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# Insights into the O : C-dependent mechanisms controlling the evaporation of $\alpha$ -pinene secondary organic aerosol particles

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Abstract. The volatility of oxidation products of volatile organic compounds (VOCs) in the atmosphere is a key factor to determine if they partition into the particle phase contributing to secondary organic aerosol (SOA) mass. Thus, linking volatility and measured particle composition will provide insights into SOA formation and its fate in the atmosphere. We produced  $\alpha$ -pinene SOA with three different oxidation levels (characterized by average oxygen-to-carbon ratio;  $\overline{O:C} = 0.53, 0.69, \text{ and } 0.96$ ) in an oxidation flow reactor. We investigated the particle volatility by isothermal evaporation in clean air as a function of relative humidity (RH < 2 %, 40 %, and 80 %) and used a filter-based thermal desorption method to gain volatility and chemical composition information.

We observed reduced particle evaporation for particles with increasing  $\overline{O:C}$  ratio, indicating that particles become more resilient to evaporation with oxidative aging. Particle evaporation was increased in the presence of water vapour and presumably particulate water; at the same time the resistance of the residual particles to thermal desorption was increased as well. For SOA with  $\overline{O:C} = 0.96$ , the unexpectedly large increase in mean thermal desorption temperature and changes in the thermogram shapes under wet conditions (80 % RH) were an indication of aqueous phase chemistry. For the lower  $\overline{O:C}$  cases, some water-induced composition changes were observed. However, the enhanced evaporation

under wet conditions could be explained by the reduction in particle viscosity from the semi-solid to liquid-like range, and the observed higher desorption temperature of the residual particles is a direct consequence of the increased removal of high-volatility and the continued presence of lowvolatility compounds.

# 1 Introduction

Secondary organic aerosol (SOA) accounts for a major fraction of the global atmospheric aerosol burden (Hallquist et al., 2009; Jimenez et al., 2009). Understanding the mechanism of formation and properties of SOA is therefore of utmost importance to estimate its effects on climate, air quality and human health. SOA - by definition - is formed when low-volatility oxidation products of volatile organic compounds (VOCs) deposit onto existing particles or form new particles. While the scientific community has made significant improvements in characterizing the precursors and gas phase oxidation products at the molecular level with new measurement techniques (e.g. Ehn et al., 2014; Lopez-Hilfiker et al., 2014; Mohr et al., 2017), the particle phase is still proving to be complicated. There are technical challenges as chemical composition analysis techniques typically operate with liquid or gaseous samples, and thus aerosol particle samples need to be extracted or desorbed which can introduce artefacts. In addition, there are a multitude of compounds in aerosol particles (Goldstein and Galbally, 2007), and the particle phase is far from static once it is formed. Many different chemical processes can occur in the particle phase, especially if water is present; e.g. larger molecules may be formed by polymerization reactions or there may be oxidation with oxidants taken up from the gas phase (Herrmann et al., 2015; Kroll and Seinfeld, 2008). In addition, aerosol particles interact with the gas phase constantly. Compounds will partition between the gas and particle phase depending, e.g., on their vapour pressures and concentrations in the different phases (Donahue et al., 2006).

Gas-particle partitioning has been described in detail and applied to model SOA formation (Pankow, 1994a, b; Pankow et al., 2001). Donahue et al. (2006) bypassed the issue of having to define the phase diagram for each compound in the SOA system by introducing the volatility basis set (VBS) framework which utilizes the effective saturation concentration ( $C^*$ ) to group the compounds into volatility classes (or bins). This framework can then be used to follow changes in partitioning, and thus bulk volatility, with time and to predict SOA formation, e.g. in global modelling (e.g. Bergström et al., 2012; Jimenez et al., 2009). More comprehensive schemes have since been developed linking volatility with other properties, e.g. oxidation level represented by the elemental oxygen-to-carbon  $\overline{O:C}$  ratio (Donahue et al., 2011, 2012) or a certain functionality (Krapf et al., 2016).

In principle, VBS distributions can be derived from any type of volatility measurements. Besides the SOA yield studies, the most common approaches are to measure the evaporation of SOA particles at elevated temperatures (typically 30-300 °C) or by removing the gas phase around the particles through isothermal dilution. Thermal methods include measurements of particle mass loss after passing through a thermodenuder (An et al., 2007; Cappa and Wilson, 2011; Huffman et al., 2008; Kolesar et al., 2015) and methods where particles are collected and then desorbed with a heated carrier gas flow (e.g. the filter inlet for gas and aerosol sampling unit, FIGAERO; Lopez-Hilfiker et al., 2014). Typically, the remaining SOA mass or volume fraction is determined or the changes in chemical composition with temperature are monitored. With any thermal method, there is the potential of thermal decomposition at elevated temperatures prior to detection, and some studies report strong indication of decomposition and particle phase reactions (Hall and Johnston, 2012; Kolesar et al., 2015; Lopez-Hilfiker et al., 2015; Stark et al., 2017). This may lead to an overestimation of the (semi-)volatile fraction of the SOA (Lopez-Hilfiker et al., 2015; Schobesberger et al., 2018). This issue may be constrained by combining isothermal dilution and thermodenuder measurements (Grieshop et al., 2009; Louvaris et al., 2017).

Vaden et al. (2011) proposed an isothermal dilution method utilizing a differential mobility analyser (DMA) for

particle size selection coupled to an evaporation-residence time chamber at room temperature with active carbon as an absorber for the evaporating vapours. This method avoids issues associated with thermal decomposition in thermodenuders and wall loss artefacts in smog chamber experiments. It enables studying evaporation times of up to 2 days. With this method, it was shown that dry particles evaporate much slower than expected from SOA yield experiments (Vaden et al., 2011). Combining this type of data with detailed process modelling suggested that mass transport limitations in the most likely semi-solid, and thus highly viscous, dry particles were responsible for the observed slow evaporation (Yli-Juuti et al., 2017). In this context, particulate water can play a crucial role as it can act as a plasticizer, reducing particle viscosity and thus enhancing evaporation (Wilson et al., 2015; Yli-Juuti et al., 2017). Aqueous phase chemical reactions may also have an impact on SOA volatility, enhancing the formation of larger molecules or, conversely, hydrolysing especially peroxy compounds. To distinguish between the physical (viscosity decrease) and chemical (aqueous phase reactions) role of particulate water in particle evaporation, it is necessary to investigate the chemical composition of SOA particles during isothermal evaporation at ambient temperatures under a range of RH conditions. Although the thermal desorption in FIGAERO-CIMS (chemical ionization mass spectrometer) may cause decomposition as described above, it can still be useful for this purpose by combining measurements of fresh SOA particles and of particles that have been left to evaporate for long times. Whereas D'Ambro et al. (2018) left the collected SOA particles on the FIGAERO filter in a clean nitrogen flow for up to 24 h, we utilized our residence time chamber (RTC) to probe the particle phase composition after different evaporation times.

In this study we investigated the volatility of SOA particles formed from  $\alpha$ -pinene at different oxidation levels with a combination of isothermal evaporation and thermal desorption. In addition, we monitored the evolution of the chemical composition of the residual particle after evaporation using FIGAERO–CIMS. Special focus was on the role of particulate water in the evaporation behaviour to determine whether water simply affected the rate transfer through the particles or also induced chemical reactions in the particles. Understanding particle phase processes is crucial to improving our ability to model SOA formation and to predict its behaviour in the atmosphere.

# 2 Methods

# 2.1 Measurements

The experiments were conducted in the same fashion as described in Yli-Juuti et al. (2017) with the main difference that here SOA was formed in a potential aerosol mass reactor (PAM; Aerodyne Research Inc., Kang et al., 2007; Lambe et al., 2011) through ozonolysis and photooxidation of  $\alpha$ pinene instead of pure ozonolysis. A schematic of the experimental set-up is shown in Fig. S1 in the Supplement.  $\alpha$ -Pinene vapour was introduced to the PAM reactor by flowing nitrogen over a vial of  $\alpha$ -pinene ( $\geq$  98 % purity, Sigma Aldrich) placed in a diffusion source resulting in a mixing ratio of 190 parts per billion by volume (ppb). The  $\alpha$ -pinene mixing ratio input to the PAM reactor was 190 ppb as monitored with a high-resolution time-of-flight proton transfer reaction mass spectrometer (PTR-MS, Ionicon model 8000). Concentrations of  $\alpha$ -pinene were checked periodically at the outlet of the PAM reactor and were < 1 ppb for all settings (i.e. the precursor reacted completely). The PAM reactor was operated at a constant temperature of 27 °C and RH of 40 %. Hydroxyl radicals (OH) were formed from ozone  $(O_3)$  photolysis and reaction of the generated  $O(^1D)$  with water vapour inside the PAM reactor. The O3 concentration and the irradiation level in the PAM reactor were varied to create SOA with a range of effective oxidation ages (see Tables 1 and S1). The polydisperse SOA was characterized with a scanning mobility particle sizer (SMPS; TSI Inc., model 3082+3775) and a high-resolution time-of-flight aerosol mass spectrometer (AMS; Aerodyne Research Inc., DeCarlo et al., 2006).

