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Ref: Ms. No. JEMA-D-18-00511: Former title: “**New insights on CH₄ mass transfer characterization in stirred tank bioreactors devoted to CH₄ bioconversion**”, current title: “**A systematic comparison of two empirical gas-liquid mass transfer determination methodologies to characterize methane biodegradation in stirred tank bioreactors**”, submitted for publication in the special issue ISEAC5-Asia in **Journal of Environmental Management**.

Dear editor,

The authors would like to thank you for the attention given to our manuscript during this peer-review process. We have carefully addressed all reviewers and editor comments, and therefore, the quality of the paper has been significantly improved. More specifically:

REVIEWER 1

Muñoz et al. compared the potential of two experimental methodologies for the estimation of $k_{la}CH_4$ at multiple stirring rates (400, 600 and 800 rpm). These two methodologies are somewhat useful. Although authors have used a simple methodology to accurately describe CH₄ mass transport in continuous bioreactors. I recommend it for the publication only after major changes for the following reasons:

The authors acknowledge the positive feedback from Reviewer 1

1. Research highlights need to be re-refined.

The research highlights were reworded in order to better reflect the main outcomes of the present study in accordance to Reviewer 1 request (see highlight section).

*k_{laO_2} determination based on SO_3^{2-} oxidation accurately described CH₄ mass transfer
Microbial CH₄ uptake accelerated CH₄ mass transport similarly to SO_3^{2-} oxidation
The gassing-out methodology underestimated CH₄ mass transport
Maximum CH₄ elimination capacities of $62 \pm 5 \text{ g CH}_4 \text{ m}^{-3} \text{ h}^{-1}$ were achieved at 800 rpm*

2. Abstract and conclusions are written very poorly, I would suggest authors' to modify carefully. Provide more systematic explanation and important results. A complete re-write is necessary.

The abstract and conclusion sections were significantly modified in order to highlight the main results and provide a more systematic explanation of the findings obtained as requested by Reviewer 1. In addition, the length of both sections was reduced to ≥ 150 words as requested by the guest editor (current pages 2 and 14-15).

Abstract (150 words)

“This study aimed at systematically comparing the potential of two empirical methods for the estimation of the volumetric CH₄ mass transfer coefficient (k_{lCH_4}), namely gassing-out and oxygen transfer rate (OTR), to describe CH₄ biodegradation in a fermenter operated with a methanotrophic consortium at 400, 600 and 800 rpm. The k_{lCH_4} estimated from the OTR methodology accurately predicted the CH₄ elimination capacity (EC) under CH₄ mass transfer limiting conditions regardless of the stirring rate (~ 9 % of average error between empirical and estimated ECs). Thus, empirical CH₄-ECs of 37.8±5.8, 42.5±5.4 and 62.3±5.2 g CH₄ m⁻³ h⁻¹ vs predicted CH₄-ECs of 35.6±2.2, 50.1±2.3 and 59.6±3.4 g CH₄ m⁻³ h⁻¹ were recorded at 400, 600 and 800 rpm, respectively. The rapid Co²⁺-catalyzed reaction of O₂ with SO₃²⁻ in the vicinity of the gas-liquid interphase during OTR determinations, mimicking microbial CH₄ uptake in the biotic experiments, was central to accurately describe the k_{lCH_4} .”

Conclusions (119)

“This work empirically validated a simple methodology to accurately describe CH₄ mass transport and biodegradation in continuous stirred tank bioreactors based on OTR estimations. The rapid O₂ uptake at the gas-liquid interphase mediated by sulfite and Co²⁺, mimicking microbial CH₄ uptake during continuous CH₄ treatment, supported the superior performance of the OTR methodology for the description of k_{lCH_4} compared to the classical gassing-out methodology. Thus, the empirical CH₄-ECs of 37.8±5.8, 42.5±5.4 and 62.3±5.2 g CH₄ m⁻³ h⁻¹ recorded at 400, 600 and 800 rpm, respectively, were comparable to the CH₄-ECs of 35.6±2.2, 50.1±2.3 and 59.6±3.4 g CH₄ m⁻³ h⁻¹ predicted by the OTR methodology. This finding will help in the design of suspended growth bioreactors devoted to CH₄ bioconversion.”

4. Authors should compare two methods. Figure 2 is simply comparison at three rates (400, 600 and 800 rpm), how about 1000 rpm or 1200 rpm? What would happen at high rates.

The authors agree with Reviewer 1 on the additional value of the information provided by tests conducted at 1000 and 1200 rpm. Unfortunately, these experiments were not conducted based on the fact that these agitation rates would be hardly applicable in full scale fermenters. In our humble opinion, the main findings of this study were validated with the experiments carried out at 400, 600 and 800 rpm. An explanatory remark was included in current page 5.

“These aeration rates were selected as potential operational conditions implemented in large-scale fermenters devoted to the bioconversion of CH₄.”

3. The authors' interpretation of the methodologies is weak. Many sentences are merely left to the speculation of the scientific community. Provide more sound discussion on the mechanisms.

The interpretation and discussion of the results obtained was significantly improved as requested by Reviewer 1, 3 and 4. More sound discussions of the mechanisms underlying O₂ mass transfer during the gassing-out and sulphite oxidation experiment, and CH₄ mass transfer during the biotic experiments, was also provided. More specifically:

Current Page 9

“The type of O_2 sensor (polarographic vs optical) used in the gassing-out methodology did not significantly ($p \leq 0.05$) influence the measurement of k_{laO_2} (average error < 2.75 %), which confirmed that the significant ($p \leq 0.05$) differences”

Current page 10

“At this point it should be highlighted that moderate salinity media, despite slightly reducing the solubility of gases, typically result in enhanced gas-liquid k_{la} induced by the higher gas-liquid interfacial areas (Rodero et al., 2018).”

“Figure 2, which depicts the O_2 profiles in the liquid phase during the gassing-out and OTR experiments, clearly shows the higher mass transfer gradients (estimated locally from the slope of the decrease in dissolved O_2 concentration at the gas-liquid interphase) despite the overall difference in O_2 concentration between the interphase and the bulk liquid ($200 \text{ mg L}^{-1} - 0 \text{ mg L}^{-1}$) remained similar. These higher gradients mediated by the rapid oxidation of O_2 in the presence of SO_3^{2-} and Co^{2+} entailed a superior O_2 mass transfer, which mathematically resulted into higher k_{laO_2} (obtained from equations 1 and 3). Finally, it should be highlighted that the k_{laCH_4} here recorded were similar to those reported in vertical tubular loop (70 h^{-1}) and airlift external loop reactors (85 h^{-1}), but higher than those observed in stirred tank, airlift bubble column and U-loop reactors ($12\text{-}26 \text{ h}^{-1}$) (Stone et al., 2017).”

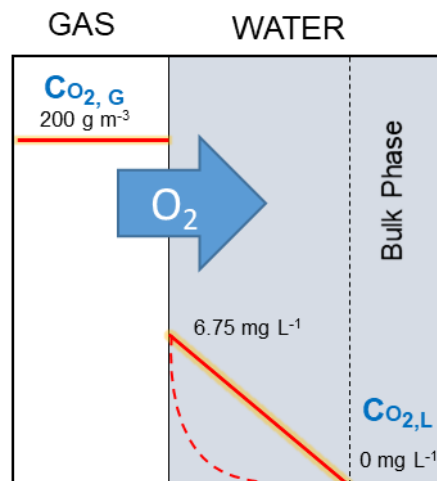


Figure 2. O_2 concentration profiles in the liquid phase during the gassing-out (continuous line) and OTR (dashed line) methods.

Current page 11

“At this point it should be also stressed that high k_{laCH_4} values are required in CH_4 treating bioreactors in order to compensate for the low concentration differences imposed by the high Henry law constant of CH_4 (~ 30 at $25 \text{ }^\circ\text{C}$) (Lopez et al., 2013). Thus, the higher the k_{laCH_4} values, the lower the bioreactor volumes and consequently, the lower the investment costs.”

“This implies that the removal of CH_4 recorded at 400, 600 and 800 rpm can be attributed to the biocatalytic action of the microbial consortium present in the fermenter, which converts CH_4 to CO_2 , H_2O and new biomass (Lopez et al., 2014).”

$$CH_4 - EC = kLa_{CH_4} \times \left(\frac{C_{CH_4,in}}{H_{CH_4}} - C_{L,CH_4} \right) \quad (Eq. 7)$$

“Where H_{CH_4} represents the dimensionless Henry law constant of CH_4 (28.7 at 30 °C) and C_{L,CH_4} the aqueous concentration of CH_4 in the bulk phase, which is negligible under the mass transfer limiting scenario here encountered. In fact, no significant differences ($p \leq 0.05$) between the empirical and estimated CH_4 -ECs were found at 400 and 800 rpm.”

“In this context, Cu positively regulates the activity of the enzymes pMMO and sMMO (particulate and soluble CH_4 monooxygenase, respectively) and controls the expression of their genes. Most methanotrophs grow optimally at Cu concentrations lower than $270 \mu g L^{-1}$, though previous enzymatic assays have demonstrated that sMMO in type II methanotrophs is properly synthesized at low Cu concentrations (below $50 \mu g L^{-1}$) (Graham et al., 1993; Bender and Conrad, 1995; Semrau et al., 2010).”

“Thus, the rapid CH_4 biodegradation in the vicinity of the gas-liquid interphase mediated by methanotrophic bacteria likely supported an enhanced CH_4 mass transfer (induced by the higher local gradient close to the gas-liquid interphase)”

“3.3 Practical applications and future research perspectives

A renewed attention on suspended growth bioreactors for CH_4 abatement (i.e. stirred tank, airlift and bubble columns) has emerged in the past 5 years due to the popularization of biorefineries based on the feasible bioconversion of CH_4 into added-value bioproducts such as biopolymers, ectoine, exopolysaccharides, protein, etc. (Cantera et al., 2017; Karthikeyan et al., 2015; Khmelenina et al., 2015; Pieja et al., 2012). These novel biorefineries, which will be based both on dilute CH_4 emissions from abandoned landfills or coal mines and biogas from solid waste management, will create value out of GHG mitigation and envisage a new bioeconomy based on CH_4 . In this context, suspended growth systems represent nowadays the most popular bioreactor platform capable of supporting a cost-competitive methanotrophic biomass harvesting and bioproduct downstream, although recent attempts to produce added-value products from residual CH_4 in immobilized bioreactors have been reported (Cantera et al., 2016; Karthikeyan et al., 2017). Thus, the characterization of CH_4 mass transport from residual emissions or biogas streams into the microbial community is central to dimension the bioreactor and downstream processes. However, the absence of reliable commercial dissolved CH_4 sensors has hindered the development of accurate methodologies for the determination of kLa_{CH_4} , whose characterization has been conducted based on either classical kLa_{O_2} determination methodologies without a further validation for CH_4 or complex experimental methodologies difficult to implement at large scale (Yu et al., 2010; Rocha-Rios et al., 2011). Therefore, the results herein obtained validated the SO_3^{2-} oxidation methodologies as the most accurate method to estimate kLa_{CH_4} , which is expected to support in the design of cost-effective CH_4 biorefineries. For instance, figure S3 depicts a simple model simulation of the CH_4 outlet concentration as a function of the CH_4 inlet concentration and the kLa_{CH_4} estimated at the

three stirring rates tested. This methodology entails the purchase of SO_3^{2-} and Co^{2+} , and will result in SO_4^{2-} as a final product in the cultivation medium after the experiment. Further research should focus on validating this methodology in high-performance off gas treatment bioreactors such as Taylor flow or two-phase partitioning bioreactors, which are foreseen as the only bioreactor configuration capable of supporting an effective CH_4 bioconversion at a reduced energy consumption (Cantera et al., 2017). In addition, this methodology could be also applied to the design of algal-bacterial photobioreactors devoted to biogas upgrading (CO_2 and H_2S removal), where both chemically and biologically enhanced mass transport has been identified (Rodero et al., 2018, Yang et al., 2016a,b, Zhu, 2015, Zhu et al., 2016)."

Detailed Comments

Page 2

Line 29: what do authors means by "accurately predicted the CH_4 elimination capacity"?

The original sentence was modified as recommended by Reviewer 1 in order to quantify the error between the empirical and estimated methane elimination capacities in the fermenter (current page 2).

"The kla_{CH_4} estimated from the OTR methodology accurately predicted the CH_4 elimination capacity (EC) under CH_4 mass transfer limiting conditions regardless of the stirring rate (~ 9 % of average error between empirical and estimated ECs)"

Page 5

Line 29: "The kla_{O_2} was recorded at an aeration rate of 0.42 L min^{-1} under three stirring rates (400, 600 and 800 rpm) ? Why chose three stirring rates? Should carefully explain in detail.

These aeration rates were selected as potential operational conditions implemented in large-scale fermenters devoted to the bioconversion of CH_4 . An explanatory remark was included in page 5 to avoid further misunderstandings:

"These aeration rates were selected as potential operational conditions implemented in large-scale fermenters devoted to the bioconversion of CH_4 ."

Page 6

Line 1: "O₂ concentration at the beginning of the test (DO = 0 mg L⁻¹ at t = 0 s)", DO = 0 ?

Please, kindly note that the cultivation broth in the fermenter was always initially degassed with Argon prior to each gassing-out experiment (former page 5 lines 32-35), which supported the negligible dissolved oxygen concentrations reported at $t = 0 \text{ s}$ ($DO = 0 \text{ mg L}^{-1}$). An explanatory remark was included in this revised manuscript in order to avoid further misunderstandings (current page 6):

“....., $C_{O_2,L}$ the dissolved O_2 concentration at time t , and $(C_{O_2,L})_0$ the dissolved O_2 concentration at the beginning of the test ($DO = 0 \text{ mg L}^{-1}$ at $t = 0 \text{ s}$, since the cultivation broth was initially degassed with argon).”

Page 7

Line 43: Analytical methods is too simple and should be explained and supplemented. Do not refer to literatures in this section. Provide brief explanation on how the authors did the work in real situation in the lab.

The authors agree with Reviewer 2 on the fact that this section can be systematically improved by providing the description of the dissolved oxygen sensors, and the detailed methodologies for the determination of biomass and sulphite concentrations (current page 8-9):

“A polarographic sensor (Applikon, The Netherlands) and a luminescent/optical dissolved O_2 sensor (Intellical™ LDO101, Hach, USA) were used for the measurement of the dissolved O_2 concentration in the cultivation broth during the gassing-out experiments.”

The determination of SO_3^{2-} concentration involved the careful sampling of 5 mL of cultivation broth to minimize their contact with ambient air. The sample was then transferred to a 50-ml Erlenmeyer containing 1 ml of sulfuric acid and 0.1 g of NH_2SO_3H crystals. After 1 ml of a starch solution (1%) was added as indicator, the sample was titrated with standard $KI-KIO_3$ titrant until a faint permanent blue color developed. Similarly, a sample of distilled water was processed as the blank control. The SO_3^{2-} concentration was then calculated according to the following equation 6:

$$[SO_3^{2-}][M] = \frac{[\text{Titrant Volume for Sample (ml)} - \text{Titrant volume for Blank (ml)}]}{\text{Volumen Sample (ml)}} \quad (\text{Eq. 6})$$

Biomass concentration was determined by filtering 10 ml of cultivation broth in dry pre-weighed 0.2 μm acetate cellulose filters, which were then allowed to dry overnight and weighted again.”

Page 10

Line 43: How unique and useful is this "first time" of authors compared to other study? It should be summarized and explained in the text with proper comparisons. A new Table has to be prepared.

To the best of our knowledge, this was the first work validating a simple methodology to accurately describe CH_4 mass transport in continuous bioreactors based on OTR estimations. However, since other methodologies have been used to characterize the gas-liquid methane mass transport in bioreactors (see Table S1 in the *supplementary material* section), the term “for the first time” was deleted in this revised version of the manuscript. An explanatory remark in the introduction section (page 4) and a new table (Table S1 in *supplementary materials*) were included in this revised version of the manuscript in order to provide a fairer and more detailed overview of the previous studies on the topic.

“Table S1 presents the main studies devoted to the characterization of the volumetric CH_4 mass transfer coefficients in bioreactors”

1. Compilation of previous studies on the determination of kl_{CH_4}

Table S1. Previous studies on the characterization of the gas-liquid CH_4 mass transport in bioreactors			
Bioreactor Configuration	Methodology	kl_{CH_4} (h^{-1})	Reference
Immobilized soil bioreactor & Fermenter	kl_{O_2} based on the dynamic method combined with gas phase CH_4 analysis Model validation	16.3 \pm 0.67	Yu et al. (2010)
Airlift bioreactor with internal recirculation	kl_{O_2} based on OTR No validation of CH_4 -EC was achieved	28.6 -200*	Rocha-Rios et al. (2011)
Two-phase partitioning fermenter	kl_{O_2} based on static method No validation of CH_4 -EC was achieved	6.2-35.4*	Rocha-Rios et al. (2010)
250 mL bottle	kl_{CH_4} based on CH_4 absorption experiments No methodology validation	22.7-41.9	Lee et al. (2015)

*- Estimated from kl_{O_2} using equation 4.

Lee, S.Y., Mo, K.S., Choi, J.H., Hur, N.H., Kim, Y.K., Oh, B.K., Lee, J., 2015. Enhancement of CH_4 -water mass transfer using methyl-modified mesoporous silica nanoparticles. *Korean Journal of Chemical Engineering*. 32(9), 1744-1748.

Rocha-Rios, J., Muñoz, R., Revah, S., 2010. Effect of silicone oil fraction and stirring rate on methane degradation in a stirred tank reactor. *Journal of Chemical Technology and Biotechnology*. 85, 314-319.

Rocha-Rios, J.M., Quijano, G., Thalasso, F., Revah, S., Muñoz, R., 2011. Methane biodegradation in a two-phase partitioning airlift reactor with gas recirculation. *Journal of Chemical Technology and Biotechnology*. 86(3), 353-360.

Yu, Y., Ramsay, J.A., Ramsay, B.A., 2006. On-Line estimation of dissolved methane concentration during methanotrophic fermentations. *Biotechnology and Bioengineering*. 95, 788-793.

REVIEWER 2

This study contributes to the scientific literature of the methanotrophic fermentation; also the findings of this study can be useful for engineers and scientist that use mathematical modelling in order to design a methanotrophic bioreactor or to estimate the performance of a methanotrophic process. However, the authors need to elaborate more on several key parameters that have investigated.

The authors acknowledge the positive feedback from Reviewer 2

1) The title can be more specific and can describe in a higher accuracy the experiments and the comparison.

The title of the paper was modified in accordance to Reviewer 2 request to “*A systematic comparison of two empirical gas-liquid mass transfer determination methodologies to characterize methane biodegradation in stirred tank bioreactors*” (current page 1)

2) At the abstract the aim of this study need to be clearly stated.

The aim of the study was clearly stated in the first sentence of the revised abstract (current page 2)

“This study aimed at systematically comparing the potential of two empirical methods for the estimation of the volumetric mass transfer coefficient of CH₄ (k_{lCH_4}), namely gassing-out vs oxygen transfer rate (OTR), to describe CH₄ biodegradation in a fermenter operated with a methanotrophic consortium at 400, 600 and 800 rpm.”

