

1 **Title:** Short term changes in moisture content drive strong changes in Normalized Difference
2 Vegetation Index and gross primary productivity in four Arctic moss communities

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11 **Abstract**

12 Climate change is currently altering temperature and precipitation totals and timing in
13 Arctic regions. Moss communities constitute much of the understory in Arctic vegetation, and
14 as poikilohydric plants moss are highly sensitive to timing and duration of moisture levels. Here
15 we investigate the role of moisture content on NDVI, red and near-infrared reflectance, and
16 gross primary productivity (GPP) of two sphagnum and two pleurocarpus moss community
17 types during two separate drying experiments. For both experiments, blocks of moss were
18 collected near Imnavait Creek, Alaska, saturated to full water capacity, and then allowed to air
19 dry before being re-saturated. Drying of blocks was conducted in a translucent outdoor tent
20 during the first experiment and under indoor climate-controlled conditions during the second.
21 Community NDVI (experiment 1 and 2), and GPP (experiment 2) were measured at regular
22 intervals during the dry-down and after rewetting. In both experiments, moss NDVI sharply
23 declined between 80% and 70% moisture content for sphagnum moss communities (NDVI
24 change = -0.17 to -0.2), but less so for pleurocarpus moss communities (NDVI change = -0.06 to
25 -0.12). Changes in NDVI were largely the result of increases in reflectance in red wavelengths.
26 Peak GPP for all community types in the second experiment (1.31 to $2.08 \mu\text{mol m}^{-2} \text{s}^{-1}$) occurred
27 at 80% moisture content and declined significantly as moisture content decreased. Rates of GPP
28 continued to decline below 80% moisture content until near zero as moss reached a steady
29 weight (air dry) over a period of 84 hours, while NDVI values declined slowly between 70%
30 hydration and fully air dry. Re-saturation caused NDVI to increase in both sphagnum (NDVI
31 change = +0.18 to +0.23) and pleurocarpus (NDVI change = +0.10 to +0.17) communities. Only
32 sphagnum communities showed GPP resuming ($0.824 \mu\text{mol m}^{-2} \text{s}^{-1}$) after 24 hours. The strong

33 changes in NDVI and mismatch of moss NDVI values and GPP with moisture content
34 fluctuations indicate that using NDVI as a proxy for productivity in Arctic vegetation
35 communities may be problematic and underscores the need for quantification of moss
36 community coverage, composition, and moisture content.

37

38

39 **Introduction**

40 Arctic regions have experienced significant warming over the past several decades
41 (Overland 2002, Chapin et al. 2005, IPCC 2013) with substantial ecosystem consequences
42 (Hinzman et al. 2005). In some regions of the Arctic, changes in precipitation patterns,
43 increased evapotranspiration, and falling water tables associated with climate change have
44 reduced pond size and number (Riordan et al. 2006, Andreson and Lougheed 2015) and
45 affected soil moisture available to tundra plants (Roulet et al. 1992, IPCC 2013). While
46 precipitation models for the Arctic generally suggest an increase in precipitation, most of this
47 increase is distributed over winter and fall (as snow), while summer would remain relatively
48 unchanged (Kattsov et al. 2007). Arctic plant communities are shifting to vegetation types and
49 resulting carbon dynamics that reflect drier conditions (Oechel et al. 1992, Chapin et al. 1995,
50 Mack et al. 2011). Mosses constitute an important component of Arctic vegetation
51 communities, particularly in the understory, and may contribute substantially to ecosystem
52 carbon fluxes (Shaver and Chapin 1991, Douma et al. 2007, Campioli et al. 2009, Turestky et al.
53 2010, Olivas et al. 2011, Zona et al. 2011). As a result of warming, recent vegetation
54 measurements in the Arctic have shown a decrease in moss cover (Chapin et al. 1995, Molau
55 and Alattalo 1998, Elmendorf et al. 2012, Hollister et al. 2015, Hobbie et al. 2017).

56 Decreases in available water have a particularly marked effect on the carbon balance in
57 moss-dominated communities (Titus et al. 1983, Rydin and McDonald 1985, Alm et al. 1999,
58 Komulainen et al. 1999). As poikilohydric plants, mosses do not have the ability to actively
59 control water loss and therefore are highly susceptible to changes in water availability (Van
60 Breemen 1995, Proctor and Tuba 2002). Consequently the moisture content of mosses varies

61 widely and rapidly compared to that of tundra vascular plants, and most mosses are more
62 resilient to periodic drying, or even complete desiccation, than vascular plants (Levitt 1956,
63 Proctor and Tuba 2002). Mosses have an optimal water content for peak photosynthetic rates
64 that does not necessarily occur at full saturation (Ueno and Kanda 2006, Van Gaalen et al. 2007,
65 Harris 2008). Photosynthetic rates of moss are particularly sensitive to drying because cellular
66 water content is a crucial limiting factor in the light reactions (Skre and Oechel 1981). As water
67 loss becomes more severe, photosynthesis declines significantly (Schipperges and Rydin 1998).
68 Even after re-saturation there can be a substantial lag before activity resumes (Oliver and
69 Bewley 1984, Green and Lange 1995, Charron and Quatrano 2009, de Carvalho et al. 2012).
70 Multiple drying and wetting cycles have a negative effect on the photosynthesis of moss due to
71 lengthy recovery times (McNeil and Waddington 2003). As a result, prolonged periods of
72 warmer, drier conditions have the potential to adversely affect moss growth and
73 photosynthesis (Potter et al. 1995, Dorrepaal et al. 2004, Proctor et al. 2007) and therefore
74 ecosystem productivity (Turetsky et al. 2012).

75 Accompanying variation in moss moisture content are changes in apparent coloration of
76 some mosses. For example, some sphagnum species exhibit a markedly lighter appearance as a
77 result of moisture loss (Van Breeman 1995), suggesting the possibility that desiccation may
78 have implications for plant reflectance and remote sensing indices. Remote sensing has
79 provided an effective method for determination of important environmental parameters on
80 large spatial and temporal scales that, using conventional methods, would be otherwise cost
81 and time prohibitive (Kerr and Ostrovsky, 2003). These resource limitations are exacerbated in
82 Arctic regions as a result of the remoteness and scale of study areas. One metric commonly

83 used is the Normalized Difference Vegetation Index (NDVI) that takes advantage of the strong
84 absorbance of the red and strong reflectance in the near-infrared region of the electromagnetic
85 spectrum by green plants (Kriegler et al. 1969). Changes in growing season length and available
86 moisture associated with climate change have been shown to alter remotely sensed NDVI, a
87 measure strongly correlated with green biomass (Jia et al. 2003, Reidel et al., 2005, Gamon et
88 al. 2013), shrub cover (Stow et al. 2007, Walker et al. 2012), and community productivity (Harris
89 2008). Increases in Arctic vegetation cover have been associated with increases in peak season
90 NDVI measurements (Laidler et al. 2008, Chen et al. 2009, Kushida et al. 2009), particularly
91 when moisture is high (Riedel et al. 2005, Huemrich et al. 2010). However, if short-term
92 changes in water content result in significant changes in reflectance of such an important
93 ecosystem component as mosses, considerable uncertainty will be introduced into remote sensing-
94 derived estimates of green biomass and productivity. The role of moisture content on
95 photosynthesis and spectral reflectance has been investigated for a few moss species (Potter et
96 al. 1995, Van Breeman 1995, Dorrepaal et al. 2003, Proctor et al. 2007); however no studies
97 have addressed the role of water content on both of these properties simultaneously for
98 different Arctic bryophyte communities

99 Here, we investigate how variation in plant water content affects NDVI, red and near-
100 infrared reflectance, and gross primary productivity (GPP) of four moss communities through a
101 full cycle from saturation, dry down, and re-saturation. We hypothesize that NDVI will be
102 greatest at high but not fully saturated water contents and decrease with drying at rates
103 specific to each community type. Re-saturation will quickly re-establish initial NDVI values.
104 Gross Primary Productivity will peak at levels below full saturation, similar to those reported in

105 previous studies (Ueno and Kanda 2006, Van Gaalen et al. 2007, Harris 2008), and decrease
106 strongly with drying. We also expect that air drying of moss to constant weight will cause a
107 delay in re-establishment of initial GPP values after re-saturation.

