

Spatial mixing of mitochondrial lineages and greater genetic diversity in some invasive populations of the American mink (Neovison vison) compared to native populations

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1	Spatial mixing of mitochondrial lineages and greater genetic diversity in some invasive populations
2	of the American mink (Neovison vison) compared to native populations
3	
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16	

- 17 Abstract
- 18

19 The genetic characteristics of introduced populations have a relevant impact on their ability to 20 establish and spread. The American mink (Neovison vison), native to North America, is an important 21 invasive species in the Iberian Peninsula. Here, we used mitochondrial DNA sequences data to investigate 22 the genetic diversity and phylogeographic structure of invasive vs. native populations of this species. We 23 also evaluated whether genetic diversity in invasive populations could be explained by the genetic 24 characteristics of the native sources from which they derived. Phylogenetic analysis revealed two major 25 lineages in the native range, which indicated a clear separation between western and eastern populations. 26 On the contrary, we found no evidence of genetic structure in the invasive range. This was probably the 27 result of the diverse origins of the released specimens and the rapid expansion and encounters of the 28 introduced populations. We detected spatial mixing of both North American lineages in several sampling 29 localities of the north central area of the Iberian Peninsula, giving rise to high levels of genetic diversity 30 in some areas compared to North American populations. This could potentially lead to higher fitness of 31 these individuals and thus increase the population viability and invasiveness of this species. These results 32 point to the need to better study the populations in which lineages mix and, if necessary, intensify control 33 efforts in them. 34 35 Keywords: American mink, mitochondrial genes, genetic diversity, native range, Iberian Peninsula.

36 1. Introduction

37

38 Identifying the biological attributes of successful invaders is among the most pressing questions yet to 39 be answered (Sol et al. 2012). Although there is evidence that species differ in their invasion potential, 40 controversy exits about the characteristics that differentiate successful and unsuccessful invaders 41 (Blackburn et al. 2009). Other studies of introduced species show that genetics might play an integral role 42 in the success of an invasion, and they suggest that this role is determined by the quality of the genetic 43 variation introduced, rather than the total quantity of genetic variation (see Dlugosch et al. 2015). The 44 genetic diversity of founding populations, along with potential increase in genetic diversity from 45 subsequent introductions, may be important to the successful establishment and spread of introduced 46 species (Suarez and Tsutsui 2008; Sol et al. 2012). Nevertheless, there is no consensus on whether 47 invaders stemming from multiple native origins are more successful than those from single populations, 48 or whether demographic bottlenecks may limit the invasion success of a species (Edelaar et al. 2015; 49 Estoup 2016). 50

51 Overall, introduced populations lose genetic variation relative to their source populations (Dlugosch 52 and Parker 2008). However, numerous introduction events and a greater number of source populations 53 may buffer against such genetic losses (Kolbe et al. 2004; Dlugosch and Parker 2008; Uller and Leimu 54 2011). In fact, in some cases successful invaders show large increases in genetic diversity in the 55 introduced range, apparently caused by mixing of populations from different source regions (Genton et al. 56 2005; Novak and Mack 1993). Therefore, it is crucial to obtain information on the phylogeographic 57 structure and genetic diversity throughout the native and introduced ranges to understand the factors 58 affecting genetic diversity during invasion (e.g. multiple introduction events or admixture levels among 59 populations) and the interaction between genetic diversity and invasive potential (Edelaar et al. 2015).

60

61 The American mink (*Neovison vison*) is a successful invader with a native range restricted to North 62 America and with invasive populations distributed across Europe, Asia and South America (Bonesi and 63 Palazón 2007; Dunstone 1993). In Europe, this species was introduced for commercial fur farming during 64 the 20th century. Populations established by escaped individuals are now naturalised in fresh water and 65 coastal ecosystems (Bonesi and Palazón 2007), with detrimental consequences for the abundance of many 66 native species of prey and competitors of conservation importance (Bonesi et al. 2004; Harrington et al. 67 2009; Macdonald and Harrington 2003; Melero et al. 2012a), including critically endangered species such 68 as the European mink (Mustela lutreola) (Maran et al. 2016) and the Pyrenean desman (Galemys 69 pyrenaicus) (Fernandes et al. 2008). Fur farms were established in many countries and harboured mink 70 stocks from different parts of their native range. Because of the large scale and long-term nature of this 71 farming, the number of escapees, which started to breed in the wild and became the founder individuals 72 of feral populations, was very high (Bonesi and Palazón 2007). In Spain, mink farming started at the end 73 of the 1950s and the number of farms reached a peak in the 1980s with about 400 farms concentrated 74 mainly in the North West (Galicia) (Ruiz-Olmo et al. 1997). Today, feral mink populations are widely 75 distributed across almost all of the north and centre of the Iberian Peninsula. During the last few decades, 76 these populations experienced a rapid expansion and increased the number of nuclei in almost every

colonised area (Ruiz-Olmo et al. 1997; Santulli 2014). Although the number of mink farms in Spain was
considerably reduced at the end of the 1980s, mink escapes from the last active farms are still ongoing
(Bonesi and Palazón 2007).

