REVISITING NICHE FUNDAMENTALS WITH TUKEY DEPTH

J. Orestes Cerdeira, Tiago Monteiro-Henriques, M. João Martins, Pedro C. Silva, Diogo Alagador, Aldina M. A. Franco, Manuel L. Campagnolo, Pedro Arsénio, Francisca C. Aguiar, Mar Cabeza

Jorge Orestes Cerdeira

Departamento de Matemática and Centro de Matemática e Aplicações, Faculdade de Ciências e Tecnologia, Universidade Nova de Lisboa, Quinta da Torre, 2829-516 Caparica, Portugal.

jo.cerdeira@fct.unl.pt

Tiago Monteiro Henriques

Centro de Estudos Florestais, CEF, Instituto Superior de Agronomia, Universidade de Lisboa, Tapada da Ajuda, 1349-017 Lisboa, Portugal.

Centro de Investigação e de Tecnologias Agroambientais e Biológicas, CITAB, Universidade de Trás-os-Montes e Alto Douro, Apartado 1013, 5001-801 Vila Real, Portugal.

tmh@isa.ulisboa.pt

M. João Martins

Centro de Estudos Florestais, CEF, Instituto Superior de Agronomia, Universidade de Lisboa, Tapada da Ajuda, 1349-017 Lisboa, Portugal. mjmartins@isa.ulisboa.pt

Pedro C. Silva

Centro de Estudos Florestais, CEF, Instituto Superior de Agronomia, Universidade de Lisboa, Tapada da Ajuda, 1349-017 Lisboa, Portugal. pcsilva@isa.ulisboa.pt

Diogo Alagador

CIBIO/InBio-UE: Centro de Investigação em Biodiversidade e Recursos Genéticos, Universidade de Évora, 7000-890 Évora, Portugal. alagador@uevora.pt

Aldina M. A. Franco

School of Environmental Sciences, University of East Anglia, Norwich NR4 7TJ, UK. a.franco@uea.ac.uk

Manuel L. Campagnolo

Centro de Estudos Florestais, CEF, Instituto Superior de Agronomia, Universidade de Lisboa, Tapada da Ajuda, 1349-017 Lisboa, Portugal. mlc@isa.ulisboa.pt

Pedro Arsénio

Linking Landscape, Environment, Agriculture and Food, LEAF, Instituto Superior de Agronomia, Universidade de Lisboa, Tapada da Ajuda, 1349-017 Lisboa, Portugal. arseniop@isa.ulisboa.pt

Francisca C. Aguiar

Centro de Estudos Florestais, CEF, Instituto Superior de Agronomia, Universidade de Lisboa, Tapada da Ajuda, 1349-017 Lisboa, Portugal. fraguiar@isa.ulisboa.pt

Mar Cabeza

Centre of Excellence in Metapopulation Research, Department of Biosciences, University of Helsinki, Finland.

mar.cabeza@helsinki.fi

Corresponding author: Jorge Orestes Cerdeira

Type of article: Research article

Running headline: Revisiting niche fundamentals

ABSTRACT

1

20

21

22

- 1. The first attempts to describe species ecological niches were simple geometric

 procedures that depict the niche boundaries directly from environmental data. The

 convex hull was one of such procedures, popular for its simplicity, clear

 ecological rational and precise definition of the niche. However, it lacked the

 ability to differentiate areas of the niche with different probabilities of occurrence

 according to environmental suitability.
- 8 2. We incorporate the Tukey depth, a mathematical tool to measure the centrality of 9 a point within a cloud of points on a multidimensional space, in the convex hull 10 approach to (i) propose a new procedure (CH-Tukey) to estimate species' 11 environmental suitability, and (ii) estimate niche overlap coherently. In addition 12 to a clear ecological rational and simplicity the CH-Tukey procedure has a 13 number of attractive features: use of presence-only data; independence from 14 background data; invariance to scale; robustness to outliers; and the 15 decomposition of the niche into a finite number of iso-suitability levels. 16 permitting the computation of consistent overlap indices. We illustrate the use of 17 CH-Tukey, using occurrence data of the main *Quercus* species and subspecies 18 from Western Mediterranean Europe, comparing its outputs with BIOCLIM and 19 MaxEnt.
 - 3. Results showed distinct niche geometries among the different approaches.

 BIOCLIM produced rectilinear niches reflecting the assumption that ecological variables are independent in their action on the species. CH-Tukey, relaxing this assumption, adjusts niche outer boundary and the inner suitability levels to the

- 1 known occurrences. MaxEnt produced unbounded niche geometries, showing
 2 abrupt shifts in the species response to the environmental variables.
- 4. The niche predictions obtained with geometric approaches, BIOCLIM and
 CH-Tukey, are simpler but better aligned with Hutchinson's niche concept than
 those obtained with MaxEnt, this latter showing ecologically implausible
 relationships with the environmental variables. CH-Tukey and the related overlap
 measures provide an adequate tool to explore niche properties and speciesenvironment relationships.

Keywords: Depth functions, environmental envelope, niche modelling, geometric procedures, Hutchinsonian niche, niche overlap, Quercus species.

INTRODUCTION

Hutchinson's niche concept (Hutchinson, 1957) inspired ecological analyses immediately after its proposal (Whittaker & Levin, 1975), with numerous applications and methods emerging after Hutchinson's seminal paper. Applications are broad, including: studies of resource partitioning in ecological communities (Schoener, 1974), methods on predictive vegetation mapping (Franklin, 1995), explorations of the relation between genome changes and niche differentiation (Rocap *et al.*, 2003), reflections on plant coexistence theories (Silvertown, 2004), or even in taxonomy (Raxworthy *et al.*, 2007). Several other associated ecological concepts followed from Hutchinson's niche concept, e.g.: the guild (Root, 1967), niche breadth and niche overlap (see Colwell & Futuyma, 1971; Hurlbert,

- 1 1978), niche width (see Roughgarden, 1972), the regeneration niche (Grubb, 1977), the
- 2 ontogenetic niche (Werner & Gilliam, 1984), persistence niche (Bond & Midgley, 2001),
- 3 niche conservatism/dynamics (Peterson et al., 1999; Pearman et al., 2008). The uses of
- 4 the Hutchinson's n-dimensional hypervolume concept across ecology and evolution have
- 5 been recently surveyed and discussed by Blonder (2017).
- 6 In recent years, the most conspicuous application of the niche concept has been the
- 7 plethora of methods developed to estimate the suitability of ecological conditions to
- 8 species, or eventually, species distributions (Guisan & Zimmermann, 2000; Franklin,
- 9 2010; Ahmed et al., 2015). Generalized Linear Models (GLM), Generalized Additive
- 10 Models (GAM), Maximum Entropy Modelling (MaxEnt), Mahalanobis distance, point-
- to-point using Gower metric (DOMAIN), Ecological-Niche Factor Analysis (ENFA),
- among others, produce continuous predictions, raising further challenges. For instance,
- they depend on thresholds to delineate a precise niche boundary (see Elith *et al.*, 2006;
- 14 Tsoar et al., 2007; Blonder et al., 2014). Some methods, such as GLM, GAM, MaxEnt,
- 15 Genetic Algorithm for Rule-set Production (GARP) and artificial neural networks, even
- after applying user-defined suitability/probability thresholds, may produce unlimited
- 17 hypervolumes (see Blonder et al., 2014). Although some of those methods fit data well,
- they may exhibit species-environment relationships with lack of ecological plausibility
- 19 (Brewer et al., 2016). Addressing this critical issue and aiming at a better relationship
- with ecological theory, Brewer et al. (2016) proposed a new parametric approach, the
- 21 Plateau climate envelope. A non-parametric procedure, also ecologically meaningful, is
- 22 BIOCLIM, the very first geometric approach to the niche. Revisiting and improving such
- 23 geometric approaches can result interesting, as will be shown hereafter.

