**1** Reducing the Effects of Vegetation Phenology on Change Detection in

# 2 Tropical Seasonal Biomes

- 3 Eduarda Martiniano de Oliveira Silveira<sup>1</sup>\*, Fernando Del Bon Espírito-
- 4 Santo<sup>2</sup>, Fausto Weimar Acerbi-Júnior<sup>1</sup>, Lênio Soares Galvão<sup>3</sup>, Kieran
- 5 Daniel Withey<sup>4</sup>, George Alan Blackburn<sup>4</sup>, José Márcio de Mello<sup>1</sup>, Yosio
- 6 Edemir Shimabukuro<sup>3</sup>, Tomas Domingues<sup>5</sup> and José Roberto Soares
- 7 Scolforo<sup>1</sup>
- 8 <sup>1</sup>Forest Science Department, Federal University of Lavras (UFLA), Lavras, Brazil,
- 9 3037, dudalavras@hotmail.com; fausto@dcf.ufla.br; josemarcio@dcf.ufla.br;
- 10 jscolforo@dcf.ufla.br
- <sup>11</sup> <sup>2</sup>School of Geography, Geology and the Environment, University of Leicester, Leicester,
- 12 UK, LE1 7RH, f.delbon@gmail.com
- 13 <sup>3</sup>National Institute for Space Research (INPE), São José dos Campos, Brazil, 3037,
- 14 122270, lenio.galvao@inpe.br; yosio@dsr.inpe.br
- <sup>4</sup>Lancaster Environment Centre (LEC), Lancaster University, Lancaster, UK, LA1 4YQ,
- 16 kieranwithey@gmail.com; alan.blackburn@lancaster.ac.uk
- 17 <sup>5</sup>Biology Department, University of São Paulo (USP-FFCLRP), Ribeirão Preto, Brazil,
- 18 tdomingu@gmail.com
- 19 \**Corresponding author*
- 20
- 21
- 22
- 23

# Reducing the Effects of Vegetation Phenology on Change Detection in Tropical Seasonal Biomes

26	Tropical seasonal biomes (TSBs), such as the savannas (Cerrado) and semi-arid
27	woodlands (Caatinga) of Brazil, are vulnerable ecosystems to human-induced
28	disturbances. Remote sensing can detect disturbances such as deforestation and
29	fires, but the analysis of change detection in TSBs is affected by seasonal
30	modifications in vegetation indices due to phenology. To reduce the effects of
31	vegetation phenology on changes caused by deforestation and fires, we
32	developed a novel object-based change detection method. The approach
33	combines both the spatial and spectral domains of the normalized difference
34	vegetation index (NDVI), using a pair of Operational Land Imager
35	(OLI)/Landsat-8 images acquired in 2015 and 2016. We used semivariogram
36	indices (SIs) as spatial features and descriptive statistics as spectral features
37	(SFs). We tested the performance of the method using three machine-learning
38	algorithms: support vector machine (SVM), artificial neural network (ANN) and
39	random forest (RF). The results showed that the combination of spatial and
40	spectral information improved change detection by correctly classifying areas
41	with seasonal changes in NDVI caused by vegetation phenology and areas with
42	NDVI changes caused by human-induced disturbances. The use of
43	semivariogram indices reduced the effects of vegetation phenology on change
44	detection. The performance of the classifiers was generally comparable, but the
45	SVM presented the highest overall classification accuracy (92.27%) when using
46	the hybrid set of NDVI-derived spectral-spatial features. From the vegetated
47	areas, 18.71% of changes were caused by human-induced disturbances between
48	2015 and 2016. The method is particularly useful for TSBs where vegetation
49	exhibits strong seasonality and regularly spaced time series of satellite images are
50	difficult to obtain due to persistent cloud cover.

51

Keywords: remote sensing; geostatistics; seasonality; LULCC

### 52 1. Introduction

53 Tropical seasonal biomes (TSBs), such as savannas (also known as Cerrado) and semi-

54 arid woodlands (also known as *Caatinga*), cover 35% of Brazil and consist of several

55 vegetation types ranging from grasslands to forests (Silveira et al. 2018a). However,

human-induced disturbances, such as deforestation and fires, are threatening these
ecosystems (Silva et al. 2006; Hansen et al. 2013). In addition, because most of the
conservation plans focus on moist evergreen tropical forests (Hoekstra et al. 2005), less
attention has been dedicated to TSB areas (Beuchle et al. 2015).

60 TSBs experience seasonal changes in hydrological and nutrient conditions that 61 affect the spectral signature of vegetation measured by satellites (Zhang, Ross, and 62 Gann 2016). For instance, leaf area index (LAI) varies seasonally, having a maximum 63 value during the rainy season and a minimum value during the dry season. Therefore, a 64 seasonal fluctuation in the Normalized Difference Vegetation Index (NDVI) is 65 generally observed over TSBs due to leaf shedding and increasing amounts of 66 nonphotosynthetic vegetation during the dry season (Lagomasino et al. 2014). This 67 NDVI behavior represents a challenge for land use and land cover change (LULCC) 68 detection when multi-temporal images are used in the analysis.

69 Bi-temporal remote sensing images can be used to monitor vegetation and to 70 detect changes caused by human and natural processes (Verbesselt et al. 2010; Zhu, 71 Woodcock, and Olofsson 2012). However, in TSBs, phenology produces significant 72 changes in vegetation conditions affecting the spectral response of vegetation (Wright 73 and Schaik 1994). Even fixing a single period for image acquisition (rainy or dry 74 season), the effects of vegetation phenology on LULCC detection are still significant 75 due to the large seasonal and interannual variability in precipitation observed in TSBs. 76 Several methods have been proposed to reduce the effects of vegetation 77 phenology on LULCC detection using time series of satellite images. Examples are the 78 Breaks For Additive Seasonal and Trend algorithm (BFAST) (Verbesselt et al. 2010); 79 Continuous Change Detection and Classification (CCDC) (Zhu and Woodcock 2014); 80 Vegetation Change Tracker (VCT) (Huang et al. 2010); LandTrend (Kennedy, Yang,

and Cohen 2010); Vegetation Regeneration and Disturbance Estimates through Time
(VerDET) (Hughes, Kaylor, and Hayes 2017); and the Residual Trend Analysis
(RESTREND) (Evans and Geerken 2004; Ibrahim et al. 2015). These methods usually
require high-quality time series, which are not generally available over TSBs due to
persistent cloud cover. Therefore, LULCC detection in complex landscapes, like those
found in TSBs, still present a significant challenge (Healey et al. 2018).

87 Previous studies have shown that pixel-based change detection approaches can 88 benefit from including information on spatial context (G. Chen et al. 2012; Hamunyela, 89 Verbesselt, and Herold 2016). The neighborhood used to extract the spatial information 90 is often defined by a square window that is easy to implement, however, they are 91 computationally demanding (Zhu 2017), biased along their diagonals, and can straddle 92 the boundary between two landscape features, especially when a large window size is 93 used (Laliberte, Rango, and Laliberte A. 2009). Using OBIA these problems are 94 eliminated, allowing the inclusion of additional spatial information to improve remote 95 sensing applications (G. Chen et al. 2018). For example, semivariograms of geostatistics 96 have been widely used in image classification analyses (Balaguer et al. 2010; Silveira et 97 al. 2017; Wu et al. 2015) and change detection studies (Gil-Yepes et al. 2016; 98 Hamunyela et al. 2017; Silveira et al. 2018b). Thus, object-based methods that require 99 only a few satellite images to reduce the effects of vegetation phenology on LULCC 100 detection are needed to monitor TSB areas with persistent cloud cover and strong 101 seasonality.

