

Intraspecific phenotypic variation in life history traits of Daphnia galeata populations in response to fish kairomones

Verena Tams, Jennifer Lüneburg, Laura Seddar, Jan-Phillip Detampel and Mathilde Cordellier

Institut für Zoologie, Universität Hamburg, Hamburg, Germany

ABSTRACT

Phenotypic plasticity is the ability of a genotype to produce different phenotypes depending on the environment. It has an influence on the adaptive potential to environmental change and the capability to adapt locally. Adaptation to environmental change happens at the population level, thereby contributing to genotypic and phenotypic variation within a species. Predation is an important ecological factor structuring communities and maintaining species diversity. Prey developed different strategies to reduce their vulnerability to predators by changing their behaviour, their morphology or their life history. Predator-induced life history responses in Daphnia have been investigated for decades, but intra-and inter-population variability was rarely addressed explicitly. We addressed this issue by conducting a common garden experiment with 24 clonal lines of European Daphnia galeata originating from four populations, each represented by six clonal lines. We recorded life history traits in the absence and presence of fish kairomones. Additionally, we looked at the shape of experimental individuals by conducting a geometric morphometric analysis, thus assessing predator-induced morphometric changes. Our data revealed high intraspecific phenotypic variation within and between four D. galeata populations, the potential to locally adapt to a vertebrate predator regime as well as an effect of the fish kairomones on morphology of D. galeata.

Subjects Ecology, Evolutionary Studies, Freshwater Biology Keywords Phenotypic plasticity, Life history traits, Population ecology, Predator-induced response,

Daphnia, Adaptive potential

INTRODUCTION

Intraspecific phenotypic variation is crucial for the persistence of a population, since low intra-population variation increases the risk of extinction (Bolnick et al., 2011; Scheiner & Holt, 2012; Forsman, 2014). Loss of phenotypic variation can be caused by the reduction of genetic variation, for example, due to genetic drift (random loss of alleles) (Vanoverbeke & De Meester, 2010; Bolnick et al., 2011), inbreeding depression (Lynch, 1991; Swillen, Vanoverbeke & De Meester, 2015) or positive selection (Biswas & Akey, 2006). On the contrary, phenotypic variation can increase as a consequence of environmental change (biotic and/or abiotic) as well as through an

Submitted 8 March 2018 Accepted 14 September 2018 Published 17 October 2018

Corresponding author Mathilde Cordellier, mathilde.cordellier@uni-hamburg.de

Academic editor Valery Forbes

Additional Information and Declarations can be found on page 19

DOI 10.7717/peerj.5746

(cc) Copyright 2018 Tams et al.

Distributed under Creative Commons CC-BY 4.0

OPEN ACCESS

increase in genetic variation, which in turn occurs through gene flow (migration), mutation and recombination (*Griffiths et al.*, 2000). Phenotypic variation 'is the fuel that feeds evolutionary change' because natural selection acts on it (*Stearns*, 1989). Phenotypic plasticity describes the ability of genotypes to produce different phenotypes depending on the environment, helping organisms to survive and reproduce in heterogeneous environment (*Stearns*, 1989; *Agrawal*, 2001). Phenotypic plasticity implies an adaptive potential to locally adapt to a changed environment (*Stearns*, 1989). If the phenotypically plastic organism produces a modified and successful phenotype whose fitness (higher reproductive success) is higher than an unmodified phenotype, the underlying genotype contributes more to the genetic make-up of the whole population.

Predation structures whole communities (*Werner & Peacor*, 2003; *Beschta & Ripple*, 2009; *Boaden & Kingsford*, 2015; *Aldana et al.*, 2016), drives natural selection within populations (*Morgans & Ord*, 2013; *Kuchta & Svensson*, 2014) and maintains species diversity (*Estes et al.*, 2011; *Fine*, 2015). Aquatic predators release chemical substances, so called kairomones, into the surrounding waters which can be detected by their prey. Both vertebrates (*Schoeppner & Relyea*, 2009; *Stibor*, 1992) and invertebrates (*Machacek*, 1991; *Stibor & Lüning*, 1994) release kairomones, triggering specific phenotypic plastic responses such as morphological or behavioural changes (*Dodson*, 1989; *Schoeppner & Relyea*, 2009). The predator-induced defences can be highly variable within a species, depending on factors such as the predator and colonization histories (*Eklöv & Svanbäck*, 2006; *Kishida*, *Trussell & Nishimura*, 2007; *Edgell & Neufeld*, 2008).

Invertebrate as well as vertebrate predator kairomones have been shown to cause phenotypic plastic responses in *Daphnia*. These induced responses are predator specific and vary across Daphnia species. Behavioural changes such as diel vertical migration (Effertz & Von Elert, 2015) and the associated metabolic costs (Dawidowicz & Loose, 1992), aggregation and escape response (Pijanowska, 1997), reduction of ingestion and filtration rates (Rose, Warne & Lim, 2003; Beckerman, Wieski & Baird, 2007; Pestana, Baird & Soares, 2013), depth selection (Cousyn et al., 2001), increased alertness (Boersma, Spaak & De Meester, 1998) and diapause (production of resting eggs = ephippia) (Pijanowska & Stolpe, 1996) were reported for different daphnids exposed to vertebrate predator kairomones (fish). Diverse morphological changes have been shown to occur in the presence of kairomones of the invertebrate predator *Chaoborus*, such as the production of neck teeth in D. pulex (Lüning, 1995; Tollrian, 1995) or the famous helmets of D. longispina (Brett, 1992) and D. cucullata (Agrawal, Laforsch & Tollrian, 1999). Recently Herzog et al. (2016) observed a remarkable morphological change of D. barbata exposed to Triops kairomones. D. barbata changes its whole body symmetry to an S-shape, presumably to impede ingestion by their invertebrate predator. Apart from morphology, physiology and behaviour, predator kairomones were also shown to influence life history traits in different *Daphnia* species. Among others, size and fecundity, two important traits for population survival, were affected, resulting in earlier maturation (Riessen, 1999; De Meester & Weider, 1999; Weber, 2003; Castro, Consciência & Gonçalves, 2007) and smaller size (Stibor & Lüning, 1994; Castro, Consciência & Gonçalves, 2007). Size is a very important factor for survival in the face of fish predation, since small individuals are more likely to go undetected. These predator-induced responses are the result of phenotypic plasticity and their magnitude might play a role in adaptation.

The ability of Daphnia to locally adapt to different stressors has been demonstrated before, for example, for fish as a vertebrate predator (Boersma, Spaak & De Meester, 1998; Cousyn et al., 2001; Declerck & Weber, 2003) and pesticides (Jansen et al., 2011). Differences in phenotypic responses between Daphnia populations can further happen because of the high genetic divergence between lakes, which is turn due to specific colonization patterns in Daphnia (De Meester et al., 2002), monopolization effects, and overall lack of gene flow. Hence, we expect the investigated populations to exhibit population specific responses. Clonal variation of Daphnia has been regularly reported (Machacek, 1991; Castro, Consciência & Gonçalves, 2007; Beckerman, Rodgers & Dennis, 2010). Some investigators evaluated clonal variation within one population (Cousyn et al., 2001; De Meester & Weider, 1999) or among populations of Daphnia (Boersma, Spaak & De Meester, 1998; Declerck & Weber, 2003; Boeing, Ramcharan & Riessen, 2006; Hamrová, Mergeay & Petrusek, 2011; Lind et al., 2015). Others rarely used more than one or two clonal lines per population, drawing conclusions based on single clonal lines. Although intra-population variation or lack thereof is relevant to population maintenance in the face of predation pressure, the relative importance of the intra- and inter-population variation was rarely measured. We are aware of only a few studies which addressed this aspect of predator-induced responses in Daphnia. Boersma, Spaak & De Meester (1998) used four clonal lines of D. magna for each of the four population showing that the strength and combination of responsive traits can differ across genotypes (clonal lines). The clonal variation in D. pulex exposed to vertebrate (fish) and invertebrate (Chaoborus) predator kairomones was assessed for migration behaviour (Boeing, Ramcharan & Riessen, 2006) and life history traits (Lind et al., 2015) at the inter- and intra-population level. Recently, Reger et al. (2018) showed that predation drives local adaptation in phenotypic plasticity in 70 clonal lines of D. pulex.

In the present study, we assess the intraspecific phenotypic variation of European *D. galeata* in the presence of fish kairomones, by measuring shifts in life history traits as well as morphological changes in a total of 24 clonal lines (six clonal lines within each of the four populations). We expect that (i) there is intraspecific phenotypic variation, evidenced by inter-clonal variation within as well as among populations. We hypothesize that (ii) local adaptation and population divergence caused by drift and limited gene flow shape the observed phenotypic variation, and is evidenced by population specific responses. Finally, we expect that (iii) the exposure to fish kairomones affects the morphology in *D. galeata*. We hypothesize that (iv) a correlation between life history change and morphology exists. Specifically, we hypothesize that females which increased their total number of offspring in the presence of fish kairomones, change their morphology towards a bulkier shape to accommodate more eggs.

