# 1 Milling plant and soil material in plastic tubes over-estimates carbon and

# 2 under-estimates nitrogen concentrations

- 4 Stuart W. Smith<sup>a,b,c\*</sup>, A. H. Jean Robertson<sup>c</sup>, Andrew A. Meharg<sup>a</sup>, Robin J. Pakeman<sup>c</sup>, David
- 5 Johnson<sup>a</sup>, Sarah J. Woodin<sup>a</sup>, René van der Wal<sup>b</sup>
- <sup>a</sup>IBES, University of Aberdeen, St Machar Drive, Aberdeen AB24 3UU, UK.
- <sup>7</sup> <sup>b</sup>ACES, University of Aberdeen, St Machar Drive, Aberdeen AB24 3UU,UK.
- <sup>c</sup>The James Hutton Institute, Craigiebuckler, Aberdeen AB15 8QH, UK.
- 9 \* Corresponding author. E-mail address: <u>s.w.smith@abdn.ac.uk</u> (S.W. Smith).
- 10
- 11
- 12
- 13
- 14
- 15
- 16
- 17
- 18
- 19

21 Abstract

Background and aims Milling of plant and soil material in plastic tubes, such as microcentrifuge tubes, over-estimates carbon (C) and under-estimates nitrogen (N) concentrations due to the introduction of polypropylene into milled samples, as identified using Fourier-transform infra-red spectroscopy. Methods and results This study compares C and N concentrations of roots and soil milled in microcentrifuge tubes versus stainless steel containers, demonstrating that a longer milling time, greater milling intensity, smaller sample size and inclusion of abrasive sample material all increase polypropylene contamination from plastic tubes leading to overestimation of C concentrations by up to 8 % (0.08 g  $g^{-1}$ ). Conclusions Erroneous estimations of C and N, and other analytes, must be assumed after milling in plastic tubes and milling methods should be adapted to minimise such error. Keywords: Carbon, Fourier-transform infrared spectroscopy, grinding, microcentrifuge tubes, milling, nitrogen. 

### 43 Introduction

Analysis of the carbon (C) and nitrogen (N) contents of plants and soils is crucial for 44 45 assessing how climate and land-use change affect global biogeochemical cycles (Guo and Gilford 2002; Bellamy et al. 2005; Powlson et al. 2011). Estimation of ecosystem C and N 46 47 stores, inputs and losses depends upon accurate determination of C and N concentrations in ecological materials. Automated elemental analysis has become ubiquitous for C and N 48 49 determination, due to its accuracy and reliability (Kalembas and Jenkinson, 1973; Soon and 50 Abboud 1991; Lal et al. 2001). This type of analysis, based on dry oxidation, only requires a 51 2-20 mg sub-sample and thus precision is dependent on sample homogeneity (Jimenez and Ladha 1993). Homogenisation is achieved through milling, often in stainless steel grinding 52 53 jars containing stainless steel balls, with mills typically processing 1–3 samples at once (Allen 1989). However, significant time can be saved in preparation of large sets of samples 54 using microcentrifuge tubes with stainless steel balls, with tens of samples being processed 55 56 simultaneously (Warren and Adams 2004; Salvo-Chirnside et al. 2011; Nadeem et al. 2012). 57 Milling in microcentrifuge tubes avoids the loss of material which occurs through cleaning of 58 steel grinding jars between each use and is thus ideal for small quantities of plant and soil 59 material (e.g. roots, decomposed litter, soil fauna). The use of disposable containers also 60 minimizes any cross-sample contamination. However, during an investigation of C loss from 61 root litter we found that the average C concentration was 5.45 percentage points higher (50.02 vs 44.56 %C), and N concentration 0.072 percentage points lower (0.912 vs 0.985 62 %N), in roots milled in microcentrifuge tubes as compared to steel jars (Figure 1). The 63 64 additional C was identified, using Fourier-transform infrared (FTIR) spectroscopy, to be 65 atactic-polypropylene  $((C_{15}H_{30})_n)$  and some copolymers, originating from the microcentrifuge tubes. Abraded polypropylene also increased sample mass with N free material, thus reducing 66

the overall N concentration. This study aimed to define milling procedures that would
preclude plastic contamination, investigating the effects of milling time, intensity and sample
size on measured C and N concentrations of roots and soils milled in microcentrifuge tubes
and in steel jars. Material milled by both methods was tested for polypropylene by FTIR
spectroscopy.