Here, to evaluate whether we can differentiate seasonal variations in NDVI
values due to vegetation phenology from spectral variations associated with humaninduced disturbances, we developed a novel object-based change detection (OBCD)
approach. The objective was to reduce the effects of vegetation phenology on LULCC

106 detection by combining spatial (i.e. semivariogram indices - SIs) and spectral

107 information (i.e. spectral features - SFs). Our method does not require time series of

108 satellite images because it exploits the spatial and spectral domains of NDVI, calculated

109 from a pair of Operational Land Imager (OLI)/Landsat-8 images. Specifically, we tested

110 the approach with three machine learning algorithms (MLAs), including support vector

111 machine (SVM), artificial neural network (ANN) and random forest (RF) algorithms, to

112 classify areas that experienced changes caused by vegetation phenology and human-

113 induced disturbances

#### 114 **2. Study Area**

115 The study area is located in the north of Minas Gerais (MG) state, Brazil (Figure 1). In 116 this area, the TSBs include Brazilian savannas (*Cerrado*) and semi-arid woodlands

117 (*Caatinga*) (Figure 1a) (Scolforo et al. 2015). The study area is covered by the path 219

and row 71 of the Worldwide Reference System version 2 (WRS-2) (Figure 1b). From a

119 total of 32,000 km<sup>2</sup>, 50% of the area is covered by native vegetation (Figure 1c)

120 (Carvalho et al. 2006).

121

#### [Figure 1 near here]

122 The diversity of vegetation types in the study area is well documented, ranging 123 from savanna grasslands and woodland savannas to semideciduous and deciduous 124 forests (Ferreira et al. 2004). Low shrubs to small patches of tall dry forests are 125 therefore observed (Figure 2) (Santos et al. 2012). The study area has experienced 126 extensive land-cover change (Espírito-Santo et al. 2016), resulting from the 127 implementation of cattle grazing and establishment of pastures. In general, the native 128 vegetation has been converted into areas of pasture or croplands (Sano et al. 2010). 129 The climate is tropical with rainfall concentrated in October to May. The peak of 130 the dry season in August has close to zero rainfall and air humidity less than 20% with

	131	high seasonality	y (Peel,	Finlayson,	and McMahon	2006).	Rainfall in	this region is
--	-----	------------------	----------	------------	-------------	--------	-------------	----------------

132 extremely irregular over space and time. More than 75% of the total annual rainfall

133 occurs within three months, but interannual variation in precipitation is large and

134 droughts can last for years in areas of Caatinga (Leal et al. 2005).

135 [Figure 2 near here]

#### 136 **3. Methodology**

137 We developed a new OBCD method to detect human-induced changes in TSBs by

138 reducing the effects of vegetation phenology on change detection, combining both the

139 spatial and spectral domains of bi-temporal NDVI images. We used semivariogram

140 indices (SIs) as spatial features (Balaguer et al. 2010), as described below (see Table 1).

141 Descriptive statistics for NDVI imagery was used to represent spectral features (SFs), as

142 detailed below (see section 3.4.).

By training MLAs using the difference between the two NDVI images in terms of spatial and spectral features, we were able to classify changes caused by phenology and those caused by human-induced disturbances. The method is summarized in six steps (Figure 3), which are described in detail in the following sections.

147 [Figure 3 near here]

#### 148 3.1. Image acquisition

We used two cloud-free OLI/Landsat-8 images to calculate NDVI and test our method:
one image was obtained on June 19<sup>th</sup>, 2015 (Figure 4a), and the other was obtained on

151 Oct. 27<sup>th</sup>, 2016 (Figure 4b). They were selected from the dry and rainy seasons to

152 maximize the effects of vegetation seasonality. We used the image acquired in June

153 2015 as representative of the end of the rainy season with high NDVI values. On the

154 other hand, the image acquired in October 2016 was used as representative of the end of

the dry season with comparatively lower NDVI values due to water stress (Figure 4c).
The images were downloaded from the United States Geological Survey (USGS) with
geometric and atmospheric corrections. We used NDVI (Rouse et al. 1973) because the
spatial domain of this index has been explored in several LULCC studies (Hamunyela et
al., 2016; Silveira et al. 2018a, 2018b). However, the proposed approach may be applied
to any index.

161

#### [Figure 4 near here]

#### 162 3.2. Image segmentation

163 The first procedure in the OBCD method was image segmentation. We applied the 164 multiresolution segmentation algorithm (Baatz and Schäpe 2000) from the eCognition 165 software (Definies 2009) selecting the original bands of the OLI/Landsat-8 images 166 acquired in 2015 and 2016 (years 1 and 2). This approach has the distinct advantage of 167 considering all images during object formation, thus minimizing sliver errors and 168 potentially honoring key multi-temporal boundaries (Desclée, Bogaert, and Defourny 169 2006; Tewkesbury et al. 2015). We used the following parameters: 0.1 for shape and 0.5 170 for compactness. The most critical step is the selection of the scale parameter (SP), 171 which controls the size of the image objects. The SP sets a homogeneity threshold that 172 determines the number of neighboring pixels that can be merged together to form an 173 image object (Benz et al. 2004). The SP directly influences the size of the objects which 174 are related to the predefined semivariogram criteria (lag distance) and the minimum 175 number of pixels inside each object necessary to generate the semivariogram. We 176 adopted a trial and error approach (Duro, Franklin, and Dube 2012) to find an 177 appropriate value for SP (X. Chen et al. 2015). We ensured a minimum number of 178 samples (25 pixels) inside the objects and an adequate size to allow calculation of the 179 semivariogram. The SP (set to 250) and image segmentation results were assessed based on visual inspection of the delineated polygons (Figure 5). The objects generated were
overlapped with the NDVI images from 2015 and 2016 to extract the input data for the
OBCD method.

183

#### [Figure 5 near here]

184

# 3.3. Class definition for change detection

185 This study focused on two broad classes: (*i*) vegetation covers with seasonal 186 changes in NDVI caused by phenology (Figure 6a); and (*ii*) vegetation covers with 187 changes caused by disturbances, especially human-induced deforestation/clearing 188 (Figure 6b) and fires (Figure 6c). Historically, most of the fires detected in the area have 189 been considered human-induced events. Therefore, we did not evaluate events of natural 190 occurrence.

191 Representative areas of these two classes were identified from visual inspection 192 of the images and from available land-cover maps. Randomly stratified design was used 193 to sample these areas (Olofsson et al. 2014). We first used a land-cover map (Carvalho 194 et al. 2006) showing the native vegetation for the 2006-2008 period to mask out the 195 non-vegetated areas. Subsequently, we performed post-classification and image edition 196 using a skilled human interpreter to update the available map to 2015 (Figure 1c). Thus, 197 a dataset of 300 objects (well-distributed polygons over the vegetated areas; 150 per 198 class) was obtained. The samples were randomly divided into training (50%) and 199 validation (50%) datasets (Figure 6). 200 [Figure 6 near here]

#### 201 3.4. Feature extraction

202 We extracted spatial and spectral features based on the NDVI values inside the objects.

203 The spatial information was obtained from experimental semivariograms (Equation 1),

where  $\gamma(h)$  is the estimator of the semivariance for each distance *h*, *N*(*h*) is the number of pairs of points (pixels) separated by distance *h*, *Z*(*x*) is the value of the regionalized variable at point x, and *Z*(*x*+*h*) is the value at point (*x*+*h*):

207 
$$\gamma(h) = \left(\frac{1}{2N(h)}\right) \sum_{i=1}^{N(h)} (Z(x) - Z(x+h))^2$$
 (1)

208 Semivariance functions are characterized by three parameters: sill ( $\sigma^2$ ), range ( $\varphi$ ) 209 and nugget effect ( $\tau^2$ ). The sill is the plateau reached by the semivariance values, 210 measuring the variance explained by the spatial structure of the data. The range is the 211 distance until the semivariogram reaches the sill, reflecting the distance at which the 212 data become correlated. The nugget effect is the non-spatial component of the variance 213 composed of random sensor noise or sampling errors (Curran 1988). We attempted to 214 find an optimal lag distance to ensure that sill values would provide a concise 215 description of data variability. We fixed the number of lags as 30 pixels and the lag size 216 equivalent to the image spatial resolution (30 m), resulting in a lag distance of 900 m. 217 We extracted a set of semivariogram indices (SIs) (Balaguer et al. 2010) using 218 the feature extraction software FETEX 2.0 for object-based image analysis (Ruiz et al. 219 2011) (Table 1). These indices describe the shape of the experimental semivariograms 220 and, therefore, the properties that characterize the spatial patterns of the image objects. 221 They have been categorized according to the position of the lags used in their definition: 222 (i) near the origin and (ii) up to the first maximum.

