

Available online at BCREC Website: https://bcrec.undip.ac.id

Bulletin of Chemical Reaction Engineering & Catalysis, 13 (1), 2018, 179-186



Research Article

Comparison of Five Advanced Oxidation Processes for Degradation of Pesticide in Aqueous Solution

Augustine Chioma Affam*1, Malay Chaudhuri2, Shamsul Rahman M. Kutty3

¹Department of Civil Engineering, School of Engineering and Technology, University College of Technology Sarawak, 96000, Sibu, Sarawak, Malaysia ^{2,3}Department of Civil Engineering, Universiti Teknologi PETRONAS, Bandar Seri Iskandar, 31750 Tronoh, Perak, Malaysia

Received: 26th July 2017; Revised: 26nd September 2017; Accepted: 27th September 2017; Available online: 22nd January 2018; Published regularly: 2nd April 2018

Abstract

The study compared the technical efficiency and economic cost of five advanced oxidation processes (Fenton, UV photo-Fenton, solar photo-Fenton, UV/TiO₂/_{H2O2} and FeGAC/H₂O₂) for degradation of the pesticides chlorpyrifos, cypermethrin, and chlorothalonil in aqueous solution. The highest degradation in terms of COD and TOC removals and improvement of the biodegradability (BOD5/COD ratio) index (BI) were observed to be (i) Fenton - 69.03% (COD), 55.61% (TOC), and 0.35 (BI); (ii) UV photo-Fenton - 78.56% (COD), 63.76% (TOC) and 0.38 (BI); (iii) solar photo-Fenton - 74.19% (COD), 58.32% (TOC) and 0.36 (BI); (iv) UV/TiO₂/H₂O₂ - 53.62% (COD), 21.54% (TOC), and 0.26 (BI); and (v) the most technical efficient and cost effective process was FeGAC/H₂O₂. At an optimum condition (FeGAC 5 g/L, H₂O₂ 100 mg/L, and reaction time of 60 min at pH 3), the COD and TOC removal efficiency were 96.19 and 85.60%, respectively, and the biodegradation index was 0.40. The degradation rate constant and cost were 0.0246 min⁻¹ and \$0.74/kg TOC, respectively. The FeGAC/H₂O₂ process is the most technically efficient and cost effective for pretreatment of the pesticide wastewater before biological treatment. Copyright © 2018 BCREC Group. All rights reserved

Keywords: Fenton; UV photo-Fenton; Solar photo-Fenton; UV/TiO₂/H₂O₂; FeGAC/H₂O₂; Pesticide; biodegradability index (BI)

How to Cite: Affam, A.C., Chaudhuri, M., Kutty, S.R.M. (2018). Comparison of Five Advanced Oxidation Processes for Degradation of Pesticide in Aqueous Solution. Bulletin of Chemical Reaction Engineering & Catalysis, 13 (1): 179-186 (doi:10.9767/bcrec.13.1.1394.179-186)

Permalink/DOI: https://doi.org/10.9767/bcrec.13.1.1394.179-186

1. Introduction

Advanced oxidation process (AOPs) are commonly used for pretreatment of recalcitrant (non-biodegradable) solution or wastewater to improve biodegradability [1]. The Fenton process is typically a reaction between iron salts and hydrogen peroxide (H₂O₂) to produce hydroxyl (OH[•]) radicals. This occurs due to the catalytic decomposition of hydrogen peroxide in acidic pH [2]. In the UV photo-Fenton and solar photo Fenton processes, the formation of OH[•] radicals is doubled by photo reactions of H₂O₂ and/or Fe³⁺ producing OH[•] radical either directly or by regeneration Fe²⁺ [3]. The effective UV wavelength is from 100 to about 400 nm, whereas solar wavelength extends from about 280 to above 550 nm. The UV/TiO₂/H₂O₂ photo-

^{*} Corresponding Author. E-mail: augustine@ucts.edu.my, affamskii@yahoo.com (A.C. Affam) Telp: +60103849701

catalytic process is another alternative usually recommended due to the super photocatalytic oxidation potential, optical and electronic attributes, no possession of mass transfer property, operation at ambient conditions, low cost, large band gap, commercial availability. nonhazardous nature and photochemical stability [4-5]. The reaction mechanism of $UV/TiO_2/H_2O_2$ process has been given elsewhere [6]. $FeGAC/H_2O_2$ process was implemented in the degradation of wastewater. It employs the adsorption capacity of GAC with iron oxide coated on its surface. The additional presence of H₂O₂ enables the production of hydroxyl radicals to further increase the degradation efficiency when applied in wastewater treatment. Reuse of the GAC was up to six times [7].