80

92

81 In Europe, there is genetic information on feral American mink populations from several countries. 82 Previous studies in Spain using microsatellite data determined that American mink populations from 83 disjunct geographical areas appeared differentiated and not yet connected by gene flow (Lecis et al. 84 2008). In Scotland, Zalewski et al. (2009) and Fraser et al. (2013) found that landscape features restricted 85 gene flow within this species. Other studies in Poland found that multiple introductions determine the 86 genetic structure of American mink populations (Zalewski et al. 2010) and that these populations exhibit 87 high mitochondrial genetic diversity (Zalewski et al. 2011). A study in France using microsatellite data 88 suggested recent admixture between populations that had recently encountered each other (Bifolchi et al. 89 2010). Similarly, in Sweden it was found, also using microsatellite data, that genetic diversity decreased 90 over six consecutive years of mink culling as part of a population control programme (Zalewski et al. 91 2016).

93 Despite the importance of understanding the genetic diversity within the native populations for 94 investigating the evolutionary dynamics of invasions, no phylogeographic study has been performed so 95 far in the North American native range of the American mink. Therefore, many genetic studies performed 96 thus far are difficult to interpret. We used mitochondrial DNA (mtDNA) sequence data from populations 97 in the native range to elucidate whether invasive populations in the Iberian Peninsula have single or 98 multiple origins from the North American sources. We aimed to determine the phylogeographic structure 99 of native and invasive populations, and to evaluate whether mitochondrial genetic diversity observed in 100 established invasive populations can be explained by genetic characteristics of native populations. We 101 examined whether the genetic variability of the Iberian population represented only a small part of that in 102 its native range, as expected for many invasive species or, due to the introduction history and particular 103 origin of these populations, this genetic variability was larger. Overall, the study aims to provide 104 recommendations for mink control in the Iberian Peninsula with respect to determining key control areas. 105

- 106 2. Materials and methods
- 107

108 2.1 Sample collection

109

We used hair and tissue samples from invasive populations provided by different Autonomous
Communities of Spain. These samples were collected during population control operations conducted by
the local administrations between 2007 and 2016 as part of eradication programs and were conserved in
tubes containing absolute ethanol (ESM Appendix: Table S1). A total of 105 samples from Catalonia
(16), La Rioja (15), Galicia (12), Basque Country (20), Navarra (11), Castellon (7) and Castilla-Leon (24)

- 115 were included in the study.
- 116

117	Moreover, tissue samples of 21 specimens from North America (ESM Appendix: Table S2) were
118	obtained from different biological collections: University of Alaska Museum (12), Museum of
119	Southwester Biology (5) and Cornell University Museum of Vertebrates (4).
120	
121	2.2 DNA extraction, PCR and sequencing of mitochondrial DNA
122	
123	We extracted DNA using the QIAGEN DNeasy Blood and Tissue Kit, following the manufacturer's
124	instructions, in a final elution volume of 100 μ l. These extractions were carried out in a separated UV-
125	irradiated area with dedicated equipment. Samples were incubated in a lysis buffer with proteinase K at
126	56°C overnight to ensure maximum tissue lysis.
127	
128	For each sample, we amplified a 573 bp fragment of the 5' distal part of the mitochondrial DNA
129	control region (D-loop). For this purpose, we designed novel primers using complete mitochondrial
130	sequences of several mustelids: Neovison_tRNA_pro (5'-ATC AGC ACC CAA AGC TGA CAT TCT
131	A-3') and Neovison_Dloop_556r (5'-TGT GTA TGT CCT GTG ACC ATT GAC T-3'). PCR reactions
132	were set up in a dedicated PCR clean room that is physically separated from post-PCR working areas and
133	regularly decontaminated by UV-irradiation. PCR reactions were performed in a final volume of 25 μ l,
134	containing 2–4 μ l of genomic DNA, 1 μ M of each primer, 0.2 mM dNTPs, 0.75 units of Promega GoTaq
135	DNA polymerase and 17.5 μ M of bovine serum albumin under the following conditions: an initial
136	denaturation of 2 min at 95°C, followed by 35 cycles of denaturation (30 s at 95°C), annealing (30 s at
137	54°C) and extension (30 s at 72°C). PCR products were visualised by electrophoresis in a 1% agarose
138	SYBR-Safe (Invitrogen) stained gel.
139	
140	PCR products were purified using ExoSAP-It (Affymetrix), and both strands were sequenced using
141	the original PCR primers at Macrogen Europe (Amsterdam, The Netherlands). Sequences were inspected,
142	trimmed and assembled using Geneious Pro (Biomatters Ltd.). All sequences obtained in this study were
143	deposited in European Nucleotide Archive/GenBank under accession numbers LT854061-LT854186.
144	
145	2.3 Phylogenetic and genetic diversity analyses
146	
147	Consensus sequences were aligned using MAFFT version 7.130 (Katoh et al. 2002) with default
148	settings. Due to the presence of several gaps and ambiguous positions in a very specific region of the
149	alignment, we manually removed positions 92 to 110, thus leaving 554 bp in the alignment.
150	
151	A maximum likelihood phylogenetic tree was calculated with RAxML version 8.0 (Stamatakis 2014)
152	using a general time reversible substitution model and rate heterogeneity modelled with a gamma
153	distribution plus a proportion of invariable sites, as recommended by the program. From this tree, a
154	haplotype genealogy was generated using Haploviewer 1.0 (Salzburger et al. 2011).
155	
156	The nucleotide diversity (π) was estimated using the R package PEGAS (Population and Evolutionary

