

1 Assessment of Consequences of Notifiable Fish Disease Incursions in
2 England and Wales

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1 **ABSTRACT**

2 A consequence assessment framework was developed to evaluate the economic and
3 environmental consequences of an exotic disease in the context of supporting policy level
4 decisions on mitigation strategies. The framework adopted a semi-qualitative analysis of
5 impacts supported by expert judgement. The efficacy of the framework was illustrated via
6 assessment of the notifiable fish disease, *Gyrodactylus salaris*. In this example, the economic
7 cost of an illustrative outbreak of *G. salaris* ranged from £0.22 million to £90 million. The cost
8 of the most likely scenario (regional spread) was estimated to be £7.5 million (minimum to
9 maximum range of £2-22 million), reflecting the uncertainty in the extent of spread of the
10 parasite before detection. The environmental impacts varies by a factor of 35 between incursion
11 scenarios reflecting the number of affected catchments in the scenarios.

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13 **Key Words:** consequence assessment, *Gyrodactylus salaris*, aquatic fish disease, incursion

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1 **INTRODUCTION**

2 During the global trade of live aquatic animals and their commodities, transboundary
3 movement presents significant routes of disease introduction inflicting various economic,
4 environmental and/or socio-economic impacts (Minchin, 2007). There is a range of regulatory
5 and legislative guidance in this area, paramount being the Sanitary and Phytosanitary (SPS)
6 Agreement of the World Trade Organisation (WTO, 1995) providing basic rules for international
7 trade governing food safety, animal and plant health while maintaining a free and safe trade
8 environment among Member states. Key to this agreement is the import risk analysis (IRA)
9 framework determining the level of sanitary measures necessary to provide the safe trade of
10 particular species. Guidelines for its implementation are outlined by the World Organisation for
11 Animal Health (OIE, 2004) and suggest that IRAs assess the consequence of a disease incursion
12 and an estimation of the likelihood of pathogen introduction and establishment. Currently, the
13 UK is free of a number of notifiable diseases listed by the OIE (i.e. those that must be reported
14 by law) that are prevalent in continental European aquaculture (e.g. viral haemorrhagic
15 septicaemia and infectious haematoptic necrosis). Statutory controls are in place to protect the
16 UK's 'disease free' status; for example, live fish can only be imported from areas with the same
17 high health status. The disease control policies used for a particular outbreak (e.g. movement
18 restrictions or culling) are the main predictor of the cost of the process and so consequence
19 assessment plays an important role in evaluating the cost-benefit assessment of intervention to
20 reduce and manage risk.

21 To date, no example of a structured method for rapidly assessing the consequences of an
22 exotic disease incursion of an aquatic animal pathogen exists. However, consequence
23 assessments have been developed for other risk domains including for natural disasters –

1 volcanoes and earthquakes (Granot, 1995; Magill and Blong, 2005), mining (Hutchinson *et al.*,
2 2002) and terrestrial animal disease incursions (Moutou *et al.*, 2001; Australian Department of
3 Agriculture, 2004). As a result, in practice the consequence assessment element of IRA for fish
4 disease often receives minimal attention, sometimes reduced to a few paragraphs of narrative or
5 is excluded completely (Peeler *et al.*, 2007a). This may be due to the lack of structured
6 methodology or genuine data deficiency (Edgerton, 2002).

7 Here, we propose a method for assessing the consequences of exotic notifiable disease
8 incursions for England and Wales (E&W). The method is designed for use by the Centre for
9 Environment, Fisheries and Aquaculture Sciences (Cefas) whose role it is to oversee aquatic
10 animal health and respond to exotic disease outbreaks within E&W.

11

12 **Background**

13 There is merit in reviewing the decision context, so as to inform the discussion of the
14 proposed approach and case study later in this paper. Decisions about notifiable disease
15 incursions are complex and generally suffer from a paucity of data (Pharo, 2004). Decision-
16 makers (policy and regulatory officials, for example) usually benefit from access to high quality
17 information to inform decisions on intervention strategies to reduce risk. Usually, these
18 decisions require some justification of their benefits in the context of the costs born in the event
19 of a disease outbreak. An evaluation of economic as well as environmental impacts is valuable.
20 The proposed framework presented below operates within the aforementioned regulatory
21 systems and builds upon the SPS agreement's vision of consequence assessment (one currently
22 dominated by economic assessment (Macdiarmid and Pharo, 2003)), by broadening the
23 assessment to include environmental impacts. Consequence metrics are aligned with the fish

1 health legislation in England and Wales (AAHR, 2009), which is largely determined by the
2 European Council Directive 2006/88 on animal health requirements for aquaculture animals, and
3 assesses scale and cost with respect to farmed and wild fin fish species.