As described by Balaguer et al. (2010), the ratio between the values of the total variance and the semivariance at first lag (RVF) is an indicator of the relationship between the spatial correlation at long and short distances. The first derivative near the origin (FDO) represents the slope of the semivariogram at the first two lags. The second derivative semivariogram at the third lag (SDT) quantifies the concavity or convexity 228 level of the semivariogram at short distances, representing the heterogeneity of the 229 objects in the image. The mean of the semivariogram values up to the first maximum 230 (MFM) is an indicator of the average of the semivariogram values between the first lag 231 and the first maximum. It provides information about the changes in the data variability 232 and is related to the concavity or convexity of the semivariogram in that interval. The 233 difference between the mean of the semivariogram values up to the first maximum 234 (MFM) and the semivariance at first lag shows the decreasing rate of the spatial 235 correlation in the image up to the lags where the semivariogram theoretically tends to be 236 stabilized. Finally, the area between the semivariogram value in the first lag and the 237 semivariogram function until the first maximum (AFM) provides information about the 238 semivariogram curvature, which is also related to the variability of the data. 239 [Table 1 near here] 240 To explore the spectral information of the satellite images, we used the 241 minimum (MIN), mean (MEAN), maximum (MAX) and standard deviation (STDEV) 242 of the NDVI values inside each object. This allowed the performance of spatial and 243 spectral features to be compared and combined.

244 3.5. Change Detection using MLAs

245 After extracting the spatial and spectral features for each object, the differences in 246 NDVI values for each feature between years 1 (2015) and 2 (2016) were calculated and 247 used as input data to train the MLAs. The samples were randomly divided into training 248 (50%) and validation (50%) datasets. We used three MLAs implemented in the Waikato 249 Environment for Knowledge Analysis (WEKA 3.8 software): SVM, ANN and RF. 250 SVM has the ability to handle small training datasets, often producing higher 251 classification accuracies than traditional methods (Bovolo, Camps-Valls, and Bruzzone 252 2010; Mantero, Moser, and Serpico 2005; Wylie et al. 2018). For SVM, we used the

radial basis function (RBF) kernel, as this is known to be effective and accurate (Pereira et al. 2017; Shao and Lunetta 2012; Zuo, John, and Carranza 2011; Wu et al. 2015). To train the SVM classifier, an error parameter C (10) and a kernel parameter  $\gamma$  (0.1) were set after a series of tests and analyses of the outputs.

There are many different types of ANN, but the multilayer perceptron (MLP) is most commonly used in remote sensing (Berberoglu et al. 2000; Vafaei et al. 2018; Zhang et al. 2018). We used the ANN obtained by running the MLP function with the back-propagation algorithm (Pham, Yoshino, and Bui 2017). The main challenge associated with MLP is the adjustment of network parameters (Shao and Lunetta 2012). The learning rate, the momentum term, and iteration numbers were fixed at 0.3, 0.2 and 500, respectively (Tien Bui et al. 2016).

We also tested the non-parametric RF algorithm (Breiman 2001) because it has the ability to accommodate many predictor variables with accuracy and efficiency (Breiman 2001; DeVries et al. 2016; Ghimire, Rogan, and Miller 2010; Silveira et al. 2018a; Zhu et al. 2016). We set the number of decision trees (Ntree), to 500 (Lawrence, Wood, and Sheley 2006) and the number of variables for the best split when growing the trees (Mtry) to the default value (log of the number of features + 1) (Millard and Richardson 2015).

#### 271 **3.6.** Change Detection Evaluation

272 To evaluate our change detection using the three MLAs, we tested: (*i*) the spatial

273 domain of the NDVI images using the SIs; (ii) the spectral domain of the NDVI images

using the SFs; and (*iii*) the combination of the spatial-spectral attributes (SIs plus SFs).

275 We obtained a confusion matrix to evaluate classification accuracy for the two classes

under analysis: (a) vegetation covers with seasonal changes in NDVI caused by

277 phenology; and (b) vegetation covers with NDVI changes caused by human-induced

278 disturbances. We evaluated the overall, producer's and user's accuracies.

**4. Results** 

#### 280 4.1. Semivariogram analysis

From the use of semivariograms to quantify the spatial variability of the NDVI pixels

inside the objects, we found the maximum level of semivariance (sill  $-\sigma^2$ 

semivariogram parameter) at around 900 m. This indicated that at least 30 pixels and a

lag size equivalent to the image spatial resolution (30 m) were necessary to quantify

spatial variability of the OLI/Landsat-8 images. We detected two distinct patterns in the

semivariograms: (i) the shape and the overall data variability (sill –  $\sigma^2$ ) remained

287 constant over time with seasonal changes in NDVI caused by phenology (Figure 7a);

and (*ii*) the shape and sill increased in areas that experienced human-induced

disturbances between 2015 and 2016 (Figure 7b). These results indicated that the spatial

290 variability of NDVI quantified by semivariograms was very sensitive to changes in

291 vegetation cover caused by deforestation or fires. On the other hand, seasonal changes

in NDVI caused by vegetation phenology did not modify the shape and overall

293 variability of the semivariograms.

- 294 [Figure 7 near here]
- 295

## 4.2. Change Detection Evaluation

When we used the MLAs to classify areas with seasonal NDVI variations caused by

297 vegetation phenology and areas with NDVI variations caused by human-induced

298 disturbances, our results showed overall classification accuracies higher than 80% for

299 SVM, ANN and RF considering the spectral features and semivariogram indices (Table

300 2). Therefore, these classifiers and features were generally efficient to discriminate

301 areas of vegetation covers with seasonal changes in NDVI caused by phenology from302 other disturbance-affected areas.

303 The classification results using SFs (MIN, MEAN, MAX and STDEV) produced 304 the lowest accuracies, reaching values of 85.02%, 82.60% and 84.05% for SVM, ANN 305 and RF, respectively. The lowest user's and producer's accuracies were obtained using 306 this group of features (Table 2). In contrast, the overall classification accuracies slightly 307 improved when the semivariogram indices (RVF, FDO, SDT, MFM, DMF and AFM) 308 were included in the analysis, producing values of 87.43%, 83.09% and 85.99% for 309 SVM, ANN and RF, respectively. Thus, the semivariogram indices performance 310 slightly better than the spectral features, because they are related to the structured 311 variance of the NDVI pixel values. 312 A substantial gain in classification accuracy, reducing confusion between 313 vegetation phenological changes and human-induced disturbances, was obtained from 314 the combination of the SIs and the SFs (Table 2). The accuracies increased from 85.02 315 to 92.27%, 82.60 to 90.82% and 84.05 to 91.30% for SVM, ANN and RF, respectively. 316 The highest user's accuracy, considering both groups of features and all MLAs, was 317 observed for the class with changes controlled by vegetation phenology (95.33%). The 318 user's accuracies for this class improved significantly from 90.65 to 95.33% (SVM), 319 89.72 to 94.39 (ANN) and from 85.98 to 92.52% (RF). This was highly significant 320 because the objects with seasonal changes in NDVI presented low commission errors. 321 On the other hand, the highest producer's accuracy was observed for the class with 322 changes caused by human-induced disturbances having 94.68% for the SVM classifier. 323 [Table 2 near here] 324 The classification performance of the MLAs was generally comparable, but the

325 SVM algorithm was the most effective classifier in our TSBs. In Table 2, the SVM

326 presented the highest overall classification accuracy (92.27%). Using SFs or SIs as well 327 as the combination of these features, the accuracies were slightly superior for SVM than for ANN and RF. The differences in performance are probably due to the difficulties of 328 329 parameterization between the MLAs (García-Gutiérrez et al. 2015). SVM have been 330 frequently cited as a group of theoretically superior machine learning algorithms for 331 image classification and have been shown to perform well (Foody and Mathur 2004). 332 They appear to be especially advantageous in the presence of heterogeneous classes for 333 which only a few training samples are available (Li, Im, and Beier 2013; Wu et al. 2015). 334 The resultant SVM classification map, using the hybrid set of spatial and spectral features 335 from the OLI/Landsat-8 data, is shown in Figure 8. From the vegetated areas, 18.71% of 336 changes (331,830 ha) were caused by human-induced disturbances between 2015 and 337 2016.