108

109 **Methods**

110 ***Sample handling***

111 In two separate drying experiments, conducted in July (Exp 1) and August (Exp 2) 2016,
112 monoliths of four moss communities were collected from the low Arctic tundra near Toolik
113 Field Station (TFS) at Imnavait Creek, Alaska, USA (68.635° N, -149.349° W). The four
114 communities included two sphagnum communities, >95% *Sphagnum angustifolium* (green in
115 color), >95% *Sphagnum capilliofolium* (red in color), and two pleurocarpous communities,
116 >95% *Hylocomium splendens* and a mixed community (~50% *Aulacomnium* spp., ~30%
117 *Hylcomnium splendens*, and ~10% *Polytrichum* spp.). Four replicate 20x20x8 cm (length x width
118 x depth) blocks of each community were collected (16 total) for both experiments. Each
119 replicate was prepared by removing vascular plants and soil prior to placing them in a tray of
120 distilled water (3 cm depth) to hydrate. Moss blocks were soaked for two hours until they
121 reached full saturation and then allowed to drain for one hour to remove excess water. The
122 vertical faces of each moss block were wrapped in cellophane and then placed in a Styrofoam
123 tray to prevent uneven drying from the sides.

124 During Exp1, blocks were allowed to dry gradually in a translucent white outdoor tent to
125 allow temperatures (range 3-29°C) to track daily temperature changes. During Exp2, blocks
126 were allowed to dry gradually in the TFS Incubation Facility maintained at 23°C to minimize

127 temperature variability during the drying process. This temperature was determined to be a
128 suitable analog of natural peak season conditions (TFS Environmental Data Center, 2016).
129 Drying in the incubation facility took place under constant lighting using Hydrofarm® 1000 watt
130 lamps at a height of 1.5 m and $400 \mu\text{mol m}^{-2} \text{s}^{-1}$ (Hydrofarm, Inc., Petaluma, California, USA).
131 For both experiments, moss blocks reached ecologically dry status (air dry) after 84 (Exp2) and
132 96 (Exp1) hours and then were re-saturated with distilled water to determine drying resilience.

133 **Measurements**

134 All measurements during the drying process were taken at approximately 12 h intervals.
135 After rehydration, measurements were conducted at 4, 12 and 24 h intervals. Blocks were then
136 dried further at 50°C for 48 hours to achieve 0% water content allowing for calculation of
137 percent saturation at each measurement. To monitor water loss, block weight was measured
138 to 1 mg using an electronic balance (Ohaus Corporation, Parsippany, New Jersey, USA)
139 throughout the study period of both drying experiments. During Exp1, community reflectance
140 (350 – 1100 nm) was measured in ambient light conditions using a single channel Unispec® (PP
141 Systems, Amesbury Massachusetts, USA) at a height of 10 cm with NDVI calculated afterward
142 (see below). During the second experiment, NDVI was measured using a Trimble self-
143 illuminated, handheld GreenSeeker® crop sensing system (Trimble Navigation, Ltd., Sunnyvale,
144 California, USA) at a height of 15 cm, the minimum recommended distance. The GreenSeeker®
145 system with an internal light source was ideal for these measurements by providing accurate
146 NDVI values despite being under artificial light conditions. NDVI for both experiments was
147 calculated using the reflectance of near-infrared light (R_{774}) and red light (R_{656}) as $(R_{774} -$
148 $R_{656}) / (R_{774} + R_{656})$. Community CO_2 exchange for Exp2 was measured using a custom-made

149 transparent acrylic chamber (32 x 32 x 32 cm) with the moss block positioned on a hard flat
150 surface that formed a gas-tight seal with the chamber using weather stripping. One 12V fan
151 fixed inside the chamber insured full mixing of chamber air. The chamber was attached to a LI-
152 6400XT Portable Photosynthesis System (LI-COR Inc., Lincoln, Nebraska, USA) in closed system
153 mode. Gas exchange measurements were conducted under the Hydrofarm® 1000 watt lighting
154 systems at $900 \mu\text{mol m}^{-2} \text{s}^{-1}$. A good seal between the chamber and the flat surface was
155 determined if gas concentrations showed a steady rate of change with no fluctuations for 20 s.
156 Following the methods of Shaver et al. (2007), when a stable change in CO_2 concentrations was
157 observed, CO_2 concentration, chamber air temperature and PAR were logged every 2 s for 40 s.
158 Flux ($\mu\text{mol CO}_2 \text{s}^{-1}$) was calculated as a linear change in CO_2 concentration over time multiplied
159 by the air density (mol m^{-3}). Flux was expressed on an ecosystem area basis using the moss
160 surface area (0.04 m^{-2}). For each moss block, measurements were taken in the light (Net
161 Primary Production (NPP)) and in the dark under an opaque tarpaulin (Ecosystem Respiration
162 (ER)). Gross Primary Production (GPP) was calculated as the difference between NPP and ER,
163 assigning positive values to GPP and negative values to ER.

164

165 **Data Analysis**

166 Replicates of moss GPP and NDVI measurements were grouped by measurement time
167 and separately by 10% moisture content increments for statistical analysis. Time and moisture
168 content groupings were compared using a repeated measures analysis of variance with Tukey's
169 post-hoc analysis. In the second experiment, resilience of each moss community to re-hydration
170 after drying was determined using a drying response index (DRI) calculated as the proportion of

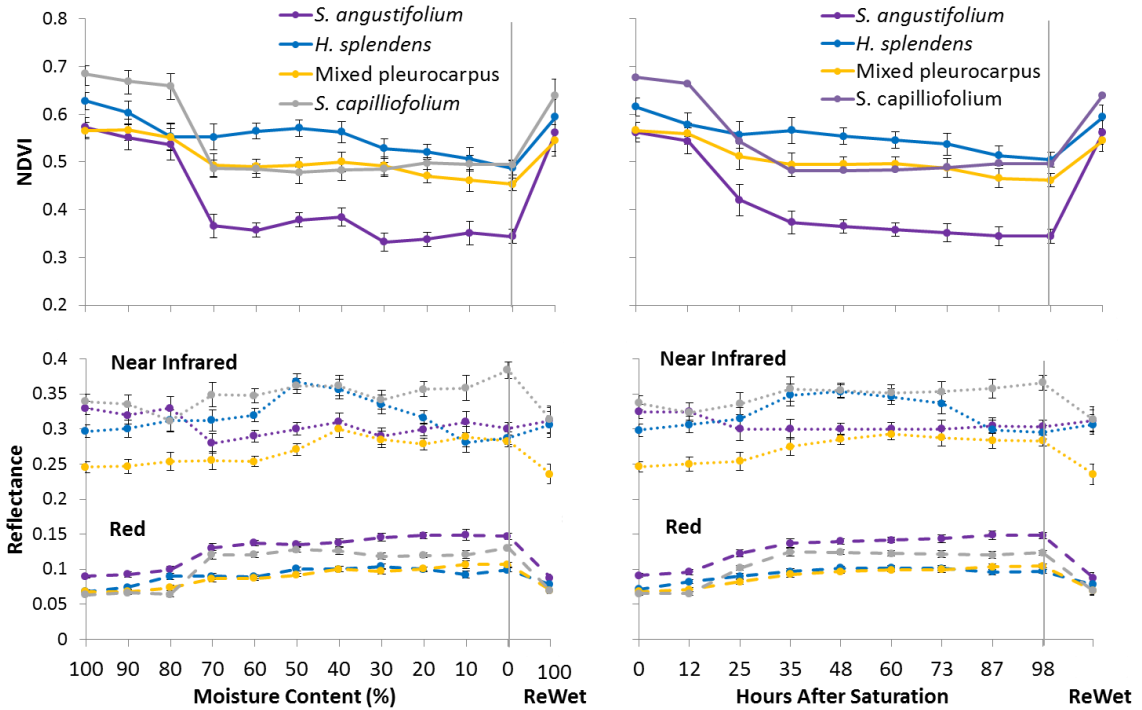
171 NDVI or GPP measured after re-saturation (V_{re-sat}) compared with the initial saturation values
172 ($V_{initial}$), $DRI = -(V_{re-sat}/V_{initial})$. All statistical tests were performed using the R statistical
173 environment (R Core Team, Vienna, Austria).

174

175 **Results**

176 During Exp1, all moss communities were fully air dried after 96 h with the largest
177 changes in NDVI occurring between 12 and 24 h of drying (Figure 1). Values of NDVI began to
178 decline with drying between 80 and 70% water content for all communities, but communities
179 differed in the magnitude of NDVI change. The sphagnum communities, *S. capilliofolium* and *S.*
180 *angustifolium*, showed the largest decline in NDVI with drying (-0.190, $p < 0.001$ and -0.230,
181 $p < 0.001$ respectively). Mixed pleurocarpus and *H. splendens* communities decreased in NDVI to
182 a lesser extent compared to the sphagnum communities but were still significant (-0.112,
183 $p = 0.021$ and -0.140, $p < 0.001$ respectively). All community NDVI values rebounded to near initial
184 saturation levels upon re-saturation.

