

1 **Calibration of pelagic stereo-BRUVs and scientific longline surveys for sampling sharks**

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25 **Abstract**

26 **1.** Our understanding of the ecology of sharks and other highly mobile marine species often relies on  
27 fishery-dependent data or extractive fishery-independent techniques that can result in catchability and  
28 size-selectivity biases. Pelagic Baited Remote Underwater stereo-Video systems (pelagic stereo-  
29 BRUVs) provide a standardised, non-destructive and fishery-independent approach to estimate  
30 biodiversity measures of fish assemblages in the water column. However, the performance of this  
31 novel method has not yet been assessed relative to other standard sampling techniques.

32 **2.** We compared the catch composition, relative abundance and length distribution of fish assemblages  
33 sampled using pelagic stereo-BRUVs and conventional scientific longline surveys. In particular, we  
34 focused on sharks of the family Carcharhinidae (requiem sharks) to assess the sampling effectiveness  
35 of this novel technique along a latitudinal gradient off the coast of Western Australia. We calibrated  
36 the sampling effort required for each technique to obtain equivalent samples of the target species and  
37 discuss the advantages, limitations and potential use of these methods to study highly mobile species.

38 **3.** The proportion of sharks sampled by pelagic stereo-BRUVs and scientific longline surveys was  
39 comparable across the latitudinal gradient. *Carcharhinus plumbeus* was the most abundant species  
40 sampled by both techniques. Longline surveys selected larger individuals of the family  
41 Carcharhinidae in comparison to the length distribution data obtained from pelagic stereo-BRUVs.  
42 However, the relative abundance estimates (catch per unit of effort) from the pelagic stereo-BRUVs  
43 were comparable to those from 5 to 30 longline hooks.

44 **4.** Pelagic stereo-BRUVs can be calibrated to standard techniques in order to study the species  
45 composition, behaviour, relative abundance and size distribution of highly mobile fish assemblages at  
46 broad spatial and temporal scales. This technique offers a non-destructive fishery-independent  
47 approach that can be implemented in areas that may be closed to fishing, and is suitable for studies on  
48 rare or threatened species.

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50 **Keywords:** Baited Remote Underwater Video, fishery-independent, method comparison, effort  
51 equivalence, gear selectivity, Carcharhinidae, pelagic fish, mid-water, behaviour, marine ecology

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## 70 **Introduction**

71 Emerging technologies are providing new options for cost-effective ecological sampling. These  
72 technical advances dramatically increase the opportunities for *in situ* ecological and behavioural  
73 research in vast and remote environments such as the open ocean (Murphy & Jenkins 2010).  
74 However, in order to understand the potential of novel techniques it is necessary to compare and  
75 calibrate them against traditional methods.

76 Studying the ecology and assessing the status of sharks is challenging due to their generally high  
77 mobility, ontogenetic shifts in habitat preference and broad geographic range (Dulvy *et al.* 2008). Our  
78 understanding on the biology and ecology of sharks and other highly mobile marine species largely  
79 relies on fishery-dependent data from commercial and recreational fisheries (Myers & Worm 2003).  
80 The use of fishery-dependent data alone can lead to sampling biases due to gear selectivity and  
81 heterogeneous fishing effort that discriminate among species and habitats (Simpfendorfer *et al.* 2002;  
82 Murphy & Jenkins 2010). Alternatively, fishery-independent surveys use more robust sampling  
83 designs, but often employ the same commercial fishing gear (e.g. longlines, gillnets, trawls) and as  
84 such, catchability and size-selectivity biases remain (McAuley, Simpfendorfer & Wright 2007).

85 Scientific longline surveys are among the most commonly used fishery-independent methods for  
86 studying the demography and ecology of shark populations (Simpfendorfer *et al.* 2002). These  
87 surveys provide measures of relative abundance, sex ratio and length distribution of a range of shark  
88 species (McAuley *et al.* 2007). Additionally, longlines allow the collection of samples for population  
89 biology studies (e.g. genetics, age, growth, reproduction, diet) and the deployment of conventional  
90 and electronic tags (Meyer, Papastamatiou & Holland 2010). However, in order to obtain length or  
91 biomass data from longline surveys, sharks must be caught, retrieved and handled out of the water.  
92 The handling of sharks aboard research vessels aims to maximise survival of individuals, but all  
93 captured fish are exposed to varying degrees of physiological stress and physical trauma that can  
94 induce pre- or post-release mortality (Skomal 2007). Consequently, these extractive techniques may

95 not be suitable for sampling rare or threatened species, or used in areas that are closed to fishing  
96 (Murphy & Jenkins 2010).

97 Baited Remote Underwater Video systems (BRUVs) provide an alternative standardised, non-  
98 extractive and fishery-independent approach that is widely used to estimate biodiversity indices and  
99 relative abundance measures of a range of marine species (Cappo *et al.* 2003; Langlois *et al.* 2012b;  
100 Santana-Garcon *et al.* 2014), including sharks (Brooks *et al.* 2011; Goetze & Fullwood 2013; White *et*  
101 *al.* 2013). This technique uses bait to attract individuals into the field of view of a camera so that  
102 species can be identified and individuals counted (Dorman, Harvey & Newman 2012). When stereo-  
103 camera pairs are used, precise length measurements can be made and biomass estimated (Harvey *et al.*  
104 2010). Pelagic stereo-BRUVs are a novel method that sets camera systems at a predetermined depth  
105 in the water column as opposed to the commonly used benthic deployment where stereo-BRUVs are  
106 set on the seafloor. This deployment design allows pelagic stereo-BRUVs to estimate the  
107 composition, relative abundance and length distribution of fish assemblages that inhabit the water  
108 column (Heagney *et al.* 2007; Santana-Garcon, Newman & Harvey 2014).

