

This is a repository copy of *Phyllosphere of staple crops under pig manure fertilization, a reservoir of antibiotic resistance genes.*

White Rose Research Online URL for this paper: http://eprints.whiterose.ac.uk/146996/

Version: Accepted Version

Article:

Zhou, S., Zhu, D., Giles, M. et al. (4 more authors) (2019) Phyllosphere of staple crops under pig manure fertilization, a reservoir of antibiotic resistance genes. Environmental Pollution, 252 (Part A). pp. 227-235. ISSN 0269-7491

https://doi.org/10.1016/j.envpol.2019.05.098

Article available under the terms of the CC-BY-NC-ND licence (https://creativecommons.org/licenses/by-nc-nd/4.0/).

Reuse

This article is distributed under the terms of the Creative Commons Attribution-NonCommercial-NoDerivs (CC BY-NC-ND) licence. This licence only allows you to download this work and share it with others as long as you credit the authors, but you can't change the article in any way or use it commercially. More information and the full terms of the licence here: https://creativecommons.org/licenses/

Takedown

If you consider content in White Rose Research Online to be in breach of UK law, please notify us by emailing eprints@whiterose.ac.uk including the URL of the record and the reason for the withdrawal request.



eprints@whiterose.ac.uk https://eprints.whiterose.ac.uk/

1 Phyllosphere of staple crops under pig manure fertilization, a

2 reservoir of antibiotic resistance genes

| 3 Shuyidan Zhou, †, ‡, # Dong Zhu, †, ‡, # Madeline Giles§, Xiao-Ru | | Shuyidan Zhou, | ,,,, # | Dong Z | ∠nu,⊺,⊥, | # Madeline | Gllesg, A | ао-ки | rang, | Tim |
|---|--|----------------|--------|--------|----------|------------|-----------|-------|-------|-----|
|---|--|----------------|--------|--------|----------|------------|-----------|-------|-------|-----|

- 4 Daniell∥,§, Roy Neilson§, Yong-Guan Zhu,*, †,‡,⊥
- 5
- 6 † Key Laboratory of Urban Environment and Health, Institute of Urban Environment,
- 7 Chinese Academy of Sciences, 1799 Jimei Road, Xiamen 361021, China.
- 8 ‡ University of the Chinese Academy of Sciences, 19A Yuquan Road, Beijing 100049,

9 China.

- 10 § Ecological Sciences, The James Hutton Institute, Dundee, DD2 5DA, Scotland, UK
- 11 || Department of Animal and Plant Sciences, The University of Sheffield, Sheffield,
- 12 S10 2TN, UK
- 13 \perp State Key Laboratory of Urban and Regional Ecology, Research Center for Eco-
- 14 Environmental Sciences, Chinese Academy of Sciences, Beijing 100085, China.

- 16 # Shuyidan Zhou and Dong Zhu contributed equally to this paper
- 17 * Corresponding author: Yong-Guan Zhu
- 18 Key Laboratory of Urban Environment and Health, Institute of Urban Environment,
- 19 Chinese Academy of Sciences, 1799 Jimei Road, Xiamen 361021, China
- 20 E-mail address: ygzhu@iue.ac.cn (Y.G. Zhu)
- 21

22 Abstract

In China, the common use of antibiotics in agriculture is recognized as a potential public 23 health risk through the increasing use of livestock derived manure as a means of 24 fertilization. By doing so this may increase the transfer of antibiotic resistance genes 25 (ARGs) from animals, to soils and plants. In this study two staple crops (rice and wheat) 26 were investigated for ARG enrichment under differing fertilization regimes. Here, we 27 applied 4 treatments, no fertilizer, mineral fertilizer, clean (reduced antibiotic practice) 28 and dirty (current antibiotic practice) pig manure, to soil microcosms planted with either 29 rice or wheat, to investigate fertilization effects on the abundance of ARGs in the 30 respective phyllospheres. For both rice and wheat, samples were collected after two 31 separate fertilization periods. In total 162 unique ARGs and 5 mobile genetic elements 32 (MGEs) were detected from all rice and wheat samples. The addition of both clean and 33 dirty manure, enhanced ARGs abundance significantly when compared to no fertilizer 34 application (P < 0.001), though clean manure enriched ARGs to a lesser extent than dirty 35 manure, in all rice and wheat samples (P < 0.001). The classes of ARGs recorded were 36 different between crops and wheat samples had a higher ARG diversity than rice. These 37 results revealed that staple crops in China such as rice and wheat may be a reservoir for 38 ARGs when clean and dirty pig manure is used for fertilization. 39

40 Capsule: Pig manure can enrich the antibiotic resistance genes in the phyllsophere of
41 rice and wheat.

42 Key words: Phyllosphere, Pig manure, Antibiotic resistance genes, Staple crops,

45 Introduction

Organic fertilizers are increasingly being used as a cost-effective alternative to mineral 46 fertilizers, as their high organic matter and mineral nutrient content generates similar 47 48 improvement in crop productivity to that of mineral fertilizer (Hao et al., 2008; Cheng et al., 2013). However, organic fertilizers frequently include animal manure or activated 49 sludge, the use of which is an efficient way to reduce agricultural waste, but manure 50 has been shown to increase the abundance of ARGs in soils (Joy et al., 2013; Xie et al., 51 2016). Both manure and activated sludge can contain incompletely metabolized 52 antibiotics, antibiotic resistant bacteria and antibiotic resistant genes (ARGs). 53

The continued increase in the abundance of antibiotic resistance genes in the environment is a recognized global public health issue (Martinez, 2008). Although ARGs generally occur within the genetic material of the carrier organisms (D'Costa et al., 2011), the anthropogenic spread of antibiotics has led to an enrichment of ARGs in the environment (Pruden et al., 2006) with a concomitant concern that such ARGs can have an impact on human and animal health.

ARGs enter the food chain through ARG contaminated crops leading to a risk of antibiotic resistance related problems in humans (Gillings, 2017). For example, studies have shown that ARGs can enter the human food chain through chicken and pork products sold at local markets (Leverstein-van Hall et al., 2011; Zhang et al., 2018),

| 64 | fish products (Antunes et al., 2018) and vegetables (Marti et al., 2013; Rahube et al., |
|----|--|
| 65 | 2014; Wu et al., 2018). Antibiotics are used widely for medical purposes, and regularly |
| 66 | in agricultural production systems, consequently large amounts of antibiotics pass |
| 67 | through waste streams into the environment (Allen et al., 2010; Marshall and Levy, |
| 68 | 2011; Zhou et al., 2013). Antibiotics released into the environment have been shown to |
| 69 | impact the regulation of antibiotic resistance genes (ARGs) and trigger a microbial |
| 70 | response that increases the mutation rate in bacteria and the sharing, through horizontal |
| 71 | gene transfer, of ARGs from antibiotic resistant bacteria (ARB) to those without |
| 72 | antibiotic resistance (Zhu et al., 2013). Consequently, individual bacteria and the wider |
| 73 | microbial community adapt to the changing presence of antibiotics by increasing the |
| 74 | number of ARGs or selecting antibiotic resistant bacteria that match the soil antibiotic |
| 75 | profile in the wider environment (Tello et al., 2012; Gillings, 2017; Zhu et al., 2017). |
| 76 | The plant phyllosphere provides a large surface area for microorganisms to inhabit |
| 77 | (Woodward and Lomas, 2004). It has been estimated that the total surface area of leaves |
| 78 | on the planet approaches approximately twice that of the land surface and can contain |
| 79 | bacterial populations of up to 10^{26} cells (Lindow and Brandl, 2003; Vorholt, 2012). Thus, |
| 80 | the phyllosphere provides a potential reservoir for micro-organisms and ARGs and |
| 81 | allows them to come in direct contact as a result of aerial spreading of animal-derived |
| 82 | manures on crops. Senescence of deciduous leaves facilitates the transfer of |
| 83 | phyllosphere adhering microorganisms to soil through leaf drop, while the exposure of |
| 84 | the phyllosphere to the atmosphere allows it to be a receiver of wind-borne |
| 85 | microorganisms (Berlec, 2012; Mhuireach et al., 2016), highlighting the transfer of |

