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1	Genetic diversity loss in a biodiversity hotspot: ancient DNA
2	quantifies genetic decline and former connectivity in a
3	critically endangered marsupial
4	Carlo Pacioni ¹ , Helen Hunt ¹ , Morten E. Allentoft ^{1,2} , Timothy G. Vaughan ³ , Adrian F.
5	Wayne ⁴ , Alexander Baynes ⁵ , Dalal Haouchar ¹ , Joe Dortch ⁶ and Michael Bunce ^{1,7*}
6	
7	¹ Ancient DNA laboratory, School of Veterinary and Life Sciences, Murdoch University,
8	Murdoch, WA 6150, Australia.
9	² Centre for GeoGenetics, Natural History Museum, University of Copenhagen, Øster
10	Voldgade 5-7, 1350 Copenhagen K, Denmark.
11	³ Department of Computer Science, University of Auckland, Auckland, New Zealand
12	⁴ Department of Parks and Wildlife, Manjimup, WA 6258, Australia.
13	⁵ Western Australian Museum, Locked Bag 49, Welshpool DC, WA 6986, Australia.
14	⁶ Archaeology, M257, The University of Western Australia, 35 Stirling Highway, Nedlands,
15	WA 6009, Australia.
16	⁷ Trace and Environmental DNA laboratory, Department of Environment and Agriculture,
17	Curtin University, Perth, WA 6845, Australia.
18	
19	* Corresponding author: michael.bunce@curtin.edu.au
20	

21 Abstract

22 The extent of genetic diversity loss and former connectivity between fragmented populations 23 are often unknown factors when studying endangered species. While genetic techniques are 24 commonly applied in extant populations to assess temporal and spatial demographic changes, it is no substitute for directly measuring past diversity using ancient DNA (aDNA). We 25 26 analysed both mitochondrial DNA (mtDNA) and nuclear microsatellite loci from 64 historical 27 fossil and skin samples of the critically endangered Western Australian woylie (Bettongia 28 *penicillata ogilbvi*), and compared them with 231 (*n*=152 for mtDNA) modern samples. In 29 modern woylie populations 15 mitochondrial control region (CR) haplotypes were identified. 30 Interestingly, mtDNA CR data from only 29 historical samples demonstrated 15 previously 31 unknown haplotypes and detected an extinct divergent clade. Through modelling, we 32 estimated the loss of CR mtDNA diversity to be between 46% and 91% and estimated this to 33 have occurred in the past 2000-4000 years in association with a dramatic population decline. 34 Additionally, we obtained near-complete 11-loci microsatellite profiles from 21 historical 35 samples. In agreement with the mtDNA data, a number of 'new' microsatellite alleles was 36 only detected in the historical populations despite extensive modern sampling, indicating a 37 nuclear genetic diversity loss greater than 20%. Calculations of genetic diversity 38 (heterozygosity and allelic rarefaction) showed that these were significantly higher in the past 39 and that there was a high degree of gene flow across the woylie's historical range. These 40 findings have an immediate impact on how the extant populations are managed and we recommend the implementation of an assisted migration program to prevent further loss of 41 42 genetic diversity. Our study demonstrates the value of integrating aDNA data into current-day 43 conservation strategies.

Keywords: ancient DNA, population bottleneck, biodiversity loss, genetic diversity, Bettongia

48 Introduction

49 Species and populations that are adversely affected by anthropogenic activities, including 50 habitat modification, are a global phenomenon (Andrén 1994; Ewers & Didham 2006). The 51 reduction in genetic diversity, often associated with population bottlenecks and habitat 52 fragmentation, can result in the loss of reproductive fitness and reduction in evolutionary 53 potential (Frankham 1996; Frankham et al. 1999). Genetic bottlenecks are most commonly 54 studied using modern molecular data and theoretical models are applied to determine their 55 likelihood and severity (Cornuet & Luikart 1996; Garza & Williamson 2001). The 56 incorporation of time-stamped ancient DNA (aDNA) data, for example obtained from 57 museum skins and fossils, can provide direct insight into a population's past diversity 58 (Ramakrishnan & Hadly 2009). The demographic history of bison (Shapiro et al. 2004), musk 59 ox (Campos et al. 2010) and the collared lemming (Brace et al. 2012), for example, have been 60 extensively studied using mitochondrial DNA (mtDNA) isolated from preserved fossil 61 material. However, the use of aDNA data to make informed conservation decisions is rare, 62 despite the fact that it can provide vital information regarding past population demography 63 (e.g. bottlenecks), former connectivity and extent of genetic diversity (see Leonard 2008 for a 64 review).

65

An Australian example of a species adversely affected by population declines is the woylie
(or brush-tailed bettong, *Bettongia penicillata ogilbyi*). Woylies are critically endangered,
nocturnal marsupials (Groom 2010; Wayne *et al.* 2008a), that feed primarily on hypogeal

69 fungi (native truffles), tubers, bulbs, seeds and invertebrates, and they weigh between 1.1 and 70 1.6 kg (Van Dyck & Strahan 2008). With the exception of five translocated sites (i.e. not 71 naturally occurring) in South Australia and New South Wales, the distribution of the woylie is 72 currently limited to the south-west region of Western Australia (WA), one of only 34 73 recognised biodiversity hotspots on the planet (Myers et al. 2000). Historically, together with 74 the subspecies B. p. penicillata, this species had a widespread distribution across southern Australia, but since European settlement (in the 18th century), it has undergone a substantial 75 population decline (Fig. 1), attributed primarily to the introduction of feral predators, 76 77 particularly red foxes (Vulpes vulpes) (Burbidge & McKenzie 1989; Start et al. 1998; Woinarski et al. 2015). Habitat fragmentation and changes in fire regimes have also been 78 79 suggested as contributing factors (Groom 2010). An example of the extent of land clearing in the woylie's former range is the "Wheat-belt" region, an area of 95,800 km² about the size of 80 Portugal, whereonly $\sim 7\%$ of the original vegetation remains in a number of small isolated 81 82 pockets (Saunders 1979).

83

84 Despite a spectacular recovery in the 1990s due to the intensification of the control of the red fox (Start et al. 1998), the woylie has undergone a second, more recent, decline, resulting in a 85 86 90% reduction in seven years from a peak of about 200,000 animals in 1999 (Wayne et al. 87 2013). The factors responsible for this decline are still not well defined (Groom 2010; Wayne 88 et al. 2015). Recent genetic studies of the extant populations using microsatellite data from 89 231 live individuals revealed four genetically distinct wild populations (Pacioni et al. 2011). 90 Two of these are found in the highly fragmented Wheat-belt region: Dryandra woodland and 91 Tutanning Nature Reserve and the other two within the Upper Warren region; Kingston and 92 Perup (Fig. 1). Current size estimates in these naturally occurring populations (excluding translocated populations) are approximately <10,000, with Tutanning Nature Reserve having
gone extinct in the wild in 2011 (Wayne *et al.* 2013, Department of Parks and Wildlife
unpublished data). Woylie translocated populations account for another ~4,000 individuals
(Wayne *et al.* 2013).

97

98 An evaluation of contemporary genetic diversity among 152 modern individuals from the four 99 remnant populations, using the mitochondrial control region (CR), revealed 15 haplotypes, 100 and expected heterozygosity ($H_{\rm E}$, n=231) values of 63-83% for 12 microsatellite loci (Pacioni 101 et al. 2011). However, it is unknown to what extent genetic diversity has been lost since the 102 collapse of the population. The use of heterozygosity values calculated only from modern 103 samples as a measure of genetic health can be problematic due to the lack of comparative 104 baseline data. A direct assessment of genetic diversity using both pre- and post-decline 105 samples is the only way to securely quantify genetic diversity changes over time with any 106 degree of accuracy. An appreciation for temporal effects is often a missing factor when 107 making conservation decisions, the woylies being a case in point where, for example, the lack 108 of baseline data makes the establishment of conservation targets difficult (Pacioni et al. 109 2013). In this study, we used the woylie as an example to investigate to what extent the 110 application of aDNA techniques can improve our understanding of the dynamics associated with wildlife declines. We characterised two genetic markers: mitochondrial DNA (control 111 112 region and cytochrome b) and microsatellites, and assessed the possible loss of the genetic 113 diversity and the degree of historical gene flow between regions. We further discuss how the 114 information obtained can inform conservation decisions.

115

116 Materials and methods

117 Skin and Bone Sampling

118 A total of 64 historical samples were sourced from the collections of the Western Australian 119 Museum (WAM). Samples included fossil bone fragments from vertebrate archaeological 120 collections (n=15), museum skins (n=28) and bone material from the vertebrate 121 palaeontological collection (n=21) ranging between ~40 and 15,600 years old (Dortch 2004a; 122 b, Table S1 supporting information). Samples represent a wide geographical area, with 123 coverage across woylie former distribution (Fig. 1). The entire WAM woylie skin collection 124 was sampled for this study. Woylie fossils, from across southern WA, were selected from 125 sites such as caves that are typically conducive to long-term DNA preservation.