109 Methodological comparisons assist in validating the utility of innovative sampling methods and to  
110 understand the advantages and limitations of different techniques. The use of benthic BRUVs to  
111 survey demersal fish assemblages has been compared to scientific fishing surveys including trawl  
112 (Cappo, Speare & De'ath 2004), trap (Harvey *et al.* 2012b), hook and line (Langlois *et al.* 2012a), and  
113 longline (Ellis & Demartini 1995). Additionally, benthic BRUVs and scientific longline surveys were  
114 compared to estimate the diversity and relative abundance of sharks in the Bahamas (Brooks *et al.*  
115 2011). These studies found that the species composition determined with baited video techniques  
116 differed to the catch of trawls (Cappo, Speare & De'ath 2004) but it was comparable to some extent to  
117 the catch of traps and longlines (Brooks *et al.* 2011; Harvey *et al.* 2012b). Estimates of relative  
118 abundance differed among techniques, especially for rare species (Brooks *et al.* 2011; Harvey *et al.*  
119 2012b). However, differences in length distribution of species taken in traps and hooks or recorded on  
120 stereo-BRUVs were not biologically significant (Langlois *et al.* 2012a). Pelagic stereo-BRUVs have

121 been used to assess pelagic fish assemblages (Santana-Garcon, Newman & Harvey 2014) but their  
122 performance has not been assessed relative to other sampling techniques.

123 The present study aims to compare the catch composition, relative abundance and length distribution  
124 of fish sampled by pelagic stereo-BRUVs and conventional scientific longlines. The sampling effort  
125 required for each technique to obtain equivalent relative abundance samples is determined for each of  
126 the target species. In particular, surveys were conducted along a latitudinal gradient off the coast of  
127 Western Australia to target sharks of the family Carcharhinidae, commonly known as requiem sharks.  
128 Given that longlines used large hooks to target sharks, we hypothesised that the methods would differ  
129 in total catch composition with pelagic stereo-BRUVs providing data on a broader range of species.  
130 However, we expect that both methods would generate comparable estimates of relative abundance  
131 and length distribution for the targeted shark species. Finally, the advantages and limitations of  
132 pelagic stereo-BRUVs and scientific longline surveys to study highly mobile species are discussed.

## 133 **Methods**

### 134 SAMPLING TECHNIQUE

135 We conducted a longline and pelagic stereo-BRUVs survey in August 2012 at 10 sites over 950 km  
136 along the coastline of Western Australia (Fig. 1). Sites were 15 to 80 km offshore at depths ranging  
137 between 35 and 106 metres, although most sites were 40 to 60 metres deep. Data were recorded from  
138 31 pelagic stereo-BRUVs and 31 scientific longline deployments targeting requiem sharks. Three  
139 replicate deployments of each method were conducted simultaneously at each site, with the exception  
140 of one site in the Houtman Abrolhos Islands where four replicates of each technique were undertaken.  
141 The number of replicates for each method was limited by logistical constraints of this research  
142 expedition. During deployment, both methods were interspersed following a straight line with a  
143 separation of at least one kilometre between deployments of either method to avoid or minimize  
144 potential overlap of bait plumes and, to reduce the likelihood of fish moving between replicates.

#### 145 *Scientific longline surveys*

146 Scientific longline surveys were conducted as part of the annual shark monitoring and tagging  
147 program of the Department of Fisheries (Western Australia). Surveys were designed to target requiem  
148 sharks. The longlines were 500 m in length and comprised ~50 J-shaped hooks of size 12/0 baited  
149 with Sea Mullet (*Mugil cephalus*; half a fish per hook) and attached to the main line via 2 m metal  
150 snoods (Fig. 2a). Lines were designed to hold hooks approximately 8 metres above the seafloor.  
151 However, hooks near the ballast remained closer to the bottom; this was confirmed during retrieval as  
152 fragments of benthos such as sponges were occasionally caught (Santana-Garcon *pers. obs.*).  
153 Longlines were set before dawn at ~5 am and soak time ranged between 2.5 and 6 hours, depending  
154 on the time required for retrieval and processing of the catch. Upon retrieval, all individuals caught  
155 were identified to the species level and their fork length (FL) was measured. Catch per unit of effort  
156 (CPUE), a measure of abundance where catch is standardised across deployments of different  
157 sampling effort, was calculated as the catch of each longline divided by the soak time (hours) and the  
158 number of hooks used. In the present study, we defined CPUE10 as the catch per hour per 10 hooks as  
159 a measure to facilitate comparison between methods (the number of hooks was chosen on the basis of  
160 results presented herein).

#### 161 *Pelagic stereo-BRUVs*

162 Pelagic stereo-BRUVs were adapted to match the deployment characteristics of longlines so that both  
163 methods sampled at similar depths, for equal periods of time and using the same bait. During the  
164 deployment of pelagic stereo-BRUVs, cameras were placed in the mid-water, approximately 8 to 10  
165 metres above the bottom (Fig. 2b). This technique uses ballast and sub-surface floats in order to  
166 anchor the systems, enabling control over the deployment depth and reducing movement from surface  
167 waves (Santana-Garcon, Newman & Harvey 2014). The camera systems consisted of two Sony CX12  
168 high definition digital cameras mounted 0.7 m apart on a steel frame and converged inwards at 8  
169 degrees to allow the measurement of fish length (Harvey *et al.* 2010). The bait consisted of 1 kg of  
170 mullet (*Mugil cephalus*; fish cut in halves) in a wire mesh basket suspended 1.2 m in front of the  
171 cameras. As for longlines, camera deployments were set before dawn, at 5 am in the morning, and  
172 soak time ranged between 2.5 and 6 hours depending on the time required for longline retrieval.

173 Videos were analysed for the full length of the deployment. A blue light (wavelength 450-465 nm)  
174 was fitted on the frame, between the cameras, in order to illuminate the field of view during the  
175 sampling hours before dawn. Blue light wavelength is thought to be below the spectral sensitivity  
176 range for many fish species (Von Der Emde, Mogdans & Kapoor 2004), and therefore, it is expected  
177 to have minimal impact on fish behaviour (Harvey *et al.* 2012a).