To conduct the isothermal evaporation measurements, two nano differential mobility analysers (NanoDMAs, TSI Inc., model 3085) were used in parallel to select a narrow distribution of monodisperse particles (mobility diameter 80 nm). Two NanoDMAs were utilized to double the available sample flow without compromising on the sheath-to-sampleflow ratio. Each NanoDMA was operated with an open loop sheath flow, and the residence time inside the NanoDMA was  $\leq 0.3$  s, which limits the diffusional mixing of gas phase compounds into the selected sample flow. This led to a high dilution of the gas phase compounds (by more than 1 order of magnitude) and resulted in a sudden shift from the gasparticle equilibrium at the NanoDMA outlet, which initiated the particle evaporation. Due to some length of tubing always required for sampling, the shortest particle evaporation time, i.e. the minimum time between exiting the NanoDMA and reaching subsequent measurement devices, was 2 s. Evaporation times of up to 160 s were achieved by simply adding tubing to the sample line and/or reducing the sample flow. Size and composition of the size-selected particles were monitored using an SMPS, an AMS, and a FIGAERO (Aerodyne Research Inc.; Lopez-Hilfiker et al., 2014) in combination with a CIMS (Aerodyne Research Inc.; Lee et al., 2014) using iodide as reagent ion. The overall RH during the evaporation experiment was controlled with the sheath flow of the NanoDMAs at three different levels: dry (RH < 2%), RH40 %, and RH80 %. For RH40 % and RH80 %, the sheath flow was humidified with a Nafion humidifier (Permapure, model PD-200T).

Longer evaporation times, ranging from 0.5 to 11 h, were achieved by passing the sample flow into a 100 L cylindri-

cally shaped stainless-steel chamber (RTC), which had an inlet at the top and an outlet at the bottom. Our earlier study showed that the stainless-steel walls of the evaporation chamber act as nearly perfect sinks for the compounds evaporating from the particles (Yli-Juuti et al., 2017). Before each evaporation experiment, the NanoDMAs, connected tubing, and RTC were flushed for at least 12 h with clean air at the RH of the next experiment. This ensured that all parts of the system had adjusted to the new conditions. RTC experiments were initiated by adding monodisperse SOA through the top inlet for 20 min while displacing clean air, which exited via the bottom outlet. At the end of filling the RTC, the average particle number concentration was  $200-1500 \text{ cm}^{-3}$  and the average particle mass concentration was  $0.1-0.6 \,\mu g \, m^{-3}$ . Then the chamber was closed off and periodically opened again to sample for 9 to 15 min intervals with the SMPS and AMS. Clean air with the same RH was admitted into the chamber during sampling intervals to maintain constant pressure and humidity, with corresponding dilution factors of typically 1.2 or less. As we base our analysis on changes in measured particle size, and not on the change in total particle mass, the particle losses and further dilution only limit the number of times it is possible to sample from the RTC. The size selection unit, the RTC, and all particle phase measurement instruments were located in a temperature-controlled room (21 °C) to minimize the effect of ambient temperature fluctuations on particle evaporation and on RH. For the RH80 % experiments, the closed-loop sheath flow of the SMPS was also humidified to ensure that the RH stayed within  $\pm 2\%$ between size selection, evaporation, and size measurements.

For FIGAERO-CIMS measurements, the above-described procedure had to be adjusted to accommodate the need for higher particle mass loadings: (1) "fresh" samples were collected directly after size selection for 20 or 30 min; (2) RTC fill times were increased from 20 to 75 min; and (3) the FIGAERO-CIMS sampled the remaining SOA particles in the RTC once after 3 to 4 hours of evaporation (data labelled "RTC" in the following). The upper limit of collected mass was estimated from the particle mass concentration and sampling time. Collected particulate material was 140–260 ng for fresh and 20-70 ng for RTC samples (Table S3). Due to the collection time needed, the fresh filter sample contained particles which had been on the filter for 0 to 20 or 30 min. As the particle evaporation starts when the gas phase is diluted in the NanoDMAs and no new particle-gas-phase equilibrium can be reached, the evaporation continues while particles are being deposited on the FIGAERO filter. Thus, some volatile compounds may have already evaporated before the thermal desorption begins and cannot be captured with this method.

## 2.2 Data analysis

The SMPS data were inverted with the Aerosol Instrument Manager software (TSI). To check the selected sizes of the NanoDMAs, ammonium sulfate particles, which are non-

OH exposure	Instrument			$\overline{\mathbf{O}:\mathbf{C}}$	$\overline{H:C}$	Osc	Average composition
Low	AMS			0.53	1.53	-0.46	
	FIGAERO	dry	fresh	0.66	1.62	-0.30	C <sub>9.7</sub> H <sub>16.0</sub> O <sub>5.8</sub>
			RTC	0.68	1.63	-0.28	C <sub>10.8</sub> H <sub>17.9</sub> O <sub>6.8</sub>
		RH80 %	fresh	0.68	1.62	-0.25	C <sub>10.2</sub> H <sub>16.8</sub> O <sub>6.5</sub>
			RTC	0.71	1.61	-0.20	$C_{10.6}H_{17.4}O_{6.8}$
Medium	AMS			0.69	1.42	-0.05	
	FIGAERO	dry	fresh	0.75	1.52	-0.03	C <sub>9.0</sub> H <sub>14.0</sub> O <sub>6.3</sub>
			RTC	0.74	1.53	-0.04	C <sub>9.3</sub> H <sub>14.4</sub> O <sub>6.4</sub>
		RH80 %	fresh	0.76	1.52	0.00	C9.0H14.0O6.4
			RTC	0.77	1.55	-0.02	$C_{9.8}H_{15.5}O_{6.9}$
High	AMS			0.96	1.26	0.63	
	FIGAERO	dry	fresh	0.84	1.46	0.23	C <sub>8.2</sub> H <sub>12.3</sub> O <sub>6.3</sub>
			RTC	0.83	1.47	0.19	C <sub>8.4</sub> H <sub>12.7</sub> O <sub>6.3</sub>
		RH80 %	fresh	0.85	1.43	0.27	C <sub>8.0</sub> H <sub>11.7</sub> O <sub>6.1</sub>
			RTC	0.84	1.46	0.22	$C_{8.5}H_{12.8}O_{6.2}$

Table 1. Average oxidation state and average molecular formula derived from FIGAERO-CIMS and AMS measurements.

volatile at room temperature, were sampled for each sheath flow setting. This actual measured size was then used as "set" size. The evaporation factor (EF) was calculated as the ratio of the measured sizes ( $D_{meas}$ ) and the set sizes ( $D_{set}$ ):

$$EF = \frac{D_{\text{meas}}}{D_{\text{set}}}.$$
 (1)

Assuming spherical particles, the volume fraction remaining (VFR) can be calculated as

$$VFR = EF^3.$$
 (2)

In the following, the evolution of VFR as a function of the residence time in the RTC will be called "evapogram".

The high-resolution AMS data were analysed with the SQUIRREL (version 1.59D) and PIKA toolkits (version 1.19, DeCarlo et al., 2006). The improved parameterization from Canagaratna et al. (2015) was used to perform the elemental analysis which yields average oxygen-tocarbon and hydrogen-to-carbon ratios for the sampled SOA  $(\overline{O:C} \text{ and } \overline{H:C} \text{ ratios})$ . The average carbon oxidation state,  $\overline{OSc}$ , was calculated following the approximation in Kroll et al. (2011):

$$\overline{\text{OSc}} = 2 \cdot \overline{\text{O}:\text{C}} - \overline{\text{H}:\text{C}}.$$
(3)

Raw data from FIGAERO desorption temperature scans were processed using tofTools, a MATLAB-based software package developed for analysing ToF-CIMS data (Junninen et al., 2010). The raw data were averaged to provide average mass spectra spaced by 20 s, and baseline correction was applied before fitting the high-resolution mass spectral data. Details about the magnitude and impact of the instrument background (filter blank measurements) on our analysis are discussed in the Supplement (Sect S1.2). The average  $\overline{O:C}$  and  $\overline{\text{H}:\text{C}}$  ratios and composition were calculated as signal weighted sums from the elemental composition defined by the sum formulas. Details are given in the Supplement (Sect. S1.3).