4 Currently there is no formal framework used by UK Government to assess the
5 consequences of exotic disease incursions. Instead, outbreak scenarios have been employed to
6 assess the likelihood of outbreaks of varying geographic scale (*personal communication* EJ
7 Peeler, Cefas); a method also used by the Australian Department of Agriculture, Fisheries and
8 Forestry (2004). The framework presented below was developed fit-for-purpose and intended to
9 provide a structured and transparent approach for ‘high-level’ assessment of fish disease, thus
10 providing strategy and policy makers a standardised perspective of consequence. Due to a lack
11 of fish disease examples in the UK, supporting data is minimal and therefore the framework
12 relies upon expert judgement. .The intention is to provide a framework that improves
13 organisational capacity for comparing the consequences of different exotic diseases, enabling
14 improved decision making. We believe this to be the first example of a framework for rapidly
15 assessing the impact of an aquatic notifiable disease incursion in an industrial/commercial
16 context and illustrate the efficacy of the framework using an example outbreak of *Gyrodactylus*
17 *salaries*, a freshwater parasite of salmonids.

18

19 **METHODS**

20

21 **Rationale for approach and treatment of consequences**

22 The literature on fish disease consequence assessment is not extensive and methods for
23 describing the impacts of aquatic disease were linked to the wider body of risk literature, most

1 notably those applied to characterising environmental harm (Pollard *et al.*, Willis *et al.*, 2004). A
2 key consideration for the successful development and implementation of the framework was the
3 alignment of consequence attributes with organisational objectives and the need for a pragmatic
4 compromise between usability and depth of description (Willis *et al.*, 2004; HCN, 1996).
5 Environmental and economic consequences used to assess the impact of a disease incursion were
6 developed through an elicitation process with experts from Cefas. Experts were chosen based on
7 their familiarity with the subject material and capacity for providing a ‘high-level’ assessment of
8 the issue such that the methodology developed would have broad utility across a range of exotic
9 diseases.

10 The main economic driver identified was the impact on aquaculture trade. However, costs
11 are also incurred during disease eradication. Disease control regulations can, for example,
12 instruct the destruction of infected stock (AAHR, 2009) leading to short term production losses
13 and subsequent reduction in future production due to loss of juveniles. These consequences are
14 readily quantifiable based on market values, as are the impacts of mitigation activities (*e.g.*
15 disinfection, site clean-up) and lost productivity (Macdiarmid and Pharo, 2003). An
16 economically-focused viewpoint of consequences (Table 1) provides easily comparable values
17 between different types of disease incursions. However, such an approach is limited in providing
18 a broader overview of the wider range of impacts associated with an incursion. Below we
19 address economic and environmental consequences in turn.

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{Table 1 about here please}

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1 Key economic aspects of aquaculture were disaggregated to provide close detail about
2 consequence. A key differentiator is “fish utility” (farmed and wild species), the rationale for
3 which is the differences in disease management activities between the two types of utility.
4 Moreover, some diseases do not affect farmed and wild species equally (Peeler *et al.*, 2007b) and
5 decision-makers require further detail to provide an accurate representation of disease
6 consequence. Further disaggregation enables the assessor to make a distinction between
7 economic impacts to infected water bodies (inf) and those suspected of infection (susp) as each
8 will initiate a different regulatory response (AAHR, 2009). A farm may fall under suspicion if it
9 has an epidemiological link with an infected farm (e.g. within the same river catchment or
10 received movements of fish). Finally, differentiation is given to farms producing fish for
11 consumption (table) versus those that produce fish for restocking of river or still-water fisheries
12 (resto). This distinction accounts for economic differences of particular mitigation strategies,
13 dependent upon a farm’s status (Table 2) as well as the end value of the fish.

