1 2	BIOC (Biodiversity and Conservation) 2013 22(10):2353-2374
3 4 5	Camilla Wellstein, Stefano Chelli, Giandiego Campetella, Sandor Bartha, Marco Galiè, Francesco Spada, Roberto Canullo
6 7	Intraspecific phenotypic variability of plant functional traits in contrasting
8	mountain grasslands habitats
9	
10	
11	C. Wellstein* - Department of Biogeography, University of Bayreuth, Universitat sstr. 30,
12	95440 Bayreuth, Germany, * e-mail: camilla.wellstein@uni-bayreuth.de
13	Bayreuth Centre for Ecology and Environmental Research BayCEER, DrHans-Frisch-Straße
14	1-3, 95448 Bayreuth, Germany
15	
16 17	S. Chelli, G. Campetella, R. Canullo - Lab. of Plant Population Ecology, School of
17 18	Environmental Sciences, University of Camerino, Via Pontoni, 5, Camerino, 62032 Macerata, Italy
10 19	italy
20	S. Bartha - Centre for Ecological Research, Institute of Ecology and Botany, Hungarian
21	Academy of Sciences, Alkotmány út 2-4, 2163 Vácrátót, Hungary
22	
23	M. Galiè - Interdipartimentale Centre of Botanical Garden, Marche Polytechnic University,
24	Via Brecce Bianche, Monte Dago, 60131 Ancona, Italy
25	
26	F. Spada - Department of Environmental Biology, University of Rome "La Sapienza",
27	Botanical Garden of Rome, Largo Cristina di Svezia 24, 00165 Rome, Italy
28 29	Abstract
30	Empirical studies that link plants intraspecific variation to environmental conditions are
31	almost lacking, despite their relevance in understanding mechanisms of plant adaptation, in
32	predicting the outcome of environmental change and in conservation. Here, we investigate
33	intraspecific trait variation of four grassland species along with abiotic environmental
34	variation at high spatial resolution (n=30 samples per species trait and environmental factor
35	per site) in two contrasting grassland habitats in Central Apennines (Italy). We test for
36	phenotypic adaptation between habitats, intraspecific trait-environment relationships within
37	habitats, and the extent of trait and environmental variation. We considered whole plant,
38	clonal, leaf, and seed traits. Differences between habitats were tested using ANOVA and
39 40	ANCOVA. Trait-environment relationships were assessed using multiple regression models
40 41	and hierarchical variance partitioning. The extent of variation was calculated using the coefficient of variation. Significant intraspecific differences in trait attributes between the
42	contrasting habitats indicate phenotypic adaptation to in situ environmental conditions.
43	Within habitats, light, soil temperature, and the availability of nitrate, ammonium, magnesium
44	and potassium were the most important factors driving intraspecific trait-environment
45	relationships. Leaf traits and height growth show lower variability than environment being
46	probably more regulated by plants than clonal traits which show much higher variability. We
47	show the adaptive significance of key plant traits leading to intraspecific adaptation of
48	strategies providing insights for conservation of extant grassland communities. We argue that
49	protecting habitats with considerable medium- and small-scale environmental heterogeneity is

- 50 important to maintain large intra-specific variability within local populations that finally can
- 51 buffer against uncertainty of future climate and land use scenarios.
- 52
- 53 Keywords
- 54 Clonal growth, Intraspecific trait variation, Leaf traits, Phenotypic adaptation, Phenotypic
- 55 plasticity, Seed mass
- 56
- 57 Abbreviations
- 58 SLA Specific leaf area
- 59 LDMC Leaf dry matter content
- 60
- 61 Nomenclature
- 62 Pignatti (1982) for taxa; Venanzoni & Kwiatkowski (1995) for the syntaxa mentioned in
- 63 Table 1.
- 64
- 65 Electronic supplementary material
- 66 The online version of this article (doi:10.1007/s10531-013-0484-6) contains supplementary
- 67 material, which is available to authorized users.
- 68
- 69
- 70 Introduction
- 71

72 Trait-based approaches increasingly contribute to link environmental changes with plant 73 community variation (e.g. Webb et al. 2010; Wellstein et al. 2011). Studies on interspecific 74 trait-environment relationships make it clear that plant species are related to environmental 75 conditions via their functional traits (e.g. Poorter et al. 2009). However, trait variation extends beyond interspecific differences being affected by intraspecific phenotypic and genetic 76 77 variation (Albert et al. 2010; Nicotra et al. 2010). Intraspecific (i.e. within-species) variation 78 is partly due to phenotypic plasticity, i.e. the production of multiple phenotypes from a single 79 genotype (e.g. Miner et al. 2005). Phenotypic plasticity gives a plant the ability to adjust its 80 performance by altering morphology and/or physiology in response to varying environmental 81 conditions throughout its lifespan. This phenotypic adaptation is also named phenotypic 82 accommodation (see West-Eberhard 2005; Badyaev 2009) and is considered to confer a 83 benefit on the organism with regard to its present relationship with its environment (Sultan 84 1987). Moreover, intraspecific variation can be caused by inheritable differences in gene 85 expression and function (i.e. epigenetic changes, addressed as microevolution, Bossdorf et al. 86 2008) and by differences in the genotype which can be induced by local long-term adaptation 87 to the environment (i.e. evolution).

88 Since plants as sessile organisms are known for their functional variation within both 89 species and individuals (Miner et al. 2005; Hulshof and Swenson 2010), it is likely that finer-90 scale trait variation plays an important role in controlling species establishment and 91 persistence (e.g. Violle et al. 2012). Actually, it has been proved that many plant populations 92 exhibit significant phenotypic variation for a range of traits within very small areas (Linhart 93 and Grant 1996), including life-history characters (Linhart 1988), pathogen and herbivore 94 resistance (Burdon 1987; Simms 1990), and nutrient allocation related to resource capture and 95 competitive ability (Turkington and Aarssen 1984; Turkington 1989). Intraspecific trait 96 variation may lead to phenotypic differentiation resulting in different abilities of plants to 97 cope with environmental change, i.e. phenotypic specialization (e.g. Bolnick et al. 2003). In 98 this context, the adaptive capacity of traits related to specific ecological processes such as 99 dispersal, establishment, competition, regeneration and flowering, might be of particular

importance. While these traits have been tested for variation across communities taking into
account relative species cover (community-weighted means (CWM), e.g. Garnier et al. 2004;
Wellstein and Kuss 2011), their intra-specific variation in response to changing environmental
conditions has rarely been tested so far. The knowledge on the mechanisms and of the extent
of plant adaptation is a prerequisite when predicting the outcome of climate and land use
changes. At the same time, there is an increasing demand to incorporate climate change issues

106 into conservation planning (e.g. Groves et al.2012; Mawdsley et al. 2009).

107 Since the analysis of trait variation along gradients and under contrasting 108 environmental conditions provides an approach to quantify intraspecific variation and niche 109 breadth of individual species (Ackerly and Cornwell 2007), it is of high interest to test for 110 phenotypic adaptation in this context. We selected key plant traits and environmental 111 conditions, which are likely candidates to stay in trait-environment relationships according to evidence in current literature (Table 2). Those traits were sampled by a large amount of 112 113 individuals, in order to cover the phenotypic variability of leaf, seed, clonal and whole plant 114 traits in four representative species of montane nutrient poor dry grassland ecosystems 115 (Sesleria nitida, Lotus corniculatus, Astragalus sempervirens, Thymus longicaulis) in two 116 habitats with contrasting environmental conditions in terms of soil chemical, physical 117 parameters and light availability.

118 We hypothesize that phenotypic variation of leaf, seed, clonal and whole plant traits is 119 adaptive at small geographical scales (i.e. few 100 m to few centimetres). In detail, we 120 hypothesize that trait attributes of species significantly differ between habitats with 121 contrasting environmental characteristics (H1), and that trait attributes are linked to 122 environmental gradients within a habitat (H2). Furthermore, we explore the relationship 123 between trait and environmental variation, searching for conservative, regulated versus highly 124 plastic traits. Finally, we discuss the implications of our findings for nature conservation and 125 climate change adaptation.

