

1

2

3

4

5

6

7

8 Artificial light pollution influences behavioral and
9 physiological traits in a keystone predator species,
10 *Concholepas concholepas*

11

12 Patricio H. Manríquez^{1,2*}, María Elisa Jara^{1,2}, María Isabel Diaz^{1,2}, Pedro A. Quijón³,
13 Stephen Widdicombe⁴, José Pulgar⁵, Karen Manríquez⁵, Diego Quintanilla-Ahumada⁵ &
14 Cristian Duarte⁵

15

16 ¹Centro de Estudios Avanzados en Zonas Áridas (CEAZA), Coquimbo, Chile. ²Laboratorio
17 de Ecología y Conducta de la Ontogenia Temprana (LECOT), Coquimbo, Chile.

18 ³Department of Biology, University of Prince Edward Island, Charlottetown, PE, Canada.

19 ⁴Plymouth Marine Laboratory, Prospect Place, West Hoe, Plymouth PL1 3DH, UK,

20 ⁵Departamento de Ecología y Biodiversidad, Facultad de Ecología y Recursos de la Vida,
21 Universidad Andrés Bello, Santiago, Chile.

22

23

24 *Corresponding author: patriciohmanriquez@gmail.com

25

26

27

28

29

30

31 **Abstract.** Artificial Light At Night (ALAN) is an increasing global problem that, despite
32 being widely recognized in terrestrial systems, has~~ve~~ been studied much less in marine
33 habitats. In this study we investigated the effect of ALAN on behavioral and physiological
34 traits of *Concholepas concholepas*, an important keystone species of the south-eastern
35 Pacific coast. We used juveniles collected in intertidal habitats that had not previously been
36 exposed to ALAN. In the laboratory we exposed them to two treatments: darkness and
37 white LED (Lighting Emitting Diodes) to test for the impacts of ALAN on prey-searching
38 behavior, self-righting time and metabolism. In the field, the distribution of juveniles was
39 observed during daylight-hours to determine whether *C. concholepas* preferred shaded or
40 illuminated microhabitats. Moreover, we compared the abundance of juveniles collected
41 during day- and night-time hours. The laboratory experiments demonstrated that juveniles
42 of *C. concholepas* seek out and choose their prey more efficiently in darkened areas. White
43 LED illuminated conditions increased righting times and metabolism. Field surveys
44 indicated that, during daylight hours, juveniles were more abundant in shaded micro-
45 habitats than in illuminated ones. However, during darkness hours, individuals were not
46 seen to aggregate in any particular microhabitats. We conclude that the exposure to ALAN
47 might disrupt important behavioral and physiological traits of small juveniles ~~in~~ this
48 species which, as a mechanism to avoid visual predators, are mainly active at night. It
49 follows that ALAN in coastal areas might modify the entire community structure of
50 intertidal habitats by altering the behavior of this keystone species.

51

52 **Keywords:** intertidal, light pollution, ecology, prey-searching behavior, self-righting,
53 metabolism

54

55

56

57

58

59

60

61

62

63

64

65

66

67 **1. Introduction**

68 Marine environments are facing a growing number of stressors associated with global
69 climate change, local human activities and the urbanization of coastal areas. In the face of
70 this proliferation of human impacts, artificial light pollution has often been overlooked
71 despite growing evidence that ALAN could pose a threat to the diversity and functioning of
72 biological communities in terrestrial (Gaston et al. 2014, Davies et al. 2016, Davies et al.
73 2017, Davies & Smyth 2017) and marine communities (Becker et al. 2013, Gaston et al.
74 2014, Davies et al. 2015, Bolton et al. 2017, Davies & Smyth 2017). For example, exposure
75 to ALAN increased the total abundance, and modified the community composition, of
76 spiders and beetles in a grassland ecosystem (Davies et al. 2017). In marine ecosystems,
77 Lorne and Salmon (2007) showed that sea turtle orientation was negatively affected by
78 ALAN, impairing the ability of hatchlings to respond to natural orientation cues. Similarly,
79 in nocturnally migrant birds ALAN altered multiple behaviors (Van Doren et al. 2017) and
80 even human health traits such as sleep, circadian timing, next-morning alertness and
81 increased risk of breast cancer has been shown to be affected by ALAN (e.g. Chang et al.
82 2015, Keshet-Sitton et al. 2015, Zielinska-Dabkowska 2018).

83

84 Shifts in spectral signatures associated with ALAN might affect visually guided
85 behaviors across a broad taxonomic group of animals (Davies et al. 2013). It has been
86 estimated that ~19% of the global land area of the world it is now affected, to some extent,
87 by ALAN (Cinzano et al. 2001, Kyba et al. 2017, Kyba 2018). Moreover, it is estimated
88 that the total area affected by this anthropogenic change in lighting technology is increasing
89 by 6% per year (Hölker et al. 2010), which suggest that this stressor might have far
90 reaching consequences. Light-emitting diodes (LEDs) are cheap, bright, highly efficient

91 and reduce energy consumption. All of which means that LEDs are rapidly becoming one
92 of the world's most important light sources (Zissis & Bertoldi 2014) and are increasingly
93 being used for lighting in both residential and commercial areas as well as the transport
94 routes between them. In the marine environment this will specifically include beachfront
95 developments, ports, marinas and shipping. Therefore, the potential impact of this change
96 to LED illumination on marine communities needs to be considered (Gaston et al. 2015).

97

98 In the marine realm, many species have evolved behavioral and morphological
99 responses to minimize visual predation (Troscianko et al. 2009, Manríquez et al. 2009). For
100 instance, some intertidal species are most active during the night as a mechanism to avoid
101 visual predators (Wells 1980). In addition, being active at night minimizes thermal abiotic
102 stress and desiccation at low tide. This is particularly advantageous for organisms
103 performing energy-demanding activities (e.g. Kennedy et al. 2000). Recently, ALAN has
104 also been shown to affect the locomotor activity, circadian rhythm and growth rate of
105 intertidal amphipods (Luarte et al. 2016) as well as the small-scale diel vertical migrations
106 of zooplankton species (Ludvigsen et al. 2018). Therefore, the modification of the natural
107 light-dark regime by ALAN in coastal environments could have important consequences
108 for the species inhabiting these areas.

109

110 The “Loco” or “Chilean abalone”, *Concholepas concholepas* (Bruguière, 1789), is a
111 keystone species (i.e. its presence maintains the structure and integrity of the community)
112 in rocky shores of the south-eastern Pacific Ocean coast (Castilla & Paine 1987, Power et
113 al. 1996, Castilla 1999). This species is an economically and ecologically important
114 component of the rocky intertidal and subtidal communities along the Chilean coast

115 (Castilla 1999). According to observations conducted under laboratory conditions with
116 intertidal individuals, *C. concholepas* prey mainly at night (Castilla et al. 1979, Castilla &
117 Guisado 1979, Castilla & Cancino 1979, Guisado & Castilla 1983). Meanwhile, studies
118 conducted using subtidal individuals indicated that *C. concholepas* prey over the entire 24-
119 h cycle (Stotz et al. 2003) suggesting that, in this species, intertidal and subtidal populations
120 display different activity patterns. Competent larvae of *C. concholepas* show a marked
121 circadian rhythm in their swimming behavior, displaying most of their activity at night
122 (Manríquez & Castilla 2001). However, it is not yet known if the behavior of benthic stages
123 of this species are also timed over the lunar or tidal cycle is unknown. Among the most
124 important prey items of *C. concholepas* are barnacles, mussels and ascidians (Stotz et al.
125 2003, Manríquez & Castilla 2018), all of which are sessile or have limited mobility.
126 Therefore, it is highly unlikely that preying at night in this species is a mechanism that
127 evolved to avoid being perceived while approaching prey. Instead, it can be argued that
128 preying at night might be a potential mechanism evolved by *C. concholepas* to avoid its
129 own visual predators: the crab *Acanthocyclus hassleri* (Manríquez et al. 2013a, 2013b), the
130 birds *Larus dominicanus* and *Haematopus ater* (Castilla & Cancino 1979), the sea otter
131 *Lontra felina* (Castilla & Bahamondes 1979), and the fish *Pimelometopon maculatus* and
132 *Syciasis sanguineous* (Viviani 1975).

133

134 Similar to most mollusks, *C. concholepas* can use chemical and visual stimuli during
135 sensory perception (Manríquez et al. 2014, Domenici et al. 2017). In this species, the
136 detection of chemical cues associated with prey and predators play an important role in
137 feeding and predation avoidance (Manríquez et al. 2013a, Manríquez et al. 2014).
138 Moreover, as in other marine gastropods, chemoreception of odor cues emanating from

139 food items, conspecifics or predators, involves the osphradium, an external sensory organ,
140 which monitors the physiochemical properties of the surrounding seawater (Huaquín &
141 Garrido 2000). The structurally simple eyes of these gastropods are situated in each
142 tentacle, and provide information on gross differences in light intensity (distinguishing light
143 and dark), regulate daily and seasonal activities, egg laying behavior, mediate phototactic
144 behavior and locomotion, and in some species, provide also visual detection of forms (Serb
145 2008, Ter Maat et al. 2012). Tentacles withdraw in response to sudden decreases in light
146 intensity, exhibiting a shadow response consisting of partial or total retraction of the body
147 into the shell and downward movement of one or both tentacles (Stoll 1972, 1976). In *C.*
148 *concholepas* specifically, shadow response is observed under field and laboratory
149 conditions once light intensity is suddenly interrupted near the cephalic region of the
150 individuals with an opaque object (Manríquez PH. pers. obs). This suggests that the
151 cephalic eyes, or other sensitive areas in the cephalic region, might play an important role
152 in detecting habitats with appropriate light illumination. Hence, ALAN is likely to have a
153 significant effect on the activity of this species.

