyl-dibromo-cyclooctane | This FR is used in styrenic polymers such as polystyrene and styrene butadiene rubber. The first time DBHCTD was reported as a FR was in an US patent in 1975. DBHCTD consists of two enantiomers. The stability of DBHCTD in abiotic matrices is mainly unknown, but it is likely to be a stable compound. | [70, 71, 73] |
| TBCT | 1,2,3,4-Tetrabromo-5-chloro-6-methylbenzene | No information found. | |
| TCBPA | Tetrachloro-bisphenol A | TCBPA is less used than TBBPA but chlorination of bisphenol-A in aqueous media can lead to its formation. Like TBBPA, TCBPA has been shown to have potential of disrupting the thyroid hormone. | [60] |
| Phosphorous flame retardants (PFRs) | | | |
| TEP | Triethyl phosphate | TEP is used in e.g. PVC, polyester resins and polyurethane foam. In total the usage of TEP in the Nordic countries was 78 tonnes in 2010. | [73, 74]. |
| TCEP | Tris(2-chloroethyl) phosphate | TCEP is an additive FR used in e.g. PVC, textile and polyurethane foam. It is considered dangerous to the environment and carcinogenic to animals. Reduced fertility for humans has been reported as a result of exposure to TCEP. Due to its toxicity, the worldwide use and production of TCEP is believed to have been phased out. Nevertheless, there are still plenty of TCEP | [73, 74] |

| | | | |
|---------------|---------------------------------------|--|-----------------|
| | | built into houses and buildings that may continue to pose a threat to human health and the environment. According to the Swedish Chemicals Agency, the use of TCEP in the Nordic countries has decreased since 2000. In 2010, Sweden and Denmark used less than 10 tonnes each, while Finland and Norway used 147 and 65 tonnes, respectively. | |
| TPP | Tripropyl phosphate | No information found. | |
| TCIPP | Tris(chloropropyl) phosphate | TCIPP was in 2000 the most important PFR in Europe (by volume). It is used in resins, latexes and foams present in e.g. furniture. TCIPP have shown low degradability in experiments and it is assumed that it might bioaccumulate. | [74] |
| TDCIPP | Tris(1,3-dichloro-2-propyl) phosphate | TDCIPP is an additive FR used for the same kind of applications as TCIPP. It is more expensive than TCIPP but also more effective. It is therefore only used when extra fire safety is needed. However, contrasting facts exist. In one study, TDCIPP was detected in 52% of the sampled couches purchased in the US after 2005 indicating a much larger use. In the Nordic countries, about 1400 tonnes were used in 2010. TDCIPP is classified as dangerous for the environment and is not easily degraded. Nevertheless, it is rapidly biotransformed in fish and is not considered to bioconcentrate to any large extent. TDCIPP has been reported to be carcinogenic. | [74], [9], [73] |
| TDBPP | Tris(2,3-dibromopropyl) phosphate | The production of TDBPP is estimated to have started about 60 years ago. It has been used as a FR on cellulose, polyester fabrics and carpets among others. In the 1970's, TDBPP was removed from the market due to its mutagenic and carcinogenic properties. There are, however, indications that TDBPP might still be in use today. | [71] |
| TNBP | Tributyl phosphate | TNBP is used in hydraulic fluids, lacquers and plastic. In total, 168 tonnes of TNBP was used in the Nordic countries in 2010. | [74], [73] |
| TTBNPP | Tris(tribromo-neopentyl) phosphate | TTBNPP is often used in polypropylene products, such as carpets and stadium seats. It is considered to be stable in the environment but can be degraded biologically. | [71] |
| TPHP | Triphenyl phosphate | TPHP is used in plastics and hydraulic fluids. When leached into water, it is rapidly adsorbed to sediment, and it is not considered persistent or bioaccumulative. TPHP have shown toxicity to water-living organisms. Among the Nordic countries, Sweden used the most TPHP in 2010, about 100 tonnes. | [74], [73] |

| | | | |
|------------------|--|---|------------|
| TBOEP | Tris(2-butoxyethyl) phosphate | TBOEP is used in floor polish, lacquers, plastic and rubber. In Sweden, 18 tonnes were used in 2010. | [74], [73] |
| EHDPP | 2-Ethylhexyl-diphenyl phosphate | EHDPP is used in e.g. hydraulic fluids. EHDPP is also used in PVC as a plasticizer. The usage in Sweden, Norway and Denmark were all below 37 tonnes/year in 2010. In Finland, however, the usage was 282 tonnes in 2010. | [74], [73] |
| TMPP | Tritolyl phosphate | TMPP is used in e.g. hydraulic fluids. TMPP is also used in PVC as a plasticizer. Less than two tons were used in each of the Nordic countries in the year 2010. | [74], [73] |
| TEHP | Tris(2-ethylhexyl) phosphate | TEHP is used in for instance PVC, paints, rubber and polyurethane foam. The main user of TEHP among the Nordic countries is Finland. In 2010, Finland used 263 tons to be compared with Sweden which used 25 tons. | [74], [73] |
| TBPP | Tris(4-tert-butylphenyl) phosphate | TBPP has been detected in 13% of the sampled couches purchased in the US year 2005 or later. It is not used in Sweden. | [9], [73] |
| BCMP-BCEP | 2,2-Bis(chloromethyl)-1,3-propanediyl-tetrakis(2-chloroethyl) bis(phosphate) | BCMP-BCEP main usage is in polyurethane foams in cars. It is considered not readily biodegradable. Less than 5000 tons were produced within EU in 2000. | [75] |

^aThe information presented in this table was collected in year 2013. More updated information may have been presented in other parts of this report.

Table S2 Data for FRs in Sweden obtained from the Swedish Chemical Agency (KemI).

| | Range of Use (1-7) | Prod grp (1-7) | Consumer grp (0-7) | Article prod grp (1-7) | Quantity grp (1-7) | Quant reduction>25% | EI_Surface water | EI_Air | EI_Soil | EI_WWTP | EI_Consumer | EI_Occupational | EI_HumTrend | EI_EnvTrend | First_Year | Latest_Year | Swedish monitoring (number of analyses above detection limits, Swedish EPA screening database) |
|---|--------------------|----------------|--------------------|------------------------|--------------------|---------------------|------------------|--------|---------|---------|-------------|-----------------|-------------|-------------|------------|-------------|--|
| 4'-PeBPOBDE208 | 0 | 0 | 1 | 7 | 0 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1994 | 2010 | |
| Acetoguanamine | 0 | 0 | 1 | 1 | 0 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1994 | 2000 | |
| Ammeline | 1 | 1 | 1 | 1 | 1 | | 1 | 1 | 1 | 1 | 1 | 1 | 0 | 0 | 2010 | 2015 | |
| BADP | 1 | 1 | 1 | 7 | 5 | | 1 | 1 | 1 | 5 | 1 | 7 | 0 | 0 | 2004 | 2015 | |
| BEH-TEBP | 1 | 1 | 7 | 6 | 4 | | 2 | 2 | 2 | 1 | 4 | 7 | 0 | 0 | 2001 | 2015 | |
| Benzoguanamine | 0 | 0 | 1 | 1 | 0 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1992 | 2007 | |
| Bis-(t-butylphenyl) phenyl phosphate | 1 | 1 | 7 | 4 | 3 | | 6 | 1 | 6 | 3 | 6 | 7 | 0 | 0 | 1995 | 2015 | |
| BTBPE | 1 | 1 | 1 | 7 | 4 | | 1 | 1 | 1 | 3 | 1 | 7 | 0 | 0 | 1993 | 2015 | |
| CDP | 4 | 3 | 1 | 3 | 5 | | 6 | 3 | 6 | 6 | 6 | 7 | 0 | 1 | 1992 | 2015 | |
| CLP1 | 0 | 0 | 7 | 6 | 0 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2003 | 2011 | 10 |
| SCCP | 5 | 4 | 6 | 6 | 5 | | 6 | 3 | 7 | 7 | 7 | 7 | 0 | -1 | 1992 | 2015 | 85 |
| MCCP | 2 | 0 | 1 | 6 | 0 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1992 | 2014 | 133 |
| DBDPE | 2 | 3 | 1 | 7 | 5 | | 1 | 1 | 1 | 5 | 1 | 7 | 1 | 1 | 1995 | 2015 | 39 |
| DBNPG | 1 | 1 | 1 | 7 | 1 | | 1 | 1 | 1 | 1 | 1 | 1 | 0 | 0 | 1999 | 2015 | |
| DBPhP | 1 | 1 | 1 | 1 | 4 | | 5 | 2 | 7 | 5 | 7 | 7 | 0 | 0 | 1997 | 2015 | 2 |
| DMP | 0 | 0 | 1 | 1 | 0 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1992 | 2012 | 1 |
| DDC-CO | 1 | 1 | 1 | 6 | 1 | | 1 | 1 | 1 | 1 | 1 | 1 | 0 | 0 | 1992 | 2015 | 30 |
| DPhBP | 1 | 1 | 1 | 1 | 4 | | 5 | 1 | 7 | 5 | 6 | 7 | 0 | 0 | 1997 | 2015 | 2 |
| EBTEBPI | 2 | 2 | 1 | 7 | 4 | | 1 | 1 | 1 | 3 | 1 | 7 | 0 | 0 | 1992 | 2015 | |
| EHDPP | 3 | 3 | 1 | 7 | 5 | | 1 | 1 | 3 | 5 | 3 | 7 | 0 | 0 | 1992 | 2015 | 173 |

| | | | | | | | | | | | | | | | | | |
|-------------------------------|---|---|---|----|---|---|---|---|---|---|---|---|---|---|------|------|-----|
| HBCDD | 1 | 0 | 1 | 7 | 0 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1993 | 2014 | 297 |
| HEEHP-TEBP | 1 | 1 | 1 | 7 | 4 | | 1 | 1 | 1 | 3 | 1 | 7 | 0 | 0 | 1993 | 2015 | |
| IDP | 3 | 2 | 1 | 7 | 5 | | 1 | 1 | 1 | 4 | 6 | 7 | 0 | 0 | 1992 | 2015 | |
| Melamine | 4 | 4 | 1 | 99 | 5 | x | 3 | 3 | 4 | 6 | 7 | 7 | 0 | 0 | 1992 | 2015 | |
| Melamine cyanurate | 3 | 4 | 1 | 7 | 5 | | 2 | 1 | 2 | 6 | 2 | 7 | 0 | 0 | 1992 | 2015 | |
| Melamine pyrophosphate | 1 | 1 | 1 | 7 | 4 | | 1 | 1 | 1 | 2 | 1 | 6 | 0 | 0 | 2003 | 2015 | |
| Pentaerythritol | 5 | 5 | 1 | 5 | 6 | | 7 | 6 | 7 | 6 | 7 | 7 | 0 | 0 | 1992 | 2015 | |
| PBDPP | 2 | 2 | 1 | 7 | 5 | | 1 | 1 | 1 | 4 | 1 | 7 | 1 | 1 | 1993 | 2015 | |
| T2CPP | 1 | 1 | 7 | 1 | 4 | | 1 | 1 | 1 | 1 | 1 | 1 | 0 | 0 | 1992 | 2015 | |
| TBBPA | 1 | 3 | 1 | 99 | 1 | x | 1 | 1 | 1 | 1 | 1 | 5 | 0 | 0 | 1992 | 2015 | 61 |
| TBBPA-BDBPE | 1 | 1 | 1 | 7 | 5 | | 1 | 1 | 1 | 4 | 1 | 7 | 2 | 2 | 1998 | 2015 | |
| TBOEP | 5 | 5 | 3 | 3 | 5 | | 5 | 3 | 7 | 6 | 6 | 7 | 0 | 0 | 1992 | 2015 | 214 |
| TBPP | 1 | 1 | 1 | 7 | 3 | | 1 | 1 | 4 | 1 | 2 | 4 | 0 | 0 | 1995 | 2015 | |
| TCEP | 2 | 1 | 1 | 6 | 4 | | 2 | 1 | 1 | 4 | 1 | 7 | 0 | 0 | 1992 | 2015 | 425 |
| TCIPP | 4 | 5 | 4 | 7 | 5 | | 4 | 3 | 4 | 6 | 5 | 7 | 0 | 0 | 1992 | 2015 | 289 |
| TMPP (o-) | 2 | 2 | 1 | 4 | 3 | | 1 | 1 | 1 | 1 | 1 | 6 | 1 | 1 | 2002 | 2015 | 1 |
| TMPP (o-, m-, p-) | 4 | 4 | 1 | 5 | 5 | | 7 | 2 | 7 | 6 | 7 | 7 | 0 | 1 | 1992 | 2015 | 111 |
| TMPP (p-) | 4 | 4 | 1 | 5 | 5 | | 7 | 2 | 7 | 6 | 7 | 7 | 0 | 1 | 1992 | 2015 | 111 |
| TDCIPP | 0 | 0 | 1 | 7 | 0 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1992 | 2010 | 237 |
| TEHP | 3 | 3 | 1 | 1 | 5 | | 5 | 5 | 5 | 5 | 5 | 7 | 0 | 0 | 1992 | 2015 | 38 |
| TEP | 4 | 4 | 2 | 4 | 5 | | 2 | 2 | 2 | 4 | 5 | 7 | 0 | 0 | 1992 | 2015 | |
| TIBP | 5 | 4 | 1 | 4 | 5 | | 6 | 2 | 7 | 6 | 6 | 7 | 0 | 0 | 1992 | 2015 | 37 |
| TMP | 0 | 0 | 1 | 1 | 0 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1993 | 1993 | 5 |
| TNBP | 5 | 4 | 1 | 1 | 5 | | 6 | 6 | 7 | 6 | 7 | 7 | 0 | 0 | 1992 | 2015 | 349 |
| TPHP | 5 | 4 | 1 | 6 | 5 | | 6 | 5 | 7 | 6 | 7 | 7 | 0 | 0 | 1992 | 2015 | 288 |
| TTBNPP | 1 | 1 | 1 | 7 | 3 | | 1 | 1 | 1 | 2 | 1 | 7 | 2 | 1 | 2005 | 2015 | |
| TTBNPP | 1 | 1 | 1 | 7 | 3 | | 1 | 1 | 1 | 2 | 1 | 7 | 2 | 1 | 2005 | 2015 | |
| TXP | 2 | 1 | 1 | 2 | 4 | | 5 | 2 | 7 | 5 | 7 | 7 | 0 | 0 | 1992 | 2015 | |
| BCMP-BCEP | 0 | 0 | 1 | 6 | 0 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1992 | 2001 | i |

Table S3 Concentrations of detected FRs in indoor air (ng m⁻³). Detection frequency in brackets.

| <u>Home</u> | <u>Type of environment, country</u> | <u>TBP</u> | <u>2,4-DBP</u> | <u>TBP-AE</u> | <u>BATE</u> | <u>BEH-TEBP</u> | <u>BTBPE</u> | <u>CLP1</u> |
|---------------------------|---|------------|----------------|---------------|-------------|-----------------|--------------|-------------|
| Bergh 2011 | Home, Sweden | | | | | | | |
| Cequier 2014 ^a | Living room, Norway | | | 0.0693 (70%) | nd | 0.0242 (19%) | nd | |
| Luongo 2015 | House air, Stockholm, Sweden | | | | | | | |
| <u>Office</u> | <u>Type of environment, country</u> | <u>TBP</u> | <u>2,4-DBP</u> | <u>TBP-AE</u> | <u>BATE</u> | <u>BEH-TEBP</u> | <u>BTBPE</u> | <u>CLP1</u> |
| Bergh 2011 | Offices and workshop, Sweden | | | | | | | |
| Schlabach, 2011 | Office building, Oslo, Norway | nd | nd | nd | nd | 0.0067-0.0074 | 0.0088-0.019 | |
| <u>Public place</u> | <u>Type of environment, country</u> | <u>TBP</u> | <u>2,4-DBP</u> | <u>TBP-AE</u> | <u>BATE</u> | <u>BEH-TEBP</u> | <u>BTBPE</u> | <u>CLP1</u> |
| Marklund 2005 | Different indoor compartments, Sweden | | | | | | | nd-0.8 |
| Green 2008 | Shopping centre, Norway | | | | | | | |
| Moskeland 2009 | Electronics store, Norway | nd | | | | nd | nd | |
| Bergh 2011 | Day care, Sweden | | | | | | | |
| Cequier 2014 ^a | School classroom, Norway | | | 0.00288 (0%) | | 0.00632 (33%) | | |
| Fromme 2014 | Daycare centers, Germany | | | | | | | |
| <u>Point source</u> | <u>Type of environment, country</u> | <u>TBP</u> | <u>2,4-DBP</u> | <u>TBP-AE</u> | <u>BATE</u> | <u>BEH-TEBP</u> | <u>BTBPE</u> | <u>CLP1</u> |
| Remberger 2014 | Point source, recycling hall, electronics | | | | | | | |

^aMaximum value

| <u>Home</u> | <u>DBDPE</u> | <u>DBE-DBCH</u> <u>(α-)</u> | <u>DBE-DBCH</u> <u>(β-)</u> | <u>DBE-DBCH</u> <u>(γ-)</u> | <u>DBE-DBCH</u> <u>(δ-)</u> | <u>DBHCT</u> <u>D</u> | <u>DDC-</u> <u>DBF</u> | <u>DDC-</u> <u>Ant</u> | <u>DOP</u> <u>P</u> | <u>DDC-CO</u> <u>(anti-)</u> | <u>DDC-CO</u> <u>(syn-)</u> |
|---------------------------------|--------------|--|---|--|--|--------------------------|---------------------------|---------------------------|------------------------|---------------------------------|--------------------------------|
| Bergh 2011 | | | | | | | | | | | |
| Cequier 2014^a | 0.963 (47%) | 4.120 (100%) | | | | nd | nd | nd | | 0.00761 (4%) | 0.00739 (2%) |
| Luongo 2015 | | | | | | | | | | | |
| <u>Office</u> | <u>DBDPE</u> | <u>DBE-DBCH</u> <u>(α-)</u> | <u>DBE-DBCH</u> <u>(β-)</u> | <u>DBE-DBCH</u> <u>(γ-)</u> | <u>DBE-DBCH</u> <u>(δ-)</u> | <u>DBHCT</u> <u>D</u> | <u>DDC-</u> <u>DBF</u> | <u>DDC-</u> <u>Ant</u> | <u>DOP</u> <u>P</u> | <u>DDC-CO</u> <u>(anti-)</u> | <u>DDC-CO</u> <u>(syn-)</u> |
| Bergh 2011 | | | | | | | | | | | |
| Schlabach, 2011 | nd-0.056 | nd-0.0023 | nd-0.00079 | nd | | | | | | nd | nd |
| <u>Public place</u> | <u>DBDPE</u> | <u>DBE-DBCH</u> <u>(α-)</u> | <u>DBE-DBCH</u> <u>(β-)</u> | <u>DBE-DBCH</u> <u>(γ-)</u> | <u>DBE-DBCH</u> <u>(δ-)</u> | <u>DBHCT</u> <u>D</u> | <u>DDC-</u> <u>DBF</u> | <u>DDC-</u> <u>Ant</u> | <u>DOP</u> <u>P</u> | <u>DDC-CO</u> <u>(anti-)</u> | <u>DDC-CO</u> <u>(syn-)</u> |
| Marklund 2005 | | | | | | | | | nd-4.8 | | |
| Green 2008 | | | | | | | | | | | |
| Moskeland 2009 | nd | | | | | | | | | | |
| Bergh 2011 | | | | | | | | | | | |
| Cequier 2014^a | 0.0206 (50%) | 0.399 (100%) | | | | nd | nd | nd | | nd | nd |
| Fromme 2014 | | | | | | | | | | | |
| <u>Point source</u> | <u>DBDPE</u> | <u>DBE-DBCH</u> <u>(α-)</u> | <u>DBE-DBCH</u> <u>(β-)</u> | <u>DBE-DBCH</u> <u>(γ-)</u> | <u>DBE-DBCH</u> <u>(δ-)</u> | <u>DBHCT</u> <u>D</u> | <u>DDC-</u> <u>DBF</u> | <u>DDC-</u> <u>Ant</u> | <u>DOP</u> <u>P</u> | <u>DDC-CO</u> <u>(anti-)</u> | <u>DDC-CO</u> <u>(syn-)</u> |
| Remberger 2014 | 220-530 | 6.7-8.3 | | | | | | | | | |

^aMaximum value

| <u>Home</u> | <u>EBTEBPI</u> | <u>EHDPP</u> | <u>EH-TBB</u> | <u>HBB</u> | <u>OBTMPI</u> | <u>PBB</u> | <u>PBB-Acr</u> | <u>PBEB</u> | <u>PBP</u> | <u>PBT</u> | <u>TBBPA</u> | <u>TBBPA-BAE</u> |
|---------------------------|----------------|---------------|---------------|---------------|---------------|---------------|----------------|--------------|------------|----------------|--------------|------------------|
| Bergh 2011 | | nd-<LOQ | | | | | | | | | | |
| Cequier 2014 ^a | | 0.00351 (62%) | | 0.297 (70%) | nd | 0.0508 (100%) | nd | 0.0306 (45%) | | 0.213 (100%) | | |
| Luongo 2015 | | nd-3.9 (3%) | | | | | | | | | | |
| <u>Office</u> | <u>EBTEBPI</u> | <u>EHDPP</u> | <u>EH-TBB</u> | <u>HBB</u> | <u>OBTMPI</u> | <u>PBB</u> | <u>PBB-Acr</u> | <u>PBEB</u> | <u>PBP</u> | <u>PBT</u> | <u>TBBPA</u> | <u>TBBPA-BAE</u> |
| Bergh 2011 | | nd-14 | | | | | | | | | | |
| Schlabach, 2011 | | | nd-0.0067 | 0.012-0.015 | | | | nd-0.00021 | nd | 0.0025-0.0028 | nd | |
| <u>Public place</u> | <u>EBTEBPI</u> | <u>EHDPP</u> | <u>EH-TBB</u> | <u>HBB</u> | <u>OBTMPI</u> | <u>PBB</u> | <u>PBB-Acr</u> | <u>PBEB</u> | <u>PBP</u> | <u>PBT</u> | <u>TBBPA</u> | <u>TBBPA-BAE</u> |
| Marklund 2005 | | | | | | | | | | | | |
| Green 2008 | | nd-0.42 (28%) | | | | | | | | | | |
| Moskeland 2009 | nd | | nd | nd | | | | nd | | nd | | nd |
| Bergh 2011 | | nd-2.2 | | | | | | | | | | |
| Cequier 2014 ^a | | 0.00503 (67%) | | 0.00652 (83%) | nd | 0.00735 (83%) | | nd | | 0.00414 (100%) | | |
| Fromme 2014 | | | | | | | | | | | | |
| <u>Point source</u> | <u>EBTEBPI</u> | <u>EHDPP</u> | <u>EH-TBB</u> | <u>HBB</u> | <u>OBTMPI</u> | <u>PBB</u> | <u>PBB-Acr</u> | <u>PBEB</u> | <u>PBP</u> | <u>PBT</u> | <u>TBBPA</u> | <u>TBBPA-BAE</u> |
| Remberger 2014 | | | | 1400-1600 | | | | nd-12 | | 14-25 | | |