The novelty of this study was the modification of classical Fenton process in order to treat pesticide wastewater. Thus iron sludge was not produced during the process, and operation cost was minimized. This study compared the technical efficiency and cost of five advanced oxidation processes various (Fenton, UV photo-Fenton, solar photo Fenton, TiO_2 photocatalytic (UV/TiO₂/H₂O₂) and FeGAC/H₂O₂) for degradation of a recalcitrant aqueous solution of the pesticides chlorpyrifos, cypermethrin, and chlorothalonil.

2. Materials and Methods

2.1 Chemicals

Hydrogen peroxide (H_2O_2) (30% w/w), ferric nitrate (Fe₃ (NO₃)₃ 9H₂O)) and ferrous sulphate heptahydrate (FeSO₄ ·7H₂O) were supplied by R&M Marketing, Essex, U.K. The TiO₂ powder (anatase, purity >99%) was obtained from Fluka. The pesticides used in the preparation of the aqueous solution were purchased from a local commercial shop. The Calgon Corporation, Pittsburgh, USA supplied the granular activated carbon (GAC) used.

2.2 Fe-Granular activated carbon (FeGAC)

The GAC was blended to reduce it to $425 \,\mu\text{m}$ size. The preparation of the FeGAC was according to a previous study [7].

2.3 Analytical methods

The BOD₅ and COD measurements were done in accordance to known Standard Methods [8]. In an effort to minimize an interference of COD values obtained, pH was raised above 10 immediately the pretreatment (during the FeGAC/H₂O₂ process) was completed. This was to enable H₂O₂ to decompose to O₂ and H₂O [9-

10]. The total organic carbon (TOC) analyzer (Shimadzu) was employed to measure TOC, while pH meter (HACH sension 4) was used in measuring pH during the entire process. Nessler Method (Method 8038) was used to measure ammonia-nitrogen (NH₃-N) [11]. Total phosphorus (TP) was measured by PhosVer 3 Method (Hach, 2002). The YSI 5000 dissolved oxygen meter was used to measure dissolved oxygen (DO). A UV lamp (230V 0.17A, 6 W, Spectronics Corp. NY, USA), with wavelength $\lambda \sim 365$ nm was used. A hand held solar radiometer (model 776E Dodge product, USA) was used for measuring outdoor solar intensity and was subsequently simulated in the reactor. Luzchem Solsim v1.2, (intensity approximately AM 1.5 solar spectrum, 300 W ceramic xenon lamp, λ 280-800 nm and maximum power 1.5 kW/m²) provided simulated solar irradiation.

2.4 Pesticide aqueous solution

Preparation of the pesticide aqueous solution was done using distilled water and was stored at 4 $^{\circ}$ C until required. Its constituents were CPF 100 mg/L, CPT 50 mg/L and CTN 250 mg/L. The COD was 1130.0 mg/L, TOC was 274.39 mg/L, and BOD₅/COD ratio was zero.

2.5 Cost estimation

Cost evaluation for UV incorporated treatment was based on electrical energy per order (EE/O) and calculated using Equation (1) similar to the study by Cañizares *et al.* [13]. This was chosen because the study considered operating cost in terms of electrical energy required by some of the processes examined.

$$EE/O(kWh/m^{3}) = \frac{pt \times 1000}{V \times 60 \log (C_{ini}/C_{fin})}$$
(1)

where, EE/O is the energy requirement (kWh/m³), p is the rated power of the lamp (kW), V is the pesticide aqueous solution volume (L), t is the half-life time in min for achieving 50% reduction of TOC, C_{ini} and C_{fin} are the initial and final concentration of the compound treated in terms of TOC, and the factor of 1000 was for conversion to cubic metre (m³) [12]. Cost of electricity depends on applicable country and this was found to be \$0.10 kWh [13].