154

155 Coastal urbanization and tourism development is followed by coastal land
156 reclamation, creation of artificial beaches (Chee et al. 2017) and beachfront lighting
157 (Hölker et al. 2010). This is particularly important in Antofagasta, northern Chile, where
158 the urban fringe is narrow and urbanization takes place near the coast (Corsín 2001). As
159 previously mentioned, the intertidal habitat is subject to a wide range of stressors including
160 ALAN (Underwood et al. 2017), so the rocky intertidal zone and the organisms inhabiting
161 there are good models to investigate the eco-physiological consequences of ALAN. In this
162 study, we conducted laboratory experiments using juveniles of *C. concholepas* to

163 investigate the potential effects of ALAN (using LED lighting) on prey searching, self-
164 righting speed and metabolism. In this particular environment, prey searching and self-
165 righting success are important traits in mediating both predator-prey interactions and the
166 ability to return to a normal posture after dislodgement, respectively. Moreover, we
167 conducted a field survey to determine whether the distribution of juveniles in shallow
168 subtidal rocky habitats was influenced by the ambient light conditions. Our hypothesis was
169 that exposure to ALAN has significant effects on behavioral and physiological traits of
170 juveniles of this keystone species. We expected that ALAN exposure would inhibit the
171 activity of small juveniles of *C. concholepas* and prompt them to incur increased metabolic
172 costs in searching for food in darkened areas. Given that overturned individuals are
173 more vulnerable to visual predators, we also expected that ALAN would speed up self-
174 righting.

175

176 **2. Methods**

177

178 **2.1 Influence of natural lighting conditions on the abundances of Loco in the field.**

179 This field sampling was conducted to explore the link between the response of small
180 juveniles of *C. concholepas* to ALAN in the laboratory and their natural abundances during
181 night hours. Daytime natural abundances on shaded or illuminated shallow subtidal
182 microhabitats might give cues about where the small juvenile of this species prefer to be
183 more active (e.g. searching for prey). Similarly, night-time abundances might help to know
184 if this pattern changes s in absence of light. The location and abundance of juvenile
185 *Concholepas concholepas* (~1.5 cm in shell length) were determined from surveys
186 conducted during the day at low tide conditions in rock boulders mainly encrusted with

187 barnacle stands. The analyzed boulders were located in subtidal shallow-water habitats
188 inside pools that remain as pockets of seawater when the tide ebbs. These surveys were
189 conducted in Antofagasta, northern Chile, at El Lenguado (23°46'S; 70°28'W) and
190 Trocadero (23°35'S; 70°23'W). In southern Chile, surveys were conducted at Calfuco
191 (39°46'S; 73°23'W) during both day and night-time hours. All the surveys were conducted
192 during austral summer months to match those months when high abundances of small
193 juvenile *C. concholepas* are present at the chosen sampling sites (Manríquez et al. 2009,
194 Manríquez et al. 2012). During the day the average (\pm SE; N; weather condition at sampling
195 time) light intensities at seawater-level were 56.26 (0.67; 9; sunny), 46.64 (2.33; 9; sunny)
196 and 3.16 (0.61; 9; partly cloudy) kilo-lux for El Lenguado, Trocadero and Calfuco,
197 respectively. At night, during the sampling, light intensity measured in Calfuco was zero.
198 Light intensities were measured using a Pro'sKit MT-4017 Light Intensity Meter. However,
199 previous studies have shown that light intensity during full moon under clear conditions
200 could be around 0.1-0.3 lux (Gaston et al. 2013, 2104).

201

202 The presence and abundance of juvenile *C. concholepas* was noted on both the upper
203 and lower surface of rocks boulders that were loose and small enough to be able to be
204 turned by the observer (\leq 20 cm in maximal length). In the pools, and only considering
205 individuals below the water level, juveniles inhabiting the upper, illuminated surface of the
206 boulders were counted before the boulder was turned over and those inhabiting the lower,
207 shaded surface were then counted. After being assessed, all the boulders were returned to
208 their original position. The number of juveniles in each of the 2 micro-habitats (i.e. upper or
209 lower side of the boulders) was standardized by searching time. When the boulders being
210 surveyed were in the shallow subtidal pools, water level was always ~10 cm above the

211 | upper surface levels of the boulders. Therefore, small juveniles of *C. concholepas* were
212 | exposed to contrasting natural lighting but not to desiccation. At each site, the same
213 | observer was used locate juvenile Locos for the entire duration of the survey, with a second
214 | researcher present to measure searching times and the size of the individuals. The traveling
215 | time between micro-habitats and the time needed to turn the loose boulders was not
216 | considered as part of the search times. In southern Chile, the sampling schedule also
217 | included night-time samplings conducted in both micro-habitats during low tides occurring
218 | just a few hours after the daytime sampling. In those surveys, a headlamp equipped with a
219 | LED-generated dim red light was used to help visualizing-observe the small juveniles.

220

221 | **2.2 Collection of individuals for experiments.** During December 2017, early juvenile
222 | stages of Locos (<1 cm in shell length) were collected from rocky intertidal platforms
223 | located next to the remains of the old rail turntable at Caleta Coloso in northern Chile,
224 | Antofagasta (23°45'S; 70°27'W). In this study, we focused our observations and
225 | experiments on small juveniles of *C. concholepas*, because they are the more abundant
226 | stages in rocky intertidal and shallow subtidal habitats. Moreover, because of their small
227 | size these juveniles are the most susceptible to be attacked by visual predators, dislodged
228 | by wave action or disrupted by environmental stressors. At the study site, there is no
229 | urbanization or beachfront lighting in place and the only artificial light sources at night are
230 | from passing car headlights on the highway that runs parallel to the coast line, 50 m away
231 | and behind a 1 m high stone wall. Therefore, it can be assumed that sessile and low
232 | mobility organisms inhabiting this site have not been exposed to direct ALAN. The
233 | juveniles were collected from habitats mainly dominated by stands of the barnacle
234 | *Notochthamalus scabrosus*. Recently settled small individuals of the mussels *Perumytilus*

235 *purpuratus* and *Semimytilus algosus* were also presents on the rocky intertidal platforms.
236 Shell coloration in *C. concholepas* depends on the color of the more abundant prey
237 available in the habitat (Manríquez et al. 2009). Therefore, all the individuals were cryptic
238 with mixed shell color (dark and light colored) and therefore ingestively conditioned (Hall
239 et al. 1982) to the prey used in the laboratory experiments (below). After collection, the
240 individuals were moved to laboratory facilities at the Universidad Católica del Norte,
241 Coquimbo, Chile (29°58'S; 71°21'W) where all the experiments were conducted. On
242 arrival, individuals were maintained in Pexiglas® aquaria with running seawater, under a
243 natural photoperiod for the austral summer months (12:12 h light:dark photoperiod without
244 artificial control for gradual changes at dawn and dusk). During the entire rearing period,
245 lighting in the laboratory was provided indirectly by fluorescent lamps and the light
246 intensity at the level of the aquariums was on average (\pm SE; N) 38.6 lux (\pm 6.5; 6). During
247 the first 2 weeks of rearing, individuals were provided with small rocks incrustated with
248 stands of the barnacle *N. scabrosus* on which to feed and then the diet was switched to
249 small individuals of the mussel *S. algosus*. At the end of each experiment (see below),
250 individuals were weighed to the nearest 0.001 g and measured to the nearest 0.01 mm using
251 an analytical balance (Adam AFA180 LC) and a digital vernier caliper (Mitutoyo 500-196-
252 30), respectively.

253

254 **2.3 Prey-searching under light-darkness conditions.** This experiment was conducted on
255 juvenile *C. concholepas* which had been acclimated to laboratory conditions for 1 week.
256 During this acclimation period, individuals were fed *ad libitum* with barnacle stands
257 collected from the same site used for *C. concholepas* collection. After acclimation, the
258 directional responses of *C. concholepas* to artificial light were evaluated in a two-choice

259 test using Plexiglas® Y-mazes. The left and right Y-maze arms were both 0.60 m long,
260 0.18 m wide and 0.10 m deep (Fig. 1). Three Y-mazes (black lateral walls, black lid and
261 white floor) were run simultaneously with both arms of each maze receiving filtered (1 µm)
262 seawater at a constant rate of 1.0 L h⁻¹. The seawater flowing into the arms of the Y-mazes
263 | dropped ~1 cm onto the surface of the water. As in a previous study (Manríquez et al.
264 2014), the displacement paths of the individual snails were verified at the end of the trials
265 by observing the remains of fine sediment stuck to the mucous track generated by *C.*
266 *concholepas* as the anterior part of their foot travels along the white Y-maze floor. The Y-
267 maze experiments were repeated on 8 consecutive days until a total of 24 replicate trails
268 were obtained. After each run, the Y-mazes were cleaned with tap water to remove any
269 chemical cue that could influence future measurements. At the beginning of each trial, a
270 small flat rock with ~200 barnacles was placed near the seawater inlets in each of the Y-
271 maze arms. The surface of the flat rock bearing the barnacle stands were always upwardly
272 oriented. In each Y-maze, one arm was lit by attaching a white LED light to the seawater
273 inlet while the other arm remained in darkness. All the LED lights used in this study
274 (Chanzon ®) were 5 mm round clear type, forward voltage of 3.2-3.6 V, 0.06 W, with a
275 view angle of 120° and a correlated color temperature (CCT) of 8,000 K. They also had the
276 characteristic light source spectral power distribution (SPD) curve, displaying a first peak at
277 460 nm with a power of 0.028 W/nm (indicating blue content) followed by the bell shape of
278 the second peak at 550 nm (indicating higher yellow content). The LED was powered by a
279 button Lithium cell battery (3V, CR1616) and oriented toward the upper surface of the rock
280 in that arm (Fig. 1a). In the dark arm, a LED was also attached to the seawater inlet but was
281 not switched on and therefore the surface of the rock was not illuminated. This was to
282 ensure that the physical presence of the lighting equipment was the same in both arms and

283 could not influence the results. During the experiments, the upper surface of the rocks was
284 always immersed in water to depth of ~ 3-4 cm. The light intensities were measured in the
285 air just above the water surface (~2 cm) and in the starting zone situated 50 cm away from
286 the inlets using a L-100 PCE Lux Meter. Above the illuminated flat stones the light
287 intensities were on average (\pm SE; N) 329.9 lux (\pm 22.4; 8). The light intensity used in this
288 experiment (and in the others, see below) was within the natural range measured during
289 daytime hours in the locality where the animals were collected (ca. 200 to 1000 lux).
290 However, in the darkened arms and in the starting zone the light intensities were on average
291 (\pm SE; N) 0.016 lux (\pm 0.001; 8) and 0.169 lux (\pm 0.011; 8), respectively. To standardize
292 hunger levels, all trial specimens were starved for 24 h before the experiment. The test
293 individuals were placed in the acclimatization zone for 1 h (~17:00) and then moved
294 carefully to the starting zone. Each trial lasted 15 h, and the final position of the snails in
295 the Y-mazes was recorded (~09:00). The final position of the juveniles in the Y-maze was
296 categorized as lighting arm, darkened arm or in the starting zone (Fig. 1a). To maintain the
297 temperature at 15 °C the Y-mazes were semi-immersed in a temperature-controlled water
298 bath.