Thermal desorption of a FIGAERO filter sample via a nitrogen gas flow heated from 25 to 210 °C yields thermograms, i.e. the total or selected ion count rate vs. desorption temperature. The sum over all ions (except the reagent ions) will be referred to as "total thermogram". Where noted, we normalized the thermograms with the time integral of the respective total thermogram to help compare thermogram shapes. We characterize the thermograms mainly by the temperature of peak desorption (thermogram maximum,  $T_{max}$ ), as is common practice (Huang et al., 2018; Lopez-Hilfiker et al., 2014). We also use the median desorption temperature  $(T_{\text{median}})$ , i.e. the temperature dividing the thermogram into two equal areas. This value may reflect the overall desorption characteristics better than  $T_{\text{max}}$  because thermograms (individual or total) may feature poorly defined peaks and contain large fractions of signal at temperatures very different from (typically higher than) T<sub>max</sub>. Integrated normalized FI-GAERO mass spectra were obtained by calculating the time integral of each ion's count rate over the full desorption cycle and then normalizing to the sum over all non-reagent ions. The above-described normalization procedures were designed to account for differences between experiments in the amount of particle mass collected on the filter, and it allows us to directly compare the thermogram shapes and the relative contributions of certain ions between different experiments while not affecting  $T_{\text{max}}$  and  $T_{\text{median}}$ .

In the CIMS instrument, the major class of ions were clusters of iodide  $(I^-)$  and organic compounds (M) in the sample flow, detected as  $[M + I]^-$ . In this study, the voltage settings in the CIMS's ion guidance elements led to a relatively

high level of ion declustering, which included the formation of ions not containing iodide and with an odd number of hydrogen atoms likely dominated by  $[M+I]^-$  ions that lost HI resulting in  $[M-H]^-$  and by other fragmentation processes described in the Supplement. These "declustered ions" accounted for 15%–25% of the total ion signal (see Supplement Sect. S1.1 for further information). We analysed the data treating the declustered and adduct ions separately. However, for plotting the integrated spectra of all observed species, all ions were included, and it was assumed that deprotonation to form  $[M-H]^-$  was the only declustering reaction. The observed ion formulas were converted into neutral compound formulas by adding the mass of H<sup>+</sup> for  $[M-H]^$ or subtracting the mass of I<sup>-</sup> for  $[M+I]^-$ .

# 3 Results and discussion

### 3.1 SOA chemical composition

For the low, medium, and high OH exposure in the PAM reactor, the O: C ratios derived from AMS data of size-selected  $\alpha$ -pinene SOA were 0.53, 0.69, and 0.96, respectively. These O: C values are representative of fresh and aged ambient SOA in monoterpene-rich environments (Aiken et al., 2008; Ng et al., 2010; Ortega et al., 2016). From here on we refer to these three cases as low-, medium-, or high- $\overline{O:C}$  experiments. Overall, FIGAERO and AMS measurements show the same trends in the  $\overline{O:C}$  ratios; however, the AMS derived values show a larger difference in  $\overline{O:C}$  ratios between the low and high OH exposures (see Table 1). With increasing overall  $\overline{O:C}$ , the FIGAERO–CIMS mass spectra show an increasing fraction of monomers (defined as compounds with 10 or fewer C atoms, i.e. compounds derivable from a single monoterpene precursor, roughly corresponding to masses < 300 Da) as shown in Fig. S2. In the high-O : C case, there is a strong increase in the contribution of smaller molecules with high O:C (see Figs. S2 and S5c), due to the dominance of fragmentation reactions at high OH exposure (Lambe et al., 2012; Palm et al., 2016). It was not expected to find such a large contribution of low molecular weight (MW) compounds such as C<sub>3</sub>H<sub>4</sub>O<sub>4</sub> at 104 Da (detected mostly as  $[C_3H_4O_4 + I]^-)$  or  $C_4H_6O_4$  at 118 Da (detected mostly as  $C_4H_5O_4^-$ ) in the particle phase. Given that the majority of compounds of this size should be too volatile to stay in the particle phase, a likely cause for the appearance of these low-MW compounds is thermal decomposition of higher-MW compounds during the desorption from the FI-GAERO filter. Then the increase in low-MW compounds at higher O: C indicates that particulate organics become overall more susceptible to thermal decomposition when SOA is formed under higher OH exposure in the PAM reactor. However, at this point we cannot determine if the increased detection of these low-MW compounds is driven by a higher degree of fragmentation reactions at high OH exposure in PAM or the thermal decomposition of higher-MW compounds in the FIGAERO.

# **3.2** Linking isothermal evaporation and thermal desorption in FIGAERO

Plots showing the VFR of  $\alpha$ -pinene SOA as a function of residence time (i.e. evapograms) are presented in Fig. 1a-c for the three O : C cases. The dependence of the evaporation rate on RH follows the trends reported in earlier studies: at dry conditions, the evaporation is substantially slower than at RH40% and RH80% conditions, for all oxidation levels (Vaden et al., 2011; Wilson et al., 2015; Yli-Juuti et al., 2017). In the studies of Wilson et al. (2015) and Yli-Juuti et al. (2017), the slower evaporation under dry conditions was related to increased diffusional limitations, due to higher viscosity than under humid conditions. We note that the evaporation rate at RH40 % is higher than at RH80 % (Fig. 1a-c). This observation is explained by the solution or Raoult effect, i.e. the decrease in the equilibrium vapour pressure over more dilute humidified particles, as demonstrated by the evaporation modelling presented in the Supplement (Sect. S1.4 and Fig. S4c) and in the study of Yli-Juuti et al. (2017). This indicates that diffusion limitations do not play a major role in aerosol evaporation at room temperature when the RH is at atmospherically relevant levels.

The dependence of the isothermal evaporation rate on the oxidation level is reported for the first time in this study. As the  $\overline{O:C}$  ratio of the produced SOA increases, the overall rate of evaporation decreases. After 6 h of evaporation, SOA particle volume decreased by only 10% under dry and  $\sim 40\%$  under RH40% conditions in the high- $\overline{O:C}$  case. For the low- $\overline{O:C}$  case, the corresponding numbers are 40% (dry) and 60% (RH40%). These trends suggest that the more highly oxygenated SOA is less volatile, as expected from thermodenuder measurements (e.g. Donahue et al., 2012).

In Fig. 1d-f, we show FIGAERO total thermograms (signal-weighted sum of the thermograms for all individual compositions) measured at different time periods during isothermal evaporation at dry and RH80% conditions. The FIGAERO filter sampling periods of each thermogram are marked by coloured boxes in the evapograms in Fig. 1ac. Fresh SOA thermograms were shifted to higher temperatures with increasing O: C, both at dry and RH80 % conditions. For the low-, medium-, and high- $\overline{O:C}$  cases, the peak evaporation temperatures, T<sub>max</sub>, were 50, 60, and 71 °C under dry conditions and 61, 70, and 92 °C at RH80 %. These shifts are in line with our isothermal evaporation measurements suggesting a decreasing vapour pressure of SOA compounds with increasing  $\overline{O:C}$  and are also consistent with a larger role of thermal decomposition during desorption, as indicated by the increased contribution of small highly oxidized molecules discussed above.

When examining the  $T_{\text{max}}$  of fresh SOA in more detail, it can be seen that at a fixed  $\overline{O:C}$  ratio,

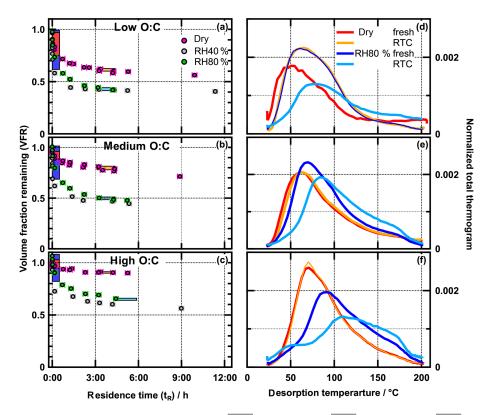


Figure 1. Evapograms (a–c) and total thermograms (d–f) for low  $\overline{O:C}$  (a, d), medium  $\overline{O:C}$  (b, e), and high  $\overline{O:C}$  (c, f). Coloured boxes in evapograms indicate FIGAERO sampling time. Thermograms are normalized with total signal area.