72

# 73 Materials and Methods

74 Two microcentrifuge types were investigated; reaction vial safe-lock 2 ml Retsch® tubes 75 (81.7 %C) and 2 ml Alpha laboratories microcentrifuge tubes (80.8 %C), compared to 10 ml 76 stainless steel jars (0.95 %C). Dried, finely chopped Molinia caerulea roots (diameter 77  $0.35\pm0.07$  mm) and two soil types, an organic iron-podzol (31.5 % sand, 61.5 % silt, 7.0 % 78 clay) and a sandy-loam (70.2 % sand, 14.8 % silt, 15 % clay) were milled on a Retsch® 79 Mixer Mill MM 400 (Retsch®, Germany), testing the effects of three variables ("treatments"): milling time (1-15 mins); milling intensity (15–30 Hz); sample size (10–60 80 81 mg). Twenty samples were milled per tube type and ten samples for stainless steel jars at 82 randomly selected intervals for each treatment. For the sample size treatment, three additional samples (< 20 mg) were milled to aid the statistical analysis. One parameter was altered at a 83 84 time; the others remained constant at 10 mins, 30 Hz or 30 mg of root per tube. Only three soil sample sizes were investigated (10.9, 28.5, 59.3 mg). Root particle size was not small 85 enough for CN analysis after 10 mins of milling in microcentrifuge tubes, so all samples were 86 re-milled in steel jars for 1 minute at 30 Hz. After milling, a 5 mg sub-sample was taken for 87 elemental analysis (Carlo-Erba NA 1500 Series 2, USA). Contamination of milled roots was 88 tested on a Bruker Vertex 70 Spectrometer (Bruker Optics, Ettlingen, Germany) comparing 89 milled samples to microcentrifuge shavings (methods as in Artz et al. 2008). Treatment 90

91 effects on percent C and N were analysed using linear models (except sample size on %C milled in microcentrifuge tubes which was fitted with a non-linear exponential function) in R 92 (version 2.10.1, R Development Core Team, 2009). There was no significant difference 93 94 between the plastic Alpha and Retsch tubes in any treatment (P>0.05), so these were grouped 95 for statistical analysis. Difference between the change in C and N concentration with treatment in microcentrifuge tubes and in stainless steel jars is indicated by the interaction 96 97 term of the model. However, for sample size the interaction could not be determined due to differences in linearity (linear for stainless steel; non-linear for microcentrifuge tube). In 98 99 order to compare the strength of the interaction term for sample size on %C with the other milling treatments (milling time and intensity) a single linear model was used on sample sizes 100 <26.5 mg, a threshold below which a linear relationship was displayed. 101

102

#### 103 **Results and Discussion**

Carbon concentrations in microcentrifuge milled roots increased with increasing milling time 104 105 and intensity, whilst %C of steel milled roots remained unchanged across both treatments (interaction terms in Table 1; Figure 2A, 2B). Milling a small quantity of sample ( $\leq 20$  mg) 106 produced the greatest polypropylene-derived C contamination (up to 8 %C or 0.08 g  $g^{-1}$ ) of 107 108 all the treatments, due to greater abrasion between the ball and tubes (Table 1; Figure 2C). Polypropylene was identified in all roots milled in microcentrifuge tubes, but not in steel 109 milled samples (Figure 3). Milling small samples in microcentrifuge tubes should be avoided 110 111 and milling time and intensity should be reduced, yet this can prevent particle size being sufficiently small for C and N analysis. Although not tested here, non-spherical balls could be 112 used to dissipate the intensity of contact between ball and microcentrifuge tube (Salvo-113 Chirnside et al. 2011); however, cones require more energy to achieve the same degree of 114

homogenization as spherical balls (Herbst and Lo 1989; Lameck et al. 2006), potentiallyresulting in similar contamination.

117

118 Carbon contamination was greater for soil than plant material, and greater for sandy-loam soil 119 than for organic iron-podzol (Figure 4), likely due to less organic matter and greater sand 120 content resulting in more abrasion of the tubes. The risk of polypropylene C contamination is 121 likely to be greater when milling abrasive material and this needs further investigation. It 122 should be noted, however, that the quantities of soil milled were small (30 mg of soil 123 occupied <1 % of a tube compared to ~22 % for roots) and, as contamination decreased with 124 increasing sample size, this may be less of a problem with much larger samples.