## 178 VIDEO ANALYSIS

179 Stereo-camera pairs were calibrated before and after the field campaign using CAL software (SeaGIS  
180 Pty Ltd) following Harvey & Shortis (1998). The video images obtained from pelagic stereo-BRUVs  
181 were analysed using the software 'EventMeasure (Stereo)' (SeaGIS Pty Ltd). All fish observed were  
182 quantified, identified to the lowest taxonomic level possible and measured. However, for this study,  
183 small pelagic fish species in the family Clupeidae and small carangids from the genus *Decapterus*,  
184 *Selar* and *Selaroides* among others were excluded from the analysis. A conservative measure of  
185 relative abundance, MaxN, was recorded as the maximum number of individuals of the same species  
186 appearing in a frame at the same time. MaxN avoids repeat counts of individual fish re-entering the  
187 field of view (Priede *et al.* 1994; Cappo *et al.* 2003). MaxN per hour was used in order to standardise  
188 sampling effort across all deployments due to variable soak times. Length measurements (FL) were  
189 made from the stereo-video imagery for each individual within 7 metres of the camera system  
190 recorded at the time of MaxN. Individuals must be measured when their body is straight which can be  
191 difficult for sharks given their swimming behaviour, as such, in order to improve the accuracy of  
192 shark measurements, the length of each individual was determined from an average of five  
193 measurements obtained in different video frames (Harvey, Fletcher & Shortis 2001).

## 194 STATISTICAL ANALYSIS

### 195 *Comparison of catch composition*

196 Differences in species composition between scientific longlines and pelagic stereo-BRUVs were  
197 tested using one-way univariate permutational analysis of variance (PERMANOVA; Anderson,  
198 Gorley & Clarke 2008). Proportional data facilitates the comparison of composition patterns sampled



199 by each method as it standardises all samples to the same scale (Jackson 1997). Hence, for each of the  
200 five species of requiem sharks recorded, we used proportional data to emphasise the contribution of  
201 each species to the total number of individuals caught per deployment and method. Proportional data  
202 were calculated from CPUE data across all replicates and were arcsine transformed to normalise  
203 possible binomial distributions (Zar 1999). Euclidean distance was used to generate the dissimilarity  
204 matrices (Anderson *et al.* 2011), P-values were obtained using permutation tests (9,999 permutations)  
205 for each individual term in the model and, Monte Carlo p-values were used to interpret the result  
206 when the number of unique permutations was less than 100 (Anderson 2001). Data manipulation and  
207 graphs across the study were undertaken using the packages ‘reshape2’(Wickham 2007), ‘plyr’  
208 (Wickham 2011) and ‘ggplot2’(Wickham 2009) in R (R Core Team 2013).

#### 209 *Catch comparison along a latitudinal gradient*

210 The ability of the two methods to describe spatial patterns along a latitudinal gradient (32° to 24° S)  
211 was compared. For each of the target species, a one-way analysis of covariance (ANCOVA) was  
212 conducted with method as a factor and latitude as a covariate. A significant interaction between  
213 latitude and method would indicate that the methods were not comparable across the latitudinal range.  
214 Analyses were based on Euclidean distance resemblance matrices calculated from arc sine  
215 transformed proportional data. Statistical significance was tested using 9,999 permutations of  
216 residuals under a reduced model.

#### 217 *Equivalence of sampling effort*

218 For the target species, the equivalent longline and pelagic stereo-BRUVs sampling effort was  
219 determined by performing a series of statistical tests on the abundance estimates obtained from  
220 BRUVs (MaxN per hour) and from a range of longline effort data sets (1-50 hooks). Random samples  
221 of our data were taken with replacement and the differences between methods were tested using  
222 univariate PERMANOVAs based on Euclidean distance resemblance matrices of the raw CPUE data,  
223 with method as a fixed factor (Anderson *et al.* 2011). P-values were obtained from 9,999 permutations  
224 using the ‘adonis’ function from the ‘vegan’ package (Oksanen *et al.* 2013) in R. This process was

225 bootstrapped (1000 times) to generate a distribution of p-values across sampling efforts for the target  
226 species.

227 Additionally, we compared the sampling precision of both techniques at the family level and for each  
228 target species. The precision of a sampling method refers to the repeatability of its measurements  
229 under unchanged conditions, it can be expressed numerically by measures of imprecision like  
230 standard deviation, variance and most commonly, as a ratio of the standard error (SE) and the mean  
231 (Andrew & Mapstone 1987). Here, we estimated precision ( $p$ ) as  $p = SE / Mean$ , where the mean and  
232 standard error were obtained from the abundance per deployment for each sampling technique.

### 233 *Comparison of length distributions*

234 For the target species, length distributions obtained from pelagic stereo-BRUVs and longline surveys  
235 were compared using kernel density estimates (KDE). The KDE method is sensitive to differences in  
236 both the shape and location of length distributions (Sheather & Jones 1991). KDE analyses were  
237 conducted using the R packages ‘KernSmooth’ (Wand 2013) and ‘sm’ (Bowman & Azzalini 2013)  
238 following the method described by Langlois et al. (2012a). For each species, the statistical analysis  
239 between the pairs of length distributions collected by each method was based on the null model of no  
240 difference and a resulting permutation test. The statistical test compared the area between the KDEs  
241 for each method to that resulting from permutations of the data into random pairs. To construct the  
242 test, the geometric mean between the bandwidths for stereo-BRUVs and longline data were calculated  
243 (Bowman & Azzalini 1997). If the data from both methods have the same distribution, the KDEs  
244 should only differ in minor ways due to within population variance and sampling effects (Langlois *et*  
245 *al.* 2012a). The ‘sm.density.compare’ function in the ‘sm’ package was used to plot the length  
246 distributions where the resulting grey band shows the null model of no difference between the pair of  
247 KDEs.