1 BIOCLIM (Nix, 1986; Busby, 1991) was one of the first attempts to formalize the 2 Hutchinsonian niche of a species and to produce maps of species potential distributions in 3 a systematic way (Booth et al., 2014). It is a simple geometric procedure, which identifies 4 the range of suitable values for each environmental variable, independently of the other 5 variables, thus, defining the niche as the hyperrectangle bounded by the minimum and 6 maximum values of the environmental variables for the species known occurrences. This 7 "range" notion within each environmental variable is implicit in Hutchinson's article and 8 is central to the niche concept (i.e. the species is expected to survive and reproduce within 9 some known pair of limiting values for each variable, and not outside those conditions). 10 To distinguish suitability levels within the niche, Nix (1986) and Busby (1991) used the 11 nested hyperrectangles obtained from percentiles of each environmental variable 12 individually. In this way they incorporated Hutchison's perception of an optimal part of 13 the niche versus a suboptimal one near the boundaries (Hutchinson, 1957). The most 14 suitable regions are those corresponding to the environmental conditions lying in the 15 innermost hyperrectangles. 16 The simple approach of BIOCLIM exhibits some attractive properties: i) performs 17 geometric operations with straightforward ecological interpretation; ii) delineates precise 18 (limited) boundaries of the niche; iii) only uses presences and does not depend on the 19 background; iv) is invariant to scale (suitability levels do not change under linear 20 transformations of data); v) is robust to outliers (while the outer regions are sensitive to 21 outliers their influence vanishes for the interior); and vi) produces ordinal predictions 22 decomposing the niche into a finite number of iso-suitability regions.

1 BIOCLIM presents however a major drawback: it relies on the assumption that ecological 2 variables are independent in their action on the species. With the aim of relaxing this 3 independence assumption, the HABITAT procedure (Walker & Cocks, 1991) replaced 4 the bounding box envelop used in BIOCLIM by the convex hull of the occurrence points. 5 The convex hull of a set of points defines the minimum convex hypervolume containing 6 all points. Hereafter the construction of the convex hull of occurrences in the 7 environmental space will be referred as the CH-approach. The CH-approach is a 8 comprehensible multidimensional geometric way to circumscribe the niche, but no 9 geometric procedure was used to discriminate suitability levels within it. However, this is 10 achievable in a similar way to the percentile procedure implemented in BIOCLIM by using 11 depth functions, which are procedures to measure the centrality of a point within a data 12 cloud on a multi-dimensional space (Zuo & Serfling, 2000). Cerdeira et al. (2014) 13 suggested the use of depth functions to estimate suitability levels for species in the CH-14 approach. Here we describe and discuss a new method that adds normalized Tukey depth 15 to the CH-approach, and give an application of this proposal to *Quercus* taxa occurring in 16 the Iberian Peninsula, illustrating it further by assessing the overlap of the produced 17 climatic niches. Finally we compare the geometry of the environmental niches produced

19

20

21

18

METHODOLOGIES

Tukey depth

22 The Tukey depth of a point (also called location depth or halfspace depth) is a measure,

by CH-approach using Tukey depth with those of BIOCLIM and MaxEnt.

23 introduced by Tukey (1975), to generalize the order of quantiles for the multivariate case.

- Consider a set P of k points of an n-dimensional space (\mathbb{R}), with >, and an arbitrary
- 2 point q. The Tukey depth of point q with respect to P is defined as $d_a(P)=m/k$, where m is
- 3 the minimum number of points to remove from P so that q is not in the convex hull of the
- 4 remaining points of P. We call m the (Tukey) interiority of point q in P.
- 5 (**Figure 1**)
- 6 Figure 1 depicts a set P of 20 points of \mathbb{R} represented by ", and three points x, y and
- 7 z. The lines passing through each of the points x, y or z, show that removing 1, 3 and 9
- 8 points from P (those that are on one of the sides of the line thought the point) will leave x,
- 9 y and z outside the convex hull of remaining points of P (those that are on the other side
- of the line). It can be easily verified that removing less points would leave x, y or z,
- respectively, inside the convex hull of the remaining points of P. We can therefore
- 12 conclude that the interiorities of points x, y and z are 1, 3 and 9, respectively, and $d_x(P) =$
- 13 1/20, $d_v(P) = 3/20$, and $d_z(P) = 9/20$.
- 14 From the definition, it follows that the maximum interiority cannot exceed

$$=\frac{-+1}{2} \tag{1}$$

- A point q with maximum interiority is a Tukey median point. Point z has interiority 9
- and, since M = 9, it is a Tukey median of P.
- Depth functions define a partition of \mathbb{R} into iso-depth regions, i.e., regions of points
- having equal depth values. Figure 2A shows the iso-depth regions corresponding to the
- set of 10 points of \mathbb{R} represented by " ". Note that the regions of points with Tukey
- 20 depth greater than or equal to m/k, , are nested, i.e., $\mathbb{R} = \neg \neg \neg \neg \neg$,

- 1 with increasing depth toward the innermost region. The region $(m \ge 1)$ could
- 2 alternatively be defined as the intersection of all convex hulls of the subsets of P with k-
- $3 \quad (m-1) \text{ points.}$
- 4 Depth values of the points of a set P can give indication on how P is arranged,
- 5 distinguishing between configurations of points that are concentrated in the "interior",
- 6 from those which occur mainly on the "margins". Silva et al. (2014) have recently
- 7 established that, if the points in P are uniformly distributed, which is a borderline
- between the two above configurations, the average $d_a(P)$, for q in P, approximates $1/2^{n+1}$.
- 9 This value, which only depends on the dimension (n) of the representation space,
- provides a sound threshold to distinguish configurations where the concentration is in the
- "interior" (average $d_a(P) > 1/2^{n+1}$), from configurations where points occur mainly on the
- 12 "margins" (average $d_a(P) < 1/2^{n+1}$).

14 Assessing environmental suitability

- Let P be the set of k points in \mathbb{R} , each point indicating the n environmental conditions of
- 16 a location where a species occurs. To every point q of P we calculate the depth $d_q(P)$ and
- divide it by M/k. With this rescaling, points with interiority M (if any) score 1. We
- propose to interpret these normalized depth values as a measure of environmental
- suitability for the species. The value 1 corresponds to environmental conditions of sites
- that belong to the most inner region of the estimated niche. Note that such
- 21 environmental conditions might not exist.

- 1 We call this the CH-Tukey procedure. Under the equilibrium assumption, areas of greater
- 2 $d_q(P)$ values correspond to the optimal parts of the niche (sensu Hutchinson 1957), i.e.
- 3 areas where species performance is expected to be higher (e.g. growth rate or species
- 4 fitness). The innermost region, where $d_a(P)$ reach its maximum value, is the set of Tukey
- 5 median points.
- 6 It should be clear that CH-Tukey procedure assumes unimodal species responses, in
- 7 relation to the considered environmental variables. More specifically, CH-Tukey
- 8 procedure assumes that the suitability of every point of the line connecting two points q_1
- 9 and q_2 of the environmental space, will be at least as large as the minimum between the
- 10 suitability of q_1 and q_2 .
- Our proposal can be viewed as a refinement of the procedure of BIOCLIM to define
- different levels of suitability, relaxing the independence assumption. Indeed, the way
- 13 BIOCLIM differentiates suitability levels in the bounding box of P, using percentiles of
- each environmental variable individually, is actually the output of a depth function that
- can be read as follows. The environmental suitability of a location corresponding to point
- 16 q on the environmental space, is twice the value of $d'_{a}(P) = m'/k$, where m' is the
- minimum number of points to remove from P so that q is not in the minimum
- 18 hyperrectangle containing the remaining points of P. The multiplication by two is a
- rescaling so that the environmental suitability of points in the most inner hyperrectangle
- 20 equals one. Figure 2B depicts the regions of points with equal d' values corresponding to
- 21 the set of 10 points of \mathbb{R} represented by " " (the same points as in Figure 2A). Note
- 22 that the regions of points for which d' is greater than or equal to m/k, ', are also
- 23 nested, i.e., $\mathbb{R} = ' \supset ' \supset \cdots \supset '$, where ' = -. The region $' (m \ge 1)$ is

