

## LETTER

# Metabolic traits predict the effects of warming on phytoplankton competition

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### Abstract

Understanding how changes in temperature affect interspecific competition is critical for predicting changes in ecological communities with global warming. Here, we develop a theoretical model that links interspecific differences in the temperature dependence of resource acquisition and growth to the outcome of pairwise competition in phytoplankton. We parameterised our model with these metabolic traits derived from six species of freshwater phytoplankton and tested its ability to predict the outcome of competition in all pairwise combinations of the species in a factorial experiment, manipulating temperature and nutrient availability. The model correctly predicted the outcome of competition in 72% of the pairwise experiments, with competitive advantage determined by difference in thermal sensitivity of growth rates of the two species. These results demonstrate that metabolic traits play a key role in determining how changes in temperature influence interspecific competition and lay the foundation for mechanistically predicting the effects of warming in complex, multi-species communities.

### Keywords

Climate change, freshwater phytoplankton, global change, interspecific competition, metabolic theory of ecology, nutrients, phosphate, physiological mismatches, temperature, trait-based ecology.

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## INTRODUCTION

Climate change is predicted to be a major cause of species extinctions over the next century (Field & Barros 2014), and a considerable threat to biodiversity (Bellard *et al.* 2012). Susceptibility to climate change will depend on species' environmental tolerances (Pacifi *et al.* 2015), with those occupying narrower thermal niches expected to be more vulnerable to climate warming (Magozzi & Calosi 2015). Recent studies have highlighted that changes in species interactions may also play an important role in mediating the impacts of climate change on populations (Dunn *et al.* 2009; Gilman *et al.* 2010; Bellard *et al.* 2012; Cahill *et al.* 2013; Field & Barros 2014). Indeed, the key drivers of global change (warming, CO<sub>2</sub> and changes in nutrient availability) are known to affect various types of species interactions, including competition (Tylianakis *et al.* 2008). Understanding how increases in temperature affect species interactions is therefore crucial to predicting the ecological consequences of future climate change (Dunn *et al.* 2009; Kordas *et al.* 2011; Bellard *et al.* 2012; Dell *et al.* 2014; Reuman *et al.* 2014; Bestion & Cote 2018).

Metabolism shapes numerous life-history traits that determine fitness, including population growth rate, abundance, mortality and interspecific interactions (Brown *et al.* 2004; Savage *et al.* 2004; Dell *et al.* 2011). Species vary widely in the way in which their metabolism and associated ecological rates respond to temperature (Kingsolver 2009; Dell *et al.* 2011). These interspecific differences in thermal performance curves (TPCs) can reflect differences in the magnitude (the elevation of the TPC), sensitivity (the relative rate of increase in performance with temperature), and/or thermal optima (the temperature at which the performance is maximised) (Kordas

*et al.* 2011; Dell *et al.* 2014; Pawar *et al.* 2015), and can greatly impact species interactions (Dell *et al.* 2014; Reuman *et al.* 2014). Recent theory suggests that differences in metabolic traits between consumers and resources can play a key role in determining the effects of temperature on trophic interactions (Dell *et al.* 2014; Gilbert *et al.* 2014; Pawar *et al.* 2015; Cohen *et al.* 2017). Despite advances in ecological theory linking the effects of temperature to metabolism and species interactions (O'Connor *et al.* 2011; Dell *et al.* 2014; Gilbert *et al.* 2014; Amarasekare 2015; Uszko *et al.* 2017), there have been very few empirical tests, and to our knowledge, no large-scale experimental study has confronted recent theoretical developments to assess whether differences in metabolic traits between species can predict how interspecific competition responds to warming.

In aquatic ecosystems, temperature and nutrients are the main drivers of phytoplankton productivity (Litchman *et al.* 2010). Phytoplankton exhibit substantial interspecific variation in their responses to temperature and nutrient availability (Eppley & Thomas 1969; Tilman 1981; Aksnes & Egge 1991; Boyd *et al.* 2013; Thomas *et al.* 2016, 2017). These interspecific variations in metabolic and nutrient acquisition traits are widely recognised as being important drivers of competition (Tilman 1981), community assembly (Bulgakov & Levich 1999; Grover & Chrzanowski 2006; Litchman *et al.* 2010; Edwards 2016) and ultimately the productivity of phytoplankton communities (Behrenfeld *et al.* 2005). However, we currently lack experimental tests of theory that can predict the dynamics of competition from differences in metabolic traits between species, which are essential components of models that forecast how the structure and functioning of phytoplankton communities respond to climate change (Follows *et al.* 2007).

Here, we address this fundamental knowledge gap by deriving a mathematical model to predict how changes in nutrients and temperature affect the outcome of interspecific competition from differences between species in the metabolic traits that characterise the TPCs of maximum growth rate and performance under nutrient limitation in phytoplankton. We parameterise our model with metabolic traits derived from six freshwater phytoplankton species and test the model's ability to predict the outcome of competition in all possible pairwise combinations of the six species in a factorial experiment, manipulating both temperature and nutrient availability.