Table 1 Background information of ecological aspects of the four European lakes of which experimental clonal lines originate from.						
Lake	Greifensee	Jordan Reservoir	Lake Constance	Müggelsee		
Abbreviation	popG	popJ	popLC	popM		
Location	Switzerland	Czech Republic	Austria, Germany, Switzerland	Germany		
GPS coordinates	47°21′20″N, 8°40′10″E	49°24′55″N, 14°39′49″E	47°37′21″N, 9°26′24″E	52°26′6″N, 13°38′6″E		
N	6	6	6	6		
Alt. (m)	435	437	395	32		
Vol. (km³)	0.1485	0.0027	48	0.0366		
Dep. (m)	34	14	254	8		
Av. Dep. (m)	18	4.5	90	4.9		
Stratification	Dimictic	Dimictic	Monomictic	Polymictic		
Fish biomass (kg/ha)	32	607.5	54	70-100		
Presence of Leuciscus sp.	Yes	Yes	Yes	Yes		

Note:

MATERIALS AND METHODS

Experimental organisms and lakes of origin

This study integrated 24 *D. galeata* clonal lines from four different locations: Lake Constance (popLC), Germany; Greifensee (popG), Switzerland; Müggelsee (popM), Germany and Jordan Reservoir (popJ), Czech Republic. These are all permanent lakes with a large water body and varying densities of planktivorous fish (Table 1).

Clonal lines were established from dormant eggs from sediment cores and have been used in previous studies (*Henning-Lucass et al., 2016*; *Herrmann et al., 2017*).

These ephippia originate from sediment layers of the recent decade (2000–2010) and were subjected to hatching stimuli described in detail in *Henning-Lucass et al. (2016*).

The clonal lines were maintained in lab cultures (18 °C, 16 h light/8 h dark cycle, food: *Acutodesmus obliquus*, medium: Aachener Daphnien Medium (ADaM) (*Klüttgen et al., 1994*)) for up to 5 years and no less than 3 years prior to the present experiment.

Fish densities and hence predation pressures are not easy to measure very precisely due to seasonal variation (*Fischer & Eckmann, 1997*), variation between years (*Kruuk, Conroy & Moorhouse, 1991*) as well as differences between lakes (*Mehner et al., 2005*). We provide here estimates of fish biomass available from the literature corresponding as closely as possible to the time period the resting eggs came from. The highest estimate of fish biomass was for the Jordan Reservoir, which is stocked for recreational purposes and the lowest for Greifensee (Table 1). Fish belonging to the *Leuciscus* genus were found in all lakes from which our clonal lines originate (*Philipp, 2006*; *Kubecka & Bohm, 1991*; *Alexander et al., 2016*; *Fischereiamt Berlin, 2013*).

Media preparation

The basic medium was ADaM for fish and *Daphnia* cultures. Two types of media were used for breeding and experimental conditions: fish kairomone and control medium. In total forty ide (*Leuciscus idus*) were maintained in an aerated, separate 200 L aquarium,

N, Number of genotypes; Alt., Altitude; Vol., Volume; Dep., Maximum depth; Av. Dep., Average depth.

in which they were fed frozen Daphnia cubes and dry food. The ide or closely related species are present in all the studied lakes (Table 1). Previous studies showed that ide elicit plastic responses in D. galeata clonal lines from Lake Constance (Sakwinska, 2002) and Greifensee (Wolinska, Löffler & Spaak, 2007). Fish medium was obtained by keeping five randomly chosen ide in an aerated 20 L aquarium for 24 h to produce fish kairomone medium. The fish were not fed in the fish medium production tank to avoid Daphnia alarm cues to be mixed with the fish kairomones. The fish kairomone media imitates a scenario of high fish density (Cousyn et al., 2001; Swillen, Vanoverbeke & De Meester, 2015). Control medium was produced in an aerated, separated aquarium and handled first, before handling of fish and fish medium. All media was filtered before use to remove faeces from predators and bacteria larger than 1.2 μm (Whatman, membrane filters, ME28, Mixed cellulose-ester, 1.2 µm). All media were supplemented with 1.0 mg C L-1, P rich Acutodesmus obliquus before use and exchanged daily (1:2) to guarantee a nutrient rich environment and a constant fish kairomone concentration. The algae concentration was calculated from photometric measurement of the absorbance rate at 800 nm.

Because fish was used to produce fish kairomone media, this experiment was subject to approval through the 'Behörde für Gesundheit und Verbraucherschutz' of the City of Hamburg. It was approved under the number 75/15.

Experimental design and procedures: life table experiment

Prior to the experiment, each clonal line was bred in kairomone-free water (control environment) and in kairomone water (fish environment) for two subsequent generations to minimize inter-individual variances. To this end, 10-15 egg-bearing females per clonal line were randomly selected from mass cultures. From these females of unknown age, neonates were collected and raised under experimental conditions and served as grandmothers (F0) for the experimental animals (F2). Neonates of the third to fifth brood carried by the F0 animals were used as breeding (F1) animals. Neonates of the third to fifth brood carried by the F1 animals were used in turn as experimental individuals (F2). A pair of neonates was introduced in the experimental vessels (50 mL glass tube) at the start of the experiment to compensate for eventual mortality. One of the individuals was randomly discarded when necessary at day 4 (t4), so that one individual remained in each vessel. This procedure was applied to F1 and F2 individuals. Fifteen replicates were used per environment and per genotype (clonal line). Sister neonates of F2 (n = 15) were collected in 70% ethanol for size measurements at day 0 (t0). Life history parameters were recorded daily during the experiment. Before media renewal, females were checked for maturation and neonates were counted, removed and preserved in ethanol every day. Adults were preserved in ethanol as well at the end of the experiment. The experiment lasted for 14 days (t14) for each experimental individual to monitor the performance of each clonal line within a fixed period of time.

Cetyl alcohol was used to break the surface tension of the media during breeding and the experiment to reduce juvenile mortality (*Desmarais*, 1997). Breeding and

experimental phases were conducted at a temperature of 20 °C and a 16 h light/8 h dark cycle in a brood chamber with a light intensity of 30% (Rumed, Type 3201D).

The experiment was conducted in three experimental rounds due to logistic reasons, and clonal lines from all four populations were included in each round (Table S1). Previous pilot studies showed that ensuring synchronicity of so many clonal lines at once were extremely difficult.

Data collection and analysis

Life history traits

Life history parameters such as age at first reproduction ('AFR') [d], number of neonates per brood per female, total number of broods per female ('broods'), total numbers of neonates per female ('offspring'), size of first clutch ('brood1') [number of neonates per female], 'survival' [%] and somatic growth rate ('SGR') [μm d⁻¹] were recorded. 'AFR' was the day of releasing the first brood from the brood pouch, with neonates swimming in the vessel. For further analysis the average value of the 15 individuals per clonal line (genotype) per environment (treatment) was calculated for each life history trait to estimate the clonal response to a kairomone (fish) vs. kairomone-free (control) environment. Survival rate was defined as the proportion of females surviving from the day of separation (t4) until the end of the experiment (t14). Reproductive rate was calculated by dividing the total number of offspring per female by the total number of broods per female. Relative fitness (w) was calculated by multiplying survival and reproductive rate of a clonal line before dividing by the maximum survival and reproductive rate of the other clonal lines within population ('relnest') and among all populations ('relclone'). Some clonal lines produced male offspring during breeding and the experiment. Males occurred at very low frequencies and were excluded from the data analysis. We aimed to test a total of 720 individuals in this experiment (24 clonal lines \times 2 environments × 15 replicates). In total we measured life history traits for 684 experimental individuals (Table S1).

Digitizing of experimental animals for 'size' and 'shape' analysis

Digital photographs of *Daphnia* preserved in ethanol were taken with a stereomicroscope (Nikon SMZ800N) at a magnification of $60\times$ for neonates (t0) and $40\times$ for adults (t14) with NIS-elements 4.3 software. All experimental individuals were photographed in lateral view (left body side up).

Measurement of body length ('size')

Body length was measured from the top of the head through the middle of the eye to the ventral basis of the spine, excluding the spine itself. 'SGR' (μ m/day) was calculated by subtracting the average length of neonates at the beginning of the experiment (t0; n=15) from the length of each adult individual at the end of the experiment (t14), divided by the complete experimental time in days. The measurement error of digitizing and measuring the length of the same individual 10 times was +/-3.24 μ m (SD).

The measurement error of measuring 10 times the length of an individual using the exact same picture was $\pm 1.67 \, \mu m$ (SD).