299

300 **2.4 Self-righting under light-darkness conditions.** This experiment was designed to
301 examine how light/darkness might influence self-righting success and speed of small
302 juvenile individuals of *C. concholepas*. This experiment was conducted one week after the
303 previous one on a different group of individuals that had been acclimated to laboratory
304 conditions for 2 weeks. To account for the existence of an endogenous rhythm that might
305 mask self-righting performance in response to light/darkness the experiment was run twice.
306 The first experimental run was conducted during daytime hours (~10:00 to 19:00), while

307 the second experimental run was conducted during the night (~22:00 to 03:00). In each
308 experimental run we used 24 different individuals. Each individual was placed in a plastic
309 chamber (5 L), filled with 3.5 L of aerated seawater and immersed in a water bath to
310 maintain the acclimation temperature (~15 °C). The entire chamber was covered with black
311 plastic to prevent light from entering the system (Fig. 1b). The opening of the chamber was
312 equipped with a modified plastic funnel used to close the chamber and to hold a white LED
313 light (the same as the one described in the previous section) (Fig. 1b). The funnel spout was
314 1.0 cm in diameter and 3.0 cm in length and provided an efficient way to stop the room
315 lighting from entering the chamber. When the white LED was on (during daytime or night
316 hours), the light intensity in the bottom of the experimental chamber was on average (\pm SE;
317 N) 485.5 lux (\pm 11.3; 8). However, when the LED was off (during daytime or night hours),
318 the light intensity was 0.068 (\pm 0.001; 8). To improve adherence during righting, the bottom
319 of the chamber was covered with a circular strip of a rough-textured rubber surface (3M™
320 Safety-Walk™ tape). Each juvenile was placed in the chamber for 5 min of acclimatization
321 and then was placed upside down in the middle of the chamber. In this way the plane of the
322 shell aperture was parallel with respect to the horizontal plane, therefore with the same
323 degree of difficulty to self-right (Manríquez et al. 2016; 2017). Self-righting time was
324 defined as the total time taken by the individual to completely return to its normal upright
325 position from the moment that the individual was placed upside down and was measured
326 using a digital stopwatch. A maximum of 15 min to assess self-righting was allowed per
327 individual. Inability to self-right within 15 min was considered as self-righting failure.
328 Regardless of experimental run (diurnal or nocturnal), during darkness a dim red LED Petzl
329 headlight (~15 lux) was used to intermittently illuminate the inside of the chamber to allow

330 the juvenile to be observed. After each trial the chamber was cleaned and filled with fresh
331 aerated seawater in preparation for the next trial.

332 **2.5 Metabolic impact of ALAN.** The effect of ALAN on oxygen consumption in small
333 juvenile *C. concholepas* was measured on a randomly selected group of 32 individuals that
334 had been acclimated to laboratory conditions for 3-4 weeks. Four experimental runs were
335 conducted, each run consisting of 4 illuminated and 4 darkened 60 mL glass respirometry
336 chambers. The illuminated chambers were lit with a white LED, the same as those used in
337 the previous experiments, placed 3 cm above each chamber (Fig. 1c). The light levels in the
338 artificially illuminated and darkened chambers, were on average (\pm SD; N) 480.8 lux
339 (\pm 12.0; 4) and zero, respectively. The darkened chambers were identical except they were
340 completely covered with sticky back vinyl tape leaving a small opening over the Oxygen
341 Sensor Spot (SP-PSt3-NAU, PreSens). The chambers were placed horizontally and semi-
342 immersed in a black Pexiglas® container connected to a refrigerated and heating water bath
343 circulator (Lab Companion RW-2025G). In this container, all the chambers were left with
344 the oxygen sensor spots (fixed at the inner side of the chambers) 0.5 cm above the water
345 level. To avoid any additional stress during the oxygen measurements, the chambers were
346 not manipulated when the fiber optical probe was ~~applied~~ **approached** to the spots. A fiber
347 optic oxygen meter (Fibox 3, PreSens) was used for all oxygen consumption
348 measurements. Calibration was performed using a Na₂SO₃ solution (0% saturation) and
349 using air bubbled seawater (100% saturation). Individuals were starved in tanks containing
350 1.0 μ m running filtered seawater for 24 h prior to the measurements. During this period
351 individuals were maintained at light intensities that were on average (\pm SE; N) 38.6 lux
352 (\pm 6.5; 6) and supplied by daylight fluorescent tubes. Metabolic measurements lasted for at
353 least 6 h, with the first 5 min removed to avoid possible manipulation effects. Special care

354 was taken to prevent oxygen levels from dropping below 70% of air saturation. Background
355 respiration was determined by measuring oxygen consumption without a snail in the
356 chamber, and that was subtracted from the experimental oxygen consumption rates. This
357 experimental setting did not allow us to test whether the return from illumination to
358 darkness reduces oxygen consumption to maintain homeostasis. The temperature during the
359 measurements was stabilized at 15 °C by connecting the container holding the respirometry
360 chambers to a refrigerated and heating water bath circulator (Lab Companion RW-2025G).
361 Wet weight of each individual was measured using an analytical balance (Adam AFA180
362 LC). To determine the presence of encrusting or shell-boring organisms in the small
363 juveniles that might influence oxygen consumption each individual was immersed in a Petri
364 dish filled with seawater and observed under a stereo-microscope (Olympus SZ 61).

365

366 **2.6 Data analysis.** Prey-searching and self-righting behavior can be affected by the size of
367 snail performing that activity. Consequently, the sizes of juvenile *C. concholepas* used in
368 the different lighting treatments were compared using a 1-way ANOVA, after the
369 assumptions of normality and homogeneity of variance were tested. This was done for both
370 the prey-searching and self-righting experiments. Field abundances (square root-
371 transformed data) of juvenile *C. concholepas* among microhabitats were analyzed by 2-way
372 ANOVAs, considering in one case microhabitat type (above or underneath the boulders)
373 and location as main factors and in the other case microhabitat type and hours of sampling
374 (day or night-time hours) as main factors. Self-righting times were compared between
375 treatments using a Kruskal-Wallis, as homogeneity of variance between treatments was
376 violated. The final position of the trial individuals (percentage) in the prey-searching Y-
377 maze experiment under contrasting lighting regimes and in the metabolism measurements

378 were analysed by a Fisher's exact-probability test. All statistical analyses were performed
379 using SPSS v.18.0 (IBM Corp., Armonk, NY, USA). At the end of the experiments all the
380 experimental individuals were returned to the same site where they were collected.

381

382 **3. Results**

383 **3.1 Influence of natural lighting conditions on the abundances of Loco in the field.** The
384 abundance of small juveniles of *C. concholepas* was significantly different between
385 microhabitats ($F_{1,28}$: 0.9307; $p < 0.00001$). Approximately 4 and 5 times more individuals
386 were found in the shaded than in the illuminated microhabitats at El Lenguado and
387 Trocadero, respectively (Fig. 2a). The same analysis found that sampling site ($F_{1,28}$: 0.05; p
388 0.3429) and the interaction with microhabitats type ($F_{1,28}$: 0.05; $p = 0.8261$) were not
389 significant. At Calfuco, ~~and~~ during day-time hours, almost 4 times more small juveniles
390 were found underneath than above the boulders (Fig. 2b). However, during night-time
391 hours the abundances were similar in both microhabitats (Fig. 2b). Overall, the abundance
392 of juvenile *C. concholepas* was significantly higher underneath than above the boulders
393 ($F_{1,28}$: 8.5653; $p = 0.0067$, Fig. 2a). Although more small juveniles were observed during
394 night-time hours (Fig. 2b), the effect of sampling period was ~~not~~ significant ($F_{1,28}$: 3.4661;
395 $p = 0.0732$). Similarly, the interaction between sampling period and microhabitat type was
396 not significant ($F_{1,28}$: 0.05; $p = 0.1702$).

397

398 **3.2 Prey-searching under light-darkness conditions.** There were no significant
399 differences (1-way ANOVA; $F_{1,46} = 0.608$; $p = 0.440$) in terms of size between the
400 individuals used in the illuminated experiments (mean \pm SE = 7.75 \pm 0.16; N=24) and those
401 used in the dark experiments (mean \pm SE = 7.91 \pm 0.14; N=24). At the end of the

402 experiments, most of the trial individuals were found in the darkened arm (Fig. 3). The
403 frequency of juvenile *C. concholepas* recorded in each of the 3 potential final positions (i.e.
404 darkened arm, white LED illuminated arm or starting position) was significantly different
405 from expected based on equal chance of being found on each positions ($\chi^2 = 64.220$; DF =
406 2; $p < 0.0001$; Fig. 3). When the juveniles were found in the darkened arms, the percentage
407 of trial individuals found in each side of the rocks (above or below) or in other areas around
408 the rocks was significantly different from expected (Table 1). In this condition the trial
409 individuals were mainly found (~71%) in the upper surface of the rocks in which the
410 barnacles were attached and no signs of mucous tracks were seen in the illuminated arm.
411 When the trial individuals were found in the illuminated arm, no signs of mucous tracks
412 were seen in the darkened arm and they were only found below the rocks (100%, Table 1).
413 Finally, ~~for~~ the trial individuals found in the starting position, no signs of mucous tracks
414 beyond that position were recorded.