## Theory

Our model predicts how interspecific differences in metabolic traits affect the competitive advantage of pairs of phytoplankton when both species are rare and colonising (co-invading) a virgin environment (or patch) (see Section S1 in supporting information for full model development). This differs from traditional resource competition (Tilman 1981) and adaptive dynamics theory (Dieckmann & Law 1996; Dieckmann 2003), in that these frameworks assume one competitor (the resident) is at population dynamics equilibrium while the other is introduced into the system at a low density. Here, we characterise scenarios where both species are rare and quantify the impact of changes in temperature and resource availability on species' relative competitive advantage. Because the two populations are initially rare, cells grow exponentially with a constant growth rate and negligible change in nutrient concentration over time until they reach an equilibrium density. Therefore, before nutrient concentration has been appreciably depleted, population growth rate of the  $i$ th species ( $i = a$  or  $b$ ) can be expressed as

$$N_i(t) = N_i(0)e^{\mu_i t}, \quad (1)$$

where  $N$  is the phytoplankton cell density (cells mL<sup>-1</sup>),  $\mu$  the realised population growth rate (d<sup>-1</sup>) and  $t$  the time (days). We model growth rate  $\mu_i$  of the  $i$ <sup>th</sup> species using the Monod eqn (Monod 1949),

$$\mu_i = \frac{\mu_{\max,i} S}{K_{S,i} + S}, \quad (2)$$

where  $\mu_{\max}$  is the maximum growth rate in nutrient-saturated conditions (d<sup>-1</sup>),  $K_S$  the half-saturation constant ( $\mu\text{mol L}^{-1}$ ) corresponds to the concentration of limiting nutrients at which the growth rate is 50% of  $\mu_{\max}$ , and measures performance at low nutrient concentrations.  $S$  is the nutrient (phosphate) concentration ( $\mu\text{mol L}^{-1}$ ). Maximum growth rate  $\mu_{\max}$  is tightly coupled to net photosynthesis rate (Geider *et al.* 1998), so its temperature dependence should follow a left-skewed unimodal function of temperature. Within the 'operational temperature range' (OTR, the temperature range typically encountered by the population, see Fig. 1)  $\mu_{\max}$  is expected to increase exponentially with temperature (Martin & Huey 2008; Angilletta 2009; Dell *et al.* 2011; Pawar *et al.* 2016). While the temperature dependence of  $K_S$  is less well known (e.g. Carter & Lathwell 1967; Ahlgren 1987; Aksnes & Egge 1991; Sterner & Grover 1998), we assume the same form of temperature dependence as  $\mu_{\max}$  (see Section S1 for a discussion of this assumption). We therefore model  $\mu_{\max}$  and  $K_S$

using the Boltzmann-Arrhenius equation (Aksnes & Egge 1991; Reuman *et al.* 2014),

$$\mu_{\max,i} = B_{0,i} \exp\left(-\frac{E_{\mu,i}}{k} \left(\frac{1}{T} - \frac{1}{T_{\text{ref}}}\right)\right) \quad (3)$$

$$K_{S,i} = K_{0,i} \exp\left(-\frac{E_{K,i}}{k} \left(\frac{1}{T} - \frac{1}{T_{\text{ref}}}\right)\right) \quad (4)$$

where  $B_{0,i}$  and  $K_{0,i}$  are the values of  $\mu_{\max,i}$  and  $K_{S,i}$  at a reference temperature  $T_{\text{ref}}$  (Kelvins) and include the scaling of  $\mu_{\max}$  and  $K_S$  with cell size (Section S1),  $E_{\mu,i}$  and  $E_{K,i}$  are the activation energies (eV) that phenomenologically quantify the relative rate of change in  $\mu_{\max}$  and  $K_S$  with temperature,  $k$  is the Boltzmann constant (eV·Kelvin<sup>-1</sup>) and  $T$  is the temperature (Kelvins). We consider the parameters of eqns (3) and (4) ( $B_{0,i}$ ,  $K_{0,i}$ ,  $E_{\mu,i}$ ,  $E_{K,i}$ ) as 'metabolic traits' that characterise how resource acquisition and growth respond to temperature.

Assuming  $N_a(0) = N_b(0)$  (starting densities are equal in experiments), we can define the competitive advantage ( $R$ ) of species  $a$  relative to species  $b$  by taking the log ratio of their abundances at time  $t$ :

$$R = \ln \frac{N_a(t)}{N_b(t)} = \mu_a - \mu_b = S \left( \frac{B_{0,a} \exp\left(-\frac{E_{\mu,a}}{k} \left(\frac{1}{T} - \frac{1}{T_{\text{ref}}}\right)\right)}{K_{0,a} \exp\left(-\frac{E_{K,a}}{k} \left(\frac{1}{T} - \frac{1}{T_{\text{ref}}}\right)\right)} + S \right. \\ \left. - \frac{B_{0,b} \exp\left(-\frac{E_{\mu,b}}{k} \left(\frac{1}{T} - \frac{1}{T_{\text{ref}}}\right)\right)}{K_{0,b} \exp\left(-\frac{E_{K,b}}{k} \left(\frac{1}{T} - \frac{1}{T_{\text{ref}}}\right)\right)} + S \right) t \quad (5)$$

(see Section S1). Thus, the value of  $R$  depends on differences in the competing species' metabolic traits, that is, on the respective parameters that define the temperature dependence of  $\mu_{\max}$  and  $K_S$  ( $B_{0,i}$ ,  $E_{\mu,i}$ ,  $K_{0,i}$ ,  $E_{K,i}$ ) between the two species. When there are no differences (the equivalent parameters are the same in both species),  $R = 0$  and both species are expected to be equally abundant at any time point  $t$ . When there are physiological mismatches,  $R \neq 0$ , the sign of  $R$  indicates which species has a competitive advantage: for  $R > 0$ , species  $a$  is expected to outnumber species  $b$  at time  $t$ , while the opposite is true for  $R < 0$ .