- 1 the intersection of all bounding boxes of the subsets of P with k-(m-1) points. Therefore,
- 2 we have: i) M'=M if n=1 and M'>M for ≥ 2 ; ii) $'\supset$, for m=1,...,M, since the
- 3 bounding box of a set of points always contains the convex hull of the same set; or,
- 4 equivalently, for any q, $d'_{a}(P) = m'/k \ge m/k = d_{a}(P)$.
- 5 (**Figure 2**)
- 6 Both procedures differentiate the outer polygon in nested regions of increasing suitability
- 7 for the interior, CH-Tukey producing tighter regions than BIOCLIM. Points outside the
- 8 outer polygon are considered unsuitable; points in the margins have small suitability,
- 9 while the largest suitability is assumed to occur in the inner regions.

Assessing the overlap of niches

10

- 12 Several indices have been proposed to assess the niche overlap between two species, by
- comparing the respective suitability (or presence/absence) on a finite number of points,
- which correspond to cells of the geographical area (see Warren *et al.*, 2008, for a survey).
- 15 However, to assess the overlap of the environmental niches of two species all points of
- the environmental space should be considered and not only the projection into the cells of
- 17 a map. This is an issue that is particularly important in climate change scenarios. Villéger
- 18 et al. (2011) proposes to evaluate the functional dissimilarity between two (or more)
- species assemblages, considering the volumes of the union and of the intersection of the
- 20 convex hulls of the points representing each species assemblage in the space of their
- 21 functional traits. This approach can be easily transposed to the evaluation of the overlap
- between the environmental niches of two species, giving an overlap index defined as the

- 1 ratio between the volume of the intersection and the volume of the union of the convex
- 2 hulls of the occurrences of both species. This measure can be interpreted as the Jaccard
- 3 index for the volume and is the natural way to estimate niche overlap when using the CH-
- 4 approach to describe the niche. It has the disadvantage of not distinguishing between
- 5 intersections occurring in regions of high suitability and those occurring in low suitability
- 6 areas of the niches.
- 7 As Tukey depth divides the niche into a finite number of iso-suitability regions, it allows
- 8 to generalize Pianka and Jaccard indices incorporating volumes and suitability.
- 9 Consider two species A and B, and the partition of the environmental space into and
- regions of (non zero) iso-depth for each species. Denote by 0 = !"#\$ < !"#\$ <
- 11 $\cdots < !$, "#\$ and 0 = !"(\$ < !"($\$ < \cdots < !$)"(\$ all possible values of suitability for
- species A and B, respectively, and let $*_+$, with $(i,j) \neq (0,0)$, be the volume of the (possibly
- disconnected) region composed by all the environments having suitability !₊"#\$ for
- species A and ! "(\$ for species B, and * arbitrary. The volumes can be arranged in a
- 15 (+ 1) by (+1) matrix $_{-} = .*_{+}$. This matrix contains all the information on
- intersections of the environmental niches, taking into account distinct levels of suitability
- for both species. Note that, if !+"#\$ and !,"(\$ only take 0-1 values, i.e. resulting from
- binary prediction models, such as the CH-approach, $_{\perp} = .*_{+,}/$ is a 2 by 2 matrix, where
- * is the volume of the intersection of the niches, and * (*) is the volume of the
- 20 niche of species A (B) not shared by species B (A).
- From M_V we define the matrix $M_A(M_B)$ of the asymmetric overlap of species B (A) on A
- 22 (B), multiplying the row i of M_V by !,"#\$ (the column j of M_V by !,"(\$). Each entry (i,j)

- of matrix $M_A(M_B)$ is the volume $*_+$ of the region with iso-suitability $!_+$ "#\$, for species A
- 2 and !"(\$ for species B, weighted by ! $_+$ "#\$ (!"(\$).
- 3 We propose to use the matrices above to define two overlap indices. The first is

$$M_{A}(i,j) M_{B}(i,j)$$

$$WJ = \frac{M_{A}(i,j) M_{B}(i,j) M_{B}(i,j)}{k_{A} k_{B}}$$

$$2M_{A}(i,j) M_{B}(i,j) M_{A}(i,j) M_{B}(i,j)$$

$$i_{0 j_{0}} i_{1 j_{1}}$$

$$k_{A} k_{B}$$

$$i_{A} k_{B}$$

$$i_{B} k_{A} k_{B}$$

- 5 Index wJ incorporates the different levels of suitability on the ratio of the volumes of the
- 6 intersection and of the union of the niches. If !₊"#\$ and !_|"(\$ only assumes values 0 or 1,

8
$$wJ = \frac{M}{M_{V}(1,1)} \frac{M}{M_{V}(1,0) M_{V}(0,1) M_{V}(1,1)}$$

which is the Jaccard index for the volume. For the particular case when suitability is

given by CH-approach, the niches are estimated by the convex hulls of the occurrences of

- both species, and wJ becomes the index of Villéger et al. (2011).
- When suitability is obtained with Tukey depth, wJ distinguishes between intersections
- occurring in regions of high suitability and those occurring in low suitability areas of the
- 15 niches.

7

- 1 Figure 3 illustrates iso-suitability regions for each of three hypothetical species A, B and
- 2 B' (top) in the same environmental space, and how they are arranged together (bottom).
- 3 The volumes of the intersections/unions of the convex hulls for A and B, and for A and
- 4 B' are exactly the same. Thus, Jaccard index returns the same value (0.2179) for both
- 5 pairs. Index wJ gives 0.3579 for pair (A, B), where intersection occurs mainly in the
- 6 regions of high suitability, and 0.1604 for pair (A, B') where the regions of high
- 7 suitability almost do not overlap.

- 9 The second overlap index we propose is a *cosine* index, which is an adaptation of Pianka
- index used to measure resource overlap between two species (Pianka, 1973). Consider
- matrices M_A and M_B as vectors of length $(+1)\times(+1)$. The cos index is defined as the
- cosine of the angle between these vectors.

$$cos = \frac{M_{A}^{k_{A}} k_{B}}{M_{A}(i,j) M_{B}(i,j)}$$

$$\sqrt{M_{A}^{k_{A}} k_{B}}$$

$$M_{A}(i,j)^{2} \sqrt{M_{A}^{k_{A}} k_{B}}$$

$$i0 \ j0 \qquad i0 \ j0$$

$$k_{A} k_{B}$$

$$S_{i}(A)S_{j}(B) \ (M_{V}(i,j))^{2}$$

$$\frac{i0 \ j0}{\sqrt{\sum_{k_A = k_B}^{k_A = k_B}} s_i(A)M_V(i,j)^2} \sqrt{\sum_{j=0}^{k_A = k_B} s_j(B)M_V} (i,j)^2$$

- When the niches do not overlap, vectors are orthogonal and *cos*=0. When high and low
- suitability occur in the same regions for both species, $\cos \approx 1$.