415

416 **3.3 Self-righting under light-darkness conditions.** There were no significant differences
417 (1-way ANOVA; $F_{1,94} = 2.269$; $p = 0.135$) in terms of size between the individuals used in
418 the diurnal experiments (mean \pm SE = 8.12 \pm 0.12; N=48) and those used in the nocturnal
419 experiments (mean \pm SE = 8.40 \pm 0.14; N=48). Moreover, no significant differences were
420 found in size between individuals exposed to white LED illumination or maintained in
421 darkened conditions during day time (1-way ANOVA; $F_{1,46} = 0.472$; $p = 0.495$) or night-
422 time hours (1-way ANOVA; $F_{1,46} = 0.364$; $p = 0.549$) trials. Regardless of whether the
423 observations were conducted during the day or at night, significantly shorter self-righting
424 times were recorded for juvenile *C. concholepas* under dark conditions (Fig. 4a-b; Kruskal-
425 Wallis, $\chi^2 = 10.29$; DF = 1; $p = 0.001$ and Kruskal-Wallis, $\chi^2 = 10.29$; DF = 1; $p = 0.001$ for

426 day and night experiments, respectively). Righting success during the day was ~75% in the
427 dark and 33% for white LED illuminated individuals. Similarly, righting success during the
428 night was ~89% in the dark and ~28% for white LED illuminated individuals.

429

430 **3.4 Metabolic impact of ALAN.** There were no significant differences (1-way ANOVA;
431 $F_{1,34} = 0.4313$; $p = 0.579$) in terms of size between the individuals used in the illuminated
432 experiments (mean \pm SE = 13.96 \pm 0.52; N=18) and those used in the dark experiments
433 (mean \pm SE = 13.68 \pm 0.41; N=18). Oxygen consumption was significantly higher (nearly
434 twice as high) in juvenile *C. concholepas* that were exposed to white LED illumination than
435 in those kept in darkness (1-way ANOVA; $F_{1,34} = 9.637$; $p = 0.004$; Fig. 5). Once the
436 chambers were opened, the juveniles inside the white LED illuminated chambers were
437 primarily found within the opaque blue screw caps away from the white LED lighting
438 source (11 of 12 individuals; $\chi^2 = 8.33$; DF = 1; $p = 0.004$). Meanwhile, once the darkened
439 chambers were opened this behavioral response was not generally observed (3 of 12
440 individuals; $\chi^2 = 3.00$; DF = 1; $p = 0.083$) and most trial individuals were found away from
441 the screw caps.

442

443 **4. Discussion**

444 Field surveys of rock boulders in shallow tide pools indicated that, during the day,
445 juvenile *Concholepas concholepas* congregate on the underside of rock boulders. However
446 at night, these juveniles were present both on top of and on the underside of the boulders.
447 Since these individuals were below the water level, they were not exposed to desiccation or
448 thermal stress. Hence, those results suggest that they were actually avoiding light. Previous
449 studies have indicated that large subtidal individuals of *C. concholepas* are active during

450 both day and night (Stotz et al. 2003). Meanwhile, laboratory studies using intertidal adult
451 individuals have indicated that they are mainly active at night (Castilla & Guisado 1979,
452 Castilla & Cancino 1979, Castilla et al. 1979, Guisado & Castilla 1983). In rocky intertidal
453 habitats, juveniles of this species remain motionless during daytime low tides, and are
454 difficult to see due to their cryptic coloration (Manríquez et al. 2009). They are mainly
455 found hiding in small cracks, crevices, on the undersides of boulders (Castilla & Cancino
456 1979), inside shells of dead barnacles (Manríquez et al. 2004) and near/beneath the pedal
457 disc of sea anemones (Moreno et al. 1993). Our study suggests the existence of a clear
458 pattern in which juveniles of *C. concholepas* are more active at night than ~~during the~~ day
459 ~~hours~~, and select darkened micro-habitats during ~~the~~ day-~~hours~~. In the rocky intertidal those
460 sites might represent micro-habitats selected by small juveniles of *C. concholepas* during
461 ebb tides to reduce desiccation and heating stress. However, our results suggest that similar
462 microhabitats immersed within tide pools can also be used as refuges from stressful or
463 inadequate lighting (i.e. scototactic behavior) and/or to avoid visual predators. If juveniles
464 of *C. concholepas* are actually avoiding lit habitats, then ALAN could modify this species'
465 distribution and prey-predation relationships, as it has been reported for other marine and
466 terrestrial species (e.g. Bird et al. 2004, Becker et al. 2013, Luarte et al 2016, Bolton et al.
467 2017, Pulgar et al. 2018).

468

469 The predation experiment showed that juvenile individuals tend to move to rocks
470 with prey placed in the darkened arm of the Y-maze rather than the lit arm containing
471 identical rocks and prey. This suggests that these organisms can distinguish between white
472 LED illuminated and darkened areas and, therefore, in shallow subtidal habitats ALAN
473 might affect the feeding activity, behavior and success of these snails at night. Given that

474 the snails were able to reach the prey in darkness, these results suggest that prey finding in
475 *C. concholepas* relies on chemical rather than on visual cues, which might be an important
476 trait in a marine keystone predator species that move and prey during hours with low
477 visibility. Absence of mucous tracks in both arms of the Y-mazes suggests that the final
478 position of the experimental individuals represents their first choice. Such preference for
479 darkened over lit areas seems natural and we are confident is not biased by the light levels
480 used in the laboratory. The average light levels used to simulate ALAN treatments (330
481 lux) and the experiments discussed below (self-righting: 486 lux and metabolic impact: 481
482 lux) are within the same order of magnitude of light intensities measured during night hours
483 at the ground level of several field sites in the Chilean littoral (Pulgar et al. 2018, Duarte
484 unp. data). Moreover, light intensities used in our laboratory experiments are also within
485 the order of magnitude of light intensities measured at night on the surface of rocky
486 intertidal habitats next to promenades illuminated by ~~lights~~ luminaires equipped with white
487 LEDs at Antofagasta (author's unpublished data).

488
489 Our findings also show that, unexpectedly, the time required for righting is negatively
490 affected by exposure to ALAN. Fast self-righting is considered an adaptive trait that
491 evolved to reduce lethal attacks by visual predators. Indeed, on wave-swept rocky shores,
492 the ability of *C. concholepas* to self-right after dislodgement and rapidly return to a normal
493 up-right position reduces the risk of predation (Manríquez et al. 2014). Therefore,
494 increased righting times in intertidal habitats affected by ALAN, might increase an
495 individual's susceptibility to a lethal attack by visual predators. Those predators are mainly
496 active at night and include some species of crab (Silva et al. 2010), birds (Dugan et al.
497 1981) and small terrestrial mammals (Navarrete & Castilla 1993). Although the negative

498 effects of ALAN on the righting of overturned snails are clear, *C. concholepas* behavioral
499 responses are complex and probably shaped by their interaction with other species. In our
500 study, exposure to ALAN was done using individuals in isolation rather than as part of an
501 interacting community. Therefore, further works addressing the impact of ALAN on multi-
502 species systems are a logical follow up to this study.

503 We found that the metabolic rate of juveniles of *C. concholepas* was higher in the
504 presence of ALAN than in a darkened environment. In our study, microbial respiration was
505 ~~minimised~~ controlled by using 0.45 µm-filtered seawater and by ensuring that, while
506 encrusting organisms, such as bryozoans, hydrozoans, ascidians and barnacles, were not
507 present oin the shell surface. Moreover, shell-boring phoronids and polychaetes were not
508 present in any of the experimental individuals. Therefore, differences in metabolic rates
509 found in this study can only be attributed to the presence or absence of ALAN. A
510 significant increase in the metabolic rate of juveniles exposed to ALAN suggests an
511 increased demand for energy which otherwise would be allocated to other activities such as
512 predator escape behavior or prey searching. This suggests that exposure to ALAN is a
513 potentially powerful stressor during night-time that might have a selection effect on this
514 species by impairing other energy-demanding activities, an issue that should be addressed
515 in future studies. This also points towards the existence of a metabolic regulatory response
516 in which light may lead to increased metabolic rates and hence ATP demand. Therefore, by
517 minimizing day displacement and avoiding illuminated areas, *C. concholepas* may reduce
518 metabolic rates. Our experimental setting did not allow us to test whether the return to
519 darkness reduced oxygen consumption to maintain homeostasis. Future studies are required
520 to investigate whether acute metabolic changes in response to ALAN induce a metabolic
521 rebound once the individuals are back to not stressful conditions. If *C. concholepas* are

522 conditioned to feed at night hours, exposure to ALAN would increase their normal energy
523 requirements and, in order to meet this increased demand, they may start to undertake
524 foraging activities with higher levels of boldness than normal, such as start foraging in the
525 light. Such activities could make them more vulnerable to visual predators, reducing
526 survival and impacting upon population success. The same applies to individuals that are
527 dislodged when exposed to ALAN, where longer self-righting times will also make them
528 more vulnerable to visual predators.

529

530 At the end of the metabolic measurements, the individuals maintained in the white
531 LED illuminated chambers were usually found inside the screw cap away from the light
532 source. In contrast, this behavior was rarely observed when the metabolism was measured
533 in individuals within darkened chambers. This is consistent with other observations
534 showing that individuals of *C. concholepas* avoided lit areas and moved towards darkened
535 areas in both the respirometry chambers and y-mazes. In nature, such microhabitats might
536 be the underside of boulders submerged in an intertidal pool during low tides. This
537 preferential movement to dark areas, or aversion to bright ones, suggests the existence of a
538 scototactic behavior that has been previously reported in fish (Maximino et al. 2010, Blazer
539 & Rosemberg 2012, Thompson et al. 2016). Scototactic behavior has been suggested as a
540 mechanism used by gastropods to return to coastal habitats after being displaced offshore
541 (Chelazzi & Vannini 1976). Studies conducted in mice have also found that short-term
542 exposure to low levels of night-time fluorescent light increases their metabolic expenditure
543 (Borniger et al. 2014). Consistently with that, in some species of fish, dark conditions
544 decrease the individual's energetic requirements (Parker 2002). Therefore, we suggest that
545 the scototactic behavior in this species can be modulated by lighting and might have further

546 | consequences ~~for~~ other important traits, such as prey finding, predation susceptibility and
547 | metabolism.

