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Pilot-scale continuous recycling of growth medium for the mass culture of a halotolerant *Tetraselmis* sp. in raceway ponds under increasing salinity: A novel protocol for commercial microalgal biomass production.

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Abstract

1 The opportunity to recycle possibility of recycling microalgal culture 2 medium for further cultivation is often hampered by salinity increases from 3 evaporation and fouling by dissolved and suspended substances particulate 4 matter. In this study, the impact of culture re-use after electro-flocculation of 5 seawater-based medium on growth and biomass productivity of a the 6 halotolerant green algal strain of *Tetraselmis* sp., MUR 233, was investigated in

7	pilot-scale open raceway ponds over five months. Despite a salinity increase
8	from 5.5 to 12% (w/v) NaCl, Tetraselmis MUR 233 grown on naturally DOC-
9	enriched recycled medium produced 48 to 160% more ash free dry weight
10	(AFDW) biomass daily per unit pond area than when grown on non-recycled
11	medium. A peak productivity of 37.5 \pm 3.1 g AFDW.m ⁻² .d ⁻¹ was reached in the
12	recycled medium upon transition from ~14 to ~7% NaCl. The combination of
13	high biomass-yielding mixotrophic growth under high salinity has been proven
14	to be a successful sustainable cultivation strategy.
15	Keywords
16	Marine microalgae: recycling: electro-flocculation: mixotrophy.
17	1. Introduction
18	Currently , there is substantial interest in the utilisation of microalgae towards
19	the mitigation of the present and future world crises in food, fuel and
20	environment. Microalgae are very diverse and represent a rich source of
21	phytochemicals which can be used in human food, animal feed, aquaculture, for
22	pharmaceutical, cosmetic and health food products, and conceivably for biofuel
23	and its associated by-products (Brennan & Owende, 2010; Pulz & Gross, 2004;
24	Spolaore et al., 2006). As an alternative form of agriculture, microalgal
25	microalgae cultivation has much commercial appeal because it can potentially
26	yield better biomass productivity rates than land crops over similar land areas
27	(Posten & Schaub, 2009; Schenk et al., 2008). Furthermore, this it offers the
28	possibility to use resources that are otherwise under-utilised (e.g. non-arable
29	land, saline water, wastewater, etc.) or that are accumulating to polluting levels

30 (e.g. excess nutrients leading to eutrophication of waterbodies, build-up of CO₂

in the atmosphere which causes the greenhouse effect, etc.).

To date, much of the successful mass cultivation of microalgae is limited to 32 the production of high value products, high revenues of which offset the high. 33 capital and operational expenditures incurred in generating and processing the 34 biomass. Some of the most cost-prohibitive components of microalgae biomass 35 production are directly associated with the large volume of water, which needs 36 to be processed for cultivation and harvesting (Borowitzka & Moheimani, 2013; 37 Fon Sing et al., 2013; Molina Grima et al., 2003). According to the life-cycle 38 39 assessments of microalgal microalgae cultivation reported by Clarens et al. 40 (2010), Flesch et al. (2013) and Yang et al. (2011), the cost-effectiveness of production could be substantially improved by minimising the water and nutrient 41 footprint through the continuous recycling of the culture medium and by using 42 non-potable sources of water. However, while culture medium recycling seems 43 to confer certain advantages, ultimately the possibility prospect of culture reuse 44 depends largely on the suitability of the harvested water for further continuous 45 cultivation. This is because as opposed to freshly made medium, the recycled 46 47 medium, if untreated, potentially carries over and accumulates all of the 48 dissolved chemical compounds and suspended particles remaining after the harvesting process. For instance, cell wall debris, contaminating organisms 49 50 (e.g. other algal species, bacteria, etc.), dissolved organic compounds and other potentially-growth inhibiting chemicals released from the cells commonly 51 foul the return water. If the water is left untreated prior to returning to the ponds. 52 53 these chemical compounds and particles can quickly lead to increased bacterial

54 activity and culture deterioration (Ben-Amotz, 1995; Chini-Zittelli et al., 1999; Rodolfi et al., 2003). In addition to this, the gradual increase in inorganic salts 55 (salinity) due to evaporation must also be considered, especially in open 56 57 cultivation systems where brackish or saline water is used. This change in 58 salinity can potentially impact on the culture in two ways, namely: by a gradual 59 dominance of more halotolerant species of microalgae, and/or in steady decline in the density of the desired microalga due to its inability to cope with osmotic 60 changes and changing salt ratios. Another potential challenge of recycling 61 medium with increasing salinity is the precipitation of calcium salts, especially in 62 calcium-laden water, thereby causing loss of alkalinity and other minerals such 63 as iron and phosphorus (Shimamatsu, 2004). 64