338

#### [Figure 8 near here]

#### 339 **5. Discussion**

#### 340 5.1. Remote sensing change detection in TSBs

341 We proposed a new object-based method to detect changes caused by either vegetation 342 phenology or human-induced disturbances in TSBs, based on the differences over time 343 in spatial (semivariogram indices) and spectral features (descriptive statistics for 344 NDVI). Spatial and spectral features were used to train MLAs (SVM, ANN and RF). 345 Our results showed that the combination of both group of features produced the highest 346 overall classification, producer's and user's accuracies. 347 The method is an alternative to detect changes in TSBs, because it does not 348 require high-quality time series, which are sometimes difficult to obtain due to cloud 349 cover. This method could be used to improve the accuracy of LULCC maps, thus

350 providing better inputs for the assessment of atmospheric emissions derived from 351 deforestation and fires (Mouillot et al. 2014). TSBs present a conspicuous seasonal 352 contrast between the rainy and dry seasons (Ferreira and Huete 2004), which is 353 challenging for change detection. The seasonality of TSBs makes the use of optical 354 remote sensing difficult in some periods of the years due to cloud-cover and vegetation 355 phenology. Most of the change detection algorithms that are based on two dates of 356 Landsat images may reduce the influence of vegetation phenology on the analysis by 357 fixing data acquisition to a given period (Lu et al. 2004; Zhu, Woodcock, and Olofsson 358 2012). However, in TSBs in eastern Brazil, even fixing a pair of dates to the rainy or 359 dry season, the confounding effects of vegetation phenology on change detection persist 360 because of the irregular patterns of precipitation observed over space and time.

361 Some remote sensing studies have mapped deforestation and fire in TSBs 362 (Achard et al. 2014; Beuchle et al. 2015; Libonati et al. 2015; Hansen et al. 2013). For 363 example, Beuchle et al. (2015) provided information on historical and recent vegetation 364 cover changes in the Cerrado from central Brazil and the Caatinga from northeastern 365 Brazil based on the analysis of Landsat images from 1990 to 2010. For the Cerrado, 366 they estimated that 117,870 km<sup>2</sup> of vegetation was lost during the studied period, while for the Caatinga they reported a loss of 25,335 km<sup>2</sup>. When these results were compared 367 368 to LULCC estimates provided by other projects, such as the *Conservation and* 369 Sustainable Use of Brazilian Biological Diversity Project (PROBIO) and Deforestation 370 Monitoring in Brazilian Biomes Project (PMDBBS), some divergences were observed 371 (Beuchle et al. 2015). Although there were several factors that could introduce 372 differences in these estimates (e.g., spatial resolution, class definition), our findings 373 showed that the confounding effects of vegetation phenology on change detection

374 should be further considered as an important factor to avoid overestimation of human-375 induced disturbances.

# 376 5.2. Classification and change detection using the spatial-spectral domains of 377 NDVI

378 The spatial domain has been recently used to detect changes in tropical regions. The 379 phenological influence on data analysis is reduced when NDVI values are spatially 380 normalized in a pixel-based change detection approach (Hamunyela, Verbesselt, and 381 Herold 2016). The influence is also reduced when geostatistical features (spatial 382 domain) are incorporated into the analysis of bi-temporal NDVI images in an object-383 based change detection approach (Silveira et al. 2018b). Although the integration 384 between remote sensing and geostatistical theory was consolidated in the late 1980s, 385 only recent studies have demonstrated that the semivariogram (a geostatistical tool) has 386 strong potential for LULCC detection (Acerbi Junior et al. 2015; Gil-Yepes et al. 2016; Silveira et al. 2018a, 2018b). 387

388 Our study has demonstrated that the combination of spectral features and 389 semivariogram indices derived from bi-temporal NDVI images reduced the effects of 390 vegetation phenology on vegetation change detection. Misclassifications of seasonal 391 NDVI changes caused by vegetation phenology as those caused by human-induced 392 disturbances were therefore reduced. We found that LULCC areas caused by 393 deforestation or fires provided singular semivariograms with higher values for the sill 394 parameter than ones associated with vegetation phenology in savannas and semi-arid 395 woodlands. These results are in agreement with several previous studies that used 396 spatial information to detect changes (e.g. Acerbi Junior et al. 2015; Sertel, Kaya, and 397 Curran 2007; Silveira et al. 2018a, 2018b).

398 Acerbi Junior et al. (2015) analyzed the potential of semivariograms generated 399 from NDVI values to detect changes in Brazilian savannas. Their results showed a very 400 clear trend, where the shape of semivariograms, and the sill and range parameters were 401 different when deforestation occurred and were similar when there was no change in 402 land cover, which was consistent with our findings. Silveira et al. (2018a, 2018b) 403 highlight the importance of considering spatial information for change detection in 404 Brazilian savannas in the absence of a dense time series of remote-sensing images. 405 When using individual spatial features (e.g. sill parameter and the AFM index) the 406 change detection results were improved considerably compared with the spectral 407 features and image differencing technique. These results demonstrated that the 408 semivariograms derived from NDVI images are not affected by phenological changes. 409 Here, by including SIs that provided information near the origin (RVF, FDO and 410 SDT) and up to the first maximum (MFM, DMF and AFM), we obtained sufficient 411 separability between the classes of vegetation changes caused by phenology and human-412 induced disturbances. By combining SIs with SFs, the misclassification of these two 413 classes was reduced, as expressed by overall classification accuracies close to 90% for 414 the three classifiers (SVM, ANN and RF) (Table 2).

#### 415 **6.** Conclusions

We have proposed a new OBCD method to detect and distinguish vegetation changes caused by phenology from those caused by human-induced disturbances in Brazilian TSBs with pronounced seasonality. We reduced the effects of vegetation phenology on change detection by combining features from both the spatial and spectral domains of NDVI satellite images. The spatial variability of NDVI is not affect by vegetation seasonality, favoring the addition of semivariogram indices to reduce the impact of seasonality for detecting deforestation or fires using bi-temporal Landsat 423 images.

424 Compared with the other classifiers tested with this method, SVM presented a 425 slightly higher overall classification accuracy (92.27%) when using the hybrid set of 426 NDVI-derived spectral and spatial features. Finally, our study highlights that the 427 combination of the spatial and spectral attributes reduces the requirement for dense time 428 series of satellite imagery throughout multiple phenological cycles to detect LULCC in 429 TSBs. In these areas, vegetation exhibits strong seasonality and regularly-spaced 430 satellite images are difficult to obtain due to persistent cloud-cover. Future studies 431 should aim to evaluate further the proposed method, including its sensitivity to class and 432 intensity of disturbance, and its applicability to other TSBs.

#### 433 Acknowledgments

- 434 The authors would like to thank the *Coordenação de Aperfeiçoamento de Pessoal de*
- 435 *Nível Superior Brasil* (CAPES) for financing part of this study (Finance Code 001).
- 436 The authors are grateful for comments and suggestions by the anonymous reviewers.

#### 437 **Disclosure statement**

438 No potential conflict of interest.

#### 439 **References**

- 440 Acerbi Junior, F.W., E.M.O. Silveira, J.M. Mello, C.R. Mello, and J.R.S. Scolforo.
- 2015. "Change Detection in Brazilian Savannas Using Semivariograms Derived
  from NDVI Images." *Ciência e Agrotecnologia* 39 (2): 103–9.
  doi:10.1590/S1413-70542015000200001.
- 444 Achard, F., R. Beuchle, P. Mayaux, H. J. Stibig, C. Bodart, A. Brink, S. Carboni, et al.
  445 2014. "Determination of Tropical Deforestation Rates and Related Carbon
  446 Losses from 1990 to 2010." *Global Change Biology* 20 (8): 2540–54.
- 447 doi:10.1111/gcb.12605.