126 DNA Extraction

127 DNA extractions from historical samples were carried out in a dedicated aDNA laboratory 128 located at Murdoch University (Perth, Australia), minimising contamination from PCR 129 amplicons and modern DNA in accordance with established aDNA guidelines (Gilbert et al. 130 2005). Skin samples with attached hair (~5x5 mm) were diced using a sterile scalpel blade 131 and incubated overnight at 55°C with rotation, in 500 µl of digest buffer (20 mM Tris; pH 8, 1% SDS, 5 mM CaCl₂, 10 mM dTT, and 10 mM EDTA) with the addition of 10 mg of 132 133 Proteinase K (Invitrogen). DNA was bound and eluted from the digest buffer using DNeasy 134 blood and tissue kit according to the manufacturer's recommendations (Qiagen). Samples of 135 bone and teeth were grounded to powder using a Dremel tool (part no. 114; Germany) set at a 136 low speed. Approximately 100 mg of bone powder was incubated in a bone digest buffer 137 (containing; 20 mM Tris; pH 8, 1% SDS, 10 mM dTT, 0.48 M EDTA and 10 mg of 138 Proteinase K) overnight at 55°C with rotation. Following digestion, all tissue samples were 139 centrifuged briefly to pellet debris and the supernatant transferred to a Vivaspin 500 tube with 140 a 30,000 Dalton molecular weight cut-off membrane (Sartorius Stedim Biotech) and

141 centrifuged at 13,000 rpm, until \sim 50 µl of supernatant remained. This was then mixed with 142 five volumes of PBI buffer (Qiagen) and DNA was extracted using a DNeasy tissue and blood 143 kit (Qiagen) according to the manufacturer's recommendations.

144

145 DNA quantification

Following extraction, each sample was assayed for DNA preservation using quantitative PCR (qPCR) assays. The ability to successfully genotype nuclear aDNA loci is largely dependent on template copy number (Allentoft *et al.* 2011), accordingly qPCR assays represent an important step in determining relative preservation across a set of samples such as collected here.

151

The mtDNA primers targeting the 12s rRNA region (12s woylie macro 302F, primer 152 153 sequence CGTAAAGCGTGTTTAAGCC and 12s woylie macro 425R, primer sequence 154 CTGTAGTGTATTCAGCAAA; PCR product ~120 bp) were tested and shown to function well in a SYBR-bead qPCR assay and therefore used to screen all samples. The PCR reagents 155 156 and protocol were optimised with a final 25 µl PCR reaction consisting of 2 µl template 157 DNA, 3.0 mM MgCl₂ (Fisher Biotech, FB), 1x PCR buffer (FB), 0.5 mM each dNTP's (Astral 158 Scientific), 0.16 µmoles of each primer (12s 302F/425R), 0.6 µl of SYBR Green dye (Stock 159 1/10,000 dilution; Invitrogen), 0.25 U AmpliTag Gold polymerase (Life Technologies) and 1 160 µl (10 mg/mL) bovine serum albumin (BSA) (Fisher Biotech). PCR conditions were an initial hot start at 95°C for 5 min followed by 40 cycles of 95°C for 45 sec, 57°C for 45 sec and 161 162 72°C for 45 sec. The same qPCR assay was used to assess the relative quantity of DNA in all 163 the historical DNA extracts, to establish the relative copy number of each sample. While many of the "failed" samples still vielded DNA (Table S1), they were excluded from the 164

analyses because of problems associated with allelic dropout of microsatellite loci — a
phenomenon commonly encountered with degraded DNA (Allentoft *et al.* 2011; Bouzat *et al.*167 1998; Taylor *et al.* 1994).

168

169 DNA amplification

170 A ~400 bp fragment of the CR and a 370 bp fragment of the cytochrome b gene were 171 amplified using primers designed to target specific mtDNA regions (Table S2, Supporting 172 information). Reactions typically consisted of a 25 µl reaction containing 2 µl template DNA, 173 2 mM MgSO₄, 1x PCR buffer, 0.25 mM each dNTP's, 8 pmoles of each primer, 0.25 U HiFi 174 Tag polymerase (Invitrogen), 1 µl (10 mg/mL) bovine serum albumin and template DNA. 175 PCR conditions were: a hot start at 95°C for 5 min followed by 40 cycles of 95°C for 45 sec, 56°C/58°C (depending on primer set used) for 45 sec and 68°C for 45 sec. DNA was 176 sequenced in both directions using dye terminator cycle sequencing chemistry (3730xl 177 178 sequencer; Applied Biosystems via Macrogen - BigDye ver 3.1). Sequences were aligned in 179 Geneious Pro (BioMatters). PCR repetitions using varying amounts of input DNA were 180 conducted on individuals with unique mutations in order to check sequence fidelity. In 181 addition, those fragments with unique mutations were cloned using the p-GEM "TA" vector 182 system (Promega) to verify these rare variants and to investigate whether they could be 183 accounted for by post-mortem DNA damage (primarily the deamination of cytosine, causing a 184 C to T change during PCR amplification) (Briggs et al. 2007; Willerslev & Cooper 2005).

185

186 The microsatellite loci targeted in this study were previously used to characterise modern 187 woylie genetic diversity (Table S3; Pacioni & Spencer 2010), with the exclusion of Y151 due 188 to large variation in amplicon product size at this locus. All microsatellite amplifications were conducted as described above except that Amplitaq Gold polymerase (Applied Biosystems)
was used and 1.25 mM MgCl₂. Primers for two loci (Y175 and Bt76) were redesigned closer
to the microsatellite repeating unit in order to shorten the allele length and reduce allelic
dropout.

193

194 DNA fragment analysis of the amplified microsatellite products was carried out using a 5-dve 195 system on an Applied Biosystems 3730 DNA analyser. The PCR product size was 196 determined using an internal size standard (Genescan TM-500 Liz; Applied Biosystems) and 197 fragments were scored using GeneMarker software (v1.5, Soft Genetics). Genotyping was 198 repeated at least three times for homozygotes, with the addition of different starting amounts 199 of DNA and were accepted only when all three repeats were homozygous for the same allele 200 and, similarly, heterozygotes were repeated twice and accepted when both were heterozygous 201 for the same alleles. We calculated the frequencies of falsely scoring an individual as 202 homozygous (P_(fh), Gagneux et al. 1997) and then the probability of falsely scoring an 203 individual (per locus) given our protocols (i.e. $P_{(fh)}^{3}$). Two modern samples were co-run in 204 each PCR to ensure direct comparability between the modern and historical datasets, and the 205 allele scores of microsatellites from redesigned primer pairs were adjusted accordingly.

206

207 Mitochondrial DNA data analysis

Haplotypes from the historical samples were identified with DNAsp (Librado & Rozas 2009), which was also used to calculate haplotype (h) and nucleotide diversity (π) (Nei 1987). We applied the Stirling probability distribution to calculate the posterior probability of the expected total number of (historical) haplotypes (Dixon 2006). As a result, we calculated the probability that additional haplotypes were yet to be sampled, and estimated the expected total 213 number of historical haplotypes, based on the observations and sample size of the historical 214 data. Sequences from modern data with complete coverage of the region amplified from 215 historical samples (n=146) were obtained from Pacioni *et al.* (2011), and their h and π values 216 were compared with those obtained from historical samples using a *t*-test. Rarefaction curves 217 from both historical and modern datasets were calculated using Vegan (Oksanen et al. 2013) 218 in R v2.15 (R DevelopmentCore Team 2015) and plotted to allow graphical comparison of haplotype richness between the two datasets. The haplotype richness obtained with the 219 220 rarefaction approach allows for a direct comparison between the "modern" and "ancient" 221 datasets because it standardises the sampling effort. To further explore the expected number 222 of historical haplotypes, we fitted the two- and three-parameter asymptotic exponential 223 models to the haplotype richness data, compared them with the F-test, and then predicted, 224 using the most supported model, the number of historical haplotypes that we would have 225 obtained had we had available the same number of samples as for the modern data. Spatial-226 temporal changes in the mtDNA genetic diversity were also visualised with a multi-227 dimensional parsimony network built in TempNet v1.8 (Prost & Anderson 2011).

228

229 The CR and cytochrome b sequence alignments were analysed with the Bayesian methods 230 implemented in BEAST v1.5.3 (Drummond & Rambaut 2007) to obtain the phylogenetic 231 relationships. The general time-reversible (GTR) model of nucleotide substitution with 232 invariant and gamma sites (I+G), was used as determined by Modeltest v3.7 (Posada & 233 Crandall 1998). A strict molecular clock was assumed. The northern bettong (Bettongia 234 tropica, GENBANK accession numbers AF287895-AF287906) is a sister taxon to B. 235 *penicillata* and was therefore used to root the tree. To place the woylie radiation in a temporal 236 context, the cytochrome b sequences, along with other marsupial sequences (Genbank 237 accession numbers in Table S4), were modeled in BEAST using the GTR+I+G substitution 238 model, assuming a strict molecular clock and Yule process speciation, and enforcing two 239 published calibrations (Westerman et al. 2004). A calibration of 23.8 million years (95% CI 240 23.0-25.9), for the divergence between the Macropodinae and the Potoroinae, and of 8.3 241 million years (95% CI 5-12) for divergence of Bettongia were imposed. Three runs of ten 242 million generations (sampling every 1000 trees) were conducted and combined using 243 LogCombiner v1.5.3. After a burn in of 1000 trees per run, a consensus tree was generated 244 using TreeAnnotator v1.5.3.