Clearly more exhaustive research should be carried out to enhance and 65 optimise current processes used to grow monocultures of microalgae in 66 67 recycled culture medium. Considering the challenges of medium recycling difficulties of recycling culture media and the need to use non-potable water for 68 improving the economics and sustainability of microalgal biomass production, 69 the real challenge lies in the ability to sustain monocultures of microalgae for as 70 long as possible in a recycled culture medium of potentially increasing salinity 71 72 without any loss in biomass productivity. and quality. This study investigates the effect of culture medium recycling on culture health and biomass productivity of 73 74 a halotolerant strain of *Tetraselmis* sp. grown at increasing salinity and continuously for long periods in an open raceway ponds under outdoor field 75 76 conditions. Thus, the proposed protocol is a direct contribution to the

- 77 development of cost-efficient and sustainable production of biomass from saline
- microalgae in the field by a rational recycling of the culture medium.
- 79

80 2. Materials and methods

81 **2.1. Location and microalgal species**

The experiment was carried out from August to December 2012 in outdoor open raceway mixed ponds in a remote area in a semi-arid climate at Karratha, Western Australia, Australia (20S 45'47.72", 116E 44'9.88"). *Tetraselmis* sp. MUR 233, originally sourced from the Murdoch University Algae Culture Collection (Perth, Western Australia) was used as test organism. This alga was maintained in semi-continuous cultivation mode in outdoor raceway ponds at the same location for at least two years prior to this experiment.

89 **2.2. Ponds**

Two 2 m² (M1 and M2) above-ground fibreglass open raceway ponds (2 90 x 1 x 0.4 m, L x W x H) and two $25m^2$ (P4 and P5) in-ground 40 mm HDPE-91 92 lined raceway ponds, the only ponds available at the time of the experiment, were employed in this study. The cultures in M1 and M2 were used as controls, 93 94 where the harvested portion of the cultures was replaced with fresh medium 95 whereas recycled medium was used in P4 and P5 ponds (experimental 96 treatments) (**Figure 1**). The use of the smaller ponds as controls instead of 97 having a control pond at each scale were dictated by the need for a minimum volume for the downstream processing, and the small ponds alone would not 98 have met this requirement. Based on data gathered from previous long-term 99

100 continuous cultivation in both sets of ponds, revealed no significant differences in growth and biomass productivities between the pond sizes were found 101 102 (Isdepsky, unpublished data) and therefore, it was deemed reasonable to proceed with the current experimental setup. 103 To limit operational differences between the ponds to the effect of the 104 recycled medium only, all cultures were kept at an operating depth of 20 cm. 105 were mixed with a surface velocity of 20 cm.s⁻¹ and were harvested at 50% of 106 the culture volume. Also, the salinity of M1 and M2 ponds was maintained equal 107 to that of P4 and P5 ponds by salt addition. to limit operational differences 108 between the ponds to the effect of the recycled medium only. 109

110

111 **2.3. Culture conditions**

The target starting cell density was ca. 40x10⁴ cells.mL⁻¹. The pH of all 112 ponds was maintained at pH 7.2 \pm 0.3 with food grade CO₂ using a pH-stat 113 114 system. The CO₂ supply was switched off between 20:00 and 07:00 to avoid CO₂ loss due to a reduction in water level upon culture harvest. The CO₂ supply 115 It was switched back on the following morning after the harvested volume has 116 been replaced. The starting culture medium salinity was set to 5.5% (w/v) NaCl 117 by adding commercial grade pool salt (Lake Deborah Natural Australian Lake 118 Salt) to raw seawater collected locally in Karratha. Nutrients in the form of 119 120 commercial grade sodium nitrate and potassium di-hydrogen phosphate were added on a daily basis to provide a nominal concentration of 35 mg L^{-1} of NO₃⁻¹ 121 and 2 mg.L⁻¹ of PO_4^{3-} in the culture medium. Additional nutrients were added to 122 the recycled medium to maintain the NO_3^{-1} and PO_4^{3-1} concentration in the ponds. 123

124	Water losses from evaporation and processing (~10% from the centrifugation
125	step) were replenished with unfiltered raw seawater only (~3.5% NaCl (w/v)), as
126	shown in Figure 1 .