126 127

128 Materials and methods

129

130 Study area and site selection

131

132 The Nature Reserve 'Montagna di Torricchio' (Central Apennines, Italy, online resource 133 ESM 1a) provides areas of montane grasslands under different environmental conditions. The 134 ones considered in this study are located in contrasting habitats, on north and on south-facing 135 slopes along the SW-NE orientated valley (Val di Tazza), between 1,100 and 1,200 m a.s.l. 136 Mean annual precipitation reaches 1,250 mm and mean annual temperature is around 11 °C 137 (Halassy et al. 2005). Jurassic-Cretaceous limestone (scaglia rosata) prevails in the area. The 138 Reserve is under protection regime since 1970; consequently the grazing activities are 139 forbidden since that time. Previously, the grasslands were grazed mostly by sheeps and cows, 140 sometimes also by small groups of horses. We selected two study sites with an area of about 1 141 ha each representing the contrasting environmental conditions of the north and south slope 142 (online resource ESM 1b; Table1). The north-facing slope (habitat N) is covered with a dense 143 grassland assigned to the association S. nitidae–Brometum erecti, here a semi-mesic 144 secondary community originated by the destruction of a former beech forest. The south-facing 145 slope (habitat S) hosts open grassland with a more scanty cover, assigned to Asperulo 146 purpureae-Brometum erecti, a more xerophilic, mostly secondary community replacing both 147 a beech forest as well as the local uppermost fringe of a mixed sub-Mediterranean open forest 148 dominated by Ostrya carpinifolia and Quercus spp. Paleobotanical and archaeological 149 evidence (see Branch 2012; Barker et al. 1991, for reviews) suggest the onset of the

deforestation of the beech forest at higher elevations close to the summits in most of the
northern and central Apennines, to date back to the time of the spread of pastoralism and the
establishment of the altitudinal transhumance in the area, at ~ 4,700 cal. years BP, a period
with warmer and drier conditions than at Middle Holocene.

154 Nevertheless, rocky outcrops, poorly developed, shallow, skeletal soils or solifluction 155 on steeper slopes occur all over the study area, due to its geological assessment (Kwiatkowski 156 and Venanzoni1994). These sites, located far below the climatic treeline at this latitude 157 (1,900–2,000 m a.s.l.) could hardly carry a close canopy, if any, of forest trees, even in 158 absence of human disturbance. They can therefore be suggested as primary stands of 159 "permanent" communities of more or less xerophilic grasslands, already growing "in situ" 160 before the Copper- and Bronze Age deforestation, from which the present day, widespread, 161 secondary grasslands developed. This grassland/forest mosaic close to summits and its 162 changes induced by a long history of grazing can be considered as a general pattern all over 163 the study area.

- 164
- 165 Study species

166

167 From a list of species occurring in both habitats we selected four perennial representatives of 168 montane grasslands: S. nitida ssp. nitida (Poaceae), T. longicaulis (Lamiaceae), L. 169 corniculatus L. (Fabaceae), A. sempervirens (Fabaceae). Due to their more or less accentuated 170 degree of orophytism, they can be assumed as persistent component of these grasslands 171 during the late-glacial/holocene. The frequency of each species in each habitat is given in the 172 online resource ESM 1c. These species represent different life forms in the community, i.e. 173 grass, woody forb, legume, and dwarf shrub. The grass S. nitida is an efficient colonizer on 174 scree and rocky outcrops; it has a secondary root system which is restricted to the upper soil 175 layer. The species exhibits a high capacity of clonal growth by epigeogenous rhizomes. S. 176 nitida is very abundant and dominant in the habitat N and is abundant but scattered as thick 177 tussocks in the habitat S. The woody forb T. longicaulis develops a tap root and forms a dense 178 mat with thin stems, often prostrate and creeping (Pignatti 1982). It has a high capacity of 179 vegetative spread by horizontal above ground stems. T. longicaulis is very abundant in habitat 180 S but is also widespread in habitat N where it is scattered in small patches on rocky outcrops 181 and on shallow soil. L. corniculatus, one of the most important legume species in dry and 182 nutrient poor grasslands, is less abundant on both slopes but shows a similar distribution 183 pattern. The species can grow clonally by hypogeogenous rhizomes (own observations from 184 the present study), it develops tap roots and has a high capacity to withstand soil erosion and 185 is a highly efficient accumulator of nitrogen, for which it is used also in agricultural 186 management (Carter et al. 1997). The dwarf shrub A. sempervirens is a spiny, long-living 187 species (over many decades) growing in summit grasslands on limestone (Pignatti 1982). It is 188 not capable of clonal growth; it produces deep-reaching tap roots which enable the species to 189 withstand erosion and to reach deeper water reserves in the soil. A. sempervirens is quite rare 190 and has the same density on both slopes; it grows on patches with deeper soil or alternatively 191 uses cracks in the rocks for the development of its tap root. 192

- 193 Sampling of plant traits and environmental parameters
- 194

195 For each species in each habitat we randomly selected thirty robust, well grown and adult

196 individuals without symptoms of pathogen or herbivore attack. All field sampling of plant and

197 environmental parameters was done from June 26th to 30th in 2010. In each selected

198 individual we measured the following traits that are associated with different life-history

199 processes (see Table 2): (i) leaf traits (specific leaf area [SLA]; leaf dry matter content

200 [LDMC]), (ii) whole plant traits (height; horizontal stem length), (iii) clonal traits (no. of 201 bifurcations/nodes per cm; distance between bifurcations/nodes) (see Table 2). As an 202 exception, seed traits (iiii) (i.e. seed mass and seed germination), were measured from a 203 random rate of seeds collected from all individuals in each habitat. The trait sampling 204 followed standard procedures (Cornelissen et al. 2003; Kleyer et al. 2008). In particular, leaf 205 traits were measured on three leaves per individual; after the measurement of the fresh weight 206 and area, the leaves were oven-dried (80  $^{\circ}$ C for 48 h) to measure their dry weight. The plant 207 height and horizontal stem length were measured in the field, while the clonal traits were 208 determined in the laboratory on the harvested individuals. Seeds were air-dried and manually 209 cleaned; seed mass was calculated using a balance with accuracy 0.0001 g. Germination tests 210 were carried out at Germoplasm Bank for ex situ conservation of anfiadriatic plant species 211 (ASSB), according to the international seeds germination protocols (ISTA 1999). Seeds were 212 sown in Petri dishes with 1 % of bacteriological agar and germination was tested at 20  $^{\circ}$ C 213 constant temperature and 12/12 h photoperiod. We tested four replicates of 25 seeds for each 214 species per habitat; this represents a commonly used amount of seeds in ecological studies. 215 The germination tests were completed after 30 days.

For the functional traits, only 9 % of the values are lacking. This originated from the absence of seeds in *L. corniculatus* in the habitat N during the sampling season. In habitat S seeds were usually present.

Additionally, in order to adjust analyzes for plant size, we sampled the biomass of all individuals of *L. corniculatus* and *T. longicaulis*. We collected the entire individual, i.e. above and belowground biomass; samples were oven-dried at 80 °C for 48 h and weighted (accuracy: 0.001 g). In case of *S. nitida*, it was not reasonable to adjust for plant size, since the species does not develop huge individuals because clonal offspring disintegrates after some years resulting in no differences appearing with age. In case of *A. sempervirens* we were not allowed to collect the plant specimen for biomass measurements in the nature reserve.

We used a sample size (30 individuals per species per habitat) 3 times larger than the one required by standard protocols (e.g. Cornelissen et al. 2003) for measuring plant traits in order to take into account the effects of environmental heterogeneity of the habitats. Soil physical and chemical parameters and light availability were measured at the exact position of each sampled plant individual. Soil samples were taken only after the measurements of traits in the field.

232 The sunlight reaching the canopy [Photosynthetic Active Radiation (PAR)] was 233 measured using a PAR/LAI Ceptometer (LP80 AccuPAR—Decagon Devices, inc.). The 234 percentage of soil moisture and the soil temperature (°C) were measured using a humidity and 235 temperature probe meter (Tr 46908). For the analysis of soil chemical parameters, a soil 236 sample (3 cm diameter, usually 10 cm depth; on rocky outcrops the soils were more shallow) 237 was taken at the exact position of each plant individual. We determined the pH value and 238 conductivity (in water), the total nitrogen (Nt) and total carbon (Ct) content (CNA analyzer 239 FlashEA 1112, Thermoquest) and the CaCO<sub>3</sub> (Calcimeter 08.53, Eijkelkamp) according to 240 Scheibler (Hoffmann1991). The C/N ratio was calculated considering that total carbon minus 241 the CaCO<sub>3</sub>-carbon gives the organic carbon in the soil.

247 Data analyses

Significant differences in environmental variables between species and habitats (at the place
of growth of individuals of each species) were evaluated by two-way ANOVA, followed by
post hoc test (Tukey HSD). Main effect of habitat, species and their interactions are given in
Table 3.

For *S. nitida* and *A. sempervirens*, ANOVA was used to test for the significance of differences in functional trait values between the two habitats. For *L. corniculatus* and *T. longicaulis*, ANCOVA was used to adjust for the effect of plant size (biomass) and to test for the significance of differences in functional trait values between the two habitats (e.g. McCarthy and Enquist 2007).