We can assess the relative importance of the metabolic traits characterising nutrient-limited and resource-saturated growth for predicting competitive advantage by comparing the full model for  $R$  (eqn 5) to a simplified version that assumes nutrient saturation:

$$R_{\infty} = \lim_{S \rightarrow \infty} R(S) = \left( B_{0,a} \exp\left(-\frac{E_{\mu,a}}{k} \left(\frac{1}{T} - \frac{1}{T_{\text{ref}}}\right)\right) \right. \\ \left. - B_{0,b} \exp\left(-\frac{E_{\mu,b}}{k} \left(\frac{1}{T} - \frac{1}{T_{\text{ref}}}\right)\right) \right) t. \quad (6)$$

In this case, species  $a$  will grow faster than species  $b$  if  $R_{\infty} > 0$ , and therefore if

$$\ln \frac{B_{0,a}}{B_{0,b}} > \frac{E_{\mu,a} - E_{\mu,b}}{k} \left(\frac{1}{T} - \frac{1}{T_{\text{ref}}}\right). \quad (7)$$

Here, the trade-off between normalisation constants ( $B_{0,a}$ ,  $B_{0,b}$ ) and activation energies ( $E_{\mu,a}$ ,  $E_{\mu,b}$ ) is explicit. At  $T = T_{\text{ref}}$ , the winner is determined by the ratio of the

normalisation constants (the right hand side of the inequality becomes zero). Species  $a$  will gain competitive advantage when  $B_{0,a} > B_{0,b}$ . However, as  $T$  increases or decreases from  $T_{\text{ref}}$ , the relative importance of the activation energies increases, and at sufficiently large  $|T - T_{\text{ref}}|$ , the competitive dominant is entirely determined by the differences in  $E_{\mu}$ : when  $T \gg T_{\text{ref}}$ , the species with the higher  $E_{\mu}$  has the advantage, while when  $T \ll T_{\text{ref}}$ , the species with the lower  $E_{\mu}$  will be dominant (e.g. Fig. S1A). For narrower temperature ranges, such as those discussed in this study, the competitive advantage is determined by differences in both normalisation constants and activation energies. This trade-off between the normalisation constants and the activation energies in shaping how the competitive advantage changes with warming is similar (but temperature specific) to the trade-off functions central to adaptive dynamics (Dieckmann & Law 1996; Dieckmann 2003).

The sign of  $R$  and  $R_{\infty}$  can change with temperature – a “reversal” in the competitive advantage indicates that one species can outcompete the other only within a specific temperature range (e.g. Fig. 3; Fig. S1B and Section S1). Thus, our model makes the following key predictions: (1) differences in individual species’ metabolic traits can predict competitive advantage between pairs of species at a given temperature; (2)  $R_{\infty}$  will approximate  $R$  in predictive power at higher nutrient concentrations, but  $R$  will better predict competitive advantage at lower nutrient concentrations; and (3) the competitive advantage will reverse with warming if the species with lower performance at low temperature ( $B_0$ ) has a sufficiently higher thermal sensitivity ( $E_{\mu}$ ).

## METHODS

### Study design

We used competition experiments among pairs of six phytoplankton species to test the model’s predictions (see Fig. S2A for a flow chart of the experimental design). We first determined the temperature dependence of  $\mu_{\text{max}}$  and  $K_S$  for each species independently, which we then used to parameterise the model, allowing us to generate predictions on the competitive advantage for each species pair as a function of temperature and nutrient concentration. We then competed the six species in all pairwise combinations at two temperatures and three nutrient concentrations.

### Species and culture conditions

The six phytoplankton species are the naturally co-occurring freshwater green algae, *Ankistrodesmus nannoselene*, *Chlamydomonas moewusii*, *Chlorella sorokiniana*, *Monoraphidium minutum*, *Scenedesmus obliquus* and *Raphidocelis subcapitata* (Fritschie *et al.* 2014). We chose these six species because they have similar cell sizes and can be cultured on the same media [standard COMBO culture medium without animal trace elements (Kilham *et al.* 1998)]. By choosing species with similar cell sizes, we aimed to minimise the effect of size on differences in metabolic traits (Section S1). Strains of each species were ordered in October 2015 from the CCAP (Table S2A), and grown on COMBO medium in semi-continuous culture at 15 °C on a 12 : 12 light-dark cycle with a light intensity of

90  $\mu\text{mol m}^{-2} \text{s}^{-1}$ , transferring cultures weekly to keep them in exponential phase of growth until the start of each experiment.