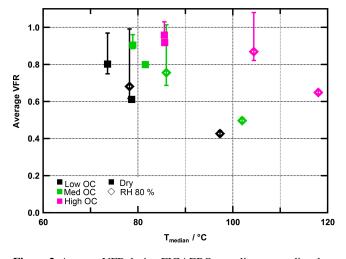
 $T_{\text{max}}$  (RH80%) >  $T_{\text{max}}$  (dry). This trend can be explained with the evapogram measurements: the particles evaporate more quickly at higher humidity as seen in the evapograms; hence a larger fraction of higher-volatility compounds is already lost during the 20 or 30 min period of FIGAERO filter collection before the thermal desorption. Thus, the collected residual particles are less volatile, characterized by a higher  $T_{\text{max}}$ . When sampling from the RTC after longer evaporation time, the VFR is even lower (i.e. a larger volume fraction has been lost due to evaporation). Correspondingly, the thermogram peak is shifted further toward higher temperatures in all studied  $\overline{O:C}$  cases, again indicating an increasing fraction of lower volatility compounds in the residual particles.

Only for the high-O: C case is an absolute increase in the amount of material desorbing at > 150 °C also observed when comparing the fresh SOA at RH80 % and dry conditions (see non-normalized thermograms in Fig. S3). Because the estimated organic mass loadings on the filter were comparable, this indicates that when the high- $\overline{O:C}$  particles, generated in the PAM at RH40 %, are exposed to elevated RH (RH80 %), compounds with high desorption (and/or decomposition) temperatures are formed in the particle phase. We will corroborate this suggestion below.

In Fig. 2, VFR is plotted as a function of  $T_{\text{median}}$ . The figure visualizes two phenomena: generally,  $T_{\text{median}}$  and VFR are positively correlated with the  $\overline{O:C}$  ratio. As laid out

above, this observation is explained by the negative relationship of  $\overline{O:C}$  and volatility. At the same time, however, for a certain  $\overline{O:C}$  ratio, VFR and  $T_{\text{median}}$  are negatively correlated. As mentioned above, this can be explained by the properties of the residual particles after a certain period of evaporation. We will explore this further now, together with the possibility of water-induced particle phase reactions.

In low- and medium- $\overline{O:C}$  cases, the trends of the VFR vs.  $T_{\text{median}}$  behaviour are comparable, and the increase in  $T_{\rm median}$  is clearly associated with the decreasing VFR, regardless of the RH and hence water content of the particles. The behaviour in the high- $\overline{O:C}$  case is different and cannot be explained by the evaporation of higher-volatility material alone. For the high-O : C cases,  $T_{\text{median}}$  of fresh SOA increases from 86 to 104 °C between the dry and RH80 % case despite a change of only 9% in VFR. Combining this observation with the fact that there is an absolute increase in material desorbed >  $150 \,^{\circ}\text{C}$  suggests that in the high- $\overline{\text{O}:\text{C}}$ cases the particle phase water alters the SOA particle composition, resulting in an increased resistance to thermal desorption or decomposition (i.e. large change in  $T_{\text{median}}$ ) even if the particles lost only a small volume fraction due to isothermal evaporation (i.e. small change in VFR). We note that these composition changes are not clearly visible in the average O: C or OSc values (see Table 1), and we will elaborate on possible reactions in Sect. 3.4.

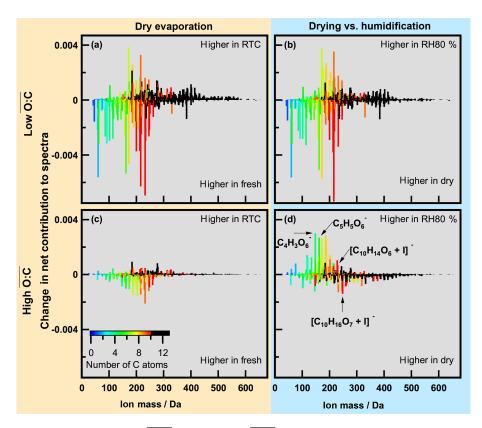


**Figure 2.** Average VFR during FIGAERO sampling vs. median desorption temperature ( $T_{median}$ ) for all experiments. Colours indicate  $\overline{O:C}$  ratios. Measurements under dry conditions are marked with squares, those under RH80% with diamonds. Error bars indicate minimum and maximum VFR observed during sampling time.

# 3.3 Residual particle composition during evaporation under dry and humid conditions

During the evaporation the initial  $\overline{O:C}$  changed very little (Table 1). This is consistent with earlier observations reported by Yli-Juuti et al. (2017), who interpreted this as evidence for the presence of low-volatility oligomers in the particles. These should have very similar O:C ratios to the corresponding monomers. To examine detailed changes in particle composition along the isothermal evaporation at dry conditions, we show the difference between the normalized integrated FIGAERO-CIMS mass spectra measured at the beginning (fresh) and after 3 to 4h of isothermal evaporation (RTC) in Fig. 3 (panels a and c for the low- and high-O: C case, respectively). To investigate the changes in composition due to the humidification, panels (b) and (d) in Fig. 3 show the differences between FIGAERO-CIMS mass spectra measured at dry and RH80% conditions in the beginning (fresh SOA) of the isothermal evaporation. As the low-O: C SOA particles evaporate (Fig. 3a), a clear decrease in the fractional contribution of low-MW compounds (< 300 Da,  $\sim$  monomers) is observed, whereas the contribution of compounds with MW > 300 Da (predominantly dimers) increases. Correspondingly, the contribution of compounds with C > 10 increases with evaporation, while that of C < 7 decreases. The relative contributions of intermediate masses are more likely to increase during evaporation if they contain more oxygen atoms (Fig. S5a). It is not possible to tell if these  $C_7-C_9$  compounds really remain in the particles or if they are simply thermal decomposition products of the more abundant dimers. However, this suggests that lighter and/or less oxidized molecules are indeed lost more readily during isothermal evaporation in the RTC than the heavier dimers and more oxidized compounds, which are expected to have very low vapour pressure (Mohr et al., 2017). The more detailed investigation of changes in the mass spectra (Fig. S5a and b, c and d) shows some indications of particle phase water-driven chemical transformation both for low and medium  $\overline{O:C}$ , but the differences are not as clear as in the high-O: C case (Fig. S5e and f). It should be noted that in the low- $\overline{O:C}$  case the molecules affected by particle phase water account for approximately 10% of the total signal. Therefore overall, the enhanced evaporation during FIGAERO filter collection under wet conditions is very similar to the evaporation happening under dry conditions in the RTC, and the water-driven chemistry plays only a minor role in low- $\overline{\mathbf{O}:\mathbf{C}}$  cases. This points to particulate water mainly reducing the viscosity and thus accelerating the mass transport in the particles as described in Yli-Juuti et al. (2017). In the high-O: C case (Fig. 3c), there is also a relative decrease in masses < 300 Da with isothermal evaporation under dry conditions, but the overall picture is less clear, consistent with very few changes in VFR and in the sum thermogram shape in this case (high-O: C dry; Fig. 1c and d). Conversely, humidifying fresh high-O: C SOA particles leads to an increase in masses < 200 Da (Fig. 3d), which is a very different behaviour compared to the low-O : C cases (Fig. 3a and b) or to the isothermal evaporation of high-O: C SOA particles at dry conditions (Fig. 3c). Again, this suggests changes in particle composition upon humidification in the high-O: C case. The mass fraction of compounds showing water-driven chemical transformation makes up approximately 30 % of the signal in high- $\overline{O:C}$  cases. This should be taken into account when process level modelling of systems comparable to the high- $\overline{O:C}$  case is considered.