^aMaximum values

| <u>Home</u> | <u>TBBPA-BDBPE</u> | <u>TBCO</u> | <u>TBOEP</u> | <u>TBP-DBPE</u> | <u>TBX</u> | <u>TCEP</u> | <u>TCIPP</u> | <u>TMPP (o-)</u> | <u>TMPP (m-)</u> | <u>TMPP (p-)</u> |
|---------------------------|--------------------|-------------|---------------|-----------------|---------------|---------------|-----------------|------------------|------------------|------------------|
| Bergh 2011 | | | nd-4.5 | | | nd-28 | 2.4-64 | | | nd-<LOQ |
| Cequier 2014 ^a | | nd | 0.0182 (100%) | 0.132 (40%) | 2.830 (38%) | 0.0101 (98%) | 0.462 (100%) | 0.000644 (57%) | | |
| Luongo 2015 | | | nd-10 (5%) | | | nd-233 (65%) | 1.3-1179 (100%) | | | |
| <u>Office</u> | <u>TBBPA-BDBPE</u> | <u>TBCO</u> | <u>TBOEP</u> | <u>TBP-DBPE</u> | <u>TBX</u> | <u>TCEP</u> | <u>TCIPP</u> | <u>TMPP (o-)</u> | <u>TMPP (m-)</u> | <u>TMPP (p-)</u> |
| Bergh 2011 | | | nd-73 | | | nd-140 | 16-240 | | | nd-1.0 |
| Schlabach 2011 | | | | nd-0.0017 | | | | | | |
| <u>Public place</u> | <u>TBBPA-BDBPE</u> | <u>TBCO</u> | <u>TBOEP</u> | <u>TBP-DBPE</u> | <u>TBX</u> | <u>TCEP</u> | <u>TCIPP</u> | <u>TMPP (o-)</u> | <u>TMPP (m-)</u> | <u>TMPP (p-)</u> |
| Marklund 2005 | | | nd-55 | | | 0.4-730 | 10-570 | | | |
| Green 2008 | | | 8-55 (100%) | | | 2.7-23 (100%) | 10-49 (100%) | | | |
| Moskeland 2009 | nd | | | nd | | | | | | |
| Bergh 2011 | | | nd-380 | | | 7.8-230 | 1.3-72 | | | nd-<LOQ |
| Cequier 2014 ^a | | nd | 0.0201 (100%) | 0.0106 (50%) | 0.00292 (17%) | 0.0213 (100%) | 0.0251 (100%) | nd | | |
| Fromme 2014 | | | nd-1279 | | | nd-33 | nd-45 | | | |
| <u>Point source</u> | <u>TBBPA-BDBPE</u> | <u>TBCO</u> | <u>TBOEP</u> | <u>TBP-DBPE</u> | <u>TBX</u> | <u>TCEP</u> | <u>TCIPP</u> | <u>TMPP (o-)</u> | <u>TMPP (m-)</u> | <u>TMPP (p-)</u> |
| Remberger 2014 | | | | | | | | | | |

^aMaximum values

| <u>Home</u> | <u>TDCIPP</u> | <u>TEBP-Anh</u> | <u>TEEdP</u> | <u>TEHP</u> | <u>TEP</u> | <u>TIBP</u> | <u>TNBP</u> | <u>TPHP</u> | <u>TPP</u> | <u>BCMP-BCEP</u> |
|---------------------------|-----------------|-----------------|--------------|-------------|-----------------|---------------|----------------|-------------------|------------|------------------|
| Bergh 2011 | nd-17 | | | nd | 3.2-16 | 3.0-66 | 3.5-45 | nd-0.8 | | |
| Cequier 2014 ^a | 0.0106 (98%) | | | | | | 0.124 (100%) | 0.00165 (89%) | | |
| Luongo 2015 | | | | | 0.84-297 (100%) | 2.4-53 (100%) | 2.5-94 (100%) | nd-25 (15%) | | |
| <u>Office</u> | <u>TDCIPP</u> | <u>TEBP-Anh</u> | <u>TEEdP</u> | <u>TEHP</u> | <u>TEP</u> | <u>TIBP</u> | <u>TNBP</u> | <u>TPHP</u> | <u>TPP</u> | <u>BCMP-BCEP</u> |
| Bergh 2011 | nd-73 | | | nd | 0.7-91 | 4.4-13 | nd-100 | nd-2.7 | | |
| Schlabach 2011 | | | | | | | | | | |
| <u>Public place</u> | <u>TDCIPP</u> | <u>TEBP-Anh</u> | <u>TEEdP</u> | <u>TEHP</u> | <u>TEP</u> | <u>TIBP</u> | <u>TNBP</u> | <u>TPHP</u> | <u>TPP</u> | <u>BCMP-BCEP</u> |
| Marklund 2005 | nd-150 | | Excluded | nd-14 | | | nd-120 | nd-23 | nd-8.4 | |
| Green 2008 | nd-18 (50%) | | | | | nd-7.9 (75%) | 8.2-16 (100%) | 2300-47000 (100%) | | nd |
| Moskeland 2009 | | nd | | | | | | | | |
| Bergh 2011 | nd-30 | | | nd | 0.8-20 | nd-63 | 3.7-320 | nd-0.9 | | |
| Cequier 2014 ^a | 0.000140 (100%) | | | | | | 0.00457 (100%) | 0.000234 (100%) | | |
| Fromme 2014 | | | | | | | nd-80 | | | |
| <u>Point source</u> | <u>TDCIPP</u> | <u>TEBP-Anh</u> | <u>TEEdP</u> | <u>TEHP</u> | <u>TEP</u> | <u>TIBP</u> | <u>TNBP</u> | <u>TPHP</u> | <u>TPP</u> | <u>BCMP-BCEP</u> |
| Remberger 2014 | | | | | | | | | | |

^aMaximum values

Table S4 Concentrations of detected FRs in indoor dust. Detection frequency in brackets.

| <u>Home</u> | <u>Year</u> | <u>Unit</u> | <u>Type of environment, country</u> | <u>TBP-AE</u> | <u>BATE</u> | <u>BEH-TEBP</u> | <u>BTBPE</u> |
|---------------------------|-------------|-------------------|---|---------------|-------------|-----------------|---------------|
| Van den Eede 2011 | 2011 | ug/g | House dust, Belgium | | | | |
| Bergh 2011 | 2011 | ug/g | Home, Sweden | | | | |
| Abdallah 2014 | 2014 | ug/g | Houses | | | | |
| Cequier 2014 ^a | 2014 | ug/g | Living room, Norway | nd | nd | 0.809 (100%) | 0.0419 (92%) |
| He 2015 | 2015 | ug/g | Rural home, China | | | | |
| He 2015 | 2015 | ug/g | Urban home, China | | | | |
| Hoffman 2015 | 2015 | ug/g | Homes, US | | | | |
| Luongo 2015 | 2015 | ug/g | House dust (2008), Stockholm, Sweden | | | | |
| Brommer 2015 | 2015 | ug/g | Living room, UK | | | | |
| Langer 2016 | 2016 | ug/g | Homes, Denmark | | | | |
| <u>Office</u> | <u>Year</u> | <u>Unit</u> | <u>Type of environment, country</u> | <u>TBP-AE</u> | <u>BATE</u> | <u>BEH-TEBP</u> | <u>BTBPE</u> |
| Bergh 2011 | 2011 | ug/g | Offices and workshop, Sweden | | | | |
| Abdallah 2014 | 2014 | ug/g | Office, Egypt | | | | |
| Brommer 2015 | 2015 | ug/g | Office, UK | | | | |
| <u>Public place</u> | <u>Year</u> | <u>Unit</u> | <u>Type of environment, country</u> | <u>TBP-AE</u> | <u>BATE</u> | <u>BEH-TEBP</u> | <u>BTBPE</u> |
| Marklund 2003 | 2003 | ug/g | Different indoor compartments, such as e.g., home, hospital, prison, Sweden | | | | |
| Bergh 2011 | 2011 | ug/g | Day care, Sweden | | | | |
| Van den Eede 2011 | 2011 | ug/g | Dust from different types of stores, Belgium | | | | |
| Abdallah 2014 | 2014 | ug/g | Public places, such as e.g., restaurants, Egypt | | | | |
| Cequier 2014 ^a | 2014 | ug/g | School classroom, Norway | nd | nd | 0.151 (100%) | 0.0530 (100%) |
| Fromme 2014 | 2014 | ug/g | Daycare centers, Germany | | | | |
| Remberger 2014 | 2014 | ug/g | Public places and homes, Sweden | | | | |
| He 2015 | 2015 | ug/g | Urban college dormitory, China | | | | |
| Brommer 2015 | 2015 | ug/g | Classroom, UK | | | | |
| Langer 2016 | 2016 | ug/g | Daycare centers, Denmark | | | | |
| <u>Point source</u> | <u>Year</u> | <u>Unit</u> | <u>Type of environment, country</u> | <u>TBP-AE</u> | <u>BATE</u> | <u>BEH-TEBP</u> | <u>BTBPE</u> |
| Marklund 2003 | 2003 | ng/m ² | Dust from computer, Sweden | | | | |
| Mai 2009 | 2009 | ug/g dw | Dust from e-waste processing area, China | | | | 0.0146-0.232 |
| Remberger 2014 | 2014 | ug/g | Dust in new car (2012) | | | | |
| Remberger 2014 | 2014 | ug/g | Dust, recycling hall, electronics | | | | |
| Abdallah 2014 | 2014 | ug/g | Cars, Egypt | | | | |
| Brommer 2015 | 2015 | ug/g | Car, UK | | | | |
| He 2015 | 2015 | ug/g | E-waste workshop, China | | | | |

| <u>Homes</u> | <u>Unit</u> | <u>CLP1</u> | <u>DBDPE</u> | <u>DBE-DBCH</u> | <u>DBHCTD</u> | <u>DDC-DBF</u> | <u>DDC-Ant</u> | <u>DOPP</u> | <u>DDC-CO (anti-)</u> | <u>DDC-CO (syn-)</u> | <u>EHDPP</u> |
|---------------------------|-------------------|-------------|---------------|-----------------|---------------|----------------|----------------|-------------|-----------------------|----------------------|--------------------|
| Van den Eede 2011 | ug/g | | | | | | | | | | |
| Bergh 2011 | ug/g | | | | | | | | | | nd-1.8 |
| Abdallah 2014 | ug/g | | | | | | | | | | 0.052 ±0.032 (55%) |
| Cequier 2014 ^a | ug/g | | 4.460 (96%) | 0.172 (96%) | nd | nd | nd | | 0.590 (92%) | 0.311 (92%) | 5.900 (100%) |
| He 2015 | ug/g | | | | | | | | | | 0.06-1.28 |
| He 2015 | ug/g | | | | | | | | | | 0.03-3.47 |
| Hoffman 2015 | ug/g | | | | | | | | | | |
| Luongo 2015 | ug/g | | | | | | | | | | nd-20 (84%) |
| Brommer 2015 | ug/g | | | | | | | | | | 0.18-130 |
| Langer 2016 | ug/g | | | | | | | | | | nd-11 |
| <u>Office</u> | <u>Unit</u> | <u>CLP1</u> | <u>DBDPE</u> | <u>DBE-DBCH</u> | <u>DBHCTD</u> | <u>DDC-DBF</u> | <u>DDC-Ant</u> | <u>DOPP</u> | <u>DDC-CO (anti-)</u> | <u>DDC-CO (syn-)</u> | <u>EHDPP</u> |
| Bergh 2011 | ug/g | | | | | | | | | | nd-73 |
| Abdallah 2014 | ug/g | | | | | | | | | | 0.043 ±0.021 (60%) |
| Brommer 2015 | ug/g | | | | | | | | | | 0.15-81 |
| <u>Public places</u> | <u>Unit</u> | <u>CLP1</u> | <u>DBDPE</u> | <u>DBE-DBCH</u> | <u>DBHCTD</u> | <u>DDC-DBF</u> | <u>DDC-Ant</u> | <u>DOPP</u> | <u>DDC-CO (anti-)</u> | <u>DDC-CO (syn-)</u> | <u>EHDPP</u> |
| Marklund 2003 | ug/g | nd-0.11 | | | | | | nd-5.1 | | | |
| Bergh 2011 | ug/g | | | | | | | | | | 0.2-160 |
| Van den Eede 2011 | ug/g | | | | | | | | | | |
| Abdallah 2014 | ug/g | | | | | | | | | | 0.049 ±0.019 (36%) |
| Cequier 2014 ^a | ug/g | | 0.360 (83%) | 0.010 (100%) | nd | nd | nd | | 0.00925 (100%) | 0.00313 (83%) | 79.000 (100%) |
| Fromme 2014 | ug/g | | | | | | | | | | 0.30-95.6 |
| Remberger 2014 | ug/g | | 0.140-8.100 | nd-0.0013 | | | | | | | |
| He 2015 | ug/g | | | | | | | | | | nd-0.57 |
| Brommer 2015 | ug/g | | | | | | | | | | 0.30-470 |
| Langer 2016 | ug/g | | | | | | | | | | nd-540 |
| <u>Point sources</u> | <u>Unit</u> | <u>CLP1</u> | <u>DBDPE</u> | <u>DBE-DBCH</u> | <u>DBHCTD</u> | <u>DDC-DBF</u> | <u>DDC-Ant</u> | <u>DOPP</u> | <u>DDC-CO (anti-)</u> | <u>DDC-CO (syn-)</u> | <u>EHDPP</u> |
| Marklund 2003 | ng/m ² | nd | | | | | | 130-450 | | | |
| Mai 2009 | ug/g dw | | nd-0.139 | | | | | | | | |
| Remberger 2014 | ug/g | | 92 | nd | | | | | | | |
| Remberger 2014 | ug/g | | 20.000-23.000 | 0.0013-0.0028 | | | | | | | |
| Abdallah 2014 | ug/g | | | | | | | | | | 0.066 ±0.053 (50%) |
| Brommer 2015 | ug/g | | | | | | | | | | 0.29-11 |
| He 2015 | ug/g | | | | | | | | | | 0.15-2.39 |

^aMaximum values

| <u>Homes</u> | <u>Unit</u> | <u>EH-TBB</u> | <u>HBB</u> | <u>OBTMPI</u> | <u>PBB</u> | <u>PBB-Acr</u> | <u>PBEB</u> | <u>PBP</u> | <u>PBT</u> | <u>TBBPA</u> | <u>TBBPA-BDBPE</u> |
|---------------------------|-------------------|---------------|---------------|---------------|----------------|----------------|---------------|------------|---------------|--------------|--------------------|
| Van den Eede 2011 | ug/g | | | | | | | | | | |
| Bergh 2011 | ug/g | | | | | | | | | | |
| Abdallah 2014 | ug/g | | | | | | | | | | |
| Cequier 2014 ^a | ug/g | 0.245 (58%) | 0.00894 (50%) | nd | 0.00464 (40%) | 0.0113 (13%) | 0.00800 (33%) | | 0.0161 (94%) | | |
| He 2015 | ug/g | | | | | | | | | | |
| He 2015 | ug/g | | | | | | | | | | |
| Hoffman 2015 | ug/g | | | | | | | | | | |
| Luongo 2015 | ug/g | | | | | | | | | | |
| Brommer 2015 | ug/g | | | | | | | | | | |
| Langer 2016 | ug/g | | | | | | | | | | |
| <u>Office</u> | <u>Unit</u> | <u>EH-TBB</u> | <u>HBB</u> | <u>OBTMPI</u> | <u>PBB</u> | <u>PBB-Acr</u> | <u>PBEB</u> | <u>PBP</u> | <u>PBT</u> | <u>TBBPA</u> | <u>TBBPA-BDBPE</u> |
| Bergh 2011 | ug/g | | | | | | | | | | |
| Abdallah 2014 | ug/g | | | | | | | | | | |
| Brommer 2015 | ug/g | | | | | | | | | | |
| <u>Public places</u> | <u>Unit</u> | <u>EH-TBB</u> | <u>HBB</u> | <u>OBTMPI</u> | <u>PBB</u> | <u>PBB-Acr</u> | <u>PBEB</u> | <u>PBP</u> | <u>PBT</u> | <u>TBBPA</u> | <u>TBBPA-BDBPE</u> |
| Marklund 2003 | ug/g | | | | | | | | | | |
| Bergh 2011 | ug/g | | | | | | | | | | |
| Van den Eede 2011 | ug/g | | | | | | | | | | |
| Abdallah 2014 | ug/g | | | | | | | | | | |
| Cequier 2014 ^a | ug/g | 0.00572 (67%) | 0.00527 (67%) | nd | 0.000682 (50%) | nd | 0.000103 (0%) | | 0.00106 (67%) | | |
| Fromme 2014 | ug/g | | | | | | | | | | |
| Remberger 2014 | ug/g | | nd-0.020 | | | | nd | | 0.00072-0.002 | | |
| He 2015 | ug/g | | | | | | | | | | |
| Brommer 2015 | ug/g | | | | | | | | | | |
| Langer 2016 | ug/g | | | | | | | | | | |
| <u>Point sources</u> | <u>Unit</u> | <u>EH-TBB</u> | <u>HBB</u> | <u>OBTMPI</u> | <u>PBB</u> | <u>PBB-Acr</u> | <u>PBEB</u> | <u>PBP</u> | <u>PBT</u> | <u>TBBPA</u> | <u>TBBPA-BDBPE</u> |
| Marklund 2003 | ng/m ² | | | | | | | | | | |
| Mai 2009 | ug/g dw | | | | | | | | | | nd |
| Remberger 2014 | ug/g | | 0.08 | | | | nd | | 0.042 | | |
| Remberger 2014 | ug/g | | 1.200-8.200 | | | | 0.0091-0.016 | | 0.0054-0.064 | | |
| Abdallah 2014 | ug/g | | | | | | | | | | |
| Brommer 2015 | ug/g | | | | | | | | | | |
| He 2015 | ug/g | | | | | | | | | | |

^aMaximum values

| <u>Homes</u> | <u>Unit</u> | <u>TBCO</u> | <u>TBOEP</u> | <u>TBP-DBPE</u> | <u>TBX</u> | <u>TCEP</u> | <u>TCIPP</u> | <u>TMPP (o-)</u> | <u>TMPP (m-)</u> | <u>TMPP (p-)</u> |
|---------------------------|-------------------|-------------|--------------------|-----------------|-------------|--------------------|--------------------|------------------|------------------|------------------|
| Van den Eede 2011 | ug/g | | 0.36-67.6 (100%) | | | nd-2.65 (86%) | 0.19-73.7 (100%) | nd-5.07 (97%) | | |
| Bergh 2011 | ug/g | | 0.6-30 | | | nd-33 | 0.7-11 | | | nd-3.0 |
| Abdallah 2014 | ug/g | | 0.086 ±0.125 (25%) | | | 0.049 ±0.049 (55%) | 0.053 ±0.045 (45%) | | | |
| Cequier 2014 ^a | ug/g | nd | 128.000 (100%) | 0.0214 (69%) | 0.0888 (6%) | 4.630 (98%) | 40.100 (100%) | 16.200 (92%) | | |
| He 2015 | ug/g | | 0.03-1.76 | | | 0.05-9.36 | 0.24-10.7 | nd-3.65 | | |
| He 2015 | ug/g | | nd-3.05 | | | 1.55-9.70 | 0.16-2.93 | nd-7.74 | | |
| Hoffman 2015 | ug/g | | | | | | | | | |
| Luongo 2015 | ug/g | | nd-107 (98%) | | | nd-808 (97%) | 1.21-98 (100%) | nd-31 (85%) | | |
| Brommer 2015 | ug/g | | | | | nd-28 | 3.7-100 | nd-14 b | | |
| Langer 2016 | ug/g | | nd-1300 | | | nd-230 | nd-100 | | | nd-18 |
| <u>Office</u> | <u>Unit</u> | <u>TBCO</u> | <u>TBOEP</u> | <u>TBP-DBPE</u> | <u>TBX</u> | <u>TCEP</u> | <u>TCIPP</u> | <u>TMPP (o-)</u> | <u>TMPP (m-)</u> | <u>TMPP (p-)</u> |
| Bergh 2011 | ug/g | | 4.5-960 | | | 1.3-260 | 3.4-120 | | | nd-2.9 |
| Abdallah 2014 | ug/g | | 0.263 ±0.515 (30%) | | | 0.061 ±0.042 (55%) | 0.119 ±0.196 (55%) | | | |
| Brommer 2015 | ug/g | | | | | nd-160 | 3.6-230 | nd-5.3 b | | |
| <u>Public places</u> | <u>Unit</u> | <u>TBCO</u> | <u>TBOEP</u> | <u>TBP-DBPE</u> | <u>TBX</u> | <u>TCEP</u> | <u>TCIPP</u> | <u>TMPP (o-)</u> | <u>TMPP (m-)</u> | <u>TMPP (p-)</u> |
| Marklund 2003 | ug/g | | 14-5300 | | | 0.19-94 | 0.47-73 | | | |
| Bergh 2011 | ug/g | | 31-4100 | | | 2.5-150 | 0.8-12 | | | nd-13 |
| Van den Eede 2011 | ug/g | | 0.20-55.7 (100%) | | | nd-5.46 (93%) | 0.58-24.4 (100%) | nd-12.5 (93%) | | |
| Abdallah 2014 | ug/g | | 0.311 ±0.425 (64%) | | | 0.277 ±0.189 (64%) | 0.232 ±0.178 (73%) | | | |
| Cequier 2014 ^a | ug/g | nd | 163.000 (100%) | 0.000707 (50%) | nd | 6.160 (100%) | 2.740 (100%) | 0.333 (50%) | | |
| Fromme 2014 | ug/g | | 1.61-4711 | | | 0.10-8.3 | 0.71-47.0 | | | |
| Remberger 2014 | ug/g | | | | | | | | | |
| He 2015 | ug/g | | 0.04-0.77 | | | 2.78-20.8 | 0.06-2.30 | nd | | |
| Brommer 2015 | ug/g | | | | | nd-8.3 | 1.7-210 | nd-5.8 b | | |
| Langer 2016 | ug/g | | nd-11000 | | | nd-1800 | nd-350 | | | nd-36 |
| <u>Point sources</u> | <u>Unit</u> | <u>TBCO</u> | <u>TBOEP</u> | <u>TBP-DBPE</u> | <u>TBX</u> | <u>TCEP</u> | <u>TCIPP</u> | <u>TMPP (o-)</u> | <u>TMPP (m-)</u> | <u>TMPP (p-)</u> |
| Marklund 2003 | ng/m ² | | 170-940 | | | 210-220 | 220-370 | | | |
| Mai 2009 | ug/g dw | | | | | | | | | |
| Remberger 2014 | ug/g | | | | | | | | | |
| Remberger 2014 | ug/g | | | | | | | | | |
| Abdallah 2014 | ug/g | | 0.284 ±0.274 (30%) | | | 0.198 ±0.195 (50%) | 0.513 ±0.475 (55%) | | | |
| Brommer 2015 | ug/g | | | | | nd-8.7 | 2.4-370 | nd-5.6 b | | |
| He 2015 | ug/g | | 0.04-0.81 | | | 0.18-1.56 | 0.11-22.3 | 0.52-46.6 | | |