Geometric morphometric analysis of the 'shape' of the body

Since the morphology of *Daphnia* does not allow the assignment of many landmarks, we decided to use the semilandmark approach. Semilandmarks are a set of individual landmarks which are interpolated to represent the curve of a structure (Zelditch et al., 2004). Landmarks and semilandmarks were assigned on a subset of digital images of adult experimental individuals (max. n = 10 per clonal line and environment, with a total of 459 individuals; Table S1) according to Zelditch et al. (2004). In total three landmarks and 115 semilandmarks were assigned on each individual photograph. The first landmark was appointed to the tip of the rostrum, the second in the middle of the eye and the third at the ventral basis of the spine. In our study the first curve consisted of 70 interpolated landmarks (=semilandmarks) along the dorsal body outline, starting at the first landmark and ending on the dorsal basis of the spine. The second set of semilandmarks consisted of 45 semilandmarks along the ventral body outline, starting at landmark three and ending opposite of the dorsal basis of antenna. After the assignment of landmarks and semilandmarks, X and Y coordinates were recorded using TpsDig2 (Rohlf, 2015). A General Procrustes analysis (GPA) was performed using the package 'geomorph' in R (Adams & Otárola-Castillo, 2013). The measurement variance for assigning landmarks and semilandmarks of an individual using the exact same picture was <0.0001. Investigators of 'shape' measurements worked with a blind data set, not knowing which individual belongs to which group (environment, genotype and population).

Statistical analysis

All statistical analyses for life history traits were performed and all figures were created using R version 3.3.1 (*R Core Team, 2016*). For the generalized linear mixed models the package '*lme4*' was used (*Bates et al., 2015*). Visualization of life history traits were performed by using the package '*geplot2*' (*Wickham, 2016*). For the geometric morphometric analysis the package '*geomorph*' was used (*Adams & Otárola-Castillo, 2013*). The visualization of shape differences was performed with the R package '*shapes*' (*Dryden, 2017*). R scripts are provided in Supplementary Materials.

To compare life history traits between the different populations in the presence and absence of fish kairomones, we applied generalized linear mixed effect models for each trait, except 'shape.' Visual inspection of residual plots as well as the Shapiro–Wilk-test revealed deviations from homoscedasticity for each trait, supporting the decision to use nonparametric models for statistical analysis. Hence, error distributions were assigned individually per trait. We used 'Environment \times Population' and 'Genotype \times Environment' as fixed categorical factors in our models. To account for genotype differences among populations, we included 'Clone' nested within 'Population' as a random factor. To account for experimental rounds, we added 'round' as a random factor. We checked for the necessity of random slopes and intercepts by the application of likelihood ratio tests of the

full model with the effect in question against the model without the effect in question, finally resulting in a general random slope and intercept model for 'Environment \times Population' (response \sim treatment * pop + (treatment | pop/clone) + (1 | round)) and 'Genotype \times Environment' (response \sim treatment * clone + (1 | round)). Statistical significances for life history traits were obtained by using the function Anova (model, type = 2)) which performs a Wald Chi-Square test.

To assess shape variation we used the principal component analysis after the GPA in the R package 'geomorph.' Subsequently the statistical analysis was done with Procrustes ANOVA and pairwise tests to reveal statistically relevant 'shape' differences between groups.

RESULTS

Effect of genotype origin on predator-induced response in life history traits: 'Environment \times Population' effect

Reaction norms of life history traits per population (Fig. 1) as well as their boxplots (Fig. 2) show intra- and inter-population variation of life history traits. The statistical analysis revealed that the factor 'Environment' affected the life history traits 'AFR' and 'broods' in a significant manner. The factor 'Population' affected 'offspring,' 'brood1,' 'relnest' and 'relclone,' but there was no significant interaction effect of 'Environment × Population' for any life history trait (Table 2A). The visualization of growth differences between environments and populations differences of somatic growth rate (dSGR, Fig. 3) showed that all clonal lines from popG had a negative growth rate in the fish-exposed environment, resulting in a smaller body size. Four out of six clonal lines from popJ had a negative somatic growth rate, while clonal lines from popLC and popM vary in somatic growth rates across environments.

The fittest population in control environment was popJ (w = 1), followed by popM (w = 0.83), popLC (w = 0.78) and popG (w = 0.67). In fish environment a small change of positions occurred for popLC and popM. Here the decreasing order was popJ (w = 1), followed by popLC (w = 0.80), popM (w = 0.77) and popG (w = 0.63) among all populations. Further details of relative fitness for each clonal line within their population can be found in Tables 3A and 3B.

Effect of 'Genotype \times Environment' interaction on life history traits

In the model comprising the factors 'Genotype' and 'Environment,' most of the traits were significantly affected by both factors as well as their interaction: 'AFR,' 'broods,' 'offspring,' 'SGR' and 'size' (Table 2B). The traits 'brood1,' 'relnest' and 'relclone' were only affected by the factor 'Genotype' (Table 2B). Reaction norms for each life history trait of each clonal line can be found in the Supplementary Material (Figs. S2–S7).

Effect of fish kairomones on the morphological trait 'shape'

A total of 83% of 'shape' variation was explained by the first four principal components (PC1 = 42%, PC2 = 24%, PC3 = 11% and PC4 = 6%) (Fig. S1).

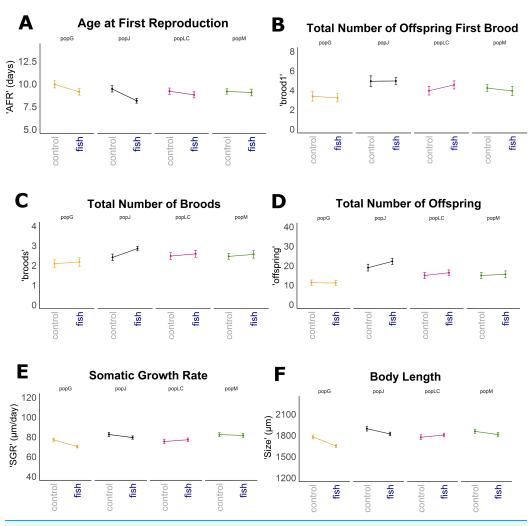


Figure 1 Reaction norms for selected life history traits showing population differences (mean +/-SE). Population Greifensee (popG, yellow), population Jordan Reservoir (popJ, black), population Lake Constance (popLC, magenta) and population Müggelsee (popM, green). (A) Age at first reproduction ('AFR'). (B) Total number of offspring first brood ('brood1'). (C) Total number of broods ('broods'). (D) Total number of offspring ('offspring'). (E) Somatic growth rate ('SGR'). (F) Body length ('size').

Full-size DOI: 10.7717/peerj.5746/fig-1

The geometric morphometric analysis revealed a significant 'Environment' effect on 'shape' (Df = 455, F = 5.93, Pr(>)F = 0.001). Differences in shape occurred within each population except for popG (Table 4A) as well as among populations (Table 4B). There was a significant interaction effect of 'Environment × Population' on 'shape' (Df = 3, F = 1.9163, Pr(>)F = 0.019). The p-value matrix revealed a statistical significance difference within popLC between environments (p = 0.035; Table 4C). There was a statistical significant 'Genotype × Environment' effect on 'shape' (Df = 411, F = 3.2947, Pr(>)F = 0.047). Experimental 'round' was included in the geometric morphometric analysis revealing its statistical significant effects on the morphological trait 'shape.'

The shape of females with many offspring (n > 22 = upper quartile of total number of offspring) differed significantly among populations in the control environment, but not in

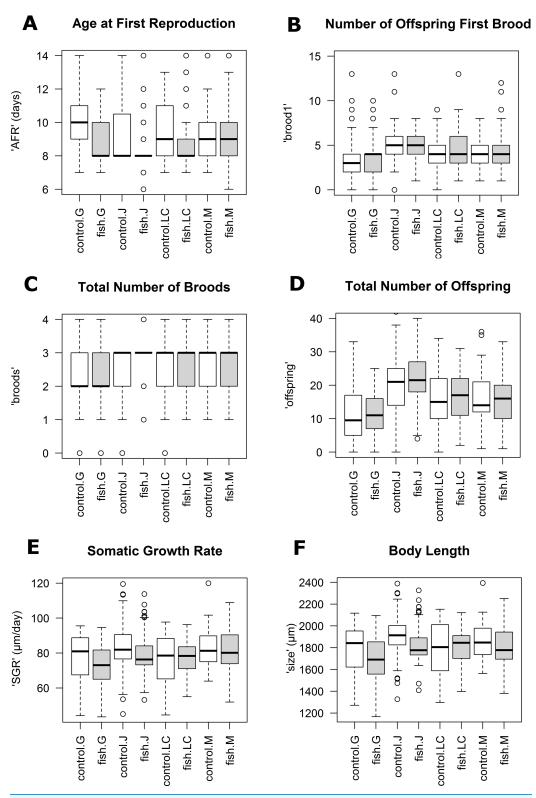


Figure 2 Boxplots for selected life history traits showing population differences (median +/- SD). (A) Age at first reproduction ('AFR'). (B) Total number of offspring first brood ('brood1'). (C) Total number of broods ('broods'). (D) Total number of offspring ('offspring'). (E) Somatic growth rate ('SGR'). (F) Body length ('size').