For seed germination, ANOVA was used since no biomass data were available for plants from which seeds were collected. Prior to statistical analysis, the traits values were log transformed when conditions of normality were not met or in order to improve homogeneity of variances. In all analyses, the level of significance was p<0.05.

262 We applied linear regression and hierarchical variance partitioning to evaluate the 263 intraspecific trait-environment relationship for each trait per species and habitat. For each trait 264 (response variable), we fitted the full model of all measured environmental parameters, 265 leading to 12 distinct models per species. In case a model revealed a significant trait-266 environment relationship we applied hierarchical variance partitioning and subsequent 267 bootstrapping (package *relaimpo*, Grömping 2006) in order to evaluate the relative 268 importance of explanatory environmental variables on a certain trait (Grömping 2006; Murray 269 and Conner 2009). Prior to analyses, data were log, square root or arcsin transformed when 270 necessary. For linear regression, ANOVA and ANCOVA we used the package *nlme*.

271 Additionally, we use the coefficient of variation (CV; SD divided by the mean) as a 272 relative measure of phenotypic variability in order to assess and compare the degree of trait 273 and environmental variation within habitats. The CV has frequently been used in the context 274 of environmental and trait variation (e.g. Lemke et al. 2012). We calculated the CV for each 275 trait of a species and for each environmental parameter (measured at the place of growth of 276 the individuals of each species) within a habitat (Table 5) as well as throughout species and 277 habitats (Table 6). We used quartiles based on data of all parameters for all species within and 278 throughout habitats (data not shown) to categorize four classes of variation (CV) for 279 environmental parameters as well as for traits, i.e. low, medium, high, and very high CV 280 (online resource ESM 2).

281 Our further investigation of the relationships of trait- and environmental variation 282 (CV–CV graphs, Fig. 1) is based on the results of linear regression and hierarchical variance 283 partitioning as we investigated the most important, significant trait-environment relation-ships 284 (Table 5). Each CV–CV graph contains a reference line which allows distinguishing if a trait 285 varies more than expected from the null model. The null model is that in evolution the trait 286 variability follows the variability of the environment. Values of traits that lay above the 287 reference line indicate higher trait variability, values below the line indicate lower variability 288 than expected (Fig. 1).

All statistical analyses were performed using R (Version 2.13.1) (R Development CoreTeam 2010).

- 291
- 292

293 Results

294

295 Phenotypic differences between habitats

296

297 All tested environmental factors, i.e. light availability (PAR), water availability (soil

298 moisture), nutrient availability (plant available phosphorous, plant available potassium,

299 magnesium, nitrate, ammonium, C/N ratio, soil pH, soil conductivity), and soil temperature,

- 300 were significantly different between the two habitats (Table 3). This contrast was mirrored by 301
- significant differences in mean values of leaf traits, whole plant traits, clonal traits and seed 302 traits within species between the two habitats (Table 4). The grass S. nitida and the dwarf
- 303 shrub A. sempervirens exhibited significantly different trait attributes for all tested traits. The
- 304 legume L. corniculatus showed significant differences in 2/3 of the traits, but no
- 305 differentiation in the clonal traits (number of nodes and distance between nodes) was
- 306 observed. The woody forb T. longicaulis differed in half of the traits but no differentiation in
- 307 LDMC, in clonal traits (number of nodes and distance between nodes), and in germination
- 308 rate was observed (Table 4).
- 309
- 310 Intraspecific trait-environment relationships within habitats
- 311

312 Based on linear regression and the results of hierarchical variance partitioning, throughout 313 habitats and traits the most important, significant drivers of trait-environment relationships in 314 sequence are soil temperature, NO<sub>3</sub>, light, NH<sub>4</sub>, magnesium and potassium (Table 5, online 315 resource ESM 3).

316 In at least one of the study species, these parameters strongly and significantly affected 317 at least one trait. Twice as much trait-environment relationships were seen in the north-facing 318 slope compared to the south-facing slope. The SLA showed an increase with increasing 319 nitrate availability In A. sempervirens on the south-facing slope; the LDMC showed an 320 increase with increasing magnesium availability but a decrease with increasing potassium 321 availability in T. longicaulis on the north-facing slope. Height growth was negatively 322 associated with soil temperature in A. sempervirens on the south-slope, in L. corniculatus on both slopes, and positively influenced by light and the C/N ratio in the later species on the 323 324 south-slope. On the north-facing slope, the horizontal stem length of the clonal S. nitida was 325 positively associated with the availability of magnesium, but negatively associated with the 326 availability of potassium and phosphorus, and with conductivity; in the clonal T. longicaulis, 327 stem length was negatively associated with the availability of ammonium. On the north-facing 328 slope, in *T. longicaulis*, the number of nodes per cm responded positively to light and 329 negatively to nitrate, potassium and soil moisture; the distance between the nodes was 330 positively associated with the availability of potassium, phosphorous and magnesium. On the 331 south-facing slope the distance of nodes in L. corniculatus exerted a positive association with 332 soil temperature and negative one with the pH (Table 5). 333

334

Environmental and intraspecific variability within habitats 335

336 Detailed information on the plasticity of traits and environmental variables (measured as the

- 337 CV) is provided in the online resource (ESM 2). Generally, the plasticity of LDMC was very
- 338 low, of SLA and horizontal stem length was intermediate, and of number of
- 339 bifurcations/nodes and distance between bifurcations/nodes was very high (Table 6).
- 340 Variability of nitrate and phosphorous was very high; variability of the C/N ratio, pH, and soil 341 temperature was very low; the other environmental variables showed intermediate variability
- 342 (Table 6).

343 We assessed the relationships between the variation of traits (CV-traits) and of the 344 environmental factors (CV-environmental parameter) emerging as the most influential based 345 on results of linear regression and hierarchical variance partitioning (Table 5, online resource 346 ESM 3). As the variable soil temperature showed a very low variation (Table 6, online

- 347 resource ESM 2) it was not considered. Consequently, CV-CV graphs are displayed for light,
- 348 nitrate (NO<sub>3</sub>), ammonium (NH<sub>4</sub>), magnesium (P) and potassium (K) which exhibited

349

intermediate to high variation (Fig. 1). From the CV-traits to CV-environment comparison 350 (Fig. 1) it appears that the variable traits SLA, LDMC, and height growth have CV values that 351 are similar to or smaller than the CV values of the environmental factors. The highly variable 352 clonal traits, by contrast, appear to have largely higher CV values than the related environmental factors.

- 353
- 354 355
- 356 Discussion
- 357

358 In our study, significant intraspecific differences in plant functional traits between the 359 contrasting habitats indicate phenotypic functional adaptation to in situ environmental 360 conditions, supporting hypothesis H1 for all traits investigated. While species adapted their 361 trait attributes between the two habitats in the same way, they showed species-specific 362 responses to environmental factors within habitats. Environmental differences within habitats 363 were less pronounced. However, soil temperature, light, nitrate, ammonium, magnesium and 364 potassium emerged to be important drivers of intraspecific trait-environment relationships 365 even at the scale of few meters. Leaf traits, horizontal stem length and clonal traits responded 366 significantly positively or negatively to the availability of nutrients, while plant height 367 responded negatively to soil temperature and positively to light availability, supporting H2 for 368 these traits. The variation of leaf traits and plant height was lower than the environmental 369 variation indicating that those traits are more regulated by the plant than clonal traits. The 370 latter, i.e. horizontal stem length, number of bifurcations/nodes and distance between 371 bifurcations/nodes, exhibited much larger trait variation than environmental variation. 372 According to Grassein et al. (2010, see review of Schellenberg and Pontes (2012)) species 373 strategy is defined by both trait values and trait plasticity. Our findings confirm to some 374 extent the conclusions of Grassein et al. (2010) on LDMC as we found the same trait to be 375 controlled compared to other traits. However, in our study, the plasticity of this trait was still 376 high enough to enable adjustment to environmental factors. Our findings encourage 377 investigating trait control at the intraspecific level.

378 In more detail, our results show that SLA and plant height are not only characterized 379 by considerable variation between communities and species, as already assessed (Westoby 380 1998; Grime 2002; Poorter et al. 2006; Pierce et al. 2007), but that ecologically significant 381 variation is also occurring at the intraspecific level (Tables 4, 5). In line with our results, 382 Wilson et al. (1999) found in a survey of 769 herbaceous species of the British flora that SLA 383 exhibits strong variation within populations. With respect to phenotypic adaptation, all four 384 study species showed significantly higher values of SLA and significantly higher height (with 385 the exception of height variation in *T. longicaulis*) in the north-exposed habitat. Our findings 386 support the view that higher plant height and higher SLA in the species in the north-exposed 387 habitat might be related to the significantly higher availability of all investigated nutrients (P, 388 K, Mg, NH<sub>4</sub>, NO<sub>3</sub>) along with higher soil moisture and conductivity. Within the south-facing 389 habitat, the SLA proved to be significantly positively influenced by nitrate availability in A. 390 sempervirens (Table 5). This finding is in line with experimental results of Al Haj Khaled et 391 al. (2005) which demonstrated that nitrogen availability positively affects SLA attributes 392 within species.