125

Milled roots had a significantly lower average N concentration (across all treatments) in 126 microcentrifuge tubes (0.621 %N) compared to steel jars (0.695 %N) (Table 1; Figure 2; D, 127 128 E, F). The lower N content corresponds with our initial observations (Figure 1) and is assumed to be due to dilution of N by the addition of plastic to the milled sample mass. 129 Unlike root C, there was no significant interaction between milling method and milling time, 130 intensity, or sample size (Table 1). This is due to variability in %N, which was much greater 131 than variability in %C, even in steel milled samples. This likely reflects natural %N variation 132 133 within perennial root tissues of different ages (Robinson and Rorison 1988) and/or greater analytical error at the low root N concentrations. Soil N concentrations were lower for 134 organic-iron podzol samples milled in microcentrifuge tubes compared to steel jars. This 135 effect was not detected in sandy-loam soil, despite its increased %C, due to low N 136 concentrations (~0.05 %) being poorly detected on the elemental analyser (data not shown). 137

#### 139 Conclusions

Milling small amounts of soil and plant material in microcentrifuge tubes risks over-140 estimating %C and potentially under-estimating %N. To reduce sample contamination, 141 milling of material in microcentrifuge tubes should be kept to as short duration, as low 142 intensity and as large a sample size as possible for achieving the required particle size. 143 Alternatively, the use of plastic tubes in milling should be avoided as, without quantification 144 on a case-by-case basis, C contamination must be assumed. Any other analyte will be 145 146 underestimated in samples milled in plastic containers that are abraded during milling. Use of erroneous C, N and other analyte concentrations could have large implications for 147 calculation of element budgets and, indeed, for any biological studies involving elemental 148 analysis. 149

150

### 151 Acknowledgements

We thank Ken Cruickshank, Angela Main, Rachael Hill, Elaine Runge and Marcel Junker for
their skilled technical assistance. SWS was funded by a Biotechnology and Biological
Sciences Research Council studentship.

155

## 156 **References**

Allen SE (1989) Chemical Analysis of Ecological Materials. 2<sup>nd</sup> ed., Blackwell Scientific,
Cambridge, UK.

160	Artz RE, Chapman SJ, Robertson AHJ, Potts JM, Laggoun-De´farge JM, Gogo S, Comont L,
161	Disnar JR, Francez AJ (2008) FTIR spectroscopy can be used as a screening tool for organic
162	matter quality in regenerating cutover peatlands. Soil Biol Biochem 40: 515-527.
163	
164	Bellamy PH, Loveland PJ, Bradley RI, Lark RM, Kirk GJD (2005) Carbon losses from all
165	soils across England and Wales 1978–2003. Nature 437: 245–248.
166	
167	Guo LB, Gifford RM (2002) Soil carbon stocks and land use change: a meta-analysis. Glob
168	Change Biol 8:345-360.
169	
170	Herbst JA, Lo YC (1989) Grinding efficiency with balls or cones as media. Int J Miner
171	Process 26, 141-151.
172	
173	Jimenez RR, Ladha JK (1993) Automated elemental analysis: A rapid and reliable but
174	expensive measurement of total carbon and nitrogen in plant and soil samples. Comm Soil
175	Sci Plan 24: 1897-1924
176	

177 Kalembasa SJ, Jenkinson DS (1973) Comparative study of titrimetric and gravimetric

methods for the determination of organic carbon in soil. J Sci Food Agri 24: 1085-1090

180	Lal R, Kimble JM, Follett RF, Stewart BA (2001). Assessment Methods for Soil Carbon.
181	CRC Press, Boca Raton, FL, USA.
182	
183	Lameck NS, Kiangi KK, Moys MH (2006) Effects of grinding media shapes on load
184	behaviour and mill power in a dry ball mill. Miner Eng 19: 1357–1361.

185

186 Nadeem M, Mollier A, Morel C, Vives A, Prud'homme L, Pellerin S (2012) Maize (Zea

187 mays L.) endogenous seed phosphorus remobilization is not influenced by exogeneous

188 phosphorus availability during germination and early growth stages. Plant Soil 357: 13 -24.

189

190 Powlson DS, Whitmore AP, Goulding WT (2011) Soil carbon sequestration to mitigate

191 climate change: a critical re-examination to identify the true and the false, Eur J Soil Sci 62:

192 42 - 55.

193

R Development Core Team (2009) R: A language and environment for statistical computing.
R Foundation for Statistical Computing, Vienna, Austria. ISBN 3-900051-07-0, URL
http://www.R-project.org.