ILLUSTRATIVE EXAMPLE

- 3 The climatic suitability of *Quercus* taxa in the Iberian Peninsula
- 4 As an example of application of the above-described CH-Tukey procedure, we computed
- 5 climatic suitability for *Quercus* taxa occurring in the Iberian Peninsula.
- 6 Quercus spp. are widespread in the Northern Hemisphere (Manos et al., 1999). In the
- 7 Iberian Peninsula *Quercus* taxa occur in most of the territory being an important
- 8 economic and conservation target (see Bugalho et al., 2011; Correia et al., 2017). For
- 9 example, several of the local economies are dependent on cork extraction, cork represents
- 10 2.5% of Portugal exports, and acorns from *Quercus rotundifolia* and *Q. suber* are
- particularly important in many animals' diet (Rodríguez-Estévez et al., 2009; Bugalho et
- 12 al., 2011). Q. robur is an important forestry species. Q. faginea and Q. pyrenaica are
- valuable to local communities for the production of firewood (Gonçalves et al., 2012;
- 14 Cerqueira *et al.*, 2013).
- 15 In this paper we focus our analyses on six species/subspecies: Q. rotundifolia (= Q. ilex
- subsp. ballota), Q. suber, Q. pyrenaica, Q. faginea subsp. faginea, Q. faginea subsp.
- 17 broteroi and Q. robur. Occurrences were obtained from the Proyecto Anthos (Anthos,
- 18 2011) and from the Herbarium "João de Carvalho e Vasconcellos" (Lisbon, LISI) and
- include a mix of point and grid-based data of several natures and types. Positional
- accuracy of the species observations data was analysed and improved when appropriate.
- 21 We used four climatic variables known to affect and limit species distributions,
- downloaded from the Worldclim website (Hijmans et al., 2005;

- 1 http://www.worldclim.org) at the 1km × 1km grid cell resolution. Two variables involve
- 2 temperature (isothermality and the minimum temperature of coldest month) and two
- 3 variables involve precipitation (annual precipitation and the precipitation of driest
- 4 quarter).
- 5 For each selected taxon, we applied the CH-Tukey and the BIOCLIM procedures in order
- 6 to produce geographical maps distinguishing suitability levels. The Tukey depth of each
- site within the Iberian Peninsula (ca. $1 \text{ km} \times 1 \text{ km}$ resolution), with respect to the
- 8 occurrences of each taxon, was determined using the function *depth* from the R package
- 9 depth version 2.0-0 (Genest et al., 2012). The depth values were then rescaled dividing
- by 0" -3\$/23, where k is the number of occurrences of the taxon, thus obtaining a
- 11 normalized Tukey depth. The BIOCLIM suitability associated to each site was obtained
- using the functions bc and predict from the R package dismo version 1.1-1 (Hijmans et
- 13 *al.*, 2016).
- 14 The maps clearly show that the inland non-white areas for CH-Tukey, (Figure 4) are
- 15 contained in the inland non-white areas for BIOCLIM, ' (Figure 5), denoting the fact
- that the convex hull is a tighter envelope than the hyperrectangle.
- 17 A robust validation of these models imply controlled experiments using a small set of
- 18 factors considered limitative of species fitness. Given that these experiments are typically
- 19 time-consuming, expensive and therefore are seldom performed, expert knowledge is a
- valuable source of information to discuss niche predictions.
- 21 The obtained spatial extents are quite consistent with experts' knowledge (see e.g. Rivas-
- 22 Martínez & Sáenz Laín, 1991), specially for the CH-Tukey approach (e.g. Q. robur in the
- NW of the Peninsula, O. faginea subsp. broteroi in the SW, O. suber mainly in the W,

- 1 while the other three species occur mostly in the interior parts of the Peninsula). For the
- 2 BIOCLIM procedure such spatial coincidence is not so evident and can only be
- 3 approximated if the lower levels of suitability are discarded.
- 4 For some taxa (Q. suber, Q. rotundifolia and both Q. faginea subspecies) the maps show
- 5 a low density of occurrences in regions of predicted high suitability. This may result,
- 6 among other possible reasons, from: i) missing relevant environmental variables (e.g. soil
- 7 characteristics or other climatic variables) in the model; ii) incomplete sampling of
- 8 occurrences in the corresponding geographic areas; iii) some ecological factors that lead
- 9 to the exclusion of the species from that area (e.g. competition, parasitism, disease); or
- even iv) local extinction due to human overuse of *Quercus* forests, conversion to
- agricultural fields, managed forests, pastures or urban fabric. In the case of *Q. faginea*
- subsp. faginea and Q. suber it is quite plausible that a soil-related environmental variable
- is missing in the model, as the former is known to prefer base-rich or clay-rich soils,
- while the latter is calcifugous (Franco, 1990). In the case of *Q. faginea* subsp. broteroi,
- 15 the biogeographic disjunction that the occurrence points show (i.e. one cloud of
- occurrences concentrated in the west of Portugal and a second cloud in the central part of
- 17 Spain) may lead to an unoccupied niche core. The findings of Silva et al. (2014),
- mentioned above, can be used to detect those less frequent distribution patterns within the
- 19 niche, previously to modelling.
- 20 In general, with the CH-Tukey approach, the maximum suitability obtained for each of
- 21 the studied *Quercus* taxa is relatively low. This is not problematic a priori and might give
- 22 further structural information on the species niche and on the study area, as: i) the
- 23 maximum suitability level of 1, might not be achieved for some configurations of

1 occurrence points in the environmental space; ii) when achieved, the volume of the 2 region with suitability 1 can be very small; iii) even when the volume of the Tukey 3 median region is conspicuous, it may correspond to combinations of the n variables that 4 do not come into existence within the study area. 5 (Figure 4) 6 (Figure 5) 7 8 Quantifying the environmental niche overlaps of Iberian Quercus taxa 9 In the space defined by the four environmental variables (isothermality, minimum 10 temperature of coldest month, annual precipitation and precipitation of driest quarter), we uniformly generated $N=45\times10^6$ points within the bounding box of the occurrences of all 11 12 taxa, and computed the Tukey suitability of each of these points with respect to the 13 occurrences of each taxon. To assess niche overlap for each pair of taxa, A and B, we estimated the volumes $*_+$, for "4, 6\$ \neq "0,0\$, as the number of points with Tukey 14 suitability $!_+$ "#\$, for taxon A, and Tukey suitability $!_-$ "(\$, for taxon B, divided by N. 15 16 Results for the two proposed overlap indices (wJ and cos) are presented in Table 1. 17 (Table 1) 18 Both wJ and cos showed that Q. robur and Q. pyrenaica presented the highest overlap, 19 while Q. robur and Q. faginea subsp. broteroi presented the lower overlap. 20 In order to visualize graphically the relative positions of occurrences and iso-depth 21 regions of the pairs of taxa that, in the 4-dimensional space, were more separated (Q. 22 robur vs. Q. faginea subsp. broteroi), and less separated (Q. robur vs. Q. pyrenaica)

- 1 (Table 1), we projected the occurrences of each of these taxa in the 2-dimensional
- 2 environmental space of variables minimum temperature of the coldest month and
- 3 precipitation of the driest quarter (Figure 6). Computing the overlap indices in this 2-
- 4 dimensional space, we obtained wJ = 0.9177 and cos = 0.3450 for the first pair (Q. robur,
- 5 Q. pyrenaica) and wJ = 0.3274 and cos = 0.0283 for the second pair (Q. robur, Q. faginea
- 6 subsp. broteroi). These values agree with the graphs of Figure 6, where it is shown a clear
- 7 overlap of the niches of Q. robur and Q. pyrenaica while the niche of Q. faginea subsp.
- 8 broteroi is mainly concentrated in the peripheral areas of the Q. robur niche.
- 9 In general, Table 1 shows values that are coherent with the known regional co-occurrence
- patterns of the studied *Quercus* (see e.g. Rivas-Martínez & Sáenz Laín, 1991; Capelo et
- 11 al., 2007). For example, Q. robur co-occurs frequently with Q. pyrenaica in the highlands
- of the northwest of the Iberian Peninsula, and with Q. suber in the lowlands. Q. pyrenaica
- 13 co-occurs with O. rotundifolia in the continental plateaux of the Iberian Meseta and with
- 14 Q. suber in the warmer areas of the centre-south. Q. rotundifolia and Q. suber co-occur in
- the SW of the Peninsula. The referred pairs correspond both to the highest values of the
- indices shown in Table 1 and to intermingled vegetation series mapped as a mosaic in the
- 17 work of Capelo *et al.* (2007).