448 Baatz, M., and A. Schäpe. 2000. "Multiresolution Segmentation : An Optimization 449 Approach for High Quality Multi-Scale Image Segmentation." Journal of 450 Photogrammetry and Remote Sensing 58 (3–4): 12–23. 451 Balaguer, A., L. A. Ruiz, T. Hermosilla, and J. A. Recio. 2010. "Definition of a 452 Comprehensive Set of Texture Semivariogram Features and Their Evaluation for 453 Object-Oriented Image Classification." Computers and Geosciences 36 (2): 454 231-40. doi:10.1016/j.cageo.2009.05.003. 455 Benz, U. C., P. H., G. Willhauck, I. Lingenfelder, and M. Heynen. 2004. "Multi-456 Resolution, Object-Oriented Fuzzy Analysis of Remote Sensing Data for GIS-457 Ready Information." ISPRS Journal of Photogrammetry and Remote Sensing 58 458 (3-4): 239-58. doi:10.1016/j.isprsjprs.2003.10.002. 459 Berberoglu, S., C. D. Llovd, P. M. Atkinson, and P. J. Curran. 2000. "The Integration of 460 Spectral and Textural Information Using Neural Networks for Land Cover 461 Mapping in the Mediterranean." Computers and Geosciences 26 (4): 385-96. 462 doi:10.1016/S0098-3004(99)00119-3. 463 Beuchle, R., R. C. Grecchi, Y. E. Shimabukuro, R. Seliger, H. D. Eva, E. Sano, and F. 464 Achard. 2015. "Land Cover Changes in the Brazilian Cerrado and Caatinga 465 Biomes from 1990 to 2010 Based on a Systematic Remote Sensing Sampling 466 Approach." Applied Geography 58: 116–27. doi:10.1016/j.apgeog.2015.01.017. 467 Bovolo, F., G. Camps-Valls, and L. Bruzzone. 2010. "A Support Vector Domain 468 Method for Change Detection in Multitemporal Images." Pattern Recognition 469 Letters 31 (10): 1148–54. doi:10.1016/j.patrec.2009.07.002. 470 Breiman, Leo. 2001. "Random Forests." Machine Learning 45 (1): 5-32. 471 doi:10.1023/A:1010933404324. 472 Carvalho, L. M. T., J. R. S. Scolforo, A. D. Oliveira, J. M. Mello, L. T. Oliveira, F. W. 473 Acerbi Junior, H. C. Cavalcanti, and R. Vargas Filho. 2006. Procedimentos de 474 Mapeamento". In Mapeamento e Inventário da Flora e dos Reflorestamentos de 475 Minas Gerais; UFLA: Lavras, Brazil, 2006; pp. 37–57. 476 Chen, G., G. J. Hay, L. M. T. Carvalho, and M. A. Wulder. 2012. "Object-Based 477 Change Detection." International Journal of Remote Sensing 33 (14): 4434-478 4457. doi:10.1127/1432-8364/2011/0085. 479 Chen, G., Q. Weng, G. J. Hay, and Y. He. 2018. "Geographic Object-Based Image 480 Analysis (GEOBIA): Emerging Trends and Future Opportunities." GIScience & 481 Remote Sensing 55 (2): 159-182. (0). doi:10.1080/15481603.2018.1426092.

482	Chen, X., D. Yang, J. Chen, and X. Cao. 2015. "An Improved Automated Land Cover
483	Updating Approach by Integrating with Downscaled NDVI Time Series Data."
484	Remote Sensing Letters 6 (1): 29–38. doi:10.1080/2150704X.2014.998793.
485	Curran, P. J. 1988. "The Semivariogram in Remote Sensing: An Introduction." Remote
486	Sensing of Environment 24 (3): 493–507. doi:10.1016/0034-4257(88)90021-1.
487	Definies A.G. 2009. "Definiens ECognition Developer 8 User Guide.". Munich,
488	Germany.
489	Desclée, B., P. Bogaert, and P. Defourny. 2006. "Forest Change Detection by Statistical
490	Object-Based Method." Remote Sensing of Environment 102 (1-2): 1-11.
491	doi:10.1016/j.rse.2006.01.013.
492	DeVries, B., A. K. Pratihast, J. Verbesselt, L. Kooistra, and M. Herold. 2016.
493	"Characterizing Forest Change Using Community-Based Monitoring Data and
494	Landsat Time Series." PLoS ONE 11 (3): 1-25.
495	doi:10.1371/journal.pone.0147121.
496	Duro, D. C., S. E. Franklin, and M. G. Dube. 2012. "A Comparison of Pixel-Based and
497	Object-Based Image Analysis with Selected Machine Learning Algorithms for
498	the Classification of Agricultural Landscapes Using SPOT-5 HRG Imagery."
499	Remote Sensing of Environment 118: 259–72. doi:10.1016/j.rse.2011.11.020.
500	Espírito-Santo, M. M., M. E. Leite., J. O. Silva, R. S. Barbosa, A. M. Rocha, F. C.
501	Anaya, and M. G. V. Dupin. 2016. "Understanding Patterns of Land-Cover
502	Change in the Brazilian Cerrado from 2000 to 2015." Philosophical
503	Transactions of The Royal Society B Biological Sciences (371): 1–11.
504	doi:10.1098/rstb.2015.0435.
505	Evans, J., and R. Geerken. 2004. "Discrimination between Climate and Human-Induced
506	Dryland Degradation." Journal of Arid Environments 57: 535-54.
507	doi:10.1016/S0140-1963(03)00121-6.
508	Ferreira, L. G., and A. R. Huete. 2004. "Assessing the Seasonal Dynamics of the
509	Brazilian Cerrado Vegetation through the Use of Spectral Vegetation Indices."
510	International Journal of Remote Sensing 25 (10): 1837–60.
511	doi:10.1080/0143116031000101530.
512	Ferreira, L. G., H. Yoshioka, A. Huete, and E. E. Sano. 2004. "Optical Characterization
513	of the Brazilian Savanna Physiognomies for Improved Land Cover Monitoring
514	of the Cerrado Biome: Preliminary Assessments from an Airborne Campaign

515 over an LBA Core Site." Journal of Arid Environments 56 (3): 425-447. 516 doi:10.1016/S0140-1963(03)00068-5. 517 Foody, G. M., and A.Mathur. 2004. "A Relative Evaluation of Multiclass Image 518 Classification by Support Vector Machines." IEEE Transactions on Geoscience 519 and Remote Sensing 42 (6): 1335-1343. doi:10.1109/TGRS.2004.827257. 520 García-Gutiérrez, J., D. Mateos-García, M. Garcia, and J. C. Riquelme-Santos. 2015. 521 "An Evolutionary-Weighted Majority Voting and Support Vector Machines 522 Applied to Contextual Classification of LiDAR and Imagery Data Fusion." 523 Neurocomputing 163: 17-24. doi:10.1016/j.neucom.2014.08.086. 524 Ghimire, B., J. Rogan, and J. Miller. 2010. "Contextual Land-Cover Classification: 525 Incorporating Spatial Dependence in Land-Cover Classification Models Using 526 Random Forests and the Getis Statistic." Remote Sensing Letters 1 (1): 45–54. 527 doi:10.1080/01431160903252327. 528 Gil-Yepes, J. L., L. A. Ruiz, J. A. Recio, A. Balaguer-Beser, and T. Hermosilla. 2016. 529 "Description and Validation of a New Set of Object-Based Temporal 530 Geostatistical Features for Land-Use/Land-Cover Change Detection." ISPRS 531 Journal of Photogrammetry and Remote Sensing 121: 77–91. 532 doi:10.1016/j.isprsjprs.2016.08.010. 533 Hamunyela, E., J. Reiche, J. Verbesselt, and M. Herold. 2017. "Using Space-Time 534 Features to Improve Detection of Forest Disturbances from Landsat Time 535 Series." Remote Sensing 9 (6):1-17. doi:10.3390/rs9060515. 536 Hamunyela, E., J. Verbesselt, and M. Herold. 2016. "Using Spatial Context to Improve 537 Early Detection of Deforestation from Landsat Time Series." Remote Sensing of 538 Environment 172: 126-38. doi:10.1016/j.rse.2015.11.006. 539 Hansen, M. C., P. V. Potapov, R. Moore, M. Hancher, S. A. Turubanova, A. Tyukavina, 540 D. Thau, et al. 2013. "High-Resolution Global Maps of 21st-Century Forest 541 Cover Change." Science 134 (2011): 850–53. doi:10.1126/science.1244693. 542 Healey, S. P., W. B. Cohen, Z. Yang, C. K. Brewer, E. B. Brooks, N. Gorelick, A. J. 543 Hernandez, et al. 2018. "Mapping Forest Change Using Stacked Generalization: 544 An Ensemble Approach." Remote Sensing of Environment 204: 717–28. 545 doi:10.1016/j.rse.2017.09.029. 546 Hoekstra, J. M., T. M. Boucher, T. H. Ricketts, and C. Roberts. 2005. "Confronting a 547 Biome Crisis: Global Disparities of Habitat Loss and Protection." Ecology 548 Letters 8 (1): 23–29. doi:10.1111/j.1461-0248.2004.00686.x.