3) Page 2 Line 12 : Which biorefineries? Please be specific – which industries? [UTF-8?]"“has raised increasing attention in biorefineries”

Please note that the abstract in this revised manuscript was reduced from 230 words to 150 words by request of the guest editor. Thus, the introductory sentences of the original abstract (including former line 12 in page 2) were deleted and this comment does not longer apply in the revised abstract.

4) At the introduction more information are needed regarding the " k_{lCH_4} , namely gassing-out vs Oxygen Transfer Rate (OTR)" Why is it important ?

Additional information was provided in the introduction section of this revised manuscript regarding the relevance of k_{lCH_4} in accordance to Reviewer 2 request (current page 4)

“Unfortunately, the validation of empirical methodologies for the determination of the volumetric CH₄ mass transfer coefficient (k_{lCH_4}) has been poorly addressed in literature (García-Ochoa et al., 2009; Rocha-Ríos et al., 2010). Indeed, the lack of cheap and reliable dissolved CH₄ sensors along with the low solubility and limited industrial interest of CH₄ to date have restricted the development of accurate methodologies for the characterization of k_{lCH_4} . Table S1 presents the main studies devoted to the characterization of the volumetric CH₄ mass transfer coefficients in bioreactors.”

“This study aims at comparing and validating, for the first time, two experimental methodologies for the estimation of k_{lCH_4} (gassing-out vs oxygen transfer rate methods) in

order to accurately describe CH₄ abatement in a stirred tank reactor operated with a methanotrophic consortium at multiple stirring rates. This information is expected to provide the basis for a correct bioreactor design for CH₄ bioconversion into added-value products.”

5) What are the literature estimated values for question 4? How was it calculated?

A new table (Table S1) compiling the most recent attempts to characterize the gas-liquid mass transfer in CH₄ treating bioreactors, along with the methodologies used to estimate kl_{CH_4} , was added in the *supplementary materials* section accordingly.

1. Compilation of previous studies on the determination of kl_{CH_4}

Table S1. Previous studies on the characterization of the gas-liquid CH₄ mass transport in bioreactors			
Bioreactor Configuration	Methodology	kl_{CH_4} (h⁻¹)	Reference
Immobilized soil bioreactor & Fermenter	kl_{O_2} based on the dynamic method combined with gas phase CH ₄ analysis Model validation	16.3 ±0.67	Yu et al. (2010)
Airlift bioreactor with internal recirculation	kl_{O_2} based on OTR No validation of CH ₄ -EC was achieved	28.6 -200*	Rocha-Rios et al. (2011)
Two-phase partitioning fermenter	kl_{O_2} based on static method No validation of CH ₄ -EC was achieved	6.2-35.4*	Rocha-Rios et al. (2010)
250 mL bottle	kl_{CH_4} based on CH ₄ absorption experiments No methodology validation	22.7-41.9	Lee et al. (2015)

*- Estimated from kl_{O_2} using equation 4.

Lee, S.Y., Mo, K.S., Choi, J.H., Hur, N.H., Kim, Y.K., Oh, B.K., Lee, J., 2015. Enhancement of CH₄-water mass transfer using methyl-modified mesoporous silica nanoparticles. *Korean Journal of Chemical Engineering*. 32(9), 1744-1748.

Rocha-Rios, J., Muñoz, R., Revah, S., 2010. Effect of silicone oil fraction and stirring rate on methane degradation in a stirred tank reactor. *Journal of Chemical Technology and Biotechnology*. 85, 314-319.

Rocha-Rios, J.M., Quijano, G., Thalasso, F., Revah, S., Muñoz, R., 2011. Methane biodegradation in a two-phase partitioning airlift reactor with gas recirculation. *Journal of Chemical Technology and Biotechnology*. 86(3), 353-360.

Yu, Y., Ramsay, J.A., Ramsay, B.A., 2006. On-Line estimation of dissolved methane concentration during methanotrophic fermentations. Biotechnology and Bioengineering, 95, 788-793.

In addition, a brief statement was included in the revised *Results and Discussion* section in order to compare the k_{lCH_4} herein obtained with those previously reported in literature (current page 10-11)

“Finally, it should be highlighted that the k_{lCH_4} here recorded were similar to those reported in vertical tubular loop (70 h^{-1}) and airlift external loop reactors (85 h^{-1}), but higher than those observed in stirred tank, airlift bubble column and U-loop reactors ($12\text{-}26\text{ h}^{-1}$) (Stone et al., 2017).”

6) What is the impact of $K_{la}\text{ CH}_4$ in bioreactors? Several examples from the literature need to be stated because CH_4 is poorly water soluble.

An explanatory remark describing the impact of k_{lCH_4} on bioreactor design was included in the *Results and Discussion* section of this revised manuscript in accordance to Reviewer 2 request (current page 11).

“At this point it should be also stressed that high k_{lCH_4} values are required in CH_4 treating bioreactors in order to compensate for the low concentration gradients imposed by the high Henry law constant of CH_4 (~ 30 at $25\text{ }^\circ\text{C}$) (Lopez et al., 2013). Thus, the higher the k_{lCH_4} values, the lower the bioreactor volumes and consequently, the lower the investment costs.”

7) Page 4 line 12 a reference is needed. [UTF-8]"â€|using a biorefinery approach"

A recent reference was included in former page 4 line 12 in accordance to Reviewer 2 request (current page 4).

“Indeed, CH_4 can be bioconverted into biopolymers, ectoine, protein, exopolysaccharides, etc., using a biorefinery approach (Cantera et al., 2017).”

Cantera, S., Frutos, O., López, J.C., Lebrero, R., Muñoz, R., 2017. Technologies for the bio-conversion of GHGs into high added value products: current state and future prospects. In: R. Álvarez (Ed.), Carbon Footprint and the Industrial Life Cycle, Springer, Cham, pp. 359-388.

Material and methods:

8) Apart from *Sphingobacterium* sp. CZ-UAM what are the other strains of the methanotrophic consortium. If the authors do not know the microbial composition they can report the inoculum source for of the methanotrophic consortium.

Unfortunately, a detailed microbial characterization of the methanotrophic consortium was not carried out. The source of the inoculum was included in this revised version of the manuscript in accordance to Reviewer 2 request (current page 5).

“The consortium was enriched from the UAM-Iztapalapa wastewater treatment plant (México City).”

9) The iodometric back-titration need to be briefly described.

A detailed description of the iodometric back-titration methodology was included in the revised *analytical methods* section in accordance to Reviewer 2 request (current page 8)

The determination of SO_3^{2-} concentration involved the careful sampling of 5 mL of cultivation broth to minimize their contact with ambient air. The sample was then transferred to a 50-ml Erlenmeyer containing 1 ml of sulfuric acid and 0.1 g of NH_2SO_3H crystals. After 1 ml of a starch solution (1%) was added as indicator, the sample was titrated with standard $KI-KIO_3$ titrant until a faint permanent blue color developed. Similarly, a sample of distilled water was processed as the blank control. The SO_3^{2-} concentration was then calculated according to the following equation 6:

$$[SO_3^{2-}][M] = \frac{[\text{Titration Volume for Sample (ml)} - \text{Titration volume for Blank (ml)}]}{\text{Volumen Sample (ml)}} \quad (\text{Eq. 6})$$

10) More information need to be provided regarding the $K_{La}CH_4 = K_{La}O_2/1.16$

This correlation was derived and empirically validated by Yu et al. (2010). This information was included in the revised version of the manuscript accordingly (current page 7).

“Finally, the volumetric mass transfer coefficient of CH_4 was estimated from k_{LaO_2} according to Yu et al. (2010) using the correlation empirically validated for stirred tank reactors based on dissolved CH_4 measurements and a dynamic methodology to characterize k_{LaO_2} (equation 4):”

Results:

11) What are the main conclusions from the first test? The statement page 9 line 4 [UTF-8?]” remains unclear in literature, despite its significant [UTF-8?] impact” do not help the reader.

The statement in former page 8 line 55–page 9 line 7 was deleted in this revised manuscript in order to avoid further misunderstandings.

12) Why MSM supplementation with Cu at 50 $\mu\text{g L}^{-1}$ was required to overcome process limitation? The authors need to explain the affinity of several methanogens for Cu.

A brief discussion to describe the role of Cu on CH_4 biodegradation and the typical levels of Cu to support the synthesis of the enzymes pMMO and sMMO was included in this revised version of the manuscript in accordance to Reviewer 2 request (current page 12).

“In this context, Cu positively regulates the activity of the enzymes pMMO and sMMO (particulate and soluble CH_4 monooxygenase, respectively) and controls the expression of their genes. Most methanotrophs grow optimally at Cu concentrations lower than 270 $\mu\text{g L}^{-1}$, though previous enzymatic assays have demonstrated that sMMO in type II methanotrophs is properly synthesized at low Cu concentrations (below 50 $\mu\text{g L}^{-1}$) (Graham et al., 1993; Bender and Conrad, 1995; Semrau et al., 2010).”

Bender, M., Conrad, R., 1995. Effect of CH₄ concentrations and soil conditions on the induction of CH₄ oxidation activity. *Soil Biology and Biochemistry*. 27(12), 1517-1527.

Graham, D.W., Chaudhary, J.A., Hanson, R.S., Arnold, R.G., 1993. Factors affecting competition between type I and type II methanotrophs in two-organism, continuous-flow reactors. *Microbial ecology*. 25(1), 1-17.

Semrau, J.D., DiSpirito, A.A., Yoon, S., 2010. Methanotrophs and copper. *FEMS microbiology reviews*. 34(4), 496-531.

13) The authors can report more studies apart from [21]. Page 10 line 21. Also a table that can contain K_{laCH_4} values from other studies, measuring methods, type of bioreactor can be included.

A table containing k_{laCH_4} values from recent studies, along with the methodologies used and the type of bioreactor, was included in the *supplementary material* section as requested by Reviewer 2.

1. Compilation of previous studies on the determination of k_{laCH_4}

<i>Table S1. Previous studies on the characterization of the gas-liquid CH₄ mass transport in bioreactors</i>			
<i>Bioreactor Configuration</i>	<i>Methodology</i>	<i>k_{laCH_4} (h⁻¹)</i>	<i>Reference</i>
<i>Immobilized soil bioreactor & Fermenter</i>	<i>klao₂ based on the dynamic method combined with gas phase CH₄ analysis Model validation</i>	<i>16.3 ±0.67</i>	<i>Yu et al. (2010)</i>
<i>Airlift bioreactor with internal recirculation</i>	<i>klao₂ based on OTR No validation of CH₄-EC was achieved</i>	<i>28.6 -200*</i>	<i>Rocha-Rios et al. (2011)</i>
<i>Two-phase partitioning fermenter</i>	<i>klao₂ based on static method No validation of CH₄-EC was achieved</i>	<i>6.2-35.4*</i>	<i>Rocha-Rios et al. (2010)</i>
<i>250 mL bottle</i>	<i>klach₄ based on CH₄ absorption experiments No methodology validation</i>	<i>22.7-41.9</i>	<i>Lee et al. (2015)</i>

*- Estimated from klao₂ using equation 4.

Lee, S.Y., Mo, K.S., Choi, J.H., Hur, N.H., Kim, Y.K., Oh, B.K., Lee, J., 2015. Enhancement of CH₄-water mass transfer using methyl-modified mesoporous silica nanoparticles. *Korean Journal of Chemical Engineering*. 32(9), 1744-1748.

Rocha-Rios, J., Muñoz, R., Revah, S., 2010. Effect of silicone oil fraction and stirring rate on methane degradation in a stirred tank reactor. Journal of Chemical Technology and Biotechnology. 85, 314-319.

Rocha-Rios, J.M., Quijano, G., Thalasso, F., Revah, S., Muñoz, R., 2011. Methane biodegradation in a two-phase partitioning airlift reactor with gas recirculation. Journal of Chemical Technology and Biotechnology. 86(3), 353-360.

Yu, Y., Ramsay, J.A., Ramsay, B.A., 2006. On-Line estimation of dissolved methane concentration during methanotrophic fermentations. Biotechnology and Bioengineering. 95, 788-793.

In addition, a brief statement was included in the revised *Results and Discussion* section in order to compare the kl_{aCH_4} herein obtained with those previously reported in literature (current page 10-11)

“Finally, it should be highlighted that the kl_{aCH_4} here recorded were similar to those reported in vertical tubular loop (70 h^{-1}) and airlift external loop reactors (85 h^{-1}), but higher than those observed in stirred tank, airlift bubble column and U-loop reactors ($12\text{-}26\text{ h}^{-1}$) (Stone et al., 2017).”

14) At the conclusion the estimated values or the absolute error between the calculated and the predicted values need to be recorded.

The CH_4 elimination capacities experimentally recorded in the continuous bioreactor and those estimated from the OTR methodology are now described in the conclusion section of this revised manuscript in accordance to Reviewer 2 request (current page 15)

“Thus, the empirical CH_4 -ECs of 37.8 ± 5.8 , 42.5 ± 5.4 and $62.3\pm 5.2\text{ g CH}_4\text{ m}^{-3}\text{ h}^{-1}$ recorded at 400, 600 and 800 rpm, respectively, were comparable to the CH_4 -ECs of 35.6 ± 2.2 , 50.1 ± 2.3 and $59.6\pm 3.4\text{ g CH}_4\text{ m}^{-3}\text{ h}^{-1}$ predicted by the OTR methodology.”

15) How can the results be used in practice, in which circumstance, which industry and how. Provide a good guideline to the reader.

The discussion section was modified in accordance to Reviewer 2 suggestion (current page 13 and 14). In addition, a practical application of the kl_{aCH_4} determined is shown in the *supplementary materials* section.

“3.3 Practical applications and future research perspectives

A renewed attention on suspended growth bioreactors for CH_4 abatement (i.e. stirred tank, airlift and bubble columns) has emerged in the past 5 years due to the popularization of biorefineries based on the feasible bioconversion of CH_4 into added-value bioproducts such as biopolymers, ectoine, exopolysaccharides, protein, etc. (Cantera et al., 2017; Karthikeyan et al., 2015; Khmelenina et al., 2015; Pieja et al., 2012). These novel biorefineries, which will be based both on dilute CH_4 emissions from abandoned landfills or coal mines and biogas from solid waste management, will create value out of GHG mitigation and envisage a new bioeconomy based on CH_4 . In this context, suspended growth systems represent nowadays the most popular bioreactor platform capable of supporting a cost-competitive methanotrophic biomass harvesting and bioproduct downstream, although

recent attempts to produce added-value products from residual CH_4 in immobilized bioreactors have been reported (Cantera et al., 2016; Karthikeyan et al., 2017). Thus, the characterization of CH_4 mass transport from residual emissions or biogas streams into the microbial community is central to dimension the bioreactor and downstream processes. However, the absence of reliable commercial dissolved CH_4 sensors has hindered the development of accurate methodologies for the determination of $k_{l\text{CH}_4}$, whose characterization has been conducted based on either classical $k_{l\text{O}_2}$ determination methodologies without a further validation for CH_4 or complex experimental methodologies difficult to implement at large scale (Yu et al., 2010; Rocha-Rios et al., 2011). Therefore, the results herein obtained validated the SO_3^{2-} oxidation methodologies as the most accurate method to estimate $k_{l\text{CH}_4}$, which is expected to support in the design of cost-effective CH_4 biorefineries. For instance, figure S3 depicts a simple model simulation of the CH_4 outlet concentration as a function of the CH_4 inlet concentration and the $k_{l\text{CH}_4}$ estimated at the three stirring rates tested. This methodology entails the purchase of SO_3^{2-} and Co^{2+} , and will result in SO_4^{2-} as a final product in the cultivation medium after the experiment. Further research should focus on validating this methodology in high-performance off gas treatment bioreactors such as Taylor flow or two-phase partitioning bioreactors, which are foreseen as the only bioreactor configuration capable of supporting an effective CH_4 bioconversion at a reduced energy consumption (Cantera et al., 2017). In addition, this methodology could be also applied to the design of algal-bacterial photobioreactors devoted to biogas upgrading (CO_2 and H_2S removal), where both chemically and biologically enhanced mass transport has been identified (Rodero et al., 2018, Yang et al., 2016a,b, Zhu, 2015, Zhu et al., 2016).”

4. Application of the $k_{l\text{CH}_4}$ for process design

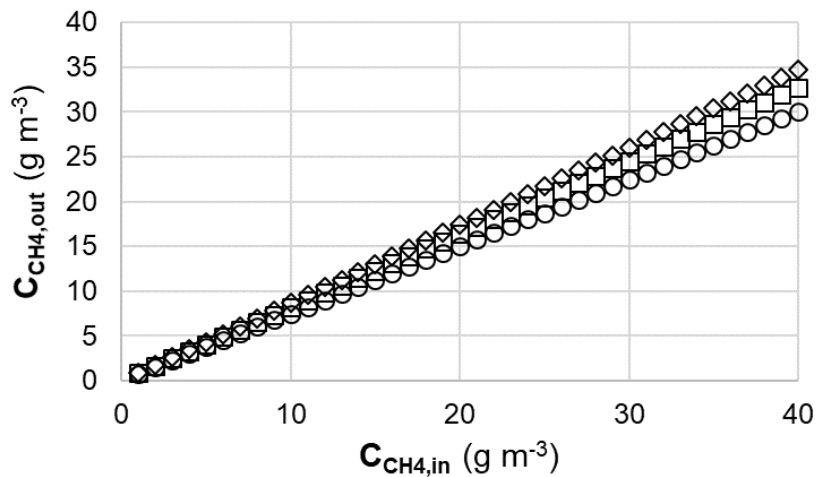


Figure S3. Model simulation of the CH_4 outlet concentration as a function of the CH_4 inlet concentration and the $k_{l\text{CH}_4}$ estimated at 400 (\diamond), 600 (\square) and 800 rpm (\circ).

The CH_4 outlet concentration, $C_{\text{CH}_4,\text{out}}$, was derived from the definition of the CH_4 elimination capacity as follows (equation S1 and S2)

$$CH_4 - EC = \frac{Q}{V} \times (C_{CH_4,in} - C_{CH_4,out}) = kla_{CH_4} \times \left(\frac{C_{CH_4,in}}{H_{CH_4}} - C_{L,CH_4} \right) \quad (Eq. S1)$$

$$C_{CH_4,out} = \frac{\left(\frac{Q}{V} \times C_{CH_4,in} \right)}{\left(\frac{kla_{CH_4}}{H_{CH_4}} + \frac{Q}{V} \right)} \quad (Eq. S2)$$

REVIEWER 3

This manuscript presents data on the CH₄ mass transfer characterization in stirred tank bioreactors devoted to CH₄ bioconversion. While the manuscript provides useful practical information, it needs to be significantly improved so that the ideas and data presented in the manuscript and data interpretation and discussion are clear. I suggested MAJOR REVISION.

The authors acknowledge the positive feedback from Reviewer 3

1. The novelty/originality shall be further justified by highlighting that the manuscript contains sufficient contributions to the new body of knowledge.

The novelty of this work was highlighted in the introduction section of this revised version of the manuscript as requested by Reviewer 3 and the guest editor (current page 4)

“Unfortunately, the validation of empirical methodologies for the determination of the volumetric CH₄ mass transfer coefficient (k_{l,CH_4}) has been poorly addressed in literature (García-Ochoa et al., 2009; Rocha-Ríos et al., 2010). Indeed, the lack of cheap and reliable dissolved CH₄ sensors along with the low solubility and limited industrial interest of CH₄ to date have restricted the development of accurate methodologies for the characterization of k_{l,CH_4} . Table S1 presents the main studies devoted to the characterization of the volumetric CH₄ mass transfer coefficients in bioreactors.