### Metabolic traits

In February 2016, we measured growth rates of each species across gradients in temperature and phosphate concentration. Each species was grown in a factorial experiment at 5 temperatures and 13 phosphate concentrations, with 3 replicates per combination, yielding 1170 cultures (Fig. S2A). We created 13 solutions of COMBO medium with different phosphate concentrations ranging from 0.01 to 50  $\mu\text{mol PO}_4^{3+} \text{L}^{-1}$  (Table S2B), a range relevant to phosphate concentrations commonly found in lakes (Downing *et al.* 2001). Small tissue culture flasks (Nunc) filled with 40 mL of each solution were inoculated with each species in monoculture at very low density (100 cells  $\text{mL}^{-1}$ ) ensuring that the increase in phosphate concentration due to the inoculum volume was minimal (0.01  $\mu\text{mol L}^{-1}$ ). Cells were then grown at 15, 20, 25, 30 and 35 °C, and 90  $\mu\text{mol m}^{-2} \text{s}^{-1}$  on a 12 : 12 light-dark cycle. Samples were shaken and their position rotated within the incubators daily during the month-long experiment. Every 2 days, 200  $\mu\text{L}$  was taken and 10  $\mu\text{L}$  of 1% sorbitol solution was added as a cryoprotectant. After 1 h of incubation in the dark, samples were frozen at  $-80$  °C until further analysis. Cell density was determined by flow cytometry (BD Accuri C6) on fast flux settings (66  $\mu\text{L min}^{-1}$ ), counting 10  $\mu\text{L}$  per sample. During the experiment, some samples failed to grow properly and were therefore removed from the subsequent analyses.

### Competition experiments

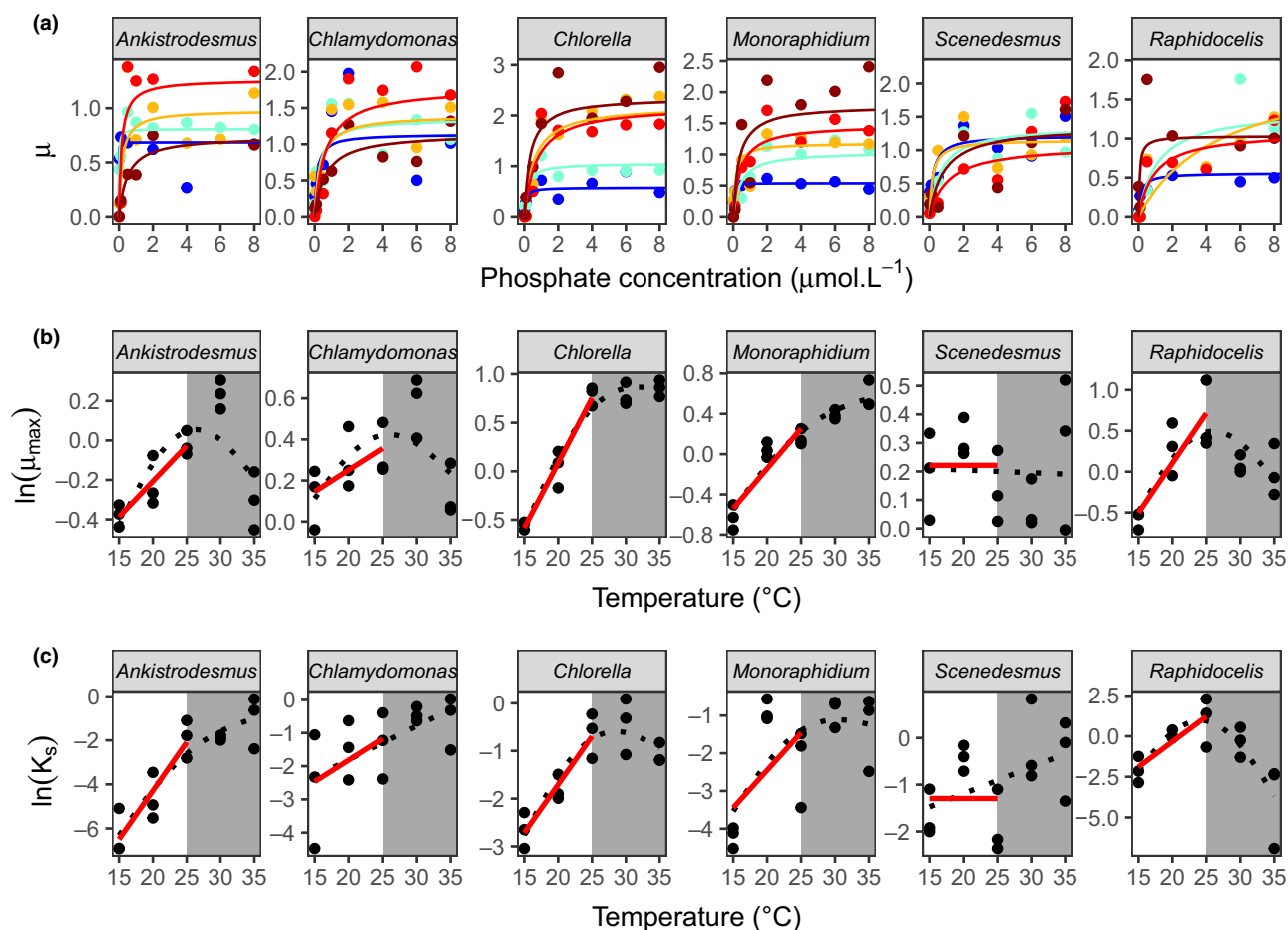
To investigate the joint effects of temperature and phosphate availability on competition, we competed species in all pairwise combinations (15 pairs) at two temperatures (15 and 25 °C; a low temperature and a temperature close to the optimum for most species, Fig. 1) and three phosphate concentrations (one saturating [30  $\mu\text{mol L}^{-1}$ ] and two limiting [1  $\mu\text{mol L}^{-1}$  and 0.1  $\mu\text{mol L}^{-1}$ ] concentrations, chosen from the Monod curves, Fig. 1), replicated 6 times (Fig. S2A), yielding 540 microcosms. We also grew the six species in monoculture at the two temperatures and three nutrient levels to train and test an algorithm for discriminating the different species in the competition trials (see Section S3 for more details). We used 24 well plates filled with 2 mL of media, inoculated them with 100 cells  $\text{mL}^{-1}$  of each species, and incubated them in the same way as described above. After 5, 14 and 23 days, a 200  $\mu\text{L}$  sample was taken and cell density was determined by flow cytometry.