548 | **5. Conclusions**

549 | Combined, our results indicate that when exposed to ALAN, small juveniles of *C.*
550 | *concholepas* showed significantly longer self-righting times, higher metabolic rates, and
551 | were less frequently found near the food items available. Moreover, in shallow subtidal
552 | habitats, small juveniles of this species preferred shaded areas during the day, but had no
553 | preference ~~during~~ night-time hours. Such evidence suggests that, due to ALAN, these
554 | juveniles become less efficient ~~at~~ finding food and more vulnerable to visual predators. A
555 | previous study found that juvenile *C. concholepas* that were stressed by a combination of
556 | elevated levels of $p\text{CO}_2$ and the presence of a predatory crab showed significantly shorter
557 | self-righting times than those maintained under control conditions (Manríquez et al.
558 | 2013b). This was not the case with individuals exposed to ALAN. Therefore, we suggest
559 | ALAN can have far-reaching impacts on this and other species of rocky intertidal
560 | communities. We conclude that in the rocky intertidal, habitat complexity and natural or
561 | artificial lighting can play an important role in micro-habitat selection by *C. concholepas*
562 | and other similar species. The reduction of darkness ~~during~~ night-time hours might alter
563 | the availability of appropriate habitat and affect those behavioral and physiological traits
564 | that are needed to make ~~individuals~~ less vulnerable to visual predators. Negative
565 | effects of night-time lighting, from the individual to the community levels, have been
566 | described in the literature (Davies et al. 2015, Luarte et al. 2016, Ludvigsen et al. 2018).
567 | However, our study is the first to report negative effects of ALAN on a keystone predator
568 | species, and thus might have negative implications for community dynamics. This far,
569 | available evidence on climate-driven effects on keystone species, and their cascading

570 effects on coastal communities, has mainly focused on the effects of temperature (e.g.
571 Sanford 1999, Harley 2012, Bonaviri et al. 2017), ocean acidification (Manríquez et al.
572 2013, 2016) or a combination of both. In the light of the results presented here, the
573 sustained growth of ALAN sources and their potential effects on keystone species and
574 associated communities can no longer be ignored. In our study, artificial control of dawn
575 and dusk was not feasible, so the potential crepuscular effect on the investigated behavioral
576 responses cannot be disregarded.

577

578 In practical terms, our results imply that beachfronts equipped with white LED
579 lighting at night might affect important behavioral (i.e. prey finding, self-righting) and
580 physiological (i.e. metabolism) traits in *C. concholepas* or other similar species inhabiting
581 intertidal habitats. This in turn can affect a whole range of species interactions and might
582 have negative consequences for intertidal communities. ~~Due to the fact that~~Because *C.*
583 *concholepas* is a keystone species, changes in the balance of biological interactions caused
584 by ALAN may, eventually, modify the whole community structure as has been described
585 already in other intertidal invertebrates (Underwood et al. 2017) and ecosystems (Frank
586 2006). The practical consequence of these results is the need for ~~the design of~~less
587 biologically disruptive lighting (see Gaston et al. 2012). We argue that the development of
588 beachfront lighting ordinance is required to help protect organisms inhabiting the coast-
589 line. For instance, sea turtle friendly lighting has been designed to protect nocturnal nesting
590 and hatching of sea turtle species (Salmon 2003), and similar initiatives may be required to
591 protect other organisms from ALAN. Further studies assessing the impacts of white LED
592 lighting on feeding rate and survival of *C. concholepas* and other benthic species will allow
593 additional predictions to be made about the effects of environmental changes on coastal

594 habitats. Since LED lighting can be highly directional, one potential solution could be to
595 direct light sources more precisely towards specific targets (i.e. coastal pedestrian paths,
596 street lighting and routes), preventing or minimizing the illumination of coastal habitats in
597 which sensitive species might inhabit.

598

599 **Acknowledgments**

600

601 The field part of this study was supported by the Fondo Nacional de Desarrollo
602 Científico y Tecnológico, FONDECYT grant No 1050841 (2005-2007) to Patricio H.
603 Manríquez (PHM). The experimental work of this study was supported by the FONDECYT
604 grant No 1171056, to Cristian Duarte (CD). During this study PHM was under the tenure of
605 the grant "Climate driven Changes in the Habitat Suitability of Marine Organisms"
606 (CLIMAR, ELAC2015/T01-0495) funded by the Network of the European Union, Latin
607 America and the Caribbean Countries on Joint Innovation and Research Activities. During
608 the writing of this article PHM and CD were under the tenure of the FONDECYT grants
609 No 1181609 and 1171056. Pedro A. Quijón acknowledges the support from a NSERC-DG.

610

611 **ORCID**

612 Patricio H. Manríquez: <https://orcid.org/0000-0001-6586-764X>

613 María Isabel Diaz: <https://orcid.org/0000-0003-2719-4696>

614 José Pulgar: <https://orcid.org/0000-0002-8816-7790>

615 Pedro A. Quijon: <https://orcid.org/0000-0002-7980-065>

616 Steve Widdicombe: <https://orcid.org/0000-0002-0598-0234>

617

618

619

620

621

622

623

624

625

626
 627
 628
 629
 630
 631
 632
 633
 634
 635
 636
 637
 638
 639
 640
 641
 642
 643
 644
 645
 646
 647
 648
 649
 650
 651
 652
 653
 654
 655

Table 1. *Concholepas concholepas*. Percentage of juvenile individuals recorded on or around the rocks placed in the selected darkened or white LED illuminated arm of the Y-maze.

Position of the trail snails	Darkened arm (%; n = 17)	White LED illuminated arm (%; n = 3)
On the rocks (Upper side-with barnacles)	70.58	0
On the rocks (Lower side-without barnacles)	23.53	100
Around the rocks	5.88	0

$\chi^2 = 66.911$; DF = 2;
 p < 0.0001

Percentages of trial snails in each position at the end of the Y-maze experiments were compared using the Fisher's exact test with equal expected proportions in each position category. n, number of snail choosing the corresponding Y-maze arm.

656
 657
 658
 659
 660
 661
 662
 663
 664
 665
 666
 667
 668
 669
 670
 671
 672
 673
 674
 675
 676
 677
 678
 679
 680
 681
 682
 683
 684
 685
 686
 687
 688
 689
 690
 691
 692
 693
 694
 695
 696
 697
 698
 699
 700
 701
 702

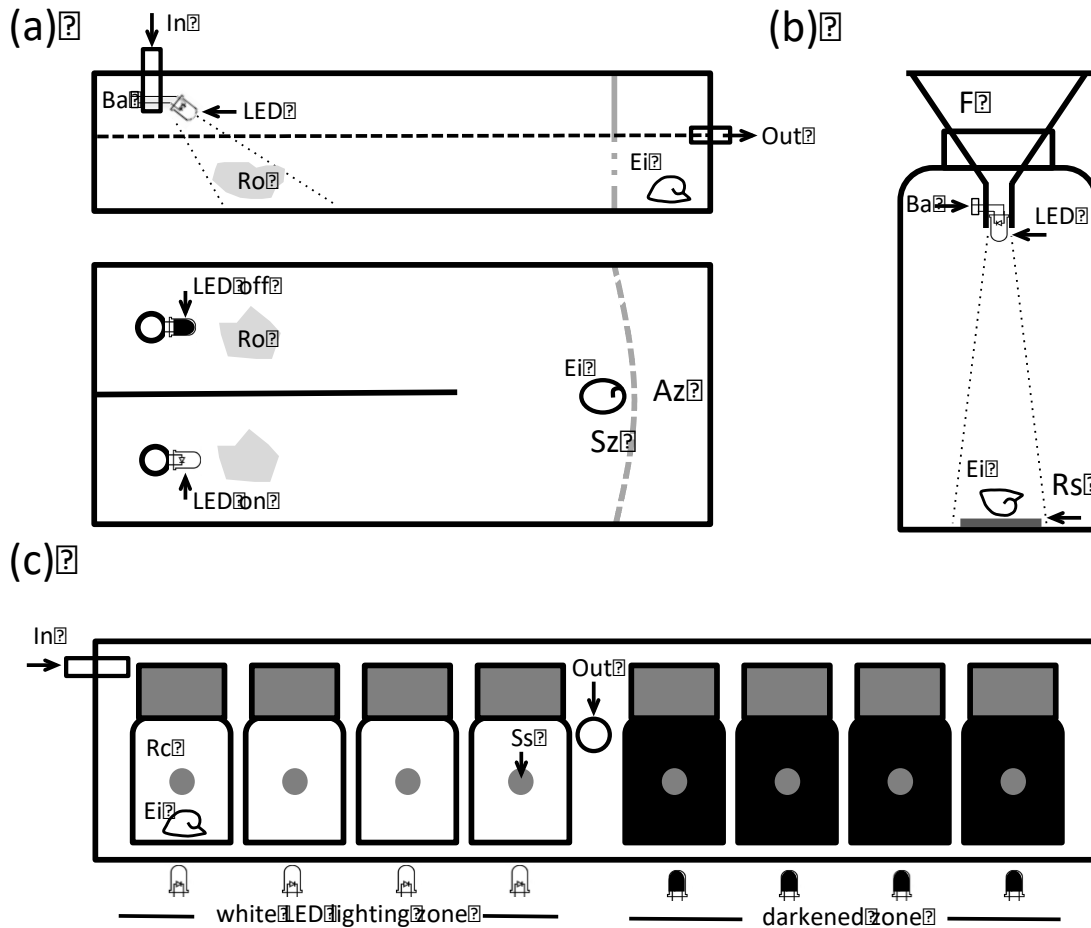


Figure 1. Schematic representation (not to scale) of the experimental setting used to evaluate the effect of white LED lighting on: (a) prey searching, (b) self-righting time and (c) metabolism of small juveniles of *C. concholepas*. In (a) a lateral view (top plot) and top view (bottom plot) of the Y-maze used to evaluate prey searching: Ba = battery; LED = white light-emitting diode; Ro = flat rock with barnacle stands; Ei = experimental individuals; Az = acclimation zone; Sz = starting zone. The dashed line depicts water level. In (b) a lateral view of the plastic chamber used to evaluate self-righting time: F = funnel; Rs = righting substrate conditioned with Safety Walk ® 3M ®. In (c) an upper view of the

703 black Plexiglas container used to hold the glass respirometry chambers (Rc) equipped with
704 oxygen sensors spots (Ss) to measure metabolism.
705

706
 707
 708
 709
 710
 711
 712
 713
 714
 715
 716
 717
 718
 719
 720
 721
 722
 723
 724
 725
 726
 727
 728
 729
 730
 731
 732
 733
 734
 735
 736
 737
 738
 739
 740
 741
 742
 743
 744
 745
 746
 747
 748
 749
 750