393 Literature evidence points out that high SLA values are positively correlated with 394 rapid leaf turnover, potential relative growth rate and photosynthetic efficiency (Cornelissen 395 et al. 2003; Kleyer et al. 2008); fast turnover of plant parts allows also a more flexible 396 response to the spatial patchiness of light availability (Grime 1994), which is a characteristic 397 of the habitat on the north-facing slope for all species (except *T. longicaulis*). The higher

density of the canopy in this habitat causes competition for light, which might modulate lightavailability giving advantage in having a higher SLA.

400 With respect to the second investigated leaf trait, i.e. the leaf dry matter content 401 (LDMC) we found the inverse response as SLA (Table 4). The LDMC was significantly 402 higher in the south-facing habitat for all species (with the exception of *T. longicaulis* whose 403 higher values were not significant). The LDMC of this species, however, showed variation in 404 response to variation in magnesium and potassium availability within the north-facing habitat 405 (Table 5). High values of LDMC correspond to a low turnover rate (Cornelissen et al. 2003; 406 Kleyer et al. 2008) and thicker leaves, better withstanding physical constraints. Both 407 characteristics help to store nutrients, representing a big advantage in unproductive 408 environments (Ryser and Urbas 2000) such as the south-facing slope. Following Gross et al. 409 (2007), this trait appears to be a better predictor for plant responses to nutrient stress than 410 SLA. Earlier literature confirms that LDMC is less variable than SLA (Garnier et al. 2001) 411 which is confirmed by our data showing lower intraspecific and interspecific variation (CV) 412 of LDMC.

413 The second investigated whole-plant trait, i.e. horizontal stem length of clonal species, 414 responded in the same way to the differentiated environmental conditions than plant height. 415 Horizontal stem length represents the diameter of individuals and thus their maximum range 416 of space occupancy. All investigated species showed significantly higher space occupancy in 417 the north-facing habitat (Table 4). These data were not sampled for A. sempervirens, but this 418 is the species with the longest horizontal stems of all examined species. Due to this, it exhibits 419 a much higher capacity of space occupancy in the N-facing habitat (several meters long) 420 compared to the S-facing ones (shorter than 1 m). The results of all study species might be 421 explained by two mechanisms. First, the higher availability of nutrients and water allowing 422 for higher growth potential and second, the space-occupancy advantage under competitive 423 conditions with lower light availability. This is in line with the findings of Tissue and Nobel 424 (1988) and Grime (2002) who state that in a dense community, the horizontal stem length 425 provides higher ability to explore new space, increasing the possibility to exploit new 426 resources and to allocate them within the organism.

Both clonal traits 'number of bifurcations/nodes per cm' and 'distance between
bifurcations/nodes' showed the highest variation in all study species. This might to some
extent be fostered by the high variation of light-, water- and nutrient-availability within
habitats. Apart from being an alternative to sexual reproduction, clonality allows rapid
colonization of open habitats, pre-emptive occupation of space by forming dense patches,
avoidance of competition implicit in fugitive growth and better foraging for resources in a
heterogeneous soil matrix (Stöcklin1992; Oborny and Bartha 1995).

434 In both investigated clonal traits, phenotypic adaptation to the contrasting environ-435 mental conditions of the north- and south-facing slope were found only in S. nitida. This 436 dominant grass showed a higher capacity of multiplication (number of bifurcations) at the 437 south-facing slope but higher capacity of space occupancy (distance between bifurcations) at 438 the north-facing slope. The space occupancy might be confined by the lower nutrient and 439 moisture availability of the habitat S while higher availability of these resources allows for 440 higher general growth capacity, which is reflected also in the horizontal stem length. This trait 441 of S. nitida was significantly positively influenced by an increased availability of magnesium 442 within the north-facing habitat. Within this habitat, T. longicaulis was seen to produce fewer 443 nodes but longer internodal stem-segments with higher nutrient availability while higher light 444 availability led to the production of more nodes.

445 In our study, the seed mass showed significant intraspecific differences between the 446 two contrasting habitats. This is surprising because seed mass differences were found to be often conservative between genera or families (Hodgson and Mackey1986; Mazer 1989; Peat
and Fitter1994; Westoby 1998).

449 Our results confirm phenotypic adaptation of seed mass in all tested species. Their 450 seed mass was significantly higher in the south-facing habitat compared to the north-facing 451 one. As summarized by Pakeman et al. (2008), it has been shown that larger seeds offer an 452 advantage (e.g. Buckley 1982) or are more common in drier environmental conditions 453 (Wright and Westoby 1999), since the seedlings of larger seeds better withstand 454 environmental hazards being reserves needed for drought-resistance mechanisms (Leishman 455 and Westoby 1994). However, Pakeman et al. (2008) found substantially more evidence that 456 seed size was higher at warmer sites. The significant differences found in our study are 457 possibly related to differences in both, temperature and water availability. However, lower 458 competition in the south-facing habitat might foster a higher investment in sexual 459 reproduction. The north-facing habitat, in contrast, exhibits higher levels of competition since 460 S. nitida forms an extraordinarily dense vegetation carpet.

461 Despite the fact that germination rate observed in our study was very low, the results 462 on the intraspecific differences of the germination rate also support our findings on seed mass. 463 Significantly higher intraspecific germination rate of seeds produced by plants in the south-464 exposed habitat are possibly linked to the higher seed mass of these seeds. Thus, the 465 production of larger seeds in the stressful south-facing habitat provides more reserves for 466 germination, in agreement with the advantages of larger seeds reported by Westoby et al. 467 (2002) and the subsequent higher survival rates of seedlings (Moles and Westoby 2006), 468 especially under various hazards including drought (Westoby 1998).

The range of trait variation is the result of the plants trial to reach equilibrium between
costs and benefits as the strategy of each plant individual is the best possible compromise
within a given environment (Reich et al. 2003). In other words, the advantage provided by
phenotypic plasticity per se, allows for adaptation to environmental conditions.

In all four species, the individuals of the south-facing habitat are better equipped to
cope with environmental stress. In fact they are characterized by lower plant height, slower
growing rate, thicker laminas, higher tissue density, lower photosynthetic efficiency, longer
leaf life span (and lower leaf turnover), more investment in structural strength, and higher
investment in seeds (stored energy) which ensure future successful performance of seed-lings
under stressful conditions.

479

480 Data limitation481

482 Both investigated habitats, i.e. the south- and north-facing slope, very likely have differences 483 in their disturbance history. Due to local topography, grazing animals had more easily access 484 to the south-facing slope before the closure of the reserve in 1970. This might explain in 485 concert with the stressful environmental conditions that some of the late-successional plant 486 species as A. sempervirens in the vegetation of the south-facing slope might have lower age. 487 Plant age might influence the performances of some of the examined traits, mainly when it 488 affects capacity of nutrient and water uptake by plant size and rooting depth. This might be 489 the case for A. sempervirens, which is younger in the south- than in the north-facing slope. 490 One representative individual of the south-facing slope was analyzed by dwarf shrub 491 chronology and appeared to be 20 years old. One of the largest individuals on the north-facing 492 slope was estimated according to annual increment of horizontal stem length and is around 60 493 years old. In case of S. nitida no differences appear with age since the species does not 494 develop huge individuals because clonal offspring disintegrate after some years. In case of L. 495 corniculatus and T. longicaulis, both of which are larger and therefore probably older on the 496 south-facing slope, we were able to adjust the testing for trait differences between habitats for

497 plant size (biomass) in all traits (see McCarthy and Enquist 2007) except seed traits.
498 However, results were largely the same when testing for apparent plasticity, i.e. without

499 adjusting for plant size which in turn relativizes the above stressed argumentation on the500 potential impact of plant size on phenotypic trait adaptation.

501 Another data limitation refers to the fact that we studied only two grassland sites (1 ha 502 extension each) and could not include further spatial replicates, i.e. further valleys with north-503 and south-facing grasslands in this highly labour-intensive work. This limits the strength of 504 conclusions drawn to the global change level and we suggest further studies in order to shed 505 more light on the evidence indicated in our study.