### Data analyses

All statistical analyses were undertaken using R v3.3.2 (R Core Team 2014).

### Metabolic traits

To characterise the effects of phosphorous availability and temperature on growth, we estimated specific growth from the time series of cell densities. Population dynamics were fitted



**Figure 1** Interspecific variation in metabolic traits. (a) Monod curves for each species, with growth rate  $\mu$  as a function of phosphate concentration ( $\mu\text{mol L}^{-1}$ ) from 15 °C (blue) to 35 °C (dark red). Points represent the mean of the three replicates, and the Monod curve is drawn from the mean parameters across the three replicates. Note that the phosphate concentration levels in the experiment range from 0.01 to 50  $\mu\text{mol L}^{-1}$  but the  $x$ -axis was cut at 8  $\mu\text{mol L}^{-1}$  for clarity. (b) Maximum growth rate  $\mu_{\text{max}}$  and (c) the half-saturation constant  $K_S$ , as functions of temperature. Red lines represent the fit of the Boltzmann-Arrhenius within the operational temperature range (15–25 °C, white area). Black dotted lines represent the fit of the GAM over the whole temperature range. See Tables S4A–D for more details about the temperature dependence of  $\mu_{\text{max}}$  and  $K_S$ .

using nonlinear least squares regression to the Buchanan three-phase linear growth model (Buchanan *et al.* 1997):

$$N_t = \begin{cases} N_0 & \text{for } t \leq t_{\text{lag}}, \\ N_0 + \mu(t - t_{\text{lag}}) & \text{for } t_{\text{lag}} < t < t_{\text{max}}, \\ N_{\text{max}} & \text{for } t \geq t_{\text{max}}, \end{cases} \quad (8)$$

where  $t_{\text{lag}}$  is the duration of the lag phase (days),  $t_{\text{max}}$  the time when the maximum population density is reached (days),  $N_0$  the  $\log_{10}$  of the initial population density [ $\log_{10}(\text{cells mL}^{-1})$ ],  $N_{\text{max}}$  the  $\log_{10}$  of the maximum population density supported by the environment [ $\log_{10}(\text{cells mL}^{-1})$ ] and  $\mu$  the specific growth rate ( $\text{day}^{-1}$ ). Fits to the Buchanan model were determined using the ‘nlsLM’ function in the ‘minpack.lm’ package (Elzhov *et al.* 2010), which uses the Levenberg-Marquardt optimisation algorithm. Parameter estimation was achieved by running 1000 different random combinations of starting parameters picked from uniform distributions and returning the parameter set with the lowest AICc score (Padfield *et al.* 2016).

The Monod equation (eqn 2, Monod 1949) was fitted to the estimates of  $\mu$  for each species at each temperature and for each of the three replicates using the ‘nlsLM’ function as above.

We used two approaches to describe the temperature dependence of  $\mu_{\text{max}}$  and  $K_S$ : the Boltzmann-Arrhenius model and generalised additive models (GAMs). First, we fitted the Boltzmann-Arrhenius model on natural log-transformed (hereafter ‘ln’)  $\mu_{\text{max}}$  and  $K_S$  within the ‘operational temperature range’, between 15 and 25 °C, using a reference temperature  $T_{\text{ref}} = 15$  °C (eqns 3 and 4) with the ‘nlsLM’ function as above. This analysis produced normalisation constants and activation energies for both  $\mu_{\text{max}}$  and  $K_S$  per species, which we then used to parameterise eqns 5 and 6 in the theory. Second, for each species, we fitted a GAM to  $\ln \mu_{\text{max}}$  and  $\ln K_S$  across the full temperature range over which the TPCs are typically unimodal using a basis dimension of 3 and the ‘ts’ type of basis-penalty smoother with the ‘mgcv’ package.