127

128 2.4. Harvesting and medium recovery

On harvesting days, 50% of the culture volume from P4 and P5 was 129 pumped to a 3600 L proprietary electro-flocculation unit and processed for a 130 maximum period of 2 h. Floating flocs formed during the electro-flocculation 131 132 process (see Lee et al. (2013)) were harvested and transferred into a 250 900 L conical tank for biomass settling prior to centrifugation using a T10 Evodos 133 centrifuge. Thereafter, the clarified portion (i.e. supernatant) of the water in the 134 electro-flocculation unit was pumped into a 2500 L open-top conical tank, left to 135 stand overnight or longer without any chemical treatment and then pumped 136 back to the ponds on the next harvesting day (Figure 1). Any suspended flocs 137 remaining in the supernatant were gravity-settled in conical tanks for further 138 biomass concentration and collection for centrifugation. 139

140 **2.5. Analytical procedures**

Cell counts were performed daily with an improved Neubauer
haemocytometer after fixing the cells with Lugol's iodine solution. NO₃⁻ and
PO₄³⁻ concentrations were determined using a DataLine photometer and
Aquaspex© reagent kits (Aquaspex, South Australia). A digital hand-held Atago
refractometer (model PAL-106S) was used to measure the salinity (% NaCl
(w/v)) of the culture medium. Total dissolved organic carbon (DOC) was

147	analysed at the Marine and Freshwater Research Laboratory, Murdoch
148	University, Western Australia. DOC (0.45 μm filtered water samples) was
149	measured as non-purgeable organic carbon (NPOC) following
150	acidification using the high temperature combustion non-dispersive infrared gas
151	analysis method (MAFRL Method 6000) using a Shimadzu Corporation TOC-V-
152	CSH Organic Carbon Analyser. For dry weight and ash free dry weight (AFDW)
153	determinations, culture samples were filtered on GF/C Whatman filter papers
154	pre-combusted at 450 °C. Filtered samples were dried at 70 °C for 3 h overnight
155	and cooled to room temperature over activated silica gel under vacuum prior to
156	weighing. Dried samples were then combusted at 450 $^{\circ}$ C for 3 h and
157	subsequently cooled to room temperature over activated silica gel under
158	vacuum prior to weighing. The AFDW and ash contents were calculated as a
159	percentage of the dry weight of the filtered samples.

160 **2.6. Statistical analysis**

161 One-Way ANOVA and One-Way repeated measures ANOVA analysis was

used to determine significant differences between treatments (α =0.05%). An

163 ANOVA analysis based on ranks was performed whenever the normality

164 Shapiro Wilk test or equal variance tests failed.

165 **3. Results and Discussion**

166

3.1. Long-term cultivation under increasing salinity

167 The semi-continuous culturing of *Tetraselmis* MUR 233 in both non-recycled 168 and recycled media lasted for almost five months (127 days) without any major 169 interruptions or culture loss. Throughout this period, the cultures benefited from

abundant sunlight, night and day temperatures above 10 and 28°C respectively,
and no rain (Figure S1 in Supporting Information). For the purpose of the
experiment and due to technical constraints that prevented the assessment of
the impact of variation in solar intensity and temperature on growth and
biomass productivity, any variations in these two parameters were considered
as minor compared to the impact of increasing salinity and medium recycling on
culture performance.

Evaporation losses were on average 20 L.m⁻².d⁻¹ in all ponds over the 177 experimental period, which resulted in a salinity increase from 5.5 up to 14.0% 178 179 NaCl in the recycled medium. For the entire cultivation period, the daily average 180 nutrients remaining in both sets of ponds amounted to 36% of the overall nitrate input and 26% of the overall phosphate input. An average volume of 920 L of 181 water (including seawater makeup for evaporation and process losses) was 182 recirculated on a daily basis to each 25 m² pond, which equates to ~115 kL of 183 culture water in each pond over 127 days being recycled instead of being 184 discarded. 185

186 Throughout the experiment, minimal contamination by other 187 microorganisms was observed in the culture and the harvested biomass. Diatoms, filamentous cyanobacteria and the ciliated protozoan Euplotes sp. 188 189 (Figures S2a-c in Supporting Information) were occasionally detected, but rarely in noticeable quantities $(<1\times10^4 \text{ cells.mL}^{-1})$ which might significantly 190 impact the overall long-term culture guality. Microscope observation and 191 turbidity measurements of the recycled water indicated that the return water 192 193 was colourless and clear of any suspended particles. Examination of the

194	flocculated cells under the microscope revealed that the electro-flocculation
195	process actually resulted in the entrapment of the cells within a gelatinous
196	matrix (Figures S2di-diii in Supporting Information) and did not impart any
197	noticeable physical damage or changes in cell shape.
198	The cultivation periods for M1, M2, P4 and P5 are shown in Figures 2a,
199	b, c and d respectively. This was characterised by four stages labeled I, II, III
200	and IV, each of which corresponded with harvesting frequencies of 2, 3, 4 and
201	variable days, and mean salinity ranges of 5-9, 8-12, 9-14 and 6–9% NaCl
202	respectively. It is worth noting that After Stage III, a portion of the culture water
203	was discarded after stage III and substituted with raw seawater to lower the
204	salinity for the last stage of cultivation (i.e. stage IV).