197

Robinson D, Rorison IH (1988) Plasticity in grass species in relation to nitrogen supply.
Funct Ecol 2, 249-257.

200

201 Salvo-Chirnside E, Kane S, Kerr LE (2011) Protocol: high throughput silica-based

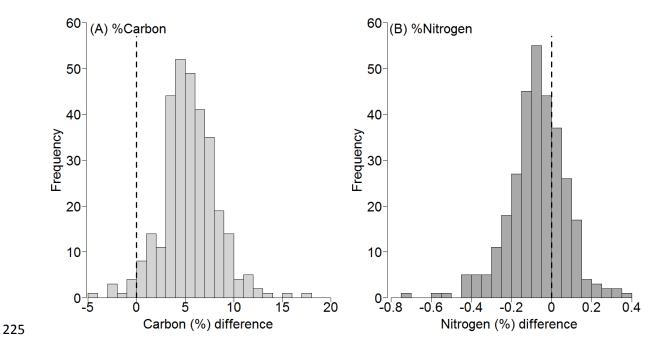
202 purification of RNA from Arabidopsis seedlings in a 96-well. Plant Methods 7: 40.

205 determination. Comm Soil Sci Plan 22: 943 – 954.

- 207 Warren CR, Adams MA (2004) What determines rates of photosynthesis per unit nitrogen in
- 208 Eucalyptus seedlings? Funct Plant Biol 31: 1169 1178.

- \_\_\_\_

## 224 Figures



**Figure 1.** The shift in absolute percent carbon (**A**) and nitrogen (**B**) in partially decomposed root litter attributed to milling samples in microcentrifuge tubes. C and N difference values represent stainless steel jar milled material subtracted from microcentrifuge milled material. Dotted lines represent no difference due to milling method. Milling in microcentrifuge tubes significantly increased %C (paired t-test:  $t_{309} = 33.798$ , p <0.001) and decreased %N (paired t-test:  $t_{309} = -8.757$  p <0.001) compared to stainless steel milled samples.

232

233

234

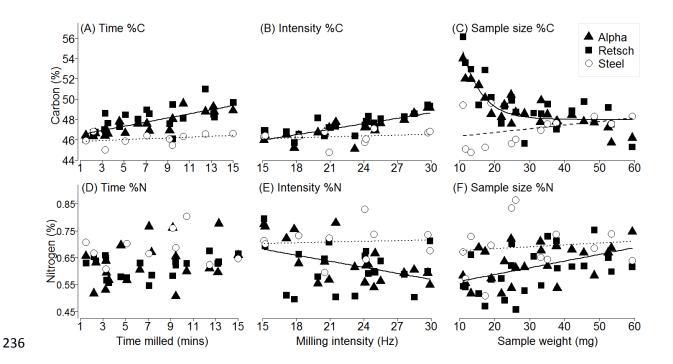
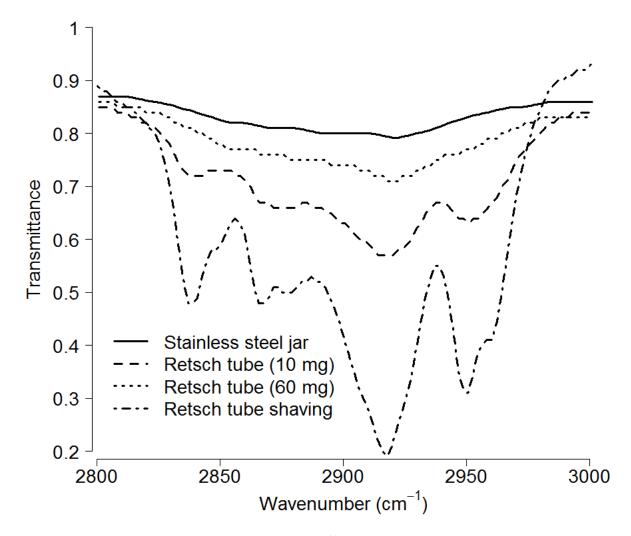
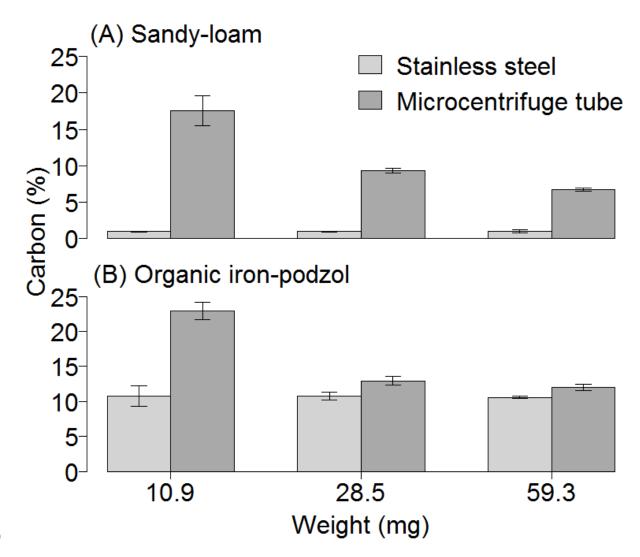