#### Competition

The flow cytometer returned side scatter (SSC), forward scatter (FSC), green (FL1), orange (FL2), red (FL3) and blue (FL4) fluorescence values that can be used to define a species’ morphology and pigment composition. We used these



quantities to quantify cell identity and thus estimate the relative abundances of each species in pairwise competition experiments. We separated the dataset into three, one for training the discrimination algorithm, one for testing its efficiency at separating species pairs and one for the actual competition trials. The training dataset was used to establish pairwise discrimination functions between pairs of species, using three different procedures: a linear discriminant analysis, a random forest analysis and a recursive partitioning and regression tree analysis (Section S3). These different discriminant functions were then applied to the testing dataset to determine the accuracy of the various discrimination algorithms in differentiating between pairs of species by creating *in silico* competition experiments (Section S3). The linear discriminant analysis predicted the correct cell identity of each species in the *in silico* pairwise experiments with 78% accuracy and was chosen for application to the competition dataset (Fig. S3A and Table S3A). Results were robust to the statistical method used to discriminate between species (Section S6).

After determining species identity for each competition trial, we computed cell density and calculated the competitive advantage,  $R$ , of species  $a$  relative to species  $b$  by taking the  $\ln$  ratio of their densities (cells  $\text{mL}^{-1}$ ) at time  $t$ , and adding one to each of the species' densities to account for instances when one species had become locally extinct. We also computed a binary competitive advantage where species  $a$  (respectively, species  $b$ ) was competitively dominant for  $R > 0$  (respectively,  $R < 0$ ).

## RESULTS

### Metabolic traits

The responses of growth rate to phosphate concentration were well fit by the Monod equation (Fig. 1a). The half-saturation constant,  $K_S$ , and the maximum growth rate,  $\mu_{\max}$ , varied with temperature, and the temperature response of these traits differed between species (Fig. 1b and c; Tables S4A–D). Maximum growth rate exhibited unimodal temperature dependence in *Ankistrodesmus*, *Chlamydomonas* and *Raphidocelis* (Fig. 1b, Table S4C). In *Chlorella* and *Monoraphidium*,  $\mu_{\max}$  increased with temperature but did not reach a peak by 35 °C, while  $\mu_{\max}$  in *Scenedesmus* exhibited negligible temperature dependence (Fig. 1b, Table S4C).  $K_S$  increased with temperature for *Ankistrodesmus*, *Chlamydomonas* and *Monoraphidium*, while the response was unimodal for *Chlorella* and *Raphidocelis* and there was no discernible trend for *Scenedesmus* (Fig. 1c, Table S4D).

### Interspecific competition

The competitive advantage depended on temperature, nutrient conditions and the identity of the species pair (Fig. 2). For instance, for the pair *Ankistrodesmus*–*Chlorella*, *Ankistrodesmus* dominated the competition at lower temperatures while *Chlorella* dominated at higher temperatures, except at very low nutrient concentrations. For some species pairs, one species dominated across all temperatures and nutrient concentrations – e.g. *Monoraphidium* always outcompeted *Raphidocelis*.

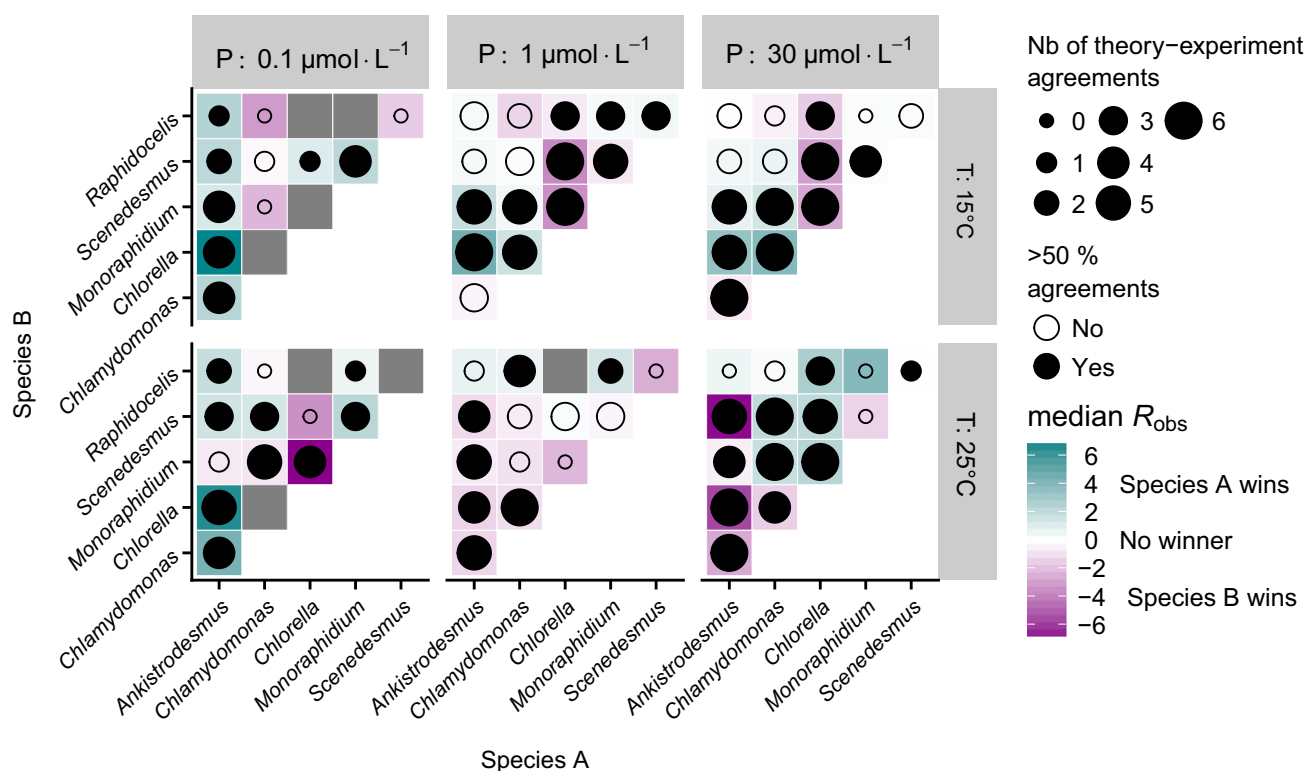
The theoretical competitive advantage  $R$  (eqn 5) correctly predicted 72% of the experimental outcomes (Table 1). The predictability of the competitive advantage did not differ between temperatures, but it varied with nutrient concentration and depended on species identity (Table 1). Eighty-seven per cent of the interactions involving *Chlorella* were correctly predicted, while those involving *Raphidocelis* were the most difficult to predict (only 52%). Indeed, removing interactions involving *Raphidocelis* increased the overall predictive power of the model to 77%. The model correctly predicted 86% of the observed reversals in competitive advantage across temperatures at the high nutrient conditions, while it was unable to predict reversals at lower nutrient levels (Table 2). Consistent with the theory, these reversals are due to the differences in metabolic traits between species leading to the crossing of growth rate TPCs between two competing species (eqn 7; Fig. 3). Assuming nutrient-saturated conditions ( $R_{\infty}$ , eqn 6) decreased the predictive power of the model (Table 1). Accounting for interspecific differences in the temperature dependence of  $K_S$  substantially improved predictions at the very low nutrient concentrations.