157	Genetics Analysis System, Paradis 2010). To analyse the variation in nucleotide diversity across the
158	Iberian Peninsula, we used a visualization method that does not require the pre-definition of populations.
159	The nucleotide diversity was estimated at each sampling location by using all samples present within 30
160	km of the location. This allowed the estimation of genetic diversity from a sufficient number of samples
161	at each point, yet the resolution was sufficient to distinguish regional differences in genetic diversity.
162	QGIS (http://qgis.osgeo.org) was used to plot the nucleotide diversity in the map. In the case of the North
163	American native range, it was not possible to apply this analysis due to the limited number of samples.
164	Therefore, we selected three areas with a sufficient number of samples: North West (British Columbia,
165	six samples), South East (Texas, three samples) and East (New York, four samples), in order to estimate
166	the nucleotide diversity in these specific areas.
167	
168	3. Results
169	
170	3.1 Mitochondrial phylogeographic analysis
171	
172	A total of 16 different haplotypes of the mitochondrial control region were detected in 126 American
173	mink, 11 of which were recorded in North America and 7 in the Iberian Peninsula (ESM Appendix: Fig.
174	S1). The maximum likelihood phylogenetic tree showed two large clades found throughout western and
175	eastern North America, respectively (Fig. 1). The haplotype genealogy reconstructed from the maximum
176	likelihood tree revealed the same two large groups (Fig. 2a). As indicated in this genealogy, two
177	mutations unambiguously differentiate both clades. The map of samples clearly shows the geographical
178	separation of the two clades in North America (Fig. 2b). In the Iberian Peninsula, we found more samples
179	belonging to the western than to the eastern clade of North America (68 vs. 37 samples, respectively).
180	However, samples belonging to both clades were detected together in four regions of the north central
181	area of the Iberian Peninsula: Basque Country, La Rioja, Navarra and Avila (South of Castilla-Leon) (Fig.
182	2c).
183	
184	In addition, the phylogeographic structure was also different at a finer scale in the native and invasive
185	ranges. While there was evidence of a geographical pattern for the individuals collected from the native
186	range, with individual haplotypes concentrated in a single locality or close localities (Fig. 1), many
187	haplotypes were widely dispersed in the Iberian Peninsula (Fig. 1, Fig. S1), showing that haplotypes are
188	not informative with regards to the geographical origin of any sample.
189	
190	3.2 Mitochondrial genetic diversity
191	
192	The global genetic diversity of the American mink estimated with the control region was relatively
193	low for all samples, with a value of nucleotide diversity of $\pi = 0.68\%$, whereas the nucleotide diversity
194	for North America and the Iberian Peninsula was 0.78% and 0.64%, respectively (haplotype diversity was
195	0.81, 0.92 and 0.76, respectively). However, the nucleotide diversity varied considerably when different
196	localities and populations were considered separately. Among the three areas of the native range in which

- 197 π was calculated separately, the population of British Columbia showed the highest nucleotide diversity
- 198 (0.53%) whereas New York (0.12%) and Texas (0%) registered lower values. In agreement with the
- spatial mixing of clades in the central areas of the Iberian Peninsula, the map derived from the genetic
- 200 diversity values of the samples around each locality clearly showed maximum levels of genetic diversity
- 201 in these regions (Fig. 3). When values of several sampling localities were averaged, the Basque
- 202 population concentrated the highest nucleotide diversity (0.61%), followed by Avila (0.50%) and La
- 203 Rioja (0.25%). From these areas, genetic diversity decreased towards the remainder of the mink
- distribution, reaching values of 0% in some areas.
- 205

206 4. Discussion

207

208 Our results show that the establishment of the invasive American mink population in the Iberian 209 Peninsula is the product of the introduction of individuals from the two main genetic lineages found in 210 western and eastern North America. In the invasive range, these lineages registered a complete lack of 211 phylogeographic structure. This was probably the result of the additive effect of multiple continuous 212 introductions, and the rapid expansion and encounters of the introduced populations. Overall, genetic 213 diversity was lower in the introduced than in the native populations, although in some introduced 214 populations diversity was higher to any one native population, at least with respect to the reduced native 215 area analysed so far. These results provide evidence for the potential of introduced populations to 216 maintain or increase the genetic diversity.