14

15 {Table 2 about here please}

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17 Environmental consequences are less tangible and difficult to quantify, though remain
18 vital for ensuring a well informed decision. Capture of environmental information relies upon a
19 qualitative assessment and expert judgement. Here, we determined that the environmental
20 consequences needed to be further fragmented to consider environmental and ecological
21 attributes that emphasised the impact on the wider community of aquatic species. Whilst these
22 were by no means inclusive of all possible environmental consequences, the process of

1 considering such broad consequences as ‘knock-on ecological effects’ will provide much greater
2 value to decision makers than a simple statement of expected monetised stock at risk.

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4 **Populating the framework**

5 The framework was populated by extrapolating economic information from market prices
6 and outbreak scenarios, as well as the opinion of experts within Cefas. Experts were asked to
7 determine the extent of economic harm arising from an outbreak. Where available, stochastic
8 distributions were used to represent the variability in cost estimates and take into account the size
9 of a potential outbreak, the size of aquaculture farms and cost variability. These distributions
10 were gathered from in house Cefas data and from expert extrapolation of predicted outcomes. To
11 account for an incursion that affects more than one farm or catchment, stochastic values were
12 multiplied by the number of farms or catchments affected in line with the outputs of the outbreak
13 scenarios. Aggregating all economic values provided an overall assessment of economic
14 consequence.

15 Environmental consequences are difficult to quantify and therefore a semi-quantitative
16 approach was adopted using scalar values linked to qualitative descriptors (Table 3). This
17 allowed aggregation of all environmental consequences into a single semi-quantitative output,
18 thus enabling a relative comparison between two or more disease agents (Granot, 1995;
19 Australian Department of Agriculture, 2004). Attribute weighting was introduced to assign an
20 importance to each environmental descriptor – often due to regulation (e.g. protected species).
21 Weights ranged from 1 (low importance) to 5 (high importance) and were assigned by Cefas
22 experts informed by AAHR guidelines (AAHR, 2009). Ranges were assigned using expert
23 consultation and with reference to existing outbreak scenarios. To illustrate the efficacy of the

1 framework an assessment of the consequences of an incursion of *Gyrodactylus salaris* was
2 performed.

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4 {Table 3 about here please}

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6 **RESULTS**

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8 **Case study - an illustrative assessment for *Gyrodactylus salaris***

9 An assessment of the likely impact of *Gyrodactylus salaris*, a freshwater parasite of
10 salmonids, was used to explore the efficacy and appropriateness of the framework. *G. salaris* is a
11 freshwater parasite not found in the UK. The limited available data (Mackenzie, 1993) indicates
12 that Atlantic salmon populations in the UK are highly susceptible to *G. salaris* and thus, if
13 infected, are likely to experience population declines similar to those observed in Norway
14 following introduction of the parasite from Sweden (Johnsen and Jensen, 1991). In consultation
15 with experts at Cefas, four outbreak scenarios for *G. salaris* were developed and likelihoods
16 ascribed (Table 2). The economic consequence of an outbreak of *G. salaris* for the defined
17 attributes is presented in Table 3. The outbreak scenarios and economic consequences were then
18 combined to provide a total economic cost (minimum, maximum and most likely) for each
19 outbreak scenario (Table 4).

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22 {Table 4 about here please}

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1 Elicitation sessions of experts from within Cefas were used to develop a general
2 weighting for each environmental consequence. Weightings were considered fixed for all
3 disease outbreaks. Experts assigned scores for all environmental attributes of an incursion of *G.*
4 *salaris* (Table 5). The magnitude of the environmental consequence was a product of the
5 weighting and ranking score, the sum of which gives an overall environmental consequence
6 score for *G. salaris* (OCS) of 50. This was then scaled by the number of affected catchments to
7 give a consequence score for each scenario (Table 6). Five economic attributes (Table 2) were
8 considered by experts as being ‘cost independent of pathogen’, meaning the stochastic
9 distributions did not significantly vary with the pathogen of concern. The main cost for
10 surveillance and supervision was manpower, which is unlikely to vary substantially between
11 pathogens (diagnostic costs may vary but comprise a relatively small proportion of the total
12 costs) and therefore the cost to the competent authority does not change. The cost of disinfection,
13 loss of stock, restocking and loss of trade for the farm also do not vary between pathogens. As an
14 example, farms carry out the same disinfection process no matter what type of pathogen has
15 infected stocks. Similarly, the cost of destroying an entire stock is also unlikely to be affected by
16 different pathogens.