In addition to the binary competition outcome, we also tested the model's ability to quantitatively predict the magnitude of  $R$  and found a significant correlation between the predicted and observed  $R$  (Fig. S7A, Table S7A), which became stronger when excluding *Raphidocelis* (Tables S7B–C). This result suggests that the simple metabolic model can be used to predict how environmental changes alter the relative abundance of species as well as the binary outcome of pairwise interactions.

## DISCUSSION

Understanding how changes in temperature and nutrients affect competitive interactions among phytoplankton is critical for predicting how environmental change will shape the structure and functioning of aquatic ecosystems. We tackled this challenge by developing, parameterising and testing a model that predicts competition among phytoplankton from differences in the 'metabolic traits' that characterise the TPCs of maximum growth rate and performance under nutrient limitation. Our analyses demonstrate that the competitive advantage of six species of freshwater phytoplankton under changing temperatures and nutrients can be predicted with information on just four metabolic traits.

In our experiments, the response of growth rate to phosphorous availability was well fit by the Monod equation. The parameters characterising this functional response to resource availability were temperature dependent. Over a broad range of temperatures (15–35 °C) both the maximum growth rate ( $\mu_{\max}$ ) and the half-saturation constant ( $K_S$ ) exhibited nonlinear temperature dependence, consistent with Senft *et al.* (1981). However, within the operational temperature range (OTR), the temperature dependence of both  $\mu_{\max}$  and  $K_S$  was well fit by the exponential Boltzmann-Arrhenius equation. This result is interesting *per se* as, compared to  $\mu_{\max}$ , the temperature dependence of  $K_S$  is poorly understood (see Section S1). Our results support the positive temperature dependence expected by some theoretical studies (Goldman &



**Figure 2** Predicting competitive advantage from metabolic traits. The colour indicates the identity of the competitively dominant species and strength of competitive advantage after 14 days (median  $R_{\text{obs}}$  over 6 replicates; see Fig S3B for  $R_{\text{obs}}$  by replicate). The circles show the agreement of the model predictions with the experimental outcomes (size: number of replicates correctly predicted; colour: more than half of the replicates correctly predicted, see Table 1). If the cell density was too low to accurately predict a winner, we dropped the replicate. Thus, the number of replicates per pair, temperature and nutrient conditions is not always 6. Eight competition trials were dropped because all replicates had too low a cell density. These are shown as grey tiles. The total number of replicates is  $N = 361$ .

Carpenter 1974; Aksnes & Egge 1991; Reuman *et al.* 2014). For both  $\mu_{\text{max}}$  and  $K_S$ , the activation energies and normalisation constants (value of the trait at a reference temperature) differed among the six phytoplankton species.

We used these empirically determined metabolic traits to parameterise our model to predict the effects of changes in temperature and nutrients on the relative competitive advantage of each species in competition with each of the others and tested the outcome against a factorial experiment, manipulating temperature and nutrient availability. Our experiment revealed that species' relative competitive advantage changed substantially with temperature and nutrients. Comparing the model's predictions to the experimental results demonstrated that differences in metabolic traits were a good predictor of the relative competitive advantage of a species in pairwise competition, with the full model correctly predicting 72% of the experimental outcomes. Accounting for the effects of temperature on nutrient-limited growth kinetics ( $R$ ) was important for predicting species' competitive advantages under very low nutrient concentrations ( $0.1 \mu\text{mol PO}_4^{3+} \text{L}^{-1}$ ), but as nutrient concentration increased, knowledge of differences in the temperature dependence of  $\mu_{\text{max}}$  was sufficient to predict the effects of warming at intermediate ( $1 \mu\text{mol PO}_4^{3+} \text{L}^{-1}$ ) and high ( $30 \mu\text{mol PO}_4^{3+} \text{L}^{-1}$ ) nutrient concentrations.