^aMaximum values

| <u>Homes</u> | <u>Unit</u> | <u>TDCIPP</u> | <u>TEEdP</u> | <u>TEHP</u> | <u>TEP</u> | <u>THP</u> |
|---------------------------|-------------------|---------------------|--------------|-------------|---------------|-------------|
| Van den Eede 2011 | ug/g | nd-6.64 (97%) | | | nd | |
| Bergh 2011 | ug/g | 2.2-27 | | nd-0.2 | nd | |
| Abdallah 2014 | ug/g | 0.147 ±0.164 (65%) | | | | |
| Cequier 2014 ^a | ug/g | 6.920 (100%) | | | | |
| He 2015 | ug/g | nd-2.77 | | 0.08-1.85 | 0.03-0.41 | |
| He 2015 | ug/g | nd-9.63 | | 0.03-1.37 | 0.02-0.24 | |
| Hoffman 2015 | ug/g | 0.197-39.530 (100%) | | | | |
| Luongo 2015 | ug/g | nd-12 (81%) | | nd-46 (2%) | nd-4.3 (84%) | nd-5.8 (6%) |
| Brommer 2015 | ug/g | 0.06-14 | | | | |
| Langer 2016 | ug/g | nd-860 | | nd-11 | | |
| <u>Office</u> | <u>Unit</u> | <u>TDCIPP</u> | <u>TEEdP</u> | <u>TEHP</u> | <u>TEP</u> | <u>THP</u> |
| Bergh 2011 | ug/g | 3.3-91 | | nd-0.3 | nd-0.3 | |
| Abdallah 2014 | ug/g | 0.099 ±0.137 (70%) | | | | |
| Brommer 2015 | ug/g | nd-51 | | | | |
| <u>Public places</u> | <u>Unit</u> | <u>TDCIPP</u> | <u>TEEdP</u> | <u>TEHP</u> | <u>TEP</u> | <u>THP</u> |
| Marklund 2003 | ug/g | 0.20-67 | 0.16-12 | 0.06-13 | | |
| Bergh 2011 | ug/g | 3.9-150 | | nd-0.7 | nd-4.7 | |
| Van den Eede 2011 | ug/g | nd-56.2 (93%) | | | nd-0.37 (53%) | |
| Abdallah 2014 | ug/g | 0.601 ±0.572 (91%) | | | | |
| Cequier 2014 ^a | ug/g | 6.140 (100%) | | | | |
| Fromme 2014 | ug/g | | | nd-4.85 | | |
| Remberger 2014 | ug/g | | | | | |
| He 2015 | ug/g | 0.06-3.71 | | nd-0.32 | 0.03-0.94 | |
| Brommer 2015 | ug/g | 0.04-10 | | | | |
| Langer 2016 | ug/g | nd-320 | | nd-3.8 | | |
| <u>Point sources</u> | <u>Unit</u> | <u>TDCIPP</u> | <u>TEEdP</u> | <u>TEHP</u> | <u>TEP</u> | <u>THP</u> |
| Marklund 2003 | ng/m ² | 170-290 | 290-560 | nd | | |
| Mai 2009 | ug/g dw | | | | | |
| Remberger 2014 | ug/g | | | | | |
| Remberger 2014 | ug/g | | | | | |
| Abdallah 2014 | ug/g | 0.087 ±0.076 (50%) | | | | |
| Brommer 2015 | ug/g | 0.11-740 | | | | |

^aMaximum values

Table S5 Concentrations of detected FRs in outdoor air ($\mu\text{g m}^{-3}$). Detection frequency in brackets.

| | <u>Brief description, country</u> | <u>2/3-BP</u> | <u>2,4,6-TBP</u> | <u>2,4-DBP</u> | <u>2,6-DBP</u> | <u>4-BP</u> |
|------------------------|--|---------------|--------------------|----------------|----------------|-------------|
| Marklund 2005 | Background air, Finland | | | | | |
| Green 2008 | Urban areas, Mainland Norway | | | | | |
| Green 2008 | Remote area, Svalbard and mainland, Norway | | | | | |
| Moskeland 2009 | Norway screening | | nd | | | |
| Mai 2009 | Pearl river delta, China | | | | | |
| Arp 2010 | Atm. Particles, city, Norway | | | | | |
| Möller 2010 | Sea air, Arctic to Antarctica | | | | | |
| Schlabach, 2011 | Background and urban, Sweden/Norway/Denmark | | nd-27 | nd-21 | | |
| Xie 2011 | Sea air from the Atlantic and Southern Ocean | | | | | |
| Remberger 2014 | Background sites in Sweden and Finland | | | | | |
| Remberger 2014 | Diffuse sources | | | | | |
| Haglund 2015 | Background sites in Sweden and Finland | 0.495-9.278 | 0.145-1.626 (100%) | 0.209-13.315 | 0.031-0.271 | 0.210-7.908 |

| | <u>TBP-AE</u> | <u>BATE</u> | <u>BEH-TEBP</u> | <u>BTBPE</u> | <u>DBDPE</u> | <u>DBE-DBCH (α-)</u> | <u>DBE-DBCH (β-)</u> |
|-----------------------|---------------|-------------|-----------------|--------------------|----------------|--|---------------------------------------|
| Marklund 2005 | | | | | | | |
| Green 2008 | | | | | | | |
| Green 2008 | | | | | | | |
| Moskeland 2009 | | | nd | nd | nd | | |
| Mai 2009 | | | | 3.83-67.4 | 402-3578 | | |
| Arp 2010 | | | | | | | |
| Möller 2010 | | | | | | | |
| Schlabach 2011 | nd-0.27 | nd-0.051 | nd-1.7 | nd-2.2 | nd-44 | 0.039-13 | nd-11 |
| Xie 2011 | | | | | | | |
| Remberger 2014 | | | | | nd | nd | |
| Remberger 2014 | | | | | nd | 0.19-0.62 | |
| Haglund 2015 | | | nd-0.087 (92%) | 0.011-0.395 (100%) | nd-0.470 (83%) | 0.017-0.504 (100%) | 0.010-0.141 (100%) |

| | <u>DBE-DBCH (γ-)</u> | <u>DBE-DBCH (δ-)</u> | <u>DDC-CO (anti-)</u> | <u>DDC-CO (syn-)</u> | <u>EHDPP</u> | <u>EH-TBB</u> | <u>HBB</u> |
|----------------|--|--|-----------------------|----------------------|---------------|---------------------|--------------------|
| Marklund 2005 | | | | | | | |
| Green 2008 | | | | | nd-1100 (33%) | | |
| Green 2008 | | | | | nd-260 (29%) | | |
| Moskeland 2009 | | | | | | nd | nd-10.2 |
| Mai 2009 | | | | | | | |
| Arp 2010 | | | | | | | 4.3 |
| Möller 2010 | | | nd-1.01 | nd-4.1 | | | |
| Schlabach 2011 | nd-18 | | 0.065-120 | 0.039-42 | | nd-1.4 | nd-2.3 |
| Xie 2011 | | | | | | | 0.92 (median) |
| Remberger 2014 | | | | | | | nd-0.21 |
| Remberger 2014 | | | | | | | nd-0.25 |
| Haglund 2015 | | | 0.0096-0.064 (100%) | 0.0090-0.068 (100%) | | 0.0032-0.036 (100%) | 0.017-0.091 (100%) |

| | <u>PBEB</u> | <u>PBP</u> | <u>PBT</u> | <u>TBBPA</u> | <u>TBBPA-BAE</u> | <u>TBBPA-BDBPE</u> | <u>TBOEP</u> | <u>TBP-DBPE</u> | <u>TCEP</u> |
|----------------|----------------------|------------|----------------------|--------------|------------------|--------------------|--------------|-----------------|-----------------|
| Marklund 2005 | | | | | | | nd | | 1.6 |
| Green 2008 | | | | | | | nd-340 (50%) | | 510-3700 (100%) |
| Green 2008 | | | | | | | nd-150 (14%) | | nd-270 (14%) |
| Moskeland 2009 | | | nd | | nd | nd | | nd | |
| Mai 2009 | | | | | | 131-1240 | | | |
| Arp 2010 | | | nd | | | | | | |
| Möller 2010 | | | | | | | | | |
| Schlabach 2011 | | nd-1.5 | nd-4.4 | nd-284 | | | | nd-3.2 | |
| Xie 2011 | | | 0.01 (median) | | | | | 0.56 (median) | |
| Remberger 2014 | | | nd | | | | | | |
| Remberger 2014 | | | nd-0.06 | | | | | | |
| Haglund 2015 | 0.0013-0.0080 (100%) | | 0.0019-0.0090 (100%) | | | | | | |

| | <u>TCIPP</u> | <u>TMPP (p-)</u> | <u>TDCIPP</u> | <u>TEHP</u> | <u>TIBP</u> | <u>TMP</u> | <u>TNBP</u> | <u>TPHP</u> | <u>TPP</u> | <u>BCMP-BCEP</u> |
|-----------------------|-----------------|------------------|---------------|-------------|-----------------|------------|-----------------|---------------|------------|------------------|
| Marklund 2005 | 810 | nd | 20 | nd | | 24 a | 280 | 12000 | nd | |
| Green 2008 | 240-3700 (100%) | | nd-72 (33%) | | 320-4400 (100%) | | 300-3700 (100%) | nd-1000 (33%) | | nd-5200 (17%) |
| Green 2008 | nd-330 (14%) | | nd-250 (57%) | | nd-230 (86%) | | nd | nd | | nd |
| Moskeland 2009 | | | | | | | | | | |
| Mai 2009 | | | | | | | | | | |
| Arp 2010 | | | | | | | | | | |
| Möller 2010 | | | | | | | | | | |
| Schlabach 2011 | | | | | | | | | | |
| Xie 2011 | | | | | | | | | | |
| Remberger 2014 | | | | | | | | | | |
| Remberger 2014 | | | | | | | | | | |
| Haglund 2015 | | | | | | | | | | |

Table S6 Concentrations of detected FRs in atmospheric deposition.

| | <u>Brief information, country</u> | <u>Unit</u> | <u>DBDPE</u> | <u>DBE- DBCH (α-)</u> | <u>DBE- DBCH (β-)</u> | <u>DDC- CO (anti-)</u> | <u>DDC- CO (syn-)</u> | <u>HBB</u> | <u>PBEB</u> | <u>PBT</u> |
|-----------------------|--|---|--------------|---------------------------|---------------------------|--------------------------------|-------------------------------|------------|-------------|------------|
| Marklund 2005 | Wet and dry deposition, Background, Finland | ng m ⁻² month ⁻¹ | | | | | | | | |
| Marklund 2005 | Snow from ground, close to road | ng/L | | | | | | | | |
| Marklund 2005 | Snow from ground, close to airport | ng/L | | | | | | | | |
| Marklund 2005 | Snow from ground, background, Sweden | ng/L | | | | | | | | |
| Regnery 2009 | Urban rain, Germany | ng/L | | | | | | | | |
| Regnery 2009 | Background rain, Germany | ng/L | | | | | | | | |
| Regnery 2009 | Background snow, Germany | ng/L | | | | | | | | |
| Newton 2013 | Wet and dry deposition (2009-2010), Abisko, Sweden | ng m ⁻² month ⁻¹ | | | 3.1 ±3.6 | | 22 ±21 | | | |
| Newton 2013 | Wet and dry deposition (2009-2010), Krycklan, Sweden | ng m ⁻² month ⁻¹ | | | 3.5 ±2.8 | | 1.1 ±0.52 | | | |
| Remberger 2014 | Background, Swedish westcoast | ng m ⁻² day ⁻¹ | nd | | nd | | | nd | nd | nd |

| | <u>Unit</u> | <u>TBOEP</u> | <u>TCEP</u> | <u>TCIPP</u> | <u>TMPP (p-)</u> | <u>TDCIPP</u> | <u>TEHP</u> | <u>TIBP</u> | <u>TMP a</u> | <u>TNBP</u> |
|-----------------------|-------------------------|---------------|---------------|-----------------|------------------|---------------|-------------|-----------------|--------------|-----------------|
| Marklund 2005 | ng/m ² month | nd | 16500 | 15300 | nd | nd | nd | | 33 | 6900 |
| Marklund 2005 | ng/L | 4-12 | 7-12 | 110-170 | nd | 8-230 | nd-130 | | nd-10 a | 11-20 |
| Marklund 2005 | ng/L | 7-94 | 29-39 | 100-210 | 260-9900 | 4-15 | 1-95 | | 11-28 a | 2100-23000 |
| Marklund 2005 | ng/L | 2 | 7 | 68 | nd | 29 | nd | | nd a | 19 |
| Regnery 2009 | ng/L | 25 (median) | 73 (median) | 743 (median) | | 7 (median) | | 244 (median) | | 203 (median) |
| Regnery 2009 | ng/L | 3-39 (median) | 3-39 (median) | 30-387 (median) | | 2-24 (median) | | 42-123 (median) | | 37-133 (median) |
| Regnery 2009 | ng/L | 4-21 (median) | 4-21 (median) | 20-83 (median) | | 5-40 (median) | | 39-196 (median) | | 15-192 (median) |
| Newton 2013 | ng/m ² month | | | | | | | | | |
| Newton 2013 | ng/m ² month | | | | | | | | | |
| Remberger 2014 | ng/m ² day | | | | | | | | | |

| | <u>Unit</u> | <u>TPHP</u> | <u>TPP</u> |
|-----------------------|--|-------------|------------|
| Marklund 2005 | ng m ⁻² month ⁻¹ | nd | nd |
| Marklund 2005 | ng/L | 4-68 | nd |
| Marklund 2005 | ng/L | 120-830 | nd-2 |
| Marklund 2005 | ng/L | 4 | nd |
| Regnery 2009 | ng/L | | |
| Regnery 2009 | ng/L | | |
| Regnery 2009 | ng/L | | |
| Newton 2013 | ng m ⁻² month ⁻¹ | | |
| Newton 2013 | ng m ⁻² month ⁻¹ | | |
| Remberger 2014 | ng m ⁻² day ⁻¹ | | |

Table S7 Concentrations of detected FRs in water (ng L⁻¹).

| | <u>Brief information, country</u> | <u>TBP</u> | <u>TBP-AE</u> |
|-------------------------------|---|------------|---------------|
| Andresen 2004 | Rivers, Germany | | |
| Andresen 2004 | WWTP effluent, Germany | | |
| Marklund 2005 | Influent/effluent WWTP, Sweden | | |
| Martinez-Carballo 2007 | River water, Austria | | |
| Green 2008 | Influent/effluent WWTP, Norway | | |
| Moskeland 2009 | Screening in Norway, including point sources | nd | nd |
| Möller 2010 | Sea water, Arctic to Antarctica | | |
| Arp 2010 | Wastewater, seepage water, near suspected sources, Norway | | |
| Regnery 2010 | Urban surface waters, Germany | | |
| Regnery 2010 | Rural surface water, Germany | | |
| Regnery 2010 | Subalpine water, Germany | | |
| Xie 2011 | Seawater from the Atlantic and Southern Ocean | | nd |
| Möller 2011 | Seawater, European Arctic | | |
| Lacorte 2012 | River water, Spain | | |
| Andersson 2013 | WWTP, influent/effluent, Norway | | |
| Cristale 2013a | River Aire, UK | | |
| Cristale 2013b | Rivers, Spain | | |
| Remberger 2014 | Stormwater, diffuse sources, Gothenburg | | |
| Remberger 2014 | WWTP, influent/effluent, Gothenburg/Borås | | |
| Remberger 2014 | Point source, extinguishing water | | |
| Gustavsson 2016 | Swedish river screening for Naturvårdsverket | nd-20* | |

| | <u>BEH-TEBP</u> | <u>BTBPE</u> |
|------------------------|---|--|
| Andresen 2004 | | |
| Andresen 2004 | | |
| Marklund 2005 | | |
| Martinez-Carballo 2007 | | |
| Green 2008 | | |
| Moskeland 2009 | nd | nd-107.0 |
| Möller 2010 | | |
| Arp 2010 | | |
| Regnery 2010 | | |
| Regnery 2010 | | |
| Regnery 2010 | | |
| Xie 2011 | | |
| Möller 2011 | nd-0.0013 (dissolved, 25%), 0-0.00012 (particulate, 6%) | nd (dissolved), 0-0.000002 (particulate, 6%) |
| Cristale 2012 | | |
| Andersson 2013 | | nd |
| Cristale 2013a | | |
| Cristale 2013b | nd | nd |
| Remberger 2014 | | |
| Remberger 2014 | | |
| Remberger 2014 | | |
| Gustavsson 2016 | | nd-4.7 |

| | <u>DBDPE</u> | <u>DBE-DBCH</u> | <u>DBHCTD</u> | <u>DOPP</u> | <u>DDC-CO (anti-)</u> | <u>DDC-CO (syn-)</u> | <u>EBTEBPI</u> | <u>EHDPP</u> | <u>EH-TBB</u> |
|------------------------|------------------|-------------------|---------------|-------------|-----------------------|----------------------|----------------|--------------|---------------|
| Andresen 2004 | | | | | | | | | |
| Andresen 2004 | | | | | | | | | |
| Marklund 2005 | | | | nd-2000 | | | | | |
| Martinez-Carballo 2007 | | | | | | | | | |
| Green 2008 | | | | | | | | 250-710 | |
| Moskeland 2009 | nd-185.7 | | | | | | nd | | nd |
| Möller 2010 | | | | | nd-0.0004 | nd-0.0009 | | | |
| Arp 2010 | | | | | | | | | |
| Regnery 2010 | | | | | | | | | |
| Regnery 2010 | | | | | | | | | |
| Regnery 2010 | | | | | | | | | |
| Xie 2011 | | | nd | | | | | | nd |
| Möller 2011 | | | | | | | | | |
| Cristale 2012 | | | | | | | | | |
| Andersson 2013 | nd-5.1 (average) | 0.6-5.3 (average) | | | | | | | |
| Cristale 2013a | | | | | | | | | |
| Cristale 2013b | nd | | nd | | | | | nd-46 | nd |
| Remberger 2014 | 250-1500 | nd | | | | | | | |
| Remberger 2014 | nd-420 | nd | | | | | | | |
| Remberger 2014 | 330-1800 | nd | | | | | | | |
| Gustavsson 2016 | | | | | | nd-12 | | nd-9.2* | nd-24 |

| | <u>HBB</u> | <u>PBB-Acr</u> | <u>PBEB</u> | <u>PBT</u> | <u>TBBPA</u> | <u>TBBPA-BAE</u> |
|------------------------|---|----------------|-------------|------------|--------------|------------------|
| Andresen 2004 | | | | | | |
| Andresen 2004 | | | | | | |
| Marklund 2005 | | | | | | |
| Martinez-Carballo 2007 | | | | | | |
| Green 2008 | | | | | | |
| Moskeland 2009 | 0.1-19.1 | | nd-1.3 | nd-7.5 | | nd-2.0 |
| Möller 2010 | | | | | | |
| Arp 2010 | 0.40-15.37 | | nd-0.94 | nd-5.63 | | |
| Regnery 2010 | | | | | | |
| Regnery 2010 | | | | | | |
| Regnery 2010 | | | | | | |
| Xie 2011 | nd-0.02 | | | nd | | |
| Möller 2011 | nd-0.000003 (dissolved, 13%), 0-0.000002 (particulate, 19%) | | | nd | | |
| Cristale 2012 | | | | | | |
| Andersson 2013 | | | | | | nd-0.46 |
| Cristale 2013a | nd-0.76 | | nd-0.40 | | | |
| Cristale 2013b | nd | | nd | nd | | |
| Remberger 2014 | 2.2-22 | | nd-91 | nd-2.2 | | |
| Remberger 2014 | nd | | nd-9.9 | nd-2.4 | | |
| Remberger 2014 | 11-1200 | | nd-16 | nd-4.2 | | |
| Gustavsson 2016 | nd-0.13 | nd-2.6 | | nd-2.5 | nd-62 | |

| | <u>TBBPA-BDBPE</u> | <u>TBOEP</u> | <u>TBP-DBPE</u> | <u>TBX</u> | <u>TCBPA</u> | <u>TCEP</u> |
|------------------------|--------------------|----------------|--|------------|--------------|----------------|
| Andresen 2004 | | 10-200 | | | | 13-130 |
| Andresen 2004 | | 500 | | | | 130 |
| Marklund 2005 | | 3100-35000 | | | | 90-450 |
| Martinez-Carballo 2007 | | 24-500 | 0 | | | 13-130 |
| Green 2008 | | 1600-9200 | | | | 1600-2500 |
| Moskeland 2009 | nd-159.6 | | nd | | | |
| Möller 2010 | | | | | | |
| Arp 2010 | | | | | | |
| Regnery 2010 | | nd-53 (median) | | | | 23-61 (median) |
| Regnery 2010 | | nq | | | | 3 (median) |
| Regnery 2010 | | <LOQ-31 (mean) | | | | 6-33 (mean) |
| Xie 2011 | | | nd-0.00005 (median) | | | |
| Möller 2011 | | | nd-0.0003 (dissolved, 81%), nd (particulate) | | | |
| Cristale 2012 | | | | | | 320 |
| Andersson 2013 | nd-18 | | | | | |
| Cristale 2013a | | | | | | 119-316 |
| Cristale 2013b | | nd-4600 | nd | | | nd-330 |
| Remberger 2014 | | | | | | |
| Remberger 2014 | | | | | | |
| Remberger 2014 | | | | | | |
| Gustavsson 2016 | | | | nd-0.022* | nd-56* | nd-14 |

| | <u>TCIPP</u> | <u>TMPP (o-)</u> | <u>TMPP (p-)</u> | <u>TDCIPP</u> | <u>TEBP-Anh</u> | <u>TEHP</u> | <u>TEP</u> | <u>TIBP</u> | <u>TMP</u> |
|-------------------------------|-----------------|------------------|------------------|---------------|-----------------|-------------|------------|----------------|------------|
| Andresen 2004 | 20-200 | | | 13-50 | | | | 30-100 | |
| Andresen 2004 | 400 | | | 120 | | | | | |
| Marklund 2005 | 1100-24000 | | | 130-450 | | nd-130 | | | nd-584 |
| Martinez-Carballo 2007 | 33-170 | nd | | nd-19 | | nd-<LOQ | 13-51 | | |
| Green 2008 | 1700-2900 | | | 86-820 | | | | 210-410 | |
| Moskeland 2009 | | | | | nd | | | | |
| Möller 2010 | | | | | | | | | |
| Arp 2010 | | | | | | | | | |
| Regnery 2010 | 85-126 (median) | | | | | | | 8-10 (median) | |
| Regnery 2010 | 7-18 (median) | | | | | | | nd-9 (median) | |
| Regnery 2010 | 31-312 (mean) | | | | | | | <LOQ-11 (mean) | |
| Xie 2011 | | | | | | | | | |
| Möller 2011 | | | | | | | | | |
| Cristale 2012 | 220 | | | 30 | | | | | |
| Andersson 2013 | | | | | | | | | |
| Cristale 2013a | 113-26050 | | | 62-149 | | | | | |
| Cristale 2013b | nd-1800 | nd-9.2 | | nd-200 | | nd-4 | | nd-1200 | |
| Remberger 2014 | | | | | | | | | |
| Remberger 2014 | | | | | | | | | |
| Remberger 2014 | | | | | | | | | |
| Gustavsson 2016 | nd-30 | nd-1.4* | nd-11 | nd-48 a | nd-67 | nd-48 a | | | |

| | <u>TNBP</u> | <u>TPHP</u> | <u>TPP</u> | <u>TTBNPP</u> |
|-------------------------------|----------------|-------------|------------|---------------|
| Andresen 2004 | 30-120 | | 40 | |
| Andresen 2004 | | | | |
| Marklund 2005 | 360-52000 | 41-290 | | |
| Martinez-Carballo 2007 | 20-110 | nd-10 | | |
| Green 2008 | 160-1800 | 1700-14000 | | |
| Moskeland 2009 | | | | |
| Möller 2010 | | | | |
| Arp 2010 | | | | |
| Regnery 2010 | 17-32 (median) | | | |
| Regnery 2010 | nd-5 (median) | | | |
| Regnery 2010 | <LOQ-7 (mean) | | | |
| Xie 2011 | | | | |
| Möller 2011 | | | | |
| Cristale 2012 | | 20 | | |
| Andersson 2013 | | | | |
| Cristale 2013a | | 6.3-22 | | |
| Cristale 2013b | nd-370 | nd-35 | | |
| Remberger 2014 | | | | |
| Remberger 2014 | | | | |
| Remberger 2014 | | | | |
| Gustavsson 2016 | nd-24 | nd-66* | | nd-3.6* |