540	Human C. C. N. Coward, I. C. Massle, N. Thomas, 7. 7hy and I. F. Masslenger, 2010
549	Huang, C., S. N Goward, J. G. Masek, N. Thomas, Z. Zhu, and J. E. Vogelmann. 2010.
550	"Remote Sensing of Environment An Automated Approach for Reconstructing
551	Recent Forest Disturbance History Using Dense Landsat Time Series Stacks."
552	Remote Sensing of Environment 114 (1): 183–98. doi:10.1016/j.rse.2009.08.017.
553	Hughes, M. J., S. D. Kaylor, and D. J. Hayes. 2017. "Patch-Based Forest Change
554	Detection from Landsat Time Series." Forests 8 (166): 1-22.
555	doi:10.3390/f8050166.
556	Ibrahim, Y. Z, H. Balzter, J. Kaduk, and C. J. Tucker. 2015. "Land Degradation
557	Assessment Using Residual Trend Analysis of GIMMS NDVI3g, Soil Moisture
558	and Rainfall in Sub-Saharan West Africa from 1982 to 2012." Remote sensing 7:
559	5471–94. doi:10.3390/rs70505471.
560	Kennedy, R. E, Z. Yang, and W. B Cohen. 2010. "Remote Sensing of Environment
561	Detecting Trends in Forest Disturbance and Recovery Using Yearly Landsat
562	Time Series : 1 . LandTrendr — Temporal Segmentation Algorithms." Remote
563	Sensing of Environment 114 (12): 2897–2910. doi:10.1016/j.rse.2010.07.008.
564	Lagomasino, D., R. M. Price, D. Whitman, P. K. E. Campbell, and A. Melesse. 2014.
565	"Estimating Major Ion and Nutrient Concentrations in Mangrove Estuaries in
566	Everglades National Park Using Leaf and Satellite Reflectance." Remote
567	Sensing of Environment 154: 202–18. doi:10.1016/j.rse.2014.08.022.
568	Laliberte, A. S., and A. Rango. 2009. "Texture and Scale in Object-Based Analysis of
569	Subdecimeter Resolution Unmanned Aerial Vehicle (UAV) Imagery." IEEE
570	Transactions on Geoscience and Remote Sensing 47 (3): 761–70.
571	doi:10.1109/TGRS.2008.2009355.
572	Lawrence, R. L., Shana D. W., and R. L. Sheley. 2006. "Mapping Invasive Plants Using
573	Hyperspectral Imagery and Breiman Cutler Classifications (RandomForest)."
574	<i>Remote Sensing of Environment</i> 100 (3): 356–62. doi:10.1016/j.rse.2005.10.014.
575	Leal, I. R, J. M. C. Silva, M. Tabarelli, and T. E. Lacher. 2005. "Changing the Course
576	of Biodiversity Conservation in the Caatinga of Northeastern Brazil."
577	<i>Conservation Biology</i> 19 (3): 701–6. doi:10.1111/j.1523-1739.2005.00703.x.
578	Li, M., J. Im. and C. Beier. 2013. "Machine Learning Approaches for Forest
579	Classification and Change Analysis Using Multi-Temporal Landsat TM Images
580	over Huntington Wildlife Forest " Giscience & Remote Sensing 50 (4): 361-84
581	doi:Doi 10 1080/15481603 2013 819161
<b>U</b> U1	

Libonati, R., C. C. Camara, A. W. Setzer, F. Morelli, and A. E. Melchiori. 2015. "An 582 583 Algorithm for Burned Area Detection in the Brazilian Cerrado Using 4 Mm 584 MODIS Imagery." Remote Sensing 7 (11): 15782-803. doi:10.3390/rs71115782. 585 Lu, D., P. Mausel, E. Brondízio, and E. Moran. 2004. "Change Detection Techniques." 586 International Journal of Remote Sensing 25 (12): 2365–2407. 587 doi:10.1080/0143116031000139863. 588 Mantero, P., G. Moser, and S. B. Serpico. 2005. "Partially Supervised Classification of 589 Remote Sensing Images through SVM-Based Probability Density Estimation." 590 IEEE Transactions on Geoscience and Remote Sensing 43 (3): 559–70. 591 doi:10.1109/TGRS.2004.842022. 592 Millard, K., and M. Richardson. 2015. "On the Importance of Training Data Sample 593 Selection in Random Forest Image Classification: A Case Study in Peatland Ecosystem Mapping." Remote Sensing 7 (7): 8489-8515. 594 595 doi:10.3390/rs70708489. 596 Mouillot, F., M. G. Schultz, C. Yue, P. Cadule, K. Tansey, P. Ciais, and E. Chuvieco. 597 2014. "Ten Years of Global Burned Area Products from Spaceborne Remote 598 Sensing-A Review: Analysis of User Needs and Recommendations for Future 599 Developments." International Journal of Applied Earth Observation and 600 Geoinformation 26 (1): 64-79. doi:10.1016/j.jag.2013.05.014. 601 Olofsson, P., G. M. Foody, M. Herold, S. V. Stehman, C. E. Woodcock, and M. A 602 Wulder. 2014. "Good Practices for Estimating Area and Assessing Accuracy of 603 Land Change." Remote Sensing of Environment 148: 42-57. 604 doi:10.1016/j.rse.2014.02.015. 605 Peel, M. C., B. L. Finlayson, and T. A. McMahon. 2006. "Updated World Map of the 606 K oppen-Geiger Climate Classification." Meteorologische Zeitschrift 15: 259-607 63. doi:10.1127/0941-2948/2006/0130. 608 Pereira, A. A., J. M.C. Pereira, R. Libonati, D. O., A. W. Setzer, F. Morelli, F. 609 Machado-Silva, and L. M. T. Carvalho. 2017. "Burned Area Mapping in the 610 Brazilian Savanna Using a One-Class Support Vector Machine Trained by 611 Active Fires." Remote Sensing 9 (11). doi:10.3390/rs9111161. 612 Pham, T. D., K. Yoshino, and D. Tien Bui. 2017. "Biomass Estimation of Sonneratia 613 Caseolaris (l.) Engler at a Coastal Area of Hai Phong City (Vietnam) Using 614 ALOS-2 PALSAR Imagery and GIS-Based Multi-Layer Perceptron Neural

- 615 Networks." *GIScience and Remote Sensing* 54 (3): 329–53.
- 616 doi:10.1080/15481603.2016.1269869.
- Rouse, J. W., R. H. Hass, J.A. Schell, and D.W. Deering. 1973. "Monitoring Vegetation
  Systems in the Great Plains with ERTS." *Third Earth Resources Technology Satellite (ERTS) Symposium* 1: 309–17. doi:citeulike-article-id:12009708.
- Ruiz, L. A., J. A. Recio, A. Fernández-Sarría, and T. Hermosilla. 2011. "A Feature
  Extraction Software Tool for Agricultural Object-Based Image Analysis." *Computers and Electronics in Agriculture* 76 (2): 284–96.

623 doi:10.1016/j.compag.2011.02.007.