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18 {Table 5 about here please}

19 {Table 6 about here please}

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21 The economic attributes dependent on a particular pathogen included ‘economic impact
22 of a decrease in wild population levels’, ‘national market reaction’ and ‘loss of international
23 trade’. These attributes required re-estimation by experts for different pathogens because of the

1 need to account for changes in market values associated with different species. Re-estimation
2 may be necessary for diseases that affect only farmed and not wild species (or vice versa) and
3 this is to take account for effects on recreational fishing and tourism.

4

5 **DISCUSSION**

6 The method assesses economic and environmental consequences of a disease incursion in
7 the context of UK/EU regulation. Consequences to human health are excluded from the
8 assessment as there are no clear links to human health implications of finfish disease (OIE,
9 2009). Broader social consequences are captured from an economic perspective when assessing
10 the attribute ‘economic impact of a decrease in wild population levels’, which assesses the
11 economic loss due to a decrease on recreational fishing and tourism caused by a decline in wild
12 population levels. However, the attributes considered here do not consider the social impact on
13 the farming population or perception of the industry after an event.

14 From a pragmatic perspective the method assesses economic and environmental
15 consequences separately, which has the advantage of enabling extraction of specific information
16 dependent upon a user’s data requirements. This avoids the contentious topic of combining
17 quantitative and qualitative data (Granot, 1995). Our rationale was to provide decision makers
18 with a transparent presentation of data rather than a reductionist single score of consequence.
19 However, decision makers may wish to have economic and environmental consequences
20 combined, possibly by using a matrix, as a single value to more easily compare the impact of
21 different diseases. This is possible only where economic and environmental consequences are
22 commensurable and can be achieved by monetisation of non-marketed goods (i.e.
23 environmental). Stated preference methods, such as contingent valuation method (CVM) or

1 choice experiments generate data that describe a respondent's preferences and determines
2 society's valuation of environmental attributes (Green *et al.*, 1998; Veisten, 2007). Monetising
3 environmental consequences is challenging due to the small size of the freshwater fisheries
4 sector, which limits the volume of non-biased data one can obtain. Also, economic valuation
5 techniques may fail to capture broader values (Fischhoff, 2005) and there is a trend in economic
6 valuation to acknowledge the impossibility of monetising certain environmental services (TEEB,
7 2010). Therefore, we prefer a rapid evaluation of environmental consequences (as presented)
8 providing a depth of information that better informs decision processes.

9 The score for the environmental impact of an exotic disease incursion can be used to rank
10 the importance of exotic diseases, once the framework has been applied to the most prevalent
11 and/or severe exotic diseases. Such ranking may aid decision-makers in the allocation of
12 resources between competing activities; for example, in the selection of exotic pathogens for
13 targeted surveillance. A robust ranking of the diseases based on their likely environmental
14 impact (and information on the economic consequences and likelihood of introduction) is
15 generally sufficient for this purpose. Approximate monetisation of environmental impacts would
16 not significantly improve the evidence base for decision making.

17 Expert opinion is a useful for eliciting reliable risk data and has been shown to provide
18 valuable information regarding climatic impacts to aquatic species (McDaniels *et al.*, 2010).
19 Similar to climatic impact, the consequences of new disease outbreaks in England and Wales are
20 generally unknown and therefore suffer from a paucity of data. In this context, expert judgement
21 is often the only source of reliable, informed data and previous research has demonstrated the
22 efficacy of multi-disciplinary Delphi panels (including fish health specialists, economists and
23 ecologists) that are useful for collating information and generating consensus (Peeler *et al.*,

1 2006). However, given the small size of the sector and the demand for rapid assessment, access
2 to a range of experts is not frequently feasible. Under this constraint, a small panel of experts on
3 finfish disease and IRA was used to complete the assessment for *G. Salaries*, with their main role
4 being to provide expert synthesis and extrapolation of disease data taken from countries where
5 the disease is present. An obvious limitation is the potential for poor translation of
6 environmental impacts between regions or the introduction of bias. Though access to numerous
7 experts is advantageous, pragmatic compromise suggests that assessment be completed using the
8 resources available.