microorganisms across the critical zone i.e. the atmosphere, phyllosphere and soil. Thus,
ARGs can be potentially ubiquitous on the plant phyllosphere, for example, on the
phyllosphere of field grown vegetables (Zhu et al., 2016; Chen et al., 2017b). However,
it is noted that to date the phyllosphere has not been recognized as an explicit source of
ARGs in agricultural systems (Chen et al., 2018). Therefore, it is critical to establish a
knowledge base about ARGs in the plant phyllosphere especially in an agricultural
context.

Questions remain as to whether staple crops such as rice and wheat which are grown in 93 94 a continuous rotation in China can transfer ARGs to the environment. Rice and wheat have contrasting agronomic management (flooded or not-flooded) which may drive 95 differences in the structure of phyllosphere microbial communities and abundance of 96 97 ARGs. As these crops contribute a high proportion of the global human diet (Tilman et al., 2002) there is an imperative to identify their interactions, if any, with antibiotic 98 resistant bacteria (ARB) /ARGs. More than 25% of farmland in China is under rice 99 100 cultivation (Rahman et al., 2018). Current studies on rice focus on soil and hydrological systems and typically overlook the risk of phyllosphere ARGs. 101

The aims of this study were therefore to 1) characterize the abundance and diversity of ARGs in the phyllosphere of rice and wheat; 2) investigate the effects of manure application on phyllosphere ARG communities; 3) determine if there is a difference in ARG abundance between manure from farms with standard and reduced antibiotic use; 4) identify differences, if any, in ARG abundances between rice and wheat; 5) explore the contribution of the bacterial community to shifts, if any, in phyllosphere ARGs. We hypothesize that 1) rice and wheat phyllospheres would exhibit different ARG profiles;
2) the addition of manure from farms with reduced antibiotic use results in lower
phyllosphere ARG enrichment than the addition of manure from farms using standard
antibiotic use; and 3) Shifts in the community composition of the bacterial community
can be accounted for by changes in ARG composition.

113 2. Materials and methods

114 2.1 Soil and manure properties

Soil, a sandy loam, was collected to a depth of 20 cm from a farm under a rice and
wheat rotation in Ningbo, Zhejiang, China (29° 47′N, 121° 21′E). Soil characteristics
are listed in Table S2. Soil was sieved through a 5 mm sieve to remove stubble, roots
and stones. Mineral fertilizer treatments comprised of pure nitrogen fertilizer (Urea)
(21 g m⁻²), P₂O₅ (7.5 g m⁻²) and K₂O (12 g m⁻²) for rice, while urea (12.6 g m⁻²), P₂O₅
(2.4g m⁻²) and K₂O (10.0 g m⁻²) were applied before wheat planting, and then urea N
(5.4 g m⁻²) was applied during shoot elongation stage.

Two organic fertilizer treatments were applied, a "dirty" slurry comprising pig manure from a farm using standard antibiotic practice and a "clean" slurry comprising pig manure from a farm using reduced antibiotic practice. The properties of each slurry are listed in Table S3. The clean slurry was added 9.47 g/pot (N%: 3.36) and the dirty slurry was added 11.007 g/pot (N%: 2.89).

128 2.2 Experimental design and crop cultivation

Sixteen treatments were established: no fertilizer, mineral fertilizer, clean manure and dirty manure for each of two crops (rice and wheat) grown in rotation and two fertilization periods per crop rotation. For each fertilization period there were 5 replicates per treatment. Phyllosphere samples were collected twice for each crop rotation: at grain filling stage and before harvest. Eighty microcosms were established each containing 3.5 kg wet soil (water holding capacity for rice 100%, for wheat 65%), Microcosms were made of polyvinyl chloride with a diameter of 15 cm and height of

136 23 cm, with water drainage effected through a small hole in the bottom of pot.

Rice (cv. Yonguou 12) and wheat (cv. Yangmai 20) were germinated before planting in 137 microcosms using the following methods: rice seeds were field grown and transplanted 138 to the microcosms 28 days after sowing. Microcosms were flooded, seedlings planted, 139 and the soil held at 100% water holding capacity (WHC). At rice tilling, approximately 140 141 2 weeks before harvest, soil in the microcosms was dried through to harvest. Wheat seeds were sterilized in a 10 % hydrogen peroxide solution for 15 min and stored at 4°C 142 for 1 week prior to planting in the microcosms. Wheat was grown at 65 % WHC. 143 Yongyou 12 is a three-line indica-japonica hybrid super rice and Yangmai 20, a main 144 middle-early mature wheat variety. 145

146

147 2.3 DNA extraction from the phyllosphere of rice and wheat

148 DNA was extracted according to the method described in (Zhu et al., 2016). Prior to

149 DNA extraction, 5 g of either rice or wheat leaf was weighed into a 250 ml conical flask

containing 100 ml of 0.01 M, phosphate-buffered saline (pH=7.4), flasks were sonicated for 7 minutes before being shaken for 1 hour at 180 rpm, with samples held at 30 °C during this process. Phosphate buffer was initially filtered through a nylon gauze followed by filtration through a 0.22 μ M cellulose membrane. Target DNA was extracted from the filters using a FastDNA Spin Kit for Soil (MP Biomedicals, CA) and quality checked using spectrophotometer analysis (NanoDrop ND-1000, Thermo Scientific, Waltham, MA). DNA was stored at -20 °C prior to analysis.

157

158 2.4 Illumina Sequencing and bioinformatics analysis

The V4-V5 hypervariable region of the 16S rRNA gene was used to analyse the 159 structure of the phyllosphere bacterial community, using primers 515F: 160 GTGCCAGCMGCCGCGG and 907R: CCGTCAATCMTTTRAGTTT (Turner et al., 161 1999). To each PCR tube 1µL 10 µM 515F primer, 1µL 10 µM 907R primer, 0.81µL 162 bovine serum albumin (BSA, 20 mg mL⁻¹), 21.2 µL sterile water, 25µL TAKARA 163 Premix TaqTM (Ex TaqTM Version 2.0 plus dye, No. RR902A) and 1µL 20ng µL⁻¹ DNA 164 template was added. PCR conditions were 95°C for 5 minutes, followed by 165 amplification for: 25 cycles of 30s at 94°C, 35s at 58°C and 30s at 72°C. TIANGEN 166 universal DNA purification kits (TIANGEN biotech, Beijing, China) were used to clean 167 PCR products. Purified products were then normalized to 200 ng DNA before being 168 sequenced by Novogene (Beijing, China) using a Illumina Hiseq 2500 platform. The 169 Quantitative Insights into Microbial Ecology (QIIME) pipeline (Caporaso et al., 2010b; 170 Chen et al., 2017c) was used to produce OTUs, with 97 % OTU using UCLUST (Edgar, 171

172 2010). OTUs were aligned using one representative sequence per OTU, using PyNAST

aligner (Caporaso et al., 2010a). RDP classifier which uses the Greengenes data base

174 (version 13.8, 16S rRNA gene database) (McDonald et al., 2012; Langille et al., 2013),

175 was used to assigned taxonomic identity and relative abundance of OTUs. Alpha

176 diversity of samples was calculated using QIIME.