- Sano, E. E, R. R., J. L. S. Brito, and L. G. Ferreira. 2010. "Land Cover Mapping of the
  Tropical Savanna Region in Brazil." *Environmental Monitoring and Assessment*166: 113–24. doi:10.1007/s10661-009-0988-4.
- Santos, R. M., A. T. Oliveira-Filho, P. V. Eisenlohr, L. P. Queiroz, D. B. O. S. Cardoso,
  and M. J. N. Rodal. 2012. "Identity and Relationships of the Arboreal Caatinga
  among Other Floristic Units of Seasonally Dry Tropical Forests (SDTFs) of
  North-Eastern and Central Brazil." *Ecology and Evolution* 2 (2): 409–28.
  doi:10.1002/ece3.91.
- Scolforo, H. F., J. R. S. Scolforo, C. R. Mello, J. M. Mello, and A. C. Ferraz Filho.
  2015. "Spatial Distribution of Aboveground Carbon Stock of the Arboreal
  Vegetation in Brazilian Biomes of Savanna, Atlantic Forest and Semi-Arid
- 635 Woodland." *PLoS ONE* 10 (6): 1–20. doi:10.1371/journal.pone.0128781.
- 636 Sertel, E., S. Kaya, and P. J. Curran. 2007. "Use of Semivariograms to Identify
  637 Earthquake Damage in an Urban Area." *IEEE Transactions on Geoscience and*638 *Remote Sensing* 45 (6): 1590–94. doi:10.1109/TGRS.2007.894019.
- 639 Shao, Y., and R. S. Lunetta. 2012. "Comparison of Support Vector Machine, Neural
  640 Network, and CART Algorithms for the Land-Cover Classification Using
- 641 Limited Training Data Points." *ISPRS Journal of Photogrammetry and Remote*642 *Sensing* 70: 78–87. doi:10.1016/j.isprsjprs.2012.04.001.
- 643 Silva, J. F., M. R. Fariñas, J. M. Felfili, and C. A. Klink. 2006. "Spatial Heterogeneity,
  644 Land Use and Conservation in the Cerrado Region of Brazil." *Journal of*645 *Biogeography* 33 (3): 536–48. doi:10.1111/j.1365-2699.2005.01422.x.
- 646 Silveira, E. M. O., I. T. Bueno, F. W. Acerbi Júnior, J. M. Mello, J. R. S. Scolforo, and
- 647 M. A Wulder. 2018a. "Using Spatial Features to Reduce the Impact of

648	Seasonality for Detecting Tropical Forest Changes from Landsat Time Series."
649	Remote Sensing 10 (808): 1-21. doi:10.3390/rs10060808.
650	Silveira, E. M.O., J. M. Mello, F. W. Acerbi Júnior, and L. M. T. Carvalho. 2018b.
651	"Object-Based Land-Cover Change Detection Applied to Brazilian Seasonal
652	Savannahs Using Geostatistical Features." International Journal of Remote
653	Sensing 39 (8): 2597–2619. doi:10.1080/01431161.2018.1430397.
654	Silveira, E.M.O., M.D. Menezes, F.W. Acerbi Júnior, M.C. N. S. Terra, and J.M. Mello.
655	2017. "Assessment of Geostatistical Features for Object-Based Image
656	Classification of Contrasted Landscape Vegetation Cover." Journal of Applied
657	Remote Sensing 11 (3): 1-15. doi:10.1117/1.JRS.11.036004.
658	Tewkesbury, A. P., A. J. Comber, N. J. Tate, A. Lamb, and P. F. Fisher. 2015. "A
659	critical synthesis of remotely sensed optical image change detection techniques.'
660	Remote Sensing of Environment 160: 1-14. Doi.10.1016/j.rse.2015.01.006.
661	Tien Bui, D., T. A. Tuan, H. Klempe, B. Pradhan, and I. Revhaug. 2016. "Spatial
662	Prediction Models for Shallow Landslide Hazards: A Comparative Assessment
663	of the Efficacy of Support Vector Machines, Artificial Neural Networks, Kernel
664	Logistic Regression, and Logistic Model Tree." Landslides 13 (2): 361-78.
665	doi:10.1007/s10346-015-0557-6.
666	Vafaei, S., J. Soosani, K. Adeli, H. Fadaei, H. Naghavi, T. D. Pham, and D. Tien Bui.
667	2018. "Improving Accuracy Estimation of Forest Aboveground Biomass Based
668	on Incorporation of ALOS-2 PALSAR-2 and Sentinel-2A Imagery and Machine
669	Learning: A Case Study of the Hyrcanian Forest Area (Iran)." Remote Sensing
670	10 (2): 1-21. doi:10.3390/rs10020172.
671	Verbesselt, J., R. Hyndman, G. Newnham, and D. Culvenor. 2010. "Detecting Trend
672	and Seasonal Changes in Satellite Image Time Series." Remote Sensing of
673	Environment 114 (1): 106-15. doi:10.1016/j.rse.2009.08.014.
674	Wright, S. J., and C. P. Van Schaik. 1994. "Light and the Phenology of Tropical Trees."
675	The American Naturalist 143 (1): 192–99.
676	Wu, X., J. Peng, J. Shan, and W. Cui. 2015. "Evaluation of Semivariogram Features for
677	Object-Based Image Classification." Geo-Spatial Information Science 18 (4):
678	159–170. doi:10.1080/10095020.2015.1116206.
679	Wyle, B. K., N. J. Pastick, J. J. Picote, and C. A. Deering. 2018. "Geospatial data
680	mining for digital raster mapping." GIScience & Remote Sensing 1-24.
681	Doi:10.1080/15481603.2018.1517445.

682	Zhang, C., X. Pan, H. Li, A. Gardiner, I. Sargent, J. Hare, and P. M. Atkinson. 2018. "A
683	Hybrid MLP-CNN Classifier for Very Fine Resolution Remotely Sensed Image
684	Classification." ISPRS Journal of Photogrammetry and Remote Sensing 140:
685	133-144. doi:10.1016/j.isprsjprs.2017.07.014.
686	Zhang, K., M. Ross, and D. Gann. 2016. "Remote Sensing of Seasonal Changes and
687	Disturbances in Mangrove Forest : A Case Study from South Florida."
688	<i>Ecosphere</i> 7: 1–23. doi:10.1002/ecs2.1366.
689	Zhu, Z. 2017. "Change Detection Using Landsat Time Series : A Review of
690	Frequencies, Preprocessing, Algorithms, and Applications." ISPRS Journal of
691	Photogrammetry and Remote Sensing 130: 370–84.
692	doi:10.1016/j.isprsjprs.2017.06.013.
693	Zhu, Z., Y. Fu, C. E. Woodcock, P. Olofsson, J. E. Vogelmann, C. Holden, M. Wang, S.
694	Dai, and Y. Yu. 2016. "Including Land Cover Change in Analysis of Greenness
695	Trends Using All Available Landsat 5, 7, and 8 Images: A Case Study from
696	Guangzhou, China (2000–2014)." Remote Sensing of Environment 185: 243–57.
697	doi:10.1016/j.rse.2016.03.036.
698	Zhu, Z., C. E. Woodcock, and P. Olofsson. 2012. "Continuous Monitoring of Forest
699	Disturbance Using All Available Landsat Imagery." Remote Sensing of
700	Environment 122: 75–91. doi:10.1016/j.rse.2011.10.030.
701	Zhu, Z., and C. E. Woodcock. 2014. "Continuous Change Detection and Classification
702	of Land Cover Using All Available Landsat Data." Remote Sensing of
703	Environment 144: 152-71. doi:10.1016/j.rse.2014.01.011.
704	Zuo, R., E. John, and M. Carranza. 2011. "Computers & Geosciences Support Vector
705	Machine: A Tool for Mapping Mineral Prospectivity." Computers and
706	Geosciences 37 (12): 1967-75. doi:10.1016/j.cageo.2010.09.014.
707	
708	
709	
710	
711	
712	
713	

- 714 Table 1. Semivariogram indices (SIs) calculated from the NDVI values inside the
- objects near the origin (\*) or up to the first maximum (\*\*). The semivariogram features 715
- $\{(h1, \gamma 1), (h2, \gamma 2) \dots (hn, \gamma n)\}$  are the points of the experimental semivariogram, as 716
- 717 described by Balaguer et al. (2010). The lags {hi, h2 ... hn} are equally spaced.
- 718 Variance is the value of the total variance of the pixels belonging to the object. hmax\_1

\_\_\_\_\_

- 719 represents the location of the first local maximum, while  $\gamma(\text{hmax } 1)$  is the first local
- 720 maximum semivariance.