18 (**Figure 6**)

Comparing niche geometries

- 20 In order to illustrate and compare niche geometries produced by different approaches, we
- compared outputs from CH-Tukey to BIOCLIM, and also MaxEnt (Phillips et al., 2006).
- 22 MaxEnt is one of the most used modelling algorithms, to obtain species environmental
- suitability maps (Ahmed *et al.*, 2015). We ran these three models using the two variables

- 1 referred above (minimum temperature of the coldest month and precipitation of the driest
- 2 quarter) and depicted their outputs both in the environmental and in the geographic space
- 3 (Figure 7) for *Q. robur*, *Q. pyrenaica* and *Q. faginea* subsp. broteroi. Default MaxEnt
- 4 parameterization was used.
- 5 Inspecting Figure 7 (B, C, D) we can visualize how BIOCLIM relies on the assumption
- 6 that ecological variables are independent in their action on the species, producing
- 7 rectilinear niche geometries, while CH-Tukey relaxes it, adjusting its outer boundary and
- 8 the inner suitability levels to the known occurrences. Both BIOCLIM and CH-Tukey are
- 9 built upon the "range" notion, implicit in Hutchinson's seminal work, i.e., if you select
- 10 two points from the species niche representing environmental conditions known to be
- suitable for that species, you should not expect a decrease in suitability between these two
- points, at least while only those variables are considered. MaxEnt neglects this, producing
- 13 not only unbounded geometries, but also allowing (visible for the three tested taxa) a
- decrease in suitability between two points of the niche. Although relying on unimodal
- 15 geometries, BIOCLIM and CH-Tukey approximate better Hutchinson's notion of optimal
- versus suboptimal niche regions than the geometry produced by MaxEnt, which proposes
- 17 high suitability values for unbounded or extrapolative areas (see "clamping" in Merow et
- 18 al., 2013; check Q. faginea subsp. broteroi output in Figure 7D, values of min
- 19 Temperature greater than 10.0 °C or smaller than -10.0 °C and Precipitation around 50
- 20 mm). This is in agreement with the known limitations of correlative models concerning
- 21 extrapolation (Merow et al., 2013). According to the same authors, MaxEnt can produce
- 22 intricate model outputs of hard ecological interpretation, especially when regularization is
- used. This is apparent in the sudden shifts in suitability observed in the MaxEnt output

1 (within the species niche). Even if some shifts could relate to interesting or interpretable

2 facts (e.g. coming from the background data), we believe that they deviate this modelling

technique from the seminal ecological theory.

4

6

7

8

9

10

11

12

13

14

15

16

17

18

19

20

21

22

3

5 (**Figure 7**)

A thorough and careful setting of the input parameters in MaxEnt can improve the ecological interpretation of the output (Merow et al., 2013), but such parameterization is not always achievable using presence-only data coming from usual databases. Even so, MaxEnt produces geographical maps that are closely linked to the known occurrences of points (see Figure 7A for the occurrence points of each taxa), which is expected from a correlative approach. This tends to hinder fundamental discussions, such as: i) niche boundaries and niche optimality; ii) interpolation vs. extrapolation of the observations; and iii) possible discrepancies between niche optimality and the geographic distribution (e.g. niche is highly suitable according to the studied environmental variables, but there are no occurrences in the corresponding geographic regions, possibly indicating that other important variables lack in the model). Incorporating putative niche voids directly in the modelling approaches has been considered useful, especially while addressing niches using real-world examples (Walker & Cocks, 1991; Blonder et al., 2014). Nevertheless, even when dealing with real-world examples, general niche premises and properties should hold (e.g. the "range" notion, niche boundaries and optimality) and these can be consistently addressed and explored

using appropriate methodological approaches such as BIOCLIM and CH-Tukey.

CONCLUDING REMARKS

3 The CH-Tukey procedure allows the construction of environmental suitability maps, as 4 shown for the selected *Quercus* taxa, differentiating iso-suitability levels within the CH-5 approach. Such levels are drawn from occurrence data only, in the selected environmental 6 hyperspace, using normalized Tukey depth, and can, subsequently, be used to derive 7 overlap measures coherently. 8 Carpenter et al. (1993) listed four disadvantages linked with the CH-approach, two 9 related to computational difficulties which are no longer a problem for current computers. 10 The other two were related to the tightly constrained envelope produced by the CH-11 approach and the exclusion of many similar sites (close to, but outside, the convex hull 12 boundary). But this may be arguable since the convex hull is a natural boundary between 13 interpolation and extrapolation (see Hirzel & Le Lay, 2008). We stress that the CH-Tukey 14 procedure is appropriate to interpolate from the input data, as the model simply considers 15 unsuitable all environmental combinations outside the convex hull (i.e. it does not 16 extrapolate). Nevertheless, similar sites in the immediate outside neighbourhood of the 17 convex hull, which are considered unsuitable in the CH-Tukey procedure, could be 18 modelled separately with other techniques, in a coupled approach, assuming that 19 extrapolation is acceptable. 20 As shown in the maps obtained with CH-Tukey methodology, it is possible that higher 21 suitability regions have low density of occurrences (see e.g. Q. faginea subspecies). This 22 results from the combinatorial-geometrical nature of the approach that does not rely so 23 strongly on point density to predict suitability. As the procedure is geometrically

- 1 interpretable, hypotheses can be sketched from the obtained patterns, given the used
- 2 environmental variables, and tested subsequently. This interpretability ensures a seamless
- 3 relationship with Hutchinsonian niche concept, a feature frequently asked by ecologists
- 4 and epistemologists (Austin, 2002; Elith & Leathwick, 2009; Brewer et al., 2016).
- 5 In conclusion, the CH-Tukey procedure presents several relevant features that can be
- 6 considered advantages in some ecological niche modelling contexts, such as:
- 7 i) The modelling process is easily interpretable geometrically, with precise definition of
- 8 the estimated niche boundaries.
- 9 ii) It uses only presence data and does not depend on the background data.
- 10 iii) The suitability levels are invariable with respect to the ecological variables units.
- iv) The regions of the multivariate space corresponding to high Tukey depth values are
- 12 robust to outliers, in accordance to what occurs with the univariate median of a set of
- values.
- 14 v) It produces ordinal predictions, decomposing the niche into a finite number of iso-
- suitability levels, permitting the computation of consistent overlap indices.
- vi) A tight envelope is adjusted to the points cloud and, consequently, the independence
- of the ecological variables action on the species is not assumed (i.e., allowing for
- interactions between variables); moreover, the output of the procedure results uniquely
- 19 from interpolations of the input data.
- The main disadvantage of the CH-Tukey procedure is that depth computation is time
- 21 demanding, and heuristic approximations may be required depending on the number of

- 1 occurrence points used as input data and on the spatial resolution of the predictions.
- 2 Heuristics are already implemented in package *depth* (Genest *et al.*, 2012).

4

ACKNOWLEDGMENTS

- 5 This work was supported by the Portuguese Foundation for Science and Technology
- 6 (FCT) through the projects UID/MAT/00297/2013, CMA (Centro de Matemática
- 7 Aplicada); UID/AGR/00239/2013, CEF (Centro de Estudos Florestais) under
- 8 FEDER/POCI; and PTDC/AAC-AMB/ 113394/2009. T. Monteiro-Henriques was
- 9 partially funded by a post-doctoral grant within Project SUSTAINSYS: Environmental
- 10 Sustainable Agro-Forestry Systems NORTE-07-0124-FEDER-0000044 (ON.2-QREN-
- 11 FEDER and PIDDAC-FCT-MEC). D. Alagador was funded through the project
- 12 PTDC/AAG-GLO/3979/2014 via FEDER through the COMPETE program and national
- funding. D. Alagador and F. C. Aguiar are also supported with FCT post-doctoral
- 14 fellowships (SFRH/BPD/104077/2014 and SFRH/BPD/112417/2015, respectively).