9 Overall, the framework provides a relative, rather than absolute, assessment of
10 consequence based upon best available data and expertise. Decision makers must be careful to
11 consider all available information because, like all risk tools, this framework will not provide a
12 definitive answer to the questions they raise. The case of a *G. salaris* disease outbreak served to
13 illustrate the efficacy of the consequence assessment framework. The economic cost of an
14 outbreak of *G. salaris* may range from £0.22 million to £90 million. The cost of the most likely
15 scenario (regional spread) was £7.5 million (minimum to maximum range of £2-22 million). The
16 range of costs reflects primarily the uncertainty about the extent of spread of the parasite before
17 detection. Similarly, the environmental impact varies by a factor of 35 which reflects the number
18 of affected catchments in the scenarios. The environmental consequences provide a non-
19 monetary value and therefore, to provide usable comparison to other diseases this method needs
20 to be consistently applied to all the major notifiable fish pathogens. In general, this case study
21 demonstrates the ease in which a formal, structured and transparent consequence assessment may
22 be generated and when applied to a variety of diseases will provide equitable evidence for
23 decision makers.

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CONCLUSION

The need to assess the consequence and risk of a fish disease incursion is well recognized. This paper presents a standardised approach for assessing consequence, suggesting suggested framework comprised of two components (economic and environmental). Challenged by a severe lack of data, the assessment was informed by experts from Cefas who provided technical and regulatory expertise. The methodology developed in this paper may be used for assessment and relative comparison of consequences for exotic aquatic animal disease incursions and thus, will is a useful tool for aquatic animal health management.

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REFERENCES

Australian Department of Agriculture, Fisheries and Forestry. 2004. Generic Import Risk Analysis for Pig Meat. Canberra, Australia.

Disease interactions and pathogen exchange between farmed and wild aquatic animal populations – a European network. 2006. Risk assessment and predictive modeling – a review of their application in aquatic animal health. Available at: <http://www.dipnet.info/docs/doc.asp?id=43>.

1
2 Edgerton, B.F. 2002. Hazard analysis of exotic pathogens of potential threat to European
3 freshwater crayfish. *B Fr Piscic* 3:801-820.
4 Fischhoff, B. 2005. Cognitive processes in stated preference methods. In: K.G. Maler and J.R.
5 Vincent (eds), *Handbook of Environmental Economics*, pp 937-968. Elsevier, North Holland.
6 Granot, H. 1995. Proposed scaling of communal consequences of disaster. *Disaster Prev*
7 *Manage.* 4:5-13.
8 Green, D., Jacowitz, K.E., Kahneman, D., and McFadden, D. 1998. Referendum contingent
9 valuation, anchoring and willingness to pay for public goods. *Resour Energy Econ.* 20:85-
10 116.
11 Health Council of the Netherlands (HCN). 1996. Risk is more than just a number. Committee on
12 risk measures and risk assessment. Publication no. 1996/03E. The Hague, The Netherlands.
13 Hutchinson, D.J., Phillips, C., and Cascante, G. 2002. Risk considerations for crown pillar
14 stability assessment for mine closure planning. *Geotech Geol Eng.* 20:41-63.
15 Johnsen, B.O., and Jensen, A.J. 1991. The *Gyrodactylus* story in Norway. *Aquaculture.* 98:289-
16 302.
17 MacDiarmid, S.C., and Pharo, H.J. 2003. Risk analysis: assessment, management and
18 communication. *Rev Sci Tech O.I.E.* 22:397-408.
19 MacKenzie, K. 1993. Susceptibility of native Scottish stocks of salmon to *Gyrodactylus salaris*.
20 In: K.T. Pirquet (ed), *Proceedings of the 10th Annual Meeting of the Aquaculture*
21 *Association of Canada*, pp105-107. Charlottetown, Canada.
22 Magill, C., and Blong, R. 2005. Volcanic risk ranking for Auckland, New Zealand. II: Hazard
23 consequences and risk calculation. *B Volcanol.* 67:340-349.

1 McDaniels, T., Wilmont, S., Healey, M., and Hinch, S. 2010. Vulnerability of Fraser River
2 sockeye salmon to climate change: A life cycle perspective using expert judgment. *J Environ*
3 *Manage.* 91:2771-2780.

4 Minchin, D. 2007. Aquaculture and transport in a changing environment: overlap and links in the
5 spread of alien biota. *Mar Pollut Bull.* 55:302-313.

6 Moutou, F., Dufour, B., and Ivanov, Y. 2001. A qualitative assessment of the risk of introducing
7 foot and mouth disease into Russia and Europe from Georgia, Armenia and Azerbaijan. *Rev*
8 *Sci Tech O.I.E.* 3:723-730.