Table S8 Concentrations of detected FRs in sediment (ng g⁻¹ dw, unless other is stated).

| | | <u>TBP</u> | <u>2,4-DBP</u> | <u>TBP-AE</u> | <u>BATE</u> | <u>BDPhP</u> | <u>BEH-TEBP</u> |
|-------------------------------|--|-------------------|----------------|---------------|-------------|--------------|-------------------|
| Martinez-Carballo 2007 | River sediment, Austria | | | | | | |
| Green 2008^a | Sediment, point sources, Norway | | | | | | |
| Green 2008^a | Recipient waters, Norway | | | | | | |
| Mai 2009 | Sediment from Pearl river delta, China | | | | | | |
| Evenset 2009 | Svalbard | | | | | | |
| Kolic 2009 | Lake Ontario, US/Canada | | | | | | |
| Moskeland 2009 | Screening in Norway, including point sources | nd-3.3 | | nd | | | nd |
| Guerra 2010 | Llobregat river basin, Spain | | | | | | |
| Arp 2010 | Sediment and seepage sediment, near suspected sources Norway | | | | | | |
| Leonards 2011 | Sediment close to WWTP and remote areas, Norway | | | | | nd | |
| Schlabach 2011 | Sediment, Nordic countries | nd-7.8 | nd-2.9 | nd | nd | | nd-3.3 |
| Klosterhaus 2012 | San Fransisco bay, US | | | | | | nd |
| Andersson 2013 | Sediment near WWTP, Norway | | | | | | |
| Cristale 2013b | Rivers, Spain | | | | | | nd |
| Kaasa 2013 | Terrestrial/freshwater | nd | | | | | 0.11 ±0.03 (100%) |
| Kaasa 2013 | Marine | 2.47 ±0.51 (100%) | | | | | nd |
| Barón 2014 | Sediment, Spain | | | | | | |
| Remberger 2014 | Background, Kosterfjorden, Strömstad | | | | | | |
| Remberger 2014 | Diffuse sources | | | | | | |

^aUnit: µg/kg LOI weight

| | <u>BTBPE</u> | <u>DBPhP</u> | <u>DBDPE</u> | <u>DBE-DBCH (α-)</u> | <u>DBE-DBCH (β-)</u> | <u>DBE-DBCH (γ-)</u> | <u>DBE-DBCH (δ-)</u> | <u>DBHCTD</u> | <u>DDC-DBF</u> |
|--------------------------------|------------------|--------------|-------------------|--|---------------------------------------|--|--|---------------|----------------|
| Martinez-Carballo 2007 | | | | | | | | | |
| Green 2008 ^a | | | | | | | | | |
| Green 2008 ^a | | | | | | | | | |
| Mai 2009 | 0.05-21.9 | | 38.8-364 | | | | | | |
| Evenset 2009 | | | | | | | | | |
| Kolic 2009 | 1.6 | | | | | | | | |
| Moskeland 2009 | nd-4.5 | | nd-1.8 | | | | | | |
| Guerra 2010 | | | 4.8-24 | | | | | | |
| Arp 2010 | | | | | | | | | |
| Leonards 2011 | | nd | | | | | | | |
| Schlabach 2011 | nd-0.25 | | nd-2.4 | nd-0.3 | nd-0.15 | | nd-0.63 | | |
| Klosterhaus 2012 | nd-0.06 (50%) | | nd | | | | | | |
| Andersson 2013 | nd-1.0 (average) | | nd | | | nd | | | |
| Cristale 2013b | nd | | nd-435 | | | | | nd | |
| Kaasa 2013 | | | 2.08 ±0.76 (100%) | | | | | | |
| Kaasa 2013 | | | 0.24 ±0.21 (100%) | | | | | | |
| Barón 2014 | | | nd-31.5 (85%) | | | | | | nd-1.91 (21%) |
| Remberger 2014 | | | nd | nd | | | | | |
| Remberger 2014 | | | nd | nd | | | | | |

^aUnit: $\mu\text{g}/\text{kg}$ LOI weight

| | <u>DDC-Ant</u> | <u>DDC-CO (anti-)</u> | <u>DDC-CO (syn-)</u> | <u>EBTEBPI</u> | <u>EHDPP</u> | <u>EH-TBB</u> | <u>HBB</u> | <u>PBEB</u> | <u>PBP</u> | <u>PBT</u> | <u>TBBPA</u> |
|--------------------------------|----------------|-----------------------|----------------------|----------------|--------------|---------------|------------|--------------|------------|------------|--------------|
| Martinez-Carballo 2007 | | | | | | | | | | | |
| Green 2008 ^a | | | | | 320-1500 | | | | | | |
| Green 2008 ^a | | | | | 140-560 | | | | | | |
| Mai 2009 | | | | | | | | | | | |
| Evenset 2009 | | | | | | | | | | | nd |
| Kolic 2009 | | 120 | 34 | | | | | | | | |
| Moskeland 2009 | | | | nd | | nd | nd-1.8 | nd-0.1 | | nd-0.3 | |
| Guerra 2010 | | | | | | | nd-2.4 | nd-9.6 | | | |
| Arp 2010 | | | | | | | nd-1.33 | nd-0.028 | | nd-0.22 | |
| Leonards 2011 | | | | | nd-15 (60%) | | | | | | |
| Schlabach 2011 | | 0.0049-2.5 | 0.0035-0.99 | | | nd-0.21 | nd-0.019 | nd-0.046 | nd | nd-2.7 | nd-16 |
| Klosterhaus 2012 | | 0.06-0.6 (100%) | 0.03-0.3 (100%) | | | nd | nd | nd-0.1 (50%) | | | |
| Andersson 2013 | | | | | | | | | | | |
| Cristale 2013b | | | | | nd-63 | nd | nd | nd | | nd | |
| Kaasa 2013 | | | | | | | | | nd | | |
| Kaasa 2013 | | | | | | | | | nd | | |
| Barón 2014 | nd-1.04 (52%) | nd-1.50 (97%) | nd-0.73 (94%) | | | | nd | nd | | | |
| Remberger 2014 | | | | | | | nd | nd | | nd | |
| Remberger 2014 | | | | | | | nd-0.52 | nd-0.19 | | nd-0.02 | |

^aUnit: $\mu\text{g}/\text{kg}$ LOI weight

| | <u>TBBPA-BAE</u> | <u>TBBPA-BDBPE</u> | <u>TBOEP</u> | <u>TBP-DBPE</u> | <u>TCBPA</u> | <u>TCEP</u> | <u>TCIPP</u> | <u>TMPP (o-)</u> | <u>TMPP (m-)</u> | <u>TMPP (p-)</u> |
|--------------------------------|-------------------|--------------------|-----------------|-----------------|--------------|--------------|--------------|------------------|------------------|------------------|
| Martinez-Carballo 2007 | | | 2.4-130 | | | nd-160 | <LOQ-1300 | nd-39 | | |
| Green 2008 ^a | | | 540-2900 | | | 27-5500 | 490-24000 | | | |
| Green 2008 ^a | | | nd-3100 | | | nd-1600 | 63-16000 | | | |
| Mai 2009 | | nd-2300 | | | | | | | | |
| Evenset 2009 | | | | | | | | | | |
| Kolic 2009 | | | | | | | | | | |
| Moskeland 2009 | nd-2.4 | nd | | nd | | | | | | |
| Guerra 2010 | | | | | | | | | | |
| Arp 2010 | | | | | | | | | | |
| Leonards 2011 | | | 0.69-100 (100%) | | | nd-8.5 (85%) | nd-54 (70%) | nd-1.5 (25%) | | nd-288 (80%) |
| Schlabach 2011 | | | | nd | | | | | | |
| Klosterhaus 2012 | | | | | | | | | | |
| Andersson 2013 | nd-0.81 (average) | nd | | | | | | | | |
| Cristale 2013b | | | nd | nd | | nd-9.7 | nd-365 | nd-84 | | |
| Kaasa 2013 | | | | | | | | | | |
| Kaasa 2013 | | | | | | | | | | |
| Barón 2014 | | | | | | | | | | |
| Remberger 2014 | | | | | | | | | | |
| Remberger 2014 | | | | | | | | | | |

^aUnit: $\mu\text{g}/\text{kg}$ LOI weight

| | <u>TDCIPP</u> | <u>TEBP-Anh</u> | <u>TEHP</u> | <u>TEP</u> | <u>TBP</u> | <u>TNBP</u> | <u>TPHP</u> | <u>BCMP-BCEP</u> |
|--------------------------------|---------------|-----------------|-------------|------------|--------------|--------------|--------------|------------------|
| Martinez-Carballo 2007 | nd-<LOQ | | nd-140 | <LOQ-81 | | <LOQ-50 | nd-160 | |
| Green 2008 ^a | nd-8800 | | | | nd-1100 | 210-4300 | 900-1500 | nd-2800 |
| Green 2008 ^a | 63-870 | | | | 62-470 | 67-480 | nd-370 | nd |
| Mai 2009 | | | | | | | | |
| Evenset 2009 | | | | | | | | |
| Kolic 2009 | | | | | | | | |
| Moskeland 2009 | | nd | | | | | | |
| Guerra 2010 | | | | | | | | |
| Arp 2010 | | | | | | | | |
| Leonards 2011 | nd-1.0 (40%) | | nd-46 (60%) | | nd-2.5 (80%) | nd-6.7 (65%) | nd-6.8 (75%) | nd |
| Schlabach 2011 | | | | | | | | |
| Klosterhaus 2012 | | | | | | | | |
| Andersson 2013 | | | | | | | | |
| Cristale 2013b | nd-8.7 | | nd-290 | | nd-8.4 | nd-13 | nd-23 | |
| Kaasa 2013 | | | | | | | | |
| Kaasa 2013 | | | | | | | | |
| Barón 2014 | | | | | | | | |
| Remberger 2014 | | | | | | | | |
| Remberger 2014 | | | | | | | | |

^aUnit: $\mu\text{g}/\text{kg}$ LOI weight

Table S9 Concentrations of detected FRs in sludge (ng g⁻¹ dw, unless other is stated).

| | <u>Brief description, country</u> | <u>TBP</u> | <u>2,4-DBP</u> | <u>TBP-AE</u> | <u>BATE</u> | <u>BDPhP</u> | <u>BEH-TEBP</u> |
|-------------------------------|---|------------|----------------|---------------|-------------|--------------|-----------------|
| Marklund 2005 | Sewage sludge, Sweden | | | | | | |
| Chu 2005 | WWTP sludge, Canada | | | | | | |
| Green 2008^a | WWTP sludge, Norway | | | | | | |
| Moskeland 2009 | Screening in Norway, sewage sludge | nd | | nd | | | nd |
| Mai 2009 | Sewage sludge from Pearl river delta, China | | | | | | |
| Arp 2010 | Wastewater sludge, near suspected sources, Norway | | | | | | |
| Schlabach 2011 | Sludge, Nordic countries | nd-101 | nd-40 | nd-27 | nd-4.1 | | nd-42 |
| Andersson 2013 | WWTP sludge, Norway | | | | | | |
| Remberger 2014 | WWTP sludge, Gothenburg/Borås | | | | | | |
| Barón 2014 | Sewage sludge, Spain | | | | | | |

^aUnit: µg/kg LOI weight

| | <u>BTBPE</u> | <u>DBPhP</u> | <u>DBDPE</u> | <u>DBE-DBCH (α-)</u> | <u>DBE-DBCH (β-)</u> | <u>DBE-DBCH (γ-)</u> | <u>DBE-DBCH (δ-)</u> | <u>DBHCTD</u> | <u>DDC-DBF</u> |
|-------------------------------|-------------------|--------------|-------------------|----------------------|----------------------|----------------------|----------------------|---------------|----------------|
| Marklund 2005 | | | | | | | | | |
| Chu 2005 | | | | | | | | | |
| Green 2008^a | | | | | | | | | |
| Moskeland 2009 | nd-2.08 | | nd-8.7 | | | | | | |
| Mai 2009 | 0.31-1.66 | | 266-1995 | | | | | | |
| Arp 2010 | | | | | | | | | |
| Schlabach 2011 | nd-3.9 | | nd-160 | nd-4.7 | nd-2.6 | nd-1.7 | | | |
| Andersson 2013 | 0.7-1.7 (average) | | 1.9-6.3 (average) | | 0.6-1.4 (average) | | | | |
| Remberger 2014 | | | 63-190 | nd | | | | | |
| Barón 2014 | | | <LOQ-124 | | | | | | nd-0.24 |

^aUnit: µg/kg LOI weight

| | <u>DDC-Ant</u> | <u>DDC-CO (anti-)</u> | <u>DDC-CO (syn-)</u> | <u>EBTEBPI</u> | <u>EHDPP</u> | <u>EH-TBB</u> | <u>HBB</u> | <u>PBEB</u> | <u>PBP</u> | <u>PBT</u> | <u>TBBPA</u> |
|--------------------------------|----------------|-----------------------|----------------------|----------------|--------------|---------------|------------|-------------|------------|------------|--------------|
| Marklund 2005 | | | | | 320-4600 | | | | | | |
| Chu 2005 | | | | | | | | | | | 2.1-28.3 |
| Green 2008 ^a | | | | | 462-1200 | | | | | | |
| Moskeland 2009 | | | | nd | | nd | 0.12-0.6 | nd | | nd | |
| Mai 2009 | | | | | | | | | | | |
| Arp 2010 | | | | | | | 0.34-0.39 | nd | | nd | |
| Schlabach 2011 | | 0.05-25 | nd-14 | | | nd-2.6 | nd-0.72 | nd-0.13 | nd-3.5 | nd-5.2 | nd-59 |
| Andersson 2013 | | | | | | | | | | | |
| Remberger 2014 | | | | | | | 0.49-1.6 | 0.43-0.64 | | 0.14-0.6 | |
| Barón 2014 | nd-0.60 | <LOQ-11.9 | 0.85-11.2 | | | | | | | | |

^aUnit: µg/kg LOI weight

| | <u>TBBPA-BAE</u> | <u>TBBPA-BDBPE</u> | <u>TBOEP</u> | <u>TBP-DBPE</u> | <u>TCBPA</u> | <u>TCEP</u> | <u>TCIPP</u> | <u>TMPP (o-)</u> | <u>TMPP (m-)</u> | <u>TMPP (p-)</u> |
|--------------------------------|------------------|--------------------|--------------|-----------------|--------------|-------------|--------------|------------------|------------------|------------------|
| Marklund 2005 | | | nd-1900 | | | 6.6-110 | 61-1900 | | | |
| Chu 2005 | | | | | 0.14-0.54 | | | | | |
| Green 2008 ^a | | | 1200-2200 | | | nd | 650-944 | | | |
| Moskeland 2009 | nd | nd | | nd | | | | | | |
| Mai 2009 | | 238-8946 | | | | | | | | |
| Arp 2010 | | | | | | | | | | |
| Schlabach 2011 | | | | nd-120 | | | | | | |
| Andersson 2013 | nd | nd | | | | | | | | |
| Remberger 2014 | | | | | | | | | | |
| Barón 2014 | | | | | | | | | | |

^aUnit: µg/kg LOI weight

| | <u>TDCIPP</u> | <u>TEBP-Anh</u> | <u>TEHP</u> | <u>TEP</u> | <u>TBP</u> | <u>TNBP</u> | <u>TPHP</u> | <u>BCMP-BCEP</u> |
|--------------------------------|---------------|-----------------|-------------|------------|------------|-------------|-------------|------------------|
| Marklund 2005 | 3.3-260 | | | | 27-2700 | 39-850 | 52-320 | |
| Chu 2005 | | | | | | | | |
| Green 2008 ^a | 110-330 | | | | 52-81 | 69-270 | 13-1100 | nd |
| Moskeland 2009 | | nd | | | | | | |
| Mai 2009 | | | | | | | | |
| Arp 2010 | | | | | | | | |
| Schlabach 2011 | | | | | | | | |
| Andersson 2013 | | | | | | | | |
| Remberger 2014 | | | | | | | | |
| Barón 2014 | | | | | | | | |

^aUnit: $\mu\text{g}/\text{kg}$ LOI weight

Table S10 Concentrations of detected FRs in soil (ng g⁻¹ dw). Detection frequency in brackets.

| | <u>Brief information, country</u> | <u>TBP</u> | <u>BEH-TEBP</u> | <u>BTBPE</u> | <u>DBDPE</u> | <u>PBP</u> | <u>TBBPA-BDBPE</u> |
|-------------------|---|------------|-----------------|--------------|--------------|------------|--------------------|
| Mai 2009 | Farmland soil, China | | | 0.02-0.11 | 17.6-35.8 | | 17.3-60.4 |
| Mai 2009 | Farmland soil from e-waste processing area, China | | | 0.07-6.19 | nd-4.56 | | nd |
| Kaasa 2013 | Soil, Norway | nd | 1.04 (100%) | | nd | nd | |

Table S11 Concentrations of detected FRs in plants.

| | <u>Brief description, country</u> | <u>Unit</u> | <u>TBP</u> | <u>2,4-DBP</u> | <u>TBP-AE</u> | <u>BATE</u> | <u>BEH-TEBP</u> | <u>BTBPE</u> | <u>DBDPE</u> | <u>DBE-DBCH (α-)</u> |
|-----------------------|--|-------------|------------|----------------|---------------|-------------|-----------------|--------------|--------------|--|
| Moskeland 2009 | Moss, Energy recycling, Norway | ng/g dw | nd | | nd | | nd | nd | nd-0.1 | |
| Moskeland 2009 | Needles, Energy recycling, Norway | ng/g ww | nd | | nd | | nd | nd | nd-0.1 | |
| Arp 2010 | Pine needles, near suspected sources, Norway | ng/g ww | | | | | | | | |
| Arp 2010 | Moss, near suspected sources, Norway | ng/g dw | | | | | | | | |
| Schlabach 2011 | Moss, near incineration, Faroe islands | ng/g ww | nd-0.46 | nd-0.53 | nd | nd | nd-0.039 | 0.056-0.15 | 0.14-0.34 | 0.0029-0.0046 |

| | <u>Unit</u> | <u>DBE-DBCH (γ-)</u> | <u>DBE-DBCH (δ-)</u> | <u>DDC-CO (anti-)</u> | <u>DDC-CO (syn-)</u> | <u>EBTEBP I</u> | <u>EH-TBB</u> | <u>HBB</u> | <u>PBEB</u> | <u>PBT</u> | <u>TBBP A</u> |
|-----------------------|-------------|--|--|-----------------------|----------------------|-----------------|---------------|--------------|---------------|---------------|---------------|
| Moskeland 2009 | ng/g dw | | | | | nd | nd | nd-0.1 | nd | nd | |
| Moskeland 2009 | ng/g ww | | | | | nd | nd | nd-0.05 | nd | nd | |
| Arp 2010 | ng/g ww | | | | | | | 0.05 | nd | nd | |
| Arp 2010 | ng/g dw | | | | | | | < LOQ | nd | nd | |
| Schlabach 2011 | ng/g ww | | nd | 0.04-0.12 | 0.02-0.05 | | nd | 0.0076-0.011 | 0.0038-0.0059 | 0.0031-0.0032 | nd |

| | <u>Unit</u> | <u>TBBPA-BAE</u> | <u>TBBPA-BDBPE</u> | <u>TBP-DBPE</u> | <u>TEBP-Anh</u> |
|-----------------------|-------------|------------------|--------------------|-----------------|-----------------|
| Moskeland 2009 | ng/g dw | nd | nd | nd | nd |
| Moskeland 2009 | ng/g ww | nd | nd-0.16 | nd | nd |
| Arp 2010 | ng/g ww | | | | |
| Arp 2010 | ng/g dw | | | | |
| Schlabach 2011 | ng/g ww | | | nd-0.0039 | |

Table S12 Concentrations of detected FRs in mammals. Detection frequency in brackets.