245

Next, we evaluated changes of woylie population size over time using coalescent-based 246 247 approaches as these have proven efficient in detecting demographic changes (Peery et al. 248 2012). We analysed the woylie CR mtDNA data using BEAST v.1.8 (Drummond et al. 2012) 249 running 200 million iterations and discarding the first 10% as burn in. We used the estimated 250 dates of the aDNA samples and calibrated the root with a lognormal prior using the results of 251 the analysis of the cytochrome b dataset described above. We used a lognormal prior also for 252 the clock rate while leaving default options for the remaining priors. Using the Akaike 253 information criterion through Markov chain Monte Carlo (AICM, Baele et al. 2012) with 1000 254 bootstraps, we compared three demographic models: constant population size, exponential 255 growth and Bayesian Skyline Plot (BSP). Appropriate mixing and estimated sample size in the 256 analyses was checked with Tracer v.1.6 (Rambaut & Drummond 2007).

257

We also used a structured coalescent sampler to analyse these data to account for potential problems arising from population structuring (Ho & Shapiro 2011; Navascués *et al.* 2010). The historical data for these analyses were limited to samples that were geographically close to the 261 extant populations (n=13). Using Migrate-n v3.6.8 (Beerli 2006), we compared migration 262 models of different levels of complexity: we started from a full model where the four 263 populations (Tutanning, Dryandra, Kingston and Perup) were kept separate. Because of their 264 relative geographic position, direct migration from Tutanning to any of the Upper Warren 265 populations (and vice versa) was prevented (i.e. animals from Tutanning could only migrate to 266 Dryandra and vice versa). We considered two three-population models where in one the two 267 populations from the Wheat-belt region (Dryandra and Tutanning) were merged into one 268 population and Perup and Kingston kept separate and, in the second, the Perup and Kingston 269 were merged and the Wheat-belt populations were kept separate. We further simplified the 270 migration model by considering only two populations (one for each region) and lastly we 271 simulated a panmictic population. These analyses were conducted running 10 replicates of 400 272 million iterations, sampling every 2000 trees and discarding 30% of the trees as burn-in. Each 273 analysis ran four chains with the default static heating scheme using slice sampling (Neal 2003) 274 and with a gamma prior for both population parameters (Θ : 0.2, 0.06, M: 500, 0.5). Eventually, 275 models were compared using the log Bayes Factor (LBF) calculated with the Bezier marginal 276 likelihoods obtained by thermodynamic integration. Once the migration model was selected, 277 we used Migrate-n to generate BSPs. Appropriate mixing and estimated sample size in 278 Migrate-n analyses as well as post-analysis data plotting was carried out with the R package 279 mtraceR (https://github.com/carlopacioni/mtraceR). To evaluate the extent of the bias 280 introduced in BEAST where substructuring was not explicitly taken into account, we generated 281 a BSP using BEAST with the same input file used for Migrate-n (i.e. limiting the historical 282 samples to geographically close to the extent populations, n=13) and graphically compared the 283 results. Where relevant, we repeated the demographic analysis with modern data only to 284 evaluate whether the addition of aDNA data had substantially changed our results.

285

286 Microsatellite data analysis

287 The microsatellite data do not strictly conform to the requirements for testing the Hardy-288 Weinberg Equilibrium, as they do not represent a population in the same space and time 289 (Frankham et al. 2002). Too few samples were available from one location during the same 290 time to test meaningfully whether genotype frequencies were in Hardy-Weinberg proportions. 291 Therefore, the data prevented identifying and correcting genotyping errors based on 292 deviations from Hardy-Weinberg proportions, as well as deficiencies and excesses of 293 particular genotypes using these standard approaches (e.g. Kalinowski 2006). However, 294 allelic dropout was identified from PCR replications of historical samples.

295

296 Descriptive measures of genetic diversity were calculated in GENALEX v6.2 (Peakall & 297 Smouse 2006), which included measures of allelic frequency per locus, observed 298 heterozygosity (H_o) and observed number of alleles (N_A).

299

300 Historical data were compared with modern data obtained from Pacioni et al. (2011) to 301 identify alleles that were not present in the modern data (unique alleles). Because differences 302 in sampling intensity can bias comparison of genetic diversity, with datasets containing larger 303 number of samples expected to contain more alleles (Kalinowski 2004), we used rarefaction 304 to compensate for the differences in sampling effort and calculated the number of alleles 305 present in an increasing sample size (from 1 to 20 for the historical dataset and from 1 to 210 306 samples for the modern data) using PopGenKit (Rioux Paquette 2011) in R v2.15 (R 307 DevelopmentCore Team 2015) and plotted the rarefaction curves for each dataset. Similarly to what we did with the mtDNA data, we fitted the two- and three-parameter asymptotic 308

309 exponential models to the haplotype richness data, compared them with the F-test, and then 310 predicted, using the most supported model, the number of historical haplotypes that we would 311 have obtained had we had available the same number of samples as for the modern data. 312 Moreover, we compared allelic richness, obtained sampling 11 diploid individuals, between 313 the modern and historic data with the non-parametric Wilcoxon signed-rank test (Kalinowski 314 2004). We also compared the individual IH between the two datasets with Mann-Whitney 315 test. Furthermore, to investigate possible differences in connectivity between regions we 316 estimated the number of migrants using the "private allele" (the alleles found exclusively in 317 one population) method (Barton & Slatkin 1986) and used GENEPOP v4.2 (Rousset 2008) 318 for this purpose. We selected this method because its assumptions are not violated by 319 heterochronous data. However, we acknowledge that the number of migrants estimated with 320 this approach is possibly biased due to the fact that most private alleles would also be rare 321 and, therefore, may be lost during population declines (i.e. the number of migrants calculated 322 from the modern dataset would be inflated). For the latter analysis, data from the two woylie 323 populations in the Upper Warren region were merged because records did not specify whether 324 historical samples were collected from the eastern or western compartment of the Upper 325 Warren forest. One locus was omitted from all statistical comparisons: Y112, due to high 326 rates of allelic dropout evident from PCR repetitions (see below). All univariate statistical 327 analyses were performed with SPSS v19 (IBM. 2010) or R v2.15 (R DevelopmentCore Team 328 2015).

329

Evidence of population bottlenecks for each extant population was investigated, using modern microsatellite data, by testing for excess in heterozygosity (Cornuet & Luikart 1996) and mode-shift (Luikart & Cornuet 1998), with the program BOTTLENECK (Piry *et al.* 1999). A 333 mixed model of microsatellite mutation was assumed, with single step mutations accounting 334 for 95% of all mutation events and a variance among multiple steps of 12 (Piry et al. 1999). 335 We used a Wilcoxon signed-rank test to test statistical significance (Piry et al. 1999). We also 336 used the M-ratio method with M P VAL, using recommended mutation parameters: 0.2 for 337 the proportion of one-step mutations (p_s) , 3.5 as the average size of non one-step mutations (Δ_{\circ}) and a mutation rate (µ) of 5×10^{-4} /locus/generation (Garza & Williamson 2001). It was 338 assumed that the population census sizes (Nc) prior to European settlement were similar to 339 340 those estimated at the peak density observed after the commencement of the fox control 341 program, immediately prior to the most recent declines (i.e. a conservative and best available 342 estimate of pre-European populations) (Groom 2010). Following Frankham (1995), we approximated effective population size (Ne) to be 10% of Nc. Tutanning pre-European 343 344 settlement Ne was assumed to be similar to the Ne at Dryandra. This test is anticipated to 345 produce positive results for at least 100 generations post-reduction (Garza & Williamson 346 2001).

347

Lastly, we used Migrate-n to estimate demographic changes over time after having determined an appropriate migration model with the same approach as outlined for the mtDNA data. Microsatellite data were analysed with an exponential prior (Θ : 300; M: 1000) and running 4 replicates of 100 million iterations, sampling every 1250 trees and discarding 25% of the trees as burn-in. Each analysis ran four chains with the default static heating scheme using slice sampling (Neal 2003).

³⁵⁵ Results

356 DNA recovery and quantification

357 Of the 64 samples tested, including historical skins and fossils bones, DNA was successfully 358 isolated from 49 samples (76.5% success, Table S1). However, we specifically focused on a 359 subset of 29 well preserved samples, as determined by relative qPCR cycle threshold (C_T) 360 values (C_T is the number of cycles that are needed to obtained a pre-determined fluorescent 361 intensity and are inversely proportional to the initial amount of template used in the 362 reactions), spanning the entire former range of the woylie. There were two reasons to focus on 363 this subset of samples: firstly, we wanted to minimise the effect of microsatellite allelic drop 364 out often encountered with low copy number DNA (Allentoft et al. 2011); secondly, samples 365 with good DNA preservation are less prone to artefacts such as contamination and DNA 366 damage.