## 248 **Results**

### 249 *Comparison of catch composition*

250 Scientific longline surveys used 1,671 baited hooks (125 hours) and caught 236 individuals of 18  
251 different species. Pelagic stereo-BRUVs recorded 123 hours of video in 31 deployments with a total  
252 of 124 individuals of 20 species identified. The numerous small pelagic fish (TL < 250 mm) observed  
253 in the video were not included in the species count or in the analyses. Teleost species were almost  
254 exclusively sampled by pelagic stereo-BRUVs, while the semi-pelagic sharks were sampled by both  
255 methods (Fig. 3). Due to the deployment design of longlines, a proportion of the hooks adjacent to the  
256 ballast were set close to the bottom and, consequently, benthic sharks were almost exclusively  
257 sampled by this method.

258 The target shark species caught included sandbar *Carcharhinus plumbeus*, tiger *Galeocerdo cuvier*,  
259 blacktip *C. limbatus/tilstoni* and milk *Rhizoprionodon acutus* sharks, and *Carcharhinus spp\**. The  
260 latter combines four requiem species that could not be confidently distinguished across all videos  
261 (bronze whaler *C. brachyurus*, dusky *C. obscurus*, spinner *C. brevipinna* and spot-tail *C. sorrah* sharks).  
262 The common blacktip *C. limbatus* and the Australian blacktip *C. tilstoni* sharks are also combined  
263 here as there are no external morphological features that distinguish these species (Harry *et al.* 2012).  
264 For each of the target species, there was no significant difference in the proportion sampled by either  
265 method ( $P > 0.05$ ; Fig. 4). The sandbar shark *C. plumbeus* was the most abundant species for both  
266 methods followed by *Carcharhinus spp\** (Figs. 3 and 4). Using pelagic stereo-BRUVs, the third and  
267 fourth most abundant species were *G. cuvier* and *C. limbatus/tilstoni*, whereas with longlines these  
268 species were the fourth and third most abundant respectively. *Rhizoprionodon acutus* was the least  
269 abundant and it was rarely recorded by either method.

#### 270 *Catch comparison along a latitudinal gradient*

271 Pelagic stereo-BRUVs and scientific longline surveys showed similar patterns of abundance for all  
272 target species along the 950 km latitudinal gradient (Fig. 5). The ANCOVA revealed no significant  
273 interaction between method and latitude for any of the target species ( $P > 0.05$ ). The species  
274 proportional abundance did not differ significantly between methods whereas latitude had a  
275 significant effect on the distribution of *C. plumbeus* with a greater abundance present in the northern

276 sites (Table 1). *Rhizoprionodon acutus* and *C. limbatus/tilstoni* also showed a significant effect of  
277 latitude in their distribution as they were only recorded north of Shark Bay, the most northern  
278 sampling sites (~24° S). Although strong patterns were apparent for *G. cuvier* and the species  
279 complex (*Carcharhinus spp\**), there was no significant effect of latitude or method.

#### 280 *Equivalence of sampling effort*

281 PERMANOVA tests on bootstrapped CPUE data and a range of sampling efforts indicated that the  
282 relative abundance of requiem sharks obtained from each camera system (MaxN per hour) is  
283 statistically comparable ( $P > 0.05$ ) to a sample obtained from 5 to 30 hooks with similarities peaking  
284 at 12 hooks (Fig. 6). This range of effort equivalence for requiem sharks is largely driven by *C.*  
285 *plumbeus*, the most abundant species in this study. The range of equivalence varied among species,  
286 for *C. plumbeus* effort equivalence ranged between 3 and 30 hooks, with similarities peaking at 10  
287 hooks. *Carcharhinus spp\** and *G. cuvier* showed no significant difference between methods when  
288 MaxN per hour was compared to the catch of 1 to 50 hooks but similarities peaked at 24 and 21  
289 hooks, respectively. Results for *C. limbatus/tilstoni* and *R. acutus* were inconclusive due to the low  
290 abundance recorded with both techniques.

291 Precision estimates of pelagic stereo-BRUVs and scientific longline surveys were similar for the  
292 Carcharhinidae family, *C. plumbeus*, *G. cuvier* and *C. limbatus/tilstoni* (Table 2). For *Carcharhinus*  
293 *spp\** and *R. acutus*, estimates obtained from longline surveys were more precise. Note that, as the  
294 values of  $p$  decrease, the precision of the sampling technique improves. We found that both  
295 techniques were considerably less precise at sampling uncommon species compared to the more  
296 abundant species. Precision values of 1 indicate that individuals of that species were only recorded in  
297 one deployment.

#### 298 *Comparison of length distributions*

299 There were no significant differences in the shape of the length distributions sampled with both  
300 methods for the family Carcharhinidae and, at the species level, for *C. plumbeus*, *Carcharhinus spp\**  
301 and *G. cuvier*. However, there were significant differences in the location (i.e. mean length) of the

302 length distributions for the Carcharhinidae and, at the species level, for *C. plumbeus*. For these taxa,  
303 longline surveys were more selective of larger individuals (Table 3, Fig. 7). Standard error bands are  
304 wide for those species with small sample sizes; therefore the interpretation of the results should be  
305 undertaken with caution.

## 306 **Discussion**

307 We demonstrated that pelagic stereo-BRUVs provide an alternative non-lethal method of sampling  
308 sharks that can be calibrated with standard methods such as scientific longline surveys. The  
309 proportion of Carcharhinidae species sampled by pelagic stereo-BRUVs and scientific longline  
310 surveys was comparable across the study. Pelagic stereo-BRUVs provided a comparable estimate of  
311 Carcharhinidae species that is proportional to longline surveys, whilst also providing abundance  
312 information on other teleost species that were not targeted or captured by longlines due to the  
313 selectivity of the hooks. Longlines sampled a greater proportion of benthic shark species due to the  
314 deployment design that set hooks adjacent to the ballast in close proximity to the benthos.