15

16

AUTHORS' CONTRIBUTIONS

- 17 JOC, TMH, MJM and PCS conceived the ideas and designed methodology; MJM, TMH
- and MLC dealt with computational issues; TMH, PA, MJM and FCA collected, analysed
- and interpreted the data; JOC, TMH, MJM and PCS led the writing of the manuscript.
- 20 DA, AF and MC revised it and contributed to the development of ideas. All authors
- 21 contributed critically to the manuscript.

REFERENCES

- 2 Ahmed, S.E., McInerny, G., O'Hara, K., Harper, R., Salido, L., Emmott, S. & Joppa, L.N.
- 3 (2015) Scientists and software surveying the species distribution modelling
- 4 community. Diversity and Distributions, **21**, 258–267.
- 5 Anthos (2011) Sistema de información de las plantas de España. Real Jardín Botánico,
- 6 CSIC- Fundación Biodiversidad. Downloaded in November 16, 2011.
- Austin, M.P. (2002) Spatial prediction of species distribution: an interface between
- 8 ecological theory and statistical modelling. *Ecological Modelling*, **157**, 101–118.
- 9 Blonder, B. (2017) Hypervolume concepts in niche- and trait-based ecology. *Ecography*,
- **40**, 001–013.
- Blonder, B., Lamanna, C., Violle, C. & Enquist, B.J. (2014) The n-dimensional
- hypervolume. *Global Ecology and Biogeography*, **23**, 595–609.
- Bond, W.J. & Midgley, J.J. (2001) Ecology of sprouting in woody plants: the persistence
- niche. Trends in Ecology & Evolution, **16**, 45–51.
- Booth, T.H., Nix, H.A, Busby, J.R. & Hutchinson, M.F. (2014) BIOCLIM: the first
- species distribution modelling package, its early applications and relevance to
- most current MAXENT studies. *Diversity and Distributions*, **20**, 1–9.
- Brewer, M.J., O'Hara, R.B., Anderson, B.J., & Ohlemüller, R. Plateau: A New Method
- for Ecologically Plausible Climate Envelopes for Species Distribution Modelling.
- 20 *Methods in Ecology and Evolution*, **7**, 1489–1502.

- 1 Bugalho, M.N., Caldeira, M.C., Pereira, J.S., Aronson, J. & Pausas, J.G. (2011)
- 2 Mediterranean cork oak savannas require human use to sustain biodiversity and
- 3 ecosystem services. Frontiers in Ecology and the Environment, 9, 278–286.
- 4 Busby, J.R. (1991) BIOCLIM A Bioclimate Analysis and Prediction System. Nature
- 5 Conservation: Cost Effective Biological Surveys and Data Analysis (ed. by C.R.
- 6 Margules and M.P. Austin), pp. 64–68. CSIRO.
- 7 Capelo, J., Mesquita, S., Costa, J.C., Ribeiro, S., Arsénio, P., Neto, C., Monteiro-
- 8 Henriques, T., Aguiar, C., Honrado, J., Espírito-Santo, D. & Lousã, M.F. (2007)
- 9 A methodological approach to potential vegetation modeling using GIS
- techniques and phytosociological expert-knowledge: application to mainland
- 11 Portugal. *Phytocoenologia*, **37**, 399–415.
- 12 Carpenter, G., Gillison, A.N. & Winter, J. (1993) DOMAIN: a flexible modelling
- procedure for mapping potential distributions of plants and animals. *Biodiversity*
- 14 *and Conservation*, **2**, 667–680.
- 15 Cerdeira, J.O., Monteiro-Henriques, T., Martins, M.J., Silva, P.C., Alagador, D. &
- Franco, A.M.A. (2014) Mathematical contributions to link biota with
- 17 environment. *Journal of Vegetation Science*, **25**, 1148–1153.
- 18 Cerqueira, M., Gomes, L., Tarelho, L. & Pio, C. (2013) Formaldehyde and acetaldehyde
- 19 emissions from residential wood combustion in Portugal. *Atmospheric*
- 20 *Environment*, **72**, 171–176.
- 21 Colwell, R. & Futuyma, D. (1971) Measurement of Niche Breadth and Overlap. *Ecology*,
- **52**, 567–576.

- 1 Correia, R.A., Bugalho M.N., Franco A.M.A. & Palmeirim J.M. Contribution of Spatially
- 2 Explicit Models to Climate Change Adaptation and Mitigation Plans for a Priority
- 3 Forest Habitat. Mitigation and Adaptation Strategies for Global Change, in press
- 4 1–16.
- 5 Elith, J., Graham, C.H., Anderson, R.P., Dudík, M., Ferrier, S., Guisan, A., Hijmans, R.J.,
- 6 Huettmann, F., Leathwick, J.R., Lehmann, A., Jin Li, Lohmann, L.G., Loiselle,
- 7 B.A., Manion, G., Moritz, C., Nakamura, M., Nakazawa, Y., McC. Overton, J.,
- 8 Peterson, A.T. & Phillips, S.J. (2006) Novel methods improve prediction of
- 9 species' distributions from occurrence data. *Ecography*, **29**, 129–151.
- 10 Elith, J. & Leathwick, J.R. (2009) Species Distribution Models: Ecological Explanation
- and Prediction Across Space and Time. Annual Review of Ecology, Evolution, and
- 12 *Systematics*, **40**, 677–697.
- 13 Franco, J. do A. (1990) 3. Quercus L.*. Flora iberica: Plantas vasculares de la Península
- 14 Ibérica e Islas Baleares (ed. by S. Castroviejo, M. Laínz, G. López González, P.
- Montserrat, F. Muñoz Garmendia, J. Paiva, and L. Villar), pp. 15–36. Real Jardín
- 16 Botánico Consejo Superior de Investigaciones Científicas, Madrid.
- 17 Franklin, J. (1995) Predictive vegetation mapping: geographic modelling of biospatial
- patterns in relation to environmental gradients. *Progress in Physical Geography*,
- 19 **19**, 474–499.
- 20 Franklin, J. (2010) Mapping Species Distributions: Spatial Inference and Prediction,
- 21 Cambridge University Press, Cambridge.

- 1 Genest, M., Masse, J.-C. & Plante, J.-F. (2012) depth: Depth functions tools for
- 2 *multivariate analysis (version 2.0-0)*, http://CRAN.R-project.org/package=depth.
- 3 Gonçalves, C., Alves, C. & Pio, C. (2012) Inventory of fine particulate organic
- 4 compound emissions from residential wood combustion in Portugal. *Atmospheric*
- 5 *Environment*, **50**, 297–306.
- 6 Grubb, P.J. (1977) The maintenance of species-richness in plant communities: the
- 7 importance of the regeneration niche. *Biological Reviews*, **52**, 107–145.
- 8 Guisan, A. & Zimmermann, N.E. (2000) Predictive habitat distribution models in
- 9 ecology. *Ecological Modelling*, **135**, 147–186.
- Hijmans, R.J., Cameron, S.E., Parra, J.L., Jones, P.G. & Jarvis, A. (2005) Very High
- Resolution Interpolated Climate Surfaces for Global Land Areas. *International*
- 12 *Journal of Climatology*, **25**, 1965–1978.
- Hijmans, R.J., Phillips, S., Leathwick, J. & Elith, J. (2016) dismo: Species distribution
- 14 *modeling*, http://CRAN.R-project.org/package=dismo.
- Hirzel, A.H. & Le Lay, G. (2008) Habitat suitability modelling and niche theory. *Journal*
- 16 *of Applied Ecology*, **45**, 1372–1381.
- Hurlbert, S. (1978) Measurement of Niche Overlap and Some Relatives. *Ecology*, **59**, 67–
- 18 77.
- 19 Hutchinson, G.E. (1957) Concluding remarks. Cold Spring Harbor Symposia on
- 20 *Quantitative Biology*, **22**, 415–427.