3.2. Analysis of specific growth rate, biomass productivity and AFDW/ ash content

The specific growth rate, the AFDW biomass productivity and the AFDW/ 207 Ash content of the dry biomass obtained for each aforementioned stages in 208 which the long-term cultivation experiment under increasing salinity was carried 209 out are shown in Figures 3 and 4 Table 1 and Figure 3. During the first fifty-210 211 three days of culturing (i.e. stage I in Figure 2) the cell density in the control 212 ponds M1 and M2 (Figures 2a and b) remained relatively stable between 40x10⁴ and 80x10⁴ cells.mL⁻¹. At a harvesting frequency of every two days and 213 salinity range of 5-9% NaCl, the control cultures had an average specific growth 214 rate of 0.35 \pm 0.02 d⁻¹ and mean biomass productivity of 15.4 \pm 0.7 g AFDW.m⁻ 215 ².d⁻¹ (stage I in Figures 3a and b Table 1). In contrast, despite having the same 216

217	initial starting cell densities, the cultures receiving the recycled medium, P4 and
218	P5, reached a higher cell density at a faster rate, resulting in twice as many
219	cells as the control ponds by the 38th day of cultivation (stage Ia in Figures 2c
220	and d). This boost in the standing biomass concentration correlated with a
221	significantly higher AFDW content (P<0.001) (Figures 4c and d Figures 3c
222	and d), a rate of growth 11% faster than in the control ponds and an 11%
223	improvement in the AFDW biomass productivity to yield a mean of 26.9 \pm 1.9 g
224	AFDW.m ^{-2} .d ^{-1} over the initial salinity range. Because the rate of growth of the
225	cells was faster than the harvest frequency, there was a gradual increase in the
226	baseline cell density, which did not match that of the control ponds anymore.
227	Consequently, the cultures in P4 and P5 were harvested consecutively on the
228	37th and 38th days in an attempt to bring the starting cell density back to
229	40x10 ⁴ cells.mL ⁻¹ (stage 1b in Figures 2c and d). This was in turn followed in
230	turn by a 26% increase in growth rate (i.e. from 0.39 \pm 0.06 d ⁻¹ to 0.49 \pm 0.02 d ⁻
231	¹) in the more dilute culture in the recycled medium for the next 15 days (stage
232	Ib in Figure 3a Table 1) as a result of better light penetration. This spike in
233	growth rate was followed by a decline in AFDW content of the biomass, such
234	that overall, the biomass productivity during that particular period of cultivation
235	slightly declined to 23.8 \pm 3.0 g AFDW.m ⁻² .d ⁻¹ (stage lb in Figure 3b Table 1).

In the course of the stage II of this experiment, (i.e. between the 53rd
and 78th days), all four cultures were harvested every 3 days. Both sets of
cultures maintained a steady growth pattern as the salinity gradually increased
from 8 to 12% NaCl. There were no signs of culture deterioration in any of the
ponds, and the cultures receiving the recycled medium outperformed the control

cultures by close to 50% in terms of AFDW biomass productivity (stage II in
Figure 3b Table 1). This was reflected in a higher range of cell densities and
AFDW content as compared to the previous days of culture. No adverse effect
associated with medium recycling on the continuous culturing of *Tetraselmis*MUR 233 could be observed.

For the following thirty-four days of the experiment (i.e. stage III), the limit 246 of salinity tolerance of the cultures was investigated by increasing the residence 247 time between harvests to 4 days, which allowed for a longer evaporation period, 248 249 and thus higher salinities in the ponds. The cultures in the control ponds M1 and M2 maintained a steady growth pattern until ~12% NaCI (stage III in Figures 2a 250 and b), at which point the cultures appeared paler and not as healthy as before. 251 It was therefore decided to maintain those cultures at 11% NaCl for as long as 252 possible. Likewise, there was a gradual decline in cell numbers in experimental 253 254 ponds P4 and P5 up to a salinity of 14% NaCl, at which point the cell densities were too low to sustain the same harvesting frequency (stage III in Figures 2c 255 and d). The salinity was temporarily brought down to ca. 12% NaCl with raw 256 seawater to revive the cultures before again bringing the salinity up to 14% 257 NaCl through evaporation. Once again, the cell densities in the recycled 258 medium gradually dipped. During the period of extreme salinity, the average 259 growth rates in the recycled medium were 24% slower than those in the control 260 261 ponds (stage III in Figure 3a Table 1), but because the decline in biomass (i.e. AFDW content (w/v) was not as significant as for the growth rates (P<0.001) 262 263 (Figures 4a and b Figures 3a and b) and because the average biomass concentration in the experimental ponds was 42% higher than in the control 264