Full-size DOI: 10.7717/peerj.5746/fig-2

Table 2 Results of general linear mixed effect models (GLMM) on various life history traits.

(A) The effect of individual origin ('Environment × Population').

		$\underline{\textbf{Environment} \times \textbf{Population}}$		opulation
Life history trait		Chi-square	Df	Pr(>Chi-square)
Age at first reproduction ('AFR')	Environment	5.0261	1	0.0250
	Population	3.5870	3	0.3097
	Environment \times Population	1.9912	3	0.5742
Total number of broods ('broods')	Environment	3.9718	1	0.0463
	Population	7.5636	3	0.0560
	Environment × Population	1.4309	3	0.6983
Total number of offspring	Environment	1.5044	1	0.2200
('offspring')	Population	17.1803	3	0.0006
	$Environment \times Population$	1.9182	3	0.5860
Total number of offspring first	Environment	0.0065	1	0.9358
brood ('brood1')	Population	10.3740	3	0.0156
	Environment × Population	0.6623	3	0.8031
'Survival'	Environment	0.2127	1	0.6447
	Population	0.2403	3	0.9708
	$Environment \times Population$	0.0000	3	1
Relative fitness within	Environment	0.2608	1	0.6096
populations ('relnest')	Population	9.8864	3	0.0196
	Environment × Population	0.1236	3	0.9889
Relative fitness among	Environment	0.3751	1	0.5402
populations ('relclone')	Population	13.6158	3	0.0035
	$Environment \times Population$	0.3250	3	0.9553
Somatic growth rate ('SGR')	Environment	3.3442	1	0.0674
	Population	5.2943	3	0.1514
	Environment × Population	5.7855	3	0.1225
Body length ('size')	Environment	3.5277	1	0.0604
	Population	2.1413	3	0.5436
	Environment × Population	5.2355	3	0.1553

(B) The effect of 'Genotype \times Environment' interaction.

	Genotype × Environment		onment
	Chi-square	$\mathbf{D}f$	Pr(>Chi-square)
Environment	41	1	<0.001
Genotype	253	23	<0.001
$Genotype \times Environment$	146	23	<0.001
Environment	11	1	<0.001
Genotype	114	23	<0.001
Genotype \times Environment	64	23	<0.001
Environment	5	1	<0.001
Genotype	988	23	<0.001
$Genotype \times Environment$	175	23	<0.001
	Genotype Genotype × Environment Environment Genotype Genotype × Environment Environment Genotype	Environment 41 Genotype 253 Genotype × Environment 146 Environment 11 Genotype 114 Genotype × Environment 64 Environment 5 Genotype 988	$ \begin{array}{c ccccccccccccccccccccccccccccccccccc$

(Continued)

Table 2 (continued).

(B) The effect of 'Genotype \times Environment' interaction.

		$\textcolor{red}{\textbf{Genotype}} \times \\$	Enviro	onment
Life history trait		Chi-square	Df	Pr(>Chi-square)
Total number of offspring first	Environment	0.0267	1	0.870
brood ('brood1')	Genotype	116	23	<0.001
	Genotype \times Environment	34	23	0.060
'Survival'	Environment	3	1	0.0944
	Genotype	28	23	0.1991
	Genotype \times Environment	8	23	0.9987
Relative fitness within	Environment	2.6	1	0.1076
populations ('relnest')	Genotype	302	23	<0.001
	$Genotype \times Environment$	NA	NA	NA
Relative fitness among	Environment	2.6	1	0.1041
populations ('relclone')	Genotype	235	23	<0.001
	Genotype \times Environment	NA	NA	NA
Somatic growth rate ('SGR')	Environment	20	1	<0.001
	Genotype	1,289	23	<0.001
	Genotype \times Environment	147	23	<0.001
Body length ('size')	Environment	50	1	<0.001
	Genotype	1,613	23	<0.001
	Genotype \times Environment	179	23	<0.001

Note

Significant values (p < 0.05) are highlighted in bold. Values are rounded.

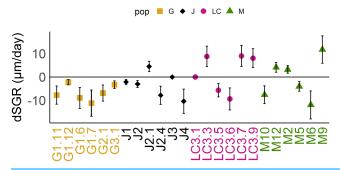


Figure 3 Differences of somatic growth rate (dSGR). Differences of somatic growth rate (dSGR) as µm per day (mean +/- SD); calculated as: mean of SGR (fish) minus mean SGR (control) equals dSGR per clonal line, sorted by populations. Each clonal line is displayed in its population specific colour. Population Greifensee (popG, yellow), population Jordan Reservoir (popJ, black), population Lake Constance (popLC, magenta) and population Müggelsee (popM, green).

Full-size DOI: 10.7717/peerj.5746/fig-3

the fish environment. There was no association between 'shape' and a high number of 'offspring' in the fish environment. Further analysis revealed that the 'shape' of females with many offspring did not differ significantly between environments within each population.

Visualization revealed an overall 'shape' change towards a smaller body. In detail, a homogenous change from all directions to a smaller body form was found for popG

Table 3 Relative fitness (w) within and among populations.

(A) Relative fitness within populations for genotype means.

		w within popu	lation ('relnest')	w among popu	w among populations ('relclone')		
Population	Clone	Control	Fish	Control	Fish		
G	G1.11	0.53	0.84	0.36	0.50		
	G1.12	0.35	0.66	0.24	0.40		
	G1.6	0.46	0.31	0.32	0.19		
	G1.7	0.95	0.86	0.65	0.51		
	G2.1	0.81	0.86	0.56	0.52		
	G3.1	1.00	1.00	0.69	0.60		
J	J1	0.73	0.75	0.73	0.75		
	J2	0.64	0.68	0.64	0.68		
	J2.1	0.50	0.69	0.50	0.69		
	J2.4	1.00	1.00	1.00	1.00		
	J3	0.67	0.70	0.67	0.70		
	J4	0.63	0.55	0.63	0.55		
LC	LC3.1	0.73	0.59	0.55	0.45		
	LC3.3	0.56	0.63	0.42	0.47		
	LC3.5	0.78	0.96	0.59	0.72		
	LC3.6	1.00	1.00	0.75	0.75		
	LC3.7	0.46	0.54	0.35	0.41		
	LC3.9	0.78	0.95	0.59	0.71		
M	M10	0.72	0.97	0.43	0.66		
	M12	0.87	0.71	0.52	0.48		
	M2	0.98	0.86	0.59	0.59		
	M5	0.95	1.00	0.57	0.69		
	M6	0.82	0.88	0.50	0.60		
	M9	1.00	0.78	0.60	0.54		

(B) Range of relative fitness among populations for genotype means.

Population	w control	w fish
G	0.24-0.69	0.19-0.60
J	0.50-1.00	0.55-1.00
LC	0.35-0.75	0.41-0.75
M	0.43-0.60	0.54-0.69

Note:

Fittest genotype or population (w = 1.0) is highlighted in bold.

(Fig. 4B). Within popJ the overall shape change towards a smaller body size was shown with the strongest change in the head area (bending of the thin plate spline) and an anterior–posterior direction (Fig. 4C). Within popLC the head position changed from dorsal to ventral direction, while a small change of the tail area from a ventral to dorsal direction (Fig. 4D) occurred. Within popM the overall shape change towards a smaller body size was shown in the head area from a dorsal to ventral direction and in the tail area from a ventral to dorsal direction (Fig. 4E).

Table 4 Results of geometric morphometric analysis.

(A) p-values of 'Environment' effect on 'shape' differences within populations.

Population	Df	F	Pr (> <i>F</i>)
G	1	0.39	0.843
J	1	5.87	0.001
LC	1	3.79	0.007
M	1	3.16	0.009

(B) p-value Matrix of 'Environment' effect on 'shape' among populations.

-	G	J	LC	M
G	_	0.013	0.167	0.013
J	0.013	_	0.143	0.001
LC	0.167	0.143	_	0.003
M	0.013	0.001	0.003	_

(C) p-value Matrix of the interaction of 'Environment × Population' on 'shape.'