217

218 4.1 Phylogeographic patterns

219

220 Phylogenetic analysis revealed two large native phylogeographic lineages, which indicated a clear 221 separation between the western and eastern populations in North America. Therefore, the relationship 222 between the mtDNA haplotypes and the location from which they were sampled provides evidence of 223 phylogeographic structure among populations of American mink in their native range. Unfortunately, the 224 lack of information about sequence data of mtDNA of the American mink in North America did not allow 225 our results to be compared with those of similar studies. Other studies of plant and animal species from 226 North America have illustrated a complex phylogeographic history of the biota of this region (Shafer et 227 al. 2010; Soltis et al. 2006). The data currently available suggest that vicariance, dispersal and the 228 existence of refugia have worked in concert to produce the species' distribution and genetic divergence 229 patterns of the region (Brunsfeld et al. 2001). Future research focused on filling the gaps by sampling at 230 locations between the two clades seen in this work is required to achieve a deeper understanding of the 231 evolutionary history of the American mink in its native range.

232

The pattern of spatial structure recorded in North America allowed for a more precise identification of the origin of invasive populations in the Iberian Peninsula. Our analysis revealed that there were introductions from both western and eastern regions in North America. Additionally, the distribution of the invasive haplotypes in the Iberian Peninsula suggests that at least the north central area of the Iberian

- Peninsula (populations of the Basque Country, La Rioja, Navarra and Avila) experienced multiple
 introductions from both regions of North America, while the western and eastern Iberian populations
 (Galicia, Catalonia and Castellon) originated from introductions from western North America, contrary to
 what was previously believed (Palazón et al. 1997). This situation most likely occurred as a result of the
 multiple origins of mink from different farms and the fast range expansion of the occupied area since
 their introduction in the 1960s-1970s. It is also likely that farms received founder individuals from
- 243 244

245 The occurrence of the two main mitochondrial lineages in the north central area of the Iberian 246 Peninsula suggests that the intermixing of individuals of both lineages may potentially take place. 247 However, it will be necessary to perform analysis with multiple genomic markers in the future to 248 determine whether individuals belonging to these two lineages effectively interbreed and, if so, to 249 estimate admixture levels in these populations (Pritchard et al. 2000). It is also important to note that, due 250 to the limited information derived from the mitochondrial DNA, which only reflects the maternal lineage, 251 the existence of some admixture levels in other Iberian populations cannot be discarded until genomic 252 analyses are performed. In fact, a previous study in the Iberian Peninsula based on microsatellites showed 253 that the highest admixture levels between populations were found in the Basque Country population, 254 which in principle is consistent with our results (Lecis et al. 2008). However, since no native specimens 255 were available in that study, it was not possible to determine how the identified populations were related 256 to the source populations. This confirms that it is crucial to obtain samples from the native range to shed 257 light on the relationship between invasive and native populations and to identify potential source 258 populations.

259

261

260 4.2 Genetic diversity in the native versus invasive range

different source populations in the native range.

262 Based on mitochondrial DNA control region sequence data, our results provide evidence that the 263 introduced American mink populations in the Iberian Peninsula have overall less genetic diversity than 264 native mink populations in North America. These results are in agreement a priori with the "genetic 265 paradox" (Allendorf and Lundquist 2003; Estoup et al. 2016), which predicts that low genetic variability 266 can be expected in invasive species because of the loss of rare alleles, as a result of the founder effect and 267 population bottlenecks (Allendorf and Lundquist 2003; Grapputo et al. 2005). Nevertheless, when 268 analysed at the population level within each range, nucleotide diversity in the population of the Basque 269 Country and La Rioja in the Iberian Peninsula was found to be greater than that recorded for North 270 American populations. The pattern registered here is consistent with other studies (e.g. Edelaar et al. 271 2015; Genton et al. 2005; Roman and Darling 2007), providing another example of an invasive species 272 retaining genetic diversity. Similarly, previous studies have demonstrated that the genetic diversity of 273 invasive populations in their new range may be increased through the admixture of lineages from multiple 274 native populations (Keller and Taylor 2010; Kolbe et al. 2008). As already indicated, the pattern observed 275 in the Iberian Peninsula could be the result of multiple introductions of mink deriving from different local 276 lineages of North America, their rapid expansion, and the encounter of lineages in the north central area

of the Iberian Peninsula that may potentially hybridise. However, we cannot discard that cross-breeding
of the different North American subspecies or populations took place within farms (Belliveau et al. 1999;
Dunstone 1993).