ponds, the overall biomass productivity remained high at 31.9 ± 2.7 g AFDW.m⁻ 265 ².d⁻¹, which was twice as much as that achieved in the control ponds (Figure 3b 266 Table 1). Therefore, it appears that the limit of salinity tolerance of *Tetraselmis* 267 MUR 233 is close to 12 % NaCl but more importantly, it also seems that the 268 cells are slightly more tolerant to the high salinity when grown in the recycled 269 medium. A likely reason for this difference could be the differences in DOC 270 content between the fresh and recycled medium media (see Section 3.3). 271 The final stage of the experiment (i.e. stage IV) consisted of returning the 272 273 culture salinity to ca. 7% NaCl with raw seawater over two harvesting periods in 274 all ponds to determine the rate and extent of culture recovery from the prolonged high salinity treatment. The M1 culture recovered quickly from the 275 high salinity treatment with much higher cell densities than in the previous five 276 days. However, despite the increase in cell numbers, the growth rates and 277 AFDW productivities remained at the same level as those obtained in stage I 278 (Figure 3 Table 1). A breakdown in the pH-stat system in M2 resulted in poor 279 culture recovery and as the culture was left undisturbed (i.e no harvest and 280 change of medium), the culture salinity increased due to evaporation and 281 eventually the culture could no longer be maintained. Dilution of the 282 experimental cultures (P4 and P5) with raw seawater to bring the salinity to 7% 283 NaCl resulted in a steady recovery in the cell number and culture appearance. 284 The mean growth rate of 0.34 \pm 0.05 d⁻¹ of all ponds over this brief period at 285 lower salinity was comparable to that obtained in stage I. In contrast, the AFDW 286 content of the biomass in P4 and P5 increased by 33% in comparison to that 287 obtained in stage I of the experiment. The concomitant increase in cell density 288

and AFDW content in these two ponds consequently led to a biomass productivity of 37.5 ± 3.1 g AFDW.m⁻².d⁻¹, which is the absolute maximum achieved throughout the experiment.

Up until the beginning of stage IV of the experiment, the inorganic portion 292 (ash content) of the biomass in the experimental ponds P4 and P5 was 6% less 293 than that obtained in the control ponds M1 and M2 (Figure 4 Figure 3). This 294 difference was small enough to be statistically significant (P<0.001), which 295 means that overall, the cells in the recycling culture medium contained slightly 296 297 more organic carbon. Upon returning the culture salinity to 7% NaCl with raw 298 seawater, the mean differences in declines in ash content was much more significant in the experimental ponds P4 and P5 (from 72.0 % to 61.2 % ash, 299 (P < 0.001)) than in the control (from 76.5% to 72.4% ash). It thus appears that 300 the cultures having undergone culture recycling treatment are in a much better 301 condition to grow faster under the return of normal culture conditions. 302

303 In the light of the results obtained with respect to salinity tolerance, it can be concluded that One of the most fundamental observations from this study is 304 305 that Tetraselmis sp. MUR 233 demonstrated true halotolerance traits in that it adapted extremely well to both gradual and sudden changes in salinity. This is 306 not atypical of the *Tetraselmis* genus; in fact, there is supporting evidence that 307 308 certain *Tetraselmis* species and strains are equipped with a highly efficient Na⁺ pump (Popova & Balnokin, 2013; Strizh et al., 2004) and a highly adaptable 309 osmolyte regulatory osmoregulatory mechanism to cope with rapid and gradual 310 changes in salinity (Hellebust, 1976; Kirst, 1977; Kirst, 1988). However, it is 311 quite unusual to observe exceptionally good growth and biomass productivities 312

313	at the high salinities tested in this experiment and to our knowledge, it is
314	believed that this current study is the first ever to report the growth of
315	Tetraselmis continuously in actual outdoor conditions in large quantities over a
316	wide salinity range without any significant loss in biomass. It appears that the
317	combination of constant abundant sunlight, high temperatures, adequate supply
318	of inorganic nutrients (CO ₂ , NO ₃ ⁻ and PO ₄ ³⁻) were conducive towards providing
319	adequate energy to the cells to combat and cope with the high salinity-stress.
320	Using the recycled medium as growth medium seems to have provided with an
321	additional benefit, the exact cause and effect of which is yet to be identified.