185 Decreases in NDVI were largely driven by increases in reflectance of red light (Figure 1),
186 with the largest increases in the two sphagnum communities (*S. capilliofolium* +0.058 $p = 0.025$
187 and *S. angustifolium* +0.067 $p = 0.017$) compared with *H. splendens* (+0.039 $p = 0.231$) and the
188 mixed pleurocarpus (+0.038 $p = 0.234$). Near-infrared reflectance for all communities was mixed
189 with drying; some communities increased and some decreased, but no changes were significant
190 (0.009-0.045). Red and near-infrared reflectance returned to near initial saturation levels upon
191 re-saturation after only a few minutes.



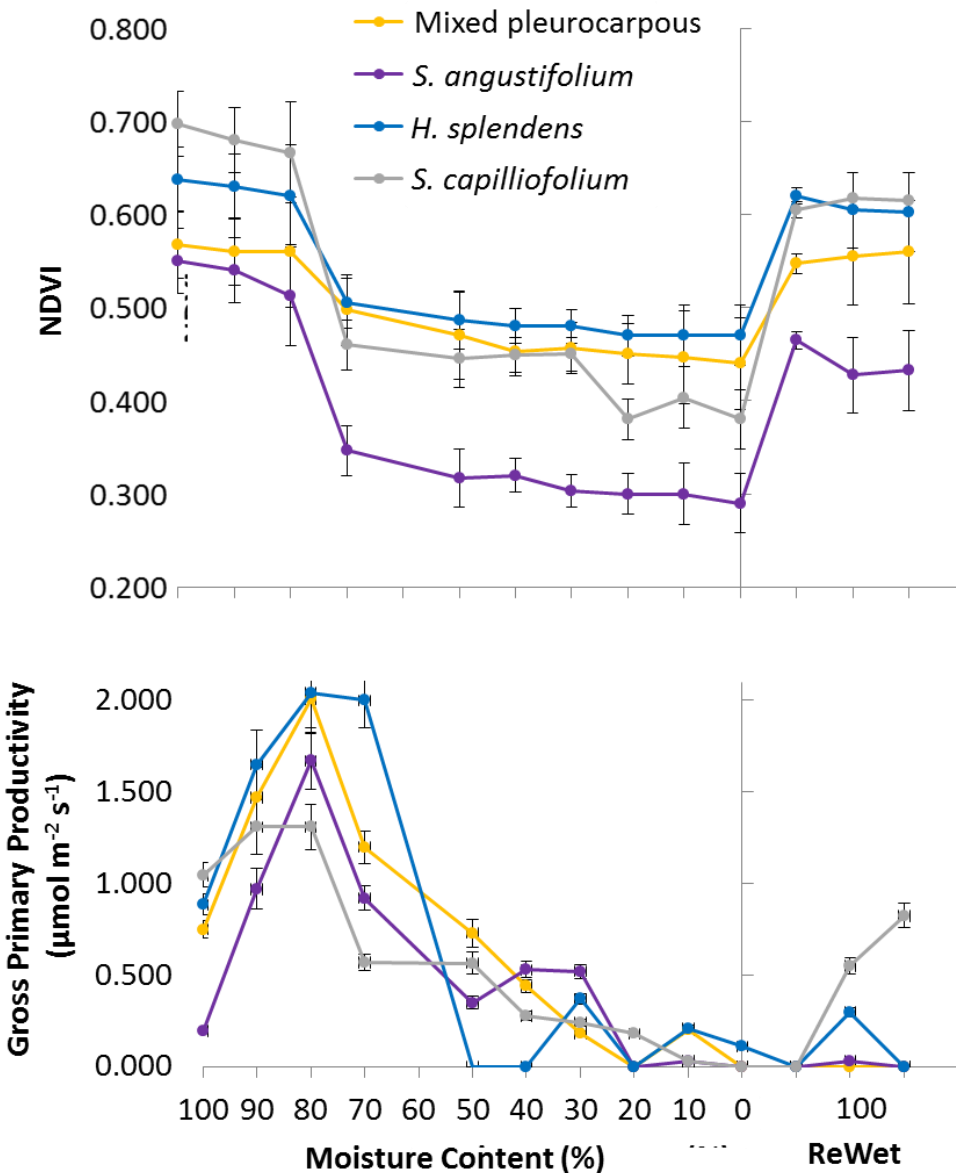
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193 **Figure 1:** NDVI (solid line), near-infrared (dotted line), and red (dashed line) of four
 194 communities by percent moisture content (left panels) and hours after initial saturation (right
 195 panels) during drying and after re-saturation during experiment 1.

196

197 In Exp2, all communities took approximately 84 hours to reach air dry and the lowest
 198 NDVI values (0.29 to 0.47, Figure 2). The largest decrease in NDVI observed as water contents
 199 declined occurred between 12 and 36 h post-saturation for all communities. *Sphagnum*
 200 *capilliofolium* and *H. splendens* communities had the highest NDVI measurements (0.70 and
 201 0.64 respectively) at initial full saturation, while the *S. angustifolium* and mixed pleurocarpus
 202 were lower (0.55 and 0.57, Figure 2, Table 1). The NDVI of all community types was generally
 203 stable from 100% to 80% saturation, followed by an abrupt decline between 80 and 70%, after
 204 which there was steady but slow decline in NDVI to fully air dry. The largest decreases in NDVI
 205 between 80% and 70% saturation were found for the two sphagnum communities (*S.*

206 *capilliofolium* -0.19, $P < 0.001$ and *S. angustifolium* -0.16, $P < 0.001$), while the decreases for the
 207 mixed pleurocarpus and *H. splendens* communities were less, albeit still significant (-0.06, $P =$
 208 0.038 and -0.09, $P = 0.014$ respectively). NDVI of all communities increased strongly upon re-
 209 saturation (+0.17 to +0.23, all $p < 0.001$).



210
 211 **Figure 2:** Gross primary productivity and NDVI of four communities by percent moisture
 212 content during drying and three measurement times after re-saturation during experiment 2.

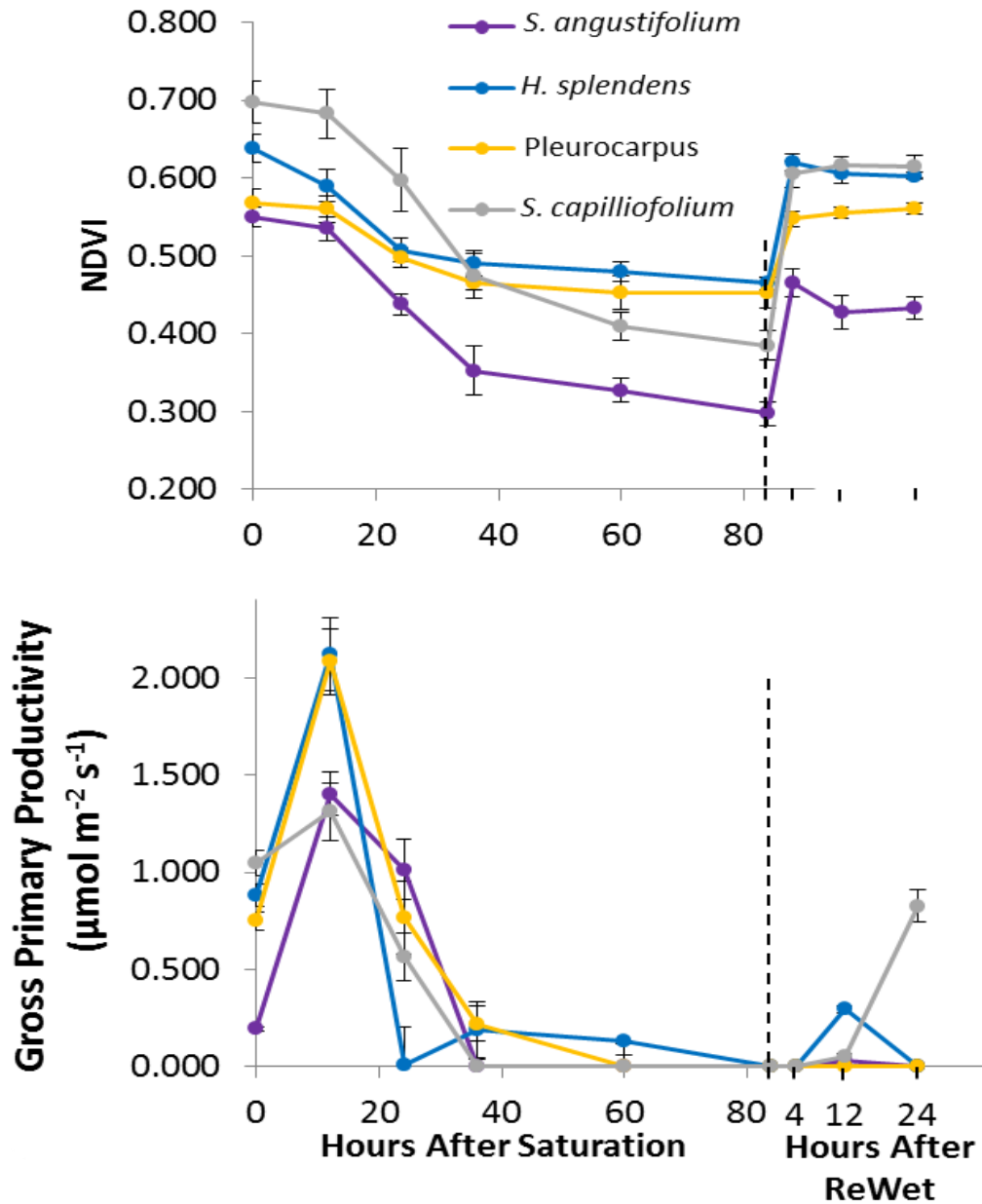
214 **Table 1:** Gross primary productivity and NDVI of four communities by percent moisture content during drying and three
 215 measurement times after re-saturation compared using repeated measures analysis of variance with Tukey's post-hoc analysis,
 216 along with the initial saturated and final oven dry weight for each of the four community types during experiment 2. Letters denote
 217 statistically significant differences ($p < 0.05$) between moisture content measurements.