This study aims at comparing and validating, for the first time, two experimental methodologies for the estimation of k_{l,CH_4} (gassing-out vs oxygen transfer rate methods) in order to accurately describe CH₄ abatement in a stirred tank reactor operated with a methanotrophic consortium at multiple stirring rates. This information is expected to provide the basis for a correct bioreactor design for CH₄ bioconversion into added-value products.”

In other words, literature survey is not sufficient. Therefore, I suggest that you should read the following references:

- a) Photosynthetic CO₂ uptake by microalgae for biogas upgrading and simultaneously biogas slurry decontamination by using of microalgae photobioreactor under various light wavelengths, light intensities, and photoperiods. Applied Energy
- b) Review Microalgal cultivation with biogas slurry for biofuel production. Bioresource Technology
- c) The effects of various LED lighting strategies on simultaneous biogas upgrading and biogas slurry nutrient reduction by using of microalgae *Chlorella* sp. Energy

d) Biorefinery as a promising approach to promote microalgae industry: an innovative framework. Renewable and Sustainable Energy Reviews

The literature survey was improved in this revised version of the manuscript as recommended by Reviewer 3. The recommended references were included in a new 1.5 page section entitled “**3.3 Practical applications and future research perspectives**” (current page 14)

“In addition, this methodology could be also applied to the design of algal-bacterial photobioreactors devoted to biogas upgrading (CO₂ and H₂S removal), where both chemically and biologically enhanced mass transport has been identified (Rodero et al., 2018, Yang et al., 2016a,b, Zhu, 2015, Zhu et al., 2016).”

2. More discussion and literature reference should be provided in the results and discussions. A lot of discussion is missing.

The interpretation and discussion of the results obtained was significantly improved as requested by Reviewer 3, 2 and 1. More sound discussions of the mechanisms underlying O₂ mass transfer during the gassing-out and sulphite oxidation experiment, and CH₄ mass transfer during the biotic experiment, was also provided. More specifically:

Current Page 9

“The type of O₂ sensor (polarographic vs optical) used in the gassing-out methodology did not significantly ($p \leq 0.05$) influence the measurement of k_{lao_2} (average error < 2.75 %), which confirmed that the significant ($p \leq 0.05$) differences”

Current page 10

“At this point it should be highlighted that moderate salinity media, despite slightly reducing the solubility of gases, typically result in enhanced gas-liquid k_{la} induced by the higher gas-liquid interfacial areas (Rodero et al., 2018).”

“Figure 2, which depicts the O₂ profiles in the liquid phase during the gassing-out and OTR experiments, clearly shows the higher mass transfer gradients (estimated locally from the slope of the decrease in dissolved O₂ concentration at the gas-liquid interphase) despite the overall difference in O₂ concentration between the interphase and the bulk liquid (200 mg L⁻¹ – 0 mg L⁻¹) remained similar. These higher gradients mediated by the rapid oxidation of O₂ in the presence of SO₃²⁻ and Co²⁺ entailed a superior O₂ mass transfer, which mathematically resulted into higher k_{lao_2} (obtained from equations 1 and 3). Finally, it should be highlighted that the k_{lACH_4} here recorded were similar to those reported in vertical tubular loop (70 h⁻¹) and airlift external loop reactors (85 h⁻¹), but higher than those observed in stirred tank, airlift bubble column and U-loop reactors (12-26 h⁻¹) (Stone et al., 2017).”

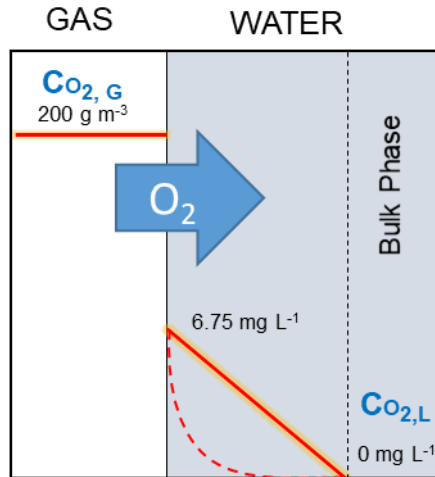


Figure 2. *O₂ concentration profiles in the liquid phase during the gasing-out (continuous line) and OTR (dashed line) methods.*

Current page 11

“At this point it should be also stressed that high k_{lCH_4} values are required in CH_4 treating bioreactors in order to compensate for the low concentration differences imposed by the high Henry law constant of CH_4 (~ 30 at $25\text{ }^\circ\text{C}$) (Lopez et al., 2013). Thus, the higher the k_{lCH_4} values, the lower the bioreactor volumes and consequently, the lower the investment costs.”

“This implies that the removal of CH_4 recorded at 400, 600 and 800 rpm can be attributed to the biocatalytic action of the microbial consortium present in the fermenter, which converts CH_4 to CO_2 , H_2O and new biomass (Lopez et al., 2014).”

Current page 12

$$CH_4 - EC = k_l a_{CH_4} \times \left(\frac{C_{CH_4, in}}{H_{CH_4}} - C_{L, CH_4} \right) \quad (Eq. 7)$$

“Where H_{CH_4} represents the dimensionless Henry law constant of CH_4 (28.7 at $30\text{ }^\circ\text{C}$) and C_{L, CH_4} the aqueous concentration of CH_4 in the bulk phase, which is negligible under the mass transfer limiting scenario here encountered. In fact, no significant differences ($p \leq 0.05$) between the empirical and estimated CH_4 -ECs were found at 400 and 800 rpm.”

“In this context, Cu positively regulates the activity of the enzymes pMMO and sMMO (particulate and soluble CH_4 monooxygenase, respectively) and controls the expression of their genes. Most methanotrophs grow optimally at Cu concentrations lower than $270\text{ }\mu\text{g L}^{-1}$, though previous enzymatic assays have demonstrated that sMMO in type II methanotrophs is properly synthesized at low Cu concentrations (below $50\text{ }\mu\text{g L}^{-1}$) (Graham et al., 1993; Bender and Conrad, 1995; Semrau et al., 2010).”

“Thus, the rapid CH_4 biodegradation in the vicinity of the gas-liquid interphase mediated by methanotrophic bacteria likely supported an enhanced CH_4 mass transfer (induced by the higher local gradient close to the gas-liquid interphase)”

“3.3 Practical applications and future research perspectives

A renewed attention on suspended growth bioreactors for CH₄ abatement (i.e. stirred tank, airlift and bubble columns) has emerged in the past 5 years due to the popularization of biorefineries based on the feasible bioconversion of CH₄ into added-value bioproducts such as biopolymers, ectoine, exopolysaccharides, protein, etc. (Cantera et al., 2017; Karthikeyan et al., 2015; Khmelenina et al., 2015; Pieja et al., 2012). These novel biorefineries, which will be based both on dilute CH₄ emissions from abandoned landfills or coal mines and biogas from solid waste management, will create value out of GHG mitigation and envisage a new bioeconomy based on CH₄. In this context, suspended growth systems represent nowadays the most popular bioreactor platform capable of supporting a cost-competitive methanotrophic biomass harvesting and bioproduct downstream, although recent attempts to produce added-value products from residual CH₄ in immobilized bioreactors have been reported (Cantera et al., 2016; Karthikeyan et al., 2017). Thus, the characterization of CH₄ mass transport from residual emissions or biogas streams into the microbial community is central to dimension the bioreactor and downstream processes. However, the absence of reliable commercial dissolved CH₄ sensors has hindered the development of accurate methodologies for the determination of k_{lCH_4} , whose characterization has been conducted based on either classical k_{lO_2} determination methodologies without a further validation for CH₄ or complex experimental methodologies difficult to implement at large scale (Yu et al., 2010; Rocha-Rios et al., 2011). Therefore, the results herein obtained validated the SO_3^{2-} oxidation methodologies as the most accurate method to estimate k_{lCH_4} , which is expected to support in the design of cost-effective CH₄ biorefineries. For instance, figure S3 depicts a simple model simulation of the CH₄ outlet concentration as a function of the CH₄ inlet concentration and the k_{lCH_4} estimated at the three stirring rates tested. This methodology entails the purchase of SO_3^{2-} and Co^{2+} , and will result in SO_4^{2-} as a final product in the cultivation medium after the experiment. Further research should focus on validating this methodology in high-performance off gas treatment bioreactors such as Taylor flow or two-phase partitioning bioreactors, which are foreseen as the only bioreactor configuration capable of supporting an effective CH₄ bioconversion at a reduced energy consumption (Cantera et al., 2017). In addition, this methodology could be also applied to the design of algal-bacterial photobioreactors devoted to biogas upgrading (CO₂ and H₂S removal), where both chemically and biologically enhanced mass transport has been identified (Rodero et al., 2018, Yang et al., 2016a,b, Zhu, 2015, Zhu et al., 2016).”

REVIEWER 4

The research focuses on the methane mass transfer in conventional fermenters. The authors compared two experimental methodologies for the estimation of the volumetric mass mass transfer coefficient for methane, concluding that the Oxygen Transfer Rate method described the methane mass transfer coefficient. The topic of the manuscript is interesting and related to the field of the journal. The objectives, results and conclusions are clearly presented.

The authors acknowledge the positive feedback from Reviewer 4

However, in my opinion some information in the methodology part are missing and the results of the control tests need to be shown.

The methodology part was significantly improved by adding a more detailed description in the *analytical methods* section and a new statistical treatment section (current page 8-9)

“A polarographic sensor (Applikon, The Netherlands) and a luminescent/optical dissolved O₂ sensor (Intellical™ LDO101, Hach, USA) were used for the measurement of the dissolved O₂ concentration in the cultivation broth during the gassing-out experiments.

The determination of SO₃²⁻ concentration involved the careful sampling of 5 mL of cultivation broth to minimize their contact with ambient air. The sample was then transferred to a 50-ml Erlenmeyer containing 1 ml of sulfuric acid and 0.1 g of NH₂SO₃H crystals. After 1 ml of a starch solution (1%) was added as indicator, the sample was titrated with standard KI-KIO₃ titrant until a faint permanent blue color developed. Similarly, a sample of distilled water was processed as the blank control. The SO₃²⁻ concentration was then calculated according to the following equation 6:

$$[\text{SO}_3^{2-}][M] = \frac{[\text{Titrant Volume for Sample (ml)} - \text{Titrant volume for Blank (ml)}]}{\text{Volumen Sample (ml)}} \quad (\text{Eq. 6})$$

Biomass concentration was determined by filtering 10 ml of cultivation broth in dry pre-weighed 0.2 μm acetate cellulose filters, which were then allowed to dry overnight and weighted again.”

2.5 Statistical Treatment

All experiments were performed in duplicate and each assay was considered as a replicate. The results are given as the average value ± standard deviation from the duplicate assays. In addition, the results were analysed using a one-way ANOVA with significance at p ≤ 0.05. The Excel statistical package (Microsoft Corporation, USA) was used for data treatment.”

The gassing-out control test in the presence of sulphite during the *kla*_{O₂} determination, along with the control abiotic test in the continuous bioreactor, are now provided in the *supplementary materials* section as Figure S1 and Figure S2.

2. Influence of stirring rate on the volumetric mass transfer coefficient of O₂

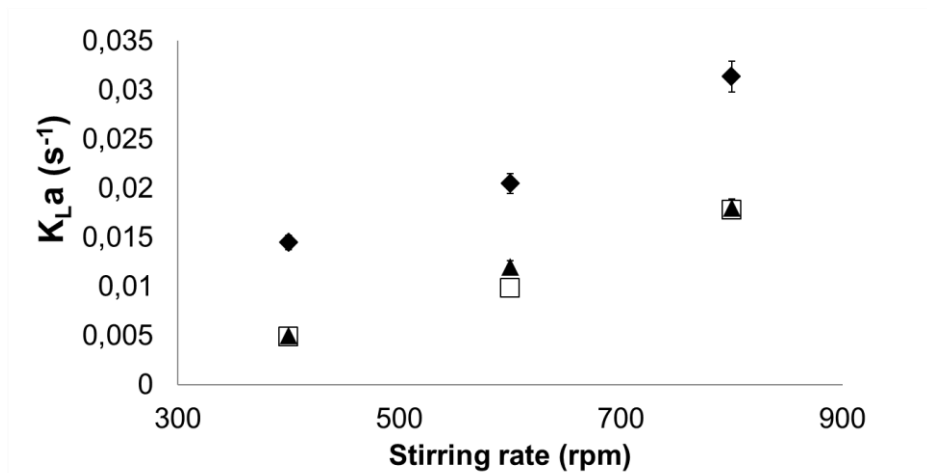


Figure S1. Influence of the stirring rate on the *kla* of O₂ determined by the OTR method (◆), gassing-out (□) and control gassing-out (▲).

3. Abiotic Test

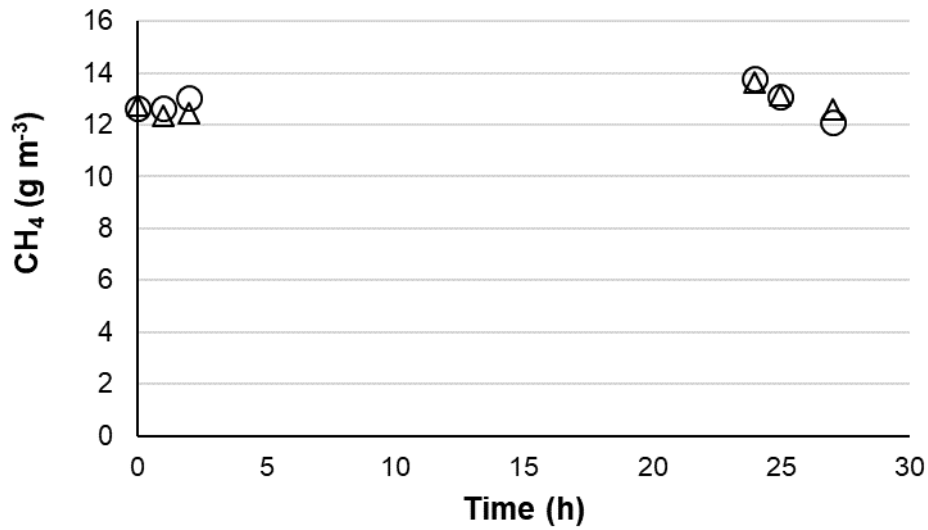
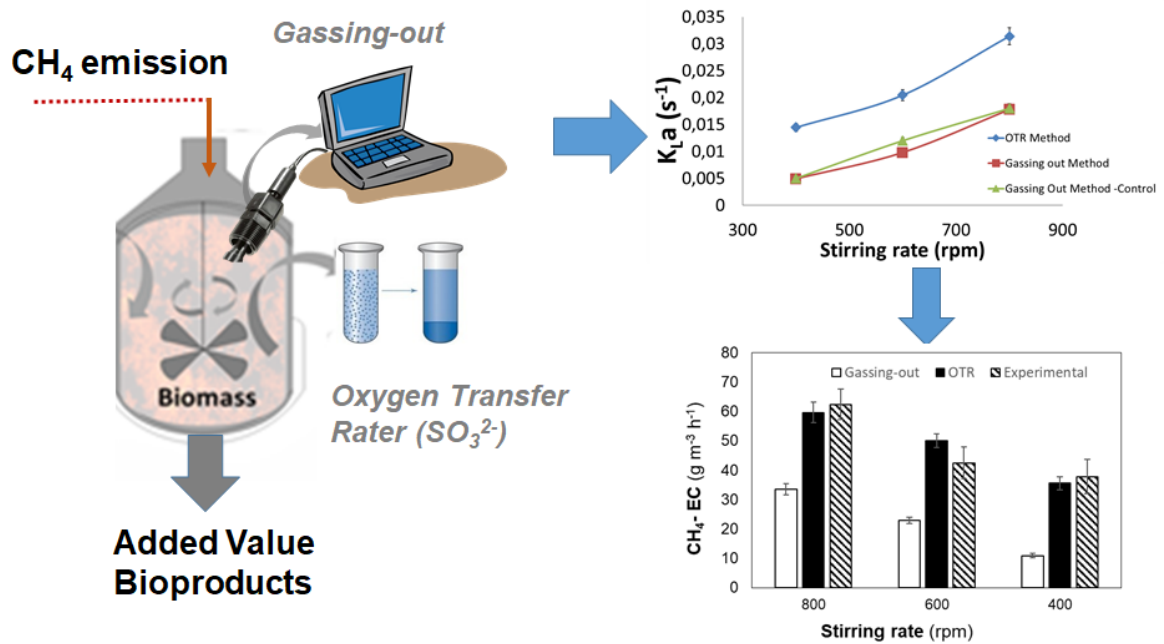


Figure S2. Time course of the inlet (o) and outlet (Δ) CH₄ concentration in the continuous bioreactor during the abiotic test.

1) Graphical abstract: images not clear. Moreover, the organization flow of the graphical abstract is not coherent with the text. From the way it is drawn, the gassing-out method seems the best method, which is not. The graph "k_La vs stirring rate" is not clear, missing units and legend.

The quality of the images, the organization flow of the information and the quality of figures provided in the revised graphical abstract were significantly improved in accordance to the suggestions from Reviewer 4 and the guest editor.



2) Third Highlight need to be rephrased: "Microbial methane uptake accelerated methane transport", where is this explained in the text?

The highlights were rephrased as requested by Reviewer 1 and Reviewer 4 as follows (see *Highlights* section):

k_la_{O₂} determination based on SO₃²⁻ oxidation accurately described CH₄ mass transfer
Microbial CH₄ uptake accelerated CH₄ mass transport similarly to SO₃²⁻ oxidation
The gassing-out methodology underestimated CH₄ mass transport
Maximum CH₄ elimination capacities of 62 ± 5 g CH₄ m⁻³ h⁻¹ were achieved at 800 rpm

Please note that the original Highlight 3 was supported by the discussion provided in former page 9 lines 43-56: “Overall, the results here obtained confirmed that the classical gassing-out methodology for the estimation of *k_la_{CH₄}* clearly underestimated CH₄ mass transfer under continuous biotic operation of the fermenter. Thus, the instantaneous CH₄ biodegradation at the gas-liquid interphase mediated by methanotrophic bacteria likely supported an enhanced CH₄ mass transfer, similarly to the acceleration in the O₂ mass transfer recorded in the OTR experiments. “

3) kLa should be defined also in the abstract.

The acronym *k_la* was defined in the revised abstract section in accordance to Reviewer 4 suggestion (current page 2).

“This study aimed at systematically comparing the potential of two empirical methods for the estimation of the volumetric mass transfer coefficient of CH₄ (*k_la_{CH₄}*),...”

4) Lacking reference of the two methodologies (Page 5 Lines 24-26).

The corresponding references to the gassing-out and OTR methodologies were included in this revised version of the manuscript in accordance to Reviewer 4 request (current page 5)

“A 3.5 L ez-control fermenter (Applikon, The Netherlands) equipped with two rushton turbines and containing 2.5 L of MSM was used for the determination of the volumetric mass transfer coefficient of O₂ (*k_la_{O₂}*) using the gassing-out and the oxygen transfer rate (OTR) methods (Quijano et al., 2009, Estrada et al., 2014b).”