| | <u>Brief description, country</u> | <u>Unit</u> | <u>TBP</u> | <u>2,4-DBP</u> | <u>TBP-AE</u> | <u>BATE</u> |
|-------------------------|---|-------------|--------------------|----------------|---------------|-------------|
| McKinney 2010 | Adipose from polar bear, Western Hudson Bay, Canada | ng/g lw | | | | |
| Sagerup 2010 | Arctic fox liver, Norwegian Arctic | ng/g ww | nd | | nd | |
| Sagerup 2010 | Polar bear plasma, Norwegian Arctic | ng/g ww | nd | | nd | |
| Kaasa 2013 | Moose liver, Norway | ng/g ww | 80.7 ±44.6 (89%) | | | |
| Kaasa 2013 | Mouse liver, Norway | ng/g ww | 53.6 ±43.4 (88%) | | | |
| Kaasa 2013 | Shrew liver, Norway | ng/g ww | 27.1 ±6.9 (100%) | | | |
| Kaasa 2013 | Polar bear plasma, Norwegian Arctic | ng/mL | 25.7 ±14.7 (100%) | | | |
| Sagerup 2010 | Ringed seal liver, Norwegian Arctic | ng/g ww | 0.050 ±0.023 (50%) | | nd | |
| Klosterhaus 2012 | Harbor seal blubber Adults, San Frasisco Bay, US | ng/g lw | | | | |
| Klosterhaus 2012 | Harbor seal blubber Pups, San Frasisco Bay, US | ng/g lw | | | | |
| Kaasa 2013 | Harbor seal liver | ng/g ww | 164 ±84 (100%) | | | |
| Kaasa 2013 | Ringed seal plasma, arctic | ng/mL | 31.2 ±32.3 (100%) | | | |

| | <u>Unit</u> | <u>BdPhP</u> | <u>BEH-TEBP</u> | <u>BTBPE</u> | <u>Chlordene Plus</u> | <u>SCCPs</u> | <u>MCCPs</u> |
|-------------------------|-------------|--------------|--------------------|--------------|-----------------------|-------------------|------------------|
| McKinney 2010 | ng/g lw | | | | | | |
| Sagerup 2010 | ng/g ww | | nd | | nd | | |
| Sagerup 2010 | ng/g ww | | nd | | nd | | |
| Kaasa 2013 | ng/g ww | | nd | | | | |
| Kaasa 2013 | ng/g ww | | nd | | | | |
| Kaasa 2013 | ng/g ww | | nd | | | | |
| Kaasa 2013 | ng/mL | | 0.15 ±0.16 (95%) | | | 3.99 ±2.91 (95%) | 2.20 ±1.84 (95%) |
| Sagerup 2010 | ng/g ww | | 0.573 ±0.198 (60%) | | nd | | |
| Klosterhaus 2012 | ng/g lw | | | nd | | | |
| Klosterhaus 2012 | ng/g lw | | | nd | | | |
| Kaasa 2013 | ng/g ww | | 0.10 (10%) | | | | |
| Kaasa 2013 | ng/mL | | 0.04 (10%) | | | 4.96 ±2.70 (100%) | 2.91 ±2.39 (90%) |

| | <u>Unit</u> | <u>DBDPE</u> | <u>DBE-DBCH (α-)</u> | <u>DBE-DBCH (β-)</u> | <u>DBE-DBCH (γ-)</u> | <u>DBE-DBCH (δ-)</u> | <u>DBPhP</u> | <u>DDC-DBF</u> |
|------------------|-------------|------------------------|--|---------------------------------------|--|--|--------------|----------------|
| McKinney 2010 | ng/g lw | nd-2.0 | | | | | | |
| Sagerup 2010 | ng/g ww | nd | | | | | | |
| Sagerup 2010 | ng/g ww | nd | | | | | | |
| Kaasa 2013 | ng/g ww | 0.40 \pm 0.09 (100%) | | | | | | |
| Kaasa 2013 | ng/g ww | 11.9 \pm 5.7 (100%) | | | | | | |
| Kaasa 2013 | ng/g ww | 25.5 \pm 9.1 (100%) | | | | | | |
| Kaasa 2013 | ng/mL | 6.98 \pm 9.11 (100%) | | | | | | |
| Sagerup 2010 | ng/g ww | nd | | | | | | |
| Klosterhaus 2012 | ng/g lw | | | | | | | |
| Klosterhaus 2012 | ng/g lw | | | | | | | |
| Kaasa 2013 | ng/g ww | 12.9 \pm 6.8 (100%) | | | | | | |
| Kaasa 2013 | ng/mL | 5.36 \pm 1.94 (100%) | | | | | | |

| | <u>Unit</u> | <u>Dec 604B</u> | <u>DDC-CO (anti-)</u> | <u>DDC-CO (syn-)</u> | <u>DPhBP</u> | <u>EBTEBPI</u> | <u>EHDPP</u> |
|------------------|-------------|-----------------|-----------------------|----------------------|--------------|----------------|--------------|
| McKinney 2010 | ng/g lw | | | | | | |
| Sagerup 2010 | ng/g ww | | | | | | |
| Sagerup 2010 | ng/g ww | | | | | | |
| Kaasa 2013 | ng/g ww | | | | | | |
| Kaasa 2013 | ng/g ww | | | | | | |
| Kaasa 2013 | ng/g ww | | | | | | |
| Kaasa 2013 | ng/mL | | | | | | |
| Sagerup 2010 | ng/g ww | | | | | | |
| Klosterhaus 2012 | ng/g lw | | 0.06-3.3 (100%) | 0.08-3.8 (100%) | | | |
| Klosterhaus 2012 | ng/g lw | | nd-0.06 (8%) | nd-0.07 (42%) | | | |
| Kaasa 2013 | ng/g ww | | | | | | |
| Kaasa 2013 | ng/mL | | | | | | |

| | <u>Unit</u> | <u>EH-TBB</u> | <u>HBB</u> | <u>OBTMPI</u> | <u>PBEB</u> | <u>PBP</u> | <u>PBT</u> | <u>T2CPP</u> | <u>TBBPA</u> | <u>TBBPA-BAE</u> |
|-------------------------|-------------|---------------------|------------|---------------|-----------------|------------|------------|--------------|--------------|------------------|
| McKinney 2010 | ng/g lw | | nd-3.4 | | nd-1.7 | | | | | |
| Sagerup 2010 | ng/g ww | 0.975 ±0.608 (90%) | nd | | nd | | nd | | | |
| Sagerup 2010 | ng/g ww | 3.460 ±2.481 (90%) | nd | | nd | | nd | | | |
| Kaasa 2013 | ng/g ww | | | | | nd | | | | |
| Kaasa 2013 | ng/g ww | | | | | nd | | | | |
| Kaasa 2013 | ng/g ww | | | | | nd | | | | |
| Kaasa 2013 | ng/mL | | | | | nd | | | | |
| Sagerup 2010 | ng/g ww | 0.435 ±0.292 (100%) | nd | | nd | | nd | | | |
| Klosterhaus 2012 | ng/g lw | nd | nd | | 0.07-0.5 (100%) | | | | | |
| Klosterhaus 2012 | ng/g lw | nd | nd | | nd-0.2 (83%) | | | | | |
| Kaasa 2013 | ng/g ww | | | | | nd | | | | |
| Kaasa 2013 | ng/mL | | | | | nd | | | | |

| | <u>Unit</u> | <u>TBBPA-BDBPE</u> | <u>TBOEP</u> | <u>TBP-DBPE</u> | <u>TCEP</u> | <u>TCIPP</u> | <u>TMPP (o-)</u> | <u>TMPP (p-)</u> | <u>TDCIPP</u> |
|-------------------------|-------------|--------------------|--------------|-----------------|-------------|--------------|------------------|------------------|---------------|
| McKinney 2010 | ng/g lw | | | | | | | | |
| Sagerup 2010 | ng/g ww | nd | | nd | | | | | |
| Sagerup 2010 | ng/g ww | nd | | nd | | | | | |
| Kaasa 2013 | ng/g ww | | | | | | | | |
| Kaasa 2013 | ng/g ww | | | | | | | | |
| Kaasa 2013 | ng/g ww | | | | | | | | |
| Kaasa 2013 | ng/mL | | | | | | | | |
| Sagerup 2010 | ng/g ww | nd | | nd | | | | | |
| Klosterhaus 2012 | ng/g lw | | | | | | | | |
| Klosterhaus 2012 | ng/g lw | | | | | | | | |
| Kaasa 2013 | ng/g ww | | | | | | | | |
| Kaasa 2013 | ng/mL | | | | | | | | |

| | <u>Unit</u> | <u>TEBP-Anh</u> | <u>TEHP</u> | <u>TEP</u> | <u>TIBP</u> | <u>TNBP</u> | <u>TPHP</u> | <u>BCMP-BCEP</u> |
|-------------------------|-------------|-----------------|-------------|------------|-------------|-------------|-------------|------------------|
| McKinney 2010 | ng/g lw | | | | | | | |
| Sagerup 2010 | ng/g ww | nd | | | | | | |
| Sagerup 2010 | ng/g ww | nd | | | | | | |
| Kaasa 2013 | ng/g ww | | | | | | | |
| Kaasa 2013 | ng/g ww | | | | | | | |
| Kaasa 2013 | ng/g ww | | | | | | | |
| Kaasa 2013 | ng/mL | | | | | | | |
| Sagerup 2010 | ng/g ww | nd | | | | | | |
| Klosterhaus 2012 | ng/g lw | | | | | | | |
| Klosterhaus 2012 | ng/g lw | | | | | | | |
| Kaasa 2013 | ng/g ww | | | | | | | |
| Kaasa 2013 | ng/mL | | | | | | | |

Table S13 Concentrations of detected FRs in aquatic species. Detection frequency in brackets.

| | <u>Brief description, country</u> | <u>Unit</u> | <u>TBP</u> | <u>2,4-DBP</u> | <u>TBP-AE</u> | <u>BATE</u> |
|-------------------------|--|-------------|-------------------|----------------|---------------|-------------|
| Arp 2010 | Fish liver, crab, mussel, near suspected sources, Norway | ng/g ww | | | | |
| Green 2008 | Mussels, Norway | ng/g ww | | | | |
| Green 2008 | Cod liver, Norway | ng/g ww | | | | |
| Evenset 2009 | Fish, Svalbard | ng/g ww | | | | |
| Sundkvist 2009 | Freshwater perch, background, Sweden | ng/g lw | | | | |
| Sundkvist 2009 | Freshwater perch/carp, close to sources, Sweden | ng/g lw | | | | |
| Sundkvist 2009 | Marine herring, Sweden | ng/g lw | | | | |
| Sundkvist 2009 | Marine perch, Sweden | ng/g lw | | | | |
| Sundkvist 2009 | Marine mussels, Sweden | ng/g lw | | | | |
| Sundkvist 2009 | Marine eelpout, Sweden | ng/g lw | | | | |
| Sundkvist 2009 | Marine salmon, Sweden | ng/g lw | | | | |
| Mai 2009 | Fish from e-waste processing area, China | ug/g lw | | | | |
| Moskeland 2009 | Fish liver, Norway | ng/g ww | nd-55.8 | | nd | |
| Moskeland 2009 | Blue mussel, Norway | ng/g ww | nd-1.1 | | nd | |
| Moskeland 2009 | Crab, Norway | ng/g ww | 2.4-130.5 | | nd | |
| Leonards 2011 | Cod liver, Norway | ng/g ww | | | | |
| Leonards 2011 | Trout, Norway | ng/g ww | | | | |
| Leonards 2011 | Beach crab, Norway | ng/g ww | | | | |
| Leonards 2011 | Blue mussel, Norway | ng/g ww | | | | |
| Leonards 2011 | Burbot liver, Norway | ng/g ww | | | | |
| Sagerup 2010 | Whole Capelin, Norwegian Arctic | ng/g ww | nd | | nd | |
| Schlabach 2011 | Fish and mussel, Nordic countries | ng/g ww | nd-86 | nd-6.4 | nd | nd-0.00072 |
| Klosterhaus 2012 | White croaker, San Frasco Bay, US | ug/g lw | | | | |
| Klosterhaus 2012 | Shiner surfperch, San Frasco Bay, US | ug/g lw | | | | |
| Kaasa 2013 | Perch liver | ng/g ww | 42.4 ±16.1 (67%) | | | |
| Kaasa 2013 | Brown trout liver | ng/g ww | 66.2 ±39.6 (40%) | | | |
| Kaasa 2013 | Atlantic cod liver | ng/g ww | 68.8 ±35.8 (60%) | | | |
| Kaasa 2013 | Mussels | ng/g ww | 2.53 ±0.22 (100%) | | | |
| Kaasa 2013 | Atlantic cod liver, arctic | ng/g ww | 115 ±61 (70%) | | | |
| Kaasa 2013 | Polar cod, arctic | ng/g ww | nd | | | |
| Houde 2014 | Yellow perch, Canada | ug/g lw | | | | |
| Houde 2014 | Northern pike, Canada | ug/g lw | | | | |
| Houde 2014 | Muskellunge, Canada | ug/g lw | | | | |
| McGoldrick 2014 | Fish, Canada | ng/g ww | | | | |

| | | | | | | |
|-------------------------|--|---------|--|--|--|--|
| Remberger 2014 | Herring muscle, background Swedish westcoast | ng/g ww | | | | |
| Remberger 2014 | Netted dogwhelk, diffuse sources Swedish westcoast | ng/g ww | | | | |
| Malarvannan 2015 | Yellow eel, Belgium | ng/g lw | | | | |

| | <u>Unit</u> | <u>BdPhP</u> | <u>BEH-TEBP</u> | <u>BTBPE</u> | <u>Chlordene Plus</u> | <u>SCCPs</u> | <u>MCCPs</u> |
|-------------------------|-------------|--------------|--------------------|--------------|-----------------------|-------------------|--------------|
| Arp 2010 | ng/g ww | | | | | | |
| Green 2008 | ng/g ww | | | | | | |
| Green 2008 | ng/g ww | | | | | | |
| Evenset 2009 | ng/g ww | | | | | | |
| Sundkvist 2009 | ng/g lw | nd | | | | | |
| Sundkvist 2009 | ng/g lw | nd-2000 | | | | | |
| Sundkvist 2009 | ng/g lw | nd | | | | | |
| Sundkvist 2009 | ng/g lw | nd | | | | | |
| Sundkvist 2009 | ng/g lw | nd-0.5 | | | | | |
| Sundkvist 2009 | ng/g lw | <LOQ | | | | | |
| Sundkvist 2009 | ng/g lw | nd | | | | | |
| Mai 2009 | ug/g lw | | | nd-0.00015 | | | |
| Moskeland 2009 | ng/g ww | | nd | nd | | | |
| Moskeland 2009 | ng/g ww | | nd | nd | | | |
| Moskeland 2009 | ng/g ww | | nd | nd | | | |
| Leonards 2011 | ng/g ww | nd | | | | | |
| Leonards 2011 | ng/g ww | nd | | | | | |
| Leonards 2011 | ng/g ww | nd | | | | | |
| Leonards 2011 | ng/g ww | nd | | | | | |
| Leonards 2011 | ng/g ww | nd | | | | | |
| Sagerup 2010 | ng/g ww | | 0.719 ±0.292 (90%) | | nd | | |
| Schlabach 2011 | ng/g ww | | nd-0.46 | nd-0.2 | | | |
| Klosterhaus 2012 | ug/g lw | | | nd | | | |
| Klosterhaus 2012 | ug/g lw | | | nd | | | |
| Kaasa 2013 | ng/g ww | | nd | | | | |
| Kaasa 2013 | ng/g ww | | 0.04 ±0.01 (30%) | | | | |
| Kaasa 2013 | ng/g ww | | 0.14 ±0.02 (30%) | | | | |
| Kaasa 2013 | ng/g ww | | nd | | | | |
| Kaasa 2013 | ng/g ww | | 0.07 (10%) | | | 10.3 ±10.7 (100%) | 0.94 (10%) |
| Kaasa 2013 | ng/g ww | | nd | | | 2.28 (100%) | 1.51 (100%) |
| Houde 2014 | ug/g lw | | nd | nd | nd | | |

| | | | | | | | |
|------------------|---------|--|----------------------|----|----------------------|--|--|
| Houde 2014 | ug/g lw | | 0.0054 ±0.0017 (64%) | nd | 0.0012 ±0.0002 (91%) | | |
| Houde 2014 | ug/g lw | | nd-0.013 (40%) | nd | 0.0087 ±0.0073 (80%) | | |
| McGoldrick 2014 | ng/g ww | | | | | | |
| Remberger 2014 | ng/g ww | | | | | | |
| Remberger 2014 | ng/g ww | | | | | | |
| Malarvannan 2015 | ng/g lw | | | | | | |

| | <u>Unit</u> | <u>DBDPE</u> | <u>DBE-DBCH (α-)</u> | <u>DBE-DBCH (β-)</u> | <u>DBE-DBCH (γ-)</u> | <u>DBE-DBCH (δ-)</u> | <u>DBPhP</u> | <u>DDC-DBF</u> |
|------------------|-------------|-------------------|--|---------------------------------------|--|--|--------------|----------------|
| Arp 2010 | ng/g ww | | | | | | | |
| Green 2008 | ng/g ww | | | | | | | |
| Green 2008 | ng/g ww | | | | | | | |
| Evenset 2009 | ng/g ww | | | | | | nd | |
| Sundkvist 2009 | ng/g lw | | | | | | nd | |
| Sundkvist 2009 | ng/g lw | | | | | | nd-3300 | |
| Sundkvist 2009 | ng/g lw | | | | | | nd | |
| Sundkvist 2009 | ng/g lw | | | | | | nd | |
| Sundkvist 2009 | ng/g lw | | | | | | nd-1.2 | |
| Sundkvist 2009 | ng/g lw | | | | | | nd | |
| Sundkvist 2009 | ng/g lw | | | | | | nd | |
| Mai 2009 | ug/g lw | nd | | | | | | |
| Moskeland 2009 | ng/g ww | nd | | | | | | |
| Moskeland 2009 | ng/g ww | nd | | | | | | |
| Moskeland 2009 | ng/g ww | nd | | | | | | |
| Leonards 2011 | ng/g ww | | | | | | nd | |
| Leonards 2011 | ng/g ww | | | | | | nd | |
| Leonards 2011 | ng/g ww | | | | | | nd | |
| Leonards 2011 | ng/g ww | | | | | | nd | |
| Leonards 2011 | ng/g ww | | | | | | nd | |
| Sagerup 2010 | ng/g ww | nd | | | | | | |
| Schlabach 2011 | ng/g ww | nd-0.092 | nd-0.26 | nd-0.25 | nd-1.1 | | | |
| Klosterhaus 2012 | ug/g lw | nd | | | | | | |
| Klosterhaus 2012 | ug/g lw | | | | | | | |
| Kaasa 2013 | ng/g ww | 2.47 ±0.30 (100%) | | | | | | |
| Kaasa 2013 | ng/g ww | 11.1 ±8.57 (100%) | | | | | | |
| Kaasa 2013 | ng/g ww | 4.29 ±0.70 (100%) | | | | | | |
| Kaasa 2013 | ng/g ww | 0.29 ±0.10 (100%) | | | | | | |

| | | | | | | | |
|-------------------------|---------|------------------|---------|--|--|--|----------------------|
| Kaasa 2013 | ng/g ww | 5.57 ±1.38 (90%) | | | | | |
| Kaasa 2013 | ng/g ww | 0.42 (100%) | | | | | |
| Houde 2014 | ug/g lw | nd | | | | | nd |
| Houde 2014 | ug/g lw | 0.0267 (9%) | | | | | 0.0026 ±0.0010 (55%) |
| Houde 2014 | ug/g lw | nd | | | | | 0.0237 ±0.0201 (50%) |
| McGoldrick 2014 | ng/g ww | | | | | | |
| Remberger 2014 | ng/g ww | nd | nd-0.14 | | | | |
| Remberger 2014 | ng/g ww | nd | nd | | | | |
| Malarvannan 2015 | ng/g lw | | | | | | |

| | <u>Unit</u> | <u>Dec 604B</u> | <u>DDC-CO (anti-)</u> | <u>DDC-CO (syn-)</u> | <u>DPhBP</u> | <u>EBTEBPI</u> | <u>EHDPP</u> |
|-------------------------|-------------|-----------------|-----------------------|----------------------|--------------|----------------|--------------|
| Arp 2010 | ng/g ww | | | | | | |
| Green 2008 | ng/g ww | | | | | | nd |
| Green 2008 | ng/g ww | | | | | | nd |
| Evenset 2009 | ng/g ww | | | | nd | | nd-52 |
| Sundkvist 2009 | ng/g lw | | | | | | 8.9-150 |
| Sundkvist 2009 | ng/g lw | | | | | | 160-190 |
| Sundkvist 2009 | ng/g lw | | | | | | 3.0-7.5 |
| Sundkvist 2009 | ng/g lw | | | | | | 37-78 |
| Sundkvist 2009 | ng/g lw | | | | | | 14-16 |
| Sundkvist 2009 | ng/g lw | | | | | | 14000 |
| Sundkvist 2009 | ng/g lw | | | | | | 1.5 |
| Mai 2009 | ug/g lw | | | | | | |
| Moskeland 2009 | ng/g ww | | | | | nd | |
| Moskeland 2009 | ng/g ww | | | | | nd | |
| Moskeland 2009 | ng/g ww | | | | | nd | |
| Leonards 2011 | ng/g ww | | | | | | nd |
| Leonards 2011 | ng/g ww | | | | | | nd-1.1 (13%) |
| Leonards 2011 | ng/g ww | | | | | | nd-0.41 (8%) |
| Leonards 2011 | ng/g ww | | | | | | nd-0.3 (33%) |
| Leonards 2011 | ng/g ww | | | | | | nd |
| Sagerup 2010 | ng/g ww | | | | | | |
| Schlabach 2011 | ng/g ww | | nd-0.026 | nd-0.023 | | | |
| Klosterhaus 2012 | ug/g lw | | nd-0.0018 (83%) | nd | | | |
| Klosterhaus 2012 | ug/g lw | | nd-0.0037 (75%) | nd | | | |
| Kaasa 2013 | ng/g ww | | | | | | |

| | | | | | | | |
|-------------------------|---------|----------------------|----------------------|----------------------|--|--|--|
| Kaasa 2013 | ng/g ww | | | | | | |
| Kaasa 2013 | ng/g ww | | | | | | |
| Kaasa 2013 | ng/g ww | | | | | | |
| Kaasa 2013 | ng/g ww | | | | | | |
| Kaasa 2013 | ng/g ww | | | | | | |
| Houde 2014 | ug/g lw | nd | nd | nd | | | |
| Houde 2014 | ug/g lw | 0.0059 ±0.0025 (82%) | nd-0.0028 (45%) | nd-0.0091 (36%) | | | |
| Houde 2014 | ug/g lw | 0.139 ±0.130 (80%) | 0.0018 ±0.0011 (80%) | 0.0044 ±0.0025 (90%) | | | |
| McGoldrick 2014 | ng/g ww | | | | | | |
| Remberger 2014 | ng/g ww | | | | | | |
| Remberger 2014 | ng/g ww | | | | | | |
| Malarvannan 2015 | ng/g lw | | | | | | |