315 The species composition of the Carcharhinidae between the two methods was also consistent across a  
316 broad latitudinal gradient. These findings support previous studies that define baited video techniques  
317 as a suitable, standardised and non-extractive approach to study the distribution of mobile species  
318 across broad spatial scales (Langlois *et al.* 2012b; White *et al.* 2013). In the current study, *C.*  
319 *plumbeus* was the most abundant requiem shark species captured by both sampling methods. It was  
320 recorded throughout the study area, but in greater numbers at the northern sites. *Galeocerdo cuvier*  
321 and the *Carcharhinus spp\** complex also occurred throughout the study area, and showed no  
322 significant pattern along the latitudinal gradient for either method. *Carcharhinus limbatus/tilstoni* and  
323 *R. acutus* were only recorded in the most northern sites.

324 Each pelagic stereo-BRUV system yielded equivalent relative abundance estimates for requiem sharks  
325 to that of 5 to 30 hooks in scientific longlines. Effort equivalence between techniques peaked at 12  
326 hooks for requiem sharks and, at the species level, effort equivalence peaked at 10, 21 and 24 hooks  
327 for *C. plumbeus*, *G. cuvier* and *C. spp\**, respectively. Due to logistic constraints we could only deploy

328 one camera system for every longline (50 hooks). Although the target species composition and  
329 relative abundance derived from these techniques were comparable, in absolute terms longlines  
330 caught a greater number of individuals of the target species (159) than those recorded at MaxN on  
331 BRUVs (36). In addition, it should be noted that the methods differed in the area covered and in the  
332 amount of bait used. Longline shots, with a length of 500 m each and more than 10 times the amount  
333 of bait in the water, have a greater ability to attract or encounter fish than a single baited camera  
334 system (Brooks *et al.* 2011). Thus, increasing sampling effort of the non-extractive pelagic stereo-  
335 BRUVs to approximately one camera deployment for every 10 to 24 hooks is recommended to exert a  
336 sampling effort equivalent to the commonly used scientific longline surveys.

337 Precision estimates for both techniques were similar at family level and for the most abundant target  
338 species. Precision is most affected by sampling effort; thus increasing replication would rapidly  
339 enhance sampling precision (Andrew and Mapstone, 1987). In this study, the number of deployments  
340 of pelagic stereo-BRUVs per site was limited by the complexity of using two methods at once.  
341 However, future studies using this technique could deploy more camera systems simultaneously,  
342 which would rapidly boost replication without added field-time cost (Santana-Garcon, Newman &  
343 Harvey 2014) and, in turn, it would rapidly enhance their sampling precision.

344 Stereo-BRUVs remove biases due to gear selectivity, such as hook size, that are an undesired by-  
345 product of conventional fishing methods (Cappo *et al.* 2003). In the present study, longline surveys  
346 were selective towards larger individuals of the family Carcharhinidae in comparison to pelagic  
347 stereo-BRUVs. At the species level, size selectivity was only significant for *C. plumbeus*, but KDE  
348 tests for other shark species might lack power to detect differences between methods due to the small  
349 sample sizes available ( $N < 50$ ) (Bowman & Azzalini 1997). Mean fork length of *C. plumbeus* (1280  
350 mm) and *C. spp\** (1627 mm) recorded from longline surveys were larger than those recorded from  
351 stereo-BRUVs (1089 and 1534 mm). Conversely, tiger sharks were on average larger on stereo-  
352 BRUVs (2028 mm) in comparison to longlines (1823 mm). Previous studies have shown species-  
353 specific differences in the length distributions of fish sampled with stereo-BRUVs and line fishing  
354 (Langlois *et al.* 2012a), or traps (Harvey *et al.* 2012b) but the differences reported were not

355 biologically significant. In the present study, low replication was a major limitation in the analysis of  
356 length distributions; hence, further research is needed to continue exploring the differences between  
357 size selectivity of longlines and pelagic stereo-BRUVs.

358 Video techniques have proven to be non-intrusive, causing no physical trauma or physiological stress  
359 to the individuals recorded (Brooks *et al.* 2011). Despite attempts to minimise the impact on sharks  
360 caught in longline surveys, mortality does occur and the level of post-release mortality is not known  
361 (Skomal 2007). The non-destructive nature of stereo-BRUVs allows for deployment in fragile and  
362 protected areas and reduces the negative effects of extractive gears when targeting rare and threatened  
363 species (White *et al.* 2013). Additionally, remote video techniques provide a permanent record of  
364 species behaviour in their natural environment (Zintzen *et al.* 2011; Santana-Garcon *et al.* 2014). A  
365 recurrent behaviour across shark species observed during video analysis was that individuals were  
366 first observed far from the camera system but remain in the area patrolling the bait source. They  
367 approach the bait in a cautious manner over time. This behaviour suggests that longer soak times  
368 facilitate the recognition of individual features including species, sex or external markings.  
369 Nonetheless, this territorial behaviour could also prevent other individuals of the same or other  
370 species from approaching the cameras, which could affect estimates of species composition and  
371 relative abundance (Klages *et al.* 2014).