- 1 Manos, P.S., Doyle, J.J. & Nixon, K.C. (1999) Phylogeny, biogeography, and processes
- of molecular differentiation in Quercus subgenus Quercus (Fagaceae). *Molecular*
- *Phylogenetics and Evolution*, **12**, 333–349.
- 4 Merow, C., Smith, M.J., Silander Jr, J.A. 2013 A practical guide to MaxEnt for modeling
- 5 species' distributions: what it does, and why inputs and settings matter.
- 6 *Ecography*, **36**, 1058-1069.
- 7 Nix, H.A. (1986) A biogeographic analysis of Australian elapid snakes. Atlas of elapid
- 8 snakes of Australia (ed. by R. Longmore), pp. 4–15. Australian Government
- 9 Publishing Service, Canberra.
- 10 Pearman, P.B., Guisan, A., Broennimann, O. & Randin, C.F. (2008) Niche dynamics in
- space and time. *Trends in Ecology & Evolution*, **23**, 149–158.
- 12 Peterson, A.T., Soberon, J. & Sanchez-Cordero, V. (1999) Conservatism of ecological
- niches in evolutionary time. *Science*, **285**, 1265–1267.
- 14 Phillips, S.J., Anderson, R.P., Schapire, R.E. (2006) Maximum entropy modeling of
- species geographic distributions. *Ecological Modelling*, **190**, 231-259.
- Pianka, E.R. (1973) The Structure of Lizard Communities. *Annual Review of Ecology*
- 17 *and Systematics*, **4**, 53–74.
- 18 Raxworthy, C.J., Ingram, C.M., Rabibisoa, N. & Pearson, R.G. (2007) Applications of
- 19 Ecological Niche Modeling for Species Delimitation: A Review and Empirical
- 20 Evaluation Using Day Geckos (Phelsuma) from Madagascar. Systematic Biology,
- **56**, 907.

- 1 Rivas-Martínez, S. & Sáenz Laín, C. (1991) Enumeración de los Quercus de la Península
- 2 Ibérica. Rivasgodaya, 6, 101–110.
- 3 Rocap, G., Larimer, F.W., Lamerdin, J., Malfatti, S., Chain, P., Ahlgren, N.A., Arellano,
- 4 A., Coleman, M., Hauser, L., Hess, W.R., Johnson, Z.I., Land, M., Lindell, D.,
- 5 Post, A.F., Regala, W., Shah, M., Shaw, S.L., Steglich, C., Sullivan, M.B., Ting,
- 6 C.S., Tolonen, A., Webb, E.A., Zinser, E.R. & Chisholm, S.W. (2003) Genome
- 7 divergence in two Prochlorococcus ecotypes reflects oceanic niche differentiation.
- 8 *Nature*, **424**, 1042–1047.
- 9 Rodríguez-Estévez, V., García, A. & Gómez, A.G. (2009) Characteristics of the acorns
- selected by free range Iberian pigs during the montanera season. *Livestock*
- 11 *Science*, **122**, 169–176.
- Root, R.B. (1967) The Niche Exploitation Pattern of Blue-Gray Gnatcatcher. *Ecological*
- 13 *Monographs*, **37**, 317–350.
- Roughgarden, J. (1972) Evolution of Niche Width. The American Naturalist, 106, 683–
- 15 718.
- Schoener, T. (1974) Resource Partitioning in Ecological Communities. *Science*, **185**, 27–
- 17 39.
- Silva, P.C., Cerdeira, J.O., Martins, M.J. & Monteiro-Henriques, T. (2014) Data depth for
- the uniform distribution. *Environmental and Ecological Statistics*, **21**, 27–39.
- 20 Silvertown, J. (2004) Plant coexistence and the niche. Trends in Ecology & Evolution, 19,
- 21 605–611.

- 1 Tsoar, A., Allouche, O., Steinitz, O., Rotem, D. & Kadmon, R. (2007) A comparative
- 2 evaluation of presence-only methods for modelling species distribution. *Diversity*
- 3 & Distributions, **13**, 397–405.
- 4 Tukey, J.W. (1975) Mathematics and the picturing of data. *Proceedings of the*
- 5 *International Congress of Mathematicians*, **2**, 523–531.
- 6 Villéger, S., Novack-Gottshall, P.M. & Mouillot, D. (2011) The multidimensionality of
- 7 the niche reveals functional diversity changes in benthic marine biotas across
- 8 geological time. *Ecology Letters*, **14**, 561–568.
- 9 Walker, P.A. & Cocks, K.D. (1991) HABITAT: A Procedure for Modelling a Disjoint
- 10 Environmental Envelope for a Plant or Animal Species. *Global Ecology and*
- 11 Biogeography Letters, 1, 108–118.
- Warren, D.L., Glor, R.E. & Turelli, M. (2008) Environmental Niche Equivalency Versus
- 13 Conservatism: Quantitative Approaches to Niche Evolution. *Evolution*, **62**, 2868–
- 14 2883.
- Werner, E. & Gilliam, J. (1984) The Ontogenetic Niche and Species Interactions in Size
- Structured Populations. *Annual Review of Ecology and Systematics*, **15**, 393–425.
- 17 Whittaker, R.H. & Levin, S.A. eds. (1975) *Niche: Theory and Application*, Dowen,
- Hutchinson & Ross, Pennsylvania.
- 200, Y. & Serfling, R. (2000) General notions of statistical depth function. *The Annals of*
- 20 *Statistics*, **28**, 461–482.

1	BIOSKETCH
2	
3	The authors include mathematicians, ecologists and biologists from several Universities
4	and Research Centers, involved in a three-year multidisciplinary project (S-PerforMA:
5	Species Performance Modelling Algorithm - http://www.isa.utl.pt/proj/S-PerforMa/)
6	aiming at exploring issues related with the concept of niche using combinatorial and
7	geometric tools.
8	
9	
10	

TABLES

1 2

3 **Table 1.** Overlap (wJ and cos indices) of all pairs of Quercus taxa in the four dimensional

4 space defined by the used environmental variables: isothermality, minimum temperature

5 of coldest month, annual precipitation and the precipitation of driest quarter.

6

			wJ			
	faginea broteroi	0.3388	0.3238	0.1128	0.0460	0.2190
cos	0.0349	faginea faginea	0.5419	0.2802	0.1214	0.2876
	0.0700	0.2044	rotundifolia	0.8391	0.5262	0.7920
	0.0130	0.0463	0.3041	pyrenaica	0.8874	0.8810
	0.0043	0.0084	0.0516	0.3542		0.7605
	0.0772	0.0426	0.1509	0.3141	0.2079	suber

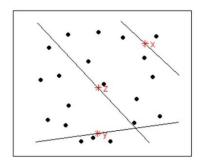


Figure 1. Illustration of the Tukey interiority of points x, y and z (represented by *), with respect to the set P of

20 points (represented by), as the number of points of P in the side of the line passing through * which has the minimum number of points. Interiorities of x, y and z are 1, 3 and 9, respectively; $d_x(P) = 1/20$, $d_y(P) = 3/20$, and $d_z(P) = 9/20$.

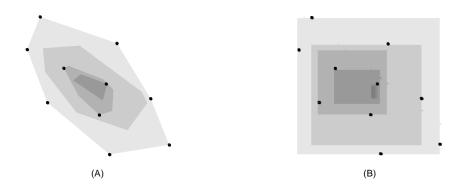


Figure 2. The 10 points represent occurrences of a species in a 2-dimensional environmental space. Shades of grey define areas of "iso-suitability". The suitability levels increase from the white (=0) to the dark grey areas. The left panel (A) refers to CH-Tukey and the right panel (B) refers to BIOCLIM.