-	G:control	G:fish	J:control	J:fish	LC:control	LC:fish	M:control	M:fish
G:control	-	0.848	0.589	0.364	0.137	0.473	0.420	0.932
G:fish	0.848	_	0.913	0.671	0.470	0.454	0.418	0.421
J:control	0.589	0.913	_	0.534	0.731	0.352	0.629	0.614
J:fish	0.364	0.671	0.534	_	0.172	0.290	0.892	0.290
LC:control	0.137	0.470	0.731	0.172	-	0.035	0.103	0.248
LC:fish	0.473	0.454	0.352	0.290	0.035	_	0.743	0.876
M:control	0.420	0.418	0.629	0.892	0.103	0.743	_	0.399
M:fish	0.932	0.421	0.614	0.290	0.248	0.876	0.399	_

Note:

Statistical significant F-values (Pr(>F) < 0.05) are displayed in bold.

DISCUSSION

Intraspecific phenotypic variation of life history traits in D. galeata

Our study revealed a significant 'Environment,' 'Genotype' as well as 'Genotype × Environment' effect for the life history traits 'AFR,' 'broods,' 'offspring,' 'SGR' and 'size.' Concordant to previous studies (*Boersma, Spaak & De Meester, 1998*; *Stibor & Lüning, 1994*), our results showed a decrease of 'AFR', a decrease of 'SGR' and a decrease of body length in the presence of fish kairomones in *D. galeata*. Out of the 24 studied genotypes, 13 matured earlier (Fig. S2) and 17 reduced their body length (Fig. S7) in the presence of fish kairomones. Indeed, early maturation and a reduced size of *Daphnia* in the presence of vertebrate predators have been reported before (*Machacek, 1991*; *Weider & Pijanowska, 1993*; *Lampert, 1993*; *Gliwicz & Boavida, 1996*). The ecological benefit lies in a successful reproduction before reaching a body size making the individual vulnerable to fish predation (*Lynch, 1980*; *Lampert, 1993*).

The 'Genotype \times Environment' effect for most of the life history traits implies that the presence or absence of certain clonal lines within one population might have an effect on overall population survival, depending on environmental factors such as predation risk. Hence, if the phenotypic diversity within one population is reduced and the

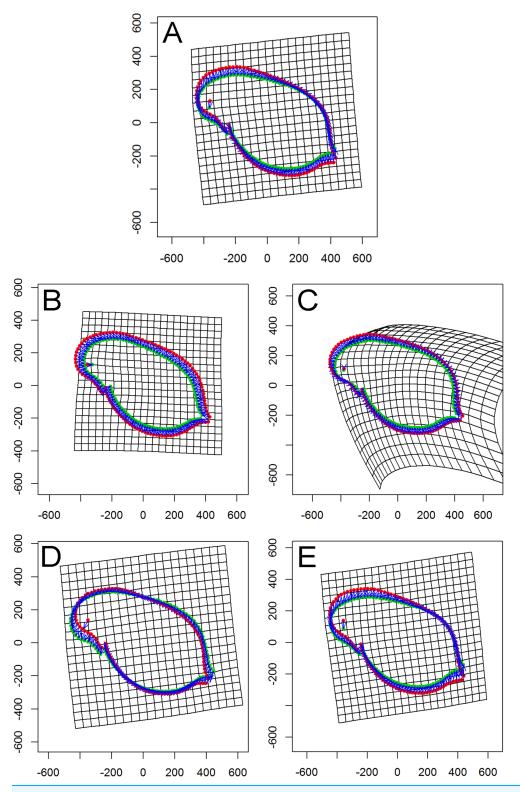


Figure 4 Thin plate spline (TPS) grids of consensus shapes of superimposed Procrustes coordinates. Control (red). Fish (green). (A) All specimens. (B) Population Greifensee (popG). (C) Population Jordan Reservoir (popJ). (D) Population Lake Constance (popLC). (E) Population Müggelsee (popM).

Full-size DOI: 10.7717/peerj.5746/fig-4

majority produces relatively less offspring in a fish environment, the result could be an overall low number of offspring in the following cohorts, which would threaten the persistence of the whole population. Notably, individuals of popG produced less offspring and less broods compared to the other three populations regardless of the environment and their relative fitness was comparatively low. Potential explanations for this relative low performance of popG could be genetic drift and inbreeding depression which have a negative effect on genetic diversity (*Vanoverbeke & De Meester*, 2010). However, low genetic variation for *D. galeata* in Greifensee was not identified (*Herrmann et al.*, 2017), making these two explanations unlikely at first glance. On the contrary, most clonal lines in Greifensee (four out of six) had a higher heterozygosity than expected, perhaps as a result of hybridization, which is known to occur in this population (*Brede et al.*, 2009). Therefore, outbreeding depression could explain lower fitness in popG and should be further investigated in a future study.

Our experimental design allowed us to assess the distribution of variance at clonal as well as population level. We thus detected phenotypic variation within each as well as among several populations independent of the environment. We identified two different strategies of phenotypic plastic responses of D. galeata by comparing the environmental effect within as well as among the populations. In popJ, the variation of a trait itself, not the change in the trait median value as a response was extremely reduced for two life history traits, 'AFR' and total number of 'broods' (Fig. 2C). Almost all individuals of popJ started to reproduce at the same age and produce the same amount of broods in fish environment, showing a striking homogeneity under stress. On the contrary, in popM the variation for 'AFR' increased, resulting in a broader range of 'AFR' in fish environment. The phenotypic variation between clonal lines was best visualized by plotting the dSGR between the environments (Fig. 3), unifying the environmental and genotype effect. All six clonal lines of popG and four out of six clonal lines of popJ decreased their somatic growth in fish environment, while the direction of response varied for popLC and popM. Overall our study with a total of 24 clonal lines revealed a broad spectrum of phenotypic variation of life history traits in European D. galeata.

To our surprise we did not find an 'Environment × Population' effect on the life history response, although we observed intra- and inter-population differences, especially between the two extremes popG and popJ. A significant 'Population' effect was found for the interdependent traits total number of offspring ('offspring'), size of first brood ('brood1'), relative fitness within ('relnest') and among populations ('relclone').

This observed significant population divergence could be due to the extreme difference of total number of offspring between popG and popJ. In general, genotypes in popJ produced the highest number of offspring among all populations. In contrast, the total number of offspring of genotypes in popG was overall lower compared to the other three populations, regardless of the environment. Even the increased number of offspring for clonal lines of popG in fish environment is lower than the numbers of offspring for clonal lines of popJ in control environment. Hence, the genotype origin ('Population') itself had little to no effect on life history traits in *Daphnia* implying that the identity of a clonal line ('Genotype') within a population seems to be more important than the origin of the clonal line per se.

In the end, we were not able to identify one main driving force influencing the phenotypic variation of life history traits in *D. galeata*, and could not infer a population specific response. Instead our study displays the complexity of the interacting factors 'Environment' and 'Genotype' to produce a variety of phenotypes within one species, thereby contributing to the understanding of intraspecific phenotypic variation.

Potential for local adaptation

Our findings allow the conclusion that there is a potential for local adaptation to predation risk in the investigated European populations of *D. galeata*. This conclusion was based on two outcomes of our study.

On one hand, we observed an extreme predator-induced life history response for popJ. The range of variation of the phenotypic response was reduced to a minimum in popJ, so that almost all individuals of the six clonal lines and 15 replicates reproduce at the very same age when exposed to fish (Fig. 2A). On top of that, we observed a similar reduction of variation for the life history trait total number of broods (Fig. 2C). These strong responses could be explained by local adaptation to the presence of fish. The Jordan Reservoir is an artificial inner city water reservoir, used for recreational purposes such as fishing since 1900 (Kubecka & Bohm, 1991) and had been regularly stocked with fish (Seda, Hejzlar & Kubecka, 2000). Therefore, D. galeata of Jordan Reservoir had the possibility to adapt to an environment with a higher predation risk for more than a century. Such micro-evolutionary changes for Daphnia species have been described in other contexts before. For instance, Jansen et al. (2011) showed that D. magna was able to evolve resistance to a pesticide (carbaryl) within experimental time. Further, Declerck, Cousyn & De Meester (2001) as well as Reger et al. (2018) showed that populations were able to locally adapt to fish kairomones (D. galeata and D. pulex, respectively). Alternatively, since the reservoir, unlike the other lakes in this study, has been created specifically with fishing in mind, differential colonization might also be the source of the observed pattern. This habitat might have been colonized only by Daphnia pre-adapted to fish, with very specific life-histories, leading to the present-day striking pattern.

On the other hand, the relative fitness of individuals of popJ suggests that females exposed to fish kairomones are fitter, concurring with results obtained by *Castro, Consciência & Gonçalves* (2007) and *Jansen et al.* (2011). Since local adaptation to a certain stressor implies a better performance in the stressful environment than without this stressor (*Lenormand et al.*, 1999; *Joshi et al.*, 2001) we suggest that the local adaptive potential exists for at least three populations because the relative fitness in the presence of fish kairomones increased overall for 14 out of 24 clonal lines (popG = 3, popJ = 4, popLC = 4, popM = 3) (Table 3A). Our results are in line with earlier studies showing the adaptive potential of phenotypic plasticity in *Daphnia* exposed to different stressors (*Yin et al.*, 2011; *Altshuler et al.*, 2011; *Reger et al.*, 2018; *Hesse et al.*, 2012).