506

507 Implications for global change and nature conservation

508

509 Our findings demonstrate phenotypic differentiation of species at medium spatial scales, i.e. 510 200 m air distance between the opposing slopes of a valley. Some species even vary their 511 traits according to the fine scale (i.e. centimeters to meters) heterogeneity in temperature. The 512 existing medium-scale climatic differences between the contrasting slopes, manifesting e.g. in 513 a difference of 8 °C in the soil temperature, are larger than the overall climatic shifts predicted 514 by coarse-scale scenarios (IPCC 2012) and partly cover the magnitude of European extreme 515 events such as mega-heatwaves (Barriopedro et al. 2011). Therefore, we expect that these 516 species can be pre-adapted (to some degree) to the overall expected environmental changes. 517 Studying medium- and fine-scale intraspecific trait variability helps to quantify the magnitude 518 of plasticity that can serve as adaptive potential of plant species. 519 Switching the focus on land-use change, the abandonment of traditional grazing

regimes in the study area resulted in succession and subsequent increase in competition in the climatically and edaphically favorable north-facing slope. The related differences in nutrient availability at medium and fine spatial scales resulted in an adjustment of functional traits, e.g. SLA and LDMC.

524 As phenotypic specialization might enable a differentiated response to land-use and 525 climate change it ultimately might increase plant fitness and survival. Nature conservation 526 should therefore protect environmental heterogeneity between and within habitats in order to 527 maintain larger intraspecific variability and thereby a variety of phenotypic specialization that 528 finally can buffer future environmental extremities due to climate and land-use changes. 529 These findings might support conservation planning with information on how and where to 530 prioritize conservation objects and how to work in situ (Beier and Brost 2010; Groves et al. 531 2012). This meets the purpose of moderating impacts of climate change and capitalizing on 532 emerging opportunities, i.e. climate change adaptation (Groves et al.2012; IPCC 2012).

- 533
- 534

## 535 Conclusions

536

Using ca. three times larger sample size than required by the standard protocol for measuring plant traits, we show the intermediate to high degree of intraspecific variability of whole plant, clonal, leaf and seed traits. We stress the adaptive significance of the key plant traits leading to intraspecific adaptation of strategies. We argue that protecting habitats with considerable medium- and small-scale environmental heterogeneity is important to maintain large intraspecific variability within local populations that finally can buffer against uncertainty of future climate and land use scenarios.

544

546 Acknowledgments

547 We thank the Analytical Chemistry of the Bayreuth Centre of Ecology and Environmental Research (BayCEER,

548 Germany) for performing the analyses of soil chemical parameters. This research was partially supported by

549 Montagna di Torricchio Nature Reserve. SB was supported by the Hungarian National Science Foundation

(OTKA K 72561). CW was supported by the Bavarian State Ministry of Sciences, Research and the Arts within
 the FORKAST project. We thank Martin Hallinger (University of Greifswald, Germany) for help with the

552 determination of age of *A. sempervirens*.

- 553
- 554

556

## 555 References

- Ackerly DD, Cornwell WK (2007) A trait-based approach to community assembly: partitioning of species trait
   values into within- and among-community components. Ecol Lett 10:135–145
- Albert CH, Thuiller W, Yoccoz NG, Douzet R, Aubert S, Lavorel S (2010) A multi-trait approach reveals the structure and the relative importance of intra vs. interspecific variability in plant traits. Funct Ecol 24:1192–1201
- Al Haj Khaled R, Duru M, Theau JP, Plantureux S, Cruz P (2005) Variation in leaf traits through seasons and N availability levels and its consequences for ranking grassland species. J Veg Sci 16:391–398
- Badyaev AV (2009) Evolutionary significance of phenotypic accommodation in novel environments: an
   empirical test of the Baldwin effect. Phil Trans R Soc B 364:1125–1141
- Barriopedro D, Fischer EM, Luterbacher J, Trigo RM, Garcia-Herrera R (2011) The hot summer of 2010:
   redrawing the temperature record map of Europe. Science 332:220–224
- Barker G, Grant A, Beavitt P, Christie N, Giorgi J, Hoare P, Leggio T, Migliavacca M (1991) Ancient and
   Modern Pastoralism in Central Italy: an Interdisciplinary Study in the Cicolano Mountains. Papers of the
   British School at Rome 59:15–88
- Baskin CC, Baskin JM (1998) Seeds: Ecology, biogeography, and evolution of dormancy and germination.
   Academic Press, San Diego
- Beier P, Brost B (2010) Use of land facets to plan for climate change: conserving the arenas, not the actors.
   Conserv Biol 24:701–710
- Bolnick DI, Svanba¨ck R, Fordyce JA, Yang LH, Davis JM, Hulsey CD, Forister ML (2003) The ecology of
   individuals: incidence and implications of individual specialization. Am Nat 161:1–28
- 577 Bossdorf O, Richards CL, Pigliucci M (2008) Epigenetics for ecologists. Ecol Lett 11:106–115
- Branch NP (2012) Early-Middle Holocene vegetation history, climate change and human activities at Lago
   Riane (Ligurian Apennines, NW Italy). Veg Hist Archeobot. Published online 21 November 2012
- 580 Buckley RC (1982) Seed size and seedling establishment in tropical arid dunecrest plants. Biotropica 14:314– 581 315
- Burdon JJ (1987) Phenotypic and genetic patterns of resistance to the pathogen Phakopsora pachyrhizi in populations of Glycine canescens. Oecologia 73:257–267
- Carter EB, Theodorou MK, Morris P (1997) Responses of Lotus corniculatus to environmental change. New
   Phytol 36:245–253
- Cornelissen JHC, Lavorel S, Garnier E, Di´az S, Buchmann N, Gurvich DE, Reich PB, ter Steege H, Morgan
   HD, van der Heijden MGA, Pausas JG, Poorter H (2003) A handbook of protocols for standardised and
   easy measurement of plant functional traits worldwide; in. Aust J Bot 51:335–380
- Garnier E, Laurent G, Bellmann A, Debain S, Berthelier P, Ducout B, Roumet C, Navas ML (2001) Consistency
   of species ranking based on functional leaf traits. New Phytol 152:69–83
- 591 Garnier E, Cortez J, Billes G, Navas ML, Roumet C, Debussche M, Laurent G, Blanchard A, Aubry D,
   592 Bellmann A, Neill C, Toussaint JP (2004) Plant functional markers capture ecosystem properties during
   593 secondary succession. Ecology 85:2630–2637
- 594 Grassein F, Till-Bottraud I, Lavorel S (2010) Plant resource-use strategies: the importance of phenotypic 595 plasticity in response to a productivity gradient for two subalpine species. Ann Bot 106:637–645
- 596 Grime JP (1994) The role of plasticity in exploiting environmental heterogeneity. In: Caldwell MM, Pearcy R
   597 (eds) Exploitation of Environmental Heterogeneity in Plants. Academic press, San Diego, pp 1–18