The new reference (Estrada et al., 2014) was included in the revised reference section

Estrada, J.M., Dudek, A., Muñoz, R., Quijano, G., 2014b. Fundamental study on gas-liquid mass transfer in a biotrickling filter packed with polyurethane foam. Journal of Chemical Technology and Biotechnology. 89(9), 1419-1424.

5) Provide brief description of all the analytics used. Do not directly refer to literatures.

A more detailed description of the analyses conducted was included in the *analytical methods* section along with a new statistical treatment section (current page 8-9)

“A polarographic sensor (Applikon, The Netherlands) and a luminescent/optical dissolved O₂ sensor (Intellical™ LDO101, Hach, USA) were used for the measurement of the dissolved O₂ concentration in the cultivation broth during the gassing-out experiments.

The determination of SO_3^{2-} concentration involved the careful sampling of 5 mL of cultivation broth to minimize their contact with ambient air. The sample was then transferred to a 50-ml Erlenmeyer containing 1 ml of sulfuric acid and 0.1 g of NH_2SO_3H crystals. After 1 ml of a starch solution (1%) was added as indicator, the sample was titrated with standard $KI-KIO_3$ titrant until a faint permanent blue color developed. Similarly, a sample of distilled water was processed as the blank control. The SO_3^{2-} concentration was then calculated according to the following equation 6:

$$[SO_3^{2-}][M] = \frac{[\text{Titrant Volume for Sample (ml)} - \text{Titrant volume for Blank (ml)}]}{\text{Volumen Sample (ml)}} \quad (\text{Eq. 6})$$

Biomass concentration was determined by filtering 10 ml of cultivation broth in dry pre-weighed 0.2 μm acetate cellulose filters, which were then allowed to dry overnight and weighted again.”

2.5 Statistical Treatment

All experiments were performed in duplicate and each assay was considered as a replicate. The results are given as the average value \pm standard deviation from the duplicate assays. In addition, the results were analysed using a one-way ANOVA with significance at $p \leq 0.05$. The Excel statistical package (Microsoft Corporation, USA) was used for data treatment.”

6) In page 6 Eq. 3, it is not defined how the OTR was determined. Please add one more equation or the correct reference. Moreover, the $CO_{2,G}$ is not defined in the text.

The parameters OTR and $CO_{2,G}$ are now defined in this revised manuscript and the corresponding reference was also included in accordance to Reviewer 4 suggestion (current page 7)

“Where OTR is calculated from the slope of the absorbed O_2 concentration (estimated from the SO_3^{2-} concentration multiplied by the O_2/SO_3^{2-} stoichiometric coefficient derived from equation 3) versus time plot, H_{O_2} stands for the dimensionless Henry law constant of O_2 (30.9 at 30 °C) and $Co_{2,G}$ represents the gas O_2 concentration in the exhaust gas (Zhao et al. 1999)”

The new reference (Zhao et al., 1999) was included in the revised reference section

Zhao, S., Kuttuva, S.G., Ju, L.K., 1999. Oxygen transfer characteristics of multiple-phase dispersions simulating water-in-oil xanthan fermentations. *Bioprocess Engineering*. 20, 313-323.

7) Please make sure the references are correct. For instance, Page 6 Line 18, ref. 15 is not describing the sulfite method, I think reference 17 is describing the method.

The citations used along the main text were double checked in accordance to Reviewer 4 request. In addition, former reference 15 was replaced by former reference 17 (current page 6)

“The OTR ($g L^{-1} h^{-1}$) from the gas phase to the MSM was estimated (in duplicate) by periodically monitoring the time course of sulfite oxidation in the fermenter according to Quijano et al. (2009).”

8) A lot of reference formatting errors. Use manual editing.

The authors apologize for the above mentioned formatting errors in the *references* section of the original manuscript. The format of the references was double checked manually in this revised submission in order to avoid further typos. The previous errors were highlighted in red colour (see *references* section).

9) Page 9 lines 24-26. The equation is not well explained.

The equation embedded in the text in former page 9 lines 24-26 (from now on equation 7) was more clearly described in the revised version of the manuscript in accordance to Reviewer 4 request (current page 11).

“These values were similar to the CH_4 -ECs estimated using the OTR $k_{La}CH_4$ at a maximum concentration gradient (equation 7): 35.6 ± 2.2 , 50.1 ± 2.3 and 59.6 ± 3.4 g CH_4 m^{-3} h^{-1} at 400, 600 and 800 rpm, respectively (Figure 2).

$$CH_4 - EC = k_{La}CH_4 \times \left(\frac{C_{CH_4,in}}{H_{CH_4}} - C_{L,CH_4} \right) \quad (Eq. 7)$$

Where H_{CH_4} represents the dimensionless Henry law constant of CH_4 (28.7 at 30 °C) and C_{L,CH_4} the aqueous concentration of CH_4 , which is negligible under the mass transfer limiting scenario here encountered.”

10) Page 9 Lines 11-16, no results are presented in the manuscript to show that abiotic methane removal was negligible. In my opinion, this is an important results considering that you are assuming that O₂ uptake at the gas-liquid inter phase is mimicking the microbial methane uptake (Page 11 lines 6-12).

The results from the abiotic test conducted in the continuous bioreactor are included as Figure S2 in the *supplementary materials* section. In addition, a brief discussion of the implications of the absence of abiotic CH_4 removal in the bioreactor was included in current page 11.

3. Abiotic Test

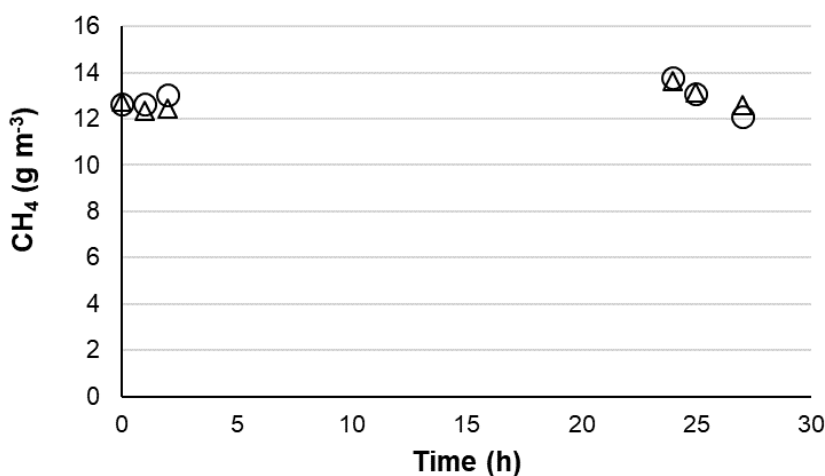


Figure S2. Time course of the inlet (○) and outlet (Δ) CH_4 concentration in the continuous bioreactor during the abiotic test.

“The abiotic test carried out in the fermenter under continuous gas phase supply revealed a negligible contribution of adsorption or photolysis to CH₄ removal in this particular experimental set-up (Figure S2). This implies that the removal of CH₄ recorded at 400, 600 and 800 rpm can be attributed to the biocatalytic action of the microbial consortium present in the fermenter, which converts CH₄ to CO₂, H₂O and new biomass (Lopez et al., 2014).”

Minor points of concern:

11) Page 10 lines 26-29, rephrase " has emerged in the past 5 years based on.... based on the bioconversion.

Former lines 26-29 in page 10 were reworded to *“A renewed attention on suspended growth bioreactors for CH₄ abatement (i.e. stirred tank, airlift and bubble columns) has emerged in the past 5 years due to the popularization of biorefineries based on the feasible bioconversion of CH₄ into added-value bioproducts such as biopolymers, ectoine, exopolysaccharides, protein, etc. (Cantera et al., 2017; Karthikeyan et al., 2015; Khmelenina et al., 2015; Pieja et al., 2012)”* in accordance to Reviewer 4 request (current page 13).

12) Page 10 Line 43. full stop missing at the end of the phrase.

The sentence in former page 10 line 43 was reworded to *“Therefore, the results herein obtained validated the SO₃²⁻ oxidation methodologies as the most accurate method to estimate $k_{lA}CH_4$, which is expected to support in the design of cost-effective CH₄ biorefineries.”* and a full stop was included in accordance to Reviewer 4 request.

13) Many parts of the results and discussion section contain only results. It will be highly beneficial for the reader, if the authors also provide more insights and discuss based on their expertise so far. Some of the results are just stated without proper explanation/scientific discussion. This section has to be revised carefully.

The interpretation and discussion of the results obtained was significantly improved as requested by Reviewer 1, 3 and 4. More sound discussions of the mechanisms underlying O₂ mass transfer during the gassing-out and sulphite oxidation experiment, and CH₄ mass transfer during the biotic experiments, was also provided. More specifically:

Current Page 9

“The type of O₂ sensor (polarographic vs optical) used in the gassing-out methodology did not significantly ($p \leq 0.05$) influence the measurement of k_{lAO_2} (average error < 2.75 %), which confirmed that the significant ($p \leq 0.05$) differences”

Current page 10

“At this point it should be highlighted that moderate salinity media, despite slightly reducing the solubility of gases, typically result in enhanced gas-liquid k_{lA} induced by the higher gas-liquid interfacial areas (Rodero et al., 2018).”

“Figure 2, which depicts the O₂ profiles in the liquid phase during the gassing-out and OTR experiments, clearly shows the higher mass transfer gradients (estimated locally from the slope of the decrease in dissolved O₂ concentration at the gas-liquid interphase) despite the overall difference in O₂ concentration between the interphase and the bulk liquid (200 mg L⁻¹ – 0 mg L⁻¹) remained similar. These higher gradients mediated by the rapid oxidation of O₂ in the presence of SO₃²⁻ and Co²⁺ entailed a superior O₂ mass transfer, which mathematically resulted into higher k_{laO₂} (obtained from equations 1 and 3). Finally, it should be highlighted that the k_{laCH₄} here recorded were similar to those reported in vertical tubular loop (70 h⁻¹) and airlift external loop reactors (85 h⁻¹), but higher than those observed in stirred tank, airlift bubble column and U-loop reactors (12-26 h⁻¹) (Stone et al., 2017).”

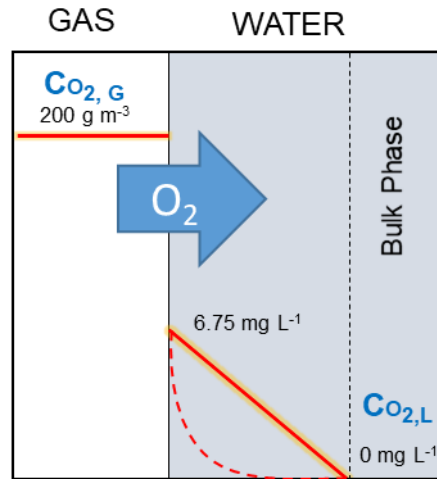


Figure 2. O₂ concentration profiles in the liquid phase during the gassing-out (continuous line) and OTR (dashed line) methods.

Current page 11

“At this point it should be also stressed that high k_{laCH₄} values are required in CH₄ treating bioreactors in order to compensate for the low concentration differences imposed by the high Henry law constant of CH₄ (~ 30 at 25 °C) (Lopez et al., 2013). Thus, the higher the k_{laCH₄} values, the lower the bioreactor volumes and consequently, the lower the investment costs.”

“This implies that the removal of CH₄ recorded at 400, 600 and 800 rpm can be attributed to the biocatalytic action of the microbial consortium present in the fermenter, which converts CH₄ to CO₂, H₂O and new biomass (Lopez et al., 2014).”

Current page 12

$$CH_4 - EC = k_{laCH_4} \times \left(\frac{C_{CH_4,in}}{H_{CH_4}} - C_{L,CH_4} \right) \quad (Eq. 7)$$

“Where H_{CH₄} represents the dimensionless Henry law constant of CH₄ (28.7 at 30 °C) and C_{L,CH₄} the aqueous concentration of CH₄ in the bulk phase, which is negligible under the mass transfer limiting scenario here encountered. In fact, no significant differences (p ≤ 0.05) between the empirical and estimated CH₄-ECs were found at 400 and 800 rpm.”

“In this context, Cu positively regulates the activity of the enzymes pMMO and sMMO (particulate and soluble CH₄ monooxygenase, respectively) and controls the expression of their genes. Most methanotrophs grow optimally at Cu concentrations lower than 270 μg L⁻¹, though previous enzymatic assays have demonstrated that sMMO in type II methanotrophs is properly synthesized at low Cu concentrations (below 50 μg L⁻¹) (Graham et al., 1993; Bender and Conrad, 1995; Semrau et al., 2010).”

“Thus, the rapid CH₄ biodegradation in the vicinity of the gas-liquid interphase mediated by methanotrophic bacteria likely supported an enhanced CH₄ mass transfer (induced by the higher local gradient close to the gas-liquid interphase)”

Current page 13-14

“3.3 Practical applications and future research perspectives

A renewed attention on suspended growth bioreactors for CH₄ abatement (i.e. stirred tank, airlift and bubble columns) has emerged in the past 5 years due to the popularization of biorefineries based on the feasible bioconversion of CH₄ into added-value bioproducts such as biopolymers, ectoine, exopolysaccharides, protein, etc. (Cantera et al., 2017; Karthikeyan et al., 2015; Khmelenina et al., 2015; Pieja et al., 2012). These novel biorefineries, which will be based both on dilute CH₄ emissions from abandoned landfills or coal mines and biogas from solid waste management, will create value out of GHG mitigation and envisage a new bioeconomy based on CH₄. In this context, suspended growth systems represent nowadays the most popular bioreactor platform capable of supporting a cost-competitive methanotrophic biomass harvesting and bioproduct downstream, although recent attempts to produce added-value products from residual CH₄ in immobilized bioreactors have been reported (Cantera et al., 2016; Karthikeyan et al., 2017). Thus, the characterization of CH₄ mass transport from residual emissions or biogas streams into the microbial community is central to dimension the bioreactor and downstream processes. However, the absence of reliable commercial dissolved CH₄ sensors has hindered the development of accurate methodologies for the determination of k_{lCH_4} , whose characterization has been conducted based on either classical k_{lO_2} determination methodologies without a further validation for CH₄ or complex experimental methodologies difficult to implement at large scale (Yu et al., 2010; Rocha-Rios et al., 2011). Therefore, the results herein obtained validated the SO_3^{2-} oxidation methodologies as the most accurate method to estimate k_{lCH_4} , which is expected to support in the design of cost-effective CH₄ biorefineries. For instance, figure S3 depicts a simple model simulation of the CH₄ outlet concentration as a function of the CH₄ inlet concentration and the k_{lCH_4} estimated at the three stirring rates tested. This methodology entails the purchase of SO_3^{2-} and Co^{2+} , and will result in SO_4^{2-} as a final product in the cultivation medium after the experiment. Further research should focus on validating this methodology in high-performance off gas treatment bioreactors such as Taylor flow or two-phase partitioning bioreactors, which are foreseen as the only bioreactor configuration capable of supporting an effective CH₄ bioconversion at a reduced energy consumption (Cantera et al., 2017). In addition, this methodology could be also applied to the design of algal-bacterial photobioreactors devoted to biogas upgrading (CO₂ and H₂S removal), where both chemically and biologically enhanced mass transport has been identified (Rodero et al., 2018, Yang et al., 2016a,b, Zhu, 2015, Zhu et al., 2016).”

REVIEWER 5

Authors present interesting experimental results highlighting the removal of methane under aerated condition.

The authors acknowledge the positive feedback from Reviewer 5

Comments

1) At least for the experimental conditions investigated in the work, they should try to simulate the observed CH₄ concentration once it enters the reactor and leaves the reactor.

A simple model simulation of the CH₄ concentration leaving the reactor was conducted based on the kl_{aCH_4} estimated as requested by Reviewer 5. This information was highlighted in this revised version of the manuscript (current page 14) and comprehensively described in the *supplementary materials* section.

“For instance, figure S3 depicts a simple model simulation of the CH₄ outlet concentration as a function of the CH₄ inlet concentration and the kl_{aCH_4} estimated at the three stirring rates tested.”

4. Application of the kl_{aCH_4} for process design

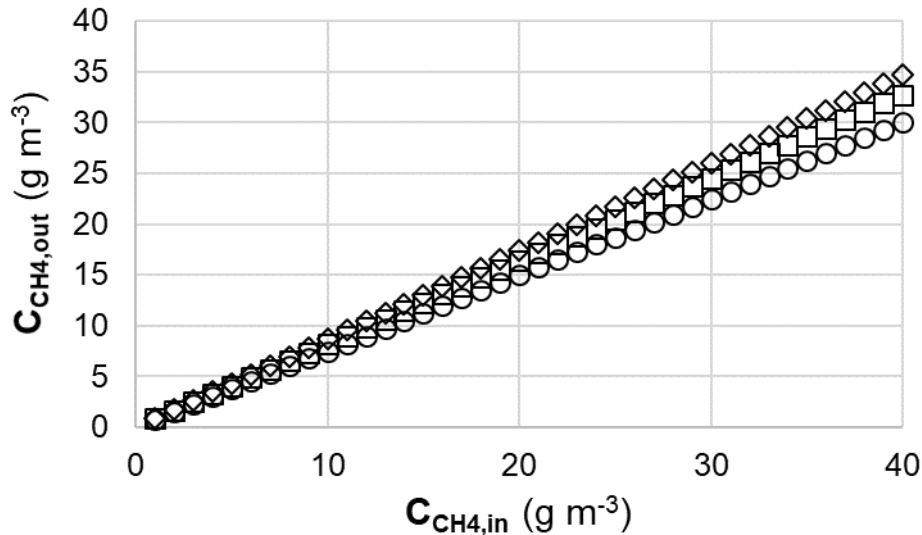


Figure S3. Model simulation of the CH₄ outlet concentration as a function of the CH₄ inlet concentration and the kl_{aCH_4} estimated at 400 (◇), 600 (□) and 800 rpm (○).

The CH₄ outlet concentration, $C_{CH_4,out}$, was derived from the definition of the CH₄ elimination capacity as follows (equation S1 and S2)

$$CH_4 - EC = \frac{Q}{V} \times (C_{CH_4,in} - C_{CH_4,out}) = kl_{aCH_4} \times \left(\frac{C_{CH_4,in}}{H_{CH_4}} - C_{L,CH_4} \right) \quad (Eq. S1)$$

$$C_{CH_4,out} = \frac{\left(\frac{Q}{V} \times C_{CH_4,in}\right)}{\left(\frac{kl a_{CH_4}}{H_{CH_4}} + \frac{Q}{V}\right)} \quad (Eq. S2)$$

2) Governing differential equations, used if any, must be stated along with factors affecting the kinetic parameters. This will help in further improvement of the quality of work.

The authors also acknowledge the relevance of the determination of the microbial kinetic parameters, but unfortunately, the study was focused on investigating process performance under mass transfer limiting conditions and steady state (where microbial kinetics are not governing the process). Similarly, since the process was studied under steady state conditions, no differential equation was used in the description of the process.

3) On page 5, line 56, in place of Mexico city, temperature should be a better choice.

The temperature used in the experiments carried out in the fermenter was included in this revised version of the manuscript in accordance to Reviewer 5 request (current page 6).