322 3.3. Analysis of dissolved organic carbon

Dissolved organic carbon (DOC) concentration in the raw seawater, 323 recycled medium, and in ponds M2 and P4 was analysed over a period of ten 324 days between the 50th and 59th days of the experiment to investigate the 325 potential influence of culture medium recycling on DOC loads. A cyclic pattern 326 327 in DOC levels was observed in both ponds whereby a drop in DOC occurred after each harvesting day which was subsequently followed by a gradual 328 329 increase in DOC up to when the culture was next harvested and refilled with fresh/recycled medium (Figure 5 Figure 4). The DOC input from the recycled 330 medium in pond P4 was on average four times more concentrated than that 331 332 added to pond M2 via the raw seawater. This resulted in an overall increase of 27% in DOC in P4 over the ten days, compared to pond M2. 333

334 DOC of the culture water was also measured on two occasions
 335 immediately before and after the electro-flocculation process to determine the

effect of electro-flocculation on DOC concentration in the return water. The
results showed that the harvesting process removed on average 25% of the
initial DOC from the water so that the residual DOC concentration in the clarified
water that was left to stand overnight in the settling conical tanks was less than
7.0 mg.L⁻¹. However, by the next harvesting/pond refilling occasion, the DOC
would have increased to more than 8.0 mg.L⁻¹ (Figure 5 Figure 4).

342 3.4. Analysis of electro-flocculation method

Given the sustained high growth and high AFDW productivities achieved 343 in the recycled medium throughout the cultivation period, it is strongly believed 344 that the electro-flocculation method could have been a major contributing factor 345 to the success of the study. The electro-flocculation process could have 346 conferred two critical advantages which allowed for sustained culture medium 347 re-use: (1) the apparent absence of cell breakage during the harvesting process 348 and (2) the partial reduction in dissolved organic carbon compounds during 349 electro-flocculation. In opposition to other harvesting methods such as 350 351 centrifugation and filtration which use centrifugal forces and pressure to 352 concentrate the microalgal cells and which general lead to cell damage, electro-353 flocculation is non-destructive as the cells are simply aggregated and entrapped in growing networks of aluminium polymers after neutralisation of cell-to-cell 354 355 charge repulsion (Lee et al., 2013; Pearsall et al., 2011). The resultant clarified water is thus much less contaminated with intracellular organic compounds, 356 reducing the risk of fouling by bacteria, cell debris and growth-inhibiting 357 substances, and therefore making the culture medium more amenable for re-358 use. The successful cultivation of *Tetraselmis* sp. in the flocculant-treated 359

360	culture medium as observed in this study is in full agreement with reports by
361	Farid et al. (2013), Wu et al. (2012) and Rwehumbiza et al. (2012), even though
362	different types of flocculants were used in these experiments.
363	The second advantage of the electro-flocculation step is that it in itself a
364	cleaning process due to its oxidative and flocculating characteristics (Koren &
365	Syversen, 1995; Sasson et al., 2009). The fact that the DOC level in the water
366	declined by over 25% immediately after the electro-flocculation step confirms
367	the sanitising effect of this harvesting method, which could have been a
368	significant contributing factor towards the healthy growth of the Tetraselmis
369	MUR 233 cells in the water recycled to the culture. At present, the exact effect
370	of the electrolysis of the culture water is unknown. However, changes in the
371	physicochemical parameters as well as in the water chemistry have been
372	observed and will be reported elsewhere.

373 3.5. Analysis of AFDW biomass productivity performance of the
 374 proposed cultivation method

One of the most important criteria for microalgal cultivation to be 375 economically viable is the ability to maintain high growth rates in the cultures to 376 377 ensure fast biomass throughput high productivities. Another key performance indicator for the commercial potential of microalgal cultivation is the biomass 378 productivity of the culture system. A specific growth rate of $\leq \geq 1$ d⁻¹ would be 379 ideal to maximise harvesting frequency and biomass output. With the simplest 380 setup and cultivation technology and, in the absence of freshwater input to 381 maintain a constant salinity, we have shown that the growth rates of *Tetraselmis* 382