372 Many species of the family Carcharhinidae are externally similar and visual identification can be  
373 difficult. Identification of individuals to species level from video alone is the main limitation of  
374 pelagic stereo-BRUVs to study requiem sharks (Santana-Garcon, Newman & Harvey 2014). Species  
375 identification can also be restricted in fishery-dependent methods, species may be misidentified or  
376 pooled under general categories (Walker 1998). Identifying features of requiem sharks are often  
377 subtle and the most important of these are tooth shape and numbers, position of the dorsal fins, colour,  
378 and the presence or absence of an interdorsal ridge. These features can be difficult to assess during the  
379 rapid processing of sharks caught on longlines and this is exacerbated when using remote video  
380 techniques. Although most species could be distinguished on video when individuals come close to  
381 the cameras, identification of some species across all replicate videos may not be possible. Another

382 constraint of video techniques, although not assessed in this study, are limitations to identifying the  
383 sex of individuals. Claspers of mature males were often visible, but identification of females and  
384 young males with uncalcified claspers was more challenging. The lack of this information could limit  
385 the use of video techniques in studies of intra-species demographics (Brooks *et al.* 2011). However,  
386 advances in high-definition digital video and automation of the identification of key morphological  
387 characteristics could improve the rates of identification of species, sex and even discrimination  
388 between individuals (Harvey *et al.* 2010; Shortis *et al.* 2013).

389 This assessment of the novel pelagic stereo-BRUVs and its comparison to the commonly used  
390 scientific longline surveys provides a better understanding of the strengths and limitations of each  
391 technique. The two methods produce comparable estimates of relative abundance and species  
392 composition for requiem sharks, and the choice of sampling technique in future should depend on the  
393 specific aims of the study. Scientific longline surveys continue to be a more appropriate approach for  
394 research targeting species that could not be confidently identified on video, or studies on population  
395 biology that require finer intra-specific information such as sex ratio or reproduction information, the  
396 collection of tissue samples (e.g. genetic and isotopic analyses), or the implantation of conventional or  
397 electronic tags (McAuley *et al.* 2007). Stereo-BRUVs, however, provide a suitable sampling method  
398 that can be calibrated to standard techniques for studies with broad spatial and temporal scales,  
399 directed at questions of species composition, behaviour, relative abundance and size distribution of  
400 fish assemblages (Watson & Harvey 2009; Langlois *et al.* 2012b; Santana-Garcon *et al.* 2014),  
401 including highly mobile species (Brooks *et al.* 2011; White *et al.* 2013; Santana-Garcon, Newman &  
402 Harvey 2014). Furthermore, studies conducted on rare or threatened species, and in areas that are  
403 closed to fishing might require a non-intrusive approach like baited video techniques (White *et al.*  
404 2013). Our study demonstrated that pelagic stereo-BRUVs can provide comparable information to  
405 longline surveys on the relative abundance and size composition of requiem sharks, and determined  
406 the required sampling effort to calibrate both methods.

407



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564 **TABLES:**

565 **Table 1.** Summary of ANCOVA tests with method as factor and latitude as covariate. Abundance  
 566 data was collected with pelagic stereo-BRUVs and scientific longline surveys along an 8-degree  
 567 latitudinal gradient. P-values in bold are statistically significant.

	<u>Latitude</u>	<u>Method</u>	La x Me
<i>Carcharhinus plumbeus</i>	<b>&lt;0.001</b>	0.878	0.108
<i>Carcharhinus spp*</i>	0.486	0.291	0.571
<i>Galeocerdo cuvier</i>	0.350	0.226	0.208
<i>Carcharhinus limbatus/tilstoni</i>	<b>0.022</b>	0.682	0.303
<i>Rhizoprionodon acutus</i>	<b>0.003</b>	0.834	0.410

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581 **Table 2.** Precision estimates for target species sampled using pelagic stereo-BRUVs and scientific  
 582 longline surveys. Precision ( $p$ ) was estimated as a ratio of the standard error and the mean abundance  
 583 per deployment. Note that lower values of  $p$  indicate better precision.

	$p$ BRUVs	$p$ Longlines
Family Carcharhinidae	0.236	0.206
<i>Carcharhinus plumbeus</i>	0.273	0.238
<i>Carcharhinus spp*</i>	0.427	0.270
<i>Galeocerdo cuvier</i>	0.376	0.331
<i>Carcharhinus limbatus/tilstoni</i>	1	0.964
<i>Rhizoprionodon acutus</i>	1	0.736

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597 **Table 3.** Summary of the lengths of target species measured on pelagic stereo-BRUVs and caught on  
 598 scientific longline surveys. Maximum (Max), minimum (Min) and mean fork length (FL) are shown  
 599 in millimetres.

Species	Max FL (mm)		Min FL (mm)		Mean FL (mm)	
	BRUVs	Longline	BRUVs	Longline	BRUVs	Longline
Family Carcharhinidae	2937	2870	587	702	1348	1367
<i>Carcharhinus plumbeus</i>	1386	1600	587	730	1089	1280
<i>Carcharhinus spp*</i>	1855	2110	1157	1534	1534	1627
<i>Galeocerdo cuvier</i>	2937	2870	1285	930	2028	1823