| | Moisture Content | | | | | | | | | | ReWet (100% Moisture) | | | Initial | Dry |
|-------------------------|--|---------|---------|---------|----------|---------|---------|----------|---------|---------|-----------------------|---------|---------|-----------|--------|
| | 100 | 99-90 | 89-80 | 79-70 | 59-50 | 49-40 | 39-30 | 29-20 | 19-10 | 9-0 | Hours After | | | Saturated | Weight |
| | % | | | | | | | | | | 4 | 12 | 24 | Weight | Weight |
| | NDVI | | | | | | | | | | | | | (g) | (g) |
| Mixed Pleurocarpus | 0.57 a | 0.56 a | 0.56 a | 0.50 ab | 0.47 b | 0.45 b | 0.46 b | 0.45 b | 0.45 b | 0.44 b | 0.55 a | 0.56 a | 0.56 a | 578.5 | 113.8 |
| <i>S. angustifolium</i> | 0.55 a | 0.54 a | 0.51 a | 0.35 b | 0.32 b | 0.32 b | 0.30 b | 0.30 b | 0.30 b | 0.29 b | 0.47 a | 0.43 a | 0.43 a | 584.0 | 101.5 |
| <i>H. splendens</i> | 0.64 a | 0.63 a | 0.62 a | 0.51 b | 0.49 b | 0.48 b | 0.48 b | 0.47 b | 0.47 b | 0.47 b | 0.62 a | 0.61 a | 0.60 a | 403.3 | 116.2 |
| <i>S. capillifolium</i> | 0.70 a | 0.68 a | 0.67 a | 0.46 b | 0.45 b | 0.45 b | 0.45 b | 0.38 c | 0.40 bc | 0.38 c | 0.61 a | 0.62 a | 0.62 a | 753.3 | 144.6 |
| | Gross Primary Productivity ($\mu\text{mol m}^{-2}\text{s}^{-1}$) | | | | | | | | | | | | | | |
| Mixed Pleurocarpus | 0.749 a | 1.472 b | 2.081 c | 1.207 b | 0.727 a | 0.443 d | 0.175 e | 0.000 f | 0.202 e | 0.000 f | 0.000 f | 0.000 f | 0.000 f | | |
| <i>S. angustifolium</i> | 0.195 a | 0.970 b | 1.671 c | 0.913 b | 0.352 ad | 0.526 d | 0.521 d | 0.000 e | 0.033 e | 0.000 e | 0.000 e | 0.034 e | 0.000 e | | |
| <i>H. splendens</i> | 0.883 a | 1.652 b | 2.042 c | 1.608 b | 0.002 d | 0.000 d | 0.374 e | 0.000 d | 0.206 e | 0.111 d | 0.000 d | 0.296 e | 0.000 d | | |
| <i>S. capillifolium</i> | 1.046 a | 1.320 b | 1.311 b | 0.571 c | 0.565 c | 0.277 d | 0.235 d | 0.184 de | 0.031 e | 0.000 e | 0.000 e | 0.552 c | 0.824 f | | |

218

219

220 All moss communities were photosynthesizing at full water saturation (0.195 to 1.046
221 $\mu\text{mol m}^{-2} \text{s}^{-1}$) and initially increased as drying began (Figure 2, Table 1). *Sphagnum capilliofolium*
222 community GPP peaked at approximately 90% saturation ($1.320 \mu\text{mol m}^{-2} \text{s}^{-1}$) while all other
223 communities peaked at 80% saturation (mixed pleurocarpus $1.332 \mu\text{mol m}^{-2} \text{s}^{-1}$, *S.*
224 *angustifolium* $1.476 \mu\text{mol m}^{-2} \text{s}^{-1}$, *H. splendens* $1.159 \mu\text{mol m}^{-2} \text{s}^{-1}$). Rates of GPP for all
225 communities decreased precipitously below 80% saturation with little or no GPP occurring at
226 fully air dry (0 to $0.111 \mu\text{mol m}^{-2} \text{s}^{-1}$). Rates of GPP peaked for all communities after 12 hours of
227 drying following the initial saturation and then continued to decline until a total of 84 hours of
228 drying. Twelve hours after re-saturation, *S. angustifolium* ($0.034 \mu\text{mol m}^{-2} \text{s}^{-1}$), *H. splendens*
229 ($0.296 \mu\text{mol m}^{-2} \text{s}^{-1}$), and *S. capilliofolium* ($0.552 \mu\text{mol m}^{-2} \text{s}^{-1}$) had regained some GPP (Figure 3).
230 Twenty four hours post saturation only the *S. capilliofolium* community showed any GPP (0.824
231 $\mu\text{mol m}^{-2} \text{s}^{-1}$).



232

233 **Figure 3:** Gross primary productivity and NDVI of four communities by hours of drying and three
 234 measurement times after re-saturation during experiment 2.

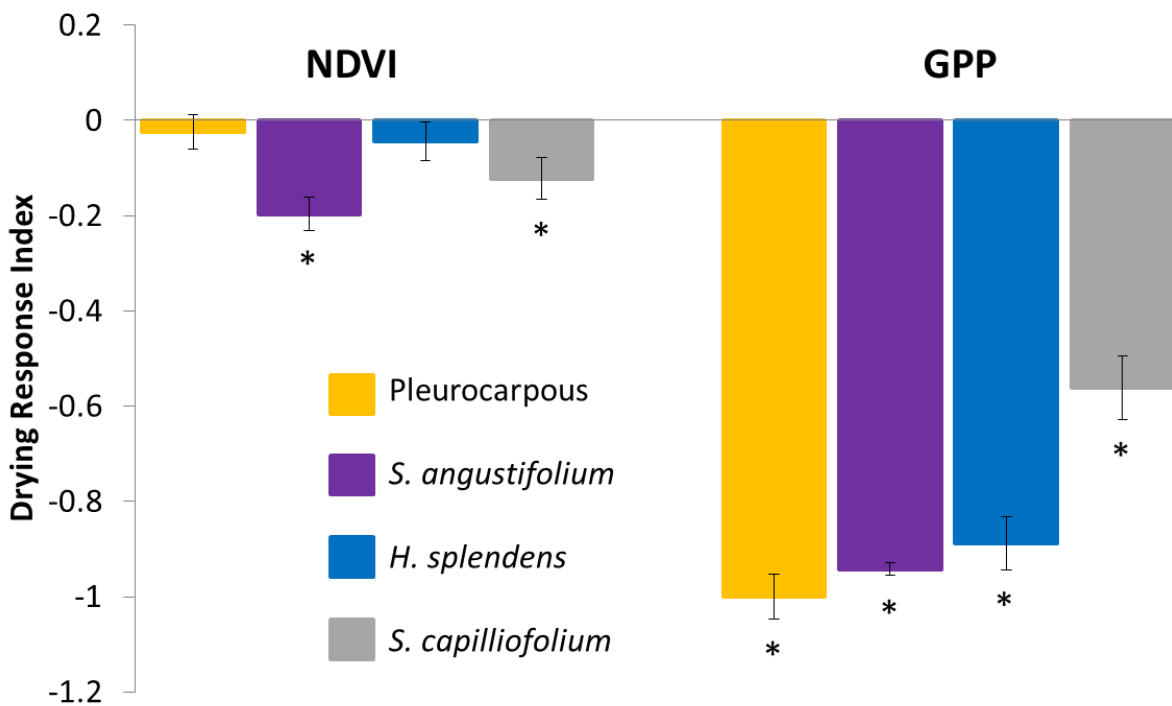
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236 Measuring drying resilience by means of the DRI showed that NDVI and GPP decreased

237 in all community types after re-saturation compared with the initial comparison (Figure 4).