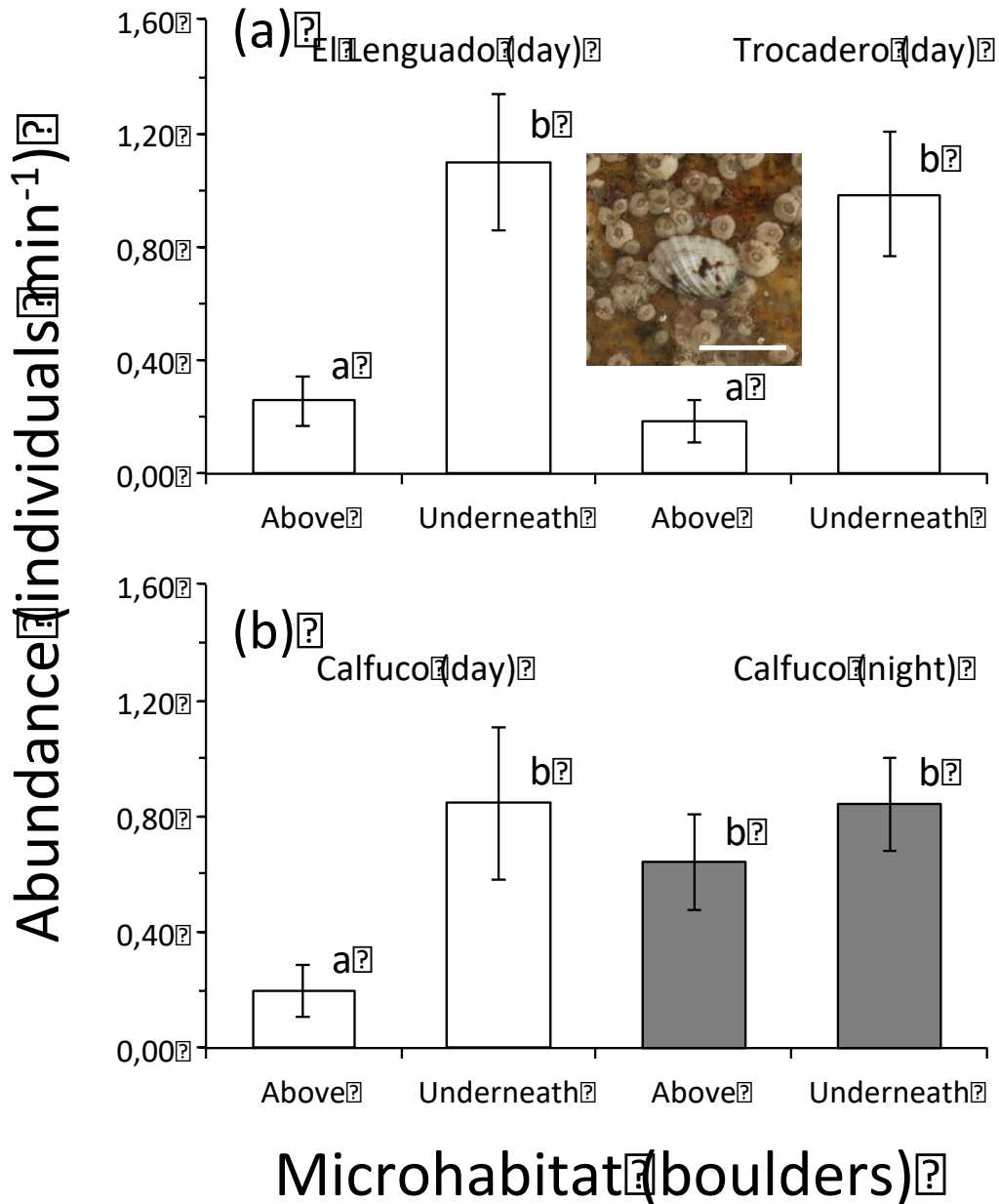


Figure 2. Average (\pm SE) abundances of small juveniles of *C. concholepas* recorded above and underneath rock boulders presents in shallow subtidal pools. In the top graphs are depicted the abundances in El Lenguado and Trocadero (a) during day-time hours at northern Chile. In the bottom graphs (b) the abundances are depicted for day (open bars) and night (filled bar) hours in Calfuco at southern Chile. Different letters indicate significant differences verified by ANOVA 2-way and Tukey's test ($p < 0.05$). In the insert picture a small juvenile of *C. concholepas* preying on a barnacle stand during day-time hours in the underneath side of a rock boulder at Antofagasta, northern Chile. Scale bar 1 cm.

751
752
753
754
755
756
757
758
759
760
761
762
763
764
765
766
767
768
769
770
771
772
773
774
775

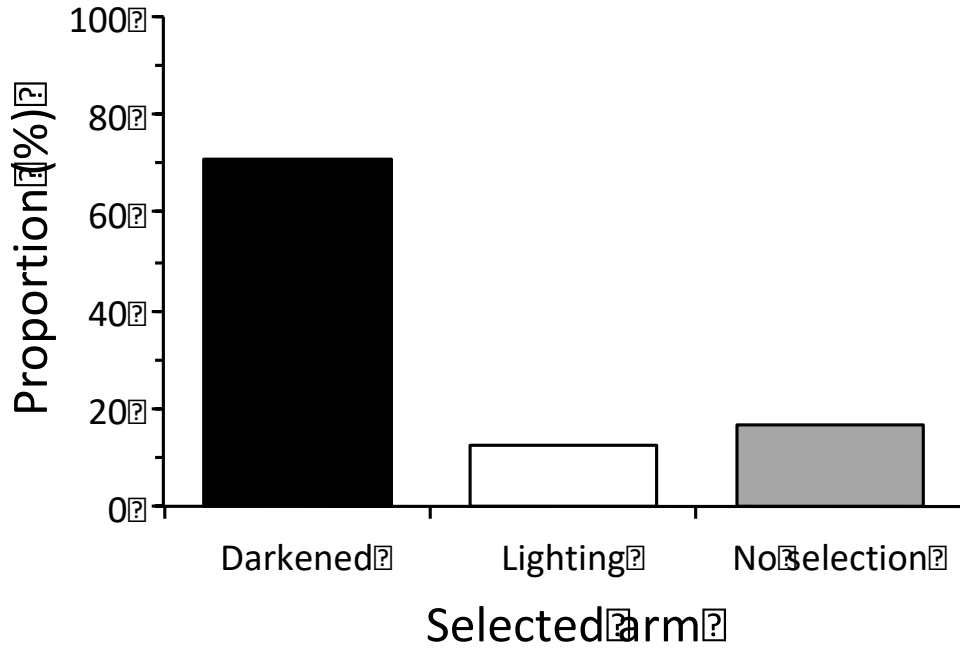


Figure 3. The effect of white LED lighting on the selected Y-maze by the small juveniles of *C. concholepas*. White and dark bars are the proportion of individuals in the white LED illuminated and darkened arm of the Y-maze, respectively. The grey bar is the proportion of individuals that were found in the starting position of the Y-mazes.

776
777
778
779
780
781
782
783
784
785
786
787
788
789
790
791
792
793
794
795
796
797
798
799
800
801

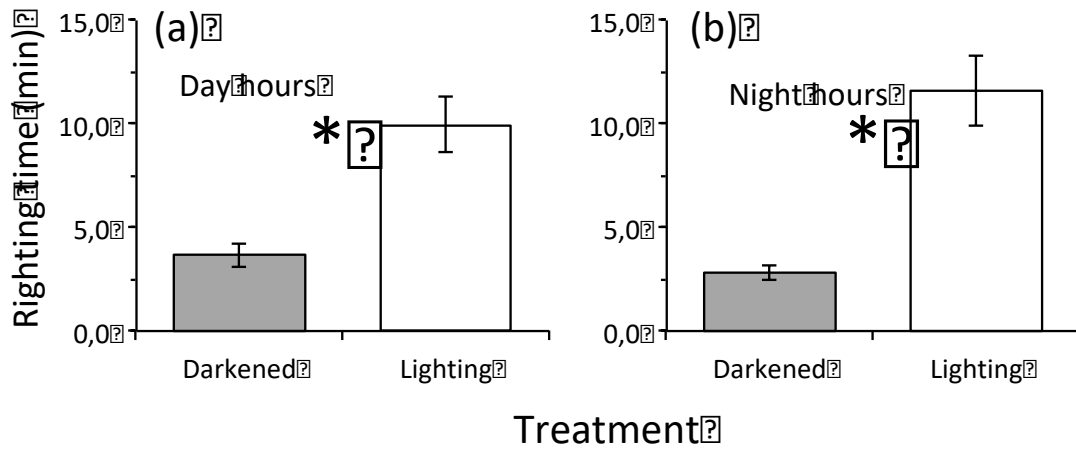


Figure 4. The effect of white LED lighting on self-righting times of small juveniles of *C. concholepas* during day (a) and night hours (b). White and grey bars are average (\pm SE) values in illuminated and darkened conditions respectively. Significant differences ($p < 0,05$) verified by 1-way ANOVA are indicated with an asterisk.

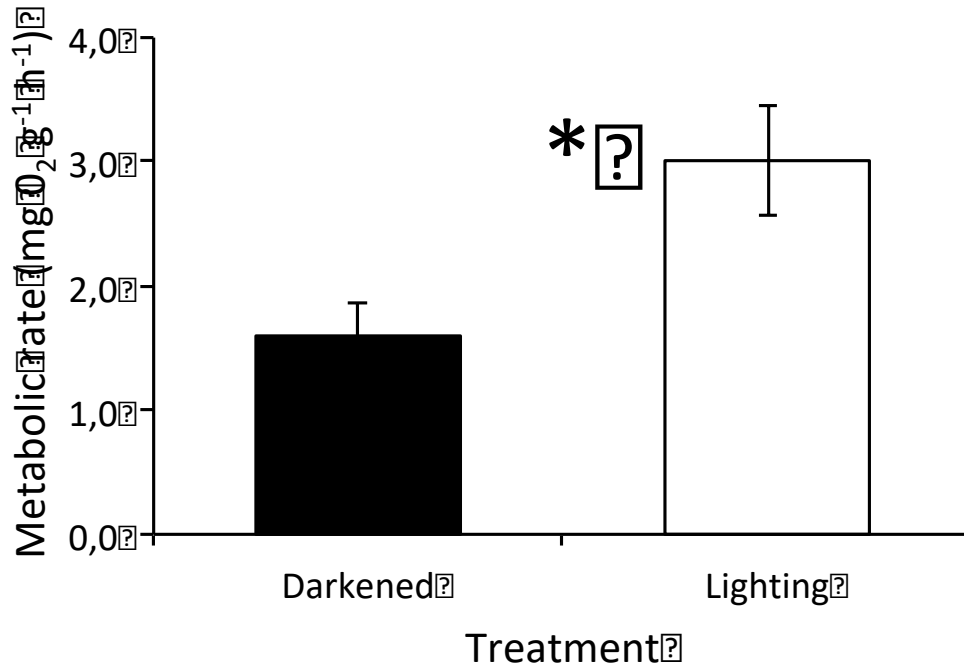


Figure 5. The effect of white LED lighting on metabolism of small juveniles of *C. concholepas*. White and grey bars are average (\pm SE) values in illuminated and darkened conditions respectively. Significant differences ($p < 0.05$) verified by 1-way ANOVA are indicated with an asterisk.