To gain a better understanding of these compositional changes related to humidification, we examined the individual desorption thermograms of a few major ions (Fig. 4):  $C_4H_3O_6^-$  (a),  $C_5H_5O_6^-$  (b), and  $[C_{10}H_{14}O_6 + I]^-$  (c), which show an increase when the particles are humidified, and  $[C_{10}H_{16}O_7 + I]^-$  (d), which exhibits a net decrease in the high- $\overline{O:C}$  RH80 % case. In the low- $\overline{O:C}$  case, only small shifts (0–6 °C) in  $T_{\text{max}}$  are observed for all four ions when RH is increased. This and the changes in the thermogram shape are consistent with the behaviour observed for the total thermograms, described and explained above. In the high-O:C case, only a small shift in  $T_{max}$  is visible for  $[C_{10}H_{16}O_7 + I]^-$  as well (Fig. 4d), but for the other ions (Fig. 4a-c), we see a clear shift in the thermograms, with  $T_{\text{max}}$  increasing for the humidified case from 63 to 84, 67 to 84, and 70.5 to 95.5 °C for  $C_4H_3O_5^-$ ,  $C_5H_5O_6^-$ , and  $[C_{10}H_{14}O_6 + I]^-$ , respectively. This behaviour is unique for the high- $\overline{O:C}$  case. The collected organic mass loading on the FIGAERO filter was comparable (within 20%) for dry and RH80 % conditions. Thus, the apparent shift in  $T_{\text{max}}$  for  $[C_{10}H_{16}O_7+I]^-$  (Fig. 4d) could be explained by volatile material (with lower  $T_{\text{max}}$ ) leaving the particles during evaporation in the same way as in the low-O: C case. But for the



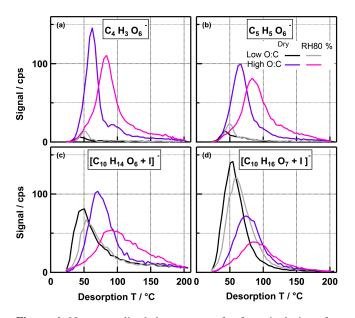
**Figure 3.** Changes in normalized spectra for low  $\overline{O:C}$  (**a**, **b**) and high  $\overline{O:C}$  (**c**, **d**). (**a**, **c**) Changes due to evaporation under dry conditions; (**b**, **d**) changes between dry and wet conditions. Colour indicates number of C atoms in the identified ions (black corresponds to C number larger than 10). Mass spectra were normalized by total signal, and then the difference was calculated.

ions in Fig. 4a–c, new material with higher  $T_{\text{max}}$  has clearly been formed. This must be triggered by the presence of larger amounts of water under RH80% conditions, via either of two scenarios: (a) there are two different isomers dominating the thermogram for the respective composition at dry vs. RH80 % conditions and the desorption temperatures of these isomers differ considerably; (b) the individual thermograms are dominated by the evaporation of monomers at dry conditions (or the decomposition of relatively unstable larger compounds into the observed compositions) but by thermal decomposition of (more stable) larger compounds at RH80 % conditions. Isomers with higher  $T_{\text{max}}$  may be formed and at the same time the isomers with lower  $T_{\text{max}}$  are lost, either through reactions or through more rapid evaporation or decomposition than in the dry case. Alternatively, other lowvolatility material is formed that thermally decomposes into the observed compounds. With this data set, we cannot exclude either of these two explanations, but the very broad shape of the RH80 % thermogram for  $[C_{10}H_{14}O_6 + I]^-$  at high temperatures (see Fig. 4c) is an indication that at least for this ion the mechanism including thermal decomposition is more likely.

#### 3.4 Possible aqueous phase processes

The FIGAERO–CIMS combined composition and thermogram measurements provided insights into the chemical composition of the residual SOA particles after humidification and/or evaporation. The observed changes in the high- $\overline{O:C}$  RH80 % case can only be explained by the formation of low-volatility compounds in the particle phase and removal of the corresponding higher-volatility compound at the same elemental composition. Thus, we briefly consider possible processes that can explain the formation of these less volatile and/or thermally more stable compounds in highly oxidized SOA at 80 %RH in the sections below.

Liquid water can have several effects on particle chemical composition. First, water may initiate hydration and hydrolysis reactions. Second, water may catalyse reactions between organics (e.g. Dong et al., 2018; Kaur and Vikas, 2017). Third, water reduces SOA viscosity (Hosny et al., 2016; Renbaum-Wolff et al., 2013), thereby reducing diffusional limitations to particle phase reactions. Whereas hydrolysis generally reduces the average molecular weight of the reactants, other processes in principle enable the formation of higher-MW but thermally labile products. Water may also enhance the uptake of  $O_3$  from the gas phase (Berke-



**Figure 4.** Non-normalized thermograms for four single ions for low- and high- $\overline{O:C}$  cases. Ions in panels (a), (b), and (c) showed a net increase in the RH80 % case, while the ion in panel (d) had a net decrease in the high- $\overline{O:C}$  ratio case. Note that the amount of SOA mass collected on the FIGAERO filter was 5%–20% higher in the RH80% cases and is also different between the  $\overline{O:C}$  cases.

meier et al., 2016; Gallimore et al., 2011), but as the average  $\overline{O:C}$  ratio did not change with increased RH, any oxidizing reaction can be excluded.

# 3.4.1 Homolysis of peroxy bonds (O-O) in organic peroxides (ROOH)

Organic hydro peroxides (ROOH) species are known to occur in SOA (Sanchez and Myers, 2000) and have been observed to decompose in the particle phase on timescales < 1 h under dark and wet ( $\sim 40\%$  RH) conditions (Krapf et al., 2016). Liquid-water-induced homolysis of the O-O bond in ROOH yields alkoxy (RO•) radicals that may initiate oligomerization reactions (Tong et al., 2016). Under dry conditions organic (hydro-)peroxides (ROOH) may be more stable and persist in the particle phase but decompose upon desorption and be detected as RO or R fragments in FIGAERO-CIMS. The oligomers (RO-OR) formed in the RH80 % case will most likely have a lower volatility than the educts, but they may still be prone to decompose upon desorption due to their relatively weak O-O bond. Thus, they would also be detected as RO or R fragments but at a higher apparent desorption T than the monomers.

# 3.4.2 Addition and accretion reactions

As generally no significant change in the O : C ratio was observed, only non-oxidative oligomerization ("accretion") reactions such as aldol addition (aldehyde + aldehyde/ketone, (hemi)acetal formation (aldehyde/ketone+alcohol), peroxy(hemi)acetal formation (aldehyde/ketone + peroxide), and esterification (alcohol + (per)acid) are likely (Herrmann et al., 2015; Kroll and Seinfeld, 2008). All these reactions have reaction pathways which are enhanced by the presence of an acid catalyst in the aqueous phase. Thus, the complete lack of water in dry particles may sufficiently prevent these reactions preserving the monomer educts. To explain our results, there would have to be an additional factor in the high-O: C case, e.g. a much higher fraction of organic peroxides and peroxy acids that form peroxy hemiacetals instead of the more stable hemiacetals. These peroxy hemiacetals would then most likely decompose at a higher temperature than the desorption temperature of the educts but be detected as low-MW compounds, while hemiacetals may be stable enough to be detected without decomposition.

#### 4 Summary and conclusions

We present the first study linking the oxidative age of  $\alpha$ pinene SOA, quantified by the  $\overline{O:C}$  ratio, and isothermal evaporation, for a wide range of RH (< 2% to 80% RH). By utilizing an RTC at room temperature and FIGAERO– CIMS as a thermal desorption technique, we were able to determine SOA volatility independent of artefacts due to thermal decomposition. At the same time, the thermal desorption data gave insights into the possible particle phase chemistry during evaporation especially under wet conditions. It has to be kept in mind though that the particles measured with FIGAERO–CIMS are always the residual particles after minutes to hours of isothermal evaporation either during filter collection or in the RTC.

We found a strong correlation between increased oxidation level of the initial particles and lower particle volatility expressed by less isothermal evaporation and higher  $T_{\text{max}}$  values in FIGAERO–CIMS thermograms. This suggests that atmospheric particles become more resistant to evaporation as they age over time, possibly increasing their lifetime in the atmosphere. This also means that the oxidation level needs to be kept in mind when investigating aerosol volatility in chamber or flow tube experiments. For example, for deriving VBS distributions from smog chamber yield experiments, care has to be taken that the oxidation level stays in the same range for all SOA mass loadings.

Increasing RH enhances particle evaporation as described in previous studies (Wilson et al., 2015; Yli-Juuti et al., 2017) while the  $T_{\text{max}}$  values of the residual particles was also increased. The observed changes under wet conditions in the low- and medium- $\overline{O:C}$  cases could be explained with the lowering of the particle phase viscosity alone, but there were some indications for water-induced changes in the chemical composition. However, the compounds exhibiting these changes accounted for approximately 10% of the detected mass, and thus these changes are minor compared to the shift in composition due to evaporation. In the high- $\overline{O:C}$  case, strong evidence for aqueous phase reactions were found with approx. 30% of the mass being affected. Further evidence for different processes happening in the high- $\overline{O:C}$  case was found in the relationship between the isothermal evaporation (as VFR) and thermal desorption (as T<sub>median</sub>). It was similar for the low- and medium- $\overline{O:C}$  cases and independent of the RH. For the high- $\overline{O:C}$  case, VFR changed very little during the isothermal evaporation at RH80 %, while a large increase in  $T_{\text{median}}$  was observed. We attribute this different behaviour to the overall different chemical composition and most likely much higher concentration of organic peroxides (ROOH) in the high- $\overline{O:C}$  case. We hypothesize that waterinduced (1) homo- and heterolytic breaking of weaker O-O bonds present in ROOH and/or (2) the formation of peroxy hemiacetals may form thermally labile oligomers with enhanced yields in high- $\overline{O:C}$  SOA at RH80 %. To further verify this explanation, direct measurements of the organic peroxide concentration for the different O:C cases would be needed but were not part of this study.