2.6 Kinetic study

The pseudo kinetic rate equation for degradation in terms of TOC removal is given by the following nth order reaction kinetics in Equation (2) [14].

$$\frac{-d[TOC]}{dt} = k * TOC^n \tag{2}$$

TOC removal may be described in accordance with the first or second order pseudo models [14]. Linearizing Equation (2) for pseudo first and second order kinetic models respectively gives Equations (3) and (4).

$$\frac{1}{[TOC]_t} - \frac{1}{[TOC]_0} = k * t \tag{3}$$

$$\frac{-d[TOC]}{dt} = k * TOC^n \tag{4}$$

where k is a pseudo-first-order rate constant, t is the irradiation time in min; C_0 the initial concentration of TOC in aqueous solution and C_t is the residual concentration of TOC at time t. The half-life $(t_{1/2})$ is known as the time needed to decrease the concentration of the TOC to half the initial value and it is calculated using Equations (4) and (5) for first and second order, respectively.

$$\frac{\frac{0.693}{k}}{[TOC]_{q * k}}$$
(4)
(5)

The plots of $\ln([TOC]_{0}/[TOC]_{i})$ and $1/[TOC]_{i}$ versus time were done in order to obtain the appropriate reaction order to describe all processes. The values of k from the slopes of the plots at different reaction time and the coefficient of determination values were obtained. The degradation of organic carbon in the pesticide aqueous solution as a function of reaction time was estimated.

2.7 Experimental procedure

In all five processes, except FeGAC/H₂O₂, laboratory study was conducted using 500 mL of the pesticide aqueous solution. The quantity of iron salt needed was added to this solution. Thereafter, an adjustment of the pH of the solution was done using either H₂SO₄ or NaOH. Stirring the solution to appropriate mix was done using a magnetic stirrer to ensure homogeneity. The required amount of H₂O₂ was thereafter added to the solution. In the UV photo-Fenton and TiO₂ photocatalytic processes, the aqueous solution was placed 5 cm under the UV light at room temperature (23 ± 2) °C). In the solar photo-Fenton process, a hand held portable solar radiation meter was used to measure the average daily irradiation and simulated inside the reactor preset to a light intensity of 0.85 kW/m² accordingly. In the UV/TiO₂/H₂O₂, the TiO₂ was placed in a beaker and H_2O_2 was added. The mixture was stirred under UV light at 23±2 °C. The FeGAC/H₂O₂ process was carried putting 200 mL pesticide aqueous solution inside a 250 mL conical flask. The FeGAC and H_2O_2 were mixed together a conical flask and they were put on an orbital shaker. At known time intervals, a conical flask was removed and a portion of the solution was filtered. In the different processes, samples were taken at known intervals of time, and were filtered by a 0.45 µm size filter before measurement of BOD₅, TOC, and COD concentrations. They were also filtered using a 0.20 um size filter so as to determine pesticide concentration by high performance liquid chromatography.

3. Results and Discussion

3.1 Degradation and biodegradability improvement

The degradation and biodegradation (BOD₅/COD ratio) improvement of the pesticide aqueous solution for the five AOPs were studied. A comparison among them was conducted in terms of pesticide degradation and improvement of the biodegradability index under the obtained best operating conditions. The ranges of operating conditions are shown in Table 1. The different operating conditions and their corresponding results are presented in Tables 2 and 3. The reaction conditions which led to best operating conditions can be seen in Table 4.

The best operating conditions for degradation and improvement of biodegradability were: Fenton process - H₂O₂/COD molar ratio 3, H₂O₂/Fe²⁺ molar ratio 10 and ambient pH 3 [15]; UV photo-Fenton process - H₂O₂/COD molar ratio 2, H₂O₂/Fe²⁺ molar ratio 25 and ambient pH 3; solar photo-Fenton process -H₂O₂/COD molar ratio 2, H₂O₂/Fe²⁺ molar ratio 25 and, ambient pH 3; UV/TiO₂/H₂O₂ process - TiO_2 concentration 1.5 g/L, ambient pH 6, H_2O_2 concentration 100 mg/L and reaction time 300 min and FeGAC/H₂O₂ - FeGAC dose 5 g/L, H₂O₂ concentration 100 mg/L and ambient pH 3 (Figure 1). It should be noted that because of the heterogeneous nature of FeGAC/H₂O₂ and UV/TiO₂/H₂O₂ due to the introduction of GAC