834 **References**

835

836 Bird, B.L., Branch, L.C., Miller, D.L., 2004. Effects of coastal lighting on foraging
837 behaviour of beach mice. *Conserv. Biol.* 18, 1435–1439.

838

839 Becker, A., Whitfield, A.K., Cowley, P.D., Järnegren, J., Næsje, T.F., 2013. Potential
840 effects of artificial light associated with anthropogenic infrastructure on the abundance and
841 foraging behaviour of estuary-associated fishes. *J. Appl. Ecol.* 50, 43–50.

842

843 Blaser, R.E., Rosemberg, D.B., 2012. Measures of anxiety in zebrafish (*Danio rerio*):
844 dissociation of black/white preference and novel tank test. *PlosOne*, 7(5), e36931.

845

846 Bolton, D., Mayer-Pinto, M., Clark, G.F., Dafforn, K.A., Brassil, W.A., Becker, A.,
847 Johnston, E.L., 2017. Coastal urban lighting has ecological consequences for multiple
848 trophic levels under the sea. *Sci. Tot. Environ.* 576, 1–9.

849

850 Bonaviri, C., Graham, M., Gianguzza, P., Shears, N.T., 2017. Warmer temperatures reduce
851 the influence of an important keystone predator. *J. Anim. Ecol.* 86, 490–500.

852

853 Borniger, J.C., McHenry, Z.D., Abi Salloum, B.A., Nelson, R.J., 2014. Exposure to dim
854 light at night during early development increases adult anxiety-like responses. *Physio.*
855 *Behav.* 133, 99–106.

856

857 Boss, K.J., 1982. Mollusca. In: *Synopsis and classification of living organisms* (ed. SB
858 Parker), pp 947-1166. New York: McGraw-Hill Book Co

859

860 Castilla, J.C., 1999 Coastal marine communities: trends and perspectives from human-
861 exclusion experiments. *Trends Ecol. Evol.* 14, 280–283.

862

863 Castilla, J.C., Bahamondes, I., 1979. Observaciones conductuales y ecológicas sobre *Lutra*
864 *felina* (Molina) 1782 (Carnivora: Mustelidae) en las zonas central y centro-norte de Chile.
865 *Arch. Biol. Med. Exp.* 12, 119–132

866

867 Castilla, J.C., Cancino, J., 1979. Main predators of *Concholepas concholepas* (mollusca:
868 gastropoda: muricidae) and preliminary observations about behavioural mechanisms of
869 escape and defense. *Biol. Pesq (Chile)*, 12, 115–123.

870

871 Castilla, J.C., Guisado, Ch., 1979. Conducta de alimentación nocturna de *Concholepas*
872 *concholepas* (Mollusca: Gastropoda: Muricidae). *Biol. Pesq. (Chile)*, 12, 125–130.

873

874 Castilla, J.C., Guisado, Ch., Cancino, J., 1979. Aspectos ecológicos y conductuales
875 relacionados con la alimentación de *Concholepas concholepas* (Mollusca: Gastropoda:
876 Muricidae). *Biol. Pesq. (Chile)*. 12, 91–97.

877

878 Castilla, J.C., Paine, R.T., 1987. Predation and community organization on Eastern Pacific
879 temperate zone, rocky intertidal shores. *Rev. Chil. Hist. Nat.* 60, 131–151.

880

881 Chang, A.M., Aeschbach, D., Duffy, J.F., Czeisler, C.A., 2015. Evening use of light-
882 emitting eReaders negatively affects sleep, circadian timing, and next-morning alertness.
883 Proc. Natl. Acad. Sci. U.S.A. 112, 1232–1237.
884

885 Chee, S.Y., Othman, A.G., Sim, Y.K., Adam, A.N.M., Firth, L.B., 2017. Land reclamation
886 and artificial islands: Walking the tightrope between development and conservation. Glob.
887 Ecol. Conserv. 12, 80–95.
888

889 Chelazzi, G., Vannini, M., 1976. Researches on the coast of Somalia. The shore and the
890 dune of Sar Uanle 9. Coastward orientation after displacement in *Nerita textilis*. Dillwyn
891 (Gastropoda Prosobranchia). *Monitore Zoologico Italiano* (NS Supplemento VIII), 4, 161–
892 178.
893

894 Cinzano, P., Falchi, F., Elvidge, C.D., 2001. The first world atlas of the artificial night sky
895 brightness. Mon. Not. R. Astron. Soc. 328, 689–707.
896

897 Corsin, A., 2001. The becoming of space: a geography of liminal practices of the city of
898 Antofagasta, Chile. D. Phil. thesis, University of Oxford.
899

900 Davies, T.W., Bennie, J., Cruse, D., Blumgart, D., Inger, R., Gaston, K.J., 2017. Multiple
901 night-time light-emitting diode lighting strategies impact grassland invertebrate
902 assemblages. Glob. Chang. Biol. 23, 2641–2648.
903

904 Davies, T.W., Bennie, J., Inger, R., Hempel de Ibarra, N., Gaston, K.J., 2013. Artificial
905 light pollution: are shifting spectral signatures changing the balance of species interactions?
906 Glob. Chang. Biol. 19, 1417–12166.
907

908 Davies, T.W., Coleman, M., Griffith, K., Jenkins, S.R., 2015. Night-time lighting alters the
909 composition of marine epifaunal communities. Biol. Lett. 11, 20150080.
910

911 Davies, T.W., Duffy, J.P., Bennie, J., Gaston, K.J., 2016. Stemming the tide of light
912 pollution encroaching into marine protected areas. Conserv. Lett. 9, 164–171.
913

914 Davies, T.W., Smyth, T., 2017. Why artificial light at night should be a focus for global
915 change research in the 21st century. Glob. Chang. Biol. 24, 872–882.
916

917 Domenici, P., Torres, R., Manríquez, P.H., 2017. Effects of elevated carbon dioxide and
918 increased temperature on locomotion and the repeatability of lateralization of a keystone
919 marine mollusk. J. Exp. Biol. 220, 667–676.
920

921 Dugan, P.J., 1981. The importance of nocturnal foraging in shorebirds: a consequence of
922 increased prey activity. In Jones, N.V. & Wolff, W.J. (eds) Feeding and Survival Strategies
923 of Estuarine Organisms: 251–260. Plenum Press, New York.
924

925 Frank, K.D., 2006. Effects of artificial night lighting on moths. Ecological Consequences of
926 Artificial Night Lighting (ed. by C. Richard T. Longcore), pp. 305–344. Island Press,
927 Washington, District of Columbia.

928 Gaston, K.J., Davies, T.W., Bennie, J., Hopkins, J., 2012. Reducing the ecological
929 consequences of night-time light pollution: options and developments. *J. Appl. Ecol.* 49,
930 1256–1266.
931

932 Gaston, K.J., Bennie, J., Davies, T.W., Hopkins, J., 2013. The ecological impacts of
933 nighttime light pollution: a mechanistic appraisal. *Biol. Rev.* 88, 912–927.
934

935 Gaston, K.J., Duffy, J.P., Gaston, S., Bennie, J., Davies, T.W., 2014. Human alteration of
936 natural light cycles: causes and ecological consequences. *Oecologia.* 176, 917–931.
937

938 Gaston, K.J., Visser, M.E., Hölker, F., 2015. The biological impacts of artificial light at
939 night: the research challenge. *Philos. Trans. R. Soc. London [Biol].* 370, 20140133.
940

941 Guisado, Ch., Castilla, J.C., 1983. Aspects of the ecology and growth of an intertidal
942 juvenile population of *Concholepas concholepas* (Mollusca: Gastropoda: Muricidae) at Las
943 Cruces, Chile. *Mar. Biol.* 73, 99–103.
944

945 Hall, S.J., Todd, C.D., Gordon, A.D., 1982. The influence of ingestive conditioning on the
946 prey species selection in *Aeolidia papillosa* (Mollusca: Nudibranchia). *J. Anim. Ecol.* 51,
947 907–921.
948

949 Hölker, F., Moss, T., Griefahn, B., Kloas, W., Voigt, C.C., Henckel, D., Hänel, A.,
950 Kappeler, P.M., Völker, S., Schwoppe, A., Franke, S., Uhrlandt, D., Fischer, J., Klenke, R.,
951 Wolter, C., Tockner, K. 2010. The dark side of light: a transdisciplinary research agenda for
952 light pollution policy. *Ecol. Soc.* 15(4), 13.
953

954 Huaquín, L.G., Garrido, J. 2000. Morphology and discussion of the possible role of the
955 osphradium in *Concholepas concholepas* (Brugière, 1789) (Neogastropoda: Muricidae). *J.*
956 *Med. Appl. Malacol.* 10, 145–155.
957

958 Kennedy, F., Naylor, E., Jaramillo, E. 2000. Ontogenetic differences in the circadian
959 locomotor activity rhythm of the talitrid amphipod crustacean *Orchestoidea tuberculata*.
960 *Mar. Biol.* 137, 511–517.
961

962 Keshet-Sitton, A., Or-Chen, K., Yitzhak, S., Tzabary, I., Haim, A., 2015. Can avoiding
963 light at night reduce the risk of breast cancer? *Integr. Cancer Ther.* 15, 145–152.
964