Predation risk and morphological changes

In general, we did not observe any predator-induced extreme morphological changes such as the formation of helmets for fish kairomone exposed *Daphnia* as those reported for

D. lumholtzi (Laforsch & Tollrian, 2004). We presented here the first study using the geometric morphometric analysis to measure morphometric changes to an environmental factor in D. galeata, hence complementing the traditional approaches (life history traits and behaviour). Our morphometric analysis revealed that the presence of fish kairomones had an effect on the body shape of D. galeata. However, no overall pattern was recognizable among the populations and no effect was observed at all for popG. Instead we observed different changes of 'shape' in each population. We suggest that the morphological trait 'shape' is phenotypically plastic due to high clonal variation, which is consistent with the results reported by Dlouhá et al. (2010) and Zuykova, Bochkarev & Katokhin (2012). The difference between experimental rounds for 'shape' could be attributed to this observed high clonal variation which we observed in all life history traits as well.

We hypothesized that life history change and morphological change are correlated, meaning that females with a higher number of offspring (n > 22, upper quartile of observed total number of offspring) would change their 'shape' towards a bulkier body form to accommodate a greater number of offspring within their brood pouch. This correlation was found only in individuals in control environment and not for individuals in fish environment. Changing shape of the body might come along with some drawbacks: the bulkier the shape, the higher the detection risk by the predator and the slower the swimming ability due to drag. In fact, fish prey size-selectively on Daphnia, meaning that larger Daphnia are preyed upon more often than smaller Daphnia (Weber & Van Noordwijk, 2002; Beckerman, Rodgers & Dennis, 2010). Since fish prey on faster swimming individuals of Daphnia (O'Keefe, Brewer & Dodson, 1998), being a slow swimming Daphnia would be beneficial. Alternatively, accommodating more offspring without changing the shape of the body might be achieved through the production of smaller offspring, as was shown by Castro, Consciência & Gonçalves (2007), among others (Lampert, 1993). In line with previous studies showing a predator-induced reduction in neonate size, we can speculate that this is also the case in our experiment and plan to further explore this dimension.

CONCLUSIONS

The study presented here focused on the assessment of intraspecific phenotypic variation in *D. galeata*. By comparing the range of phenotypic response of 24 clonal lines originating from four populations, we contribute to the understanding of the effect of environmental change (shifts in predator regime) on intraspecific phenotypic variation at the population level. We observed high clonal variation in all studied life history traits at the intra- and inter-population level, leading to the suggestion that single genotype studies on *Daphnia* might deliver biased conclusions.

ACKNOWLEDGEMENTS

We thank Jens Oldeland, Bob O'Hara and Suda Parimala Ravindran for valuable statistical advice. Additionally, we thank Michael Engelmohn, Tatjana Usinger and Anne Ehring for their help during *Daphnia* breeding and the experiment. We would like

to thank Lisa Gottschlich for testing and confirming geometric morphometric measurements and Thomas Mehner for his input on fish density calculation. Earlier versions of this manuscript greatly benefited from the comments of three anonymous reviewers.

ADDITIONAL INFORMATION AND DECLARATIONS

Funding

This work was supported by the Volkswagen Foundation (Grant No. 86030). The funders had no role in study design, data collection and analysis, decision to publish, or preparation of the manuscript.

Grant Disclosures

The following grant information was disclosed by the authors: The Volkswagen Foundation: 86030.

Competing Interests

The authors declare that they have no competing interests.

Author Contributions

- Verena Tams conceived and designed the experiments, performed the experiments, analysed the data, prepared figures and/or tables, authored or reviewed drafts of the paper, approved the final draft.
- Jennifer Lüneburg performed the experiments, approved the final draft.
- Laura Seddar performed the experiments, approved the final draft.
- Jan-Phillip Detampel analysed the data, prepared figures and/or tables, approved the final draft.
- Mathilde Cordellier conceived and designed the experiments, contributed reagents/materials/analysis tools, authored or reviewed drafts of the paper, approved the final draft.

Animal Ethics

The following information was supplied relating to ethical approvals (i.e., approving body and any reference numbers):

Animal handling and experiments were in accordance with the ethical standards (approved for the execution of experiments on vertebrates No: 75/15).

Data Availability

The following information was supplied regarding data availability: The raw data are provided in the Supplemental Files.

Supplemental Information

Supplemental information for this article can be found online at http://dx.doi.org/10.7717/peerj.5746#supplemental-information.

REFERENCES

- **Adams DC, Otárola-Castillo E. 2013.** geomorph: an R package for the collection and analysis of geometric morphometric shape sata. *Methods in Ecology and Evolution* **4(4)**:393–399 DOI 10.1111/2041-210X.12035.
- **Agrawal AA. 2001.** Phenotypic plasticity in the interactions and evolution of species. *Science* **294**:321–326 DOI 10.1126/science.1060701.
- **Agrawal AA, Laforsch C, Tollrian R. 1999.** Transgenerational induction of defences in animals and plants. *Nature* **401(6748)**:60–63 DOI 10.1038/43425.
- Aldana M, Maturana D, Pulgar J, García-Huidobro MR. 2016. Predation and anthropogenic impact on community structure of boulder beaches. *Scientia Marina* 80(4):543–551 DOI 10.3989/scimar.04444.27A.
- Alexander TJ, Vonlanthen P, Périat G, Raymond JC, Degiorgi F, Seehausen O. 2016. Artenvielfalt und Zusammensetzung der Fischpopulation im Bodensee. Kastanienbaum: Projet Lac, Eawag.
- Altshuler I, Demiri B, Xu S, Constantin A, Yan ND, Cristescu ME. 2011. An integrated multi-disciplinary approach for studying multiple stressors in freshwater ecosystems: *Daphnia* as a model organism. *Integrative and Comparative Biology* 51(4):623–633 DOI 10.1093/icb/icr103.
- Bates D, Mächler M, Bolker B, Walker S. 2015. Fitting linear mixed-effects models using lme4. *Journal of Statistical Software* 67(1):1–48 DOI 10.18637/jss.v067.i01.
- Beckerman AP, Rodgers GM, Dennis SR. 2010. The reaction norm of size and age at maturity under multiple predator risk. *Journal of Animal Ecology* **79**(5):1069–1076 DOI 10.1111/j.1365-2656.2010.01703.x.
- Beckerman AP, Wieski K, Baird DJ. 2007. Behavioural versus physiological mediation of life history under predation risk. *Oecologia* 152(2):335–343 DOI 10.1007/s00442-006-0642-6.
- **Beschta RL, Ripple WJ. 2009.** Large predators and trophic cascades in terrestrial ecosystems of the Western United States. *Biological Conservation* **142(11)**:2401–2414 DOI 10.1016/j.biocon.2009.06.015.
- **Biswas S, Akey JM. 2006.** Genomic insights into positive selection. *Trends in Genetics* **22(8)**:437–446 DOI 10.1016/j.tig.2006.06.005.
- **Boaden AE, Kingsford MJ. 2015.** Predators drive community structure in coral reef fish assemblages. *Ecosphere* **6(4)**:1–33 DOI 10.1890/ES14-00292.1.
- **Boeing WJ, Ramcharan CW, Riessen HP. 2006.** Clonal variation in depth distribution of *Daphnia pulex* in response to predator kairomones. *Archiv Für Hydrobiologie* **166(2)**:241–260 DOI 10.1127/0003-9136/2006/0166-0241.
- Boersma M, Spaak P, De Meester L. 1998. Predator-mediated plasticity in morphology, life history, and behavior of *Daphnia*: the uncoupling of responses. *American Naturalist* 152(2):237–248 DOI 10.1086/286164.
- Bolnick DI, Amarasekare P, Araújo MS, Bürger R, Levine JM, Novak M, Rudolf VHW, Schreiber SJ, Urban MC, Vasseur DA. 2011. Why intraspecific trait variation matters in community ecology. *Trends in Ecology & Evolution* 26(4):183–192 DOI 10.1016/j.tree.2011.01.009.
- Brede N, Sandrock C, Straile D, Spaak P, Jankowski T, Streit B, Schwenk K. 2009. The impact of human-made ecological changes on the genetic architecture of *Daphnia* species. *Proceedings of the National Academy of Sciences of the United States of America* 106(12):4758–4763 DOI 10.1073/pnas.0807187106.
- **Brett MT. 1992.** *Chaoborus* and fish-mediated influences on *Daphnia longispina* population structure, dynamics and life history strategies. *Oecologia* **89(1)**:69–77 DOI 10.1007/BF00319017.