“Where $C_{O_2L}^$ stands for the dissolved O_2 concentration at saturation (6.75 mg L^{-1} at Mexico city at $30 \text{ }^\circ\text{C}$), ...”*

4) on page 7: line 50, basis for these temperatures should be provided.

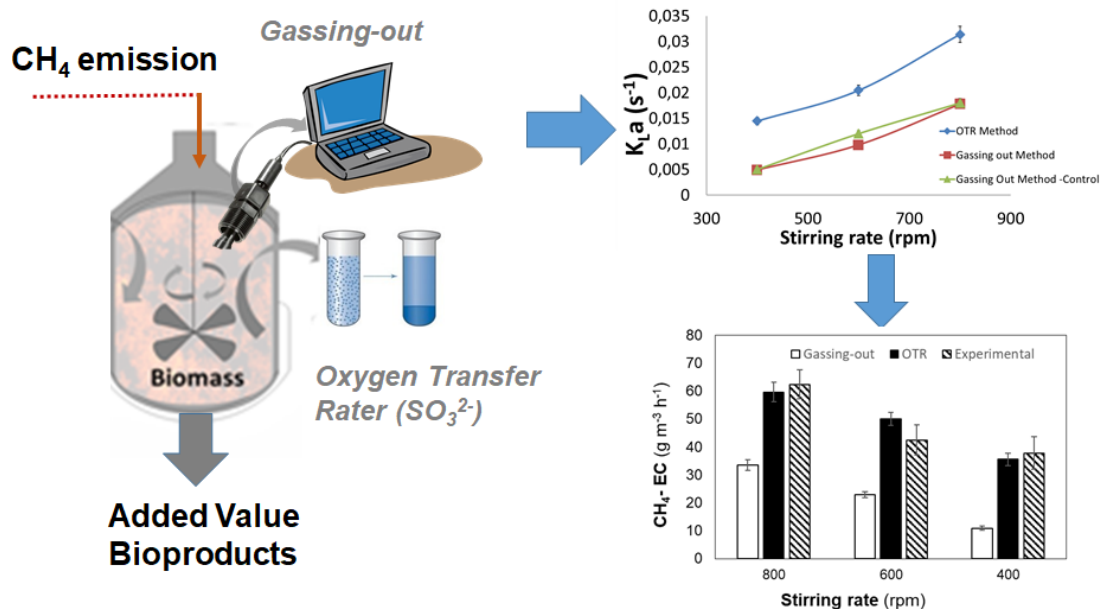
The temperatures set in the GC-TCD method in the injector, oven and detector were recommended by the manufacturer of the column for the separation of permanent gases. This information was provided in the revised manuscript in order to avoid any further misunderstanding (current page 8)

“The injector, oven and detector temperatures were maintained at 50, 40, 115 $^\circ\text{C}$, respectively, as recommended by the column manufacturer for the separation of permanent gases. “

GUEST EDITOR PROBLEMS

1) Check the quality of the image provided in the graphical abstract. The graph should be readable and understandable.

The quality of the graphical abstract was significantly improved (Arial font also used) in this revised submission as requested by the guest editor (and also by Reviewer 4).



2) Use times new roman, 12 font size in the document.

Times New Roman 12 is used along this revised manuscript in accordance to the request from the guest editor.

3) Use Arial font type for the graphs.

Arial font was used in the text and numbers contained in figures 1, 2 and 3 (former figure 2) as requested by the guest editor.

Figure 1

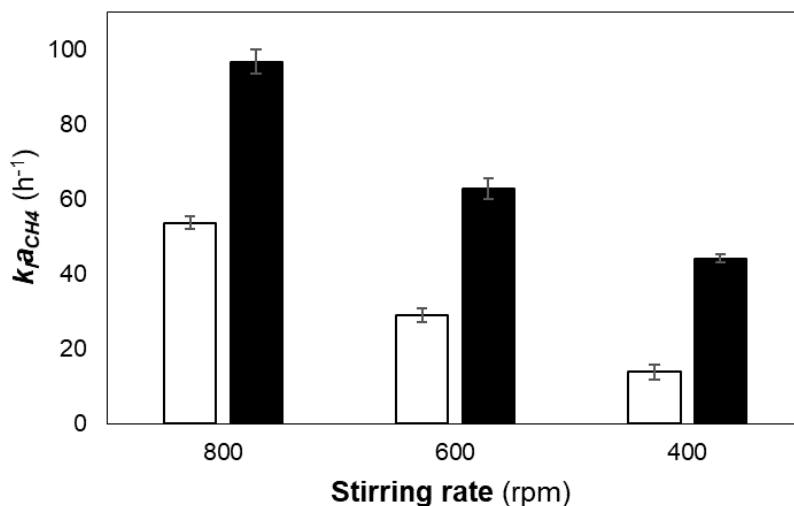


Figure 2

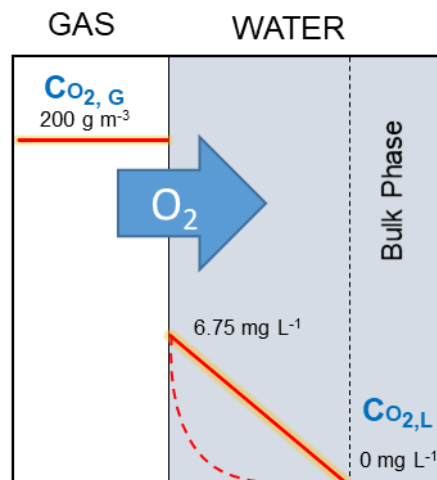
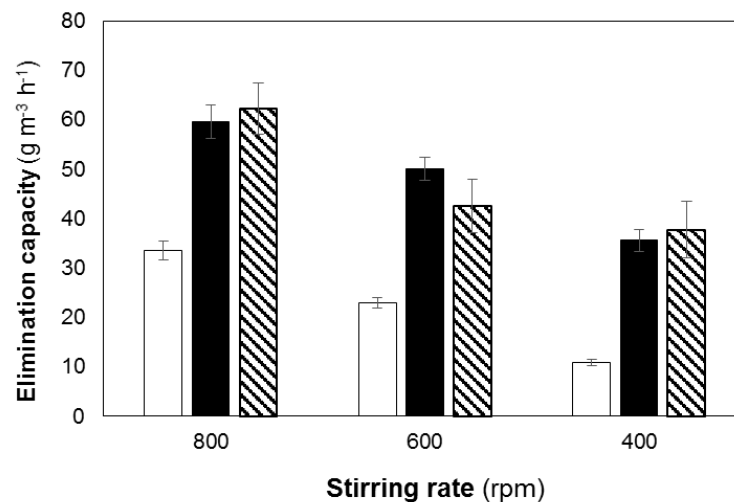


Figure 3



4) Write CH₄ as methane in the title.

The title of the paper was modified in accordance to Reviewer 2 request to “*A systematic comparison of two empirical gas-liquid mass transfer determination methodologies to characterize methane biodegradation in stirred tank bioreactors*”, which does not contain the acronym “CH₄” (current page 1)

5) Write the research gaps of this topic in 3-4 sentences.

The current research gaps in this topic are now clearly identified in the revised version of the introduction section (current page 4) as requested by the guest editor:

“Unfortunately, the validation of empirical methodologies for the determination of the volumetric CH₄ mass transfer coefficient (k_{LCH_4}) has been poorly addressed in literature (García-Ochoa et al., 2009; Rocha-Ríos et al., 2010). Indeed, the lack of cheap and reliable dissolved CH₄ sensors along with the low solubility and limited industrial interest of CH₄ to date have restricted the development of accurate methodologies for the characterization of

k_lCH₄. Table S1 presents the main studies devoted to the characterization of the volumetric CH₄ mass transfer coefficients in bioreactors.”

6) Write the novelty of this work in 3-4 sentences.

The novelty of this work was highlighted in the introduction section of this revised version of the manuscript as requested by the guest editor and Reviewer 3 (current page 4)

“This study aims at comparing and validating, for the first time, two experimental methodologies for the estimation of k_lCH₄ (gassing-out vs oxygen transfer rate methods) in order to accurately describe CH₄ abatement in a stirred tank reactor operated with a methanotrophic consortium at multiple stirring rates. This information is expected to provide the basis for a correct bioreactor design for CH₄ bioconversion into added-value products.”

7) Provide the manufacture details of all equipment's, sensors, instruments used and their operational conditions clearly. Some are missing.

While the details of the fermenter and GC-TCD, along with their operational conditions, were initially provided in the original manuscript (former page 5 lines 19-29, former page 7 line 45-53), additional information of the sensors and analytical methodologies used in the study was included in current page 8-9:

“A polarographic sensor (Applikon, The Netherlands) and a luminescent/optical dissolved O₂ sensor (Intellical™ LDO101, Hach, USA) were used for the measurement of the dissolved O₂ concentration in the cultivation broth during the gassing-out experiments.

The determination of SO₃²⁻ concentration involved the careful sampling of 5 mL of cultivation broth to minimize their contact with ambient air. The sample was then transferred to a 50-ml Erlenmeyer containing 1 ml of sulfuric acid and 0.1 g of NH₂SO₃H crystals. After 1 ml of a starch solution (1%) was added as indicator, the sample was titrated with standard KI-KIO₃ titrant until a faint permanent blue color developed. Similarly, a sample of distilled water was processed as the blank control. The SO₃²⁻ concentration was then calculated according to the following equation 6:

$$[SO_3^{2-}][M] = \frac{[\text{Titrant Volume for Sample (ml)} - \text{Titrant volume for Blank (ml)}]}{\text{Volumen Sample (ml)}} \quad (\text{Eq. 6})$$

Biomass concentration was determined by filtering 10 ml of cultivation broth in dry pre-weighed 0.2 μm acetate cellulose filters, which were then allowed to dry overnight and weighted again.”

8) Information on statistical analysis will be useful for the reader. Also, provide such information in the results and discussions section.

A new section on statistical data treatment (and its associated analyses) was included in both the *Material & Methods* and *Results & Discussion* sections of this revised version of the manuscript as requested by the guest editor (current page 9).

“2.5 Statistical Treatment

All experiments were performed in duplicate and each assay was considered as a replicate. The results are given as the average value \pm standard deviation from the duplicate assays. In addition, the results were analysed using a one-way ANOVA with significance at $P \leq 0.05$. The Excel statistical package (Microsoft Corporation, USA) was used for data treatment.”

The revised *Results & Discussion* section also included the main findings of the statistical analysis as requested by the guest editor (current page 9 and 12)

“The type of oxygen sensor (polarographic vs optical) used in the gassing-out methodology did not significantly ($p \leq 0.05$) influence the measurement of k_{lO_2} (average error $< 2.75\%$), which confirmed that the significant ($p \leq 0.05$) differences observed in k_{lCH_4} between the OTR and gassing-out methodologies were not caused by the delay in the response of the polarographic sensor.”

“In fact, no significant differences ($p \leq 0.05$) between the empirical and estimated CH_4 -ECs were found at 400 and 800 rpm.”

9) Write the abstract in less than 150 words and also the conclusions in less than 150 words.

The abstract and conclusion sections were significantly modified in order to highlight the main results and provide a more systematic explanation of the findings obtained as requested by Reviewer 1. In addition, the length of both sections was reduced to 150 words as requested by the guest editor (current pages 2 and 14-15).

Abstract (150 words)

“This study aimed at systematically comparing the potential of two empirical methods for the estimation of the volumetric CH_4 mass transfer coefficient (k_{lCH_4}), namely gassing-out and oxygen transfer rate (OTR), to describe CH_4 biodegradation in a fermenter operated with a methanotrophic consortium at 400, 600 and 800 rpm. The k_{lCH_4} estimated from the OTR methodology accurately predicted the CH_4 elimination capacity (EC) under CH_4 mass transfer limiting conditions regardless of the stirring rate ($\sim 9\%$ of average error between empirical and estimated ECs). Thus, empirical CH_4 -ECs of 37.8 ± 5.8 , 42.5 ± 5.4 and 62.3 ± 5.2 $g\ CH_4\ m^{-3}\ h^{-1}$ vs predicted CH_4 -ECs of 35.6 ± 2.2 , 50.1 ± 2.3 and 59.6 ± 3.4 $g\ CH_4\ m^{-3}\ h^{-1}$ were recorded at 400, 600 and 800 rpm, respectively. The rapid Co^{2+} -catalyzed reaction of O_2 with SO_3^{2-} in the vicinity of the gas-liquid interphase during OTR determinations, mimicking microbial CH_4 uptake in the biotic experiments, was central to accurately describe the k_{lCH_4} .”

Conclusions (119)

“This work empirically validated a simple methodology to accurately describe CH_4 mass transport and biodegradation in continuous stirred tank bioreactors based on OTR estimations. The rapid O_2 uptake at the gas-liquid interphase mediated by sulfite and Co^{2+} , mimicking microbial CH_4 uptake during continuous CH_4 treatment, supported the superior performance of the OTR methodology for the description of k_{lCH_4} compared to the classical gassing-out methodology. Thus, the empirical CH_4 -ECs of 37.8 ± 5.8 , 42.5 ± 5.4 and 62.3 ± 5.2 $g\ CH_4\ m^{-3}\ h^{-1}$ recorded at 400, 600 and 800 rpm, respectively, were comparable to the CH_4 -ECs of 35.6 ± 2.2 , 50.1 ± 2.3 and 59.6 ± 3.4 $g\ CH_4\ m^{-3}\ h^{-1}$ predicted by the OTR methodology.

This finding will help in the design of suspended growth bioreactors devoted to CH₄ bioconversion.”

10) Do not use [1], [2] numbering style for references. We kindly suggest you to write the names of the authors in the text.

The style to refer to the citations in the main text was modified as requested by the guest editor.

11) Do not use UNWANTED upper case alphabets in the text without any reason. There are several such mistakes in the entire document.

The terms “Oxygen Transfer Rate”, “Equation”, “Helium” and “Rushton” now read “*oxygen transfer rate*”, “*equation*”, “*helium*”, “*rushton*” along the main text as recommended by the guest editor.

12) Provide more details of the analytical techniques, the columns, the elution time, etc.

The *Analytical methods* section was significantly improved in this revised version of the manuscript by providing the description of the dissolved oxygen sensors, and the detailed methodologies for the determination of biomass and sulphite concentrations (current page 8-9):

“A polarographic sensor (Applikon, The Netherlands) and a luminescent/optical dissolved O₂ sensor (Intellical™ LDO101, Hach, USA) were used for the measurement of the dissolved O₂ concentration in the cultivation broth during the gassing-out experiments.

The determination of SO₃²⁻ concentration involved the careful sampling of 5 mL of cultivation broth to minimize their contact with ambient air. The sample was then transferred to a 50-ml Erlenmeyer containing 1 ml of sulfuric acid and 0.1 g of NH₂SO₃H crystals. After 1 ml of a starch solution (1%) was added as indicator, the sample was titrated with standard KI-KIO₃ titrant until a faint permanent blue color developed. Similarly, a sample of distilled water was processed as the blank control. The SO₃²⁻ concentration was then calculated according to the following equation 6:

$$[SO_3^{2-}][M] = \frac{[\text{Titrant Volume for Sample (ml)} - \text{Titrant volume for Blank (ml)}]}{\text{Volumen Sample (ml)}} \quad (\text{Eq. 6})$$

Biomass concentration was determined by filtering 10 ml of cultivation broth in dry pre-weighted 0.2 μm acetate cellulose filters, which were then allowed to dry overnight and weighted again.”

13) What is fast kinetics? How fast is fast?

The presence of cobalt in the aqueous phase increases the kinetics of sulphite oxidation as previously reported by Zhao et al. (1999). This information is now included in the *Material & Method* section of the revised manuscript (current page 6)

“The kinetics of O_2 absorption were recorded after the addition of the catalyst (5 mL of a 2.5×10^{-4} M $CoSO_4$ solution, which supports an instantaneous SO_3^{2-} oxidation) until complete SO_3^{2-} depletion according to equation 2. “

In addition, the reference to the original method of determination of oxygen mass transfer rate was included in the revised paper (current page 10)

“The extremely fast (almost instantaneous) kinetics of this reaction (faster than any gas-liquid diffusional mechanisms in the mass transfer process) in the presence of Co^{2+} likely accelerated O_2 uptake from the gas phase in the fermenter (Zhao et al. 1999),..”

Zhao, S., Kuttuva, S.G., Ju, L.K., 1999. Oxygen transfer characteristics of multiple-phase dispersions simulating water-in-oil xanthan fermentations. Bioprocess Engineering 20, 313- 323

14) The results and discussions section should be divided according to the specific objectives of this study.

The *Results and Discussion* section in this revised manuscript was divided into three subsections in accordance to the specific objectives of the study (current page 9 and 10)

3.1 Volumetric mass transfer coefficient determination

3.2 CH_4 abatement under continuous operation: methodology validation

3.3 Practical applications and future research perspectives

15) Write a new section on practical applications and future research perspectives, before the conclusions section. In this new section, the authors should explain, how the results of this work can be applied in the successful operation of bioreactors, what are the limitations, what future work can be done, etc. Write at least one page of information.

A new section discussing the practical applications of the results herein obtained and future research perspectives, including all items requested by the guest editor, was included in this revised version of the manuscript (current pages 13-14). In addition, a practical application of the kl_{CH_4} determined is shown in the *supplementary materials* section.

“3.3 Practical applications and future research perspectives

A renewed attention on suspended growth bioreactors for CH_4 abatement (i.e. stirred tank, airlift and bubble columns) has emerged in the past 5 years due to the popularization of biorefineries based on the feasible bioconversion of CH_4 into added-value bioproducts such as biopolymers, ectoine, exopolysaccharides, protein, etc. (Cantera et al., 2017; Karthikeyan et al., 2015; Khmelenina et al., 2015; Pieja et al., 2012). These novel biorefineries, which will be based both on dilute CH_4 emissions from abandoned landfills or coal mines and biogas from solid waste management, will create value out of GHG mitigation and envisage a new bioeconomy based on CH_4 . In this context, suspended growth systems represent nowadays the most popular bioreactor platform capable of supporting a cost-competitive methanotrophic biomass harvesting and bioproduct downstream, although

recent attempts to produce added-value products from residual CH_4 in immobilized bioreactors have been reported (Cantera et al., 2016; Karthikeyan et al., 2017). Thus, the characterization of CH_4 mass transport from residual emissions or biogas streams into the microbial community is central to dimension the bioreactor and downstream processes. However, the absence of reliable commercial dissolved CH_4 sensors has hindered the development of accurate methodologies for the determination of k_{laCH_4} , whose characterization has been conducted based on either classical k_{laO_2} determination methodologies without a further validation for CH_4 or complex experimental methodologies difficult to implement at large scale (Yu et al., 2010; Rocha-Rios et al., 2011). Therefore, the results herein obtained validated the SO_3^{2-} oxidation methodologies as the most accurate method to estimate k_{laCH_4} , which is expected to support in the design of cost-effective CH_4 biorefineries. For instance, figure S3 depicts a simple model simulation of the CH_4 outlet concentration as a function of the CH_4 inlet concentration and the k_{laCH_4} estimated at the three stirring rates tested. This methodology entails the purchase of SO_3^{2-} and Co^{2+} , and will result in SO_4^{2-} as a final product in the cultivation medium after the experiment. Further research should focus on validating this methodology in high-performance off gas treatment bioreactors such as Taylor flow or two-phase partitioning bioreactors, which are foreseen as the only bioreactor configuration capable of supporting an effective CH_4 bioconversion at a reduced energy consumption (Cantera et al., 2017). In addition, this methodology could be also applied to the design of algal-bacterial photobioreactors devoted to biogas upgrading (CO_2 and H_2S removal), where both chemically and biologically enhanced mass transport has been identified (Rodero et al., 2018, Yang et al., 2016a,b, Zhu, 2015, Zhu et al., 2016).”