Our data suggest that the degree of thermal decomposition in FIGAERO–CIMS and its impact on derived volatility most likely depends on the initial composition of the SOA and may be changed by the presence of particulate water. Another recent study has shown that chemical composition changes induced by the presence of acidic inorganic seeds may also produce low-volatility, but thermally instable compounds, which can only be detected as their decomposition products with FIGAERO–CIMS (Riva et al., 2019). This highlights the benefit of isothermal methods for studying SOA particle volatility.

The SOA particles studied here had  $\overline{O:C}$  ratios comparable with atmospheric SOA (typically in the 0.3–1.0 range, e.g. Aiken et al., 2008; Ng et al., 2010; Ortega et al., 2016), but the high-O: C case in particular was probably not fully representative of atmospheric SOA as extremely high O<sub>3</sub> and OH radical exposure levels were applied in the PAM reactor. This may have led to a larger degree of fragmentation of the precursor molecules than expected in the atmosphere, i.e. the particle phase is dominated by  $C_{4-6}$  compounds instead of the expected  $C_{8-10}$  compounds and their oligomers (Lambe et al., 2012). Some of the suggested aqueous phase reactions may be more likely for short carbon chain compounds (e.g. glyoxal like chemistry). Also, we can only speculate that there was a larger fraction of peroxy compounds in the particles in the high- $\overline{O:C}$  case as we had no direct measurement. However, recent studies with ambient SOA have shown that these particles can contain large amounts of environmental persistent radicals and are prone to form C- and O-centred organic radicals when wetted, which can start oligomerization reactions in the particle phase (Arangio et al., 2016). So, although we formed the high- $\overline{O:C}$  SOA in a PAM reactor under extreme conditions, the particles produced and their behaviour allowed us to study processes which are most likely atmospherically relevant. Hence, our study substantially increases the understanding of complicated and inadequately studied particle phase processes and the results highlights the importance of water-driven chemistry in SOA.

*Data availability.* The data set is available upon request from Angela Buchholz (angela.buchholz@uef.fi) or Annele Virtanen (annele.virtanen@uef.fi).

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*Competing interests.* The authors declare that they have no conflict of interest.

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### References

- Aiken, A. C., DeCarlo, P. F., Kroll, J. H., Worsnop, D. R., Huffman, J. A., Docherty, K. S., Ulbrich, I. M., Mohr, C., Kimmel, J. R., Sueper, D., Sun, Y., Zhang, Q., Trimborn, A., Northway, M., Ziemann, P. J., Canagaratna, M. R., Onasch, T. B., Alfarra, M. R., Prevot, A. S. H., Dommen, J., Duplissy, J., Metzger, A., Baltensperger, U., and Jimenez, J. L.: O/C and OM/OC Ratios of Primary, Secondary, and Ambient Organic Aerosols with High-Resolution Time-of-Flight Aerosol Mass Spectrometry, Environ. Sci. Technol., 42, 4478–4485, https://doi.org/10.1021/es703009q, 2008.
- An, W. J., Pathak, R. K., Lee, B.-H., and Pandis, S. N.: Aerosol volatility measurement using an improved thermodenuder: Ap-

plication to secondary organic aerosol, J. Aerosol Sci., 38, 305–314, https://doi.org/10.1016/j.jaerosci.2006.12.002, 2007.

- Arangio, A. M., Tong, H., Socorro, J., Pöschl, U., and Shiraiwa, M.: Quantification of environmentally persistent free radicals and reactive oxygen species in atmospheric aerosol particles, Atmos. Chem. Phys., 16, 13105–13119, https://doi.org/10.5194/acp-16-13105-2016, 2016.
- Bergström, R., Denier van der Gon, H. A. C., Prévôt, A. S. H., Yttri, K. E., and Simpson, D.: Modelling of organic aerosols over Europe (2002–2007) using a volatility basis set (VBS) framework: application of different assumptions regarding the formation of secondary organic aerosol, Atmos. Chem. Phys., 12, 8499–8527, https://doi.org/10.5194/acp-12-8499-2012, 2012.
- Berkemeier, T., Steimer, S. S., Krieger, U. K., Peter, T., Pöschl, U., Ammann, M., and Shiraiwa, M.: Ozone uptake on glassy, semi-solid and liquid organic matter and the role of reactive oxygen intermediates in atmospheric aerosol chemistry, Phys. Chem. Chem. Phys., 18, 12662–12674, https://doi.org/10.1039/C6CP00634E, 2016.
- Canagaratna, M. R., Jimenez, J. L., Kroll, J. H., Chen, Q., Kessler, S. H., Massoli, P., Hildebrandt Ruiz, L., Fortner, E., Williams, L. R., Wilson, K. R., Surratt, J. D., Donahue, N. M., Jayne, J. T., and Worsnop, D. R.: Elemental ratio measurements of organic compounds using aerosol mass spectrometry: characterization, improved calibration, and implications, Atmos. Chem. Phys., 15, 253–272, https://doi.org/10.5194/acp-15-253-2015, 2015.
- Cappa, C. D. and Wilson, K. R.: Evolution of organic aerosol mass spectra upon heating: implications for OA phase and partitioning behavior, Atmos. Chem. Phys., 11, 1895–1911, https://doi.org/10.5194/acp-11-1895-2011, 2011.
- D'Ambro, E. L., Schobesberger, S., Zaveri, R. A., Shilling, J. E., Lee, B. H., Lopez-Hilfiker, F. D., Mohr, C., and Thornton, J. A.: Isothermal evaporation of α-pinene ozonolysis SOA: volatility, phase state, and oligomeric composition, ACS Earth Sp. Chem., 2, 1058–1067, https://doi.org/10.1021/acsearthspacechem.8b00084, 2018.
- DeCarlo, P. F., Kimmel, J. R., Trimborn, A., Northway, M. J., Jayne, J. T., Aiken, A. C., Gonin, M., Fuhrer, K., Horvath, T., Docherty, K. S., Worsnop, D. R., and Jimenez, J. L.: Field-Deployable, High-Resolution, Time-of-Flight Aerosol Mass Spectrometer, Anal. Chem., 78, 8281–8289, https://doi.org/10.1021/ac061249n, 2006.
- Donahue, N. M., Robinson, A. L., Stanier, C. O., and Pandis, S. N.: Coupled partitioning, dilution, and chemical aging of semivolatile organics, Environ. Sci. Technol., 40, 2635–2643, https://doi.org/10.1021/es052297c, 2006.
- Donahue, N. M., Epstein, S. A., Pandis, S. N., and Robinson, A. L.: A two-dimensional volatility basis set: 1. organic-aerosol mixing thermodynamics, Atmos. Chem. Phys., 11, 3303–3318, https://doi.org/10.5194/acp-11-3303-2011, 2011.
- Donahue, N. M., Kroll, J. H., Pandis, S. N., and Robinson, A. L.: A two-dimensional volatility basis set – Part 2: Diagnostics of organic-aerosol evolution, Atmos. Chem. Phys., 12, 615–634, https://doi.org/10.5194/acp-12-615-2012, 2012.
- Dong, Z.-G., Xu, F., and Long, B.: The energetics and kinetics of the CH<sub>3</sub>CHO + (CH<sub>3</sub>)<sub>2</sub>NH/CH<sub>3</sub>NH<sub>2</sub> reactions catalyzed by a single water molecule in the atmosphere, Comput. Theor. Chem., 1140, 7–13, https://doi.org/10.1016/J.COMPTC.2018.07.013, 2018.