177

178 2.5 HT-qPCR quantification of Antibiotic resistant genes.

We investigated the diversity and abundance of ARGs using High-throughput qPCR 179 with a Wafergen SmartChip Real-time PCR system (Wafergen, Fremont, CA). All 180 samples were diluted to 20 ng μ L⁻¹ and run in triplicate. 285 ARG genes were targeted 181 using 296 primer sets (Su et al., 2015) (Table S1). Additionally these primer sets can 182 183 also detect 16S rRNA gene, 8 transposases, class 1 integron-integrase gene (*intl1*) (Stokes et al., 2006) and the clinical related class 1 integron-integrase gene (cintl1) 184 (Gillings et al., 2014). A total of 5184 Nano-wells on the Smart Chip platform provided 185 186 reaction sites for PCR. A 100nl PCR reaction mix was used per well, this mix contained 1 × LightCycler 480 SYBR Green I Master, 1mg mL⁻¹ bovine serum albumin, nuclease-187 free PCR-grade water, 500 nM of each 296 primers and 20 ng μ L⁻¹ of each DNA sample. 188 PCR conditions were: enzyme activation at 95°C for 10 minutes followed by 40 cycles 189 of amplification: denaturation at 95°C for 30s, annealing at 60°C for 30s (Chen et al., 190 2017). An amplicon range of 1.8 to 2.2 was set as the efficiency range, with anything 191 outside this range discarded, including those with multiple melting peaks. qPCR results 192 were analyzed using Smartchip qPCR software (Wafergen, Fremont, CA). The 193

| 194 | detection limit was set at a threshold cycle of 31 (C_T) and amplification only considered |
|---|--|
| 195 | as positive if all three technical replicates showed a positive result. Relative gene copy |
| 196 | number and a normalized gene copy per bacteria were calculated by the following |
| 197 | equations (Chen et al., 2016): |
| 198 | |
| 199 | Relative gene copy number = $10^{(31-CT)/(10/3)}$ |
| 200 | Normalized ARG copy number |
| 201 | = (Relative ARG copy number / Relative 16S rRNA gene copy number) × 4.1 |
| 202 | |
| 203 | 4.1 was considered to be the average number of 16S rRNA gene copies per bacterium |
| 204 | based on the Ribosomal RNA Operon Copy Number Database (Klappenbach et al., |
| | |
| 205 | 2001). |
| 205 206 | 2001). |
| | 2001).2.6 Statistical analysis |
| 206 | |
| 206 207 | 2.6 Statistical analysis |
| 206 207 208 | 2.6 Statistical analysisStatistical analysis was performed in the R environment. The R package, Vegan |
| 206 207 208 209 | 2.6 Statistical analysisStatistical analysis was performed in the R environment. The R package, Vegan(Oksanen et al., 2018) was used for PERMANOVA (Adonis), Principal Coordinate |
| 206 207 208 209 210 | 2.6 Statistical analysisStatistical analysis was performed in the R environment. The R package, Vegan(Oksanen et al., 2018) was used for PERMANOVA (Adonis), Principal CoordinateAnalysis (PCoA), Canonical Correlation Analysis (CCA), Redundancy Analysis |
| 206 207 208 209 210 211 | 2.6 Statistical analysis Statistical analysis was performed in the R environment. The R package, Vegan (Oksanen et al., 2018) was used for PERMANOVA (Adonis), Principal Coordinate Analysis (PCoA), Canonical Correlation Analysis (CCA), Redundancy Analysis (RDA), Variation partitioning analysis (VPA), Inverse Simpson index and Shannon |
| 206 207 208 209 210 211 212 | 2.6 Statistical analysis Statistical analysis was performed in the R environment. The R package, Vegan (Oksanen et al., 2018) was used for PERMANOVA (Adonis), Principal Coordinate Analysis (PCoA), Canonical Correlation Analysis (CCA), Redundancy Analysis (RDA), Variation partitioning analysis (VPA), Inverse Simpson index and Shannon evenness score. The protocol for choosing either CCA or RDA was based on the axis |

| 216 | Graphs were constructed using ggplot 2 3.1 (Wickham et al., 2018) and heatmaps were |
|-----|---|
| 217 | built by using the R heatmap package (Galili et al., 2018) Pearson correlation |
| 218 | coefficient analysis and analysis of variance (ANOVA) was conducted using SPSS 21. |
| 219 | Pie charts and bar charts were produced in Excel 2016 and originlab 2018, respectively. |
| 220 | Bar charts show the mean value of 5 replicates with standard errors (SE) calculated in |
| 221 | Excel 2016. PD whole tree, observed species, chao1 and Shannon analysis of the |
| 222 | bacterial diversity among rice and wheat samples was calculated using QIIME and |
| 223 | visualized with orginlab 2018. |

225 **3. Results**

3.1 Diversity and abundance of ARGs in the phyllosphere of rice and wheat.

Across all rice and wheat leaf samples a total of 162 unique ARGs and 5 mobile genetic 227 elements MGEs were detected. In each sample the total number of ARGs detected 228 ranged from 32 to 105 (Figure 1A), while on average there were 3 MGEs per sample, 229 230 these included major antibiotic resistance classes. Four types of antibiotic resistant mechanisms were detected: antibiotic deactivate (42%), cellular protection (17%), 231 efflux pump (34%) and other unknown mechanisms (7%) (Figure 1B). 232 The normalized abundance of ARGs in rice and wheat samples ranged from 0.0044 to 233 0.047 and 0.03 to 0.29 copies per cell, respectively (Figures 2A, B). ARG composition 234 for both rice and wheat varied between, organic and mineral fertilizer treatments along 235 236 the x axis which accounted for 37.76% and 29.26% of the variation for rice and wheat samples, respectively. (Figures 2C, D). 237 The composition of ARGs varied significantly between the phyllospheres of rice and 238 239 wheat (P < 0.001, PERMANOVA). Aminoglycoside, MGEs, multidrug and others resistance genes were increased in samples (P < 0.05, ANOVA) from the first rice 240 fertilization, compared to the second rice fertilization. While the abundance of Beta-241 242 Lactamase and multidrug resistant dominated in the second fertilization of rice (Figure S1A). In wheat, MGEs (P <0.05, ANOVA) and Beta Lactamase (P <0.001, ANOVA) 243 were significantly raised in wheat samples from the first fertilization. Aminoglycoside 244 245 abundance also increased between first and second fertilizations of wheat (Figure S1B) (*P* <0.001, ANOVA). 246

Inverse Simpson and Shannon indices for ARGs of first fertilization samples of rice were higher than those from the second fertilization (Figure 2, P < 0.001), while the diversity from the wheat phyllosphere was higher after the second fertilization than the first (Figure S1, P < 0.001). Overall, wheat phyllosphere samples had a greater diversity of ARGs than the rice phyllosphere (Figure S2).