367

Skin samples had relatively better DNA preservation than fossil bone, with a mean ΔC_T value 368 369 of 4.24 or 18.9 times more DNA (25.55±5.90 SE compared to 29.77±5.79). Likewise, 370 younger samples yielded more mtDNA quantities than older samples (Table S1. See also 371 Hunt 2010). Although there was substantial variation in preservation, as expected (Allentoft 372 et al. 2012), there was a general trend with older specimens having lower relative DNA 373 concentrations, presumably due to a longer exposure to hydrolytic and enzymatic processes 374 (Allentoft et al. 2012). Lastly, as previously observed (Allentoft et al. 2011), there was an overall trend between the C_T values and the percentage of microsatellite loci that could be 375 376 typed (Fig. S1).

378 Mitochondrial DNA analysis

379 The lengths of the CR sequences obtained from the 29 historical samples varied between 250 380 and 400 bp (excluding primers). While all of these were used for phylogenetic analysis (see 381 below), we limited statistical analysis to 23 historical samples of ~400 bp (Table S1). Among 382 these 23 historical samples (of which ~70% were around 100 years old or less, see Table S1), 383 we identified a total of 18 haplotypes. Six samples possessed four CR haplotypes that are still 384 present in extant populations. The remaining 15 haplotypes have not previously been 385 observed despite intensive sampling of the modern woylie populations (Pacioni et al. 2011). 386 These differences are further highlighted in the multidimensional network (Fig. 2).

387

Haplotype and nucleotide diversity of the sequences obtained from historical data was 0.976 (SD=0.02) and 0.039 (SD=0.004), which was substantially higher than any of the values calculated from the extant wild populations (range: h=0.26-0.82; $\pi=0-0.022$. Pacioni *et al.* 2013), and significantly higher than that obtained from the modern dataset (h=0.735, SD=0.001, p<0.001; $\pi=0.022$, SD=0.002, p<0.001).

393

394 The rarefaction curve computed from the historical dataset had a much higher trajectory and 395 did not plateau, suggesting that there were many unsampled historical haplotypes (Fig. 3a). 396 The three-parameter asymptotic exponential model was a significantly better fit (p < 0.002) 397 than the two-parameter model and predicted an expected number of haplotypes equal to 42.4 if 146 historical samples had been used (equation: $nh = 43.97 - 0.02 e^{3.78 S}$, where nh is 398 399 expected number of haplotypes, and S is sample size). Modelling the aDNA data with the 400 Stirling probability distribution suggested that the number of historical haplotypes was 43, 401 with the lower 95% confidence interval being 28 (95% CI: 28-173, Fig. 3b).

402

403 The phylogenetic reconstruction of the combined historic (n=29) and modern data 404 (Haplotypes A-K) revealed three highly supported (>98% posterior probability) clades within 405 the woylie radiation (Fig. 4). These CR clades and haplotypes were also observed in the 406 cytochrome b phylogeny confirming topological consistency between mtDNA regions (Fig. 407 S2). Two historical haplotypes (samples MB816 and MB823) constituted a distinct clade, 408 which was not present within the extensively sampled modern populations, suggesting that 409 this particular clade is likely extinct. The other two clades consisted of an admixture of both 410 modern and historical haplotypes and spanned a broad geographical area.

411

412 Molecular clock analysis based on the cytochrome b dataset dated the most recent common 413 ancestor of the western Australian B. penicillata as living between 0.59 and 1.75 million years 414 ago (mean=1.2). This parameter was therefore included in the coalescent-based demographic 415 analysis of CR with a lognormal prior of mean 1,000,000 and SD=0.32 (in real space). Of the 416 three demographic models analysed, the BSP was the model with the best fit according to 417 differences in AICM estimates (constant population size: 2642.05, SE=0.08; exponential 418 growth: 2644.47, SE=0.08; BSP: 2572.52, SE=0.04). Based on this analysis, the woylie 419 population size for 2006-2008 was 99.86% lower than the historical population size (Fig. 5, 420 range: 99.3-99.97%). The mutation rate was estimated to be approximately 10% per million 421 years, placing the beginning of the population decline between ~4,600 and ~2,300 (95% 422 HPD) years ago.

The comparison of migration models with Migrate-n favoured the most complex model with four distinct populations (LBF=5.9 with the next most likely model; Table 1), therefore we

summed the parameter values to obtain the species-wide demographic changes over time.
Taking into account migration between populations, woylie numbers declined by a 89.1%
compared to historical population size (Fig. 6a, range: 87.1-90.7).

429

430 Microsatellite data analysis

431 We implemented several quantitative and qualitative measures to improve the reliability of 432 the microsatellite dataset. Following our strict approach, we only considered 21 historical 433 samples, from which we obtained near-complete 11-loci microsatellite profiles (Table S1). 434 From the PCR repetitions, we calculate a $P_{(fh)}$ for the locus Y112 of 36.4% and it was therefore excluded from all statistical comparisons. The remaining loci had lower P_(fh) (range: 435 436 0-16.3%, Table 2). Hence, our protocol had less than 0.5% probability of falsely scoring an 437 individual homozygous (except for the locus Y112, for which it was ~5%). Despite the 438 limited number of historical individuals with full microsatellite profiles we observed nine 439 individuals with at least one "unique allele" (i.e. alleles that were not present in modern data, 440 Table 2). Levels of heterozygosity and average allelic richness (N_{4R}) varied across loci (Table 441 2) but were significantly higher in the historical dataset (N_{AR} : Wilcoxon tests: Z=-2.191, 442 p=0.028; IH: Mann-Whitney test: Z=-2.45, p=0.014, Table 2, Fig. 3c). The three-parameter 443 asymptotic exponential model was a marginally better fit (p=0.044) to the rarefaction data 444 than the two-parameter model and predicted an expected mean number of alleles equal to 20.7 if 210 historical samples had been used (equation: $na = 20.74 - 0.88 e^{2.85 S}$, where *na* is 445 expected number of alleles, and S is sample size). The numbers of migrants were consistently 446 447 higher when compared with estimates calculated with the modern data alone (Table 3).

No evidence of a bottleneck was detected using the heterozygosity excess and the mode shift
analytical approaches. The M-ratio method, however, detected a reduction in population size
in all modern populations except for that from Perup, the eastern population in the Upper
Warren region.

453

454 Migrate-n strongly favoured a panmictic population (LBF>50,000 with the next most likely 455 model). Because the commencement of the decline is less clear in Migrate-n microsatellite 456 BSP (Fig. 6a), we calculated the extent of the decline from 2Θ because accuracy may 457 decrease beyond this point (P. Beerli personal communication) and detected a 56.5% 458 population decline in recent time.

459

460 Discussion

461 The aim of this study was to overlay data generated from historical samples on those from 462 extensively sampled modern populations (Pacioni et al. 2011), and evaluate whether 463 application of aDNA techniques enhances our understanding of the possible genetic impact of 464 declines in wildlife. Specifically, we set out to characterise ancient mitochondrial and 465 microsatellite DNA to address two key questions for conservation management, namely 466 quantifying the potential loss of genetic diversity and to assess the degree of former 467 connectivity between extant populations remaining within the species' highly fragmented 468 distribution. While modern data alone can be used to model past demographic events, it is not 469 as powerful as using direct measurements of genetic diversity based on ancient DNA. The 470 woylie exemplifies a common situation in that it is a critically endangered species where 471 genetics has been largely ignored in its management for the past 20 years, despite specific

472 recommendations in the woylie recovery plan (Start *et al.* 1995). This exclusion has prevented 473 the establishment of management practices that seek to maximise the level of genetic 474 diversity from amongst remaining populations. Lack of the genetic component also makes it 475 impossible to evaluate whether assisted migration of individuals from different populations 476 would be a viable management strategy.

477

478 *Genetic diversity*

479 The inclusion of aDNA data allowed us to conclude that the woylie underwent a dramatic 480 genetic loss. This finding is somewhat in contrast with a previous assessment based only on 481 modern data. In fact, modern levels of genetic diversity were considered relatively high 482 (Pacioni et al. 2011). However, after the addition of data from historical samples, the 483 rarefaction analysis and the estimation of the most likely number of haplotypes (n=43, 95%) CI 28-173 Fig. 3b) confirmed a mtDNA genetic loss of ~65% (range: 46-91%, considering 484 485 that only 15 haplotypes are still detectable in modern populations compared to a range of 28-486 173 expected historical haplotypes). These findings are also supported by comparisons of the 487 historical and modern microsatellite data where we found a reduction of at least 20%.