3. Application of the k_{laCH_4} for process design

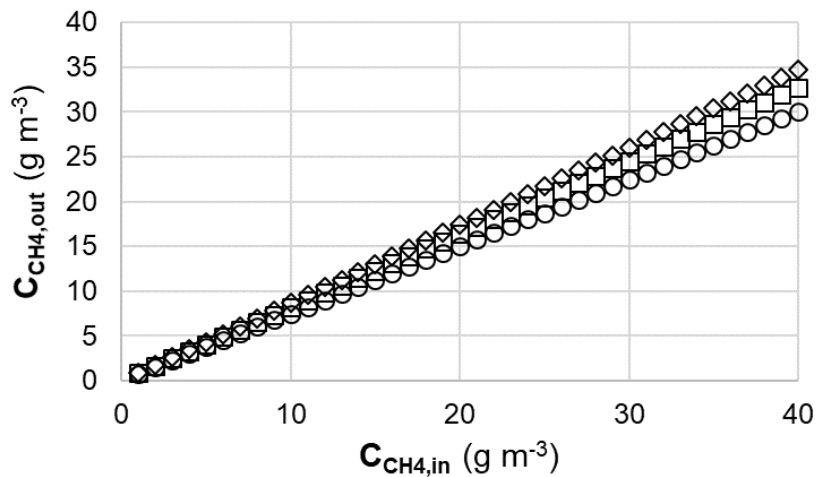


Figure S3. Model simulation of the CH_4 outlet concentration as a function of the CH_4 inlet concentration and the k_{laCH_4} estimated at 400 (\diamond), 600 (\square) and 800 rpm (\circ).

The CH_4 outlet concentration, $C_{\text{CH}_4,\text{out}}$, was derived from the definition of the CH_4 elimination capacity as follows (equation S1 and S2)

$$CH_4 - EC = \frac{Q}{V} \times (C_{CH_4,in} - C_{CH_4,out}) = kla_{CH_4} \times \left(\frac{C_{CH_4,in}}{H_{CH_4}} - C_{L,CH_4} \right) \quad (Eq. S1)$$

$$C_{CH_4,out} = \frac{\left(\frac{Q}{V} \times C_{CH_4,in} \right)}{\left(\frac{kla_{CH_4}}{H_{CH_4}} + \frac{Q}{V} \right)} \quad (Eq. S2)$$

16) References

- Should be carefully formatted. There are a lot of unwanted capital letters, small letters, etc. Please check LINE by LINE.
- There are SEVERAL punctuation errors.
- Journal names are not written in a consistent manner.
- Check a recent issue of JEMA or read the author guidelines carefully.
- Check how to refer to conference proceedings.
- Some author names are written without accents.
- Biological names should be written in italics.

The reference section was carefully double checked in order to avoid unwanted capital letters and punctuation errors, and to include biological names in italics and the journal names in a consistent manner (see revised *references* section with changes highlighted in red)

17) Once again, check all the formatting, spacing and sentence structure mistakes. As pointed out by the reviewers, there are several minor editing mistakes in the document.

The manuscript was double-checked prior submission to minimize the number of editing errors (typos, formatting, spacing and sentence structure mistakes) as requested by the guest editor.

18) The role of Co²⁺ and the mechanism should be discussed further.

The presence of cobalt in the aqueous phase increases the kinetics of sulphite oxidation as previously reported by Zhao et al. (1999). This information is now included in the *Material & Methoda* section of the revised manuscript (current page 6)

“The kinetics of O₂ absorption were recorded after the addition of the catalyst (5 mL of a 2.5×10⁻⁴ M CoSO₄ solution, which supports an instantaneous SO₃²⁻ oxidation) until complete SO₃²⁻ depletion according to equation 2. “

In addition, the reference to the original method of determination of oxygen mass transfer rate was included in the revised paper (current page 10)

“The extremely fast (almost instantaneous) kinetics of this reaction (faster than any gas-liquid diffusional mechanisms in the mass transfer process) in the presence of Co²⁺ likely accelerated O₂ uptake from the gas phase in the fermenter (Zhao et al. 1999),...”

Zhao, S., Kuttuva, S.G., Ju, L.K., 1999. Oxygen transfer characteristics of multiple-phase dispersions simulating water-in-oil xanthan fermentations. Bioprocess Engineering 20, 313- 323

19) What are the reasons for the under-estimation of the mass transfer values using the conventional method.

The reasons underlying the under-estimation of the mass transfer values using the conventional gassing-out methods were better explained in this revised manuscript in accordance to Reviewer 1 and the guest editor (current page 10)

“Therefore, the difference in the volumetric mass transfer coefficient of O_2 , which directly impacted on the estimation of $k_{l}a_{CH_4}$, was attributed to a mass transfer acceleration mediated by the instantaneous chemical oxidation of O_2 by SO_3^{2-} (Eq. 2). The extremely fast (almost instantaneous) kinetics of this reaction (faster than any gas-liquid diffusional mechanisms in the mass transfer process) in the presence of Co^{2+} likely accelerated O_2 uptake from the gas phase in the fermenter (Zhao et al., 1999), which rendered the classical two-films model not valid since the O_2 concentration gradients in the liquid phase were no longer linear under this particular scenario (Van Krevelen et al., 1948). Figure 2, which depicts the O_2 profiles in the liquid phase during the gassing-out and OTR experiments, clearly shows the higher mass transfer gradients (estimated locally from the slope of the decrease in dissolved O_2 concentration at the gas-liquid interphase) despite the overall difference in O_2 concentration between the interphase and the bulk liquid ($200 \text{ mg L}^{-1} - 0 \text{ mg L}^{-1}$) remained similar. These higher gradients mediated by the rapid oxidation of O_2 in the presence of SO_3^{2-} and Co^{2+} entailed a superior O_2 mass transfer, which mathematically resulted into higher $k_{l}a_{O_2}$ (obtained from equations 1 and 3)”

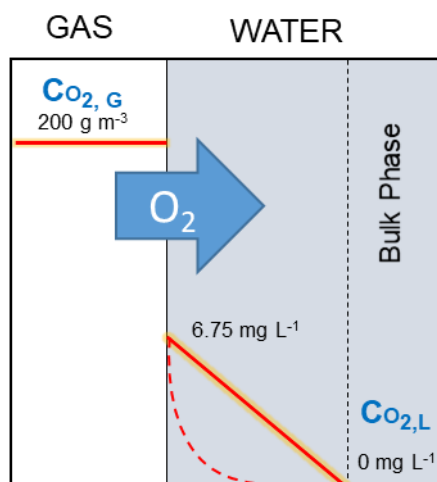


Figure 2. O_2 concentration profiles in the liquid phase during the gassing-out (continuous line) and OTR (dashed line) methods.

20) Be consistent in writing chemical names. Sometimes the authors are writing CH_4 , then sometimes, methane. Similarly, for a lot of other chemical names.

“Sulphite” was replaced by “ SO_3^{2-} ”, “oxygen” was replaced by “ O_2 ” and “methane” was replaced by “ CH_4 ” in the main text of this revised manuscript in order to be consistent with the writing of the names of chemical species.

Kind advise:

A) Provide a separate word document on "authors response to reviewers comments" and explain how all the comments were incorporated. Highlight all the textual changes in YELLOW colour. Remove all the track changes and comments from the word file.

The authors included in this document a meticulous description of the changes carried out in the main text in response to the Reviewers' queries. In fact, an itemized response to all comments of each reviewer, along with a description of the changes performed in the revised manuscript, was included in this document. The changes performed in the main manuscript and *supplementary materials* files were however highlighted in **RED** colour (for a better readability).

B) Please do not write "As suggested by the reviewer, we have addressed all the comments". Instead, the authors are advised to show, what changes were made, and in which section or page.

The authors included in this document a meticulous description of the changes carried out in the main text in response to the Reviewers' queries. In fact, an itemized response to all comments of each reviewer, along with a description of the changes performed in the revised manuscript, was included in this document.

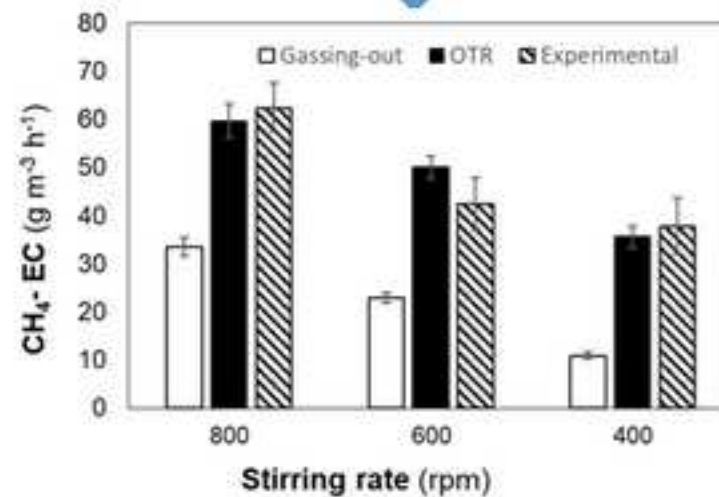
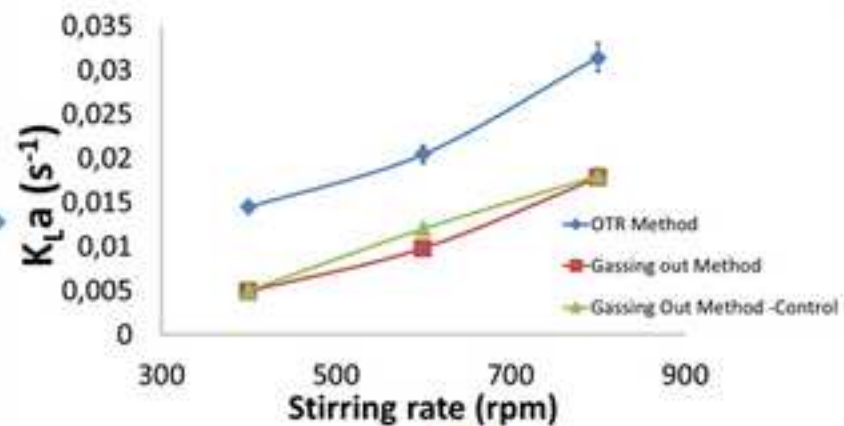
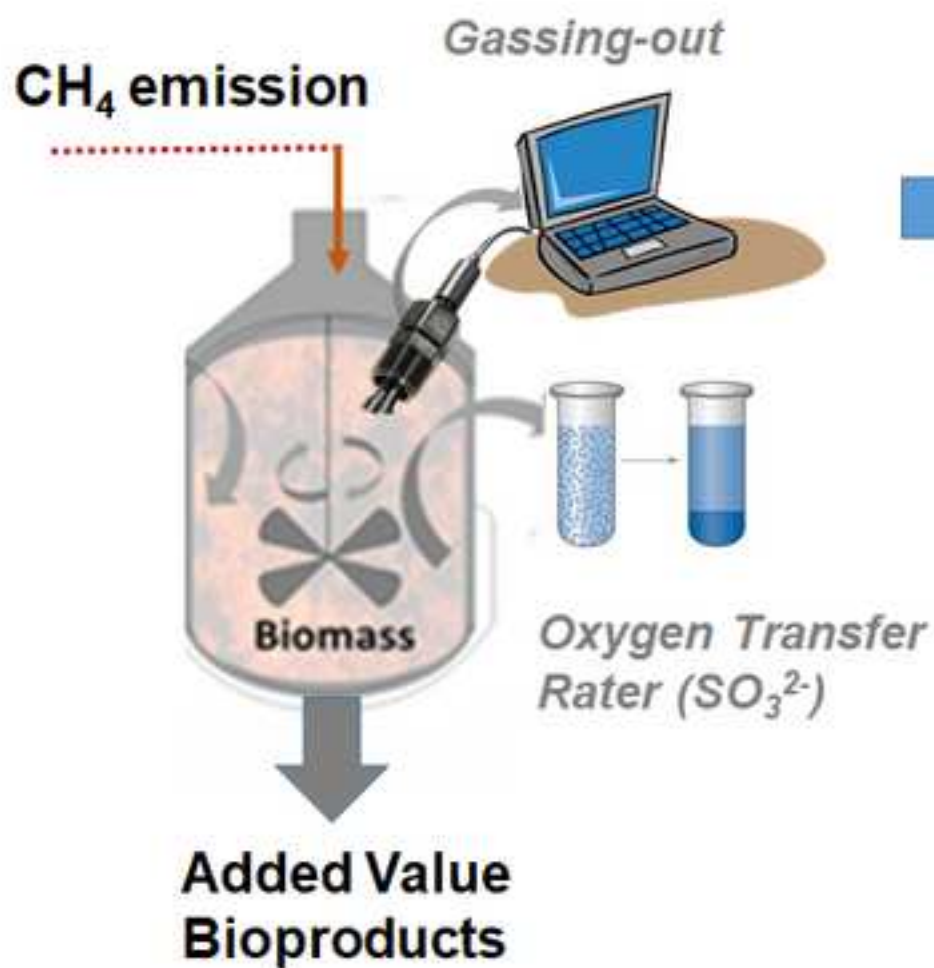
C) We strongly advise all the authors of this manuscript to carefully check the manuscript, at least 2 times, and correct all minor edits. Its very important for JEMA.

The manuscript was double-checked prior submission to minimize the number of typos as requested by the guest editor.

We hope that these modifications will comply with the requests of *Journal of Environmental Management*. Please do not hesitate to contact us at your convenience if you need further information

Raúl Muñoz

Sergio Revah Moissew



k_{lO_2} determination based on SO_3^{-2} oxidation accurately described CH_4 mass transfer

Microbial CH_4 uptake accelerated CH_4 mass transport similarly to SO_3^{-2} oxidation

The gassing-out methodology underestimated CH_4 mass transport

Maximum CH_4 elimination capacities of $62 \pm 5 \text{ g } CH_4 \text{ m}^{-3} \text{ h}^{-1}$ were achieved at 800 rpm

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4 **A systematic comparison of two empirical gas-liquid mass transfer**
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7 **determination methodologies to characterize methane biodegradation in**
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10 **stirred tank bioreactors**

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12 Muñoz^{a,b} R, Soto C^a, Zuñiga C^a, Sergio Revah^{a*}
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1
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4 **Abstract**
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6 This study aimed at systematically comparing the potential of two empirical methods for
7 the estimation of the volumetric CH₄ mass transfer coefficient (k_{laCH_4}), namely gassing-out
8 and oxygen transfer rate (OTR), to describe CH₄ biodegradation in a fermenter operated
9 with a methanotrophic consortium at 400, 600 and 800 rpm. The k_{laCH_4} estimated from the
10 OTR methodology accurately predicted the CH₄ elimination capacity (EC) under CH₄ mass
11 transfer limiting conditions regardless of the stirring rate (~ 9 % of average error between
12 empirical and estimated ECs). Thus, empirical CH₄-ECs of 37.8±5.8, 42.5±5.4 and
13 62.3±5.2 g CH₄ m⁻³ h⁻¹ vs predicted CH₄-ECs of 35.6±2.2, 50.1±2.3 and 59.6±3.4 g CH₄ m⁻³
14 h⁻¹ were recorded at 400, 600 and 800 rpm, respectively. The rapid Co²⁺-catalyzed
15 reaction of O₂ with SO₃⁻² in the vicinity of the gas-liquid interphase during OTR
16 determinations, mimicking microbial CH₄ uptake in the biotic experiments, was central to
17 accurately describe the k_{laCH_4} .
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38 **Keywords:** Biological gas treatment; Greenhouse gas; CH₄; Suspended growth bioreactor;
39 Volumetric mass transfer coefficient.
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4 **1. Introduction**
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6 The assessment of current global greenhouse gas (GHG) emissions based on the 1-year
7 pulse emission analysis has confirmed that CH₄ entails a contribution in terms of global
8 warming potential (GWP) or global temperature change potential comparable to that of
9 CO₂ for short time horizons (IPCC 2013). This fact can be attributed to the high GWP of
10 CH₄ (90 and 72 on a 10 and 20 years horizon, respectively) (Global-Change 2012).
11 However, despite the environmental relevance of CH₄ emissions, the development of cost-
12 efficient and environmentally friendly treatment technologies (especially of those intended
13 for the treatment of diluted emissions where no energy recovery is technically feasible) has
14 been scarce. Today, physical/chemical CH₄ abatement technologies, such as activated
15 carbon adsorption and incineration, are either inefficient or costly at the low concentrations
16 typically found in emissions from waste treatment facilities (except young landfills), mines
17 and animal farms, and possess a large CO₂ footprint as a result of their intensive energy use
18 (Estrada et al., 2012; Melse and Van Der Werf 2009).
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40 Biological methods represent an environmentally friendly alternative to physical/chemical
41 methods for the abatement of CH₄. Biotechnologies, which rely on the biocatalytic action
42 of specialized microorganisms, have been consistently proven as robust and efficient
43 methods for the treatment of malodorous and industrial volatile organic compounds,
44 exhibiting lower operating costs and environmental impacts than their physical/chemical
45 counterparts (Estrada et al., 2012; Ferdowsi et al., 2017; Lopez et al., 2014). However, the
46 cost-efficient application of conventional biotechnologies such as biofiltration or
47 biotrickling filtration for the abatement of CH₄ is often limited by the poor mass transfer of
48 this GHG from the gas emission to the methanotrophic community. This entails the need
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4 for large gas residence times (0.1-10 h), and therefore large bioreactor volumes and
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6 investment costs, which typically jeopardizes the economics of CH₄ abatement (Lopez et
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8 al., 2013). In this context, the bioconversion of CH₄ into added-value products in stirred
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10 tank bioreactors represents a promising alternative to enhance the economic sustainability
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12 of CH₄ abatement (Garcia-Perez et al., 2018; Karthikeyan et al., 2015; Pieja et al., 2012)
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15 Indeed, CH₄ can be bioconverted into biopolymers, ectoine, protein, exopolysaccharides,
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17 etc., using a biorefinery approach (Cantera et al., 2017). The development of cost-
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19 competitive CH₄ biorefineries, inherently limited by CH₄ supply to the microbial
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21 community, requires the development of accurate methodologies for the characterization of
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23 CH₄ gas-liquid mass transfer in conventional fermenters (Cantera et al., 2017).
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25 Unfortunately, the validation of empirical methodologies for the determination of the
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27 volumetric CH₄ mass transfer coefficient ($k_{l}a_{CH_4}$) has been poorly addressed in literature
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29 (García-Ochoa et al., 2009; Rocha-Ríos et al., 2010). Indeed, the lack of cheap and reliable
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31 dissolved CH₄ sensors along with the low solubility and limited industrial interest of CH₄ to
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33 date have restricted the development of accurate methodologies for the characterization of
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35 $k_{l}a_{CH_4}$. Table S1 presents the main studies devoted to the characterization of the volumetric
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37 CH₄ mass transfer coefficients in bioreactors.
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48 This study aims at comparing and validating, for the first time, two experimental
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50 methodologies for the estimation of $k_{l}a_{CH_4}$ (gassing-out vs oxygen transfer rate methods) in
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52 order to accurately describe CH₄ abatement in a stirred tank reactor operated with a
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54 methanotrophic consortium at multiple stirring rates. This information is expected to
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56 provide the basis for a correct bioreactor design for CH₄ bioconversion into added-value
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2. Material and methods

2.1. Microorganism and mineral salt medium

A mixed methanotrophic bacterial consortium containing *Sphingobacterium sp.* CZ-UAM (GenBankKJ411920) was used for CH₄ biodegradation purposes (Steffani-Vallejo et al., 2017). The consortium was enriched from the UAM-Iztapalapa wastewater treatment plant (México City). The composition of the mineral salt medium (MSM) used in the experiments was as follows (g l⁻¹): K₂HPO₄ 0.7, KH₂PO₄ 0.54, MgSO₄·7H₂O 1, CaCl₂·2H₂O 0.2, FeSO₄·7H₂O 0.004, NH₄CL 0.5, ZnSO₄·7H₂O 0.1, MnCl₂·4H₂O 0.03, H₃BO₃ 0.3, CoCl₂·6H₂O 0.2, CuCl₂·2H₂O 0.01, NiCl₂·6H₂O 0.02 and NaMoO₄·2H₂O 0.06 (ATCC medium 1683). All chemicals were purchased from Sigma-Aldrich (USA). CH₄ (99.9%) was obtained from Praxair (Mexico).