- Ehn, M., Thornton, J. A., Kleist, E., Sipilä, M., Junninen, H., Pullinen, I., Springer, M., Rubach, F., Tillmann, R., Lee, B., Lopez-Hilfiker, F., Andres, S., Acir, I. H., Rissanen, M., Jokinen, T., Schobesberger, S., Kangasluoma, J., Kontkanen, J., Nieminen, T., Kurtén, T., Nielsen, L. B., Jørgensen, S., Kjaergaard, H. G., Canagaratna, M., Maso, M. D., Berndt, T., Petäjä, T., Wahner, A., Kerminen, V. M., Kulmala, M., Worsnop, D. R., Wildt, J., and Mentel, T. F.: A large source of lowvolatility secondary organic aerosol, Nature, 506, 476–479, https://doi.org/10.1038/nature13032, 2014.
- Gallimore, P. J., Achakulwisut, P., Pope, F. D., Davies, J. F., Spring, D. R., and Kalberer, M.: Importance of relative humidity in the oxidative ageing of organic aerosols: case study of the ozonolysis of maleic acid aerosol, Atmos. Chem. Phys., 11, 12181–12195, https://doi.org/10.5194/acp-11-12181-2011, 2011.
- Goldstein, A. H. and Galbally, I. E.: Known and Unexplored Organic Constituents in the Earth's Atmosphere, Environ. Sci. Technol., 41, 1514–1521, https://doi.org/10.1021/es072476p, 2007.
- Grieshop, A. P., Miracolo, M. A., Donahue, N. M., and Robinson, A. L.: Constraining the volatility distribution and gas-particle partitioning of combustion aerosols using isothermal dilution and thermodenuder measurements, Environ. Sci. Technol., 43, 4750– 4756, https://doi.org/10.1021/es8032378, 2009.
- Hall IV, W. A. and Johnston, M. V.: The thermalstability of oligomers in alpha-pinene secondary organic aerosol, Aerosol Sci. Tech., 46, 983–989, https://doi.org/10.1080/02786826.2012.685114, 2012.
- Hallquist, M., Wenger, J. C., Baltensperger, U., Rudich, Y., Simpson, D., Claeys, M., Dommen, J., Donahue, N. M., George, C., Goldstein, A. H., Hamilton, J. F., Herrmann, H., Hoffmann, T., Iinuma, Y., Jang, M., Jenkin, M. E., Jimenez, J. L., Kiendler-Scharr, A., Maenhaut, W., McFiggans, G., Mentel, Th. F., Monod, A., Prévôt, A. S. H., Seinfeld, J. H., Surratt, J. D., Szmigielski, R., and Wildt, J.: The formation, properties and impact of secondary organic aerosol: current and emerging issues, Atmos. Chem. Phys., 9, 5155–5236, https://doi.org/10.5194/acp-9-5155-2009, 2009.
- Herrmann, H., Schaefer, T., Tilgner, A., Styler, S. A., Weller, C., Teich, M., and Otto, T.: Tropospheric Aqueous-Phase Chemistry: Kinetics, Mechanisms, and Its Coupling to a Changing Gas Phase, Chem. Rev., 115, 4259–4334, https://doi.org/10.1021/cr500447k, 2015.
- Hosny, N. A., Fitzgerald, C., Vyšniauskas, A., Athanasiadis, A., Berkemeier, T., Uygur, N., Pöschl, U., Shiraiwa, M., Kalberer, M., Pope, F. D., and Kuimova, M. K.: Direct imaging of changes in aerosol particle viscosity upon hydration and chemical aging, Chem. Sci., 7, 1357–1367, https://doi.org/10.1039/C5SC02959G, 2016.
- Huang, W., Saathoff, H., Pajunoja, A., Shen, X., Naumann, K.-H., Wagner, R., Virtanen, A., Leisner, T., and Mohr, C.: α-Pinene secondary organic aerosol at low temperature: chemical composition and implications for particle viscosity, Atmos. Chem. Phys., 18, 2883–2898, https://doi.org/10.5194/acp-18-2883-2018, 2018.
- Huffman, J. A., Ziemann, P. J., Jayne, J. T., Worsnop, D. R., and Jimenez, J. L.: Development and Characterization of a Fast-Stepping/Scanning Thermodenuder for Chemically-Resolved

Aerosol Volatility Measurements, Aerosol Sci. Tech., 42, 395–407, https://doi.org/10.1080/02786820802104981, 2008.

- Jimenez, J. L., Canagaratna, M. R., Donahue, N. M., Prévôt, A. S. H., Zhang, Q., Kroll, J. H., DeCarlo, P. F., Allan, J. D., Coe, H., Ng, N. L., Aiken, A. C., Docherty, K. D., Ulbrich, I. M., Grieshop, A. P., Robinson, A. L., Duplissy, J., Smith, J. D., Wilson, K. R., Lanz, V. A., Hueglin, C., Sun, Y. L., Laaksonen, A., Raatikainen, T., Rautiainen, J., Vaattovaara, P., Ehn, M., Kulmala, M., Tomlinson, J. M., Collins, D. R., Cubison, M. J., Dunlea, E. J., Huffman, J. A., Onasch, T. B., Alfarra, M. R., Williams, P. I., Bower, K., Kondo, Y., Schneider, J., Drewnick, F., Borrmann, S., Weimer, S., Demerjian, K., Salcedo, D., Cottrell, L., Griffin, R., Takami, A., Miyoshi, T., Hatakeyama, S., Shimono, A., Sun, J. Y., Zhang, Y. M., Dzepina, K., Kimmel, J. R., Sueper, D., Jayne, J. T., Herndon, S. C., Trimborn, A. M., Williams, L. R., Wood, E. C., Kolb, C. E., Baltensperger, U., and Worsnop, D. R.: Evolution of Organic Aerosols in the Atmosphere: A New Framework Connecting Measurements to Models, Science, 326, 1525-1529, https://doi.org/10.1126/science.1180353, 2009.
- Junninen, H., Ehn, M., Petäjä, T., Luosujärvi, L., Kotiaho, T., Kostiainen, R., Rohner, U., Gonin, M., Fuhrer, K., Kulmala, M., and Worsnop, D. R.: A high-resolution mass spectrometer to measure atmospheric ion composition, Atmos. Meas. Tech., 3, 1039– 1053, https://doi.org/10.5194/amt-3-1039-2010, 2010.
- Kang, E., Root, M. J., Toohey, D. W., and Brune, W. H.: Introducing the concept of Potential Aerosol Mass (PAM), Atmos. Chem. Phys., 7, 5727–5744, https://doi.org/10.5194/acp-7-5727-2007, 2007.
- Kaur, R. and Vikas: A case of a single water molecule accelerating the atmospheric reactions of hydroxyl radical at temperatures near 200 K, Chem. Phys. Lett., 685, 270–274, https://doi.org/10.1016/j.cplett.2017.07.080, 2017.
- Kolesar, K. R., Li, Z., Wilson, K. R., and Cappa, C. D.: Heating-Induced Evaporation of Nine Different Secondary Organic Aerosol Types, Environ. Sci. Technol., 49, 12242–12252, https://doi.org/10.1021/acs.est.5b03038, 2015.
- Krapf, M., El Haddad, I., Bruns, E. A., Molteni, U., Daellenbach, K. R., Prévôt, A. S. H., Baltensperger, U., and Dommen, J.: Labile Peroxides in Secondary Organic Aerosol, Chem, 1, 603–616, https://doi.org/10.1016/j.chempr.2016.09.007, 2016.
- Kroll, J. H. and Seinfeld, J. H.: Chemistry of secondary organic aerosol: Formation and evolution of low-volatility organics in the atmosphere, Atmos. Environ., 42, 3593–3624, https://doi.org/10.1016/j.atmosenv.2008.01.003, 2008.
- Kroll, J. H., Donahue, N. M., Jimenez, J. L., Kessler, S. H., Canagaratna, M. R., Wilson, K. R., Altieri, K. E., Mazzoleni, L. R., Wozniak, A. S., Bluhm, H., Mysak, E. R., Smith, J. D., Kolb, C. E., and Worsnop, D. R.: Carbon oxidation state as a metric for describing the chemistry of atmospheric organic aerosol, Nat. Chem., 3, 133–139, https://doi.org/10.1038/nchem.948, 2011.
- Lambe, A. T., Ahern, A. T., Williams, L. R., Slowik, J. G., Wong, J. P. S., Abbatt, J. P. D., Brune, W. H., Ng, N. L., Wright, J. P., Croasdale, D. R., Worsnop, D. R., Davidovits, P., and Onasch, T. B.: Characterization of aerosol photooxidation flow reactors: heterogeneous oxidation, secondary organic aerosol formation and cloud condensation nuclei activity measurements, Atmos. Meas. Tech., 4, 445–461, https://doi.org/10.5194/amt-4-445-2011, 2011.