488

The use of microsatellite loci in ancient DNA presents a challenge because of the potential problem of data fidelity associated with DNA degradation (Bourke *et al.* 2010; Morin *et al.* 2001). Amplification of microsatellite data from historical skins has been limited (Bourke *et al.* 2010), and even less frequent from fossil bones (Allentoft *et al.* 2014; Nyström *et al.* 2012). In our study, the inclusion of microsatellite data from historical samples proved to be very informative, demonstrating a significant reduction in allelic richness and heterozygosity across the loci examined. Furthermore, seven of the nine individuals with unique alleles were 496 from within or close to the woylie's current geographic distribution. Considering the large 497 number of modern samples that were profiled (n=231, Pacioni et al. 2011), it is very likely 498 that these alleles have been 'lost' from the modern populations. The three-parameter 499 asymptotic exponential model predicted at least 20.7 historical alleles, suggesting a loss of at 500 least 21.3%. Our results indicate how carefully one should interpret results based on modern 501 data alone. In fact, based on the latter dataset, only limited concerns were raised on this aspect 502 of woylie conservation (Pacioni et al. 2011), while this study clearly demonstrates a dramatic 503 overall reduction in genetic diversity. Another example of the potential bias generated by the 504 lack of baseline data, which can result in misleading management targets, is that the mean 505 genetic diversity (from modern data) of the wild woylie populations considered in this study 506 was used as a reference term to evaluate whether the genetic diversity of translocated 507 populations was satisfactory (Pacioni et al. 2013). The characterisation of the historical 508 genetic diversity, as conducted in this study, indicates that, while in principle that approach is 509 valid, the conclusions are possibly rather optimistic.

510

511 It was only possible to quantify these estimations by including ancient DNA data and, 512 considering the occurrence of allelic drop out that it is inevitable when working with aDNA, 513 these should be considered conservative.

514

515 Demographic analyses

516 Data from historical samples also allowed a better understanding of the demographic history 517 of the woylie compared with that obtained from modern samples only. Specifically, the 518 historical data helped to quantify the extent of the decline and its consequences on the genetic 519 diversity. Even with modern data alone, Migrate-n correctly detected the overall woylie 520 decline. However, in these analyses, the estimated historical population size was substantially 521 reduced as was the extent of the decline (Fig. 6). Differences in the results are even more 522 dramatic in BEAST analyses. In fact, when all historical samples for which we had time-523 stamped data (n=22) were included in the mtDNA analysis, the 95% HPD intervals were 524 reduced and this analysis indicated that the woylie experienced more than a 99% decline in 525 population size in recent times. When this analysis was repeated including only modern data 526 (n=146, Fig. 5), it was not possible to discriminate between a constant population and a 527 declining trend in the demography of the woylie. Migrate-n analysis with microsatellite data 528 detected the decline even with modern data alone. However, in this analysis too, the confidence intervals are slightly wider and the population size in recent times is mildly 529 530 underestimated. These results exemplify the potential problems in reconstructing past 531 demographic trajectories from modern data alone. It is evident that the inclusion of aDNA data can be highly informative and it is likely to have important implications for the 532 533 management of threatened species.

534

535 The lack of significant results using common bottleneck tests highlights the limitations of 536 these methods. The M-ratio tests detected genetic bottleneck in three out of four remnant wild 537 populations, while the heterozygosity excess and mode-shift approaches did not detect a 538 bottleneck at all. Various factors can influence the statistical power of the bottleneck tests, 539 including the time since the bottleneck, the nature of the post-decline recovery and the 540 effective population size pre-bottleneck (Busch et al. 2007; Cornuet & Luikart 1996; Garza & 541 Williamson 2001; Williamson-Natesan 2005). We argue that the timing of the modern 542 population sampling was too close to or contemporary with the most recent declines (i.e. after 543 1999), for them to be detected (despite the substantial loss of genetic diversity) by the

544 heterozygosity excess and mode-shift methods (Cornuet & Luikart 1996; Garza & 545 Williamson 2001; Williamson-Natesan 2005). Possibly, the demographic post-reduction 546 growth, consequent upon the fox control programs started in the late 1990s, created a 547 heterozygosity deficiency, which balanced out the heterozygosity excess following the 548 bottleneck subsequent to the European settlement. On the other hand, we consider the results 549 of the M-ratio analyses indicating a genetic bottleneck to be the outcome of substantial declines in the 20th century as opposed to the most recent declines. The M-ratio approach can 550 detect a bottleneck up to 100 generations post decline (Garza & Williamson 2001), which 551 552 would correspond to ~ 250 years in the woylie (Groom 2010; Pacioni 2010). This is also 553 supported by the observation that the new alleles found in historical samples were within the 554 allele frequency distributions of the modern populations (i.e. their loss is responsible for a 555 decrease of the M-ratio).

556

557 *Gene flow*

558 With the exception of the pair Kingston-Perup, the other modern populations are, nowadays, 559 completely isolated (Pacioni et al. 2011). The coalescent-based analyses, demonstrated that 560 historically these sites were connected, a finding that was only the subject of speculation in 561 Pacioni et al. (2011). Analysis of microsatellite data suggests that these were actually part of a 562 large panmictic population. These results are also supported by the private allele analysis, 563 despite the limitations of this approach mentioned above. In this analysis, the estimates of the number of migrants between localities increased when data from the limited number of 564 565 historical samples were included. This demonstrates a wider spatial distribution of alleles, just 566 a hundred years ago (when most historical samples were collected), than that found at present.

It is intriguing to note that the last hundred years is a time consistent with the substantialhabitat loss and fragmentation in the WA Wheat-belt region (Saunders 1979).

569

The differences between the migration models supported by mtDNA and microsatellite data are not surprising. In fact, woylies have a male-biased dispersal (Christensen 1980; Pacioni 2010), with females normally settling near to their mother's home range boundaries (Pacioni 2010). Males are, therefore, primarily responsible for long-distance gene flow, which is not detected in the mtDNA data as mtDNA is maternally inherited.

575

576 *Timing of the decline*

577 We used BEAST to attempt to establish the time of commencement of the decline, because 578 BEAST can take into account uncertainty around the evolutionary rates and calibration points 579 while reconstructing the phylogeny and the demographic history. While our sampling regime 580 (which was mid way between the "pooled" and "scattered" sampling strategies) should not be 581 prone to substantial biases (Heller et al. 2013), we acknowledge a possible bias in the 582 inference of the demographic history in the BSP, because there is a strong indication of 583 reduced gene flow between different regions within Western Australia in recent times. Habitat 584 fragmentation and isolation of the studied populations generally lead to an underestimation of 585 population size (Heller et al. 2013), and consequently the extent of the demographic decline 586 may be overestimated. The effect of fragmentation on calculations of evolutionary rates is less 587 predictable. Contrary to expectations, when using the same input data in BEAST as for the 588 Migrate-n analysis (limiting data to samples geographically representative of the four modern 589 populations), BEAST estimation of historical population size appears to be more conservative (although not substantially different), and the timing of the decline is only slightly shifted 590

591 toward the present (Fig. S3). It should be stressed that such estimates are subject to 592 uncertainties because the inter-specific mutation rates are probably lower than the intra-593 specific rates (Ho et al. 2008) and, accordingly, caution the reader to interpret these results 594 with care. We argue that, while it is difficult to pinpoint exactly when the decline commenced 595 its downward trajectory, it is clear that it started recently (i.e. less than ~4,000 years ago) and 596 has continued to the present. A possible catalyst for the start of this decline could have been 597 the arrival of the dingo (Canis lupus dingo) in Australia (3,500-5,000 years ago, Fillios et al. 598 2012; Savolainen et al. 2004), but based on the observation of the substantial loss of genetic 599 diversity compared to samples from as recently as 100 years ago, we also argue that the 600 decline has been intensified since European settlement. Additional time-stamped samples that 601 increase the time span covered by the aDNA data may help to more accurately estimate the 602 timing of woylie decline, as it remains uncertain whether European factors (foxes, cats, 603 disease), pre-European factors (e.g. dingo or fire regimes), or a combination of both, were the trigger for the decline in woylie genetic diversity and population numbers. 604

605

606 Implications for woylie conservation

The findings of this study have direct implications for the conservation and management of the critically endangered woylie in south-western Australia. Notably, evidence of a much wider former connectivity between the now fragmented habitats where the modern populations reside is highly relevant. Managing small isolated populations is challenging due to their susceptibility to stochastic events and greater effects of genetic drift (Frankham 1996; Frankham *et al.* 1999). Our findings show that there is a precedent for re-establishing gene flow between the currently isolated populations. Strategies such as natural corridors and 614 translocations could help restore gene flow to historical levels and assist in the maintenance615 of the remaining genetic diversity.

616

617 Through direct measurement of ancient genetic variability, we uncovered the extent of genetic 618 diversity loss in the woylie that occurred along with a substantial demographic reduction. A 619 very conservative estimate of genetic loss would be of 65% for mtDNA and 20% for nuclear 620 DNA, but in reality the number is likely to be far greater given the rate at which new genetic 621 variants were detected in the ancient gene pool. This loss of genetic diversity can be attributed 622 directly to the massive population decline detected with the BSP analysis. The severity and 623 tempo of the decline should act as a catalyst for more active management strategies that seek 624 to maintain the existing species-wide genetic diversity. The translocated woylie populations 625 that were established in the late 1980s (e.g. populations in South Australia, Delroy et al. 626 1986), despite being at lower genetic diversity than indigenous populations (Pacioni et al. 627 2013), may have captured some of the genetic diversity that it is now lost in Western 628 Australia and may play a role in the future genetic management of the woylie. Indeed, a 629 haplotype found in South Australian translocated populations was not found in Western 630 Australian populations, but it was identified from an historical sample from Dryandra.