- **Castro BB, Consciência S, Gonçalves F. 2007.** Life history responses of *Daphnia longispina* to mosquitofish (*Gambusia holbrooki*) and pumpkinseed (*Lepomis gibbosus*) kairomones. *Hydrobiologia* **594(1)**:165–174 DOI 10.1007/s10750-007-9074-5.
- Cousyn C, De Meester L, Colbourne JK, Brendonck L, Verschuren D, Volckaert F. 2001.

 Rapid, local adaptation of zooplankton behavior to changes in predation pressure in the absence of neutral genetic changes. *Proceedings of the National Academy of Sciences of the United States of America* 98(11):6256–6260 DOI 10.1073/pnas.111606798.
- **Dawidowicz P, Loose CJ. 1992.** Metabolic costs during predator-induced diel vertical migration of *Daphnia. Limnology and Oceanography* **37(8)**:1589–1595 DOI 10.4319/lo.1992.37.8.1589.
- Declerck S, Cousyn C, De Meester L. 2001. Evidence for local adaptation in neighbouring *Daphnia* populations: a laboratory transplant experiment. *Freshwater Biology* **46(2)**:187–198 DOI 10.1046/j.1365-2427.2001.00632.x.
- **Declerck S, Weber A. 2003.** Genetic differentiation in life history between *Daphnia galeata* populations: an adaptation to local predation regimes? *Journal of Plankton Research* **25(1)**:93–102 DOI 10.1093/plankt/25.1.93.
- De Meester L, Gómez A, Okamura B, Schwenk K. 2002. The monopolization hypothesis in dispersal-gene flow paradox in aquatic organisms. *Acta Ocologia* 23(3):121–135 DOI 10.1016/s1146-609x(02)01145-1.
- **De Meester L, Weider LJ. 1999.** Depth selection behavior, fish kairomones, and the life histories of *Daphnia hyalina* × *galeata* hybrid clones. *Limnology and Oceanography* **44(5)**:1248–1258 DOI 10.4319/lo.1999.44.5.1248.
- **Desmarais KH. 1997.** Keeping *Daphnia* out of the surface film with cetyl alcohol. *Journal of Plankton Research* **19(1)**:149–154 DOI 10.1093/plankt/19.1.149.
- **Dlouhá S, Thielsch A, Kraus RHS, Seda J, Schwenk K, Petrusek A. 2010.** Identifying hybridizing taxa within the *Daphnia longispina* species complex: a comparison of genetic methods and phenotypic approaches. *Hydrobiologia* **643(1)**:107–122 DOI 10.1007/s10750-010-0128-8.
- **Dodson SI. 1989.** The ecological role of chemical stimuli for the zooplankton: predator-avoidance behaviour in *Daphnia*. *Limnology and Oceanography* **33(6part2)**:1431–1439 DOI 10.4319/lo.1988.33.6part2.1431.
- **Dryden IL. 2017.** *shapes: Statistical Shape Analysis.* R package version 1.2.3. *Available at https://CRAN.R-project.org/package=shapes.*
- **Edgell TC, Neufeld CJ. 2008.** Experimental evidence for latent developmental plasticity: intertidal whelks respond to a native but not an introduced predator. *Biology Letters* **4(4)**:385–387 DOI 10.1098/rsbl.2008.0204.
- **Effertz C, Von Elert E. 2015.** Coupling of anti-predator defences in *Daphnia*: the importance of light. *Hydrobiologia* **798(1)**:5–13 DOI 10.1007/s10750-015-2387-x.
- **Eklöv P, Svanbäck R. 2006.** Predation risk influences adaptive morphological variation in fish populations. *American Naturalist* **167(3)**:440–452 DOI 10.1086/499544.
- Estes JA, Terborgh J, Brashares JS, Power ME, Berger J, Bond WJ, Carpenter SR, Essington TE, Holt RD, Jackson JBC, Marquis RJ, Oksanen L, Oksanen T, Paine RT, Pikitch EK, Ripple WJ, Sandin SA, Scheffer M, Schoener TW, Shurin JB, Sinclair ARE, Soulé ME, Virtanen R, Wardle DA. 2011. Trophic downgrading of planet earth. *Science* 333(6040):301–306 DOI 10.1126/science.1205106.
- **Fine PVA. 2015.** Ecological and evolutionary drivers of geographic variation in species diversity. *Annual Review of Ecology, Evolution, and Systematics* **46(1)**:369–392 DOI 10.1146/annurev-ecolsys-112414-054102.

- **Fischer P, Eckmann R. 1997.** Seasonal changes in fish abundance, biomass and species richness in the littoral zone of a large European lake, Lake Constance, Germany. *Archiv für Hydrobiologie* **139(4)**:433–448.
- **Fischereiamt Berlin. 2013.** Fische in Berlin—Bilanz der Artenvielfalt. Eds: Senatsverwaltung für Stadtentwicklung und Umwelt, Fischereiamt Berlin.
- Forsman A. 2014. Effects of genotypic and phenotypic variation on establishment are important for conservation, invasion, and infection biology. *Proceedings of the National Academy of Sciences of the United States of America* 111(1):302–307 DOI 10.1073/pnas.1317745111.
- **Gliwicz ZM, Boavida MJ. 1996.** Clutch size and body size at first reproduction in *Daphnia pulicaria* at different levels of food and predation. *Journal of Plankton Research* **18(6)**:863–880 DOI 10.1093/plankt/18.6.863.
- Griffiths AJF, Miller JH, Suzuki DT, Lewontin RC, Gelbart WM. 2000. An Introduction to Genetic Analysis. Seventh Edition. New York: W. H. Freeman.
- Hamrová E, Mergeay J, Petrusek A. 2011. Strong differences in the clonal variation of two *Daphnia* species from mountain lakes affected by overwintering strategy. *BMC Evolutionary Biology* 11(1):231 DOI 10.1186/1471-2148-11-231.
- Henning-Lucass N, Cordellier M, Streit B, Schwenk K. 2016. Phenotypic plasticity in life-history traits of *Daphnia galeata* in response to temperature—a comparison across clonal lineages separated in time. *Ecology and Evolution* 6(4):881–891 DOI 10.1002/ece3.1924.
- Herrmann M, Henning-Lucass N, Cordellier M, Schwenk K. 2017. A genotype-phenotype association approach to reveal thermal adaptation in *Daphnia galeata*. *Journal of Experimental Zoology Part A: Ecological and Integrative Physiology* 327(1):53–65 DOI 10.1002/jez.2070.
- Herzog Q, Rabus M, Wolfschoon Ribeiro B, Laforsch C. 2016. Inducible defenses with a 'twist': *Daphnia barbata* abandons bilateral symmetry in response to an ancient predator. *PLOS ONE* 11(2):e0148556 DOI 10.1371/journal.pone.0148556.
- Hesse O, Engelbrecht W, Laforsch C, Wolinska J. 2012. Fighting parasites and predators: how to deal with multiple threats? *BMC Ecology* 12(1):12 DOI 10.1186/1472-6785-12-12.
- Jansen M, Coors A, Stoks R, De Meester L. 2011. Evolutionary ecotoxicology of pesticide resistance: a case study in *Daphnia*. *Ecotoxicology* 20(3):543–551 DOI 10.1007/s10646-011-0627-z.
- Joshi J, Schmid B, Caldeira MC, Dimittrakopoulos PG, Good J, Harris R, Hector A, Huss-Danell K, Jumpponen A, Minns A, Mulder CP, Pereira JS, Prinz A, Scherer-Lorenzen M, Siamantziouras AD, Terry AC, Troumbis AY, Lawton JH. 2001. Local adaptation enhances performancs of common plant species. *Ecology Letters* 4(6):536–544 DOI 10.1046/j.1461-0248.2001.00262.x.
- **Kishida O, Trussell GC, Nishimura K. 2007.** Geographic variation in a predator-induced defense and its genetic basis. *Ecology* **88(8)**:1948–1954 DOI 10.1890/07-0132.1.
- Klüttgen B, Dulmer U, Engels M, Ratte HT. 1994. ADaM, an artificial freshwater for the culture of zooplankton. *Water Research* 28(3):743–746 DOI 10.1016/0043-1354(94)90157-0.
- Kruuk H, Conroy JWH, Moorhouse A. 1991. Recruitment to a population of otters (*Lutra lutra*) in Shetland, in relation to fish abundance. *Journal of Applied Ecology* **28(1)**:95–101 DOI 10.2307/2404116.
- **Kubecka J, Bohm M. 1991.** The fish fauna of the Jordan Reservoir, one of the oldest manmade lakes in Central Europe. *Journal of Fish Biology* **38(6)**:935–950 DOI 10.1111/j.1095-8649.1991.tb03633.x.