238 Drying response index for NDVI of the mixed pleurocarpus and *H. splendens* communities

239 decreased only slightly (-0.023 , $P = 0.101$ and -0.044 , $P = 0.081$ respectively), returning to near
 240 original saturation values. The DRI for NDVI for both sphagnum communities had significant
 241 declines in response to full drying (*S. angustifolium* -0.197 , $P = 0.013$ and *S. capilliofolium*
 242 -0.122 , $P = 0.042$). The DRI for GPP of all communities decreased after drying with mixed
 243 pleurocarpus (-1.000 , $P < 0.001$), *S. angustifolium* (-0.942 , $P < 0.001$), and *H. splendens* (-0.888 ,
 244 $P < 0.001$) declining the most and the *S. capilliofolium* community being the most resilient to
 245 drying (-0.562 , $P = 0.013$).



246
 247 **Figure 4:** Drying response indexes of four communities for GPP and NDVI during experiment 2
 248 compared using a one-way analysis of variance. Statistical significance denoted with *.

249
 250 **Discussion**

251 Previous studies have shown that drying alters spectral reflectance (Riedel et al. 2005,
 252 Huemmrich et al. 2012) and moss productivity (Skre and Oechel 1981, Ueno and Kanda 2006,

253 Harris 2008). Here we show that the reflectance index, NDVI, and GPP decline strongly with
254 moss desiccation, but they do not occur at the same rate and magnitude, resulting in a
255 mismatch between NDVI levels and productivity. Reductions in NDVI of the moss communities
256 with desiccation were very large, approaching 50% of maximum values and driven mostly by
257 increases in red light reflectance during drying. With rewetting after the strong dry down, NDVI
258 values were near peak levels while GPP was near zero.

259 ***Moss water content effect on NDVI values***

260 In recent decades, measured NDVI values have been increasing in the Arctic as
261 temperatures warm and ecosystem productivity increases (Jia et al. 2003). These changes in
262 peak season NDVI values, however, have not been increasing uniformly across spatial and
263 temporal time scales, with evidence of a slowing of the rate of increase (Bhatt et al. 2013). The
264 heterogeneity of changes in peak season NDVI values may be a result of the non-uniformity of
265 the well-documented community dominance and moss decline in the Arctic (Shaver and Chapin
266 1991, Douma et al. 2007, Campioli et al. 2009, Elmendorf et al. 2012, Hollister et al. 2015).
267 Moss communities are often an important component of Arctic understories and have been
268 shown to play a large role in community production (Duoma et al. 2007, Campioli et al. 2009)
269 and remotely-sensed spectral measurements (Walker et al 2003). Our results show that even
270 small changes in the water content of the moss understory may play a role in the slowing of
271 changes in landscape scale NDVI. Warmer, drier conditions during peak growing season may
272 artificially decrease estimates of peak season landscape scale NDVI estimates.

273 These results have important implications for remote sensing of plant biomass and
274 productivity in regions where mosses are important components of the vegetation. A general

275 assumption in the use of NDVI to estimate green biomass of plants is that NDVI is not strongly
276 affected by short-term changes in leaf water content. While this assumption is generally the
277 case for vascular plants, our results show that changes in moss water content can induce rapid
278 and large changes in NDVI with no change in biomass. Furthermore, the relationship between
279 NDVI and water content is markedly nonlinear. Variation in the water content of moss may be
280 an important source of error in models using NDVI to estimate green biomass or leaf area
281 (Oechel et al. 2000, Vourlitis et al. 2000, Shaver et al. 2007) that is then used in ecosystem
282 photosynthesis models.

283 All of the moss communities in this study followed similar patterns of NDVI reductions
284 with drying, although the magnitude of NDVI change in response to drying was community-
285 specific. As predicted, all communities had the highest NDVI values at full, initial saturation (80-
286 100% moisture content) with marked NDVI declines with drying. NDVI of all communities
287 declined sharply over a relatively narrow range of water content from 80-70% moisture
288 content, with the most substantial declines found for the two sphagnum communities. The
289 lower NDVI values and higher levels of red reflectance at lower moisture content levels may act
290 as a mechanism to minimize absorption of irradiance to prevent further evaporative water loss
291 or cellular damage (Charron and Quatrano 2009, de Carvalho et al. 2012). The *H. splendens*-
292 dominated and pleurocarpus mixed communities showed moderate increases in NDVI upon re-
293 saturation. In contrast, both sphagnum communities had abrupt (<2 minutes), significant
294 increases in NDVI values upon re-saturation. The rapidity of NDVI increases upon re-saturation
295 of sphagnum communities suggests that changes in NDVI with drying and rehydration are in
296 part a physical rather than biological response. Despite the rapid recovery of NDVI values of

297 sphagnum communities upon rewetting, values did not attain those of initial saturation, unlike
298 the pleurocarpus mixed and *H. splendens* dominated communities that recovered fully. This
299 lack of full rebound in sphagnum communities may be a result of, at least temporary,
300 physiological damage occurring in response to desiccation to fully air dry (Oliver et al. 2005;
301 Hájek and Beckett, 2008).

302 ***Moss water content effect on GPP***

303 Rapid changes in NDVI of moss communities with water content are associated with
304 large changes in GPP, albeit nonlinearly. This variability is in addition to the already substantial
305 difference in photosynthesis rates between vascular plants and mosses (Longton 1988). At
306 Barrow, Alaska, production rates of mosses are on the order 10% of that of vascular plants
307 (Oechel and Sveinbjornsson 1978), which means that photosynthesis per unit NDVI are very
308 different for vascular plants compared to those of mosses. Our results show that moss
309 communities may have relatively high NDVI values (0.55- 0.70), that if interpreted as vascular
310 plant biomass would lead to large overestimates in productivity. These mismatches
311 compromise the use of remotely-sensed NDVI data to estimate productivity in communities
312 where mosses are abundant, but information on local moisture content or precipitation are
313 lacking. Models using NDVI as a measure of productivity through estimating productivity by
314 metric such as leaf area index (LAI) are highly effective across a range of spatial scales (Shaver
315 et al. 2007, Loranty et al. 2011, Stoy et al. 2013). As spatial scale and vascular plant cover
316 increases, the proportion of moss contribution to community spectral measurements is likely to
317 decrease.

318 The magnitudes of changes in moss community GPP rates with drying were also
319 community specific. However, all communities showed moderate rates at initial saturation and
320 increased with drying to around 70-80% moisture content. This pattern of lower productivity at
321 full saturation and increasing productivity after initial drying begins is similar to results found by
322 Van Gaalen et al. (2007). A moderate amount of drying allows for air space within the plant
323 while allowing cells to retain adequate moisture for full function. All communities had a peak
324 GPP at 70-80% moisture content. Drying below 70-80% moisture content caused incremental
325 decreases in GPP dropping to near zero in all of the communities when they reached air dry.
326 Re-saturation had minimal effects on GPP, a finding consistent with previous findings that
327 showed delayed recovery of moss physiological activity with rewetting after drying (Van
328 Breeman 1995). Only the *S. capillifolium* community showed a recovery of GPP during the 24
329 hours after re-saturation.