281 High values of genetic diversity have also been registered in some mink population of Poland 282 (Zalewski et al. 2011). In this case, mean π for all feral mink analysed was 0.94%, with values for 283 different populations ranging between 0 and 1.41%. Although the mitochondrial control region fragment 284 used in that work was different to ours and a direct comparison cannot be performed, the values found in 285 some populations in Poland were much higher than in others. It is therefore likely that the same processes 286 suggested for the Iberian Peninsula (i.e. multiple introductions, rapid expansion and encounters of 287 individuals of different lineages) can be applied to other areas with invasive populations of American 288 mink.

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280

290 4.3 Implications for management

292 This study provides an understanding of the genetic structure and diversity of the populations of the 293 American mink in the Iberian Peninsula, which can be used to identify key control areas. Thus, one way 294 to improve the effectiveness of management programmes of an invasive species such as the American 295 mink is by incorporating molecular-based information into these programs (Hampton et al. 2004). In this 296 regard, the current spatial mixing of individuals from two different native lineages in the north central 297 area of the Iberian Peninsula (Basque Country, La Rioja, Navarra and Avila populations) and the 298 potential interbreeding between such lineages that may take place could translate, among other outcomes, 299 into a higher fitness in these individuals (e.g. Vandewoestijne et al. 2008). In fact, previous studies have 300 found a positive correlation between genetic diversity and colonization success as measured by the 301 reproduction rate (e.g. Crawford and Whitney 2010). Interestingly, Melero et al. (2012b) found that mink 302 of the Basque Country were bigger in size than mink of the rest of the Iberian Peninsula. According to 303 Melero et al. (2012b), this phenotypic variation could be due to a mixing of individuals from different 304 origins, among other possibilities. If the link between spatial mixing of mitochondrial lineages, genome 305 admixture and higher fitness is demonstrated for some populations, then higher efforts to control and 306 eradicate the American mink should be undertaken in such populations in order to impede or minimise 307 their expansion.

308

Current controls of the American mink in the Iberian Peninsula focus on reducing their effects on ecosystems and biodiversity (i.e., areas with local populations of threatened species, wetlands where they can cause high mortalities in waterfowl and protected natural areas, etc.). In this regard, it is important to consider that the areas of the Iberian Peninsula with presence of both American mink lineages coincide with critical areas for the conservation of two of the most threatened mammals of the Iberian Peninsula, the European mink and the Pyrenean desman. The American mink compete successfully for habitat and

- 315 prey with the European mink (Sidorovich et al. 1999; Sidorovich and MacDonald 2001), whose
- 316 population of the North of Spain and South West of France is highly isolated from the rest of the

distribution (Michaux et al. 2005). On the other hand, the populations of Pyrenean desman of the Central
System have experienced a strong reduction in the last few years and are threatened with extinction.
Although the reasons for this decline are not well known, predation by the American mink may have
contributed to it (Fernandes et al. 2008). Thus, the presence of the two native lineages of the American
mink in the areas with these threatened species should be taken into account in future conservation plans
due the putative higher expansive potential of the American mink in them.

323

324 The results of this work are in line with previous studies suggesting that mink control should also be 325 directed to minimise the contact between genetically differentiated populations (Bifolchi et al. 2010; 326 Fraser et al. 2013; Lecis et al. 2008; Robertson and Gemmell 2004), and so counteract a possible increase 327 in mink fecundity and dispersal (Melero et al. 2015; Oliver et al. 2016). The present study has shown how 328 molecular-based information obtained from the invasive range, together with that from the native range, 329 was critical to characterise the sources from which the Iberian populations of the American mink 330 originated. Further studies are now necessary to obtain more detailed information, particularly in areas 331 where admixture of lineages may potentially occur, to help prioritise efforts to control and manage this 332 invasive species.

333

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335

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349 6. References

- 350
- Allendorf FW, Lundquist LL (2003) Population Biology, Evolution, and Control of Invasive Species.
 Conserv Biol 17:24–30. doi: 10.1046/j.1523-1739.2003.02365.x
- Belliveau AM, Farid A, O'Connell M, Wright JM (1999) Assessment of genetic variability in captive and
 wild American mink (*Mustela vison*) using microsatellite markers. Can J Anim Sci 79:7–16. doi:
 10.4141/A97-115
- 356 Bifolchi A, Picard D, Lemaire C, Cormier JP, Pagano A (2010) Evidence of admixture between