- **Kuchta SR, Svensson EI. 2014.** Predator-mediated natural selection on the wings of the damselfly *Calopteryx splendens*: differences in selection among trait types. *American Naturalist* **184(1)**:91–109 DOI 10.1086/676043.
- **Laforsch C, Tollrian R. 2004.** Inducible defenses in multipredator environments: cyclomorphosis in *Daphnia cucullata*. *Ecology* **85(8)**:2302–2311 DOI 10.1890/03-0286.
- **Lampert W. 1993.** Phenotypic plasticity of the size at first reproduction in *Daphnia*: the importance of maternal size. *Ecology* **74(5)**:1455–1466 DOI 10.2307/1940074.
- **Lenormand T, Bourguet D, Guillemaud T, Raymond M. 1999.** Tracking the evolution of insecticide resistance in the mosquito *Culex pipiens*. *Nature* **400**(6747):861–864 DOI 10.1038/23685.
- **Lind MI, Yarlett K, Reger J, Carter MJ, Beckerman AP. 2015.** When the going gets tough, plasticity matters: the alignment between phenotypic plasticity, the major axis of genetic variation and the response to selection. *Proceedings of the Royal Society B: Biological Sciences* **282(1816)**:20151651 DOI 10.1098/rspb.2015.1651.
- **Lüning J. 1995.** Life-history responses to *Chaoborus* of spined and unspined *Daphnia pulex*. *Journal of Plankton Research* **17(1)**:71–84 DOI 10.1093/plankt/17.1.71.
- **Lynch M. 1980.** The evolution of cladoceran life hitories. *Quarterly Review of Biology* **55(1)**:23–42 DOI 10.1086/411614.
- **Lynch M. 1991.** The genetic interpretation of inbreeding depression and outbreeding depression. *Evolution* **45(3)**:622–629 DOI 10.2307/2409915.
- **Machacek J. 1991.** Indirect effect of planktivorous fish on the growth and reproduction of *Daphnia galeata*. *Hydrobiologia* **225(1)**:193–197 DOI 10.1007/bf00028397.
- Mehner T, Diekmann M, Brämick U, Lemcke R. 2005. Composition of fish communities in German lakes as related to lake morphology, trophic state, shore structure and human-use intensity. *Freshwater Biology* **50(1)**:70–85 DOI 10.1111/j.1365-2427.2004.01294.x.
- Morgans CL, Ord TJ. 2013. Natural selection in novel environments: predation selects for background matching in the body colour of a land fish. *Animal Behaviour* 86(6):1241–1249 DOI 10.1016/j.anbehav.2013.09.027.
- O'Keefe TC, Brewer MC, Dodson SI. 1998. Swimming behavior of *Daphnia*: its role in determining predation risk. *Journal of Plankton Research* 20(5):973–984 DOI 10.1093/plankt/20.5.973.
- **Pestana JLT, Baird DJ, Soares AMVM. 2013.** Predator threat assessment in *Daphnia magna*: the role of kairomones versus conspecific alarm cues. *Marine and Freshwater Research* **64(8)**:679–686 DOI 10.1071/MF13043.
- Philipp UJ. 2006. Kreisschreiben-Fischerei. Zürich, Switzerland: Baudirektion Kanton Zürich.
- Pijanowska J. 1997. Alarm signals in Daphnia? Oecologia 112(1):12-16 DOI 10.1007/s004420050277.
- **Pijanowska J, Stolpe G. 1996.** Summer diapause in *Daphnia* as a reaction to the presence of fish. *Journal of Plankton Research* **18(8)**:1407–1412 DOI 10.1093/plankt/18.8.1407.
- **R Core Team. 2016.** *R: A Language and Environment for Statistical Computing.* Vienna: R Foundation for Statistical Computing. *Available at https://www.R-project.org/*.
- **Reger J, Lind MI, Robinson MR, Beckerman AP. 2018.** Predation drives local adaptation of phenotypic plasticity. *Nature Ecology and Evolution* **2**:100–107 DOI 10.1038/s41559-017-0373-6.
- **Riessen HP. 1999.** Predator-induced life history shifts in *Daphnia*: a synthesis of studies using meta-analysis. *Canadian Journal of Fisheries and Aquatic Sciences* **56(12)**:2487–2494 DOI 10.1139/f99-155.

- Rohlf FJ. 2015. The TPS series of software. Hystrix 26(1):9-12 DOI 10.4404/hystrix-26.1-11264.
- **Rose RM, Warne MStJ, Lim RP. 2003.** Exposure to chemicals exuded by fish reduces the filtration and ingestion rates of *Ceriodaphnia cf. dubia. Hydrobiologia* **501**:215–217.
- **Sakwinska O. 2002.** Response to fish kairomone in *Daphnia galeata* life history traits relies on shift to earlier instar at maturation. *Oecologia* **131(3)**:409–417 DOI 10.1007/s00442-002-0901-0.
- **Scheiner SM, Holt RD. 2012.** The genetics of phenotypic plasticity. X. variation versus uncertainty. *Ecology and Evolution* **2(4)**:751–767 DOI 10.1002/ece3.217.
- Schoeppner NM, Relyea RA. 2009. Interpreting the smells of predation: how alarm cues and kairomones induce different prey defenses. *Functional Ecology* 23(6):1114–1121 DOI 10.1111/j.1365-2435.2009.01578.x.
- Seda J, Hejzlar J, Kubecka J. 2000. Trophic structure of nine Czech reservoirs regularly stocked with piscivorous fish. *Hydrobiologia* 429(1-3):141-149 DOI 10.1023/A:1004048415779.
- **Stearns SC. 1989.** The evolutionary significance of phenotypic plasticity. *BioScience* **39(7)**:436–445 DOI 10.2307/1311135.
- **Stibor H. 1992.** Predator induced life-history shifts in a freshwater Cladoceran. *Oecologia* **92(2)**:162–165 DOI 10.1007/bf00317358.
- **Stibor H, Lüning J. 1994.** Predator-induced phenotypic variation in the pattern of growth and reproduction in *Daphnia hyalina* (Crustacea: Cladocera). *Functional Ecology* **8(1)**:97–101 DOI 10.2307/2390117.
- **Swillen I, Vanoverbeke J, De Meester L. 2015.** Inbreeding and adaptive plasticity: an experimental analysis on predator-induced responses in the water flea *Daphnia*. *Ecology and Evolution* **5(13)**:2712–2721 DOI 10.1002/ece3.1545.
- **Tollrian R. 1995.** Predator-induced morphological defenses: costs, life history shifts, and maternal effects in *Daphnia pulex*. *Ecology* **76(6)**:1691–1705 DOI 10.2307/1940703.
- Vanoverbeke J, De Meester L. 2010. Clonal erosion and genetic drift in cyclical parthenogens—the interplay between neutral and selective processes. *Journal of Evolutionary Biology* 23(5):997–1012 DOI 10.1111/j.1420-9101.2010.01970.x.
- Weber A. 2003. More than one 'fish kairomone'? Perch and stickleback kairomones affect *Daphnia* life history traits differently. *Hydrobiologia* 498(1–3):143–150 DOI 10.1023/A:1026297106626.
- Weider LJ, Pijanowska J. 1993. Plasticity of *Daphnia* life histories in response to chemical cues from predators. *Oikos* 67(3):385–392 DOI 10.2307/3545351.
- Weber A, Van Noordwijk A. 2002. Swimming behavior of Daphnia clones: differentiation through predator infochemicals. *Journal of Plankton Research* 24:1335–1348 DOI 10.1093/plankt/24.12.1335.
- Werner EE, Peacor SD. 2003. A review of trait-mediated indirect interactions in ecological communities. *Ecology* 84(5):1083–1100 DOI 10.1890/0012-9658(2003)084[1083:AROTII]2.0.CO;2.
- Wickham H. 2016. ggplot2: Elegant Graphics for Data Anlysis. New York: Springer-Verlag.
- Wolinska J, Löffler A, Spaak P. 2007. Taxon-specific reaction norms to predator cues in a hybrid *Daphnia* complex. *Freshwater Biology* 52(7):1198–1209 DOI 10.1111/j.1365-2427.2007.01757.x.
- Yin M, Laforsch C, Lohr JN, Wolinska J. 2011. Predator-induced defense makes *Daphnia* more vulnerable to parasites. *Evolution* 65(5):1482–1488 DOI 10.1111/j.1558-5646.2011.01240.x.
- Zelditch ML, Swiderski DL, Sheets HD, Fink WL. 2004. Geometric Morphometrics for Biologists—A Primer. Amsterdam and Boston: El Sevier Academic Press.
- **Zuykova EI, Bochkarev NA, Katokhin AV. 2012.** Identification of the *Daphnia* species (Crustacea: Cladocera) in the lakes of the Ob and Yenisei river basins: morphological and molecular phylogenetic approaches. *Hydrobiologia* **715(1)**:135–150 DOI 10.1007/s10750-012-1423-3.