965 Kyba, C.C. 2018. Is light pollution getting better or worse?. *Nat. Astron.* 2(4), 267.
966

967 Kyba, C.C., Kuester, T., de Miguel, A.S., Baugh, K., Jechow, A., Hölker, F., Bennie, J.,
968 Elvidge, C.D., Gaston, K.J., Guanter, L. 2017, Artificially lit surface of Earth at night
969 increasing in radiance and extent. *Sci. Adv.* 3(11), e1701528.
970

971 Lorne, J.K., Salmon, L.M., 2007. Effects of exposure to artificial lighting on orientation of
972 hatchling sea turtles on the beach and in the ocean. *Endang. Species Res.* 3, 23–30.
973

974 Duarte, T., Bonta, C.C., Silva-Rodriguez, E.A., Quijón, P.A., Miranda, C., Farias, A.A.,
 975 Duarte, C., 2016. Light pollution reduces activity, food consumption and growth rates in a
 976 sandy beach invertebrate. *Environ. Pollut.* 216, 1147–1153.
 977
 978 Ludvigsen, M., Berge, J., Geoffroy, M., Cohen, J.H., De La Torre, P.R., Nornes, S.M.,
 979 Singh, H., Sørensen, A.J., Daase, M., Johnsen, G., 2018. Use of an autonomous surface
 980 vehicle reveals small-scale diel vertical migrations of zooplankton and susceptibility to
 981 light pollution under low solar irradiance. *Sci. Adv.* 4, eaap9887
 982
 983 Manríquez, P.H., Navarrete, S., Rosson, A., Castilla, J.C., 2004. Settlement of the
 984 gastropod *Concholepas concholepas* on shells of conspecific adults. *J. Mar. Biol. Assoc.*
 985 *U.K.*, 84, 651–658.
 986
 987 Manríquez, P.H., Lagos, N.A., Jara, M.E., Castilla, J., 2009. Adaptive shell color plasticity
 988 during the early ontogeny of an intertidal keystone predator. *Proc. Natl. Acad. Sci. U.S.A.*
 989 106, 16298–16303.
 990
 991 Manríquez, P.H., Galaz, S.P., Optiz, T., Hamilton, S., Paradis, G., Warmer, R.R., Castilla,
 992 J.C., Labra, F.A., Lagos, N.A., 2012. Geographic variation in trace-element signatures in
 993 the statoliths of near-hatch larvae and recruits of *Concholepas concholepas* (lco). *Mar.*
 994 *Ecol. Prog. Ser.* 448, 105–118.
 995
 996 Manríquez, P.H., Jara, M.E., Opitz, T., Castilla, J.C., Lagos, N.A., 2013a. Behavioural and
 997 morphological traits in the early ontogeny of *Concholepas concholepas* in response to
 998 predation risk. *Mar. Ecol. Prog. Ser.* 472, 169–183.
 999
 1000 Manríquez, P.H., Jara, M.E., Mardones, M.L., Navarro, J.M., Torres, R., Lardies, M.A.,
 1001 Vargas, C.A., Duarte, C., Widdicombe, S., Salisbury, S., Lagos, N.A., 2013b. Ocean
 1002 acidification disrupts prey responses to predator cues but not net prey shell growth in
 1003 *Concholepas concholepas* (lco). *PlosOne*, 8(7), e68643.
 1004
 1005 Manríquez, P.H., Jara, M.E., Mardones, M.L., Torres, R., Navarro, J.M., Lardies, M.A.,
 1006 Vargas, C.A., Duarte, C., Lagos, N.A., 2014. Ocean acidification affects predator
 1007 avoidance behaviour but not prey detection in the early ontogeny of a keystone species.
 1008 *Mar. Ecol. Prog. Ser.* 502, 157–167.
 1009
 1010 Manríquez, P.H., Jara, M.E., Seguel, M.E., Torres, R., Alarcon, E., Lee, M.R., 2016. Ocean
 1011 acidification and increased temperature have both positive and negative effects on early
 1012 ontogenetic traits of a rocky shore keystone predator species. *PlosOne*, 11(3), e0151920.
 1013
 1014 Manríquez, P.H., Torres, R., Matson, P.G., Lee, M.R., Jara, M.E., Seguel, M.E., Sepúlveda,
 1015 F., Pereira, L., 2017. Effects of ocean warming and acidification on the early benthic
 1016 ontogeny of an ecologically and economically important echinoderm. *Mar. Ecol. Prog. Ser.*
 1017 563, 169–184.
 1018

1019 Manríquez, P.H., Castilla, J.C., 2018 (in press). Life history, knowledge, bottlenecks, and
1020 challenges for the aquaculture of *Concholepa concholepas* (Gastropoda: Muricidae) in
1021 Chile. *J. Shellfish Res.* 37, 1–14.
1022

1023 Maximino, C., de Brito, T.M. de Mattos Dias, C.A.G., Gouveia, A., Morato, S., 2010.
1024 Scototaxis as anxiety-like behavior in fish. *Nat. Protoc.* 5, 221–228.
1025

1026 Moreno, C.A., Ascencio, G., Ibañez, S., 1993. Patrones de asentamiento de *Concholepas*
1027 *concholepas* (Bruguière) (Mollusca: Muricidae) en la zona intermareal rocosa de Valdivia,
1028 Chile. *Rev. Chil. Hist. Nat.* 66, 93–101.
1029

1030 Navarrete, S.A., Castilla, J.C. 1993. Predation by Norway rats in the intertidal zone of
1031 central Chile. *Mar. Ecol. Prog. Ser.* 92, 187–199.
1032

1033 Parker, R., 2002. *Aquaculture Science*. Delmar, a Division of Thomson Learning, Inc.
1034 621pp.
1035

1036 Pulgar, J., Zeballos, D., Vargas, J., Aldana, M., Manríquez, P.H., Manríquez, K., Quijón,
1037 P.A., Widdicombe, S., Anguita, C., Quintanilla, D., Duarte, C., 2018. Endogenous cycles,
1038 activity patterns and energy expenditure of an intertidal fish is modified by artificial light
1039 pollution at night (ALAN). *Environ. Pollut.* 244, 361–366.
1040

1041 Salmon, M., 2003. Artificial night lighting and sea turtles. *Biologist*, 50, 163–168.
1042

1043 Sanford, E., 1999. Regulation of keystone predation by small changes in ocean
1044 temperature. *Science*, 283, 2095–2097.
1045

1046 Serb, J.M., 2008. Towards developing models to study the disease, ecology, and evolution
1047 of the eye in Mollusca. *Am. Mal. Bull.* 26, 3–28.
1048

1049 Silva, A.C.F., Hawkins, S.J., Boaventura, D.M., Brewster, E., Thompson, R.C., 2010. Use
1050 of the intertidal zone by mobile predators: influence of wave exposure, tidal phase and
1051 elevation on abundance and diet. *Mar. Ecol. Prog. Ser.* 406, 197–210.
1052

1053 Stoll, C.J., 1972. Sensory systems involved in the shadow response of *Lymnaea stagnalis*.
1054 *Proceedings van de Koninklijke Nederlandse Akademie van Wetenschappen Section C,*
1055 *Biological and Medical Sciences*, 75, 342–351.
1056

1057 Stoll, C.J. 1976. Extraocular photoreception in *Lymnaea stagnalis* (L.). In: J. Sálanki., ed.,
1058 *Neurobiology of Invertebrates: Gastropoda Brain*, Tihany 1975, Akadémiai Kiadó
1059 Budapest. 487–495.
1060

1061 Stoll, C.J., Ploep, P., Veerman-Van, D., Van Der Woude, HA., 1976. Light-sensitivity in
1062 pulmonate gastropod *Lymnaea stagnalis* - Peripherally located shadow-receptors.
1063 *Proceedings of the Koninklijke Nederlandse Akademie van Wetenschappen Section C,*
1064 *Biological and Medical Sciences*, 79, 510–516.
1065

1066 Stotz, W.B., González, S.A., Caillaux, L., Aburto, J., 2003. Quantitative evaluation of the
1067 diet and feeding behavior of the carnivorous gastropod, *Concholepas concholepas*
1068 (Bruguiere, 1789) (Muricidae) in subtidal habitats in the southeastern Pacific upwelling
1069 system. *J. Shellfish Res.* 22, 147–164.
1070
1071 Ter Maat, A., Pieneman, A.W., Koene, J.M., 2012. The effect of light on induced egg
1072 laying in the simultaneous hermaphrodite *Lymnaea stagnalis*. *J. Molluscan Stud.* 78, 262–
1073 267.
1074
1075 Thompson, R.R.J., Paul, E.S., Radford, A.N., Purser, J., Mendl, M., 2016. Routine handling
1076 methods affect behaviour of three-spined sticklebacks in a novel test of anxiety. *Behav.*
1077 *Brain Res.* 306, 26–35.
1078
1079 Troscianko, T., Benton, C.P., Lovell, P.G., Tolhurst, D.J., Pizlo, Z., 2009. Camouflage and
1080 visual perception. *Philos. Trans. R. Soc. London [Biol.]*. 364, 449–461.
1081
1082 Underwood, C.N., Davies, T.W., Queirós, A.M., 2017. Artificial light at night alters trophic
1083 interactions of intertidal invertebrates. *J. Anim. Ecol.* 86, 781–789.
1084
1085 Van Doren, B.M., Horton, K.G., Dokter, A.M., Klink, H., Elbin, S.B., Farnsworth, A.,
1086 2017. High-intensity urban light installation dramatically alters nocturnal bird migration.
1087 *Proc. Natl. Acad. Sci. U.S.A.* 42, 11175–11180.
1088
1089 Viviani, C., 1975. Las comunidades marinas litorales en el norte grande de Chile.
1090 Publicación Ocasional del Laboratorio de Ecología Marina, Universidad del Norte, Iquique,
1091 Chile. 196 pp.
1092
1093 Wells, R., 1980. Activity pattern as a mechanism of predator avoidance in two species of
1094 acmaeid limpet. *J. Exp. Mar. Biol. Ecol.* 48, 151–158.
1095
1096 Zielinska-Dabkowska, K.M., 2018. Make lighting healthier. *Nature*, 553, 274–276.
1097
1098 Zissis, G., Bertoldi, P., 2014. Update on the status of the LED market. European
1099 Commission Report EU 27000 EN