According to heatmap analysis (Figure 3), the total abundance of ARGs in rice samples were lower than in wheat. However, the abundance of ARGs in rice after the first fertilization (Rice no fertilization 1 - RCF, Rice mineral fertilization 1 - RMF, Rice clean manure fertilization 1 - RCM and Rice dirty manure fertilization 1 - RDM) was higher than those after the second fertilization (Rice no fertilization 2 - RCF2, Rice mineral fertilization 2 - RMF2, Rice clean manure fertilization 2 - RCM2 and Rice dirty manure fertilization 2 - RDM2) (P < 0.001, PERMANOVA).

259 3.2 The effect of antibiotic content of manures on phyllosphere ARGs abundance

Both clean and dirty manure significantly increased the abundance of ARGs and MGEs 260 261 in the phylloshere of both wheat and rice (Figure 2A, B; P < 0.01; RCM, P < 0.05) when compared to mineral fertilizer and control treatments.. The abundance of ARGs 262 under clean manure application was lower than the abundance under dirty manure 263 application. In rice samples, Tetracycline (P < 0.001), Aminoglycoside (P < 0.001), Beta 264 Lactamase (P =0.001), MGEs (P <0.001), MLSB (P <0.05), Multidrug (P <0.05) and 265 Vancomycin (P < 0.05) were enriched in the dirty manure compared to the clean manure 266 treatment (Figure S1). Whereas, Tetracycline (P < 0.001), Beta Lactamase (P < 0.001), 267 Aminoglycoside (P < 0.05) and MLSB (P < 0.05) were more abundant in wheat samples 268 13

treated with dirty than clean manure (Figure S1). Specifically, the abundance of ARGs such as, *aadA*, *aadA1*, *ampC*, *ttgA*, *tnpA*, *blaSHV*, *fosX* and *qacEdelta1* were increased in rice phyllosphere samples when dirty manure was applied (P < 0.001, ANOVA). In the wheat phyllosphere *tnpA*, *acrR*, *aadA2*, *aadA5*, *ampC*, *cmlA1*, *vanSB*, *blaOXY and acrA* were enriched in both clean and dirty manure treatments (P < 0.05, ANOVA) (Figure 3).

275

276 3.3 Correlation between phyllosphere ARGs and bacterial communities.

2,891,103 high-quality sequences were detected from 80 samples and a total of 19,013 277 operational taxonomic units (OTUs) obtained. Proteobacteria (70.6%), Firmicutes 278 (13.5%), Actinobacteria (0.8%) and unassigned OTUs (11.6%) were the major phyla in 279 280 both rice and wheat phyllosphere samples. PD whole tree, observed species, chao1 and Shannon analysis of the bacterial communities demonstrated that the diversity of the 281 second wheat fertilization samples was the highest amongst all four groups with the 282 283 lowest associated with the second rice fertilization (Figure S3). Among the Proteobacteria, alpha and gamma Proteobacteria were the most abundant classes, 284 ranging in relative abundance from 0% to 38.9% and 0.4% to 99.9%, respectively. 285 286 A correlation existed between ARGs and the composition of bacterial communities based on the Bray-Curtis dissimilarity metrics by Procrustes analysis for rice 287 (Procrustes sum of squares $M^2=0.78$, r=0.17, P < 0.001) and wheat (Procrustes sum of 288 squares M²=0.87, r=0.34, P < 0.001) samples (Figure S4). 289

290 PCoA analysis showed differences in the composition of bacterial communities

between rice and wheat phyllospheres (Figure 4). Samples from the first and second
fertilization of rice clustered together while wheat samples from the two fertilization
periods separated along the secondary coordinate, which accounted for 19.7% of
variation (Figure 4).

295

Canonical correlation analysis (CCA) (Figure 5A) (First axis length 4.3481) was 296 conducted for rice samples, and highlighted that the dominant families (>1%) 297 (Enterobacteriaceae, Bacillaceae, Pseudomonadaceae, Rhizobiaceae, Moraxellaceae) 298 299 and MGEs were the main drivers. ARGs in the dirty manure treatments from rice for the second fertilization was positively correlated with the abundance of MGEs (P 300 =0.001, R^2 0.68) and Pseudomonadaceae respectively (P =0.016, R^2 0.26). 301 302 Redundancy analysis (RDA) (First four axis length shorter than 3) was carried out for wheat samples (Figure 5B), where ARGs in both manure (clean and dirty) treatments 303 at both first and second fertilizations were positively correlated with MGEs (P = 0.001, 304 R^2 0.70) and Pseudomonadaceae (P =0.001, R^2 0.72), but negatively correlated with 305 Moraxellaceae (P = 0.0047, $R^2 0.22$). In particular, ARGs in the clean and dirty manure 306 treatments at the first fertilization of wheat were influenced by Moraxellaceae. In 307 contrast, ARGs in the clean manure treatment at the second fertilization of wheat were 308 influenced by Pseudomonadaceae. Both MGEs and Pseudomonadaceae affected the 309 ARGs in the dirty manure treatment at the second fertilization of wheat. 310 Variation partitioning analysis (VPA) (Figure 6), showed that the total variation of 311

312 ARGs in rice from bacterial communities and MGEs was 47.7%, with a greater

| 313 | contribution from bacterial (43.6%) communities than by MGEs (4.1%) The coefficient |
|-----|--|
| 314 | between bacterial and MGEs accounted for 2.6% and 8.6% in rice and wheat samples, |
| 315 | respectively. For the wheat phyllosphere, shifts in ARG composition could similarly be |
| 316 | explained by interactions between bacterial communities and MGEs, with relative |
| 317 | contributions of each group being 30.3% and 2.7%, respectively. |
| 318 | |

319 4. Discussion

4.1 Rice and wheat phyllospheres have significantly different ARG patterns.

321 As we hypothesized, the composition of ARGs associated with rice and wheat phyllospheres were different. Although we applied the same treatments to both rice and 322 wheat (no fertilizer, mineral fertilization, clean manure and dirty manure), the pattern 323 of ARGs between these two crops were distinct. Moreover, differential selection 324 between rice and wheat formed distinct bacterial communities which further affected 325 the structure of ARGs in the phyllosphere. The diversity of the bacterial community in 326 327 the rice phyllosphere was lower than that of the wheat phyllosphere. The high soil water content during rice cultivation may be a factor in this reduced diversity. Paddy fields 328 are rain fed and the high water content can lead to both aerobic and anerobic soil 329 330 conditions, which may have impacted the diversity of ARGs (Wang et al., 2018). Furthermore, flooded water may block the contact pathway from manure amended soil 331 to the phyllosphere, and thus may affect the spread of ARGs. In addition to soil water 332 333 content, supply of nutrients such as nitrogen (Ikeda et al., 2011), carbon (Wilson and Lindow, 1994), phosphate and sulphate (Delmotte et al., 2009), changes in growth stage 334 and leaf age (Kadivar and Stapleton, 2003; Yutthammo et al., 2010), may all cause shifts 335 in ARGs. Therefore, ARG composition is likely to be driven by multiple factors. 336

337

4.2 Manure enhances ARGs in the phyllosphere of rice and wheat.

In our study, organic fertilization enriched ARGs in the phyllosphere of rice and wheat,

340 which is supported by previous studies (Chen et al., 2018; Marti et al., 2013).