631

A further conservation outcome is the characterisation of woylie genetics at the boundaries of the species' distribution. For example, our profiling of fossil material from Faure Island, at the northern limit of the former distribution, demonstrated that the woylies that once inhabited this region were genetically very similar to extant populations despite being more than 1000 km away. Faure Island and Dirk Hartog Island are important sites in the Shark Bay World Heritage Area for ecosystem restoration (DEC 2005), which involves translocations of 638 previously present species to re-establish past biodiversity in the region. Our data suggest that 639 the use of the available genetic stock is appropriate and it is not unreasonable to expect that 640 woylies might adapt easily to these relatively more arid environments.

641

642 Conclusion

643 The mode and tempo of genetic diversity loss is often difficult to quantify in species that have 644 already undergone significant declines. The south-west of Western Australia is a biodiversity 645 hotspot and therefore a critical region for conservation purposes, yet baseline data for past 646 biodiversity are limited. Moreover, many marsupials are still in decline in Australia but it is 647 difficult to quantify changes in the genetic profile of these species over time. We set out to 648 investigate how knowledge of the past might assist best-practice conservation in the present. 649 The south-west of Western Australia used to provide a rich, continuous habitat to numerous, 650 now endangered species and while we focused on only one species, it is likely that the degree 651 of genetic diversity loss is directly relevant to many endemic species in the region. Whilst 652 detailed screening of available fossils and skins was conducted in the case of the woylie, the 653 number of samples available in these kinds of studies will always be limiting. The lack of 654 samples for key localities, sampling restrictions on skin and bone voucher specimens and 655 constraints imposed by DNA preservation are all limitations when trying to quantify past 656 diversity using ancient DNA techniques. Despite these limitations, this study demonstrates 657 that the inclusion of historical material in genetic studies is beneficial for wildlife management and conservation. We demonstrated that the inclusion of aDNA data improves 658 659 the quantification of genetic diversity changes and inference of historical population 660 dynamics. Additionally, we advocate that, when technically feasible, it is important to expand 661 ancient DNA studies past the use of mtDNA in an attempt to quantify what has occurred at the nuclear DNA level; both microsatellite data and SNPs (single nucleotide polymorphisms)
can provide these insights and such approaches are aided by next generation DNA sequencing
platforms. This becomes especially relevant in species where male-biased dispersal may skew
mtDNA interpretation of diversity, connectivity and phylogeography.

666

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681 References

Allentoft ME, Collins M, Harker D, et al. (2012) The half-life of DNA in bone: measuring
decay kinetics in 158 dated fossils. *Proceedings of the Royal Society of London B: Biological Sciences*, 279, 4724-4733.

Allentoft ME, Heller R, Oskam CL, et al. (2014) Extinct New Zealand megafauna were not
in decline before human colonization. Proceedings of the National Academy of
Sciences of the United States of America, 111, 4922-4927.

- Allentoft ME, Oskam C, Houston J, *et al.* (2011) Profiling the dead: generating
 microsatellite data from fossil bones of extinct megafauna—protocols, problems,
 and prospects. *PLoS One*, **6**, e16670.
- Andrén H (1994) Effects of habitat fragmentation on birds and mammals in landscapes
 with different proportions of suitable habitat: a review. *Oikos*, **71**, 355-366.
- Baele G, Li WLS, Drummond AJ, Suchard MA, Lemey P (2012) Accurate model selection
 of relaxed molecular clocks in Bayesian phylogenetics. *Molecular Biology and Evolution*, **30**, 239-243.
- Barton NH, Slatkin M (1986) A quasi-equilibrium theory of the distribution of rare
 alleles in a subdivided population. *Heredity*, 56, 409-415.

698 Beerli P (2006) Comparison of Bayesian and maximum-likelihood inference of 699 population genetic parameters. *Bioinformatics*, **22**, 341-345.

- Bourke BP, Frantz AC, Lavers CP, et al. (2010) Genetic signatures of population change in
 the British golden eagle (Aquila chrysaetos). *Conservation Genetics*, **11**, 18371846.
- Bouzat JL, Paige KN, Lewin HA (1998) The ghost of genetic diversity past: historical DNA
 analysis of the greater prairie chicken. *The American Naturalist*, **152**, 1-6.
- Brace S, Palkopoulou E, Dalén L, *et al.* (2012) Serial population extinctions in a small
 mammal indicate Late Pleistocene ecosystem instability. *Proceedings of the National Academy of Sciences of the United States of America*, **109**, 20532-20536.
- Briggs AW, Stenzel U, Johnson PLF, et al. (2007) Patterns of damage in genomic DNA
 sequences from a Neandertal. Proceedings of the National Academy of Sciences of
 the United States of America, 104, 14616-14621.
- Burbidge AA, Johnson KA, Fuller PJ, Southgate RI (1988) Aboriginal knowledge of the
 mammals of the central deserts of Australia. Australian *Wildlife Research*, 15, 939.
- Burbidge AA, McKenzie NL (1989) Patterns in the modern decline of Western Australia's
 vertebrate fauna: causes and conservation implications. *Biological Conservation*,
 50, 143-198.
- Busch JD, Waser PM, DeWoody JA (2007) Recent demographic bottlenecks are not
 accompanied by a genetic signature in banner-tailed kangaroo rats (*Dipodomys spectabilis*). *Molecular Ecology*, **16**, 2450-2462.
- Campos PF, Willerslev E, Sher A, et al. (2010) Ancient DNA analyses exclude humans as
 the driving force behind late Pleistocene musk ox (Ovibos moschatus) population
 dynamics. Proceedings of the National Academy of Sciences of the United States of
 America, 107, 5675-5680.
- Christensen PES (1980) The biology of Bettongia penicillata Gray, 1837, and Macropus *eugenii (Desmarest, 1817) in relation to fire.* Forests Deptment of Western
 Australia Bulletin, No. 91, i-ix, 1-90, Perth.
- Cornuet JM, Luikart G (1996) Description and power analysis of two tests for detecting
 recent population bottlenecks from allele frequency data. *Genetics*, 144, 2001 2014.
- DEC (2005) Shark Bay terrestrial reserves and proposed terrestrial reserve additions
 management plan. Department of Environment and Conservation.
- Delroy LB, Earl J, Radbone I, Robinson AC, Hewett M (1986) The breeding and reestablishment of the brush-tailed bettong, *Bettongia penicillata*, in SouthAustralia. Australian *Wildlife Research*, **13**, 387-396.

- Dixon CJ (2006) A means of estimating the completeness of haplotype sampling using
 the Stirling probability distribution. *Molecular Ecology Notes*, 6, 650-652.
- Dortch J (2004a) Late Quaternary vegetation change and the extinction of Black-flanked
 Rock-wallaby (Petrogale lateralis) at Tunnel Cave, southwestern Australia.
 Palaeogeography, Palaeoclimatology, Palaeoecology, **211**, 185-204.
- Dortch J (2004b) Paleo-environmental change and the persistence of human occupation in
 south-western Australian forests. British Archaeological Reports, International
 Series 1288.
- Drummond A, Rambaut A (2007) BEAST: Bayesian evolutionary analysis by sampling
 trees. *BMC Evolutionary Biology*, **7**, 214.
- Drummond AJ, Suchard MA, Xie D, Rambaut A (2012) Bayesian phylogenetics with
 BEAUti and the BEAST 1.7. *Molecular Biology and Evolution*, **29**, 1969-1973.
- Ewers RM, Didham RK (2006) Confounding factors in the detection of species responses
 to habitat fragmentation. *Biological Reviews*, **81**, 117-142.
- Fillios M, Crowther MS, Letnic M (2012) The impact of the dingo on the thylacine in
 Holocene Australia. *World Archaeology*, 44, 118-134.
- Frankham R (1995) Effective population size/adult population size ratios in wildlife: a
 review. *Genetics Research*, 66, 95-107.
- Frankham R (1996) Relationship of genetic variation to population size in wildlife.
 Conservation Biology, **10**, 1500-1508.
- Frankham R, Ballou JD, Briscoe DA (2002) *Introduction to conservation genetics.* Cambridge University Press, Cambridge.
- Frankham R, Lees K, Montgomery ME, et al. (1999) Do population size bottlenecks
 reduce evolutionary potential? *Animal Conservation*, 2, 255-260.
- Gagneux P, Boesch C, Woodruff DS (1997) Microsatellite scoring errors associated with
 noninvasive genotyping based on nuclear DNA amplified from shed hair.
 Molecular Ecology, 6, 861-868.
- Garza JC, Williamson EG (2001) Detection of reduction in population size using data
 from microsatellite loci. *Molecular Ecology*, **10**, 305-318.
- Gilbert MTP, Bandelt H-J, Hofreiter M, Barnes I (2005) Assessing ancient DNA studies.
 Trends in Ecology & Evolution, **20**, 541-544.
- Groom C (2010) Justification for continued conservation efforts following the delisting
 of a threatened species: a case study of the woylie, *Bettongia penicillata ogilbyi*(Marsupialia: Potoroidae). *Wildlife Research*, **37**, 183-193.
- Heller R, Chikhi L, Siegismund HR (2013) The confounding effect of population structure
 on Bayesian skyline plot inferences of demographic history. *PLoS One*, **8**, e62992.
- Ho SY, Saarma U, Barnett R, Haile J, Shapiro B (2008) The effect of inappropriate
 calibration: three case studies in molecular ecology. *PLoS One*, **3**, e1615.
- Ho SYW, Shapiro B (2011) Skyline-plot methods for estimating demographic history
 from nucleotide sequences. *Molecular Ecology Resources*, **11**, 423-434.
- Hunt H (2010) A temporal assessment investigating the effects of population declines on *genetic diversity, in the critically endangered woylie (Bettongia penicillata ogilbyi).*Honours thesis, Murdoch University.
- Kalinowski ST (2004) Counting alleles with rarefaction: private alleles and hierarchical
 sampling designs. *Conservation Genetics*, 5, 539-543.