- Lambe, A. T., Onasch, T. B., Croasdale, D. R., Wright, J. P., Martin, A. T., Franklin, J. P., Massoli, P., Kroll, J. H., Canagaratna, M. R., Brune, W. H., Worsnop, D. R., and Davidovits, P.: Transitions from Functionalization to Fragmentation Reactions of Laboratory Secondary Organic Aerosol (SOA) Generated from the OH Oxidation of Alkane Precursors, Environ. Sci. Technol., 46, 5430–5437, https://doi.org/10.1021/es300274t, 2012.
- Lee, B. H., Lopez-Hilfiker, F. D., Mohr, C., Kurtén, T., Worsnop, D. R., and Thornton, J. A.: An iodide-adduct high-resolution timeof-flight chemical-ionization mass spectrometer: Application to atmospheric inorganic and organic compounds, Environ. Sci. Technol., 48, 6309–6317, https://doi.org/10.1021/es500362a, 2014.
- Lopez-Hilfiker, F. D., Mohr, C., Ehn, M., Rubach, F., Kleist, E., Wildt, J., Mentel, Th. F., Lutz, A., Hallquist, M., Worsnop, D., and Thornton, J. A.: A novel method for online analysis of gas and particle composition: description and evaluation of a Filter Inlet for Gases and AEROsols (FIGAERO), Atmos. Meas. Tech., 7, 983–1001, https://doi.org/10.5194/amt-7-983-2014, 2014.
- Lopez-Hilfiker, F. D., Mohr, C., Ehn, M., Rubach, F., Kleist, E., Wildt, J., Mentel, Th. F., Carrasquillo, A. J., Daumit, K. E., Hunter, J. F., Kroll, J. H., Worsnop, D. R., and Thornton, J. A.: Phase partitioning and volatility of secondary organic aerosol components formed from a-pinene ozonolysis and OH oxidation: the importance of accretion products and other low volatility compounds, Atmos. Chem. Phys., 15, 7765–7776, https://doi.org/10.5194/acp-15-7765-2015, 2015.
- Louvaris, E. E., Karnezi, E., Kostenidou, E., Kaltsonoudis, C., and Pandis, S. N.: Estimation of the volatility distribution of organic aerosol combining thermodenuder and isothermal dilution measurements, Atmos. Meas. Tech., 10, 3909–3918, https://doi.org/10.5194/amt-10-3909-2017, 2017.
- Mohr, C., Lopez-Hilfiker, F. D., Yli-Juuti, T., Heitto, A., Lutz, A., Hallquist, M., D'Ambro, E. L., Rissanen, M. P., Hao, L., Schobesberger, S., Kulmala, M., Mauldin, R. L., Makkonen, U., Sipilä, M., Petäjä, T.- and Thornton, J. A.: Ambient observations of dimers from terpene oxidation in the gas phase: Implications for new particle formation and growth, Geophys. Res. Lett., 44, 2958–2966, https://doi.org/10.1002/2017GL072718, 2017.
- Ng, N. L., Canagaratna, M. R., Zhang, Q., Jimenez, J. L., Tian, J., Ulbrich, I. M., Kroll, J. H., Docherty, K. S., Chhabra, P. S., Bahreini, R., Murphy, S. M., Seinfeld, J. H., Hildebrandt, L., Donahue, N. M., DeCarlo, P. F., Lanz, V. A., Prévôt, A. S. H., Dinar, E., Rudich, Y., and Worsnop, D. R.: Organic aerosol components observed in Northern Hemispheric datasets from Aerosol Mass Spectrometry, Atmos. Chem. Phys., 10, 4625– 4641, https://doi.org/10.5194/acp-10-4625-2010, 2010.
- Ortega, A. M., Hayes, P. L., Peng, Z., Palm, B. B., Hu, W., Day, D. A., Li, R., Cubison, M. J., Brune, W. H., Graus, M., Warneke, C., Gilman, J. B., Kuster, W. C., de Gouw, J., Gutiérrez-Montes, C., and Jimenez, J. L.: Real-time measurements of secondary organic aerosol formation and aging from ambient air in an oxidation flow reactor in the Los Angeles area, Atmos. Chem. Phys., 16, 7411–7433, https://doi.org/10.5194/acp-16-7411-2016, 2016.
- Palm, B. B., Campuzano-Jost, P., Ortega, A. M., Day, D. A., Kaser, L., Jud, W., Karl, T., Hansel, A., Hunter, J. F., Cross, E. S., Kroll, J. H., Peng, Z., Brune, W. H., and Jimenez, J. L.: In situ secondary organic aerosol formation from ambient pine forest air

using an oxidation flow reactor, Atmos. Chem. Phys., 16, 2943–2970, https://doi.org/10.5194/acp-16-2943-2016, 2016.

- Pankow, J. F.: An absorption model of gas/particle partitioning of organic compounds in the atmosphere, Atmos. Environ., 28, 185–188, https://doi.org/10.1016/1352-2310(94)90093-0, 1994a.
- Pankow, J. F.: An absorption model of the gas/aerosol partitioning involved in the formation of secondary organic aerosol, Atmos. Environ., 28, 189–193, https://doi.org/10.1016/1352-2310(94)90094-9, 1994b.
- Pankow, J. F., Seinfeld, J. H., Asher, W. E., and Erdakos, G. B.: Modeling the formation of secondary organic aerosol. 1. Application of theoretical principles to measurements obtained in the  $\alpha$ -pinene/,  $\beta$ -pinene/, sabinene/,  $\Delta$ 3-carene/, and cyclohexene/ozone systems, Environ. Sci. Technol., 35, 1164–1172, https://doi.org/10.1021/es001321d, 2001.
- Renbaum-Wolff, L., Grayson, J. W., Bateman, A. P., Kuwata, M., Sellier, M., Murray, B. J., Shilling, J. E., Martin, S. T., and Bertram, A. K.: Viscosity of α-pinene secondary organic material and implications for particle growth and reactivity, P. Natl. Acad. Sci. USA, 110, 8014–8019, https://doi.org/10.1073/pnas.1219548110, 2013.
- Riva, M., Heikkinen, L., Bell, D. M., Peräkylä, O., Zha, Q., Schallhart, S., Rissanen, M. P., Imre, D., Petäjä, T., Thornton, J. A., Zelenyuk, A., and Ehn, M.: Chemical transformations in monoterpene-derived organic aerosol enhanced by inorganic composition, Clim. Atmos. Sci., 2, 1–9, https://doi.org/10.1038/s41612-018-0058-0, 2019.
- Sanchez, J. and Myers, T. N.: Peroxides and Peroxide Compounds, Organic Peroxides, in: Kirk-Othmer Encyclopedia of Chemical Technology, John Wiley & Sons, Inc., Hoboken, NJ, USA, 2000.
- Schobesberger, S., D'Ambro, E. L., Lopez-Hilfiker, F. D., Mohr, C., and Thornton, J. A.: A model framework to retrieve thermodynamic and kinetic properties of organic aerosol from composition-resolved thermal desorption measurements, Atmos. Chem. Phys., 18, 14757–14785, https://doi.org/10.5194/acp-18-14757-2018, 2018.

- Stark, H., Yatavelli, R. L. N., Thompson, S. L., Kang, H., Krechmer, J. E., Kimmel, J. R., Palm, B. B., Hu, W., Hayes, P. L., Day, D. A., Campuzano-Jost, P., Canagaratna, M. R., Jayne, J. T., Worsnop, D. R., and Jimenez, J. L.: Impact of Thermal Decomposition on Thermal Desorption Instruments: Advantage of Thermogram Analysis for Quantifying Volatility Distributions of Organic Species, Environ. Sci. Technol., 51, 8491–8500, https://doi.org/10.1021/acs.est.7b00160, 2017.
- Tong, H., Arangio, A. M., Lakey, P. S. J., Berkemeier, T., Liu, F., Kampf, C. J., Brune, W. H., Pöschl, U., and Shiraiwa, M.: Hydroxyl radicals from secondary organic aerosol decomposition in water, Atmos. Chem. Phys., 16, 1761–1771, https://doi.org/10.5194/acp-16-1761-2016, 2016.
- Vaden, T. D., Imre, D., Beranek, J., Shrivastava, M., Zelenyuk, A., Beránek, J., Shrivastava, M., and Zelenyuk, A.: Evaporation kinetics and phase of laboratory and ambient secondary organic aerosol., P. Natl. Acad. Sci. USA, 108, 2190–2195, https://doi.org/10.1073/pnas.1013391108, 2011.
- Wilson, J., Imre, D., Beránek, J., Shrivastava, M., and Zelenyuk, A.: Evaporation kinetics of laboratory-generated secondary organic aerosols at elevated relative humidity, Environ. Sci. Technol., 49, 243–249, https://doi.org/10.1021/es505331d, 2015.
- Yli-Juuti, T., Pajunoja, A., Tikkanen, O. P., Buchholz, A., Faiola, C., Väisänen, O., Hao, L., Kari, E., Peräkylä, O., Garmash, O., Shiraiwa, M., Ehn, M., Lehtinen, K., and Virtanen, A.: Factors controlling the evaporation of secondary organic aerosol from a-pinene ozonolysis, Geophys. Res. Lett., 44, 2562–2570, https://doi.org/10.1002/2016GL072364, 2017.