Bulletin of Chemical Reaction Engineering & Catalysis, 13 (1), 2018, 182

and UV/TiO₂, respectively, there was no estimation using molar ratios as in the case of other AOP processes, and thus the experimental set up in both cases were uniquely different. The AOPs degraded and improved the biodegradability of the pesticide aqueous solution. The FeGAC/H₂O₂ was observed to be most effective and UV/TiO₂/H₂O₂ was least with a corresponding BOD₅/COD ratio of 0.40 and 0.26, respectively. Biodegradability index between 0.3 and 0.4 is considered a cut-off limit between biodegradable and hard-to-biodegrade

Table 1. Range of values for the AOPs

AOP process	H ₂ O ₂ /COD	H ₂ O ₂ /Fe ₂₊	TiO ₂	FeGAC	H_2O_2
	Molar ratio	Molar ratio	(g/L)	(g/L)	(mg/L)
Fenton UV photo-Fenton Solar photo-Fenton UV/TiO ₂ /H ₂ O ₂ FeGAC/H ₂ O ₂	1.0 - 4.0 1.0 - 3.5 n 1.0 - 3.5	2.0 - 150 5.0 - 150 5.0 - 150	0.5 - 2.5	1.0 - 5.0	50-300 10-300

Labic 1 Operating conditions and performances (1 encon 1 recesses)	Table 2. Op	erating co	onditions a	and	performances	(Fenton	Processes)
---	-------------	------------	-------------	-----	--------------	---------	------------

Process	H ₂ O ₂ /COD	H ₂ O ₂ / Fe ²⁺	pH	Time	COD rem.	TOC rem.	BOD ₅ /COD
	M.R	M.R		(min)	(%)	(%)	ratio
Fenton	1	50	3	60	28.91	17.85	0.17
	1.5	50	3	60	30.26	27.54	0.21
	2	50	3	60	37.31	29.73	0.24
	2.5	50	3	60	46.88	34.40	0.31
	3	50	3	60	50.23	40.26	0.33
	3.5	50	3	60	42.26	32.03	0.17
	4	50	3	60	37.88	30.09	0.14
	3	2	3	60	38.12	34.31	0.17
	3	5	3	60	45.06	47.12	0.18
	3	10	3	60	60.16	52.59	0.34
	3	25	3	60	53.32	43.66	0.32
	3	50	3	60	47.11	37.18	0.25
	3	100	3	60	30.44	31.35	0.17
	3	150	3	60	25.31	23.03	0.14
IIV Photo-Fenton	1	50	3	60	58.98	50.03	0.26
0 1 11010-1 0111011	1.5	50	3	60	62.34	54.45	0.28
	2	50	3	60	77.82	61.15	0.37
	2.5	50	3	60	53.24	43.27	0.25
	3	50	3	60	47.37	40.23	0.23
	3.5	50	3	60	35.71	35.09	0.22
	2	5	3	60	55.26	46.15	0.26
	2	10	3	60	70.75	54.29	0.34
	2	25	3	60	78.56	63.76	0.37
	2	50	3	60	74.98	57.13	0.33
	2	100	3	60	62.25	54.34	0.32
	2	150	3	60	57.96	48.13	0.31
	1	50	3	60	39.03	25.15	0.25
Color abote Forster	1.5	50	3	60	51.35	34.27	0.33
Solar-photo renton	2	50	3	60	72.08	57.56	0.33
	2.5	50	3	60	54.16	39.96	0.29
	3	50	3	60	40.17	33.27	0.23
	3.5	50	3	60	35.93	28.38	0.19
	2	25	3	60	32.41	27.33	0.19
	2	25^{-5}	3	60	48.39	33.27	0.24
	2	$\frac{-6}{25}$	3	60	61.47	47.87	0.31
	2	$\frac{-6}{25}$	3	60	74.19	58.32	0.35
	2	25	3	60	41.86	38.85	0.24
	2	25	3	60	23.80	30.84	0.19

Bulletin of Chemical Reaction Engineering & Catalysis, 13 (1), 2018, 183

effluent [16]. Degradation of a mixture of dimethoate, oxydemeton-methyl, carbaryl and methidathion pesticides by AOP treatment obtained similar results [17].

3.2 Kinetics of pesticide degradation and half-life time

The pesticide degradation kinetics and halflife time for the substrates by the five AOPs was studied. The kinetics of pesticide degradation is a vital pre-requisite for predicting the rate of substrates degradation before the design of typical wastewater treatment plants so as to optimize and lower the operating cost [18]. Several studies have reported pesticide degradation according to pseudo- first order [19-20]. To study the kinetics, the various AOPs were carried out using the operating conditions in Table 4.