- Kalinowski ST (2006) hw-quickcheck: an easy-to-use computer program for checking
 genotypes for agreement with Hardy-Weinberg expectations. *Molecular Ecology Notes*, 6, 974-979.
- Leonard JA (2008) Ancient DNA applications for wildlife conservation. *Molecular Ecology*, **17**, 4186-4196.
- Librado P, Rozas J (2009) DnaSP v5: a software for comprehensive analysis of DNA
 polymorphism data. *Bioinformatics*, 25, 1451-1452.
- Luikart G, Cornuet JM (1998) Empirical evaluation of a test for identifying recently
 bottlenecked populations from allele frequency data. *Conservation Biology*, 12,
 228-237.
- Morin P, Chambers K, Boesch C, Vigilant L (2001) Quantitative polymerase chain
 reaction analysis of DNA from noninvasive samples for accurate microsatellite
 genotyping of wild chimpanzees (Pan troglodytes verus). *Molecular Ecology*, 10, 1835-1844.
- Myers N, Mittermeier RA, Mittermeier CG, da Fonseca GAB, Kent J (2000) Biodiversity
 hotspots for conservation priorities. *Nature*, 403, 853-858.
- Navascués M, Depaulis F, Emerson BC (2010) Combining contemporary and ancient
 DNA in population genetic and phylogeographical studies. *Molecular Ecology Resources*, 10, 760-772.
- Neal RM (2003) Slice sampling. *Annals of statistics*, **31**, 705-741.
- 800 Nei M (1987) *Molecular evolutionary genetics.* Columbia University Press, New York.
- Nyström V, Humphrey J, Skoglund P, *et al.* (2012) Microsatellite genotyping reveals end
 Pleistocene decline in mammoth autosomal genetic variation. *Molecular Ecology*,
 21, 3391-3402.
- 804 Oksanen J, Blanchet GF, Kindt R, *et al.* (2013) Vegan: Community Ecology Package. R
 805 package version 2.0-8.
- Pacioni C (2010) The population and epidemiological dynamics associated with recent
 decline of woylies (Bettongia penicillata) in Australia. PhD, Murdoch University.
- Pacioni C, Spencer P (2010) Capturing genetic information using non-target species
 markers in a species that has undergone a population crash. *Australian Mammalogy*, **32**, 33-38.
- Pacioni C, Wayne AF, Spencer P (2013) Genetic outcomes from the translocations of the
 critically endangered woylie. *Current Zoology*, **59**, 294-310.
- Pacioni C, Wayne AF, Spencer PBS (2011) Effects of habitat fragmentation on population
 structure and long distance gene flow in an endangered marsupial: the woylie. *Journal of Zoology*, 283, 98-107.
- Peakall ROD, Smouse PE (2006) GENALEX 6: genetic analysis in Excel. Population
 genetic software for teaching and research. *Molecular Ecology Notes*, 6, 288-295.
- Peery MZ, Kirby R, Reid BN, *et al.* (2012) Reliability of genetic bottleneck tests for detecting recent population declines. *Molecular Ecology*, **21**, 3403-3418.
- Piry S, Luikart G, Cornuet JM (1999) Computer note. BOTTLENECK: a computer program
 for detecting recent reductions in the effective population size using allele
 frequency data. *Journal of Heredity*, **90**, 502-503.
- Posada D, Crandall KA (1998) MODELTEST: testing the model of DNA substitution. *Bioinformatics*, 14, 817-818.

- Prost S, Anderson CNK (2011) TempNet: a method to display statistical parsimony
 networks for heterochronous DNA sequence data. *Methods in Ecology and Evolution*, 2, 663-667.
- R DevelopmentCore Team (2015) R: A language and environment for statistical
 computing. R Foundation for statistical computing, Vienna, Austria. ISBN 3900051-07-0, Vienna.
- Ramakrishnan UMA, Hadly EA (2009) Using phylochronology to reveal cryptic
 population histories: review and synthesis of 29 ancient DNA studies. *Molecular Ecology*, 18, 1310-1330.
- 834 Rambaut A, Drummond AJ (2007) TRACER. http://beast.bio.ed.ac.uk/Tracer
- Rioux Paquette S (2011) PopGenKit: useful functions for (batch) file conversion and data
 resampling in microsatellite datasets. R package, version 1.0.
- Rousset F (2008) GENEPOP'007: a complete re-implementation of the GENEPOP
 software for Windows and Linux. *Molecular Ecology Resources*, **8**, 103-106.
- Saunders D (1979) The availability of tree hollows for use as nest sites by White-tailed
 Black Cockatoos. Australian *Wildlife Research*, 6, 205-216.
- Savolainen P, Leitner T, Wilton AN, Matisoo-Smith E, Lundeberg J (2004) A detailed
 picture of the origin of the Australian dingo, obtained from the study of
 mitochondrial DNA. *Proceedings of the National Academy of Sciences of the United*States of America, **101**, 12387-12390.
- Shapiro B, Drummond AJ, Rambaut A, *et al.* (2004) Rise and fall of the Beringian steppe
 bison. *Science*, **306**, 1561-1565.
- Start AN, Burbidge AA, Armstrong D (1998) A review of the conservation status of the
 woylie, *Bettongia penicillata ogilbyi* (Marsupialia: Potoroidae) using IUCN
 criteria. *CalmScience*, 2, 277-289.
- Start T, Burbidge AA, Armstrong D, Woylie Recovery Team (1995) *Woylie recovery plan*,
 2nd edn. Department of Conservation and Land Management, Como.
- Taylor AC, Sherwin WB, Wayne RK (1994) Genetic variation of microsatellite loci in a
 bottlenecked species: the northern hairy-nosed wombat Lasiorhinus krefftii. *Molecular Ecology*, 3, 277-290.
- 855 Van Dyck S, Strahan R (2008) *The mammals of Australia*, 3rd edn. New Holland
 856 Publishers, Sydney.
- Wayne AF, Friend T, Burbidge AA, Morris K, Van Weenen J (2008a) *Bettongia penicillata*www.iucnredlist.org
- Wayne AF, Maxwell M, Ward C, *et al.* (2013) The importance of getting the numbers
 right: quantifying the rapid and substantial decline of an abundant marsupial, *Bettongia penicillata. Wildlife Research*, 40, 169-183.
- Wayne AF, Maxwell M, Ward C, *et al.* (2015) Suddenand rapid decline of the abundant
 marsupial *Bettongia penicillata* in Australia. *Oryx*, **49**, 175-185.
- Wayne AF, Ward C, Maxwell M, *et al.* (2008b) Diagnosing the recent collapse of the
 woylie in southwestern Australia, 21st Australian Wildlife Management Society
 Conference, p.140.
- Westerman M, Loke S, Springer MS (2004) Molecular phylogenetic relationships of two
 extinct potoroid marsupials, *Potorous platyops* and *Caloprymnus campestris*(Potoroinae: Marsupialia). *Molecular Phylogenetics and Evolution*, **31**, 476-485.
- Willerslev E, Cooper A (2005) Review paper. ancient dna. *Proceedings of the Royal*Society of London B: Biological Sciences, 272, 3-16.