Additionally it has been reported that some ARGs originate from manure which may 341 enhance ARGs in the wider environment (Wang et al., 2018). The application of organic 342 343 fertilizers such as sewage sludge and pig manure can enrich the abundance of ARGs which may reach the phyllosphere (Jadhav et al., 2014; Rahube et al., 2014). The 344 phyllosphere provides a habitat for microbial communities that originate from soil, 345 water and air (Bulgarelli et al., 2013). As there is a clear correlation between the soil 346 and phyllosphere resistome, the use of manures on soil is likely to affect the microbial 347 structure on the leaf surface (Chen et al., 2017a; Chen et al., 2017b). 348

349

This study further found that the enrichment of ARGs in the phyllosphere was lower in 350 clean (reduced antibiotic burden) than dirty (current antibiotic practice) manure, 351 352 supporting our hypothesis that dirty manure can be a source of ARGs. While mutation of animal gut microbiota, creating antibiotic resistance will exist even in clean manure, 353 production of ARGs will be lower than that of dirty manure (Zhao et al., 2018). It has 354 355 been reported that wild mammals harbour ARGs, which also indicates that ARGs occur in wild populations where antibiotics in comparison to managed livestock are rare 356 (Mallon et al., 2002; Poeta et al., 2007; Tsukayama et al., 2018). This suggests that even 357 the use of clean manure can also increase the risk of ARGs spreading. As a result, 358 organic manure should be pretreated before application in order to mitigate the risk of 359 ARG transfer (Burch et al., 2017). Alternatively, composting and biochar have been 360 361 reported to effectively mitigate the risk of the antibiotic resistome (An et al., 2018; Gao et al., 2019). 362

Pathways such as the food chain and air circulation may account for the spread of ARGs into the environment. For example, after harvest, residual wheat leaves are used in animal feeds (Khush, 1997), which provides possible pathways for ARGs to enter the food chain and interact with the gut microbiome of livestock. The possible exchange of ARGs between the phyllosphere and atmosphere may also exist due to air movement (Bringel and Couee, 2015) and thus the phyllosphere may provide a new pathway of spreading ARGs to the wider environment.

370

4.3 Contribution of bacterial communities to ARGs composition in the phyllosphere 371 This study showed that changes in the composition of bacterial communities may be 372 responsible for shifts in ARGs in both rice and wheat samples, a finding supported by 373 374 Chen et al., (2018) and Zhao et al., (2018). These studies found that the application of manure could significantly alter the composition of bacterial communities in both the 375 soil and phyllosphere as well as increase the diversity of the resistome. While part of 376 377 the change in ARGs remains unexplained, it is probable that environmental factors that change during cultivation and the aerial deposition of bacteria induced compositional 378 changes in the bacterial communities of the phyllosphere. There are various channels 379 380 for bacteria to reach the phyllosphere, including soil, rain and air (Delmotte et al., 2009; Vorholt, 2012). The composition of bacterial communities are also associated with 381 geographic and climatic factors (Ren et al., 2014) as well as differences in leaf 382 construction between species. Machine learning may be used in the future for fast 383 recognition of potential microbial communities which could affect the environmental 384

387 Conclusion

In this study, a total of 162 unique ARGs and 5 MGEs were detected through HT-qPCR. 388 Rice and wheat phyllospheres had differing patterns of ARGs and bacterial 389 communities, indicating that multiple factors, such as plant species, diverse growth 390 conditions, nutrient supply and atmospheric movement, may affect diversity in the 391 phyllosphere. Both clean and dirty manures enhanced ARGs in the phyllosphere, with 392 dirty manure in particular causing the greatest enrichment of ARGs. As rice and wheat 393 are staple crops globally, the application of both clean and dirty manure that deliver 394 ARGs and MGEs to the food chain may pose a significant risk to human health and act 395 396 as a conduit for ARGs to reach the environment.

397

398 ACKNOWLEDGMENTS

This research was funded by the National Key Research and Development Program of China (2017YFE0107300), the National Natural Science Foundation of China (41571130063), the Strategic Priority Research Program of the Chinese Academy of Sciences (XDB15020302 and XDB15020402), The James Hutton Institute receives financial support from support from the Scottish Government, Rural and Environment Science and Analytical Services Division.

405 **References**

- 406 Allen, H.K., Donato, J., Wang, H.H., Cloud-Hansen, K.A., Davies, J., Handelsman, J., 2010. Call of the wild:
- 407 antibiotic resistance genes in natural environments. Nature Reviews Microbiology 8, 251-259.
- 408 An, X.L., Chen, Q.L., Zhu, D., Su, J.Q., 2018. Distinct effects of struvite and biochar amendment on the
- 409 class 1 integron antibiotic resistance gene cassettes in phyllosphere and rhizosphere. Science of the
- 410 Total Environment 631-632, 668-676.
- 411 Antunes, P., Campos, J., Mourão, J., Pereira, J., Novais, C., Peixe, L., 2018. Inflow water is a major source
- of trout farming contamination with Salmonella and multidrug resistant bacteria. Science of the TotalEnvironment 642, 1163-1171.
- Berlec, A., 2012. Novel techniques and findings in the study of plant microbiota: Search for plant
 probiotics. Plant Science 193, 96-102.
- 416 Bringel, F., Couee, I., 2015. Pivotal roles of phyllosphere microorganisms at the interface between plant
- 417 functioning and atmospheric trace gas dynamics. Frontiers in Microbiology 6.
- Bulgarelli, D., Schlaeppi, K., Spaepen, S., van Themaat, E.V.L., Schulze-Lefert, P., 2013. Structure and
 Functions of the Bacterial Microbiota of Plants, in: Merchant, S.S. (Ed.), Annual Review of Plant Biology,
- 420 Vol 64, pp. 807-838.
- 421 Burch, T.R., Sadowsky, M.J., LaPara, T.M., 2017. Effect of Different Treatment Technologies on the Fate
- of Antibiotic Resistance Genes and Class 1 Integrons when Residual Municipal Wastewater Solids are
 Applied to Soil. Environmental Science & Technology 51, 14225-14232.
- 424 Camacho, D.M., Collins, K.M., Powers, R.K., Costello, J.C., Collins, J.J., 2018. Next-Generation Machine
 425 Learning for Biological Networks. Cell 173, 1581-1592.
- Caporaso, J.G., Bittinger, K., Bushman, F.D., DeSantis, T.Z., Andersen, G.L., Knight, R., 2010a. PyNAST: a
 flexible tool for aligning sequences to a template alignment. Bioinformatics 26, 266-267.
- 428 Caporaso, J.G., Kuczynski, J., Stombaugh, J., Bittinger, K., Bushman, F.D., Costello, E.K., Fierer, N., Peña,
- 429 A.G., Goodrich, J.K., Gordon, J.I., Huttley, G.A., Kelley, S.T., Knights, D., Koenig, J.E., Ley, R.E., Lozupone,
- 430 C.A., McDonald, D., Muegge, B.D., Pirrung, M., Reeder, J., Sevinsky, J.R., Turnbaugh, P.J., Walters, W.A.,
- Widmann, J., Yatsunenko, T., Zaneveld, J., Knight, R., 2010b. QIIME allows analysis of high-throughput
 community sequencing data. Nature Methods 7, 335-336.
- 433 Chen, Q., An, X., Li, H., Su, J., Ma, Y., Zhu, Y.-G., 2016. Long-term field application of sewage sludge
- increases the abundance of antibiotic resistance genes in soil. Environment International 92-93, 1-10.
- 435 Chen, Q.L., An, X.L., Li, H., Zhu, Y.G., Su, J.Q., Cui, L., 2017a. Do manure-borne or indigenous soil 436 microorganisms influence the spread of antibiotic resistance genes in manured soil? Soil Biology &
- 437 Biochemistry 114, 229-237.
- 438 Chen, Q.L., Li, H., Zhou, X.Y., Zhao, Y., Su, J.Q., Zhang, X., Huang, F.Y., 2017c. An underappreciated hotspot
- of antibiotic resistance: The groundwater near the municipal solid waste landfill. Science of the TotalEnvironment 609, 966-973.
- Chen, Q.L., An, X.L., Zheng, B.X., Ma, Y.B., Su, J.Q., 2018. Long-term organic fertilization increased
 antibiotic resistome in phyllosphere of maize. Science of the Total Environment 645, 1230-1237.
- 443 Chen, Q.L., An, X.L., Zhu, Y.G., Su, J.Q., Gillings, M.R., Ye, Z.L., Cui, L., 2017b. Application of Struvite Alters
- 444 the Antibiotic Resistome in Soil, Rhizosphere, and Phyllosphere. Environmental Science & Technology
- 445 51, 8149-8157.
- 446 Cheng, W., Chen, H., Su, C., Yan, S., 2013. Abundance and persistence of antibiotic resistance genes in
- 447 livestock farms: a comprehensive investigation in eastern China. Environment International 61, 1-7.