The degradation of organic carbon in the pesticide aqueous solution as a function of reaction time is shown in Figure 2 for the five AOPs. The summary of rate constants (k), R2 and half life time are shown in Table 5. Estimation of the first and second order reactions were done to ascertain which one describes the

Table 3. Experimental operating conditions and results of UV/TiO₂/H₂O₂ and FeGAC/H₂O₂ processes

Process	H_2O_2	TiO ₂ (g/L)	FeGAC (g/L)	pН	Time (min)	COD rem. (%)	TOC rem. (%)	BOD5/COD ratio
UV/TiO ₂ /H ₂ O ₂	-	0.5		6	300	15.35	6.13	-
	-	1.0		6	300	20.76	7.38	-
	-	1.5		6	300	25.95	8.45	-
	-	2.0		6	300	22.51	7.48	-
	-	2.5		6	300	19.43	5.21	-
	50	1.5		6	300	32.69	17.79	0.22
	100	1.5		6	300	53.62	21.54	0.26
	150	1.5		6	300	42.18	18.56	0.24
	200	1.5		6	300	37.43	16.99	0.21
	300	1.5		6	300	33.39	13.52	0.17
			1			56.90	41.38	-
FeGAC/H ₂ O ₂	-		2			60.71	46.91	-
	-		3			67.96	50.15	-
	-		4			83.01	68.17	-
	-		5			85.75	72.15	
	10		5			51.68	30.05	0.29
	25		5			59.12	36.87	0.34
	50		5			91.86	70.55	0.38
	100		5			96.19	79.15	0.40
	150		5			85.93	62.71	0.35
	200		5			76.81	58.36	0.32
	300		5			74.16	54.09	0.29

Table 4. COD and TOC removal (%) and optimum operating conditions of the different AOPs processesstudied

Treatment process	COD Rem. (%)	TOC Rem. (%)	BOD ₅ / COD ratio	H ₂ O ₂ /COD molar ratio	H ₂ O ₂ /Fe ²⁺ molar ratio	TiO ₂ /H ₂ O ₂ mg/L	FeGAC/H ₂ O ₂ mg/L
Fenton	69.03	55.61	0.35	3.0	10		
UV photo-Fenton	78.56	63.76	0.38	2.0	25		
Solar photo-Fenton	74.19	58.32	0.36		25	1.5 g/L/100	
UV/TiO ₂ /H ₂ O ₂	53.62	21.54	0.26	2.0			
FeGAC/H ₂ O ₂	96.19	85.60	0.40				5 g/L/100

¹Complete degradation of pesticides occurred in 1 min in all treatment processes

² UV lamp 6 W, (λ 365nm)

Copyright © 2018, BCREC, ISSN 1978-2993

process appropriately. The reactions were found to follow pseudo-first order kinetics with rate constants and R^2 values shown in Table 5. The rate constant of UV/TiO₂/H₂O₂ and $FeGAC/H_2O_2$ was observed to be the lowest and highest, respectively. It is pertinent to mention that the highest rate constant may be due to increased OH. production rate and adsorption/catalytic properties of the FeGAC. GAC is known to decompose H_2O_2 . This occurs during the exchange of a surface hydroxyl group with hydrogen peroxide anion. The formed surface peroxide is known to have an increased oxidation potential and this provides the opportunity for another hydrogen peroxide molecule with release of oxygen and regeneration of the GAC surface [21]. This result confirms the significant effect of the five AOPs, but in particular FeGAC/H₂O₂ process capability in degrading of the pesticide in aqueous solution.



3.3 Cost requirement

The cost requirement for the five processes is one of the most important factors to consider in choosing an appropriate AOP treatment process. The overall costs are usually the sum of capital costs, operating cost and maintenance costs. For a full-scale system these costs strongly depend on the nature and the concentration of the pollutants, the flow rate of the effluent and the configuration of the reactor [13]. Efforts to develop standard procedures to estimate UV lamp electrical consumption has been reported [22]. The average prices of reagents were obtained from related study (Table 6) [12,

Table 5. AOPs kinetic constants and half-life



Figure 1. Degradation and biodegradability improvement by the various AOPs (a) Fenton, (b) UV photo Fenton, (c) Solar photo Fenton, (d) UV/TiO₂/H₂O₂, (e) FeGAC/H₂O₂ (degradation data presented here was obtained under best condition)