Description	Formula
*Ratio between the values total variance and the semivariance at first lag	of the $RVF = \frac{Variance}{\gamma l}$
*First derivative near the o	rigin $FDO = \frac{\gamma 2 - \gamma 1}{h}$
*Second derivative at third	lag $SDT = \frac{\gamma 4 - 2\gamma 3 + \gamma 2}{h^2}$
**Mean of the semivariog values up to the first maxir **Difference between the	$MFM = \frac{1}{Max_{1}} \sum \gamma i$ mean
values up to the first maximand the semivariance at first	num DMF=MFM- γi st lag
**Semivariance curvature	$AFM = \frac{h}{2} \left( \gamma 1 + 2 \left( \sum_{i=2}^{\max_{l=1}} \gamma 1 \right) + \gamma_{\max_{l}} \right) - \left( \gamma 1 \left( h_{\max_{l}} - h 1 \right) \right)$
721	
722	
723	
724	
725	
726	
727	
728	
729	
730	
731	
732	

- 733 Table 2. Confusion matrix from the classification of areas with seasonal NDVI changes
- caused by vegetation phenology and those due to human-induced disturbances. Spectral
- features (SFs), semivariogram indices (SIs) and their combination (SFs + SIs) were
- vised for change detection. The Producer's (PA), User's (UA) and overall (OA)
- 737 classification accuracies are shown for support vector machine (SVM), artificial neural
- 738 network (ANN) and random forest (RF).

Vegetation change class         PA (%)         UA (%)         PA (%)			S	F	S	SI		SF + SI	
Change caused by disturbance         88.76         79.00         94.05         79.00         94.68         90           SVM         Change caused by phenology         82.20         90.65         82.93         95.33         90.27         95.3           OA (%)         85.02         87.43         92.27           Change caused by disturbance         87.21         75.00         88.24         75.00         93.55         87.00           ANN         Change caused by phenology         79.34         89.72         79.51         90.65         88.60         94.33           OA (%)         82.6         83.09         90.82         90.82         90.82         90.82         90.82         90.82         90.82         90.82         90.82         90.82         90.82         90.82         90.82         90.82         90.82         90.82         90.82         90.83         92.53         90.83         92.53         90.83         92.53         90.83         92.53         90.83         92.53         90.83         92.53         90.83         92.53         90.83         92.53         90.83         92.53         90.83         92.53         90.83         92.53         90.83         92.53         90.83         92.53         90.83 </th <th></th> <th>Vegetation change class</th> <th>PA (%)</th> <th>UA (%)</th> <th>PA (%)</th> <th>UA (%)</th> <th>PA (%)</th> <th>UA (%)</th>		Vegetation change class	PA (%)	UA (%)	PA (%)	UA (%)	PA (%)	UA (%)	
SVM         Change caused by phenology         82.20         90.65         82.93         95.33         90.27         95.3           OA (%)         85.02         87.43         92.27           Change caused by disturbance         87.21         75.00         88.24         75.00         93.55         87.0           ANN         Change caused by phenology         79.34         89.72         79.51         90.65         88.60         94.33           OA (%)         82.6         83.09         90.82           Change caused by disturbance         84.54         82.00         87.37         83.00         91.84         90.00           RF         Change caused by disturbance         83.64         85.98         84.82         88.79         90.83         92.51           OA (%)         84.05         85.99         91.3         91.3         91.3         91.3		Change caused by disturbance	88.76	79.00	94.05	79.00	94.68	90	
OA (%)         85.02         87.43         92.27           ANN         Change caused by disturbance         87.21         75.00         88.24         75.00         93.55         87.00           ANN         Change caused by phenology         79.34         89.72         79.51         90.65         88.60         94.3           OA (%)         82.6         83.09         90.82           Change caused by disturbance         84.54         82.00         87.37         83.00         91.84         90.00           RF         Change caused by phenology         83.64         85.98         84.82         88.79         90.83         92.51           OA (%)         84.05         85.99         91.3         91.3         91.3	SVM	Change caused by phenology	82.20	90.65	82.93	95.33	90.27	95.33	
Change caused by disturbance         87.21         75.00         88.24         75.00         93.55         87.0           ANN         Change caused by phenology         79.34         89.72         79.51         90.65         88.60         94.3           OA (%)         82.6         83.09         90.82           Change caused by disturbance         84.54         82.00         87.37         83.00         91.84         90.00           RF         Change caused by phenology         83.64         85.98         84.82         88.79         90.83         92.51           OA (%)         84.05         85.99         91.3         91.3         91.3         91.3		OA (%)	85	.02	87	.43	92	.27	
ANN         Change caused by phenology         79.34         89.72         79.51         90.65         88.60         94.3           OA (%)         82.6         83.09         90.82           Change caused by disturbance         84.54         82.00         87.37         83.00         91.84         90.0           RF         Change caused by phenology         83.64         85.98         84.82         88.79         90.83         92.53           OA (%)         84.05         85.99         91.3		Change caused by disturbance	87.21	75.00	88.24	75.00	93.55	87.00	
OA (%)       82.6       83.09       90.82         RF       Change caused by disturbance       84.54       82.00       87.37       83.00       91.84       90.00         RF       Change caused by phenology       83.64       85.98       84.82       88.79       90.83       92.52         OA (%)       84.05       85.98       85.99       91.3       91.3	ANN	Change caused by phenology	79.34	89.72	79.51	90.65	88.60	94.39	
Change caused by disturbance         84.54         82.00         87.37         83.00         91.84         90.00           RF         Change caused by phenology         83.64         85.98         84.82         88.79         90.83         92.5           OA (%)         84.05         85.99         91.3		OA (%)	82	2.6	83	.09	90.82		
RF         Change caused by phenology         83.64         85.98         84.82         88.79         90.83         92.5           OA (%)         84.05         85.99         91.3		Change caused by disturbance	84.54	82.00	87.37	83.00	91.84	90.00	
OA (%) 84.05 85.99 91.3	RF	Change caused by phenology	83.64	85.98	84.82	88.79	90.83	92.52	
		OA (%)	84.05		85.99		91.3		

Figure 1. (a) Location of the study area in the state of Minas Gerais (MG), southeastern Brazil. The area is covered by savannas and semi-arid woodlands; (b) False color composite from an OLI/Landsat-8 image from 27 October 2016; (c) Land-cover map showing vegetated and non-vegetated surfaces.

- Figure 2. OLI/Landsat-8 false color composite (bands 5, 4 and 3 in RGB) from year 1
- 756 (19 June 2015) and year 2 (27 October 2016) showing examples of vegetation types
- found in the study area. (a) grassland (open grassland); (b) shrub savanna (open
- 758 grassland with sparse shrubs); (c) woodland savanna (mixed grassland, shrublands and
- trees up to seven meters in height); (d) palm swaps (riparian vegetation); (e)
- 760 semideciduous forest (semideciduous canopy foliage); and (f) deciduous forest
- 761 (predominance of deciduous trees whose loss of foliage reaches more than 50%).
- Figure 3. The six main steps in the methodology used to reduce the effects of seasonal
- 763 NDVI changes caused by vegetation phenology on the detection of changes caused by
- human-induced disturbances in tropical seasonal biomes (TSBs) in Brazil.
- Figure 4. (a) NDVI OLI/Landsat-8 image from June 19<sup>th</sup>, 2015; (b) NDVI OLI/Landsat-
- 766 8 image from Oct. 27<sup>th</sup>, 2016; (c) monthly precipitation pattern from years 2015, 2016
- and historical series of precipitation from year 1952 to 2018.
- Figure 5. Image segmentation results using 0.1 for shape, 0.5 for compactness and 250for the scale parameter (SP).
- Figure 6. The location of the training and validation samples is shown at the top of the
- figure. The OLI/Landsat-8 false color composites (bands 5, 4 and 3 in RGB) show
- examples of the classes defined for change detection analysis between the rainy and dry
- seasons of 2015 (year 1) and 2016 (year 2). Seasonal variations caused by vegetation
- phenology are shown in (a), while human-induced changes caused by deforestation and
- fires are illustrated in (b) and (c), respectively.
- Figure 7. Patterns of semivariograms generated from the NDVI values inside the objects
- for years 1 (2015) and 2 (2016): (a) NDVI changes caused by vegetation phenology –
- the shape and sill ( $\sigma^2$ ) parameters remained constant; (b) NDVI changes caused by
- human-induced disturbances the shape and sill ( $\sigma^2$ ) parameters increased.

- 780 Figure 8. Support vector machine (SVM) classification using spectral features (SFs) and
- 781 semivariogram indices (SIs), showing changes caused by vegetation phenology and
- human-induced disturbance between 2015 and 2016.