- 448 D'Costa, V.M., King, C.E., Kalan, L., Morar, M., Sung, W.W.L., Schwarz, C., Froese, D., Zazula, G., Calmels,
- F., Debruyne, R., Golding, G.B., Poinar, H.N., Wright, G.D., 2011. Antibiotic resistance is ancient. Nature
 477, 457-461.
- 451 Delmotte, N., Knief, C., Chaffron, S., Innerebner, G., Roschitzki, B., Schlapbach, R., von Mering, C.,
- 452 Vorholt, J.A., 2009. Community proteogenomics reveals insights into the physiology of phyllosphere
- bacteria. Proceedings of the National Academy of Sciences of the United States of America 106, 16428-16433.
- Edgar, R.C., 2010. Search and clustering orders of magnitude faster than BLAST. Bioinformatics 26, 24602461.
- Galili, T., O'Callaghan, A., Sidi, J., Sievert, C., 2018. heatmaply: an R package for creating interactive
 cluster heatmaps for online publishing. Bioinformatics 34, 1600-1602.
- Gao, Y.Z., Lu, C., Shen, D., Liu, J., Ma, Z., Yang, B., Ling, W.T., Waigi, M.G., 2019. Elimination of the risks
 of colistin resistance gene (mcr-1) in livestock manure during composting. Environment International
 126, 61-68.
- 462 Gillings, M.R., 2017. Lateral gene transfer, bacterial genome evolution, and the Anthropocene. Annals463 of the New York Academy of Sciences 1389, 20-36.
- Gillings, M.R., Gaze, W.H., Pruden, A., Smalla, K., Tiedje, J.M., Zhu, Y.-G., 2014. Using the class 1 integronintegrase gene as a proxy for anthropogenic pollution. The Isme Journal 9, 1269.
- Wickham. H., W.C., Henry. L., Pedersen. L. T, Takahashi. K., Wilke. C., Woo. K., RStudio, 2018. ggplot2:
 Create Elegant Data Visualisations Using the Grammar of Graphics. https://CRAN.Rproject.org/package=ggplot2.
- Hao, X.H., Liu, S.L., Wu, J.S., Hu, R.G., Tong, C.L., Su, Y.Y., 2008. Effect of long-term application of inorganic
 fertilizer and organic amendments on soil organic matter and microbial biomass in three subtropical
 paddy soils. Nutrient Cycling in Agroecosystems 81, 17-24.
- 472 Ikeda, S., Anda, M., Inaba, S., Eda, S., Sato, S., Sasaki, K., Tabata, S., Mitsui, H., Sato, T., Shinano, T.,
- 473 Minamisawa, K., 2011. Autoregulation of Nodulation Interferes with Impacts of Nitrogen Fertilization
- 474 Levels on the Leaf-Associated Bacterial Community in Soybeans. Appl Environ Microbiol 77, 1973-1980.
- Jadhav, A., Shanmugham, B., Rajendiran, A., Pan, A., 2014. Unraveling novel broad-spectrum
 antibacterial targets in food and waterborne pathogens using comparative genomics and protein
 interaction network analysis. Infection Genetics and Evolution 27, 300-308.
- Oksanen. J, F.G.B., Friendly. M., Kindt. R., Legendre. P., McGlinn. D., Minchin. R. P., R. B. O'Hara., Simpson.
 L. G., Solymos. P., M. Henry H. Stevens., Szoecs. E., Wagner. H., 2018. Vegan: Community Ecology
- 480 Package. R package version 2.5-3. https://CRAN.R-project.org/package=vegan.
- 481 Joy, S.R., Bartelthunt, S.L., Snow, D.D., Gilley, J.E., Woodbury, B.L., Parker, D.B., Marx, D.B., Li, X., 2013.
- 482 Fate and transport of antimicrobials and antimicrobial resistance genes in soil and runoff following land
- 483 application of swine manure slurry. Environmental Science & Technology 47, 12081-12088.
- Kadivar, H., Stapleton, A.E., 2003. Ultraviolet radiation alters maize phyllosphere bacterial diversity.
 Microbial Ecology 45, 353-361.
- 486 Khush, G.S., 1997. Origin, dispersal, cultivation and variation of rice. Plant Molecular Biology 35, 25-34.
- Klappenbach, J.A., Saxman, P.R., Cole, J.R., Schmidt, T.M., 2001. rrndb: the Ribosomal RNA Operon Copy
 Number Database. Nucleic Acids Res 29, 181-184.
- 489 Langille, M.G.I., Zaneveld, J., Caporaso, J.G., McDonald, D., Knights, D., Reyes, J.A., Clemente, J.C.,
- 490 Burkepile, D.E., Thurber, R.L.V., Knight, R., Beiko, R.G., Huttenhower, C., 2013. Predictive functional
- 491 profiling of microbial communities using 16S rRNA marker gene sequences. Nature Biotechnology 31,