Copyright © 2018, BCREC, ISSN 1978-2993

22-23]. Table 7 shows an estimation of the cost of operation per kg of TOC and this was calculated for the mineralization of 50% (half-life) of the initial TOC. As regards the operating costs, $UV/TiO_2/H_2O_2$ photocatalytic process was considerably the most expensive than the others owing to the nominal power of the UV lamp and the time required for pesticide degradation. However, costs could be considerably reduced when solar light is used [19]. The FeGAC/H₂O₂ process was observed to be the most attractive option for the pesticide degradation.

4. Conclusions

The five AOPs (Fenton, UV photo-Fenton, solar photo-Fenton, $UV/TiO_2/H_2O_2$, and FeGAC/H₂O₂) were all effective for degradation and improvement of the biodegradability of the pesticides aqueous solution. The technical efficiency and cost requirements of each process under the best operating conditions indicate

Table 6. Price of reagent

Reagent	Unit	Price (\$)
H_2O_2 [22]	kg	0.35
$FeSO_{4}.7H_{2}O$ [22]	kg	0.5
TiO_{2} [22]	kg	3.0
GAC [12]	g	0.001
Fe(NO ₃) ₃ . 9H ₂ O [23]	g	0.3

Table 7. Cost requirement for the studied AOPs

that UV/TiO₂/H₂O₂ was the least attractive option, whereas FeGAC/H₂O₂ was the preferred alternative. In addition, the highest rate constant obtained in FeGAC/H₂O₂ process could be due to the available adsorption sites on the FeGAC surface and catalytic property of the FeGAC and hence an increased OH. production. The study found FeGAC/H₂O₂ process to be cost effective in comparison to the other AOPs.

Acknowledgement

The authors are indebted to Universiti Teknologi PETRONAS for the provision of research facilities during this study.

References

- Oller, I., Malato, S., Sánchez-Pérez, J.A. (2010). Review: Combination of advanced oxidation processes and biological treatments for wastewater decontamination - A review, *Sci*ence of The Total Environment, 409(20): 4141-4166 (doi:10.1016/j.scitotenv.2010.08.061)
- Chamarro, E., Marco, A., Esplugas, S. (2001).
 Use of Fenton reagent to improve organic biodegradability, *Water Research*, 35 (4)1047-1051 (doi:10.1016/S0043-1354(00)00342-0)
- Pignatello, J.J., Liu, D., Houston, P. (1999)
 Evidence for an additional oxidant in the photo assisted Fenton reaction, *Environmental Science & Technology* 33(11): 1832-1839. (doi:10.1021/es980969b)

	Reagent	Fenton	UV-photo- Fenton	Solar photo- Fenton	$UV/TiO_2 H_2O_2$	FeGAC/H ₂ O ₂
Chemical re-	H_2O_2	3602	2401	2401	100	100
quirement (mg/L)	FeSO4·7H2O TiO2 Fe(NO3)3·9H2O GAC	2945	785	785	1500	$\begin{array}{c} 200\\ 5000 \end{array}$
Cost estima- tion (\$/kg TOC)	H ₂ O ₂ FeSO ₄ ·7H ₂ O TiO ₂ Fe(NO ₃) ₃ · 9H ₂ O GAC	$18.4 \\ 21.5$	24.4 5.7	24.4 5.7	2.2 3.8	0.26 0.44 0.036
Cost as per electricity EE/O (kWh/m ³)	UV	*4.0			*233	
Total cost \$/Kg TOC (half-life time)		39.9	34.1	30.1	239	0.74