330 Moss communities such as those in this study are often intermixed at relatively small
331 spatial scales across Arctic terrestrial ecosystems, implying a heterogeneous matrix of drying
332 and recovery responses. While all four communities showed strong reduction in NDVI at the
333 80% drying threshold, the responses of both sphagnum communities were substantially greater
334 than those in the pleurocarpus moss communities. To use remotely-sensed, reflectance-based
335 productivity monitoring of Arctic ecosystems, further investigation is needed on the effects of
336 intra-seasonal drying and rehydration on productivity and spectral reflectance of different moss
337 communities.

338 These results in response in moss moisture content highlight the need for repeated
339 remote sensing measurements over the same study regions with monitoring of a region's

340 recent precipitation events. Because of the remoteness and scale of Arctic regions, remotely
341 sensed data are currently the best means to investigate seasonal productivity and vegetation
342 composition shifts associated with climate change (Raynolds et al. 2008, Bhatt et al. 2010, Stow
343 et al. 2007, Walker et al. 2012). This issue is crucial in Arctic regions where mosses comprise a
344 major vegetation component, contribute substantially to ecosystem productivity (Olivas et al
345 2011), and are often a large component of total community reflectance (Hope et al. 1993). Our
346 results show that periods of little or no precipitation combined with clear skies, high
347 temperature and windy conditions have the potential to rapidly (<24 hours) lower moss water
348 content sufficiently to reduce ecosystem NDVI values that would imply low predictions of
349 ecosystem productivity even though vascular plant productivity may remain high. Remotely-
350 sensed NDVI values measured for the same area shortly before and after a precipitation event
351 may differ simply in response to moss moisture content.

352 Conditions conducive to moss desiccation are expected to increase with climate
353 warming as temperatures increase, driving greater evapotranspiration. These changes will
354 increase the frequency of moisture-induced changes in NDVI. Mosses grow in many different
355 conditions ranging from on the surface of mineral soil or even on bare rock to areas that remain
356 nearly continually wet or submerged. The frequency at which mosses desiccate is dependent in
357 part on the microtopographic conditions where they are growing as well as weather conditions.
358 Those growing on well-drained mineral soil or rock surfaces and hummocks are likely to
359 desiccate frequently, whereas others may rarely if ever desiccate. Species colonizing conditions
360 subject to frequent desiccation are likely to tolerate desiccation better than species in areas
361 that rarely dry out (Longton 1988). Sites where mosses are continually wet are less likely to

362 show rapid NDVI changes in response to drying, but species from these conditions may be more
363 susceptible to climate change-related drying in the long term.

364 **Conclusion**

365 This study reinforces the importance of understanding the moisture content of moss
366 when using remotely-sensed, reflectance techniques for monitoring productivity in Arctic
367 terrestrial systems. Reflectance measures of different communities of moss revealed species-
368 specific variation in response and resiliency to drying, therefore complicating the aggregation of
369 moss as a uniform understory in Arctic ecosystems. At similar NDVI values, GPP varied
370 depending on moss moisture content, demonstrating that moss NDVI is not an accurate proxy
371 for physiological activity of some important Arctic mosses. This study underscores the need for
372 monitoring and understanding the composition, spatial coverage, and moisture content of
373 mosses for remote sensing-based monitoring of Arctic terrestrial ecosystems. Methodologies
374 for remotely monitoring surface water content (e.g. Normalized Difference Water Index (NDWI)
375 (Gao 1996), Normalized Difference Infrared Index (NDII) (Serrano et al. 2000), among others)
376 are improving and could be useful for addressing these issues

377

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385 **Literature Cited**

- 386 Alm, J., L. Schulman, J. Walden, H. Nykänen, P. J. Martikainen and J. Silvola. 1999 Carbon
387 balance of a boreal bog during a year with an exceptionally dry summer. *Ecology*.
388 80:161–174.
- 389 Andresen, C. G. and V. L. Lougheed. 2015. Disappearing Arctic tundra ponds: Fine-scale analysis
390 of surface hydrology in drained thaw lake basins over a 65 year period (1948–2013).
391 *Journal of Geophysical Research: Biogeosciences*, 120(3): 466-479.
- 392 Bhatt, U. S., D. A. Walker, M. K. Reynolds, J. C. Comiso, H. E. Epstein, G. Jia, R. Gens, J. E. Pinzon,
393 C. J. Tucker, C. E. Tweedie, and P. J. Webber, 2010. Circumpolar Arctic tundra vegetation
394 change is linked to sea ice decline. *Earth Interactions*. 14(8): 1-20.
- 395 Bhatt, U.S., D. A. Walker, M. K. Reynolds, P. A. Bieniek, H. E. Epstein, J. C. Comiso, J. E. Pinzon, C.
396 J. Tucker, I. V. Polyakov. 2013. Recent declines in warming and vegetation greening
397 trends over pan-Arctic tundra. *Remote Sensing*. 5(9):4229-4254.
- 398 Campioli, M., R. Samson, A. Michelsen, S. Jonasson, R. Baxter, and R. Lemeur. 2009.
399 Nonvascular contribution to ecosystem NPP in a subarctic heath during early and late
400 growing season. *Plant Ecology* 202: 41–53.
- 401 Chapin, F. S., G. R. Shaver, A. E. Giblin, K. J. Nadelhoffer, J. A. and Laundre. 1995. Responses of
402 arctic tundra to experimental and observed changes in climate. *Ecology*. 76(3): 694-711.
- 403 Chapin, F. S., M. Sturm, M. C. Serreze, J. P. McFadden, J. R. Key, A. H. Lloyd, A. D. McGuire, T. S.
404 Rupp, A. H. Lynch, J. P. Schimel. and J. Beringe. 2005. Role of land-surface changes in
405 Arctic summer warming. *Science*. 310(5748): 657-660
- 406 Charron, A. J. and R. S. Quatrano. 2009. Between a rock and a dry place: the water-stressed
407 moss. *Molecular Plant*. 2(3): 478-486.
- 408 Chen, W., J. Li, Y. Zhang, F. Zhou, K. Khoeler, S. LeBlanc, R. Fraser, I. Olthof, Y. S. Zhang, and J.
409 Wang. 2009. Relating biomass and leaf area index to non-destructive measurements in
410 order to monitor changes in Arctic vegetation. *Arctic* 62:281-294.
- 411 de Carvalho, R. C., M. Catalá, J. M. da Silva, C. Branquinho. and E. Barreno. 2012. The impact of
412 dehydration rate on the production and cellular location of reactive oxygen species in an
413 aquatic moss. *Annals of Botany*. p.mcs. 180.
- 414 Dorrepaal, E., R. Aerts, J. H. Cornelissen, T. V. Callaghan, and R. S. Van Logtestijn. 2004. Summer
415 warming and increased winter snow cover affect *Sphagnum fuscum* growth, structure
416 and production in a sub-arctic bog. *Global Change Biology*. 10(1): 93-104.
- 417 Douma, J. C., M. T. Van Wijk, S. I. Lang, and G. R. Shaver. 2007. The contribution of mosses to
418 the carbon and water exchange of arctic ecosystems: quantification and relationships
419 with system properties. *Plant, Cell and Environment* 30: 1205–1215.

420 Elmendorf, S. C., G. H. R. Henry, R. D. Hollister, R. G. Bjork, N. Boulanger-Lapointe, E. J. Cooper,
421 J. H. C. Cornelissen, T. A. Day, E. Dorrepaal, T. G. Elumeeva, M. Gill, W. A. Gould, J. Harte,
422 D. S. Hik, A. Hofgaard, D. R. Johnson, J. F. Johnstone, I. S. Jonsdottir, J. C. Jorgenson, K.
423 Klanderud, J. A. Klein, S. Koh, G. Kudo, M. Lara, E. Levesque, B. Magnusson, J. L. May, J.
424 A. Mercado-Diaz, A. Michelsen, U. Molau, I. H. Myers-Smith, S. F. Oberbauer, V. G.
425 Onipchenko, C. Rixen, N. M. Schmidt, G. R. Shaver, M. J. Spasojevic, P. E. Porhallsdottir,
426 A. Tolvanen, T. Troxler, C. E. Tweedie, S. Villareal, C.-H. Wahren, X. Walker, P. J. Webber,
427 J. M. Welker, and S. Wipf. 2012. Plot-scale evidence of tundra vegetation change and
428 links to recent summer warming. *Nature Climate Change* 2:453–457.

429 Gamon, J. A., K. F. Huemmrich, R. S. Stone, C. E. Tweedie. 2013. Spatial and temporal variation
430 in primary productivity (NDVI) of coastal Alaskan tundra: Decreased vegetation growth
431 following earlier snowmelt. *Remote Sensing of Environment*. 129: 144-153.