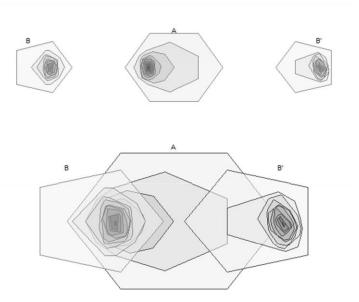


Figure 3. Niches and iso-depth regions of three species (A, B and B') in the same environmental space. Top row:

niches depicted isolated. Bottom row: overlapped niches.

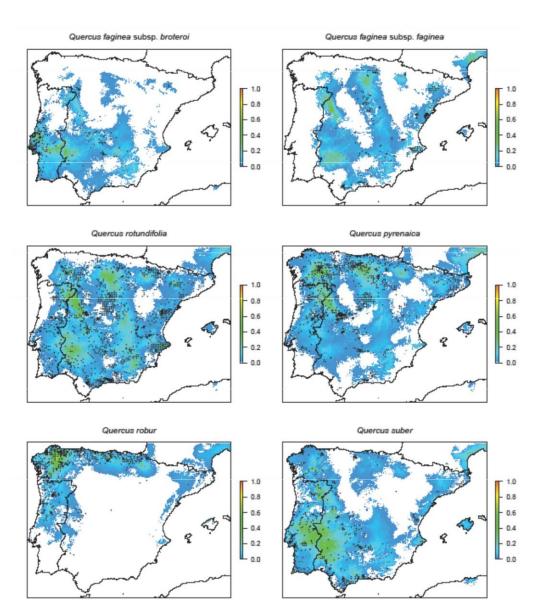


Figure 4. Environmental suitability maps for six *Quercus* taxa in the Iberian Peninsula obtained using the CH- Tukey procedure, with four climatic variables. Dots represent occurrences. The white inland areas refer to points outside the convex hull of the occurrences of the corresponding taxon, in the considered 4-dimensional environmental space.

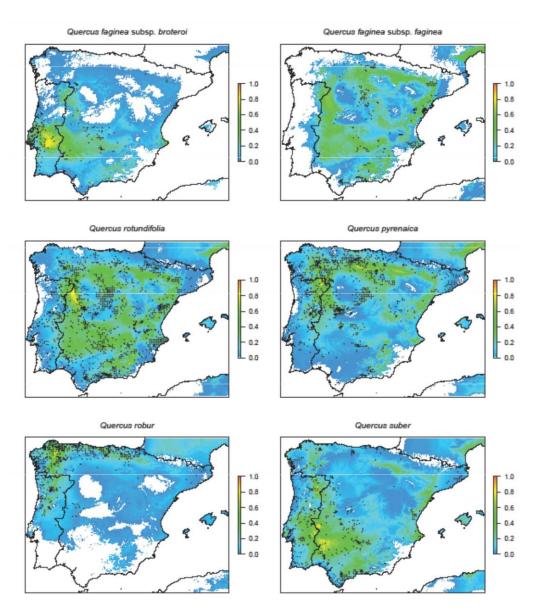


Figure 5. Environmental suitability maps for six *Quercus* taxa in the Iberian Peninsula obtained using the BIOCLIM model, with four climatic variables. Dots represent occurrences. The white inland areas refer to points outside the bounding box of the occurrences of the corresponding taxon, in the considered 4-dimensional environmental space.

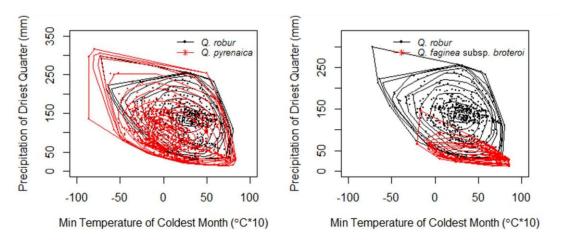


Figure 6. Occurrences and iso-depth regions of *Q. robur* and *Q. pyrenaica* in the left panel and of *Q. robur* and *Q. faginea* subsp. *broteroi* in the right panel, in the environmental space defined by the minimum temperature of the coldest month and precipitation of the driest quarter. Dots and stars correspond to occurrences, and the lines delimit iso-depth regions (regions with the same Tukey depth).

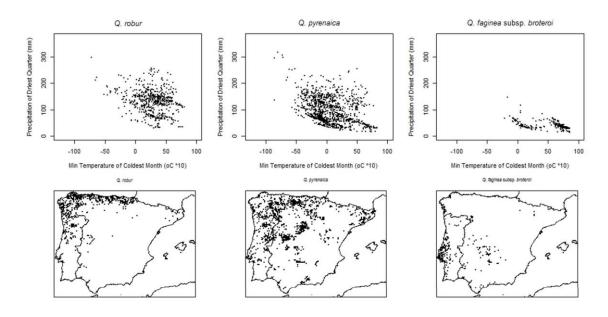


Figure 7A. Occurrences of *Q. robur*, *Q. pyrenaica* and *Q. faginea* subsp. *broteroi* in the environmental and the geographical spaces.

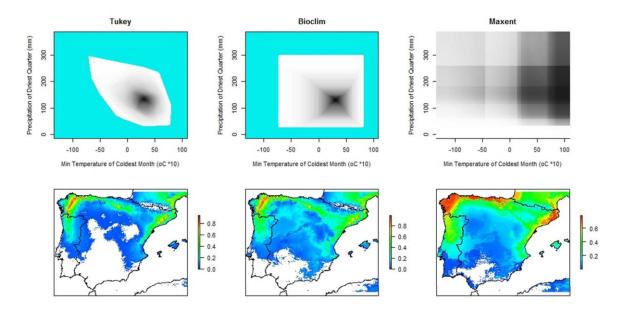


Figure 7B. Niches and suitability maps produced by CH-Tukey, BIOCLIM and MaxEnt for *Q. robur*.

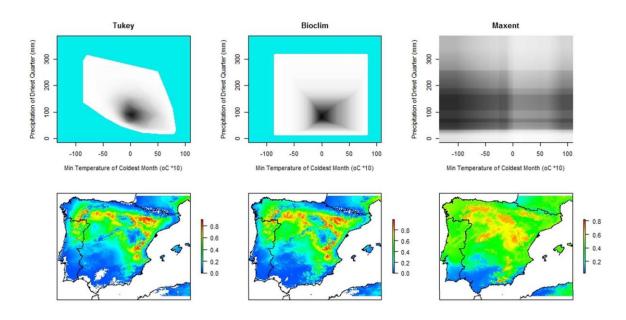


Figure 7C. Niches and suitability maps produced by CH-Tukey, BIOCLIM and MaxEnt for *Q. pyrenaica*.

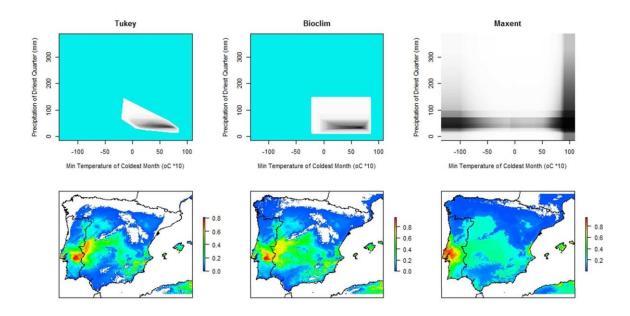


Figure 7D. Niches and suitability maps produced by CH-Tukey, BIOCLIM and MaxEnt for *Q. faginea* subsp. *broteroi*.