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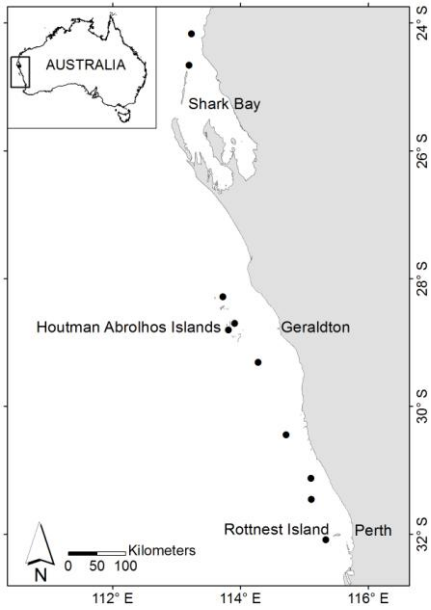
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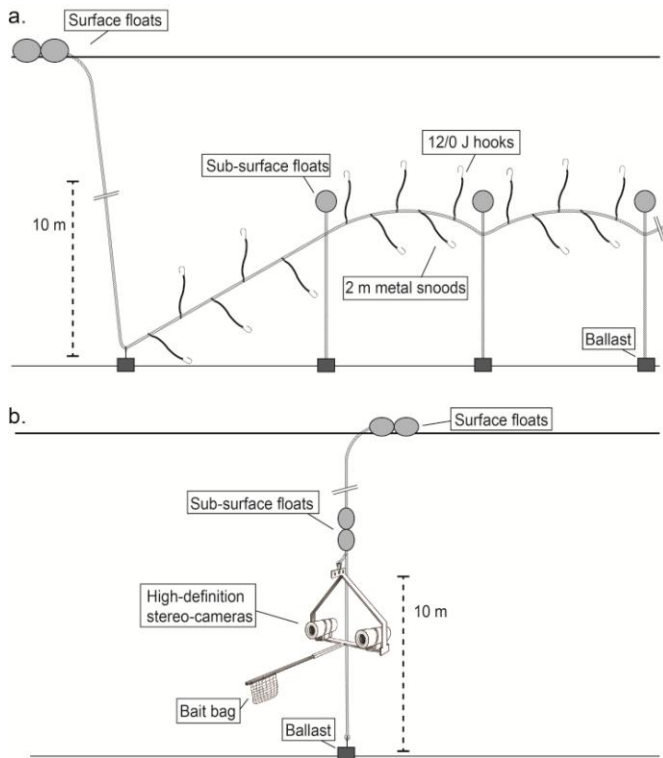
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614 **FIGURES:**



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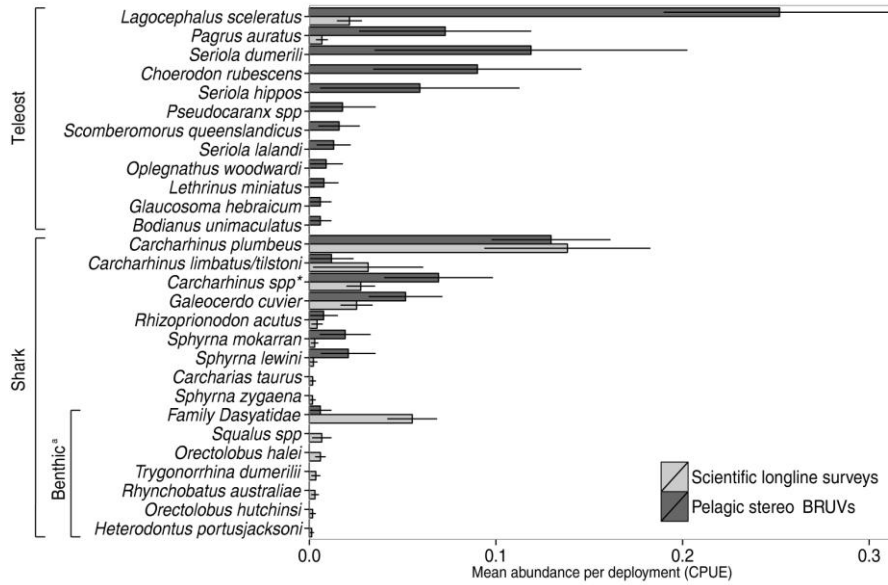
616 **Fig. 1.** Location of study sites along the coast of Western Australia.



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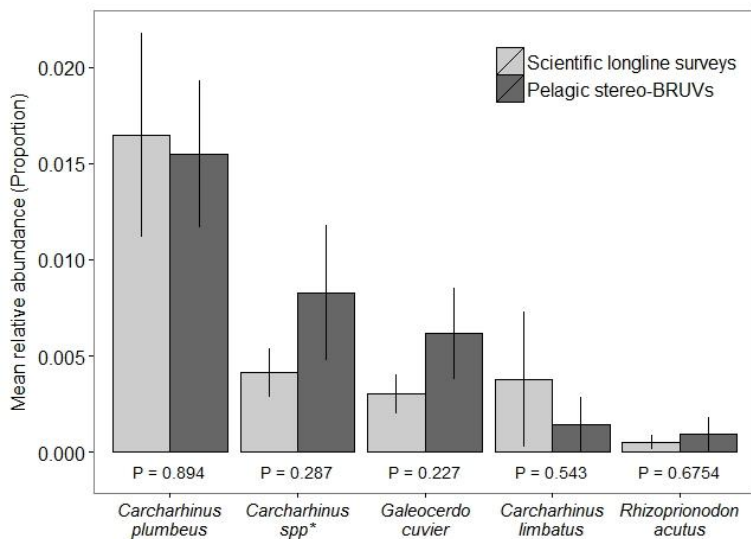
618 **Fig. 2.** Deployment design of the (a) scientific longline shots and (b) pelagic stereo-BRUVs used to

619 sample requiem sharks.