Bulletin of Chemical Reaction Engineering & Catalysis, 13 (1), 2018, 186

- [4] Lorret, O., Francová, D., Waldner, G., Stelzer Stelzer, N. (2009). W-doped titania nanoparticles for UV and visible-light photocatalytic reactions. *Appl. Catal. B* 91(1-2): 39-46 (doi:10.1016/j.apcatb.2009.05.005
- [5] Pekakis, P.A., Xekoukoulotakis, N.P., Mantzavinos, D. (2006). Treatment of textile dyehouse wastewater by TiO₂ photocatalysis, Water Research, 40(6): 1276-1286. (doi:10.1016/j.watres.2006.01.019)
- [6] Sadik, W.A., Nashed, A.W., El-Demerdash, A.G.M. (2007). Photodecolourization of ponceau 4R by heterogeneous photocatalysis, *Journal of Photochemical & Photobiology* A. 189(1):135-140
- Fan, H., Shua, H., Tajima, K. (2006). Decolorization of acid black 24 by the FeGAC/H₂O₂ process, *Journal of Hazardous Materials*, 128(2-3): 192-200
- [8] APHA, AWWA, WPCF, Standard Methods for the Examination of Water and Wastewater, 21st ed. American Public Health Association, Washington, DC, 2005.
- Talinli, I., Anderson, G.K. (1992). Interference of hydrogen peroxide on the standard COD test, Water Research, 26(1): 107-110 (doi:10.1016/0043-1354(92)90118-N)
- [10] Kang, Y.W., Cho, M.J. Hwang, K.Y. (1999). Correction of hydrogen peroxide interference on standard chemical oxygen demand test, *Water Research*, 33(5): 1247-1251 (doi:10.1016/S0043-1354(98)00315-7)
- [11] Hach. (2002). *Water Analysis Handbook* (4th Edition.) Loveland, CO: Hach Company.
- [12] Lu, C., Liu, C., Rao, G. P. (2008) Comparisons of sorbent cost for the removal of Ni2+ from aqueous solution by carbon nanotubes and granular activated carbon, *Journal of Hazardous Materials*, 151(1): 239-246 (doi:10.1016/j.jhazmat.2007.05.078)
- [13] Cañizares, P., Paz, R., Sáez, C., Rodrigo, M.A. (2009). Costs of the electrochemical oxidation of wastewaters: A comparison with ozonation and Fenton oxidation processes, *Journal of Environmental Management*, 90(1): 410-420 (doi:10.1016/j.jenvman.2007.10.010)
- [14] Körbahti, B., Artut, K. (2010). Electrochemical oil/water demulsification and purification of bilge water using Pt/Ir electrodes. *Desalination* 258: 219-228.

- [15] Affam, A.C., Chaudhuri, M., Kutty, S.R.M.(2012). Fenton treatment of combined Chlorpyrifos, cypermethrin and chlorothalonil pesticides in aqueous solution, *Journal of Environmental Science Technology*, 5(6): 407-418 (doi:10.3923/jest.2012.407.418)
- [16] Abderrazik, N.B., Al Momani, F., Sans, C., Esplugas S. (2002). Combined advanced oxidation with biological treatment. *Afinidad -Barcelona* 59(498): 141-146.
- [17] Martín M.M.B., Pérez, J.A.S., López, J.LC. Oller, I., Rodríguez, S.M.M. (2009). Degradation of a four-pesticide mixture by combined photo-Fenton and biological oxidation, *Water* $R \ e \ s \ e \ a \ r \ c \ h$. 4 3 (3): 6 5 3 - 6 6 0. (doi:10.1016/j.watres.2008.11.020)
- [18] Scott, J.P., Ollis, D.F. (1995). Integration of chemical and biological oxidation processes for water treatment: *Review and recommendation Environmental Progress & Sustainable Energy*, 14(2): 88-103 (doi: 10.1002/ep.670140212)
- Shen, Y.S., Ku, Y., Lee, K.C. (1995). The effect of light absorbance on the decomposition of chlorophenols by ultraviolet radiation and UV/H₂O₂ processes, *Water Research*, 29 (3) 907-914 (doi:10.1016/0043-1354(94)00198-G)
- [20] Zhang, Y., Xiao, Z., Chen, F., Ge, Y., Wu, J., Hu, X.(2010). Degradation behavior and products of malathion and chlorpyrifos spiked in apple juice by ultrasonic treatment, *Ultrasonics Sonochemistry*, 17(1): 72-77. (doi: 10.1016/j.ultsonch.2009.06.003)
- [21] Bansal, R.C., Donnet, J.B., Stoeckli, F. (1998). Active Carbon. New York: Marcel Dekker.
- [22] Andreozzi, R., Caprio, V., Insola, A., Marotta, R. (1999). Advanced oxidation processes (AOP) for water purification and recovery, *Catalysis Today*. 53(1): 51-59 doi:10.1016/S0920-5861(99)00102-9
- [23] Integra Chemical, http://www.integrachem.com/product_catalog _search.asp assessed on 1st January, 2013.