| | <u>Unit</u> | <u>EH-TBB</u> | <u>HBB</u> | <u>OBTMPI</u> | <u>PBEB</u> | <u>PBP</u> | <u>PBT</u> | <u>T2CPP</u> | <u>TBBPA</u> | <u>TBBPA-BAE</u> |
|-----------------------|-------------|---------------------|------------|---------------|-------------|------------|------------|--------------|--------------|------------------|
| Arp 2010 | ng/g ww | | nd | | nd | | nd | | | |
| Green 2008 | ng/g ww | | | | | | | | | |
| Green 2008 | ng/g ww | | | | | | | | | |
| Evenset 2009 | ng/g ww | | | | | | | nd-8.9 | nd | |
| Sundkvist 2009 | ng/g lw | | | | | | | | | |
| Sundkvist 2009 | ng/g lw | | | | | | | | | |
| Sundkvist 2009 | ng/g lw | | | | | | | | | |
| Sundkvist 2009 | ng/g lw | | | | | | | | | |
| Sundkvist 2009 | ng/g lw | | | | | | | | | |
| Sundkvist 2009 | ng/g lw | | | | | | | | | |
| Sundkvist 2009 | ng/g lw | | | | | | | | | |
| Mai 2009 | ug/g lw | | | | | | | | | |
| Moskeland 2009 | ng/g ww | nd | nd | | nd | | nd | | | nd |
| Moskeland 2009 | ng/g ww | nd | nd | | nd | | nd | | | nd |
| Moskeland 2009 | ng/g ww | nd | nd | | nd | | nd | | | nd |
| Leonards 2011 | ng/g ww | | | | | | | | | |
| Leonards 2011 | ng/g ww | | | | | | | | | |
| Leonards 2011 | ng/g ww | | | | | | | | | |
| Leonards 2011 | ng/g ww | | | | | | | | | |
| Leonards 2011 | ng/g ww | | | | | | | | | |
| Sagerup 2010 | ng/g ww | 0.378 ±0.240 (100%) | nd | | nd | | nd | | | |

| | | | | | | | | | |
|-------------------------|---------|---------|---------------|----|-----------|----|--------------|--|----|
| Schlabach 2011 | ng/g ww | nd-0.12 | 0.0058-0.047 | | nd-0.0044 | nd | 0.0015-0.021 | | nd |
| Klosterhaus 2012 | ug/g lw | nd | nd | | nd | | | | |
| Klosterhaus 2012 | ug/g lw | nd | nd | | nd | | | | |
| Kaasa 2013 | ng/g ww | | | | | nd | | | |
| Kaasa 2013 | ng/g ww | | | | | nd | | | |
| Kaasa 2013 | ng/g ww | | | | | nd | | | |
| Kaasa 2013 | ng/g ww | | | | | nd | | | |
| Kaasa 2013 | ng/g ww | | | | | nd | | | |
| Kaasa 2013 | ng/g ww | | | | | nd | | | |
| Houde 2014 | ug/g lw | nd | nd | | nd | | | | |
| Houde 2014 | ug/g lw | nd | 0.0032 (9%) | | | | | | |
| Houde 2014 | ug/g lw | nd | 0.0014-0.0039 | nd | | | | | |
| McGoldrick 2014 | ng/g ww | | | | | | | | |
| Remberger 2014 | ng/g ww | | 0.15-0.56 | | 0.38-0.67 | | nd | | |
| Remberger 2014 | ng/g ww | | nd-4.6 | | 0.65-3.9 | | nd | | |
| Malarvannan 2015 | ng/g lw | | | | | | | | |

| | <u>Unit</u> | <u>TBBPA-BDBPE</u> | <u>TBOEP</u> | <u>TBP-DBPE</u> | <u>TCEP</u> | <u>TCIPP</u> | <u>TMPP (o-)</u> | <u>TMPP (p-)</u> | <u>TDCIPP</u> |
|-----------------------|-------------|--------------------|--------------|-----------------|-------------|--------------|------------------|------------------|---------------|
| Arp 2010 | ng/g ww | | | | | | | | |
| Green 2008 | ng/g ww | | nd | | nd-23 | nd | | | nd |
| Green 2008 | ng/g ww | | nd | | nd | nd | | | nd |
| Evenset 2009 | ng/g ww | | nd | | nd-26 | | nd | nd | nd-8.1 |
| Sundkvist 2009 | ng/g lw | | nd | | nd-83 | 220-750 | nd | nd-43 | nd |
| Sundkvist 2009 | ng/g lw | | 240-1000 | | 39-160 | 170-770 | nd-2.5 | 22-137 | 49-140 |
| Sundkvist 2009 | ng/g lw | | nd | | 2.0-3.4 | 42-150 | nd | nd | nd |
| Sundkvist 2009 | ng/g lw | | nd | | 43-69 | 140-250 | nd | 20-23 | nd |
| Sundkvist 2009 | ng/g lw | | nd | | nd-55 | 130-1300 | nd | 11-110 | nd |
| Sundkvist 2009 | ng/g lw | | nd | | 59 | 310 | nd | 19 | nd |
| Sundkvist 2009 | ng/g lw | | nd | | 1.5 | 23 | nd | 1.8 | nd |
| Mai 2009 | ug/g lw | nd | | | | | | | |
| Moskeland 2009 | ng/g ww | nd | | nd | | | | | |
| Moskeland 2009 | ng/g ww | nd | | nd | | | | | |
| Moskeland 2009 | ng/g ww | nd | | nd | | | | | |
| Leonards 2011 | ng/g ww | | nd | | nd | nd | nd | nd | nd |

| | | | | | | | | | |
|-------------------------|---------|----|-------------|----------|---------------|---------------|----|----|--------------|
| Leonards 2011 | ng/g ww | | nd | | nd-0.21 (13%) | nd-2 (7%) | nd | nd | < 0.88 (13%) |
| Leonards 2011 | ng/g ww | | nd | | nd-1.9 (25%) | nd-8.9 (17%) | nd | nd | nd |
| Leonards 2011 | ng/g ww | | nd | | nd-0.11 (20%) | nd-0.81 (20%) | nd | nd | nd |
| Leonards 2011 | ng/g ww | | nd-411 (7%) | | nd-8.6 (7%) | nd-17 (7%) | nd | nd | nd |
| Sagerup 2010 | ng/g ww | nd | | nd | | | | | |
| Schlabach 2011 | ng/g ww | | | nd-0.049 | | | | | |
| Klosterhaus 2012 | ug/g lw | | | | | | | | |
| Klosterhaus 2012 | ug/g lw | | | | | | | | |
| Kaasa 2013 | ng/g ww | | | | | | | | |
| Kaasa 2013 | ng/g ww | | | | | | | | |
| Kaasa 2013 | ng/g ww | | | | | | | | |
| Kaasa 2013 | ng/g ww | | | | | | | | |
| Kaasa 2013 | ng/g ww | | | | | | | | |
| Kaasa 2013 | ng/g ww | | | | | | | | |
| Houde 2014 | ug/g lw | | | | | | | | |
| Houde 2014 | ug/g lw | | | | | | | | |
| Houde 2014 | ug/g lw | | | | | | | | |
| McGoldrick 2014 | ng/g ww | | nd-9.8 | | nd-3.4 | nd-<LOQ | | | nd-<LOQ |
| Remberger 2014 | ng/g ww | | | | | | | | |
| Remberger 2014 | ng/g ww | | | | | | | | |
| Malarvannan 2015 | ng/g lw | | | | | | | | |

| | <u>Unit</u> | <u>TEBP-Anh</u> | <u>TEHP</u> | <u>TEP</u> | <u>TIBP</u> | <u>TNBP</u> | <u>TPHP</u> | <u>BCMP-BCEP</u> | <u>ΣPFR</u> |
|-----------------------|-------------|-----------------|-------------|------------|-------------|-------------|-------------|------------------|-------------|
| Arp 2010 | ng/g ww | | | | | | | | |
| Green 2008 | ng/g ww | | | | nd | nd-17 | nd | nd | |
| Green 2008 | ng/g ww | | | | nd | nd | nd | nd | |
| Evenset 2009 | ng/g ww | | nd-4.6 | | nd-4.9 | nd-11.0 | 0.3-13 | | |
| Sundkvist 2009 | ng/g lw | | | | | 12-36 | 21-180 | | |
| Sundkvist 2009 | ng/g lw | | | | | 34-4900 | 100-170 | | |
| Sundkvist 2009 | ng/g lw | | | | | 3.1-7.9 | 7.1-34 | | |
| Sundkvist 2009 | ng/g lw | | | | | 16-23 | 64-81 | | |
| Sundkvist 2009 | ng/g lw | | | | | 14-20 | 18-93 | | |
| Sundkvist 2009 | ng/g lw | | | | | 120 | 400 | | |
| Sundkvist 2009 | ng/g lw | | | | | 1.6 | 4.2 | | |

| | | | | | | | | | |
|-------------------------|---------|----|---------------|--|----|---------|--------------|----|--------------|
| Mai 2009 | ug/g lw | | | | | | | | |
| Moskeland 2009 | ng/g ww | nd | | | | | | | |
| Moskeland 2009 | ng/g ww | nd | | | | | | | |
| Moskeland 2009 | ng/g ww | nd | | | | | | | |
| Leonards 2011 | ng/g ww | | nd | | nd | nd | nd-9.9 (3%) | nd | |
| Leonards 2011 | ng/g ww | | nd-0.02 (7%) | | nd | nd | nd-44 (67%) | nd | |
| Leonards 2011 | ng/g ww | | nd-0.09 (17%) | | nd | nd | nd-0.31 (8%) | nd | |
| Leonards 2011 | ng/g ww | | nd-0.25 (33%) | | nd | nd | nd | nd | |
| Leonards 2011 | ng/g ww | | nd | | nd | nd | nd | nd | |
| Sagerup 2010 | ng/g ww | nd | | | | | | | |
| Schlabach 2011 | ng/g ww | | | | | | | | |
| Klosterhaus 2012 | ug/g lw | | | | | | | | |
| Klosterhaus 2012 | ug/g lw | | | | | | | | |
| Kaasa 2013 | ng/g ww | | | | | | | | |
| Kaasa 2013 | ng/g ww | | | | | | | | |
| Kaasa 2013 | ng/g ww | | | | | | | | |
| Kaasa 2013 | ng/g ww | | | | | | | | |
| Kaasa 2013 | ng/g ww | | | | | | | | |
| Kaasa 2013 | ng/g ww | | | | | | | | |
| Houde 2014 | ug/g lw | | | | | | | | |
| Houde 2014 | ug/g lw | | | | | | | | |
| Houde 2014 | ug/g lw | | | | | | | | |
| McGoldrick 2014 | ng/g ww | | | | | nd-<LOQ | nd-<LOQ | | |
| Remberger 2014 | ng/g ww | | | | | | | | |
| Remberger 2014 | ng/g ww | | | | | | | | |
| Malarvannan 2015 | ng/g lw | | | | | | | | 0.0071-0.329 |

Table S14 Concentrations of detected FRs in birds and bird eggs. Detection frequency in brackets.

| | <u>Brief description, country</u> | <u>Unit</u> | <u>TBP</u> | <u>2,4-DBP</u> | <u>TBP-AE</u> | <u>BATE</u> |
|-------------------------|---|-------------|--------------------|----------------|---------------|-------------|
| Evenset 2009 | Seabirds, Svalbard | ng/g dw | | | | |
| Mai 2009 | Bird from e-waste processing area, China | ng/g lw | | | | |
| Leonards 2011 | Bird blood/plasma, Norway | ng/g ww | | | | |
| Leonards 2011 | Bird egg, Norway | ng/g ww | | | | |
| Sagerup 2010 | Common Eider liver, Norwegian Arctic | ng/g ww | 0.090 ±0.095 (90%) | | nd | |
| Sagerup 2010 | Brünnich's guillemot egg, Norwegian Arctic | ng/g ww | nd | | nd | |
| Sagerup 2010 | Kittiwake liver, Norwegian Arctic | ng/g ww | nd | | nd | |
| Schlabach 2011 | Eggs, Sweden/Faroe islands | ng/g ww | 0.20-1.4 | nd | nd | nd |
| Klosterhaus 2012 | Double-crested cormorant egg, San Fransisco Bay, US | ng/g lw | | | | |
| Kaasa 2013 | Herring gull egg | ng/g ww | 62.5 ±64.8 (80%) | | | |
| Kaasa 2013 | Common Eider egg | ng/g ww | 66.2 ±74.2 (90%) | | | |
| Kaasa 2013 | Glaucous gull plasma, arctic | ng/mL | 30.8 ±9.0 (100%) | | | |
| Kaasa 2013 | Kittiwake egg, arctic | ng/g ww | 52.6 ±18.8 (83%) | | | |
| Kaasa 2013 | Common eider egg, arctic | ng/g ww | 37.8 ±17.6 (58%) | | | |

| | <u>Unit</u> | <u>BdPhP</u> | <u>BEH-TEBP</u> | <u>BTBPE</u> | <u>Chlordene Plus</u> | <u>SCCPs</u> | <u>MCCPs</u> |
|-------------------------|-------------|--------------|--------------------|--------------|-----------------------|------------------|-------------------|
| Evenset 2009 | ng/g dw | | | | | | |
| Mai 2009 | ng/g lw | | | 0.07-2.41 | | | |
| Leonards 2011 | ng/g ww | nd | | | | | |
| Leonards 2011 | ng/g ww | nd | | | | | |
| Sagerup 2010 | ng/g ww | | 1.652 ±1.396 (60%) | | nd | | |
| Sagerup 2010 | ng/g ww | | 1.799 ±1.358 (70%) | | 0.664 ±0.367 (40%) | | |
| Sagerup 2010 | ng/g ww | | 0.800 ±0.356 (50%) | | nd | | |
| Schlabach 2011 | ng/g ww | | nd-0.021 | 0.019-0.042 | | | |
| Klosterhaus 2012 | ng/g lw | | nd | nd | | | |
| Kaasa 2013 | ng/g ww | | 1.99 ±2.65 (20%) | | | | |
| Kaasa 2013 | ng/g ww | | 0.04 ±0.02 (100%) | | | | |
| Kaasa 2013 | ng/mL | | 0.026 ±0.001 (17%) | | | 3.95 ±1.99 (75%) | 8.87 ±9.88 (67%) |
| Kaasa 2013 | ng/g ww | | 0.10 ±0.09 (100%) | | | 7.83 ±8.26 (67%) | 4.91 ±4.88 (100%) |
| Kaasa 2013 | ng/g ww | | 0.06 ±0.07 (58%) | | | 3.23 ±1.77 (83%) | 4.24 ±4.07 (100%) |

| | <u>Unit</u> | <u>DBDPE</u> | <u>DBE-DBCH (α-)</u> | <u>DBE-DBCH (β-)</u> | <u>DBE-DBCH (γ-)</u> | <u>DBE-DBCH (δ-)</u> | <u>DBPhP</u> | <u>DDC-DBF</u> |
|-------------------------|-------------|------------------------|--|---------------------------------------|--|--|--------------|----------------|
| Evenset 2009 | ng/g dw | | | | | | nd-0.33 | |
| Mai 2009 | ng/g lw | 9.6-124 | | | | | | |
| Leonards 2011 | ng/g ww | | | | | | nd | |
| Leonards 2011 | ng/g ww | | | | | | nd-5.6 (4%) | |
| Sagerup 2010 | ng/g ww | nd | | | | | | |
| Sagerup 2010 | ng/g ww | 0.581 (10%) | | | | | | |
| Sagerup 2010 | ng/g ww | nd | | | | | | |
| Schlabach 2011 | ng/g ww | nd-0.12 | nd-0.022 | 0.059-0.33 | | nd-0.31 | | |
| Klosterhaus 2012 | ng/g lw | | | | | | | |
| Kaasa 2013 | ng/g ww | 0.44 \pm 0.20 (100%) | | | | | | |
| Kaasa 2013 | ng/g ww | 0.33 \pm 0.11 (100%) | | | | | | |
| Kaasa 2013 | ng/mL | 6.43 \pm 2.62 (100%) | | | | | | |
| Kaasa 2013 | ng/g ww | 1.01 \pm 1.55 (100%) | | | | | | |
| Kaasa 2013 | ng/g ww | 0.82 \pm 0.62 (100%) | | | | | | |

| | <u>Unit</u> | <u>Dec 604B</u> | <u>DDC-CO (anti-)</u> | <u>DDC-CO (svn-)</u> | <u>DPhBP</u> | <u>EBTEBPI</u> | <u>EHDPP</u> |
|-------------------------|-------------|-----------------|-----------------------|----------------------|--------------|----------------|--------------|
| Evenset 2009 | ng/g dw | | | | nd | | 6.0-28 |
| Mai 2009 | ng/g lw | | | | | | |
| Leonards 2011 | ng/g ww | | | | | | nd |
| Leonards 2011 | ng/g ww | | | | | | nd-3.1 (7%) |
| Sagerup 2010 | ng/g ww | | | | | | |
| Sagerup 2010 | ng/g ww | | | | | | |
| Sagerup 2010 | ng/g ww | | | | | | |
| Schlabach 2011 | ng/g ww | | 0.012-0.057 | nd-0.026 | | | |
| Klosterhaus 2012 | ng/g lw | | nd | 0.9-1.1 (100%) | | | |
| Kaasa 2013 | ng/g ww | | | | | | |
| Kaasa 2013 | ng/g ww | | | | | | |
| Kaasa 2013 | ng/mL | | | | | | |
| Kaasa 2013 | ng/g ww | | | | | | |
| Kaasa 2013 | ng/g ww | | | | | | |

| | <u>Unit</u> | <u>EH-TBB</u> | <u>HBB</u> | <u>OBTMPI</u> | <u>PBEB</u> | <u>PBP</u> | <u>PBT</u> | <u>T2CPP</u> | <u>TBBPA</u> | <u>TBBPA-BAE</u> |
|------------------|-------------|---------------------|------------|---------------|-------------|------------|---------------|--------------|--------------|------------------|
| Evenset 2009 | ng/g dw | | | | | | | nd-2.6 | | |
| Mai 2009 | ng/g lw | | | | | | | | | |
| Leonards 2011 | ng/g ww | | | | | | | | | |
| Leonards 2011 | ng/g ww | | | | | | | | | |
| Sagerup 2010 | ng/g ww | 0.862 ±1.243 (100%) | nd | | nd | | nd | | | |
| Sagerup 2010 | ng/g ww | 1.213 ±0.984 (100%) | nd | | nd | | nd | | | |
| Sagerup 2010 | ng/g ww | 0.732 ±0.261 (90%) | nd | | nd | | nd | | | |
| Schlabach 2011 | ng/g ww | nd-0.18 | 0.023-0.03 | | nd-0.0014 | 0.12-0.43 | 0.0054-0.0063 | | nd | |
| Klosterhaus 2012 | ng/g lw | nd | nd | | nd | | | | | |
| Kaasa 2013 | ng/g ww | | | | | nd | | | | |
| Kaasa 2013 | ng/g ww | | | | | nd | | | | |
| Kaasa 2013 | ng/mL | | | | | nd | | | | |
| Kaasa 2013 | ng/g ww | | | | | nd | | | | |
| Kaasa 2013 | ng/g ww | | | | | nd | | | | |

| | <u>Unit</u> | <u>TBBPA-BDBPE</u> | <u>TBOEP</u> | <u>TBP-DBPE</u> | <u>TCEP</u> | <u>TCIPP</u> | <u>TMPP (o-)</u> | <u>TMPP (p-)</u> | <u>TDCIPP</u> |
|------------------|-------------|--------------------|--------------|-----------------|--------------|--------------|------------------|------------------|---------------|
| Evenset 2009 | ng/g dw | | nd | | nd-4.7 | | nd | nd | nd |
| Mai 2009 | ng/g lw | nd | | | | | | | |
| Leonards 2011 | ng/g ww | | nd-57 (16%) | | nd-6.1 (10%) | nd-10 (24%) | nd | nd | nd-1.9 (6%) |
| Leonards 2011 | ng/g ww | | nd | | nd-6.0 (11%) | nd-5.0 (6%) | nd | nd | nd-0.16 (5%) |
| Sagerup 2010 | ng/g ww | nd | | nd | | | | | |
| Sagerup 2010 | ng/g ww | nd | | nd | | | | | |
| Sagerup 2010 | ng/g ww | nd | | nd | | | | | |
| Schlabach 2011 | ng/g ww | | | nd | | | | | |
| Klosterhaus 2012 | ng/g lw | | | | | | | | |
| Kaasa 2013 | ng/g ww | | | | | | | | |
| Kaasa 2013 | ng/g ww | | | | | | | | |
| Kaasa 2013 | ng/mL | | | | | | | | |
| Kaasa 2013 | ng/g ww | | | | | | | | |
| Kaasa 2013 | ng/g ww | | | | | | | | |

| | <u>Unit</u> | <u>TEBP-Anh</u> | <u>TEHP</u> | <u>TEP</u> | <u>TIBP</u> | <u>TNBP</u> | <u>TPHP</u> | <u>BCMP-BCEP</u> | <u>ΣPFR</u> |
|-------------------------|-------------|-----------------|-------------|------------|-------------|-------------|--------------|------------------|-------------|
| Evenset 2009 | ng/g dw | | nd-4.6 | | nd-2.6 | nd-6.8 | 0.6-3.3 | | |
| Mai 2009 | ng/g lw | | | | | | | | |
| Leonards 2011 | ng/g ww | | nd | | nd | nd | nd-2.2 (14%) | nd | |
| Leonards 2011 | ng/g ww | | nd-8.7 (9%) | | nd | nd | nd-14 (31%) | nd | |
| Sagerup 2010 | ng/g ww | nd | | | | | | | |
| Sagerup 2010 | ng/g ww | nd | | | | | | | |
| Sagerup 2010 | ng/g ww | nd | | | | | | | |
| Schlabach 2011 | ng/g ww | | | | | | | | |
| Klosterhaus 2012 | ng/g lw | | | | | | | | |
| Kaasa 2013 | ng/g ww | | | | | | | | |
| Kaasa 2013 | ng/g ww | | | | | | | | |
| Kaasa 2013 | ng/mL | | | | | | | | |
| Kaasa 2013 | ng/g ww | | | | | | | | |
| Kaasa 2013 | ng/g ww | | | | | | | | |

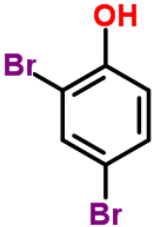
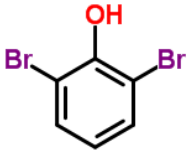
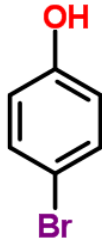
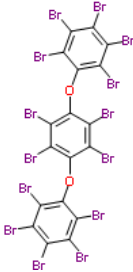
Table S15 Concentrations of detected FRs in humans. Detection frequency in brackets.