- 598 Grime JP (2002) Plant strategies, vegetation processes, and ecosystem properties, 2nd edn. Wiley, Chichester
- 599 Gross N, Suding KN, Lavorel S (2007) Leaf dry matter content and lateral spread predict response to land use 600 change for six subalpine grassland species. J Veg Sci 18:289–300
- Groves CR, Game ET, Anderson MG, Cross M, Enquist C, Ferdan a Z, Girvetz E, Gondor A, Hall KR, Higgins
   J, Marshall R, Popper K, Schill S, Shafer SL (2012) Incorporating climate change into systematic
   conservation planning. Biodivers Conserv 21:1651–1671
- 604 Grömping U (2006) Relative importance for linear regression in R: the package relaimpo. J Stat Softw 17:1–27
- Halassy M, Campetella G, Canullo R, Mucina L (2005) Patterns of functional clonal traits and clonal growth
   modes in contrasting grasslands in the central Apennines, Italy. J Veg Sci 16:29–36
- Hodgson JG, Mackey JML (1986) The ecological specialisation of dicotyledonous families within a local flora:
   some factors constraining optimization of seed size and their evolutionary significance. NewPhytol
   12:497–515
- 610 Hoffmann G (1991) Die Untersuchung von Bo<sup>--</sup>den Methodenbuch 1. VDLUFA-Verlag, Darmstadt
- Hulshof CM, Swenson NG (2010) Variation in leaf functional trait values within and across individuals and
   species: an example from a Costa Rican dry forest. Funct Ecol 24:217–223
- 613 IPCC (2012) Managing the risks of extreme events and disasters to advance climate change adaptation. A special
   614 report of working Groups I and II of the intergovernmental panel on climate change. In: Field CB, Barros
   615 V, Stocker TF et al (eds) Cambridge University Press, Cambridge, UK, and New York, NY, USA, 582 pp
- 616 ISTA (International Society of Testing Analysis) (1999) International rules for seed testing. Seed Sci Technol 617 27(supplement):201–244
- Kleyer M, Bekker RM, Knevel IC, Bakker JP, Thompson K, Sonnenschein M, Poschlod P, van Groenendael JM,
  Klimes L, Klimesová J, Klotz S, Rusch GM, Hermy M, Adriaens D, Boedeltje G, Bossuyt B, Endels P,
  Götzenberger L, Hodgson JG, Jackel A-K, Dannemann A, Ku "hn I, Kunzmann D, Ozinga WA,
  Römermann C, Stadler M, Schlegelmilch J, Steendam HJ, Tackenberg O, Wilmann B, Corne-lissen JHC,
  Eriksson O, Garnier E, Fitter A, Peco B (2008) The LEDA Traitbase: a database of plant life-history traits
  of North West European Flora. J Ecol 96:1266–1274
- Klimešová J, Klimeš L (2006) CLO-PLA3: a database of clonal growth architecture of Central-European
   plants.http://clopla.butbn.cas.cz
- Knevel IC, Bekker RM, Kunzmann D, Stadler M, Thompson K (2005) The LEDA Traitbase collecting and
   measuring standards of life-history traits of the NW European flora. Published by the University of
   Groningen, Groningen
- Kwiatkowski W, Venanzoni R (1994) Carta dei suoli della riserva naturale di Torricchio. La Riserva Naturale di Torricchio 9:15–21
- Leishman MR, Westoby M (1994) The role of seed size in seedling establishment in dry soil conditions –
   experimental evidence from semi-arid species. J Ecol 82:249–258
- Lemke I, Kolb A, Diekmann M (2012) Region and site conditions affect phenotypic trait variation in five forest
   herbs. Acta Oecol 39:18–24
- Linhart YB (1988) Intra-population differentiation in annual plants. III. The contrasting effects of intra- and inter-specific competition. Evolution 42:1047–1064
- Linhart YB, Grant MC (1996) Evolutionary significance of local genetic differentiation in plants. Annu Rev Ecol
   Syst 27:237–277
- Mawdsley JR, O'Malley R, Ojima DS (2009) A Review of climate-change adaptation strategies for wildlife
   management and biodiversity conservation. Conserv Biol 2:1080–1089
- 641 Maillette C (1992) Seasonal model of modular growth in plants. J Ecol 80:123–130
- Mazer SJ (1989) Ecological, taxonomic and life history correlates of seed mass among Indiana Dune
   angiosperms. Ecol Monogr 59:153–175
- McCarthy MC, Enquist BJ (2007) Consistency between an allometric approach and optimal partitioning theory
   in global patterns of plant biomass allocation. Funct Ecol 21:713–720
- 646 Miner BG, Sultan SE, Morgan SG, Padilla DK, Relyea RA (2005) Ecological consequences of phenotypic 647 plasticity. Trends Ecol Evol 20(12):685–692
- 648 Moles AT, Westoby M (2006) Seed size and plant strategy across the whole life cycle. Oikos 113:91–105
- Murray K, Conner MM (2009) Methods to quantify variable importance: implications for the analysis of noisy
   ecological data. Ecology 90:348–355
- Nicotra AB, Atkin OK, Bonser SP, Davidson AM, Finnegan EJ, Mathesius U, Poot P, Purugganan MD, Richards CL, Valladares F, van Kleunen M (2010) Plant phenotypic plasticity in a changing climate.
- 653 Trends Plant Sci 15:684–992
- 654 Oborny B, Bartha S (1995) Clonality in plant communities An overview. Abstracta Botanica 19:115–127

- Pakeman RJ, Garnier E, Lavorel S, Ansquer P, Castro H, Cruz P, Dolezal J, Eriksson O, Freitas H, Golodet C, Kigel J, Kleyer M, Leps J, Meier T, Papadimitriou M, Papanastasi VP, Quested H, Quetier F, Rusch G, Sternberg M, Theau JP, Thebault A, Vile D (2008) Impact of abundance weighting on the responseof seed traits to climate and land use. J Ecol 96:355–366
- Peat HJ, Fitter AH (1994) Comparative analyses of ecological characteristics of British angiosperms. Biol Rev
   660 69:95–115
- Pierce S, Luzzaro A, Caccianiga M, Ceriani RM, Cerabolini B (2007) Disturbance is the principala-scale filter
   determining niche differentiation, coexistence and biodiversity in an alpine community. J Ecol 95:698–
   706
- 664 Pignatti S (1982) Flora d'Italia. Edagricole, Bologna
- Poorter H, Pepin S, Rijkers T, De Jong Y, Evans JR, Körner C (2006) Construction costs, chemical composition,
   and payback time of high and low irradiance leaves. J Exp Bot 57:355–371
- Poorter H, Niinemets Ü, Poorter L, Wright IJ, Villar R (2009) Causes and consequences of variation in leaf mass
   per area (LMA): a meta-analysis. New Phytol 182:565–588
- R Development Core Team (2010) R: A language and environment for statistical computing. R version 2.13.1.
   Foundation for Statistical Computing. Vienna, Austria, http://www.R-project.org
- Reich PB, Wright IJ, Cavender-Bares J, Craine JM, Oleksyn J, Westoby M, Walters MB (2003) The evolution of
   plant functional variation: traits, spectra and strategies. Int J Plant Sci 164:S143–S164
- Ryser P, Urbas P (2000) Ecological significance of leaf life span among Central European grass species. Oikos
   91:41–50
- Schellenberg J, Pontes S (2012) Plant functional traits and nutrient gradients on grassland. Grass Forage Sci
   676 67:305-319
- 677 Silvertown J (2004) Plant coexistence and the niche. Trends Ecol Evol 19:605–611
- 678 Simms EL (1990) Examining selection on the multivariate phenotype: plant resistance to herbivores. Evolution
   679 44:1177–1188
- Stanton ML (1984) Seed variation in wild radish: effect of seed size on components of seedlings and adult
   fitness. Ecology 65:1105–1112
- Stöcklin J (1992) Umwelt, Morphologie und Wachstumsmuster klonaler Pflanzen eine Übersicht. Bot Helv
   102:3–21
- Sultan SE (1987) Evolutionary implications of phenotypic plasticity in plants. In: Hecht MK, Wallace B, Prance
   GT (eds) Evolutionary Biology, Vol 21. Springer, New York, pp 127–178
- Tissue DT, Nobel PS (1988) Parent-ramet connections in Agave desert: influences of carbohydrates on growth.
   Oecologia 75:266–271
- Turkington R (1989) The growth, distribution, and neighbour relationships of Trifolium repens in a permanent pasture. V. The coevolution of competitors. J Ecol 77:717–733
- Turkington R, Aarssen LW (1984) In: Dirzo R, Sarukhan J (eds) Local scale differentiation as a result of competitive interactions. Sunderland, MA, Perspectives in Plant Population Ecology. Sinauer, pp 107–127
- Venanzoni R, Kwiatkowski W (1995) Analisi integrata del paesaggio in un settore dell'Appennino Centrale
   (Riserva naturale Montagna di Torricchio). Colloq Phytosociol 24:187–201
- Violle C, Enquist BJ, McGill BJ, Jiang L, Albert CH, Hulshof C, Jung V, Messier J (2012) The return of the variance: intraspecific variability in community ecology. Trends Ecol Evol 27(4):244–252
- Walck JL, Hidayati SN, Dixon KW, Thompson K, Poschlod P (2011) Climate change and plant regeneration
   from seed. Glob Change Biol 17:2145–2161
- Webb CT, Hoeting JA, Ames GM, Pyne MI, Poff NL (2010) A structured and dynamic framework to advance
   traits-based theory and prediction in ecology. Ecol Lett 13:267–283
- Wellstein C (2012) Seed–litter–position drives seedling establishment in grassland species under recurrent drought. Plant Biol 14:1006–1010
- Wellstein C, Kuss P (2011) Diversity and frequency of clonal traits along natural and land-use gradients in grasslands of the Swiss Alps. Folia Geobot 46:255–270
- Wellstein C, Schröder B, Reineking B, Zimmermann NE (2011) Understanding species and community response
   to environmental change A functional trait perspective. Agr Ecosyst Environ 145:1–4
- West-Eberhard MJ (2005) Phenotypic accommodation: adaptive innovation due to developmental plasticity. J
   Exp Zool B 304:610–618
- Westoby M (1998) A leaf-height-seed (LHS) plant ecology strategy scheme. Plant Soil 199:213–227
- Westoby M, Falster DS, Moles AT, Vesk PA, Wright IJ (2002) Plant ecological strategies: some leading
   dimensions of variation between species. Annu Rev Ecol Syst 33:125–159
- Wilson PJ, Thompson K, Hodgson JG (1999) Specific Leaf Area and Leaf Dry Matter Content as Alter-native
   predictors of Plant Strategies. New Phytol 143:155–162
- Wright IJ, Westoby M (1999) Differences in seedling growth behaviour among species: trait correlations across
   species, and trait shifts along nutrient compared to rainfall gradients. J Ecol 87:85–97