9 Murray, A.G., and Peeler, E.J. 2005. A framework for understanding the potential for emerging
10 diseases in aquaculture. *Prev Vet Med.* 67:223-235.

11 Peeler, E.J., Murray, A.G., Thebault, A., Brun, E., Giovaninni, A., and Thrush, M.A. 2007a. The
12 application of risk analysis in aquatic animal health management. *Prev Vet Med.* 81:3-20.

13 Peeler, E.J., Murray, A.G., Thebault, A., Brun, E., Giovaninni, A., and Thrush, M.A. 2007b. The
14 application of risk analysis in aquatic animal health management. *Prev Vet Med.* 81:3-20.

15 Pharo, H. 2004. Acceptable risk in animal biosecurity import risk analysis: the New Zealand
16 experience. *Vet Ital.* 42:337-349.

17 Pollard, S.J.T., Kemp, R.V., Crawford, M., et al. 2004. Characterizing environmental harm:
18 Developments in an approach to strategic risk assessment and risk management. *Risk Anal.*
19 24:1551-1560.

20 TEEB. 2010. The economics of ecosystems and biodiversity: Ecological and economic
21 foundations. Pushpam Kumar (ed). Earthscan, London and Washington.

22 The Aquatic Animal Health (England and Wales) Regulations (AAHR). 2009. Available at:
23 http://www.opsi.gov.uk/si/si2009/uksi_20090463_en_1.

1 The Aquatic Animal Health (England and Wales) Regulations (AAHR). 2009. No.463. London,
2 HMSO.

3 The Council of the European Union. 2006. On animal health requirements for aquaculture
4 animals and products thereof, and on the prevention and control of certain diseases in aquatic
5 animals. Council Directive 2006/88/EC. Official Journal of the European Union.

6 Veisten, K. 2007. Contingent valuation controversies: Philosophic debates about economic
7 theory. *J Socio-Econ.* 36:204-232.

8 Willis, H.H., DeKay, M.L., Morgan, M.G., et al. 2004. Ecological risk ranking: Development
9 and evaluation of a method for improving public participation in environmental decision
10 making. *Risk Anal.* 24:363-378.

11 World Organization for Animal Health (O.I.E.). 2004. Aquatic Health Code. Paris, France.

12 World Organization for Animal Health (O.I.E.). 2009. Diseases notifiable to the O.I.E. Paris,
13 France..

14 World Trade Organization (W.T.O). 1995. Agreement on the Application of Sanitary and
15 Phytosanitary Measures. P 21. World Trade Organization, Geneva.

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Attributes affecting economic consequence magnitude	Definition	Who bears the cost
Farm level		
Surveillance by competent authority	Cost of sampling and diagnostics carried out by the government on infected farms and farms under suspicion of being infected to control the presence of a pathogen	Government
Disinfection of a farm	Cost of eradicating a pathogen in a farm, including killing, disposal of fish, drainage of ponds and disinfection	Farm
Supervision by competent authority	Supervision of killing and disposal of fish, drainage of ponds and disinfection of a farm	Government
Loss of stock	Economic losses due to the loss of saleable stock	Farm
Restocking	Cost of restocking the farm	Farm
Loss of trade	Loss of income due to following and movement restrictions	Farm
Catchment level		
Surveillance by competent authority	Cost of sampling and diagnostics (electrofishing wild populations) carried out by the government in affected catchments to control the presence of a pathogen and the two years of sampling to re-establish disease-free status in infected catchments (EC 2006/88)	Government
Decrease in wild fish population	Decrease of recreational fishing and tourism caused by a decrease in wild population levels	Local economy
National level		
National market reaction	Loss due to a decreased demand for farmed fish within UK	Industry
Loss of international trade	Loss of overseas market due to a decreased demand for product and eggs	Industry
Attributes affecting environmental consequence magnitude	Definition	Who bears the cost
Number of species affected	Total number of species affected by the assessed pathogen	Public
Number of threatened species affected	Number of threatened species affected by the assessed pathogen	Public
Likely level of population decline (in affected rivers)	Decline of population due to a disease incursion (taking into account the age of affected fish)	Public
Duration of population decline	Temporal scale of the decline on population levels	Public
Knock-on ecological effects	Secondary or indirect effects of a disease incursion on the ecology	Public