- Williamson-Natesan EG (2005) Comparison of methods for detecting bottlenecks from
 microsatellite loci. *Conservation Genetics*, 6, 551-562.
- Woinarski JCZ, Burbidge AA, Harrison PL (2015) Ongoing unraveling of a continental
 fauna: Decline and extinction of Australian mammals since European settlement. *Proceedings of the National Academy of Sciences of the United States of America*,
 112, 4531-4540.
- 878
- 879 Data Accessibility:
- 880 DNA sequences: Genbank accession numbers KT876917-KT876973
- 881 Microsatellite data, and BEAST and Migrate-n input and parameter files: Dryad Digital
- 882 Repository: doi:10.5061/dryad.26s85
- 883

884 Author contributions:

- 885 CP and MB conceived the experiments. AW, AB and JD assisted with sample selection,
- dating and conservation implications. CP, HH, MA, DH and TV generated and analysed the
- data and prepared figures/tables. CP, HH, MA and MB wrote the paper with input from all
- 888 co-authors.
- 889
- 890
- 891

893 Figure Legends

894 Fig. 1. Localities of historical woylie fossil and museum specimens sampled in south-west 895 Western Australia (WRS: Witchcliffe Rock Shelter). Modern populations are indicated in 896 capital letters within grey squares: UW=Upper Warren, D=Dryandra, T=Tutanning). Grey 897 area in the insert: former Bettongia penicillata distribution including both subspecies 898 Bettongia penicillata ogilbyi and B. p. penicillata (adapted from Wayne et al. 2008b) based 899 on contemporary, fossils and Aboriginal knowledge records (Burbidge et al. 1988). Images of 900 the woylie (B. p. ogilbyi: (a) an adult woylie fitted with a VHF radio collar (Photo: John 901 Lawson), (b) skin and (c) fossil bone remains).

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903 Fig. 2. Multidimensional parsimony network of mtDNA sequences from historical and 904 modern woylie samples (after removal of sequences with ambiguities). Dimensions of the 905 circles are proportional to the number of sequences for each haplotype. Margaret River is a 906 region that includes samples from WRS, Mammoth Cave and Tunnel Cave.

907

Fig. 3. a. and c. Rarefaction curves of modern and historical woylie data for (mtDNA)
haplotype richness and (microsatellite) allelic richness, respectively. Error bars indicate SE
from re-sampling iterations. b. Posterior probability distribution for expected total number of
haplotypes based on observed number of haplotypes and sampling effort in the historical
dataset.

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Fig. 4. Bayesian phylogeny of *Bettoniga penicillata ogilbyi*, showing the maximum *a posteriori* (MAP) tree generated using BEAST. A GTR+G+I substitution model was imposed on an alignment containing the 29 historical CR sequences (~400bp) and 15 modern haplotype sequences (identified with an *) from Pacioni *et al.* (2011) detected in 152 modern samples (~600bp). Posterior probabilities are shown on the selected outer nodes. The scale represents the number of nucleotide substitutions per site. Three distinct clades are indicated on the tree. The tree topology is mirrored at the cytochrome *b* locus (see Fig. S2).

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Fig. 5. Reconstruction of the woylie (*B. penicillata ogilbyi*) demographic history based on
control region of mtDNA using a Bayesian Skyline Plot analysis implemented in BEAST
v1.8 (Drummond *et al.* 2012). Shaded areas represent 95% Highest Probability Density
(HPD). Solid lines represent the medians. Population size is expressed as log of the effective
population size times the generation times.

927

Fig. 6. Reconstruction of the woylie (*B. penicillata ogilbyi*) demographic history based on
control region of mtDNA (a) and microsatellite data (b) using a Bayesian Skyline Plot
analysis implemented in Migrate-n v3.6.8 (Beerli 2006). In red analysis with modern data
alone and in blue analysis including modern and ancient DNA data. Shaded areas represent
1.96 standard deviations from parameter values.

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Tables

 Table 1. Mode (2.5 and 97.5 percentile) of number of migrants between woylie

	Kingston	Perup	Tutanning	Dryandra
Kingston		1.1 (0-6.1)	n/a	4.8 (0-23.5)
Perup	0.3 (0.1-0.5)		n/a	5.1 (0-23.5)
Tutanning	n/a	n/a		0.1 (0-12.9)
Dryandra	0.1 (0.002-0.5)	2 (0.2-6.8)	0.5 (0-10.3)	

populations (from row to column)

Table 2. Descriptive statistics of genetic diversity of microsatellite loci in the woylie

 (*Bettongia penicillata ogilbyi*) for modern (n=231)/historical (n=21) datasets. Between

 brackets SE.

Locus	H_{o} (%)	N_A	N_{AR}	U_A	$P_{(fh)}$
Bt76	85.3/85.7	19/14	10.5/11.57	2	2.2
Bt64	73.8/100	24/19	11.08/14.29	2	9.4
T17-2	68.7/90.5	16/14	10.07/11.13	1	6.5
Y175	82.2/78.9	19/13	10.57/10.5	0	14.3
Pa593	83.5/88.9	16/13	10.36/11.27	1	8
Y105	40.9/23.1	11/5	5.4/4.4	0	0
Y112 [*]	73.7/90	29/11	12.46/11	1	36.4
Pl2	70.6/68.8	10/8	6.69/6.82	1	5.8
Y170	81/90	18/14	10.51/11.56	0	16.3
Bt80	76.9/58.3	12/10	7.92/9.54	1	8.5
Pl26	42.1/54.5	5/4	3.46/4	0	0

Mean 70.8(0.5)/75.3(0.7) 16.3(2.0)/11.4(1.3) 9.0(0.8)/9.6(1) 0.8(0.2) 9.8(3)

 H_o , observed heterozygosity; N_A , number of alleles; N_{AR} , allelic richness; U_A , historical unique alleles; $P_{(fh)}$, frequencies of falsely scoring an individual as homozygous (%). * Excluded from statistical tests because of evidence of substantial allelic dropout **Table 3.** Number of migrants per generation between localities where modern woyliepopulations are located, calculated with the private allele method using 11 microsatellite loci(bold: ancient and modern DNA data, normal font: modern DNA data alone).

	Upper Warren	Tutanning	Dryandra
Upper Warren		0.52	0.91
Tutanning	0.69		0.68
Dryandra	1.25	0.73	













853 3	852	851	850	849	848	846	845	844	843	842	841	840	838	837	836	835	834	833	832	830	829	828	827	826	825	824	823	822	Sampl ID
																													e C ¹⁴ lab codes
Manjimup	Faure Island	Bernier Is., uncertain	Dryandra	Boyup Brook	Manjimup	Cuballing	Contine	location	Karridale	Karridale	Karridale	Mooterdine	Mooterdine	Contine	Pingelly	Pingelly/ Woyerling	Beverley	Chorkerup	Capel/Uralla	Mammoth Cave	Mammoth Cave	Beverley	Jerramungup	Gracefield	Dwarladine Pool	Beaufort	Dwarladine Pool	Pingelly	Location
																													Uncalibrate Age (C ¹⁴ error)
32	n/a	96	37	41	32	47	44	66	80	80	78	44	44	44	78	49	77	72	79	96	96	77	41	108	102	108	102	102	Age (year*)
Bone	Bone	Bone	Bone	Bone	Bone	Skin	Skin	Skin	Skin	Skin	Skin	Skin	Skin	Skin	Skin	Skin	Skin	Skin	Skin	Skin	Skin	Skin	Skin	Skin	Skin	Skin	Skin	Skin	Sample
27.25	29.58	31.57	25.26	29.4	17.33	16.84	18.59	28.78	32.29	ı	I	22.87	17.69	14.83	I	32.32	ı	ı	32.57	29.15	24.2	ı	19.26	26.41	34.38	25.14	29.2	I	Ct value
293	123	253	293	253	293	293	293	141	253	I	ı	293	293	293	I	293	293	I	141	293	293	I	293	253	141	253	293	I	Largest Amplicon
10			7	8	8	10	9					10	10	10						4	10			7			8		Msat typed
<			<		<	<	<					<	<	<		<				<	<		<				<		BEAST BSP
<			<		<	<	<						<	<		<							<				<		Migrate-n BSP (mtDNA)
<			<	<	<	<	<					<	<	<										<			<		Migrate-n BSP (msats)

Sample ID	C ¹⁴ lab codes	Location	Uncalibrate Age (C ¹⁴ error)	Age (year*)	Sample	Ct value	Largest Amplicon	Msat typed	BEAST BSP	Migrate-n BSP (mtDNA)	Migrate-n BSP (msats)
854		Perup River		30	Bone	28.88	293		\checkmark	\checkmark	
855		Manjimup		32	Bone	23	293	10	<	<	<
1003		Peron Peninsula		n/a	Bone	I	293				
1004		Faure Island		n/a	Bone	40.33	ı				
1005		Faure Island		n/a	Bone	I	253				
1006		Faure Island		n/a	Bone	29.8	293				
1007		Nullarbor (cave)		n/a	Bone	I	ı				
1008		Nail Tail Cave		n/a	Bone	35.37	ı				
1009		Nullarbor (cave)		n/a	Bone	I	·				
1010		Nullarbor (cave)		n/a	Bone	42.95	253				
1011		Peak Charles N.P.		n/a	Bone	I	ı				
* Years prior	to 2008. Arch	aeological dates (i.e	e. more than 150 years) are b	based on calibrated radio	carbon age est	timates of y	ears before 195	50. Betwee	n brackets ra	nge with \pm s.e.	

† Sample is positioned in an undated layer between two dated layers

Calibrations were carried out using CalPal2007_HULU calibration curve of the Cologne Radiocarbon Calibration & Paleoclimate Research Package, available at http://www.calpal-online.de/