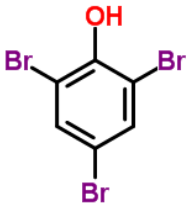
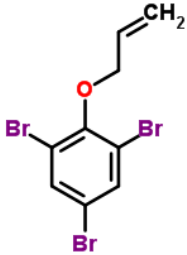
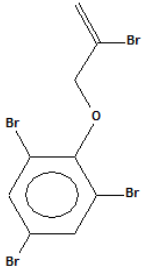
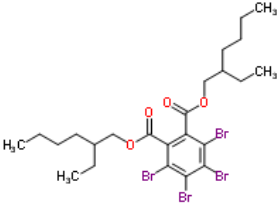
| | | <u>Unit</u> | <u>2/3-BP</u> | <u>TBP</u> | <u>2,4-DBP</u> | <u>2,6-DBP</u> | <u>4-BP</u> | <u>BdPhP</u> |
|-----------------------|--------------------|-------------|---------------|--------------------|----------------|----------------|-------------|--------------|
| Sundkvist 2009 | Human milk, Sweden | ug/g lw | | | | | | |
| Remberger 2014 | Serum, Sweden | ng/g | | | | | | |
| Haglund 2015 | Serum, Sweden | ng/g | 0.034-0.185 | 0.049-0.269 (100%) | 0.016-0.076 | 0.020-0.047 | 0.031-0.155 | nd |

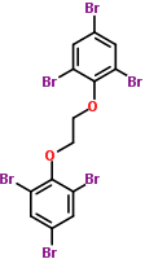
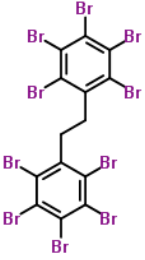
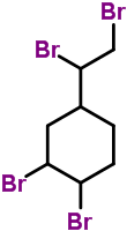
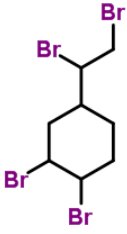
| | <u>Unit</u> | <u>BTBPE</u> | <u>DBE-DBCH (α-)</u> | <u>DBE-DBCH (β-)</u> | <u>DBPhP</u> | <u>EHDPP</u> | <u>EH-TBB</u> | <u>HBB</u> | <u>PBEB</u> | <u>PBT</u> | <u>TBOEP</u> | <u>TCEP</u> |
|-----------------------|-------------|---------------------|--|---------------------------------------|--------------|--------------|-----------------|------------|-----------------|------------|--------------|-------------|
| Sundkvist 2009 | ug/g lw | | | | | 3.5-13 | | | | | nd-63 | 2.1-8.2 |
| Remberger 2014 | ng/g | | nd | | nd | | | nd | nd-0.072 | nd | | |
| Haglund 2015 | ng/g | 0.0034-0.779 (100%) | | | | | nd-0.0055 (35%) | | nd-0.0014 (41%) | | | |

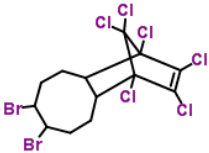
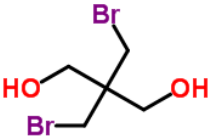
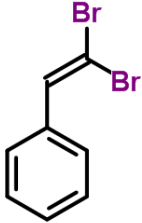
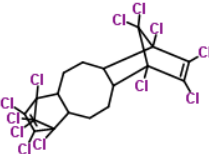
| | <u>Unit</u> | <u>TCIPP</u> | <u>TMPP (<i>o</i>-)</u> | <u>TMPP (<i>p</i>-)</u> | <u>TDCIPP</u> | <u>TNBP</u> | <u>TPHP</u> |
|-----------------------|-------------|--------------|-------------------------|-------------------------|---------------|-------------|-------------|
| Sundkvist 2009 | ug/g lw | 22-82 | nd | nd-3.7 | 1.6-5.3 | 11-57 | 3.2-11 |
| Remberger 2014 | ng/g | | | | | | |
| Haglund 2015 | ng/g | | | | | | |

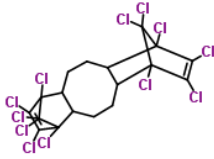
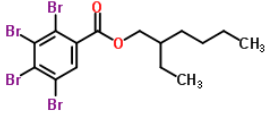
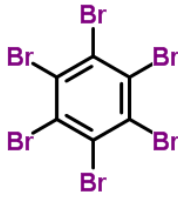
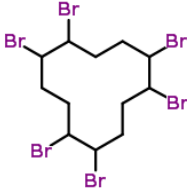
Table S16 Structure, molecular formula, CAS no., retention time and physico-chemical properties of all HFRs ($n = 46$) included in this study. Table is reused with some modifications from Gustavsson et al. (2017) [14].^a

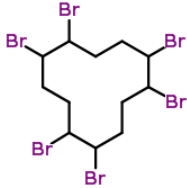
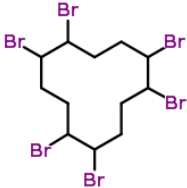
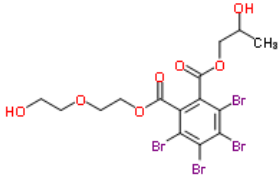
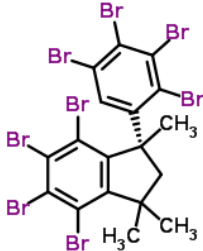
| Abbreviation | Name | Structure | Molecular formula | CAS no. | MW ^b | t_R ^c | Bp ^d | S_w ^e | $\log K_{OW}$ ^f | $\log K_{OC}$ ^g | $\log K_{OA}$ ^h | V_P ⁱ | H^j | pK_a ^k |
|----------------|--|---|---|------------|-----------------|--------------------|-------------------|--------------------|----------------------------|----------------------------|----------------------------|--------------------|-------------|---------------------|
| 2,4-DBP | 2,4-Dibromophenol |  | C ₆ H ₄ Br ₂ O | 615-58-7 | 251.9 | 3.69 | 269.7 | 3536 | 3.29 | 2.87 | 8.66 | 2.1 | 8.9E- 8 | 7.86 |
| 2,6-DBP | 2,6-Dibromophenol |  | C ₆ H ₄ Br ₂ O | 608-33-3 | 251.9 | 4.04 | 269.7 | 3536 | 3.29 | 2.95 | 8.80 | 0.57 | 8.9E- 8 | na |
| 4-BP | 4-Bromophenol |  | C ₆ H ₅ BrO | 106-41-2 | 173.0 | - | 223.0 | 1430 8 | 2.40 | 2.53 | 7.80 | 1.2 | 1.5E- 7 | na |
| 4'-PeBPOBDE208 | Pentabromophenoxy-nonabromo-diphenyl ether |  | C ₁₈ Br ₁₄ O ₂ | 58965-66-5 | 1367 | - | 808.7 | 1.4E- 6 | 16.9 | 10.38 | 26.5 | 4.9E- 18 | 6.5E- 12 | na |

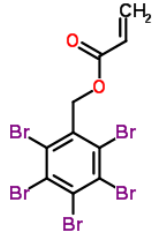
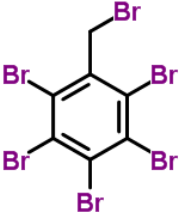
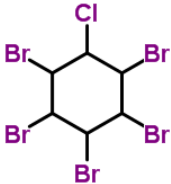
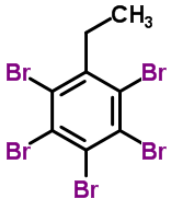
| Abbreviation | Name | Structure | Molecular formula | CAS no. | MW ^b | t_R^c | Bp^d | S_W^e | $\log K_{OW}^f$ | $\log K_{OC}^g$ | $\log K_{OA}^h$ | V_P^i | H^j | pK_a^k |
|--------------|---|---|--|------------|-----------------|---------|--------|--------------------|-------------------|-----------------|-----------------|--------------------|--------------------|-------------|
| TBP | 2,4,6-Tribromophenol |  | C ₆ H ₃ Br ₃ O | 118-79-6 | 330.8 | 6.18 | 310.1 | 788 | 4.13 ¹ | 3.38 | 10.0 | 0.040 | 3.6E- ₈ | 6.32 ± 0.23 |
| TBP-AE | Allyl 2,4,6-tribromophenyl ether |  | C ₉ H ₇ Br ₃ O | 221-913-2 | 370.9 | 6.89 | 323.2 | 1.3 | 5.59 | 4.07 | 8.6 | 0.014 | 2.7E- ₅ | na |
| BATE | 2-Bromoallyl 2,4,6-tribromophenyl ether |  | C ₉ H ₆ Br ₄ O | na | 449.8 | 8.35 | 359.1 | 0.59 | 5.98 | 4.29 | 9.7 | 9.8E- ₄ | 5.3E- ₆ | na |
| BEH-TEBP | Bis(2-ethyl-1-hexyl)tetrabromophthalate |  | C ₂₄ H ₃₄ Br ₄ O ₄ | 26040-51-7 | 706.1 | 13.56 | 539.8 | 1.9E- ₆ | 11.95 | 7.40 | 16.9 | 2.3E- ₉ | 3.0E- ₇ | na |

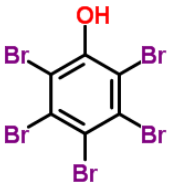
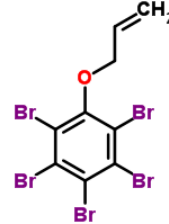
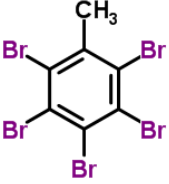
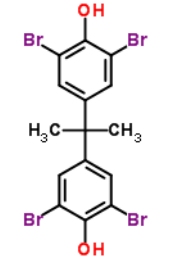
| Abbreviation | Name | Structure | Molecular formula | CAS no. | MW ^b | <i>t</i> _R ^c | <i>Bp</i> ^d | <i>S</i> _W ^e | log <i>K</i> _{OW} ^f | log <i>K</i> _{OC} ^g | log <i>K</i> _{OA} ^h | <i>V</i> _P ⁱ | <i>H</i> ^j | <i>pK</i> _a ^k |
|--------------------|--|---|---|------------|-----------------|------------------------------------|------------------------|------------------------------------|---|---|---|------------------------------------|-----------------------|-------------------------------------|
| BTBPE | 1,2-Bis(2,4,6-tribromophenoxy) ethane |  | C ₁₄ H ₈ Br ₆ O ₂ | 37853-59-1 | 687.6 | 13.2 | 502.2 | 2.2E-4 | 9.15 | 6.10 | 15.7 | 3.2E-8 | 7.3E-9 | na |
| DBDPE | 1,2-Bis(2,3,4,5,6-pentabromophenyl) ethane |  | C ₁₄ H ₄ Br ₁₀ | 84852-53-9 | 971.2 | 17.47 | 600.9 | 9.7E-7 | 13.64 | 11.84 | 19.2 | 2.5E-11 | 6.4E-8 | na |
| α-DBE-DBCH (TBECH) | 1,2-Dibromo-4-(1,2-dibromoethyl) cyclohexane |  | C ₈ H ₁₂ Br ₄ | 3322-93-8 | 427.8 | 8.16 | 336.8 | 0.92 | 5.24 | 4.55 | 8.0 | 0.014 | 4.2E-5 | na |
| β-DBE-DBCH (TBECH) | 1,2-Dibromo-4-(1,2-dibromoethyl) cyclohexane |  | C ₈ H ₁₂ Br ₄ | 3322-93-8 | 427.8 | 8.16 | 336.8 | 0.92 | 5.24 | 4.55 | 8.0 | 0.014 | 4.2E-5 | na |

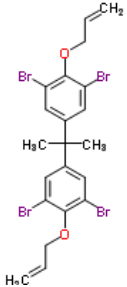
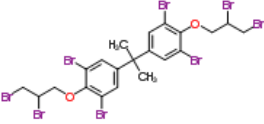
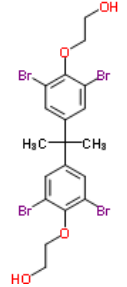
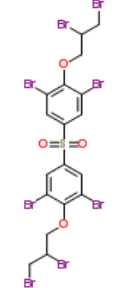
| Abbreviation | Name | Structure | Molecular formula | CAS no. | MW ^b | t_R ^c | Bp ^d | S_W ^e | $\log K_{OW}$ ^f | $\log K_{OC}$ ^g | $\log K_{OA}$ ^h | V_P ⁱ | H^j | pK_a ^k |
|--------------|---|---|---|------------|-----------------|--------------------|-------------------|--------------------|----------------------------|----------------------------|----------------------------|--------------------|--------|---------------------|
| DBHCTD | Hexachlorocyclopentadienyl dibromocyclooctane |  | C ₁₃ H ₁₂ Br ₂ Cl ₆ | 51936-55-1 | 540.8 | 11.31 | 414.8 | 1.4E-4 | 7.91 | 6.87 | 11.1 | 1.4E-5 | 1.8E-5 | na |
| DBNPG | Dibromoneopentyl alcohol |  | C ₅ H ₁₀ Br ₂ O ₂ | 3296-90-0 | 261.9 | 5.58 | 307.0 | 1015/8 | 0.85 | 0.69 | 7.84 | 8.6E-4 | 4.1E-9 | 13.6 |
| DBS | (2,2-Dibromovinyl) benzene |  | C ₈ H ₆ Br ₂ | 31780-26-4 | 261.9 | 4.39 | 273.4 | 75.8 | 3.55 | 3.08 | 5.90 | 0.78 | 1.1E-4 | na |
| anti-DDC-CO | Dechlorane Plus |  | C ₁₈ H ₁₂ Cl ₁₂ | 13560-89-9 | 653.7 | 13.91 | 486.8 | 6.5E-7 | 11.27 | 9.78 | 14.8 | 9.4E-8 | 7.4E-6 | na |

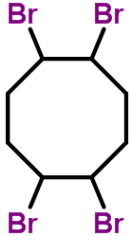
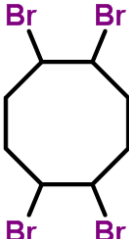
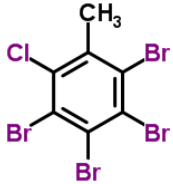
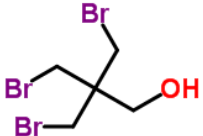
| Abbreviation | Name | Structure | Molecular formula | CAS no. | MW ^b | <i>t</i> _R ^c | <i>Bp</i> ^d | <i>S</i> _W ^e | log <i>K</i> _{OW} ^f | log <i>K</i> _{OC} ^g | log <i>K</i> _{OA} ^h | <i>V</i> _P ⁱ | <i>H</i> ^j | <i>pK</i> _a ^k |
|--------------|--|---|--|-------------|-----------------|------------------------------------|------------------------|------------------------------------|---|---|---|------------------------------------|-----------------------|-------------------------------------|
| syn-DDC-CO | Dechlorane Plus |  | C ₁₈ H ₁₂ Cl ₁₂ | 13560-89-9 | 653.7 | 13.67 | 486.6 | 6.5E-7 | 11.27 | 9.78 | 14.8 | 9.4E-8 | 7.4E-6 | na |
| EH-TBB | 2-Ethylhexyl 2,3,4,5-tetra-bromobenzoate |  | C ₁₅ H ₁₈ Br ₄ O ₂ | 183658-27-7 | 549.9 | 11.33 | 432.9 | 3.4E-3 | 8.75 | 5.70 | 12.3 | 4.6E-6 | 6.4E-6 | na |
| HBB | Hexabromo-benzene |  | C ₆ Br ₆ | 87-82-1 | 551.5 | 10.02 | 370.7 | 0.23 | 6.07 ¹ | 5.27 | 9.1 | 2.2E-6 | 2.2E-5 | na |
| α-HBCDD | Hexabromo-cyclododecane |  | C ₁₂ H ₁₈ Br ₆ | 3194-55-6 | 641.7 | 12.34 | 462.0 | 3.1E-3 | 7.74 | 6.72 | 10.5 | 2.3E-6 | 4.6E-5 | na |

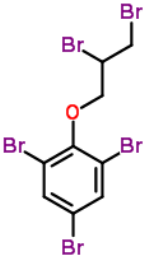
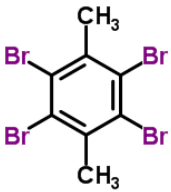
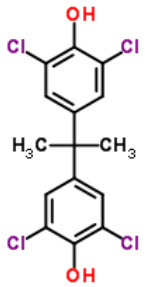
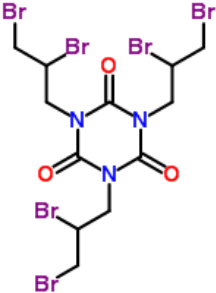
| Abbreviation | Name | Structure | Molecular formula | CAS no. | MW ^b | t_R^c | Bp^d | S_W^e | $\log K_{OW}^f$ | $\log K_{OC}^g$ | $\log K_{OA}^h$ | V_P^i | H^j | pK_a^k |
|-----------------|--|---|--|--------------|-----------------|---------|--------|--------------------|-----------------|-----------------|-----------------|---------------------|---------------------|----------|
| β -HBCDD | Hexabromo-cyclododecane |  | C ₁₂ H ₁₈ Br ₆ | 3194-55-6 | 641.7 | 12.34 | 462.0 | 3.1E- ₃ | 7.74 | 6.72 | 10.5 | 2.3E- ₆ | 4.6E- ₅ | na |
| γ -HBCDD | Hexabromo-cyclododecane |  | C ₁₂ H ₁₈ Br ₆ | 3194-55-6 | 641.7 | 12.34 | 462.0 | 3.1E- ₃ | 7.74 | 6.72 | 10.5 | 2.3E- ₆ | 4.6E- ₅ | na |
| HEEHP-TEBP | 2-(2-hydroxyethoxy) ethyl-2-hydroxy-propyl-3,4,5,6-tetrabromophthalate |  | C ₁₅ H ₁₆ Br ₄ O ₇ | 20566-35-2 | 627.9 | 11.43 | 537.5 | 769 | 3.83 | 2.00 | 17.8 | 3.2E- ₁₂ | 2.7E- ₁₆ | na |
| OBTMPI | 4,5,6,7-Tetrabromo-1,1,3-trimethyl-3-(2,3,4,5-tetrabromophenyl)-indane |  | C ₁₈ H ₁₂ Br ₈ | 1084889-51-9 | 867.5 | 15.86 | 572.2 | 8.7E- ₇ | 13.03 | 11.31 | 17.8 | 2.1E- ₁₀ | 4.6E- ₇ | na |

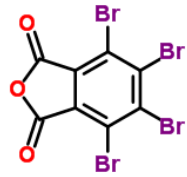
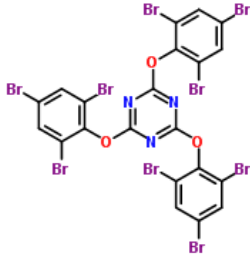
| Abbreviation | Name | Structure | Molecular formula | CAS no. | MW ^b | <i>t_R</i> ^c | <i>Bp</i> ^d | <i>S_w</i> ^e | log <i>K_{OW}</i> ^f | log <i>K_{OC}</i> ^g | log <i>K_{OA}</i> ^h | <i>V_P</i> ⁱ | <i>H^j</i> | <i>pK_a</i> ^k |
|--------------|------------------------------|---|---|------------|-----------------|-----------------------------------|------------------------|-----------------------------------|--|--|--|-----------------------------------|----------------------|------------------------------------|
| PBB-Acr | Pentabromobenzyl acrylate |  | C ₁₀ H ₅ Br ₅ O ₂ | 59447-55-1 | 556.7 | 10.91 | 411.7 | 0.13 | 6.89 | 4.67 | 12.4 | 1.8E-5 | 7.5E-8 | na |
| PBBBr | Pentabromobenzyl-bromide |  | C ₇ H ₂ Br ₆ | 38521-51-6 | 565.5 | 10.56 | 389.2 | 0.067 | 7.33 | 6.36 | 10.9 | 8.5E-5 | 6.9E-6 | na |
| PBCH | Pentabromochloro-cyclohexane |  | C ₆ H ₆ Br ₅ Cl | 87-84-3 | 513.1 | 8.90 | 373.2 | 0.45 | 4.72 ¹ | 4.10 | 9.1 | 4.6E-4 | 9.6E-7 | na |
| PBEB | Pentabromoethyl-benzene |  | C ₈ H ₅ Br ₅ | 85-22-3 | 500.6 | 9.46 | 363.2 | 0.11 | 7.48 | 6.49 | 10.0 | 6.2E-4 | 7.9E-5 | na |

| Abbreviation | Name | Structure | Molecular formula | CAS no. | MW ^b | t_R^c | Bp^d | S_W^e | $\log K_{OW}^f$ | $\log K_{OC}^g$ | $\log K_{OA}^h$ | V_P^i | H^j | pK_a^k |
|--------------|------------------------------|---|--|-----------|-----------------|---------|--------|---------|-----------------|-----------------|-----------------|---------|---------|---------------|
| PBP | Pentabromophenol |  | C ₆ HBr ₅ O | 608-71-9 | 488.6 | 9.57 | 374.8 | 34 | 5.96 | 4.39 | 12.6 | 6.8E-6 | 5.6E-9 | 4.4±0.33 |
| PBP AE | Pentabromophenyl allyl ether |  | C ₉ H ₅ Br ₅ O | 3555-11-1 | 528.7 | 9.95 | 386.0 | 0.052 | 7.37 | 5.06 | 11.1 | 1.2E-4 | 4.2E-6 | na |
| PBT | Pentabromotoluene |  | C ₇ H ₃ Br ₅ | 87-83-2 | 486.6 | 9.23 | 351.6 | 0.35 | 6.99 | 6.07 | 9.6 | 2.0E-5 | 6.0E-5 | na |
| TBBPA | Tetrabromobisphenol A |  | C ₁₅ H ₁₂ Br ₄ O ₂ | 79-94-7 | 543.9 | 12.14 | 454.6 | 4.3E-3 | 6.25 | 5.42 | 18.2 | 9.1E-7 | 2.3E-13 | 7.5/8.5 ± 0.1 |

| Abbreviation | Name | Structure | Molecular formula | CAS no. | MW ^b | t_R^c | Bp^d | S_W^e | $\log K_{OW}^f$ | $\log K_{OC}^g$ | $\log K_{OA}^h$ | V_P^i | H^j | pK_a^k |
|--------------|--|---|--|------------|-----------------|---------|--------|---------|-----------------|-----------------|-----------------|---------|---------|----------|
| TBBPA-BAE | Tetrabromo-bisphenol A bis(allyl ether) |  | C ₂₁ H ₂₀ Br ₄ O ₂ | 25327-89-3 | 624.0 | - | 508.7 | 3.4E-6 | 10.02 | 6.58 | 15.3 | 2.7E-7 | 1.3E-7 | na |
| TBBPA-BDBPE | Tetrabromo-bisphenol A-bis(2,3-dibromo-propyl ether) |  | C ₂₁ H ₂₀ Br ₈ O ₂ | 21850-44-2 | 943.6 | 14.03 | 646.9 | 9.4E-7 | 11.52 | 7.41 | 20.3 | 8.5E-13 | 4.1E-11 | na |
| TBBPA-DHEE | Tetrabromo-bisphenol A dihydroxyethyl ether |  | C ₁₉ H ₂₀ Br ₄ O ₄ | 4162-45-2 | 632.0 | 13.92 | 574.1 | 2.3E-2 | 6.78 | 3.96 | 17.9 | 5.3E-12 | 1.8E-13 | 13.8 |
| TBBPS-DBPE | Tetrabromo-bisphenol-S-bis(2,3-dibromopropyl) ether |  | C ₁₈ H ₁₄ Br ₈ O ₄ S | 42757-55-1 | 965.6 | - | 705.9 | 9.7E-7 | 9.52 | 6.33 | 21.8 | 1.0E-14 | 1.2E-14 | na |