Figure 2. Carbon concentrations (%) in *Molinia caerulea* roots for different milling times (A),
intensities (B) and sample sizes (C), and nitrogen concentrations (%) for the same treatments (D, E,
F, respectively). Microcentrifuge tube milled samples are black closed symbols with a black solid line
for the linear and non-linear model fit (Alpha and Retsch tubes combined); stainless steel jar milled
samples are open circles with a dashed line. Milling time was not significant for percent N (D) so no
line has been fitted.



243

Figure 3. FTIR spectra within the 2800 to 3000 (cm<sup>-1</sup>) wavenumber region; the CH stretching region
diagnostic of atactic-polypropylene. Spectra for 10 mg of *M. caerulea* roots milled in stainless steel
jars (solid black line) and Restch microcentrifuge tube (dashed line) and 60 mg of *M. caerulea* roots
milled in Retsch microcentrifuge (dotted line) compared to a sample of plastic from a Retsch
microcentrifuge tube (dashed and dotted line).



249

250 Figure 4. Carbon concentrations (%) in sandy-loam (A) and organic iron-podzol (B) soil for different

sample sizes milled. Microcentrifuge tube milled samples are dark-grey bars and stainless steel jar

```
252 milled samples are light-grey bars. Mean \pm 1 SE, n = 3.
```

Element	Treatment	Milling method	Predicted line	Treatment <sup>a</sup>	Milling Method	Treatment × milling method
Carbon	Time	Steel	y=45.79+0.04x	$F_{1,46} = 37.61 * * *$	$F_{1,46} = 39.70 * * *$	$F_{1,46} = 6.26*$
		Eppendorf	y=46.31+0.20x			
	Intensity	Steel	y=45.65+0.03x	$F_{1,46} = 47.74 * * *$	$F_{1,46} = 16.44 * * *$	$F_{1,46} = 9.17 * *$
		Eppendorf	y=43.28+0.24x			
	Sample Size	Steel	y=46.00+0.04x	$F_{1,25} = 15.48^{***b}$	$F_{1,25} = 39.56^{***}$	$F_{1,25} = 5.47*$
		Eppendorf	$y = 47.98(1 + \exp(-0.18x))$			
Nitrogen	Time	Steel	y=0.68+0.01x	$F_{1,46}$ = 3.03 ns	$F_{1,46} = 8.49 * *$	$F_{1,46} = 0.41 \text{ ns}$
		Eppendorf	y=0.60+0.01x			
	Intensity	Steel	y=0.6+0.24x	$F_{1,46} = 6.6*$	$F_{1,46} = 9.94 * *$	$F_{1,46}$ = 2.46 ns
		Eppendorf	y=0.79-0.01x			
	Sample size	Steel	y=0.67+0.01x	$F_{1,55} = 8.09 * *$	$F_{1,55} = 10.55 **$	$F_{1,55} = 1.31 \text{ ns}$
		Eppendorf	v = 0.54 + 0.01x			

Table 1. Fitted lines for carbon (C) and nitrogen (N) concentrations of roots milled in microcentrifuge tubes and stainless steel jars for each
 'treatment' (milling time, milling intensity, sample size).

<sup>a</sup>Significance of each factor in the model (treatment, milling method and treatment × milling method interaction) are denoted by ns not

significant, \* p<0.05, \*\* p<0.01, \*\*\* p<0.001.

<sup>b</sup>Statistics relate to the linear model for both stainless steel and microcentrifuge tube combined with sample sizes <26.5mg only, whilst the

259 predicted lines are for the full range of sample sizes.