MUR 233 in open raceway ponds that can be expected in real conditions are 383 between 0.25 and 0.50 d⁻¹. This is within the normal range of growth rates 384 reported in the literature and summarised by Griffiths & Harrison (2009) and is 385 therefore reassuring that the baseline growth rates can be easily achieved with 386 387 *Tetraselmis* MUR 233, even at high salinity. Given that the anticipated average areal biomass productivity of an open outdoor pond culture system is ~ 24-27 388 g.m⁻².day⁻¹ according to Griffiths & Harrison (2009) and Lee (2001), it was clear 389 that the AFDW biomass productivities achieved with the recycled medium in 390 this study showed a net 19.9% improvement over the expected average 391 biomass productivity. Interestingly, the differences in growth rates between the 392 control and experimental ponds are rather small, compared to the differences in 393 the AFDW biomass productivities. If the cells in the experimental ponds were 394 truly undergoing mixotrophic growth (see below), then it appears that the growth 395 rates in the ponds were not limited by the amount of carbon (inorganic and/or 396 397 organic) present in the medium, but rather by other factors, for example by light.

In terms of growing Tetraselmis MUR 233 in mineral culture medium of 398 increasing salinity, the current study demonstrated an average 15 g AFDW.m⁻ 399 2 .d⁻¹ (i.e. in the control ponds), which represents 35% less than the average 400 values obtained from other raceway pond data (Griffiths & Harrison, 2009; Lee, 401 402 2001), This could indicate that cultivation of *Tetraselmis* MUR 233 in mineral 403 medium of increasing salinity is possible but is not an economical option. It is to be noted, however, that the biomass productivity values obtained in the current 404 405 study could be excessively higher than the true mean biomass productivities

reported in the literature as these are often reported on a dry weight basis,

407 which makes biomass productivities often difficult and misleading.

The fact that pond P4 was more productive, received and contained 408 relatively higher amounts of DOC as compared to pond M2 suggests that the 409 culture was growing mixotrophically. Some Several Tetraselmis species are 410 have been shown to be capable of utilising a wide range of carbon compounds 411 412 to complement photosynthesis and under such circumstances, it is often reported that this mixotrophic growth results in improved biomass production 413 414 (Cid et al., 1992; Day & Tsavalos, 1996; Xie et al., 2001) as compared to 415 phototrophic growth (Biller et al., 2012). It certainly appears to have been the 416 case in this study, given the significant improvement in growth rate and biomass productivity in ponds P4 and P5 compared to those in ponds M1 and M2. 417 Furthermore, this improvement occurred in spite of variations in starting cell 418 419 densities and culture thickness, which under phototrophic growth, would probably have led to irregular or lower culture performance due to inconsistent 420 light penetration through the cultures. The most probable source of organic 421 carbon would have been from the partial decomposition of residual biomass that 422 would have settled at the bottom of the conical tanks prior to returning the 423 clarified water to the ponds. Lysates from the biomass can represent a rich 424 source of highly suitable and assimilable organic carbon compounds that the 425 426 living cells can readily absorb take up, a fact that Spectrova et al. (1982) successfully embraced used for the cultivation of *Dunaliella tertiolecta*. In 427 addition, the changes in the water chemistry after the electro-flocculation 428 429 process could have led to a shift in bacterial population towards the elimination

- 430 of growth-inhibiting bacteria and/or the increase in growth-promoting bacteria in
- the culture. This hypothesis, as well as the effects of DOC on growth in
- 432 *Tetraselmis* MUR 233, is being investigated further.

One of the most fundamental observations from this study is that 433 Tetraselmis sp. MUR 233 demonstrated true halotolerance traits in that it 434 adapted extremely well to both gradual and sudden changes in salinity. This is 435 436 not atypical of the *Tetraselmis* genus; in fact, there is supporting evidence that certain Tetraselmis species and strains are equipped with a highly efficient Na+ 437 438 pump (Popova & Balnokin, 2013; Strizh et al., 2004) and a highly adaptable 439 osmolyte regulatory mechanism to cope with rapid and gradual changes in salinity (Hellebust, 1976; Kirst, 1977; Kirst, 1988). However, it is guite unusual 440 to observe exceptionally good growth and biomass productivities at the high 441 salinities tested in this experiment and to our knowledge, this current study is 442 443 the first ever to report the growth of *Tetraselmis* continuously in actual outdoor conditions in large quantities over a wide salinity range without any significant 444 loss in biomass. It appears that the combination of constant abundant sunlight, 445 high temperatures, adequate supply of inorganic nutrients (CO₂, NO₃⁻ and PO₄³⁻ 446) and the availability of dissolved organic carbon compounds was conducive 447 448 towards providing adequate energy to the cells to combat and cope with the salinity stress. 449