357 differentiated genetic pools at a regional scale in an invasive carnivore. Conserv Genet 11:1-9. doi: 358 10.1007/s10592-008-9780-1 359 Blackburn TM, Cassey P, Lockwood JL (2009) The role of species traits in the establishment success of 360 exotic birds. Glob Chang Biol 15:2852-2860. doi: 10.1111/j.1365-2486.2008.01841.x 361 Bonesi L, Chanin P, Macdonald D (2004) Competition between Eurasian otter Lutra lutra and American 362 mink Mustela vison probed by niche shift. Oikos 106:19-26. doi: 10.1111/j.0030-363 1299.2004.12763.x 364 Bonesi L, Palazón S (2007) The American mink in Europe: Status, impacts, and control. Biol Conserv 365 134:470-483. doi: 10.1016/j.biocon.2006.09.006 366 Brunsfeld SJ, Sullivan J, Soltis DE, Soltis PS (2001) Comparative phylogeography of north- western 367 North America: a synthesis. In: in J. Silvertown and J. Antonovics [Eds.], Integrating ecology and 368 evolution in a spatial context, Blackwell Science, Oxford, UK. pp 319 - 339 369 Crawford KM, Whitney KD (2010) Population genetic diversity influences colonization success. Mol 370 Ecol 19:1253-1263. doi: 10.1111/j.1365-294X.2010.04550.x 371 Dlugosch KM, Parker IM (2008) Founding events in species invasions: Genetic variation, adaptive 372 evolution, and the role of multiple introductions. Mol Ecol 17:431-449. doi: 10.1111/j.1365-373 294X.2007.03538.x 374 Dlugosch KM, Anderson SR, Braasch J, Cang FA, Gillette HD (2015) The devil is in the details: genetic 375 variation in introduced populations and its contributions to invasion. Mol Ecol 24:2095-2111. doi: 376 10.1111/mec.13183 377 Dunstone N (1993). The Mink. T&AD Poyser Ltd., London 378 Edelaar P, Roques S, Hobson EA, Gonçalves Da Silva A, Avery ML, Russello MA, Senar JC, Wright TF, 379 Carrete M, Tella JL (2015) Shared genetic diversity across the global invasive range of the monk 380 parakeet suggests a common restricted geographic origin and the possibility of convergent 381 selection. Mol Ecol 24:2164–2176. doi: 10.1111/mec.13157 382 Estoup A, Ravign V, Hufbauer R, Vitalis R, Gautier M, Facon B (2016) Is There A Genetic Paradox of 383 Biological Invasion? Annu Rev Ecol Evol Syst 47:51-72. doi: 10.1146/annurev-ecolsys-121415 384 Fernandes M, Herrero J, Aulagnier S, Amori G (2008) Galemys pyrenaicus. The IUCN Red List of 385 Threatened Species: e.T8826A12934876. 386 http://dx.doi.org/10.2305/IUCN.UK.2008.RLTS.T8826A12934876.en 387 Fraser EJ, Macdonald DW, Oliver MK, Piertney S, Lambin X (2013) Using population genetic structure 388 of an invasive mammal to target control efforts - An example of the American mink in Scotland. 389 Biol Conserv 167:35-42. doi: 10.1016/j.biocon.2013.07.011 390 Genton BJ, Shykoff JA, Giraud T (2005) High genetic diversity in French invasive populations of 391 common ragweed, Ambrosia artemisiifolia, as a result of multiple sources of introduction. Mol Ecol 392 14:4275-4285. doi: 10.1111/j.1365-294X.2005.02750.x 393 Grapputo A, Boman S, Lindström L, Lyytinen A, Mappes J (2005) The voyage of an invasive species 394 across continents: Genetic diversity of North American and European Colorado potato beetle 395 populations. Mol Ecol 14:4207-4219. doi: 10.1111/j.1365-294X.2005.02740.x 396 Hampton JO, Spencer PBS, Alpers DL, Twigg LE, Woolnough AP, Doust J, Higgs T, Pluske J (2004)