	Site N	Site S
Coordinates	42°57′21.8″N-13°01′10.1″E	42°57'35.1"N-13°01'07.4"E
Elevation (m)	1200	1165
Exposition	320°N-NW	140°S-SE
Slope (°)	35	17
Soil	Rendzina on limestone	Lithosols and regosols, highly eroded soil
Vegetation	Sesleria nitidae–Brometum erecti	Asperulo purpureae-Brometum erecti
Series	Montane calcicolous Fagus sylvatica series	Montane calcicolous Ostrya carpinifolia series on steep rocky slopes

Table 1 Geo-physical and plant-sociological characteristics of the sampling sites (i.e. habitats)

Both associations belong to the order Brometalia erecti W. Koch 1926, class Festuco-Brometea Br.-Bl. et Tx. in Br.-Bl. 1949

	Life-history processes	Trait- environment links	Definition	Functional significance	Literature
Leaf traits	III CO				TO TOUL
SLA	competition moisture, area to leaf dry mass rate and mass-based maximum photosy the relationships between SLA and light temperature leaf defence structures and environment		Positively correlated with potential relative growth rate and mass-based maximum photosynthetic rate; the relationships between SLA and light condition, leaf defence structures and environmental stresses are well known. Plasticity of SLA is well documented	Kleyer et al. (2008); Poorter et al. (2006); Cornelissen et al. (2003); Wilson et al. (1999); Westoby (1998)	
LDMC	IC Establishment; Nutrients; The ratio of dry leaf mass competition temperature to fresh leaf mass		The ratio of dry leaf mass to fresh leaf mass	Used as a measure of leaf tissues density, which plays a central role in the use of nutrients by the species, determining the rate of turnover of biomass. At high values of LDMC, the rate of potential growth is low and the leaves tend to have a low turn-over to conserve nutrients, a big advantage in unproductive environments	Cornelissen et al. (2003); Kleyer et al. (2008); Ryser and Urbas (2000); Gross et al. (2007)
Whole plant trai	ts				
Plant height	Competition	Nutrients; moisture; temperature	The distance between the highest photosynthetic tissues and the base of the plant	Associated with competitive vigour, whole plant fecundity and generation time after disturbance, There are also important trade-offs between plant height and tolerance or avoidance of environmental stress	Kleyer et al. (2008); Cornelissen et al. (2003); Westoby (1998)
Horizontal stem length	Competition	Light; nutrients; moisture; temperature	Maximum length of the horizontal stems of the plant	A measure of the maximum range of space occupancy. Lateral spread of horizontal stems enables the plant to place photosynthetic and reproductive organs at the microscale to new environments. This might help to minimize competition. Additionally, maximal lateral spread of sexual and vegetative reproduction is associated with increasing distribution of new individuals	Silvertown (2004); Klimešová and Klimeš (2006)

Table 2 Traits definitions, respective life-history processes, and functional information from literature

Table 2 continued

	Life-history processes	Trait- environment links	Definition	Functional significance	Literature	
Clonal traits						
No. of bifurcations/ nodes per cm	Competition; regeneration	Nutrients; moisture; temperature	No. of bifurcations or nodes per cm of a plants stem	A measure of the multiplication capacity which is associated with the abundance of space occupancy, i.e. density/packing of individuals/ramets	Knevel et al. (2005); Maillette (1992)	
Distance between bifurcations/ nodes	tween regeneration nutrients; bifurcations or not furcations/ moisture; a plants stem des temperature		Distance between the bifurcations or nodes of a plants stem	A measure of the capacity of space occupancy by clonal growth. Lateral spread enables the plant to colonize a new substrate and avoid intraspecific competition. Limited lateral spread could be expected in situations where facilitation is important	Klimešová and Klimeš (2006)	
Seed traits						
Seed mass Establishment; Moisture; dispersal nutrients; temperature		The air dried weight of seeds	Resources stored in large seeds help the young seedling to survive and establish in the face of environmental hazards. Connection with dispersion and persistence in the soil	Kleyer et al. (2008); Cornelisser et al. (2003); Westoby (1998); Westoby et al. (2002).		
Seed germination	Establishment	Moisture; temperature	The percentage of seeds germinated under defined environmental conditions	Strongly associated with species ecology and competition. The relative performance of individual plants during the early stages of life, i.e. germination and seedlings establishment, can have important effects on subsequent adult growth and fitness	Baskin and Baskin (1998); Stanton (1984); Walck et al. (2011); Wellstein (2012)	

		Main effect of site		Main effect of speci	ies' place of growth			
	Site N + S	Site N	Site S	Site N + S				
Place of growth of individuals of respecti species	All species	All species		Sesleria nitida	Lotus corniculatus	Astragalus sempervirens	Thymus longicaulis	
Environmental variable and resources	les Mean and SD	Mean and SD	Mean and SD	Mean and SD	Mean and SD	Mean and SD	Mean and SD	
Light-PAR (%)	74.69 ± 25.72	65.85 ± 30.38 a	83.52 ± 15.78 b	48.10 ± 20.33 b	86.59 ± 21.73 a	74.17 ± 24.78 c	90.08 ± 8.65 a	
Soil temperature (°C)	$18.59 \pm 4.94$	14.62 ± 1.74 a	22.66 ± 3.68 b	15.08 ± 2.58 c	18.58 ± 4.70 a	18.74 ± 4.53 a	21.97 ± 5.05 b	
Soil moisture (%)	$10.98 \pm 6.46$	12.55 ± 6.84 a	9.37 ± 5.63 b	14.66 ± 4.81 a	13.68 ± 5.79 a	13.13 ± 3.22 a	2.41 ± 2.26 b	
pH	$6.67 \pm 0.41$	$6.60 \pm 0.50$ a	6.95 ± 0.14 b	6.90 ± 0.31 b	6.74 ± 0.60 ab	6.72 ± 0.32 a	6.73 ± 0.31 a	
Conductivity (µS/cm)	195.69 ± 61.72	220.84 ± 64.62 a	169.69 ± 46.05 b	203.20 ± 60.94 ns	182.06 ± 54.37 ns	197.14 ± 54.12 ns	201.22 ± 74.52 ns	
K (mg/100 g)	$7.42 \pm 3.65$	9.11 ± 3.67 a	5.68 ± 2.70 b	7.83 ± 3.54 ns	7.10 ± 3.18 ns	7.60 ± 4.29 ns	7.20 ± 3.60 ns	
Mg (mg/100 g)	$7.56 \pm 2.30$	$7.92 \pm 2.35 a$	6.54 ± 1.73 b	6.83 ± 1.29 b	7.49 ± 2.20 ab	7.60 ± 2.51 ab	8.31 ± 2.72 a	
P (mg/100 g)	$0.12 \pm 0.14$	$0.20 \pm 0.13$ a	$0.03 \pm 0.11 \text{ b}$	$0.10 \pm 0.11$ a	$0.09 \pm 0.11$ a	0.18 ± 0.19 b	$0.11 \pm 0.14$ a	
NH4 (mg/100 g)	$0.81 \pm 0.53$	1.09 ± 0.59 a	$0.52 \pm 0.23$ b	0.90 ± 0.49 a	$0.87 \pm 0.56$ a	$0.96 \pm 0.64 a$	$0.52 \pm 0.25$ b	
NO3 (mg/100 g)	$3.30 \pm 4.14$	4.81 ± 4.79 a	1.74 ± 2.54 b	6.01 ± 4.33 c	4.25 ± 5.41 b	2.19 ± 1.94 b	$0.72 \pm 0.48 a$	
C/N (%)	$10.81 \pm 0.78$	$11.27 \pm 0.57$ a	$10.33\pm0.67~\mathrm{b}$	$10.96 \pm 0.57 a$	10.78 ± 0.59 ab	$10.54\pm0.77~b$	$10.95 \pm 1.04$ a	
Interaction of site and	species' place of gro	wth						
Site N Sesleria nitida	Site S	Site N Lotus corniculatus	Site S	Site N Astragalus semper	Site S virens	Site N Thymus longicauli:	Site S	
Mean and SD	Mean and SD	Mean and SD	Mean and SD	Mean and SD	Mean and SD	Mean and SD	Mean and SD	
34.40 ± 18.63 a	61.80 ± 10.27 b	80.38 ± 27.61 a	93.01 ± 10.24 b	52.35 ± 15.70 a	95.26 ± 7.35 b	95.84 ± 7.26 a	84.33 ± 5.56 b	
12.60 ± 0.28 a	17.56 ± 0.87 b	$14.36 \pm 0.64 a$	23.32 ± 1.58 b	14.46 ± 0.66 a	$23.03 \pm 1.83 \text{ b}$	17.19 ± 0.72 a	26.75 ± 2.01 b	
16.58 ± 5.44 a	12.67 ± 3.05 b	17.56 ± 5.05 a	9.63 ± 2.66 b	$12.42 \pm 1.70$ ns	13.85 ± 4.14 ns	3.27 ± 1.23 a	$1.54 \pm 0.41$ b	