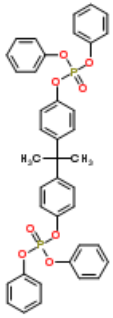
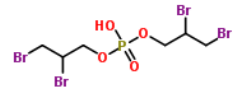
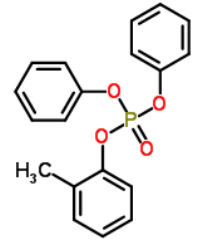
| Abbreviation | Name | Structure | Molecular formula | CAS no. | MW ^b | t_R^c | Bp^d | S_W^e | $\log K_{OW}^f$ | $\log K_{OC}^g$ | $\log K_{OA}^h$ | V_P^i | H^j | pK_a^k |
|----------------|---|---|--|------------|-----------------|---------|--------|---------|-----------------|-----------------|-----------------|--------------------|--------------------|----------|
| α -TBCO | 1,2,5,6-Tetrabromo-cyclooctane |  | C ₈ H ₁₂ Br ₄ | 3194-57-8 | 427.8 | 8.49 | 342.8 | 1.5 | 5.24 | 4.55 | 8.0 | 9.4E- ₃ | 4.2E- ₅ | na |
| β -TBCO | 1,2,5,6-Tetrabromo-cyclooctane |  | C ₈ H ₁₂ Br ₄ | 3194-57-8 | 427.8 | 8.49 | 342.8 | 1.5 | 5.24 | 4.55 | 8.0 | 9.4E- ₃ | 4.2E- ₅ | na |
| TBCT | 1,2,3,4-Tetrabromo-5-chloro-6-methylbenzene |  | C ₇ H ₃ Br ₄ Cl | 39569-21-6 | 442.2 | 8.70 | 339.4 | 0.38 | 6.74 | 5.85 | 9.1 | 3.7E- ₃ | 1.1E- ₄ | na |
| TBNPA | Tribromoneopentyl alcohol |  | C ₅ H ₉ Br ₃ O | 1522-92-5 | 324.8 | 5.84 | 299.7 | 852 | 2.25 | 1.76 | 8.5 | 5.6E- ₃ | 1.3E- ₈ | 13.7 |

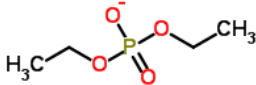
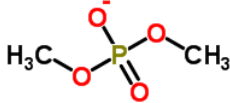
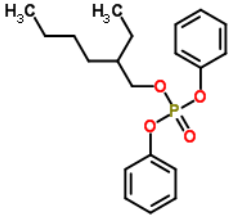
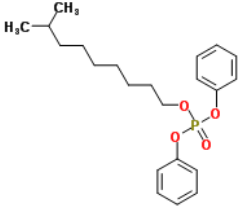
| Abbreviation | Name | Structure | Molecular formula | CAS no. | MW ^b | t_R^c | Bp^d | S_W^e | $\log K_{OW}^f$ | $\log K_{OC}^g$ | $\log K_{OA}^h$ | V_P^i | H^j | pK_a^k |
|--------------|--|---|---|------------|-----------------|---------|--------|---------|-----------------|-----------------|-----------------|---------|---------|----------|
| TBP-DBPE | 2,3-Dibromopropyl-2,4,6-tribromophenyl ether |  | C ₉ H ₇ Br ₅ O | 35109-60-5 | 530.7 | 9.88 | 393.7 | 0.080 | 6.34 | 4.49 | 11.1 | 8.3E-5 | 4.7E-7 | na |
| TBX | 2,3,5,6-tetrabromo-p-xylene |  | C ₈ H ₆ Br ₄ | 23488-38-2 | 421.8 | 8.35 | 331.9 | 0.53 | 6.65 | 5.77 | 8.8 | 5.5E-3 | 1.7E-4 | na |
| TCBPA | Tetrachloro-bisphenol-A |  | C ₁₅ H ₁₂ Cl ₄ O ₂ | 27360-90-3 | 366.1 | 10.81 | 438.3 | 0.40 | 6.22 | 4.84 | 16.2 | 3.8E-7 | 2.8E-12 | na |
| TDBP-TAZTO | 1,3,5-tris(2,3-dibromopropyl)-1,3,5-triazine-2,4,6-(1H,3H,5H)-trione |  | C ₁₂ H ₁₅ Br ₆ N ₃ O ₃ | 52434-90-9 | 728.7 | 13.54 | 669.5 | 1.4E-5 | 7.37 | 4.92 | 23.7 | 1.5E-10 | 1.2E-18 | na |

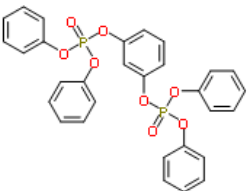
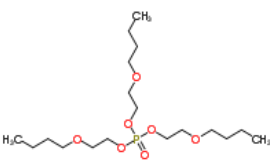
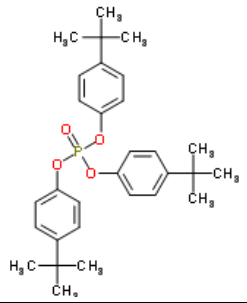
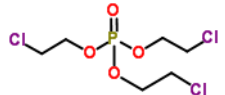
| Abbreviation | Name | Structure | Molecular formula | CAS no. | MW ^b | t_R ^c | Bp ^d | S_W ^e | $\log K_{OW}$ ^f | $\log K_{OC}$ ^g | $\log K_{OA}$ ^h | V_P ⁱ | H^j | pK_a ^k |
|--------------|--|---|--|------------|-----------------|--------------------|-------------------|--------------------|----------------------------|----------------------------|----------------------------|--------------------|---------|---------------------|
| TEBP-Anh | 3,4,5,6-Tetrabromophthalic anhydride |  | C ₈ Br ₄ O ₃ | 632-79-1 | 463.7 | 10.03 | 394.1 | 65 | 5.63 | 3.58 | 10.8 | 2.7E-6 | 1.6E-7 | na |
| TTBP-TAZ | 2,4,6-tris(2,4,6-tribromophenoxy)-1,3,5-triazine |  | C ₂₁ H ₆ Br ₉ N ₃ O ₃ | 25713-60-4 | 1067 | 18.00 | 767.7 | 1.1E-6 | 11.46 | 7.25 | 21.5 | 9.3E-17 | 2.4E-12 | na |

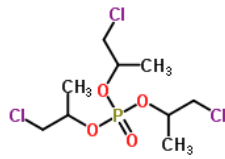
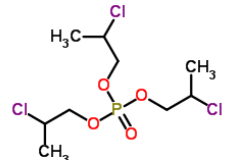
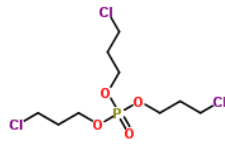
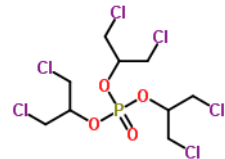
^aPhysico-chemical properties (except pK_a) were modeled using EPIsuite 4.1 (US EPA); na = not available; ^bMW = molecular weight; ^c t_R = retention time (min) using GC-conditions given in Section 2.3; ^d B_p = boiling point (°C); ^e S_W = water solubility (mg L⁻¹, 25°C); ^f K_{OW} = octanol-water partition coefficient; ^g K_{OC} = organic carbon-water partition coefficient; ^h K_{OA} = octanol-air partition coefficient; ⁱ V_p = vapour pressure; ^j H = Henry's law constant (atm m³ mole⁻¹); ^kacid dissociation values (pK_a) from Bergman et al. (2012) [13]. ^lExperimental value from EPIsuite 4.1.

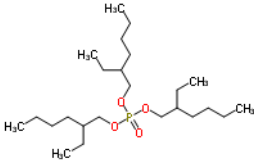
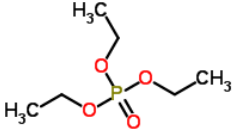
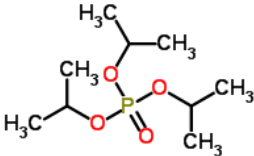
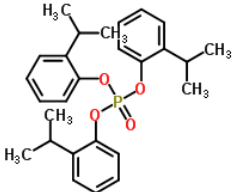
Table S17 Structure, molecular formula, CAS no., retention time and physico-chemical properties of selected OPFRs ($n = 29$). Table reused from Gustavsson et al. (2017) [14].^a

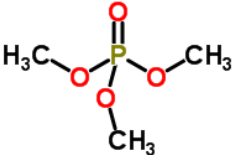
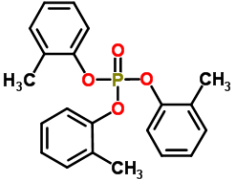
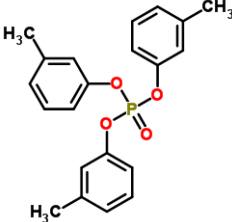
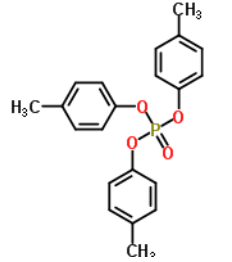
| Abbreviation | Name | Structure | Molecular formula | CAS no. | MW ^b | t_R^c | Bp^d | S_w^e | $\log K_{OW}^f$ | $\log K_{OC}^g$ | $\log K_{OA}^h$ | V_P^i | H^j | pK_a^k |
|--------------|--------------------------------------|--|---|------------|-----------------|---------|--------|---------|-----------------|-----------------|-----------------|---------|---------|----------|
| BADP | Bisphenol A bis (diphenyl phosphate) |  | C ₃₉ H ₃₄ O ₈ P ₂ | 5945-33-5 | 692.7 | - | 480.0 | 1.9E-6 | 10.0 | 6.24 | 21.7 | 2.7E-6 | 4.6E-14 | na |
| bBDBP | bis(2,3-Dibromopropyl) phosphate |  | C ₆ H ₁₁ Br ₄ O ₄ P | 5412-25-9 | 497.7 | - | 434.9 | 220 | 2.53 | 2.43 | 13.3 | 6.0E-7 | 4.3E-13 | na |
| CDP | Cresyl diphenyl phosphate |  | C ₁₉ H ₁₇ O ₄ P | 26444-49-5 | 340.3 | - | 452.9 | 1.5 | 5.25 | 3.19 | 10.3 | 1.4E-5 | 4.2E-8 | na |

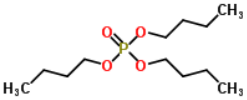
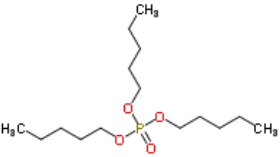
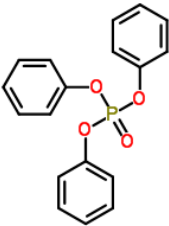

| Abbreviation | Name | Structure | Molecular formula | CAS no. | MW ^b | t_R^c | Bp^d | S_w^e | $\log K_{OW}^f$ | $\log K_{OC}^g$ | $\log K_{OA}^h$ | V_P^i | H^j | pK_a^k |
|--------------|----------------------------------|---|--|------------|-----------------|---------|--------|---------|-------------------|-----------------|-----------------|---------|---------|----------|
| mDEP/dDEP | Diethyl phosphate (mono & di) |  | C ₄ H ₁₁ O ₄ P | 598-02-7 | 154.1 | - | 258.8 | 4.0E5 | 0.32 | 1.21 | 7.57 | 1.8E-3 | 1.4E-9 | na |
| DMP | Dimethyl phosphate |  | C ₂ H ₇ O ₄ P | 813-78-5 | 126.1 | - | 222.9 | 1.0E6 | -0.66 | 0.66 | 6.84 | 46 | 7.8E-10 | na |
| EHDPP | 2-Ethylhexyl diphenyl phosphate |  | C ₂₀ H ₂₇ O ₄ P | 1241-94-7 | 362.4 | 10.00 | 443.0 | 0.18 | 5.73 ¹ | 3.87 | 8.4 | 4.5E-3 | 2.5E-7 | na |
| IDP | Isodecyl diphenyl phosphate |  | C ₂₂ H ₃₁ O ₄ P | 29761-21-5 | 390.5 | 10.82 | 466.2 | 1.7E-2 | 7.28 | 3.71 | 10.2 | 6.3E-6 | 4.4E-7 | na |

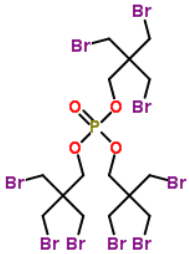
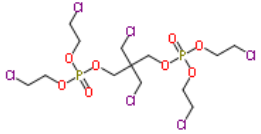
| Abbreviation | Name | Structure | Molecular formula | CAS no. | MW ^b | t_R^c | Bp^d | S_w^e | $\log_{K_{OW}}^f$ | $\log_{K_{OC}}^g$ | $\log_{K_{OA}}^h$ | V_P^i | H^j | pK_a^k |
|--------------|-------------------------------------|---|---|------------|-----------------|---------|--------|---------|-------------------|-------------------|-------------------|---------|---------|----------|
| PBDPP | Resorcinol bis (diphenyl phosphate) |  | C ₃₀ H ₂₄ O ₈ P ₂ | 57583-54-7 | 574.5 | 15.43 | 480.0 | 6.9E-3 | 7.41 | 4.80 | 18.3 | 2.7E-6 | 2.9E-13 | na |
| TBOEP | Tri(2-butoxyethyl) phosphate |  | C ₁₈ H ₃₉ O ₇ P | 78-51-3 | 398.5 | 9.99 | 433.8 | 604 | 3.00 | 2.83 | 13.1 | 1.7E-4 | 1.2E-11 | na |
| TBPP | Tris(4-tert-butylphenyl) phosphate |  | C ₃₀ H ₃₉ O ₄ P | 78-33-1 | 494.6 | 13.25 | 480.0 | 4.1E-5 | 10.43 | 6.47 | 15.0 | 2.7E-6 | 6.9E-7 | na |
| TCEP | Tris(2-chloroethyl) phosphate |  | C ₆ H ₁₂ Cl ₃ O ₄ P | 115-96-8 | 285.5 | 7.00 | 351.7 | 5597 | 1.63 | 1.83 | 5.31 | 0.052 | 3.3E-6 | na |

| Abbreviation | Name | Structure | Molecular formula | CAS no. | MW ^b | t_R^c | Bp^d | S_w^e | $\log K_{OW}^f$ | $\log K_{OC}^g$ | $\log K_{OA}^h$ | V_P^i | H^j | pK_a^k |
|--------------|---------------------------------------|---|---|------------|-----------------|---------|--------|---------|-------------------|-----------------|-----------------|---------|--------|----------|
| TCIPP | Tri(1-chloro-2-propyl) phosphate |  | C ₉ H ₁₈ Cl ₃ O ₄ P | 13674-84-5 | 327.6 | 7.14 | 365.5 | 740 | 2.59 ¹ | 2.46 | 8.2 | 7.5E-3 | 6.0E-8 | na |
| T2CPP | Tris(2-Chloropropyl) phosphate |  | C ₉ H ₁₈ Cl ₃ O ₄ P | 6145-73-9 | 327.57 | 7.14 | 346.5 | 740 | 2.89 | 2.63 | 8.50 | 7.0E-3 | 6.0E-8 | na |
| T3CPP | Tris(3-chloropropyl) phosphate |  | C ₉ H ₁₈ Cl ₃ O ₄ P | 26248-87-3 | 327.57 | 7.14 | 386.5 | 157 | 3.11 | 2.75 | 8.72 | 6.4E-4 | 6.0E-8 | na |
| TDCIPP | Tris(1,3-dichloroisopropyl) phosphate |  | C ₉ H ₁₅ Cl ₆ O ₄ P | 13674-87-8 | 430.9 | 9.57 | 458.7 | 30 | 3.65 ¹ | 3.05 | 10.6 | 3.8E-5 | 2.6E-9 | na |

| Abbreviation | Name | Structure | Molecular formula | CAS no. | MW ^b | t_R^c | Bp^d | S_w^e | $\log K_{OW}^f$ | $\log K_{OC}^g$ | $\log K_{OA}^h$ | V_P^i | H^j | pK_a^k |
|--------------|----------------------------------|---|--|------------|-----------------|---------|--------|---------|-----------------|-----------------|-----------------|---------|--------|----------|
| TEHP | Tris(2-ethylhexyl) phosphate |  | C ₂₄ H ₅₁ O ₄ P | 78-42-2 | 434.7 | 9.57 | 446.3 | 2.8E-4 | 9.49 | 6.28 | 15.0 | 8.1E-5 | 7.9E-8 | na |
| TEP | Triethyl phosphate |  | C ₆ H ₁₅ O ₄ P | 78-40-0 | 182.2 | - | 233.3 | 11525 | 0.87 | 1.47 | 6.63 | 22 | 3.6E-8 | na |
| TiPP | Triisopropyl phosphate |  | C ₉ H ₂₁ O ₄ P | 513-02-0 | 224.2 | - | 254.5 | 16352 | 2.12 | 2.20 | 6.38 | 18 | 1.4E-6 | na |
| TiPPP | Tri(2-Isopropylphenyl) phosphate |  | C ₂₇ H ₃₃ O ₄ P | 64532-95-2 | 452.5 | 11.24 | 480.0 | 4.9E-4 | 9.07 | 5.72 | 14.0 | 2.7E-6 | 2.9E-7 | na |

| Abbreviation | Name | Structure | Molecular formula | CAS no. | MW ^b | t_R^c | Bp^d | S_w^e | $\log K_{OW}^f$ | $\log K_{OC}^g$ | $\log K_{OA}^h$ | V_P^i | H^j | pK_a^k |
|----------------|----------------------------------|---|--|-----------|-----------------|---------|--------|---------|-----------------|-----------------|-----------------|---------|--------|----------|
| TMP | Trimethyl phosphate |  | C ₃ H ₉ O ₄ P | 512-56-1 | 140.1 | - | 174.2 | 1.0E6 | -0.60 | 0.669 | 5.88 | 55 | 7.2E-9 | na |
| <i>o</i> -TMPP | <i>ortho</i> -Tritolyl phosphate |  | C ₂₁ H ₂₁ O ₄ P | 1330-78-5 | 368.4 | 10.58 | 476.1 | 0.14 | 6.34 | 4.21 | 12.0 | 4.7E-6 | 5.4E-8 | na |
| <i>m</i> -TMPP | <i>meta</i> -Tritolyl phosphate |  | C ₂₁ H ₂₁ O ₄ P | 1330-78-5 | 368.4 | 10.86 | 476.1 | 0.14 | 6.34 | 4.21 | 12.0 | 4.7E-6 | 5.4E-8 | na |
| <i>p</i> -TMPP | <i>para</i> -Tritolyl phosphate |  | C ₂₁ H ₂₁ O ₄ P | 1330-78-5 | 368.4 | 11.19 | 476.1 | 0.14 | 6.34 | 4.21 | 12.0 | 4.7E-6 | 5.4E-8 | na |

| Abbreviation | Name | Structure | Molecular formula | CAS no. | MW ^b | t_R^c | Bp^d | S_w^e | $\log K_{OW}^f$ | $\log K_{OC}^g$ | $\log K_{OA}^h$ | V_P^i | H^j | pK_a^k |
|--------------|---------------------|---|--|-----------|-----------------|---------|--------|---------|-------------------|-----------------|-----------------|---------|--------|----------|
| TNBP | Tributyl phosphate |  | C ₁₂ H ₂₇ O ₄ P | 126-73-8 | 266.3 | 6.35 | 327.0 | 101 | 4 ¹ | 3.24 | 8.2 | 0.47 | 1.4E-6 | na |
| TPeP | Tripentyl phosphate |  | C ₁₅ H ₃₃ O ₄ P | 2528-38-3 | 308.4 | 7.86 | 362.9 | 2.9 | 5.29 | 3.96 | 8.8 | 2.2E-3 | 7.5E-6 | na |
| TPHP | Triphenyl phosphate |  | C ₁₈ H ₁₅ O ₄ P | 115-86-6 | 326.3 | 9.82 | 441.3 | 4.7 | 4.70 | 3.24 | 8.46 | 6.3E-5 | 3.3E-6 | na |
| TPP | Tripropyl phosphate |  | C ₉ H ₂₁ O ₄ P | 513-08-6 | 224.2 | 4.46 | 284.2 | 3474 | 1.87 ¹ | 2.06 | 6.4 | 3.1 | 6.8E-7 | na |

| Abbreviation | Name | Structure | Molecular formula | CAS no. | MW ^b | t_R ^c | Bp ^d | S_W ^e | $\log K_{OW}$ ^f | $\log K_{OC}$ ^g | $\log K_{OA}$ ^h | V_P ⁱ | H ^j | pK_a ^k |
|--------------|---|---|---|------------|-----------------|--------------------|-------------------|--------------------|----------------------------|----------------------------|----------------------------|--------------------|------------------|---------------------|
| TTBNPP | Tris(tribromoneopentyl) phosphate |  | C ₁₅ H ₂₄ Br ₉ O ₄ P | 19186-97-1 | 1020 | 5.91 | 480.0 | 1.2E-5 | 8.05 | 5.48 | 20.0 | 2.7E-6 | 2.7E-14 | na |
| BCMP-BCEP | Tetrakis(2-Chloroethyl) dichloroisopentyl diphosphate |  | C ₁₃ H ₂₄ Cl ₆ O ₈ P ₂ | 38051-10-4 | 583.0 | 12.91 | 480.0 | 33 | 3.31 | 2.86 | 15.5 | 2.7E-6 | 1.6E-14 | na |

^aPhysico-chemical properties (except pK_a) were modeled using EPIsuite 4.1 (US EPA); na = not available; ^bMW = molecular weight; ^c t_R = retention time (min) using GC-conditions given in Section 2.3; ^d B_p = boiling point (°C); ^e S_W = water solubility (mg L⁻¹, 25°C); ^f K_{OW} = octanol-water partition coefficient; ^g K_{OC} = organic carbon-water partition coefficient; ^h K_{OA} = octanol-air partition coefficient; ⁱ V_p = vapour pressure; ^j H = Henry's law constant (atm m³ mole⁻¹); ^kacid dissociation values (pK_a) from Bergman et al. (2012) [13]. ^lExperimental value from EPIsuite 4.1.