397	Molecular techniques, wildlife management and the importance of genetic population structure and
398	dispersal: a case study with feral pigs. J Appl Ecol 41:735-743. doi: 10.1111/j.0021-
399	8901.2004.00936.x
400	Harrington LA, Harrington AL, Moorhouse T, Gelling M, Bonesi L, Macdonald DW (2009) American
401	mink control on inland rivers in southern England: An experimental test of a model strategy. Biol
402	Conserv 142:839-849. doi: 10.1016/j.biocon.2008.12.012
403	Katoh K, Misawa K, Kuma K, Miyata T (2002) MAFFT: a novel method for rapid multiple sequence
404	alignment based on fast Fourier transform. Nucleic Acids Res 30:3059-3066. doi:
405	10.1093/nar/gkf436
406	Keller SR, Taylor DR (2010) Genomic admixture increases fitness during a biological invasion. J Evol
407	Biol 23:1720–1731. doi: 10.1111/j.1420-9101.2010.02037.x
408	Kolbe JJ, Glor RE, Rodríguez Schettino L, Lara AC, Larson A, Losos JB (2004) Genetic variation
409	increases during biological invasion by a Cuban lizard. Nature 431:177–181. doi:
410	10.1038/nature02807
411	Kolbe JJJ, Larson A, Losos JBB, de Queiroz K (2008) Admixture determines genetic diversity and
412	population differentiation in the biological invasion of a lizard species. Biol Lett 4:434-437. doi:
413	10.1098/rsbl.2008.0205
414	Lecis R, Ferrando A, Ruiz-Olmo J, Mañas S, Domingo-Roura X (2008) Population genetic structure and
415	distribution of introduced American mink (Mustela vison) in Spain, based on microsatellite
416	variation. Conserv Genet 9:1149-1161. doi: 10.1007/s10592-007-9428-6
417	Macdonald DW, Harrington LA (2003) The American mink: The triumph and tragedy of adaptation out
418	of context. N.Z. J. Zool. 30(4):421-441.doi: http://dx.doi.org/10.1080/03014223.2003.9518350
419	Maran T, Skumatov D, Gomez A, Põdra M, Abramov AV, Dinets V (2016) Mustela lutreola. The IUCN
420	Red List of Threatened Species 2016: e.T14018A45199861.
421	http://dx.doi.org/10.2305/IUCN.UK.2016-1.RLTS.T14018A45199861.en
422	Melero Y, Plaza M, Santulli G, Saavedra D, Gosàlbez J, Ruiz-Olmo J, Palazón S (2012a) Evaluating the
423	effect of American mink, an alien invasive species, on the abundance of a native community: is
424	coexistence possible? Biodivers Conserv 21:1795-1809. doi: 10.1007/s10531-012-0277-3
425	Melero Y, Santulli G, Gómez A, Gosàlbez J, Rodriguez-Refojos C, Palazón S (2012b) Morphological
426	variation of introduced species: The case of American mink (Neovison vison) in Spain. Mamm Biol
427	77:345–350. doi: 10.1016/j.mambio.2012.02.001
428	Melero Y, Robinson E, Lambin X (2015) Density- and age-dependent reproduction partially compensates
429	culling efforts of invasive non-native American mink. Biol Invasions 17(9): 2645-2657. doi:
430	10.1007/s10530-015-0902-7
431	Michaux JR, Hardy OJ, Justy F, Fournier P, Kranz A, Cabria M, Davison A, Rosoux R, Libois R (2005)
432	Conservation genetics and population history of the threatened European mink Mustela lutreola,
433	with an emphasis on the west European population. Mol Ecol 14:2373-2388. doi: 10.1111/j.1365-
434	294X.2005.02597.x
435	Novak S, Mack R (1993) Genetic variation in Bromus tectorum (Poaceae): comparison between native
436	and introduced populations. Heredity (Edinb) 71:167-176. doi: 10.1038/hdy.1993.121

437 Oliver MK, Piertney SB, Zalewski A, Melero Y, Lambin X (2016) The compensatory potential of 438 increased immigration following intensive American mink population control is diluted by male-439 biased dispersal. Biol Invasions 18:3047-3061. doi: 10.1007/s10530-016-1199-x 440 Palazón S, Ruiz-Olmo J, Bueno F, Bueno F, Jordan G, Palomero G, Palomero G, Munilla I, Romero R, 441 Gimenez YJ (1997) El visón americano en España. In: El visón europeo y el visón americano en 442 España. (Eds.) Palazón, S. and Ruiz-Olmo, J. Colección Técnica, Ministerio de Medio Ambiente. 443 Madrid. pp 107-114 444 Paradis E (2010) PEGAS: An R package for population genetics with an integrated-modular approach. 445 Bioinformatics 26:419-420. doi: 10.1093/bioinformatics/btp696 446 Pritchard JK, Stephens M, Donnelly P (2000) Inference of population structure using multilocus genotype 447 data. Genetics 155:945-959. doi: 10.1111/j.1471-8286.2007.01758.x 448 Reid F, Schiaffini M, Schipper J (2016) Neovison vison. The IUCN Red List of Threatened Species 2016: 449 e.T41661A45214988. http://dx.doi.org/10.2305/IUCN.UK.2016-1.RLTS.T41661A45214988.en. 450 Robertson BC, Gemmell NJ (2004) Defining eradication units to control invasive pests. J Appl Ecol 451 41:1042-1048. doi: 10.1111/j.0021-8901.2004.00984.x 452 Roman J, Darling JA (2007) Paradox lost: genetic diversity and the success of aquatic invasions. Trends 453 Ecol Evol 22:454-464. doi: 10.1016/j.tree.2007.07.002 454 Ruiz-Olmo J, Munilla I, Romero R (1997) Distribution, Status and Colonization of the American Mink 455 Mustela vison in Spain. J Wildl Res 2:30-36. 456 Salzburger W, Ewing GB, Von Haeseler A (2011) The performance of phylogenetic algorithms in 457 estimating haplotype genealogies with migration. Mol Ecol 20:1952-1963. doi: 10.1111/j.1365-458 294X.2011.05066.x 459 Santulli G (2014) Occupancy, abundance, potential distribucion and spatial competition of the critically 460 endangered european mink (Mustela lutreola) and the invasive non-native american mink 461 (Neovison vison) in the Iberian Peninsula. Dissertation, Universitat de Barcelona. 462 Shafer ABA, Cullingham CI, Côté SD, Coltman DW (2010) Of glaciers and refugia: a decade of study 463 sheds new light on the phylogeography of northwestern North America. Mol Ecol 19:4589-4621. 464 doi: 10.1111/j.1365-294X.2010.04828.x 465 Sidorovich V, Kruuk H, Macdonald D (1999) Body size, and interactions between European and 466 American mink (Mustela lutreola and M. vison) in Eastern Europe. J Zool 248:521-527. doi: 467 10.1111/j.1469-7998.1999.tb01051.x 468 Sidorovich V, MacDonald DW (2001) Density dynamics and changes in habitat use by the European 469 mink and other mustelids in connection with the American mink expansion in Belarus. Netherlands 470 J Zool 51:107-126. doi: 10.1163/156854201750210878 471 Sol D, Maspons J, Vall-Llosera M, Bartomeus I, García-Peña GE, Piñol J, Freckleton RP (2012) 472 Unraveling the life history of successful invaders. Science 337:580-583. doi: 473 10.1126/science.1221523 474 Soltis DE, Morris AB, McLachlan JS, Manos PS, Soltis PS (2006) Comparative phylogeography of 475 unglaciated eastern North America. Mol Ecol 15(14):4261-4293. doi:10.1111/j.1365-476 294X.2006.03061.x