Table 3 Differences of environmental variables at the place of growth of individuals between species and sites (i.e. habitats)

## Table 3 continued

Interaction of site and species' place of growth

Site N Sesleria nitida	Site S	Site N Lotus corniculatus	Site S	Site N Astragalus semperv	Site S virens	Site N Thymus longicaulis	Site S
Mean and SD	Mean and SD	Mean and SD	Mean and SD	Mean and SD	Mean and SD	Mean and SD	Mean and SD
6.75 ± 0.36 ns	7.06 ± 0.14 ns	$6.55 \pm 0.77$ ns	6.94 ± 0.13 ns	6.55 ± 0.36 ns	$6.88 \pm 0.14$ ns	6.55 ± 0.34 ns	6.92 ± 0.09 ns
219.07 ± 63.81 ns	186.79 ± 54.11 ns	194.18 ± 60.98 ns	$168.73 \pm 43.20$ ns	218.52 ± 50.04 a	175.76 ± 50.13 b	254.17 ± 69.76 a	148.27 ± 25.13 b
9.74 ± 3.28 a	5.86 ± 2.61 b	7.91 ± 3.19 ns	6.21 ± 2.96 ns	10.39 ± 4.22 a	4.80 ± 1.86 b	8.57 ± 3.58 a	5.84 ± 3.12 b
7.46 ± 0.93 ns	6.18 ± 1.29 ns	7.87 ± 2.56 ns	7.07 ± 1.67 ns	9.17 ± 1.97 a	6.03 ± 1.98 b	9.77 ± 2.76 a	6.86 ± 1.76 b
$0.17 \pm 0.09 a$	$0.02 \pm 0.05$ b	$0.14 \pm 0.12$ a	0.03 ± 0.06 b	0.27 ± 0.14 a	0.08 ± 0.19 b	0.22 ± 0.12 a	$0.01 \pm 0.04 \text{ b}$
$1.17 \pm 0.54$ ns	$0.64 \pm 0.19$ ns	$1.18 \pm 0.62$ ns	$0.53 \pm 0.15$ ns	$1.35 \pm 0.64$ ns	$0.56 \pm 0.31$ ns	0.68 ± 0.25 ns	$0.36 \pm 0.12$ ns
7.11 ± 4.76 ns	4.87 ± 3.57 ns	7.60 ± 5.69 a	$0.57 \pm 0.28$ b	3.35 ± 2.10 a	$1.03 \pm 0.66$ b	0.85 ± 0.59 ns	$0.58 \pm 0.31$ ns
11.23 ± 0.38 a	10.67 ± 0.60 b	10.99 ± 0.46 a	10.55 ± 0.63 b	$11.11 \pm 0.49$ a	9.97 ± 0.54 b	11.65 ± 0.64 a	$10.14 \pm 0.67$ b

Significant differences between species and sites were evaluated by two-way ANOVA followed by post hoc test (Tukey HSD). Main effect of site, species' place of growth, and their interaction are given subsequently. For the interaction effects, significant differences between sites are noted within species (not between species as this was not of interest). Significant differences are indicated by different lower case letters in bold; *ns* not significant

		Leaf traits		Whole plant traits		Clonal traits	Seed traits		
		SLA (mm <sup>2</sup> mg <sup>-1</sup> )	LDMC (mg g <sup>-1</sup> )	Height (cm)	Horizontal stem length (cm)	No. of bifurcations/ nodes per cm (n)	Distance between bifurcations/nodes (cm)	Seed mass (mg)	Seed germination (%)
Sesleria nitida	N	18.04 a	375.03 a	60.33 a	23.35 a	0.09 a	7.93 a	1.80 a	2.00 a
	S	15.77 b	408.61 b	51.00 b	8.37 b	0.46 b	2.88 b	2.30 b	8.00 b
		2.27	33.58	9.33	14.98	0.37	5.05	0.50	6.00
Lotus	Ν	38.43 a	227.47 a	20.33 a	18.69 a	0.10 a	1.99 a		
corniculatus	S	24.32 b	254.98 b	13.10 b	16.90 b	0.13 a	3.25 a	No data	No data
	$\triangle$	14.11	27.51	7.23	1.79	0.03	1.26		
Astragalus	N	17.31 a	343.48 a	12.80 a				2.30 a	2.00 a
sempervirens	S	15.26 b	371.56 b	9.17 b	No data	Non clonal	Non clonal	2.80 b	11.00 b
	Δ	2.05	28.08	3.63				0.30	9.00
Thymus	N	14.44 a	361.06 a	6.97 a	21.10 a	0.30 a	2.64 a	0.053 a	29.00 a
longicaulis	S	10.80 b	367.81 a	8.40 b	13.25 b	0.28 a	3.04 a	0.070 b	33.00 a
	Δ	3.64	6.75	1.43	7.85	0.02	0.40	0.017	4.00

Table 4 Mean trait values (untransformed data) per species in each slope (N,S), and their difference (△)

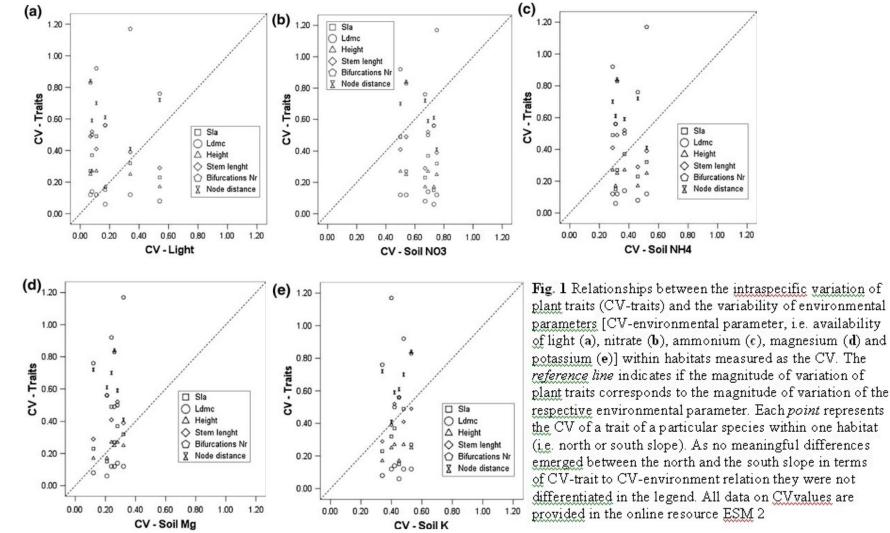
Significant differences between means (p < 0.05) were evaluated by ANOVA (all traits of *S. nitida* and *A. sempervirens*; seed traits of all species) and ANCOVA using biomass as covariate (leaf, whole plant, and clonal traits of *L. corniculatus* and *T. longicaulis*). Significant differences are indicated by different lower case letters in bold. Identical lower case letters indicate homogeneous groups, i.e. no significant difference

Table 5 Intraspecific trait-environment relationships of the study species within habitats (i.e. N: north-facing slope, S: south-facing slope)

	Lotus co	miculatus	Astragalus s	empervirens	Thymus	longicaulis	Sesleria nitida	
Habitat	s	N	s	N	s	N	s	N
SLA	n.s.	n.s.	NO3(+)*	n.s.	n.s.	n.s.	n.s.	n.s.
LDMC	n.s.	n.s.	n.s.	n.s.	n.s.	K(-)* Mg(+)*	n.s.	n.s.
Height	light(+)** temp (-)* CN(+)*	temp (-)**	temp (-)**	n.s.	n.s.	n.s.	n.s.	n.s.
Horizontal stem lenght	n.s.	n.s.	no data	no data	n.s.	NH4(-)*	n.s.	Mg(+)** K(-)* P(-)* cond(-)*
N. of bifurcations / nodes per cm	n.s.	n.s.	non cional	non clonal	n.s.	light(+)* NO3(-)* moist(-)* K(-)*	n.s.	n.s.
Distance between bifurcations / nodes	pH(-)** temp(+)*	n.s.	non cional	non clonal	n.s.	K(+)* Mg(+)* P(+)*	n.s.	n.s.

significance: \*\* - p<0.01; \* - p<0.05; . - p<0.1

relative importance: >20% dark grey; 20%>x>10% grey



1.20

 $\begin{array}{c} 722\\ 723 \end{array}$