- 492 814-821.
- 493 Leverstein-van Hall, M.A., Dierikx, C.M., Stuart, J.C., Voets, G.M., van den Munckhof, M.P., van Essen-
- 494 Zandbergen, A., Platteel, T., Fluit, A.C., van de Sande-Bruinsma, N., Scharinga, J., Bonten, M.J.M., Mevius,
- 495 D.J., Natl, E.S.G., 2011. Dutch patients, retail chicken meat and poultry share the same ESBL genes,
- 496 plasmids and strains. Clinical Microbiology and Infection 17, 873-880.
- Lindow, S.E., Brandl, M.T., 2003. Microbiology of the phyllosphere. Appl Environ Microbiol 69, 1875-1883.
- Mallon, D.J.P., Corkill, J.E., Hazel, S.M., Wilson, J.S., French, N.P., Bennett, M., Hart, C.A., 2002. Excretion
 of vancomycin-resistant enterococci by wild mammals. Emerging Infectious Diseases 8, 636-638.
- Marshall, B.M., Levy, S.B., 2011. Food Animals and Antimicrobials: Impacts on Human Health. Clinical
 Microbiology Reviews 24, 718-733.
- Marti, R., Scott, A., Tien, Y.C., Murray, R., Sabourin, L., Zhang, Y., Topp, E., 2013. Impact of Manure
 Fertilization on the Abundance of Antibiotic-Resistant Bacteria and Frequency of Detection of Antibiotic
 Resistance Genes in Soil and on Vegetables at Harvest. Appl Environ Microbiol 79, 5701-5709.
- 506 Martinez, J.L., 2008. Antibiotics and antibiotic resistance genes in natural environments. Science 321,507 365-367.
- McDonald, D., Price, M.N., Goodrich, J., Nawrocki, E.P., DeSantis, T.Z., Probst, A., Andersen, G.L., Knight,
 R., Hugenholtz, P., 2012. An improved Greengenes taxonomy with explicit ranks for ecological and
 evolutionary analyses of bacteria and archaea. Isme Journal 6, 610-618.
- 511 Mhuireach, G., Johnson, B.R., Altrichter, A.E., Ladau, J., Meadow, J.F., Pollard, K.S., Green, J.L., 2016. 512 Urban greenness influences airborne bacterial community composition. Science of the Total 513 Environment 571, 680-687.
- 514 Poeta, P., Costa, D., Igrejas, G., Rojo-Bezares, B., Saenz, Y., Zarazaga, M., Ruiz-Larrea, F., Rodrigues, J.,
- 515 Torres, C., 2007. Characterization of vanA-containing Enterococcus faecium isolates carrying tn5397-like
- and Tn916/Tn1545-like transposons in wild boars (Sus scrofa). Microbial Drug Resistance-Mechanisms
- 517 Epidemiology and Disease 13, 151-156.
- 518 Pruden, A., Pei, R.T., Storteboom, H., Carlson, K.H., 2006. Antibiotic resistance genes as emerging
 519 contaminants: Studies in northern Colorado. Environmental Science & Technology 40, 7445-7450.
- Rahman, M.M., Shan, J., Yang, P., Shang, X., Xia, Y., Yan, X., 2018. Effects of long-term pig manure
 application on antibiotics, abundance of antibiotic resistance genes (ARGs), anammox and
 denitrification rates in paddy soils. Environmental Pollution 240, 368-377.
- 523 Rahube, T.O., Marti, R., Scott, A., Tien, Y.C., Murray, R., Sabourin, L., Zhang, Y., Duenk, P., Lapen, D.R.,
- 524 Topp, E., 2014. Impact of fertilizing with raw or anaerobically digested sewage sludge on the abundance
- 525 of antibiotic-resistant coliforms, antibiotic resistance genes, and pathogenic bacteria in soil and on 526 vegetables at harvest. Applied & Environmental Microbiology 80, 6898-6907.
- Ren, G., Zhang, H., Lin, X., Zhu, J., Jia, Z., 2014. Response of phyllosphere bacterial communities to
 elevated CO2 during rice growing season. Applied Microbiology and Biotechnology 98, 9459-9471.
- 529 Stokes, H.W., Nesbø, C.L., Holley, M., Bahl, M.I., Gillings, M.R., Boucher, Y., 2006. Class 1 Integrons 530 Potentially Predating the Association with Tn402-Like Transposition Genes Are Present in a
- 531 Sediment Microbial Community. Journal of Bacteriology 188, 5722-5730.
- 532 Su, J.-Q., Wei, B., Ou-Yang, W.-Y., Huang, F.-Y., Zhao, Y., Xu, H.-J., Zhu, Y.-G., 2015. Antibiotic Resistome
- 533 and Its Association with Bacterial Communities during Sewage Sludge Composting. Environmental
- 534 Science & Technology 49, 7356-7363.
- 535 Tello, A., Austin, B., Telfer, T.C., 2012. Selective Pressure of Antibiotic Pollution on Bacteria of Importance