450 **4. Conclusions**

451 This proof-of-concept study demonstrates that (1) the expected baseline 452 productivity from a halotolerant microalgal microalgae culture grown under

453	increasing salinity in a semi-arid climate is close to and beyond consistently
454	greater than 15 g AFDW.m ⁻² .d ⁻¹ , (2) <i>Tetraselmis</i> MUR 233 can be grown
455	continuously with minimal freshwater input, and in recycled culture medium
456	without any interruption decline in biomass productivity and, (3) the electro-
457	flocculation harvesting technique circumvents the wastage of water and
458	nutrients and the need for any water treatment for medium sterilisation. This
459	novel integration of cultivation and harvesting processes opens an exciting new
460	avenue in the production of microalgal biomass in saline water.
461	

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573 574	7. Figure captions
575	Figure 1. Flow diagram for the cultivation of <i>Tetraselmis</i> MUR 233 in
576	fresh medium, the harvesting of the biomass through the electro-flocculator and
577	the Evodos centrifuge, and the subsequent return of the clarified supernatant
578	into the experimental cultures.
579	Figure 2. Cell density (filled circle) and salinity trends (area) in the
580	control treatments M1 (A) and M2 (B), and in the recycled culture medium
581	treatments P4 (C) and P5 (D). The cultures stages I(a, b), II and III represent
582	harvesting frequencies of 2, 3 and 4 days respectively and stage IV represents
583	a period of operation at low salinity with harvesting carried out whenever
584	possible. Thin arrows in graph C and D indicate two consecutive harvesting
585	days (day 37 and 38), and the bold arrow in graph B indicates a breakdown in
586	the pH-stat for the M2 culture.
587	Figure 3. Specific growth rates and AFDW biomass productivity of
588	Tetraselmis MUR 233 at the different stages of semi-continuous growth in fresh
589	medium (mean od M1 and M2 results) (white bar) and in recycled culture
590	medium (mean of P4 and P5 results) (grey bar). Mean growth rate and AFDW
591	biomass productivity obtained during culture stage lb are illustrated as filled
592	circles. Bars represent standard errors.

Figure 4. Figure 3. AFDW (filled circle) and ash content (open circle) on a DW basis of *Tetraselmis* sp. MUR 233, and salinity (area) in the control treatments M1 (A) and M2 (B) and the recycled medium treatments P4 (C) and P5 (D). The culture stages I, II, III and IV are as explained in Figure 2 caption. Figure 5. Figure 4. DOC concentrations in control pond M2 (filled circle) and experimental pond P4 (open circle) over a ten-day period in September 2012. DOC input from raw seawater into pond M2 and from the recycled medium into pond P4 on harvesting days are indicated. Bars represent standard errors for 3 replicates.

603 8. Table captions

Table 1. Specific growth rates and AFDW biomass productivity of
 Tetraselmis MUR 233 at the different stages of semi-continuous growth in fresh
 medium (mean of M1 and M2 results- control, except for stage IV where results
 from M1 only are shown) and in recycled culture medium (mean of P4 and P5
 results-experimental).











644	Table	1.

	Culture stage	Specific grow	wth rate, μ	Biomass produ	ctivity
		(d ⁻¹)		(g AFDW.m ⁻² .d	l ⁻¹)
		Control	Experimental	Control	Experimental
	la	0.35 ± 0.02	0.39 ± 0.02	15.4 ± 0.7	26.3 ± 1.9
	lb		0.49 ± 0.06	47.00 . 0.0	23.8 ± 3.0
		0.26 ± 0.03	0.32 ± 0.03	17.00 ± 0.9	25.2 ± 1.7
		0.37 ± 0.03	0.28 ± 0.03	15.50 ± 1.4	31.9 ± 2.7
C 4 F	IV	0.35 ± 0.05	0.34 ± 0.05	14.3 ± 2.1	37.5 ± 3.1
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651	Pilot-scale continuous recycling of growth medium for the mass culture of
652	a halotolerant Tetraselmis sp. in raceway ponds under increasing salinity:
653	A novel protocol for commercial microalgal biomass production.
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673	High	lights
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675 676 677		<i>Tetraselmis</i> sp. MUR 233 grew in recycled culture medium with increasing salinity
678 679 680		The salinity of the recycled culture medium ranged between 5 and 12% NaCl (w/v)
681 682		Electro-flocculation was employed as harvesting technique
683	\triangleright	Growth and AFDW biomass productivity were superior in the recycled
684		culture medium
685		