432 Gao, B.C., 1996. NDWI—A normalized difference water index for remote sensing of vegetation
433 liquid water from space. *Remote sensing of environment*. 58(3):257-266.

434 Green, T. G. A. and O. L. Lange. 1995. Photosynthesis in poikilohydric plants: a comparison of
435 lichens and bryophytes. In *Ecophysiology of Photosynthesis*. 319-341. Springer Berlin
436 Heidelberg.

437 Hájek, T., and R. P. Beckett. 2008. Effects of water content components on desiccation and
438 recovery of Sphagnum mosses. *Annals of Botany*. 101(1): 165-173.

439 Harris, A. 2008. Spectral reflectance and photosynthetic properties of Sphagnum mosses
440 exposed to progressive drought. *Ecohydrology*. 1(1): 35-42.

441 Hinzman, L. D., N. D. Bettez, W. R. Bolton, F. S. Chapin, M. B. Dyurgerov, C. L. Fastie, B. Griffith,
442 R. D. Hollister, A. Hope, H. P. Huntington, and A. M. Jensen. 2005. Evidence and
443 implications of recent climate change in northern Alaska and other arctic
444 regions. *Climatic Change*. 72(3): 251-298.

445 Hobbie, J. E., G. R. Shaver, E. B. Rastetter, J. E. Cherry, S. J. Goetz, K. C. Guay, W. A. Gould, and
446 G. W. Kling. 2017. Ecosystem responses to climate change at a Low Arctic and a High
447 Arctic long-term research site. *Ambio* 46:160–173.

448 Hollister, R. D., J. L. May, K. S. Kremers, C. E. Tweedie, S. F. Oberbauer, J. A. Liebig, T. F. Botting,
449 R. T. Barrett, and J. L. Gregory. 2015. Warming experiments elucidate the drivers of
450 observed directional changes in tundra vegetation. *Ecology and evolution*. 5(9): 1881-
451 1895.

452 Hope, A. S., J. S. Kimball, and D. A. Stow. 1993. The relationship between tussock tundra
453 spectral reflectance properties and biomass and vegetation composition. *International
454 Journal of Remote Sensing*. 14: 1861-1874.

455 Hope, A. S., J. B. Fleming, G. Vourlitis, D. A. Stow, W. C. Oechel. T. and Hack. 1995. Relating CO₂
456 fluxes to spectral vegetation indices in tundra landscapes: importance of footprint
457 definition. *Polar Record*. 31(177): 245-250.

458 Huemmrich, K. F., J. A. Gamon, C. E. Tweedie, S. F. Oberbauer, G. Kinoshita, S. Houston, A.
459 Kuchy, R. D. Hollister, H. Kwon, M. Mano, and Y. Harazono. 2010. Remote sensing of
460 tundra gross ecosystem productivity and light use efficiency under varying temperature
461 and moisture conditions. *Remote Sensing of Environment*. 114(3): 481-489.

462 Intergovernmental Panel on Climate Change Fifth Assessment Report: Climate Change. 2013.
463 The physical basis. Working group I contribution to the fifth assessment report of the
464 intergovernmental panel on climate change. T. Stocker, D. Qin, G.-K. Plattner, M. Tignor,
465 S. Allen, J. Boschung, A. Nauels, Y. Xia, V. Bex, P. Midgley (eds). Cambridge University
466 Press. Cambridge, UK.

467 Jia, G.J.; H. E. Epstein, H. E.; Walker, D. A. 2003. Greening of Arctic Alaska, 1981-2001.
468 *Geophysical Research Letters*. 30(20):2067.

469 Kattsov, V. M., Walsh, J. E., Chapman, W. L., Govorkova, V. A., Pavlova, T. V., &
470 Zhang, X. (2007). Simulation and projection of Arctic freshwater budget components by
471 the IPCC AR4 global climate models. *Journal of Hydrometeorology*, 8(3), 571-589.

472 Kerr, J. T., & Ostrovsky, M. (2003). From space to species: ecological applications for
473 remote sensing. *Trends in Ecology & Evolution*, 18(6), 299-305.

474 Komulainen, V.-M., E.-S. Tuittila, H. Vasander and J. Laine. 1999. Restoration of drained
475 peatlands in southern Finland: initial effects on vegetation change and CO₂ balance.
476 *Journal of Applied Ecology*. 36: 634–648.

477 Kriegler, F.J., W. A. Malila, R. F. Nalepka, W. Richardson. 1969. Preprocessing transformations
478 and their effects on multispectral recognition. *Proceedings of the Sixth International
479 Symposium on Remote Sensing of Environment*. University of Michigan, Ann Arbor, MI.
480 97-131.

481 Kushida, K., Y. Kim, S. Tsuyuzaki, and M. Fukuda. 2009. Spectral vegetation indices for
482 estimating shrub cover, green phytomass and leaf turnover in a sedge-shrub tundra.
483 *International Journal of Remote Sensing* 30:1651-1658.

484 Laidler, G. J., P. M. Treitz, and D. M. Atkinson. 2008. Remote sensing of arctic vegetation:
485 relations between the NDVI, spatial resolution and vegetation cover on Boothia
486 Peninsula, Nunavut. *Arctic*. 61:1-13.

487 Levitt, J. 1956. The hardiness of plants. *Soil Science*. 82(4): 346.

488 Longton, R. E. 1988. Adaptations and strategies of polar bryophytes. *Botanical journal of the
489 Linnean Society*. 98(3): 253-268.

490 Loranty, M.M., S. J. Goetz, P. S. Beck, 2011. Tundra vegetation effects on pan-Arctic albedo.
491 *Environmental Research Letters*. 6(2):024014.

492 Mack, M. C., M. S. Bret-Harte, T. N. Hollingsworth, R. R. Jandt, E. A. Schuur, G. R. Shaver, and D.
493 L. Verbyla. 2011. Carbon loss from an unprecedented Arctic tundra wildfire. *Nature*.
494 475(7357): 489-492.

495 McNeil, P. and J. M. Waddington. 2003. Moisture controls on *Sphagnum* growth and CO₂

496 exchange on a cutover bog. *Journal of Applied Ecology*. 40: 354-367.

497 Molau, U. and J. M. Alatalo. 1998. Responses of subarctic-alpine plant communities to
 498 simulated environmental change: biodiversity of bryophytes, lichens, and vascular
 499 plants. *Ambio*. 322-329.

500 Oechel, W. C. and B. Sveinbjornsson. 1978. Primary production processes in Arctic bryophytes
 501 at Barrow, Alaska. In L. L. Tieszen (Ed.), *Vegetation Production and Production Ecology*
 502 *of an Alaskan Arctic Tundra*: 269-298. New York: Springer-Verlag.

503 Oechel, W. C., S. J. Hastings, G. Vourlitis, M. Jenkins, G. Riechers, and N. Grulke. 1992. Recent
 504 change of Arctic tundra ecosystems from a net carbon dioxide sink to a source. *Nature*.
 505 361(6412): 520-523.

506 Oechel, W. C., G. L. Vourlitis, J. Verfaillie, T. Crawford, S. Brooks, E. Dumas, A. Hope, D. Stow, B.
 507 Boynton, V. Nosov, and R. Zulueta. 2000. A scaling approach for quantifying the net CO₂
 508 flux of the Kuparuk River Basin, Alaska. *Global Change Biology*. 6(S1): 160-173.

509 Olivas, P. C., S. F. Oberbauer, C. E. Tweedie, W. C. Oechel, D. Lin, and A. Kuchy. 2011. Effects of
 510 fine-scale topography on CO₂ flux components of Alaskan coastal plain tundra:
 511 Response to contrasting growing seasons. *Arctic, Antarctic, and Alpine Research*.
 512 43(2): 256-266.

513 Oliver, M. J. and J. D. Bewley. 1984. Plant Desiccation and Protein Synthesis V. Stability of Poly
 514 (A)⁻ and Poly (A)⁺ RNA during desiccation and their synthesis upon rehydration in the
 515 desiccation-tolerant moss *Tortula ruralis* and the intolerant moss *Cratoneuron filicinum*.
 516 *Plant Physiology*. 74(4): 917-922.

517 Oliver, M. J., J. Velten, B. D. Mishler. 2005. Desiccation tolerance in bryophytes: A reflection of
 518 the primitive strategy for plant survival in dehydrating habitats? *Integrative and*
 519 *Comparative Biology*. 45(5): 788-799.