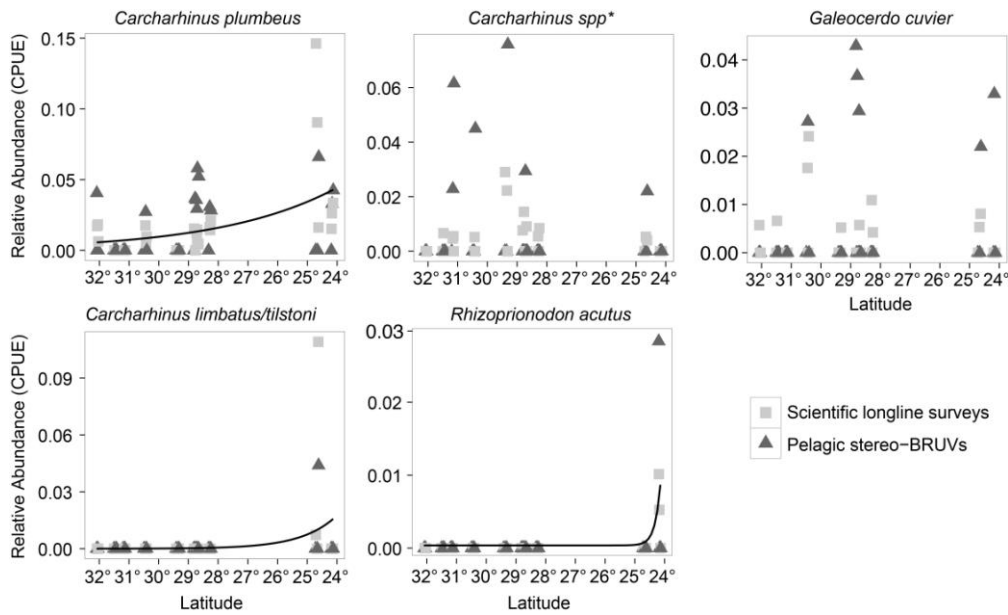
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621 **Fig. 3.** Mean relative abundance of fish species sampled using scientific longline surveys and pelagic  
 622 stereo-BRUVs. Catch per unit of effort (CPUE) is shown as catch per hour for 10 hooks (CPUE10) in  
 623 longline samples and as MaxN per hour in stereo-BRUVs. <sup>a</sup> Benthic sharks were caught in longlines  
 624 due to the deployment design setting of a proportion of hooks near the bottom.



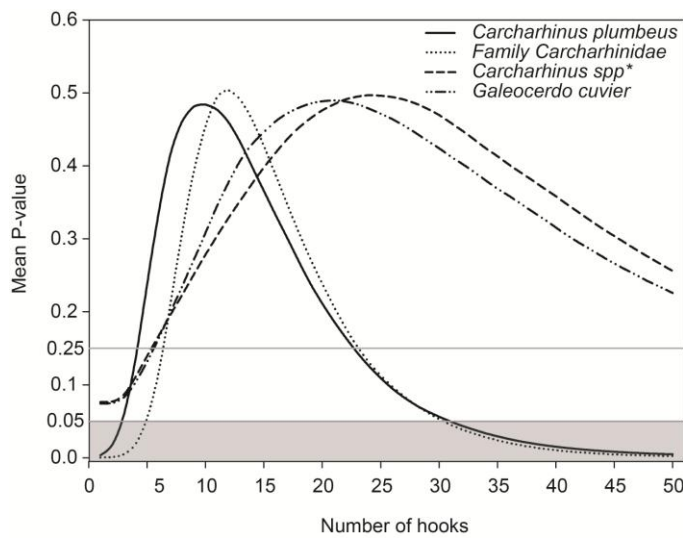
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626 **Fig. 4.** Mean species proportion of target species sampled using scientific longline surveys and  
 627 pelagic stereo-BRUVs. P-values show non-significant differences between sampling methods.



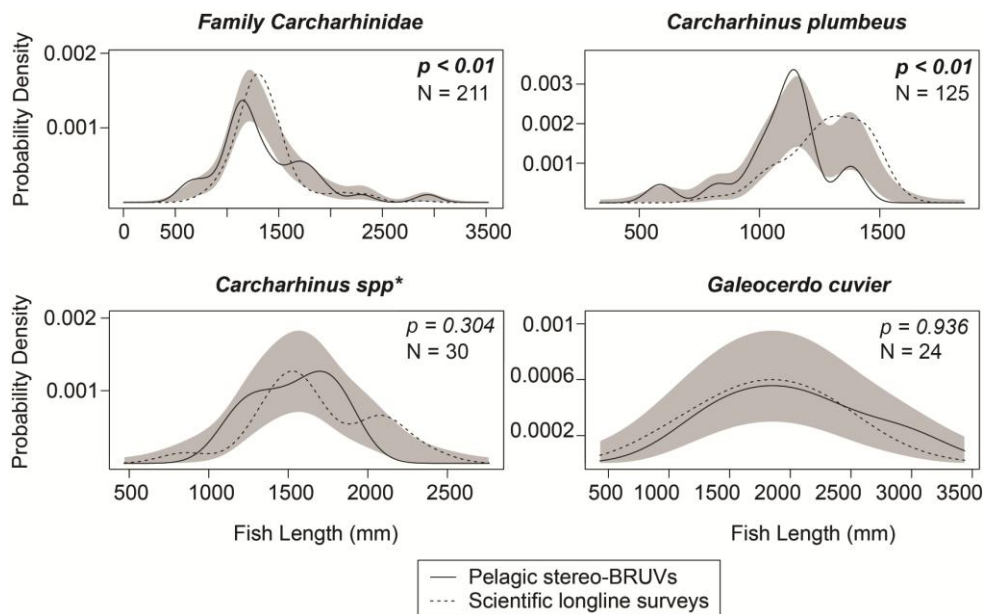
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629 **Fig. 5.** Relative abundance (CPUE) of target species along a latitudinal gradient (32° - 24° S) sampled  
 630 using scientific longline surveys and pelagic stereo-BRUVs. Trendlines illustrate the ANCOVA  
 631 result.



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633 **Fig. 6.** Equivalence of sampling effort required to estimate relative abundance of requiem sharks. Plot  
 634 shows mean p-values from one-way PERMANOVAs testing the differences between pelagic stereo-  
 635 BRUVs (MaxN per hour) and scientific longline surveys (catch\* $h^{-1}$ \*hook $^{-1}$ ) across different levels of  
 636 sampling effort (number of hooks). Values in the shaded area ( $P < 0.05$ ) are statistically significant.  
 637 Grey line is shown for reference as the general pooling cut-off ( $P > 0.25$ ).



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639 **Fig. 7.** Comparison of kernel density estimate (KDE) probability density functions for the length  
 640 distributions of requiem shark species caught by pelagic stereo-BRUVs and scientific longline  
 641 surveys. Grey bands represent one standard error either side of the null model of no difference  
 642 between the KDEs for each method.