2.2 Determination of the volumetric CH₄ mass transfer coefficient

A 3.5 L *ez*-control fermenter (Applikon, The Netherlands) equipped with two rushton turbines and containing 2.5 L of MSM was used for the determination of the volumetric mass transfer coefficient of O₂ ($k_{l}a_{O_2}$) using the gassing-out and the oxygen transfer rate (OTR) methods (Quijano et al., 2009, Estrada et al., 2014). The $k_{l}a_{O_2}$ was recorded at an aeration rate of 0.42 L min⁻¹ under three stirring rates (400, 600 and 800 rpm) at 30 °C. These aeration rates were selected as potential operational conditions implemented in large-scale fermenters devoted to the bioconversion of CH₄.

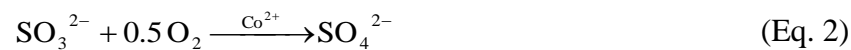
In the gassing-out method, the MSM was initially degassed inside the fermenter for 15 min with argon at 0.42 L min⁻¹ at the corresponding stirring rate (to complete dissolved O₂

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4 depletion) and the dissolved O₂ concentration (DO) was then recorded every 5 seconds
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6 following the initiation of the aeration. The tests were carried out in duplicate using both
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8 polarographic and luminescent/optical dissolved O₂ sensors calibrated before each
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10 experimental series. The $k_1 a_{O_2}$ (h⁻¹) was determined using equation 1:
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$$k_1 a_{O_2} = -\ln \left(\frac{C_{O_2,L}^* - C_{O_2,L}}{C_{O_2,L}^* - (C_{O_2,L})_0} \right) / t \quad (\text{Eq. 1})$$

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21 Where $C_{O_2,L}^*$ stands for the dissolved O₂ concentration at saturation (6.75 mg L⁻¹ at Mexico
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23 city at 30 °C), $C_{O_2,L}$ the dissolved O₂ concentration at time t , and $(C_{O_2,L})_0$ the dissolved O₂
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25 concentration at the beginning of the test (DO = 0 mg L⁻¹ at $t = 0$ s, since the cultivation
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27 broth was initially degassed with argon). In addition, a similar test series as above described
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29 was conducted with MSM supplemented with 0.03 M Na₂SO₃.
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36 The OTR (g L⁻¹ h⁻¹) from the gas phase to the MSM was estimated (in duplicate) by
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38 periodically monitoring the time course of SO₃²⁻ oxidation in the fermenter according to
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40 Quijano et al. (2009). The bioreactor was initially filled with 2.5 L of MSM containing
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42 sodium sulfite at 0.03 M and aerated for 30 minutes prior to the test. The kinetics of O₂
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44 absorption were recorded after the addition of the catalyst (5 mL of a 2.5×10⁻⁴ M CoSO₄
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46 solution, which supports an instantaneous SO₃²⁻ oxidation) until complete SO₃²⁻ depletion
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48 according to equation 2.
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4 Aqueous samples of 5 mL were drawn from the bulk phase of the fermenter and SO_3^{-2}
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6 concentration was determined by iodometric back-titration according to Zhao et al. (1999).
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9 The $k_l a_{O_2}$ was determined using equation 3:

$$k_l a_{O_2} = \frac{OTR}{\frac{C_{O_2,G}}{H_{O_2}} - C_{O_2,L}} \quad (\text{Eq. 3})$$

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16 Where OTR is calculated from the slope of the absorbed O₂ concentration (estimated from
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18 the SO₃⁻² concentration multiplied by the O₂/SO₃⁻² stoichiometric coefficient derived
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20 from equation 3) versus time plot, H_{O_2} stands for the dimensionless Henry law constant of
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22 O₂ (30.9 at 30 °C) and $C_{O_2,G}$ represents the gas O₂ concentration in the exhaust gas (Zhao et
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24 al. 1999). Finally, the volumetric mass transfer coefficient of CH₄ was estimated from $k_l a_{O_2}$
25
26 according to Yu et al. (2010) using the correlation empirically validated for stirred tank
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28 reactors based on dissolved CH₄ measurements and a dynamic methodology to characterize
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30 $k_l a_{O_2}$ (equation 4):
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$$k_l a_{CH_4} = \frac{k_l a_{O_2}}{1,16} \quad (\text{Eq. 4})$$

35 36 37 38 39 40 41 **2.3 Determination of the influence of the stirring rate on CH₄ elimination capacity**

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43 The 3.5 L *ez*-control fermenter equipped with two Rushton turbines was initially filled with
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45 2.5 L of MSM and inoculated with the bacterial consortium. The bioreactor was
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47 continuously supplied with a CH₄-laden air emission (0.42 L min⁻¹) at a CH₄ concentration
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49 of 26.9 ± 1.7 g m⁻³ and Cu-supplemented MSM (50 µg L⁻¹ of CuSO₄) at a dilution rate of
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51 0.12 d⁻¹. The temperature and pH of the cultivation broth were automatically controlled at
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53 30 °C and 7, respectively. The system was initially operated for 60 days at 800 rpm to reach
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55 a steady state biomass concentration of 2.5 ± 0.4 g L⁻¹ in order to prevent process limitation
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4 by microbial activity. The influence of the stirring rate (800, 600 and 400 rpm) on the CH₄
5 elimination capacity (EC) was evaluated by monitoring the inlet and outlet CH₄ gas
6 concentration (C_{CH₄,in} and C_{CH₄,out}, respectively) for at least 7 consecutive days at each
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9 stirring rate. The EC was calculated according to equation 5:
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$$EC = \frac{C_{CH_4,in} - C_{CH_4,out}}{C_{CH_4,in}} \quad (\text{Eq. 5})$$

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18 A control test was initially conducted for 2 days in the absence of biomass in order identify
19 any potential abiotic removal of CH₄.
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25 **2.4. Analytical methods**

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27 CH₄ gas concentration was measured in a GOW MAC series 550 GC-TCD (USA)
28 equipped with a CTR1 packed column (Alltech, USA) using helium as a carrier gas at 60
29 ml min⁻¹. The injector, oven and detector temperatures were maintained at 50, 40, 115 °C,
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A polarographic sensor (Applikon, The Netherlands) and a luminescent/optical dissolved O₂
sensor (Intellical™ LDO101, Hach, USA) were used for the measurement of the dissolved
O₂ concentration in the cultivation broth during the gassing-out experiments.

The determination of SO₃²⁻ concentration involved the careful sampling of 5 mL of
cultivation broth to minimize their contact with ambient air. The sample was then
transferred to a 50-ml Erlenmeyer containing 1 ml of sulfuric acid and 0.1 g of NH₂SO₃H
crystals. After 1 ml of a starch solution (1%) was added as indicator, the sample was

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4 titrated with standard KI-KIO₃ titrant until a faint permanent blue color developed.
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6 Similarly, a sample of distilled water was processed as the blank control. The SO₃²⁻
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8 concentration was then calculated according to the following equation 6:
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$$[SO_3^{2-}][M] = \frac{[\textit{Titrant Volume for Sample (ml)} - \textit{Titrant volume for Blank (ml)}]}{\textit{Volumen Sample (ml)}} \quad (\textit{Eq. 6})$$

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20 Biomass concentration was determined by filtering 10 ml of cultivation broth in dry pre-
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22 weighted 0.2 µm acetate cellulose filters, which were then allowed to dry overnight and
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24 weighted again.
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29 **2.5 Statistical Treatment**

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32 All experiments were performed in duplicate and each assay was considered as a replicate.
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34 The results are given as the average value ± standard deviation from the duplicate assays. In
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36 addition, the results were analysed using a one-way ANOVA with significance at $p \leq 0.05$.
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38 The Excel statistical package (Microsoft Corporation, USA) was used for data treatment.
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44 **3. Results and Discussion**

45 **3.1 Volumetric mass transfer coefficient determination**

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49 The volumetric mass transfer coefficients for CH₄ increased when increasing the stirring
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51 rate regardless of the methodology used to estimate $k_{l}a_{O_2}$ as a result of the increased
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53 turbulence and air bubble disruption caused by agitation. However, while $k_{l}a_{CH_4}$ increased
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55 from 13.7±2.0 h⁻¹ at 400 rpm to 52.6±1.7 h⁻¹ at 800 rpm when the gassing-out method was
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57 used to characterize CH₄ mass transfer, the OTR methodology resulted in an increase from
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4 44.1±1.1 h⁻¹ at 400 rpm to 96.7±3.2 h⁻¹ at 800 rpm (Figure 1). The type of O₂ sensor
5
6 (polarographic vs optical) used in the gassing-out methodology did not significantly ($p \leq$
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8 0.05) influence the measurement of $k_{l}a_{O_2}$ (average error < 2.75 %), which confirmed that
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10 the significant ($p \leq 0.05$) differences observed in $k_{l}a_{CH_4}$ between the OTR and gassing-out
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12 methodologies were not caused by the delay in the response of the polarographic sensor. In
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14 addition, the increase in MSM salinity as a result of the addition of 0.03 M Na₂SO₃ (=3.78
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16 g L⁻¹) in the OTR methodology could only explain an average 9.5 % increase in $k_{l}a_{CH_4}$ due
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18 to the reduction in bubble coalescence mediated by the slightly higher ionic strength of the
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20 liquid medium (Figure S1). Indeed, this methodology was optimized by Quijano et al.
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22 (2009) in 2009 to prevent a salinity effect of the reagent Na₂SO₃ on the determination of
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24 $k_{l}a_{O_2}$. At this point it should be highlighted that moderate salinity media, despite slightly
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26 reducing the solubility of gases, typically result in enhanced gas-liquid $k_{l}a$ induced by the
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28 higher gas-liquid interfacial areas (Rodero et al., 2018). Therefore, the difference in the
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30 volumetric mass transfer coefficient of O₂, which directly impacted on the estimation of
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32 $k_{l}a_{CH_4}$, was attributed to a mass transfer acceleration mediated by the instantaneous
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34 chemical oxidation of O₂ by SO₃²⁻ (Eq. 2). The extremely fast (almost instantaneous)
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36 kinetics of this reaction (faster than any gas-liquid diffusional mechanisms in the mass
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38 transfer process) in the presence of Co²⁺ likely accelerated O₂ uptake from the gas phase in
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40 the fermenter (Zhao et al., 1999), which rendered the classical two-films model not valid
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42 since the O₂ concentration gradients in the liquid phase were no longer linear under this
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44 particular scenario (Van Krevelen et al., 1948). Figure 2, which depicts the O₂ profiles in
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46 the liquid phase during the gassing-out and OTR experiments, clearly shows the higher
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48 mass transfer gradients (estimated locally from the slope of the decrease in dissolved O₂
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4 concentration at the gas-liquid interphase) despite the overall difference in O₂ concentration
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6 between the interphase and the bulk liquid (200 mg L⁻¹ – 0 mg L⁻¹) remained similar. These
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8 higher gradients mediated by the rapid oxidation of O₂ in the presence of SO₃⁻² and Co²⁺
9
10 entailed a superior O₂ mass transfer, which mathematically resulted into higher *kla*_{O₂}
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12 (obtained from equations 1 and 3). Finally, it should be highlighted that the *kla*_{CH₄} here
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14 recorded were similar to those reported in vertical tubular loop (70 h⁻¹) and airlift external
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16 loop reactors (85 h⁻¹), but higher than those observed in stirred tank, airlift bubble column
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18 and U-loop reactors (12-26 h⁻¹) (Stone et al., 2017). At this point it should be also stressed
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20 that high *kla*_{CH₄} values are required in CH₄ treating bioreactors in order to compensate for
21
22 the low concentration differences imposed by the high Henry law constant of CH₄ (~ 30 at
23
24 25 °C) (Lopez et al., 2013). Thus, the higher the *kla*_{CH₄} values, the lower the bioreactor
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26 volumes and consequently, the lower the investment costs.
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41 **3.2 CH₄ abatement under continuous operation: methodology validation**

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43 The abiotic test carried out in the fermenter under continuous gas phase supply revealed a
44
45 negligible contribution of adsorption or photolysis to CH₄ removal in this particular
46
47 experimental set-up (Figure S2). This implies that the removal of CH₄ recorded at 400, 600
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49 and 800 rpm can be attributed to the biocatalytic action of the microbial consortium present
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51 in the fermenter, which converts CH₄ to CO₂, H₂O and new biomass (Lopez et al., 2014).
52
53 On the other hand, the empirical CH₄ elimination capacities recorded in the biotic fermenter
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55 inoculated with the consortium at 400, 600 and 800 rpm accounted for 37.8 ± 5.8, 42.5 ±
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57 5.4 and 62.3 ± 5.2 g CH₄ m⁻³ h⁻¹. These values were similar to the CH₄-ECs estimated using
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4 the OTR $k_1a_{CH_4}$ at a maximum concentration gradient (equation 7): 35.6 ± 2.2 , 50.1 ± 2.3
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6 and 59.6 ± 3.4 g CH₄ m⁻³ h⁻¹ at 400, 600 and 800 rpm, respectively (Figure 3).
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$$CH_4 - EC = k_1a_{CH_4} \times \left(\frac{C_{CH_4,in}}{H_{CH_4}} - C_{L,CH_4} \right) \quad (\text{Eq. 7})$$

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17 Where H_{CH_4} represents the dimensionless Henry law constant of CH₄ (28.7 at 30 °C) and
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19 C_{L,CH_4} the aqueous concentration of CH₄ in the bulk phase, which is negligible under the
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21 mass transfer limiting scenario here encountered. In fact, no significant differences ($p \leq$
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23 0.05) between the empirical and estimated CH₄-ECs were found at 400 and 800 rpm. The
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25 assumption of process operation at a maximum CH₄ concentration difference [$C_{CH_4,in}/H_{CH_4} -$
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27 C_{L,CH_4}] (which implies a negligible CH₄ concentration in the cultivation broth) was
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29 validated by the fact that the process was mass transfer limited during CH₄ biodegradation
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31 at a biomass concentration of 2.5 ± 0.4 g dry weight L⁻¹ (as confirmed by the higher CH₄-
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33 ECs at higher stirring rates). Indeed, MSM supplementation with Cu at 50 µg L⁻¹ was
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35 required to overcome process limitation by microbial activity, which increased CH₄-ECs
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37 from 24.4 ± 1.7 to 62.3 ± 5.2 g CH₄ m⁻³ h⁻¹ at 800 rpm. In this context, Cu positively
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39 regulates the activity of the enzymes pMMO and sMMO (particulate and soluble CH₄
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41 monooxygenase, respectively) and controls the expression of their genes. Most
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43 methanotrophs grow optimally at Cu concentrations lower than 270 µg L⁻¹, though previous
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45 enzymatic assays have demonstrated that sMMO in type II methanotrophs is properly
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47 synthesized at low Cu concentrations (below 50 µg L⁻¹) (Graham et al., 1993; Bender and
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49 Conrad, 1995; Semrau et al., 2010). Overall, the results here obtained confirmed that the
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51 classical gassing-out methodology for the estimation of $k_1a_{CH_4}$ clearly underestimated CH₄
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4 mass transfer under continuous biotic operation of the fermenter. Thus, the rapid CH₄
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6 biodegradation in the vicinity of the gas-liquid interphase mediated by methanotrophic
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8 bacteria likely supported an enhanced CH₄ mass transfer (induced by the higher local
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10 gradient close to the gas-liquid interphase), similarly to the acceleration in the O₂ mass
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12 transfer recorded in the OTR experiments.
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15 16 17 18 <Figure 3> 19

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21 The CH₄-ECs here recorded (38-62 g m⁻³ h⁻¹) were significantly higher than those reported
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23 in biotrickling filter or two-phase airlift bioreactors operated with internal gas recirculation
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25 at similar inlet loading rates (CH₄-ECs ~ 20-25 g m⁻³ h⁻¹) (Quijano et al., 2009; Van
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27 Krevelen et al., 1948). Comparable CH₄-ECs of 35-54 g m⁻³ h⁻¹ were recently achieved in a
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29 bubble column bioreactor operated with internal gas recirculation and *Methylocystis hirsuta*
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31 as a model CH₄ degrading microorganisms (Garcia-Perez et al., 2018).
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39 **3.3 Practical applications and future research perspectives**

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41 A renewed attention on suspended growth bioreactors for CH₄ abatement (i.e. stirred tank,
42
43 airlift and bubble columns) has emerged in the past 5 years due to the popularization of
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45 biorefineries based on the feasible bioconversion of CH₄ into added-value bioproducts such
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47 as biopolymers, ectoine, exopolysaccharides, protein, etc. (Cantera et al., 2017;
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49 Karthikeyan et al., 2015; Khmelenina et al., 2015; Pieja et al., 2012). These novel
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51 biorefineries, which will be based both on dilute CH₄ emissions from abandoned landfills
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53 or coal mines and biogas from solid waste management, will create value out of GHG
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55 mitigation and envisage a new bioeconomy based on CH₄. In this context, suspended
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4 growth systems represent nowadays the most popular bioreactor platform capable of
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6 supporting a cost-competitive methanotrophic biomass harvesting and bioproduct
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8 downstream, although recent attempts to produce added-value products from residual CH₄
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10 in immobilized bioreactors have been reported (Cantera et al., 2016; Karthikeyan et al.,
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12 2017). Thus, the characterization of CH₄ mass transport from residual emissions or biogas
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14 streams into the microbial community is central to dimension the bioreactor and
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16 downstream processes. However, the absence of reliable commercial dissolved CH₄ sensors
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18 has hindered the development of accurate methodologies for the determination of $k_{lA_{CH_4}}$,
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20 whose characterization has been conducted based on either classical $k_{lA_{O_2}}$ determination
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22 methodologies without a further validation for CH₄ or complex experimental
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24 methodologies difficult to implement at large scale (Yu et al., 2010; Rocha-Rios et al.,
25
26 2011). Therefore, the results herein obtained validated the SO₃²⁻ oxidation methodologies
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28 as the most accurate method to estimate $k_{lA_{CH_4}}$, which is expected to support in the design
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30 of cost-effective CH₄ biorefineries. For instance, figure S3 depicts a simple model
31
32 simulation of the CH₄ outlet concentration as a function of the CH₄ inlet concentration and
33
34 the $k_{lA_{CH_4}}$ estimated at the three stirring rates tested. This methodology entails the purchase
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36 of SO₃²⁻ and Co²⁺, and will result in SO₄²⁻ as a final product in the cultivation medium after
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38 the experiment. Further research should focus on validating this methodology in high-
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40 performance off gas treatment bioreactors such as Taylor flow or two-phase partitioning
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42 bioreactors, which are foreseen as the only bioreactor configuration capable of supporting
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44 an effective CH₄ bioconversion at a reduced energy consumption (Cantera et al., 2017). In
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46 addition, this methodology could be also applied to the design of algal-bacterial
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48 photobioreactors devoted to biogas upgrading (CO₂ and H₂S removal), where both
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4 chemically and biologically enhanced mass transport has been identified (Rodero et al.,
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7 2018, Yang et al., 2016a,b, Zhu, 2015, Zhu et al., 2016).
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10 11 **4. Conclusions**