- to Public Health. Environmental Health Perspectives 120, 1100-1106.
- 537 Tilman, D., Cassman, K.G., Matson, P.A., Naylor, R., Polasky, S., 2002. Agricultural sustainability and
- 538 intensive production practices. Nature 418, 671-677.
- 539 Tsukayama, P., Boolchandani, M., Patel, S., Pehrsson, E.C., Gibson, M.K., Chiou, K.L., Jolly, C.J., Rogers, J.,
- Phillips-Conroy, J.E., Dantas, G., 2018. Characterization of Wild and Captive Baboon Gut Microbiota and
 Their Antibiotic Resistomes. Msystems 3, e00016-18.
- 542 Turner, S., Pryer, K.M., Miao, V.P.W., Palmer, J.D., 1999. Investigating deep phylogenetic relationships
- among cyanobacteria and plastids by small submit rRNA sequence analysis. Journal of EukaryoticMicrobiology 46, 327-338.
- 545 Vorholt, J.A., 2012. Microbial life in the phyllosphere. Nature Reviews Microbiology 10, 828-840.
- 546 Wang, F., Xu, M., Stedtfeld, R.D., Sheng, H.J., Fan, J.B., Liu, M., Chai, B.L., de Carvalho, T.S., Li, H., Li, Z.P.,
- Hashsham, S.A., Tiedje, J.M., 2018. Long-Term Effect of Different Fertilization and Cropping Systems on
 the Soil Antibiotic Resistome. Environmental Science & Technology 52, 13037-13046.
- Wilson, M., Lindow, S.E., 1994. COEXISTENCE AMONG EPIPHYTIC BACTERIAL-POPULATIONS MEDIATED
 THROUGH NUTRITIONAL RESOURCE PARTITIONING. Appl Environ Microbiol 60, 4468-4477.
- Woodward, F.I., Lomas, M.R., 2004. Vegetation dynamics simulating responses to climatic change.
 Biological Reviews 79, 643-670.
- Wu, S., Huang, J.H., Wu, Q.P., Zhang, F., Zhang, J.M., Lei, T., Chen, M.T., Ding, Y., Xue, L., 2018. Prevalence
 and Characterization of Staphylococcus aureus Isolated From Retail Vegetables in China. Frontiers in
 Microbiology 9.
- Xie, W.Y., Mcgrath, S.P., Su, J.Q., Hirsch, P.R., Clark, I.M., Shen, Q., Zhu, Y.G., Zhao, F.J., 2016. Long-Term
 Impact of Field Applications of Sewage Sludge on Soil Antibiotic Resistome. Environmental Science &
 Technology 50, 12602-12611.
- Yutthammo, C., Thongthammachat, N., Pinphanichakarn, P., Luepromchai, E., 2010. Diversity and
 Activity of PAH-Degrading Bacteria in the Phyllosphere of Ornamental Plants. Microbial Ecology 59, 357368.
- Zhang, L.N., Fu, Y., Xiong, Z.Y., Ma, Y.B., Wei, Y.H., Qu, X.Y., Zhang, H.X., Zhang, J.M., Liao, M., 2018. Highly
 Prevalent Multidrug-Resistant Salmonella From Chicken and Pork Meat at Retail Markets in Guangdong,
 China. Frontiers in Microbiology 9.
- Zhao, Y., Su, J.Q., An, X.L., Huang, F.Y., Rensing, C., Brandt, K.K., Zhu, Y.G., 2018. Feed additives shift gut
 microbiota and enrich antibiotic resistance in swine gut. Science of the Total Environment 621, 12241232.
- Zhou, L.J., Ying, G.G., Liu, S., Zhang, R.Q., Lai, H.J., Chen, Z.F., Pan, C.G., 2013. Excretion masses and
 environmental occurrence of antibiotics in typical swine and dairy cattle farms in China. Science of the
 Total Environment 444, 183-195.
- 571 Zhu, B., Chen, Q., Chen, S., Zhu, Y.G., 2016. Does organically produced lettuce harbor higher abundance
 572 of antibiotic resistance genes than conventionally produced? Environment International 98, 152-159.
- 573 Zhu, D., Zheng, F., Chen, Q.-L., Yang, X.-R., Christie, P., Ke, X., Zhu, Y.-G., 2018. Exposure of a Soil
- 574 Collembolan to Ag Nanoparticles and AgNO3 Disturbs Its Associated Microbiota and Lowers the 575 Incidence of Antibiotic Resistance Genes in the Gut. Environmental Science & Technology 52, 12748-576 12756.
- Zhu, Y.G., Gillings, M., Simonet, P., Stekel, D., Banwart, S., Penuelas, J., 2017. Microbial mass movements.
 Science 357, 1099-1100.
- 579 Zhu, Y.-G., Johnson, T.A., Su, J.-Q., Qiao, M., Guo, G.-X., Stedtfeld, R.D., Hashsham, S.A., Tiedje, J.M., 2013.

580 Diverse and abundant antibiotic resistance genes in Chinese swine farms. Proceedings of the National
 581 Academy of Sciences 110, 3435-3440.

582

583

- 584
- 585
- 586
- 587

588 Figure Legend

Figure 1. a: Number of ARGs and MGEs detected from different samples. b: The 589 percentage of antibiotic resistant mechanisms within all samples. ARGs were separated 590 into 10 classes based on the following reference antibiotic resistant genes: 591 Aminoglycosides, beta-lactams, chloramphenicol, MGEs, MLSB, Multidrug, 592 sulfonamides, tetracycline, vancomycin and other unknown. For rice samples, RCF, 593 RMF, RCM and RDM represent no fertilizer, mineral fertilization, clean and dirty 594 manure fertilization. RCF2, RMF2, RCM2 and RDM2 represent second fertilization 595 samples. Similar nomenclature applies to wheat samples. 596

597

Figure 2. Characteristics of ARGs in rice and wheat samples. a and b represent patterns 598 of normalized abundance (copy/cell) of rice and wheat, respectively. c and d depict the 599 PCoA analyses of ARGs. "R" represents Rice samples, "W" represents wheat samples. 600 (1), (2), (3), (4) represent ARGs in the manure treatments at the first fertilization of rice 601 (RCM, RDM), second fertilization of rice (RCM2, RDM2), first control (no fertilizer) 602 samples and mineral ferilitzation treatments (RCF, RMF) and second control (no 603 fertilizer) samples and mineral fertilization (RCF2, RMF2). (5), (6), (7), (8) represented 604 ARGs in manure treatments from the first fertilization of wheat (WCM, WDM), second 605 fertilization of wheat sample (WCM2, WDM2), first control (no fertilizer) and mineral 606 fertilization treatments (WCF, WMF) and second control (no fertilizer) and mineral 607 fertilization (WCF2, WMF2). 608

609

Figure 3. Heatmap analysis of ARGs in rice and wheat samples. The vertical axis lists the detected ARGs found in this study. For rice, RCF, RMF, RCM and RDM represent no fertilizer, mineral fertilization, clean and dirty manure fertilization respectively. RCF2, RMF2, RCM2 and RDM2 represent the second rice fertilization. Similar nomenclature applies to wheat samples. The order of the genes was based on their similarity abundance.

Figure 4. PCoA analysis of bacterial communities based on Bray-Curtis distance. a) represents the first and second fertilization of phyllosphere samples from rice; b) and c) represent the first and second wheat fertilization, respectively.

620

Figure 5. CCA (a) and RDA (b) analysis of the correlation among ARGs in rice and wheat samples, major microbial families (>1%) (Enterobacteriaceae, Bacillaceae, Pseudomonadaceae, Rhizobiaceae, Moraxellaceae) and MGEs in rice and wheat respectively. Label (1), (2), (3), (4) represent the ARGs in the first and second fertilization of rice and wheat samples respectively. Pseudomonadaceae and MGEs present a positive correlation in both rice and wheat samples, whereas Moraxellaceae in wheat samples shows a negative correlation.

628

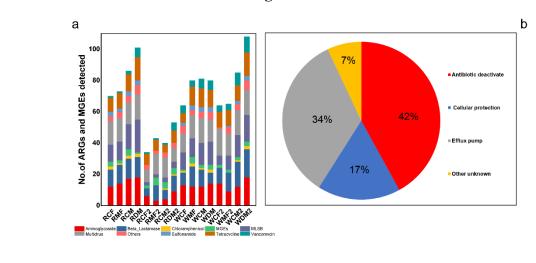
Figure 6. Variation partitioning analysis highlighting the influence of the bacterial

630 community and mobile genetic elements to the change in ARGs. For rice samples, 50.3%

of changes in ARGs were explained through the bacterial community, MGEs and their

coefficient, whereas in wheat samples this was 41.6%.

Figure 1





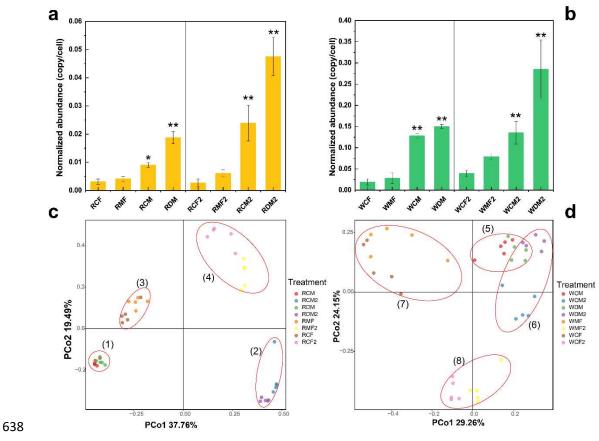


Figure 3