- 477 Stamatakis A (2014) RAxML version 8: A tool for phylogenetic analysis and post-analysis of large
 478 phylogenies. Bioinformatics 30:1312–1313. doi: 10.1093/bioinformatics/btu033
- 479 Suarez A V., Tsutsui ND (2008) The evolutionary consequences of biological invasions. Mol Ecol
 480 17:351–360. doi: 10.1111/j.1365-294X.2007.03456.x
- 481 Uller T, Leimu R (2011) Founder events predict changes in genetic diversity during human-mediated
 482 range expansions. Glob Chang Biol 17:3478–3485. doi: 10.1111/j.1365-2486.2011.02509.x
- Vandewoestijne S, Schtickzelle N, Baguette M (2008) Positive correlation between genetic diversity and
 fitness in a large, well-connected metapopulation. BMC Biol 6:46–55. doi: 10.1186/1741-7007-6485
 46
- Zalewski A, Piertney SB, Zalewska H, Lambin X (2009) Landscape barriers reduce gene flow in an
 invasive carnivore: geographical and local genetic structure of American mink in Scotland. Mol
 Ecol 18:1601–1615. doi: 10.1111/j.1365-294X.2009.04131.x
- Zalewski A, Michalska-Parda A, Bartoszewicz M, Kozakiewicz M, Brzeziński M (2010) Multiple
 introductions determine the genetic structure of an invasive species population: American mink *Neovison vison* in Poland. Biol Conserv 143:1355–1363. doi: 10.1016/j.biocon.2010.03.009
- Zalewski A, Michalska-Parda A, Ratkiewicz M, Kozakiewicz M, Bartoszewicz M, Brzeziński M (2011)
 High mitochondrial DNA diversity of an introduced alien carnivore: Comparison of feral and ranch
 American mink *Neovison vison* in Poland. Divers Distrib 17:757–768. doi: 10.1111/j.1472495 4642.2011.00767.x
- Zalewski A, Zalewska H, Lunneryd SG, André C, Mikusiński G (2016) Reduced genetic diversity and
 increased structure in American mink on the Swedish coast following invasive species control.
 PLoS One 11(6):e0157972. doi: 10.1371/journal.pone.0157972
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509	Fig. 1 Maximum-likelihood phylogenetic tree derived from the partial mitochondrial control region
510	sequences for 126 American mink, with eastern and western clades indicated. The number of individuals
511	per locality is shown in brackets. Sequences from the native range are underlined. The scale indicates
512	number of substitutions per position.
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515	Fig. 2 Phylogeographic analysis of the mitochondrial control region sequences of <i>Neovison vison</i> . (A)
516	Haplotype genealogy based on the maximum-likelihood tree. Circle sizes are proportional to the
517	haplotype frequency and small filled circles are inferred haplotypes not present in the sample. (B, C) Map
518	of native (B) and invasive (C) ranges showing the 126 samples of N. vison used in this study. Red and
519	blue colours represent the eastern and western clades, respectively. Grey areas represent the original
520	distribution of N. vison downloaded from the IUCN Red List of Threatened Species website (Reid et al.
521	2016) (B) and the occupancy area of mink in the Peninsula (MAGRAMA, 2013) (C). AL: Alaska, BC:
522	British Columbia, MT: Montana, WY: Wyoming, TX: Texas, NY: New York, GA: Galicia, CL: Castilla-
523	Leon, AV: Avila, LR: La Rioja, BC: Basque Country, NAV: Navarra, CAS: Castellon, CAT: Catalonia
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526	Fig. 3 Map plotting genetic diversity (π) in different sampling localities of <i>Neovison vison</i> in the Iberian
527	Peninsula. Colours of the sample points indicate the nucleotide diversity according to the provided scale.
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