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14 This work empirically validated a simple methodology to accurately describe CH₄ mass
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16 transport and biodegradation in continuous stirred tank bioreactors based on OTR
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18 estimations. The rapid O₂ uptake at the gas-liquid interphase mediated by SO₃⁻² and Co²⁺,
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20 mimicking microbial CH₄ uptake during continuous CH₄ treatment, supported the superior
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22 performance of the OTR methodology for the description of $k_{l}a_{CH_4}$ compared to the
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24 classical gassing-out methodology. Thus, the empirical CH₄-ECs of 37.8±5.8, 42.5±5.4 and
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26 62.3±5.2 g CH₄ m⁻³ h⁻¹ recorded at 400, 600 and 800 rpm, respectively, were comparable to
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28 the CH₄-ECs of 35.6±2.2, 50.1±2.3 and 59.6±3.4 g CH₄ m⁻³ h⁻¹ predicted by the OTR
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30 methodology. This finding will help in the design of suspended growth bioreactors devoted
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32 to CH₄ bioconversion.
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Figure captions

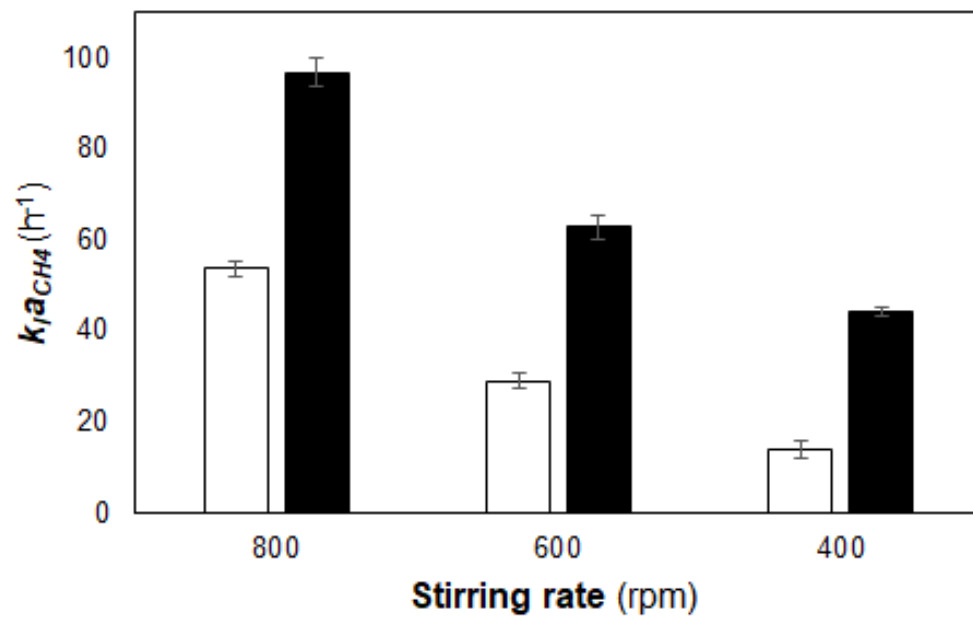
Figure 1. Influence of the stirring rate on the volumetric mass transfer coefficient of CH₄ estimated based on the gassing-out (□) and OTR (■) methods for O₂.

Figure 2. O₂ concentration profiles in the liquid phase during the gassing-out (continuous line) and OTR (dashed line) methods.

Figure 3. Influence of the stirring rate on the CH₄-EC estimated from the gassing-out (□) and OTR (■) methods, and experimentally (▣).

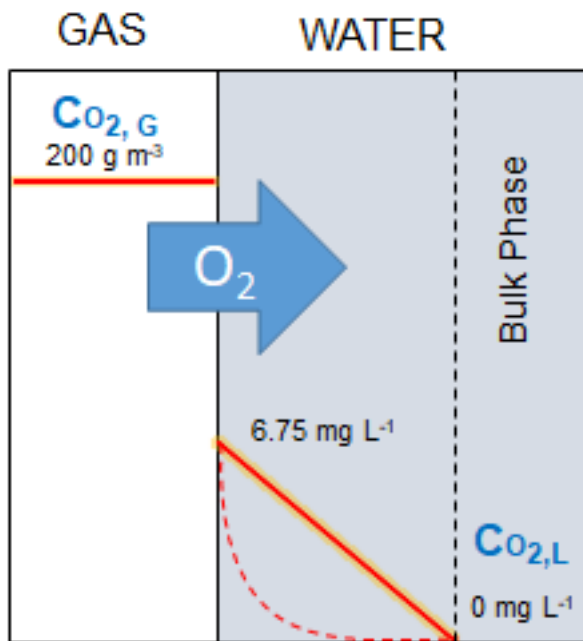
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Figure 1.



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Figure 2.



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Figure 3.

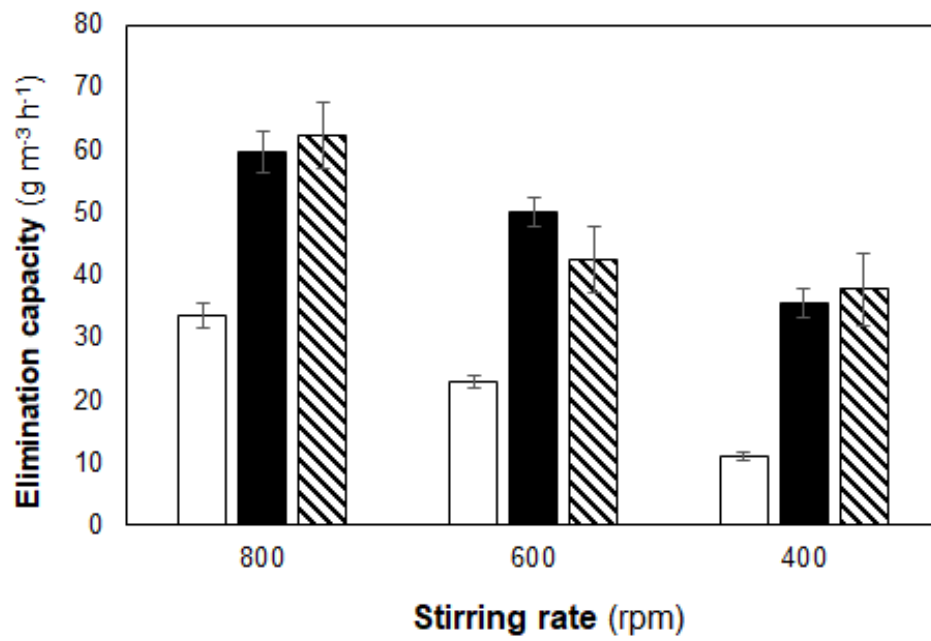


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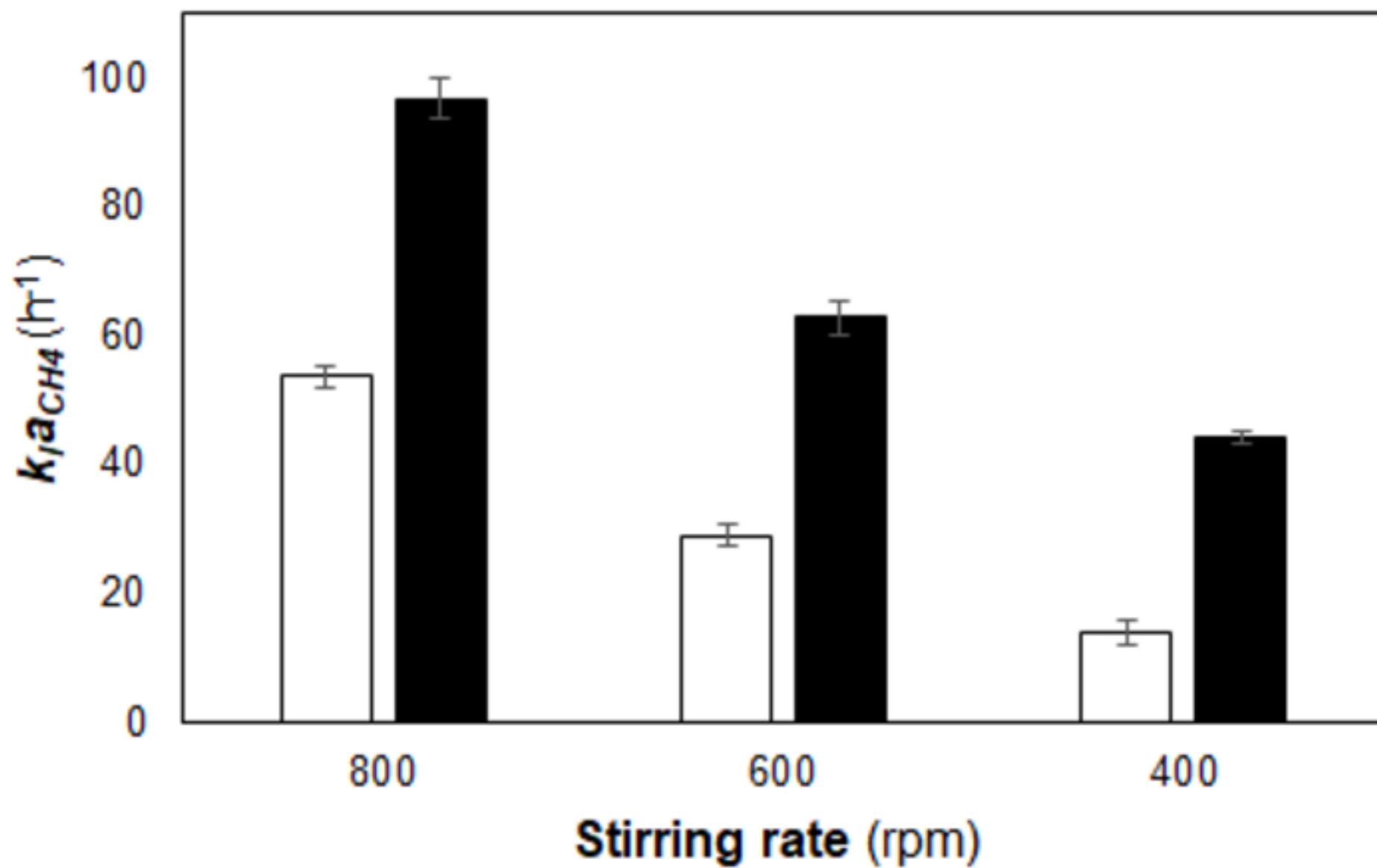


Figure 2
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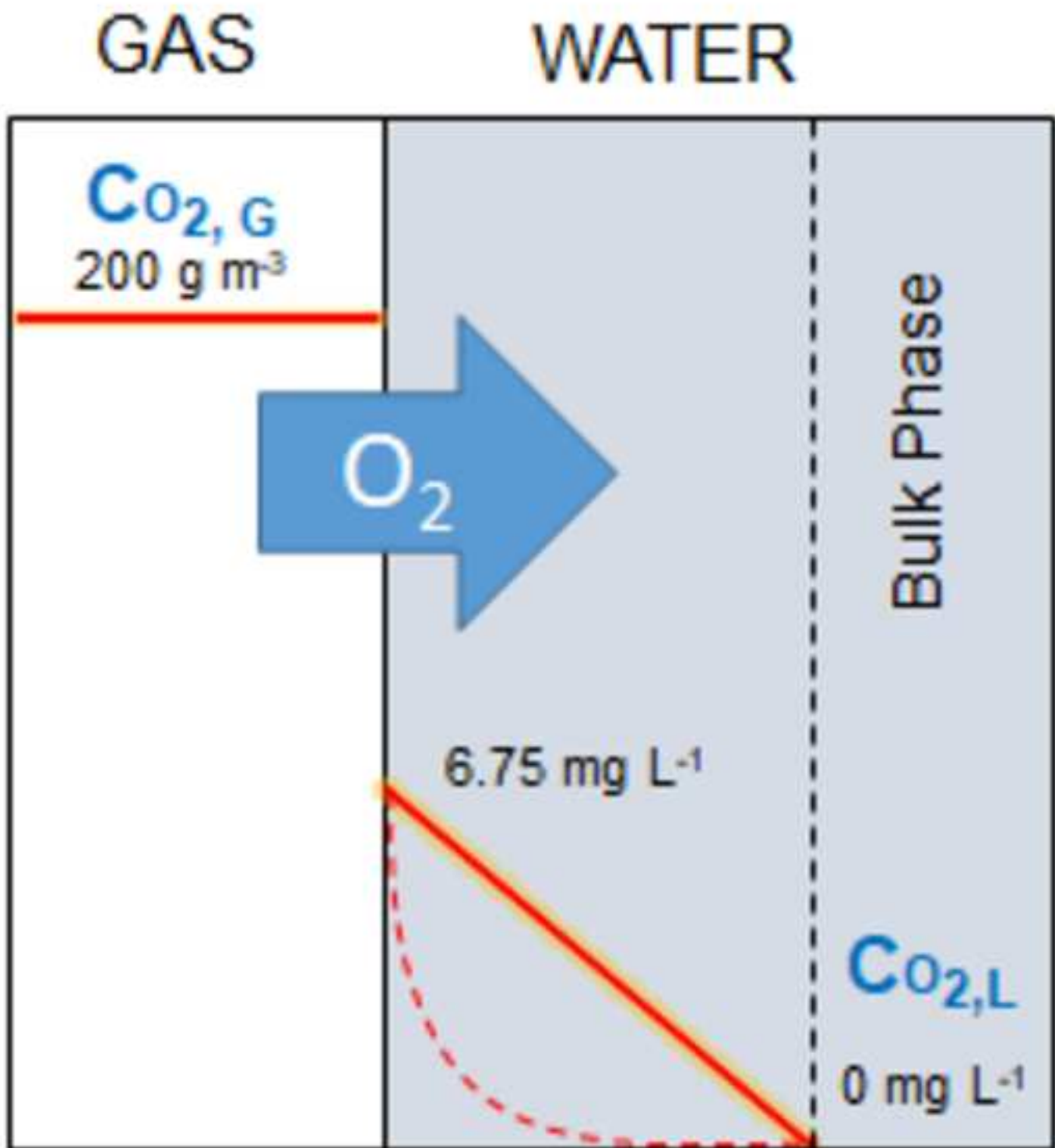
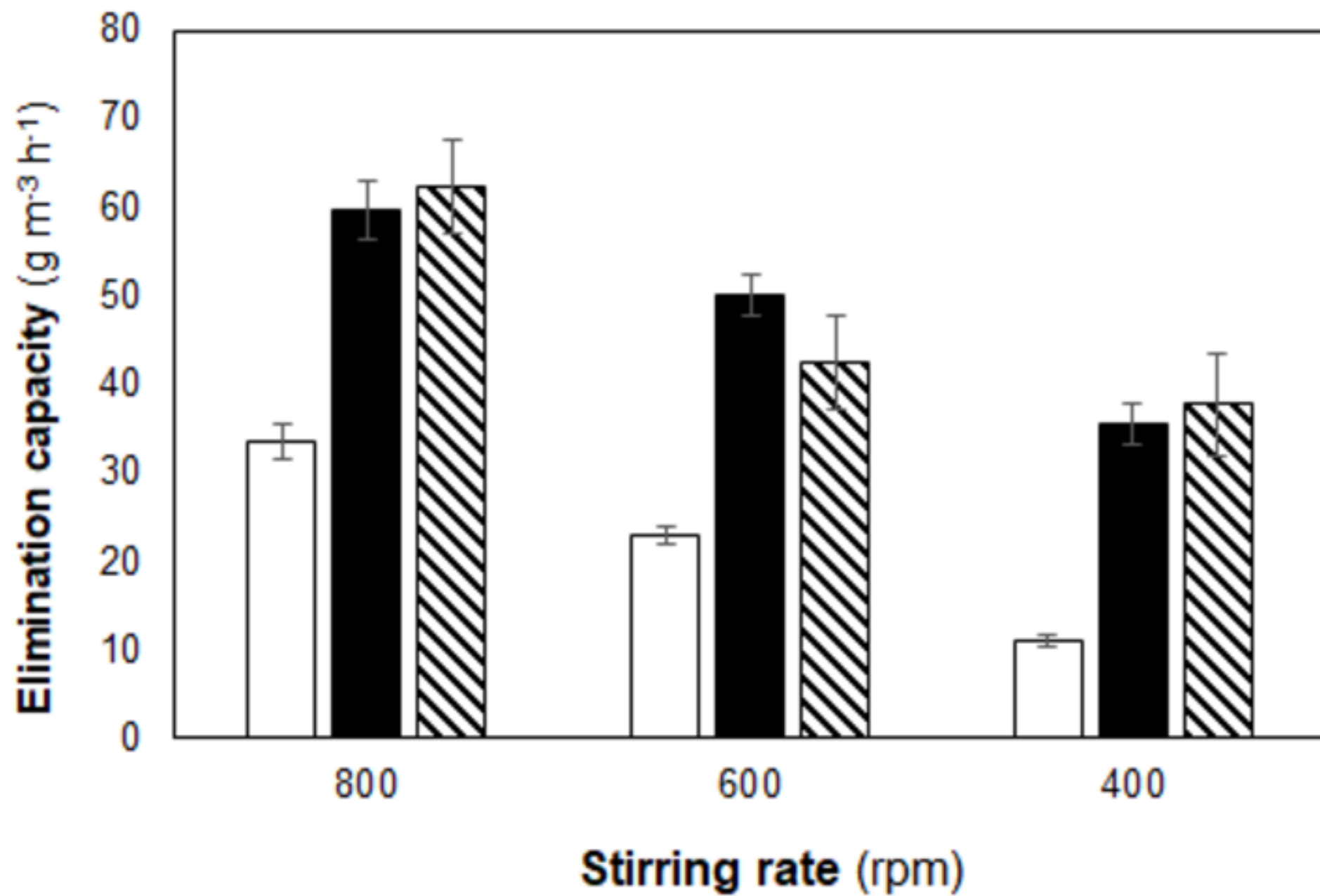


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