520 Overland J. E., M. Wang, and N. A. Bond. 2002. Recent Temperature Changes in the Western
 521 Arctic during Spring. *Journal of Climate*. 15(13): 1702-1716.

522 Potter, J. A., M. C. Press, T. V. Callaghan, and J. A. Lee. 1995. Growth responses of *Polytrichum*
 523 *commune* and *Hylocomium splendens* to simulated environmental change in the sub
 524 arctic. *New Phytologist*. 131(4): 533-541.

525 Proctor, M.C. and Z. Tuba. 2002. Poikilohydry and homoihydry: antithesis or spectrum of
 526 possibilities?. *New Phytologist*. 156(3): 327-349.

527 Proctor, M. C., M. J. Oliver, A. J. Wood, P. Alpert, L. R. Stark, N. L. Cleavitt, and B. D. Mishler.
 528 2007. Desiccation-tolerance in bryophytes: a review. *The Bryologist*. 110(4): 595-621.

529 R Core Team (2017). R: A language and environment for statistical computing. R Foundation for
 530 Statistical Computing, Vienna, Austria. URL <https://www.R-project.org/>.

531 Reynolds, M. K., J. C. Comiso, D. A. Walker and D. Verbyla. 2008. Relationship between
 532 satellite-derived land surface temperatures, arctic vegetation types, and NDVI. *Remote*
 533 *Sensing of Environment*. 112(4): 1884-1894.

534 Riedel, S. M., H. E. Epstein, and D. A. Walker. 2005. Biotic controls over spectral reflectance of
535 arctic tundra vegetation. *International Journal of Remote Sensing*. 26(11): 2391-2405.

536 Riordan, B., D. Verbyla, and A. D. McGuire. 2006. Shrinking ponds in subarctic Alaska based on
537 1950–2002 remotely sensed images. *Journal of Geophysical Research:*
538 *Biogeosciences*, 111(G4).

539 Roulet, N., T. I. M. Moore, J. Bubier, and P. Lafleur. 1992. Northern fens: methane flux and
540 climatic change. *Tellus B*. 44(2): 100-105.

541 Rydin, H. and A. J. S. McDonald. 1985. Photosynthesis in *Sphagnum* at different water contents.
542 *Journal of Bryology*. 13: 579-584.

543 Schipperges, B. and H. Rydin. 1998. Response of photosynthesis of *Sphagnum*
544 species from contrasting microhabitats to tissue water content and repeated
545 desiccation. *New Phytologist*. 149: 677-684.

546 Serrano, L., S. L. Ustin, D. A. Roberts, J. A. Gamon, J. Penuelas, 2000. Deriving water content of
547 chaparral vegetation from AVIRIS data. *Remote Sensing of Environment*, 74(3):570-581.

548 Shaver, G. R., and F. S. Chapin III. 1991. Production: biomass relationships and element cycling
549 in contrasting Arctic vegetation types. *Ecological Monographs* 61: 1–31.

550 Shaver, G. R., L. E. Street, E. B. Rastetter, M. T. Van Wijk, and M. Williams. 2007. Functional
551 convergence in regulation of net CO₂ flux in heterogeneous tundra landscapes in Alaska
552 and Sweden. *Journal of Ecology*. 95(4): 802-817.

553 Skre, O. and W. C. Oechel. 1981. Moss functioning in different taiga ecosystems in interior
554 Alaska. *Oecologia*. 48(1): 50-59.

555 Stow, D., A. Petersen, A. Hope, R. Engstrom, and L. Coulter. 2007. Greenness trends of
556 Arctictundra vegetation in the 1990s: comparison of two NDVI data sets from NOAA
557 AVHRR systems. *International Journal of Remote Sensing*. 28(21): 4807-4822.

558 Stoy, P.C., M. Williams, J. G. Evans, A. Prieto-Blanco, M. Disney, T. C. Hill, H. C. Ward, T. J. Wade,
559 L. E. Street. 2013. Upscaling tundra CO₂ exchange from chamber to eddy covariance
560 tower. *Arctic, Antarctic, and Alpine Research*. 45(2):275-284.

561 Titus, J. E., D. J. Wagner, and M. D. Stephens. 1983. Contrasting water relations of
562 photosynthesis for two *Sphagnum* mosses. *Ecology*. 64(5): 1109-1115.

563 Toolik Field Station, Environmental Data Center. 2016.
564 https://toolik.alaska.edu/edc/about/conditions_of_use.php?page=/edc/abiotic_monitoring/data_query.php
565

566 Turetsky M. R., M. C. Mack, T. N. Hollingsworth, and J. W. Harden. 2010. The role of mosses in
567 ecosystem succession and function in Alaska’s boreal forest. *Canadian Journal of Forest*
568 *Research*. 40: 1237–1264.

569 Turetsky, M. R., B. Bond-Lamberty, E. Euskirchen, J. Talbot, S. Froking, A. D. McGuire, and E. S.
570 Tuittila. 2012. The resilience and functional role of moss in boreal and arctic ecosystems.
571 *New Phytologist*, 196(1): 49-67.

- 572 Ueno, T. and H. Kanda. 2006. Photosynthetic response of the arctic semi-aquatic moss
573 *Calliergon giganteum* to water content. *Aquatic Botany*. 85: 241-243.
- 574 Van Breemen, N. 1995. How Sphagnum bogs down other plants. *Trends in Ecology & Evolution*.
575 10(7): 270-275.
- 576 Van Gaalen, K. E., L. B. Flanagan, D. R. Peddle. 2007. Photosynthesis, chlorophyll, and spectral
577 reflectance in Sphagnum moss at varying water contents. *Oecologia*. 153(1): 19-28.
- 578 Viereck, L. A. 1975. Forest ecology of the Alaska taiga. In *Proceedings, Circumpolar Conference*
579 *on Northern Ecology*, Sept 15-18, 1975. 11.
- 580 Vourlitis, G. L., W. C. Oechel, A. Hope, D. Stow, B. Boynton, J. Verfaillie, R. Zulueta, and S. J.
581 Hastings. 2000. Physiological models for scaling plot measurements of CO₂ flux across
582 an arctic tundra landscape. *Ecological Applications*. 10(1): 60-72.
- 583 Walker, D. A., H. E. Epstein, G. J. Jia, A. Balsler, C. Copass, E. J. Edwards, W. A. Gould, J.
584 Hollingsworth, J. Knudson, H. A. Maier, and A. Moody, A., 2003. Phytomass, LAI, and
585 NDVI in northern Alaska: Relationships to summer warmth, soil pH, plant functional
586 types, and extrapolation to the circumpolar Arctic. *Journal of Geophysical Research:*
587 *Atmospheres*. 108(D2).
- 588 Walker, D. A., H. E. Epstein, M. K. Reynolds, P. Kuss, M. A. Kopecky, G. V. Frost, F. J. A. Daniëls,
589 M. O. Leibman, N. G. Moskalenko, G. V. Matyshak, and O. V. Khitun. 2012. Environment,
590 vegetation and greenness (NDVI) along the North America and Eurasia Arctic transects.
591 *Environmental Research Letters*. 7(1): 015504.
- 592 Zona, D., W. C. Oechel, J. H. Richards, S. Hastings, I. Kopetz, H. Ikawa, and S. Oberbauer. 2011.
593 Light-stress avoidance mechanisms in a Sphagnum-dominated wet coastal Arctic tundra
594 ecosystem in Alaska. *Ecology*, 92(3): 633-644.

595

596 List of Figures:

597 **Figure 1:** NDVI (solid line), near-infrared (dotted line), and red (dashed line) of four
598 communities by percent moisture content (left panels) and hours after initial saturation (right
599 panels) during drying and after re-saturation during experiment 1.

600 **Figure 2:** Gross primary productivity and NDVI of four communities by percent moisture
601 content during drying and three measurement times after re-saturation during experiment 2.

602 **Figure 3:** Gross primary productivity and NDVI of four communities by hours of drying and three
603 measurement times after re-saturation during experiment 2.

604 **Figure 4:** Drying response indexes of four communities for GPP and NDVI during experiment 2
605 compared using a one-way analysis of variance. Statistical significance denoted with *.

606 List of Tables:

607 **Table 1:** Gross primary productivity and NDVI of four communities by percent moisture content
608 during drying and three measurement times after re-saturation compared using repeated
609 measures analysis of variance with Tukey's post-hoc analysis, along with the initial saturated
610 and final oven dry weight for each of the four community types during experiment 2. Letters
611 denote statistically significant differences ($p < 0.05$) between moisture content measurements.

612