1

2 Table 1. Definition of economic and environmental consequences

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1

		Outbreak scenarios			
		1.No spread	2.Local spread	3.Regional spread	4.National spread
Infected farm	Table farms	1	1	5	38
	Restocking farms	0	1	4	25
Farms under suspicion	Table farms	5	11	43	91
	Restocking farms	3	7	29	61
Affected catchments		1	1	9	35
Likelihood (%)		8	7	80	5

2

3 Table 2. Outbreak scenario for *G. salaris*: (number of affected farms, catchments and
4 likelihoods)

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Attributes affecting economic magnitude		Farmed stock stochastic magnitude distributions (£)			Wild populations stochastic magnitude distributions (£)		
		min	max	most likely value	min	max	most likely value
Farm level					Catchment level		
1. Surveillance by competent authority*		1000	10 000	2000	Sampling immediately post outbreak		
					3000	60 000	30 000
					Status re-establishment		
		8000	100 000	30 000			
2. Disinfection of a farm*	Inf/table	5000	30 000	10 000	n/a		
	Inf/resto	5000	30 000	10 000			
	Susp/table	5 000	30 000	10 000			
	Susp/resto	5 000	30 000	10 000			
3. Supervision by competent authority*		1000	5000	1500	n/a		
4. Loss of stock*	Inf/table	50 000	180 000	90 000	n/a		
	Inf/resto	50 000	180 000	90 000			
	Susp/table	0	0	0			
	Susp/resto	0	0	0			
5. Restocking*	Inf/table	50 000	180 000	90 000	n/a		
	Inf/resto	50 000	250 000	140 000			
	Susp/table	0	0	0			
	Susp/resto	0	0	0			
6. Loss of trade	Inf/table	50 000	300 000	100 000	n/a		
	Inf/resto	50 000	300 000	150 000			
	Susp/table	0	0	0			
	Susp/resto	0	200 000	75 000			
7. Economic impact of a decrease in wild population levels		n/a			10 000	500 000	75 000
National level							
8. National market reaction		0	50 000	10 000	n/a		
9. Loss of international trade		0	50 000	10 000	n/a		

1 Table 3. Economic consequence assessment framework for an outbreak of *G. salaris*. Inf –
2 farms that are infected; Susp – farms under suspicion of infection (i.e. situated in the same
3 catchment as or have received live fish from infected farms); Table – farms producing fish for
4 human consumption; Resto – farms producing live fish for restocking purposes; n/a – no
5 available information (i.e. no relevance to wild or farm stock).

6 *costs are independent of pathogen

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Outbreak scenarios	Minimum	Maximum	Most likely
1.No spread	0.23	2.39	0.77
2.Local spread	0.42	4.43	1.63
3.Regional spread	1.95	21.60	7.52
4.National spread	11.00	89.40	33.40

2

3 Table 4. Overall stochastic summation of economic consequences (£ million) for each outbreak
4 scenario

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Environmental consequences of <i>G. salaris</i>	Weighting (W)	Ranking criteria Scalar:Value	Ranking scores (S)	Magnitude (W*S)
1. Number of species affected	1	1: 1 2: 2-4 3: 5-7 4: >7	1	1
2. Number of threatened species affected	3	0: 0 1: 1 2: 2 3: 3 4: >4	1	3
3. Likely level of population decline in the affected rivers (%)	4	1: <5 2: 5-40 3: 41-70 4: >70	4	16
4. Duration of population decline (years)	5	1: <1 2: 1-10 3: 11-20 4: >20	4	20
5. Knock-on ecological effects	5	1: Negligible 2: Minor 3: Moderate 4: Severe	2	10
Overall consequence score (OCS): ($\sum W*S$)				50

6

7 Table 5. Environmental consequence assessment framework for an outbreak of *G. salaris*. Linear
8 weighting (1 – least significant; 5 – most significant) was determined by Cefas experts. Experts
9 provided ranking scores which were then multiplied by the categorical weight.

10

1

Outbreak scenarios	Number of affected catchments (IC)	Outbreak scenario score (IC * OCS)
1. No spread	1	50
2. Local spread	1	50
3. Regional spread	9	450
4. National spread	35	1750

2

3 Table 6. Overall scores of environmental consequences for an outbreak of *G. salaris* scaled by
4 number of catchments affected for each outbreak scenario

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6