For some combinations, one species was dominant across all temperatures and nutrient concentrations. In these cases,

the competitively superior species often had a higher normalisation constant for maximum growth rate (i.e.  $B_0$ ), resulting in faster realised growth rates under all conditions (Fig. 3). There were also frequent reversals of competitive advantage, particularly with changes in temperature. Temperature-driven reversals in competitive advantage were often linked to analogous reversals in the competitive advantage predicted by the model, where the superior competitor in the warm environment typically had a higher activation energy for maximum growth rate ( $E_{\mu}$ , Fig. 3). The model predicted 86% of competitive reversals at high nutrient levels. The poor predictability at low nutrient concentrations may simply reflect the fact that temperature-driven competitive reversals were generally rare under nutrient limitation ( $n = 2$ ). Indeed, the model's overall performance under nutrient-limited conditions was very good, predicting outcomes in 76% of cases. The lack of temperature-driven reversals in competitive advantage under nutrient-limited conditions suggests that normalisation constants for  $\mu_{\text{max}}$  and  $K_S$  were the main drivers of competition rather than the activation energies, perhaps because the temperature dependence of growth and resource uptake is heavily constrained at low nutrients (Thomas *et al.* 2017). Overall, these results demonstrate that metabolic traits play a central role in shaping competitive interactions among phytoplankton and highlight that particular combinations of traits consistently predict competitive advantage under warming – i.e. high  $B_0$

**Table 1** Proportion of competitive advantages correctly predicted by theory

	$R_{\infty}$		$R$		$N$
Full dataset	0.63	(0.009)	0.72	(0.000)	361
By temperature					
$T = 15\text{ }^{\circ}\text{C}$	0.66	(0.071)	0.73	(0.006)	188
$T = 25\text{ }^{\circ}\text{C}$	0.58	(0.100)	0.72	(0.003)	173
By nutrient					
$[P] = 0.1\text{ }\mu\text{mol L}^{-1}$	0.32	(0.800)	0.76	(0.061)	68
$[P] = 1\text{ }\mu\text{mol L}^{-1}$	0.64	(0.025)	0.68	(0.007)	148
$[P] = 30\text{ }\mu\text{mol L}^{-1}$	0.75	(0.004)	0.75	(0.004)	145
By species					
<i>Ankistrodesmus</i>	0.68	(0.015)	0.80	(0.000)	136
<i>Chlamydomonas</i>	0.61	(0.051)	0.70	(0.005)	138
<i>Chlorella</i>	0.78	(0.011)	0.87	(0.001)	119
<i>Monoraphidium</i>	0.60	(0.067)	0.72	(0.008)	131
<i>Scenedesmus</i>	0.58	(0.054)	0.65	(0.005)	125
<i>Raphidocelis</i>	0.42	(0.831)	0.52	(0.344)	73

Results are shown for the full dataset (including competitions at both temperatures and nutrient concentrations), by temperature, nutrient concentration and species (where only competitions involving each individual species are considered in turn). The column " $R_{\infty}$ " (eqn 6) assumes nutrient-saturated conditions, while column " $R$ " (eqn 5) explicitly captures nutrient limitation. " $N$ " indicates the number of competitions in each subset.  $P$  values indicated in parentheses were obtained by bootstrapping (see Section S5). The experimental competition data use the LDA discrimination method on the results at day 14. Analogous results for the random forest and rpart discrimination methods are shown in Tables S6A–B, and for results at day 5 and day 23 are shown in Tables S9A–B.

**Table 2** Number of observed and predicted reversals in competitive advantage between pair of species

	Observed revs.		Predicted revs. ( $R_{\infty}$ )		Predicted revs. ( $R$ )	
	Yes	No	$N$	Prop.	$N$	Prop.
Full dataset	16	23	10	0.62	9	0.56
By nutrient						
$[P]=0.1\text{ }\mu\text{mole}\cdot\text{L}^{-1}$	2	8	1	0.50	0	0.00
$[P]=1\text{ }\mu\text{mole}\cdot\text{L}^{-1}$	7	7	3	0.43	3	0.43
$[P]=30\text{ }\mu\text{mole}\cdot\text{L}^{-1}$	7	8	6	0.86	6	0.86
By species						
<i>Ankistrodesmus</i>	7	8	5	0.71	4	0.57
<i>Chlamydomonas</i>	5	9	2	0.40	2	0.40
<i>Chlorella</i>	8	3	7	0.88	7	0.88
<i>Monoraphidium</i>	5	8	4	0.80	3	0.60
<i>Scenedesmus</i>	5	9	1	0.20	1	0.20
<i>Raphidocelis</i>	2	9	1	0.50	1	0.50

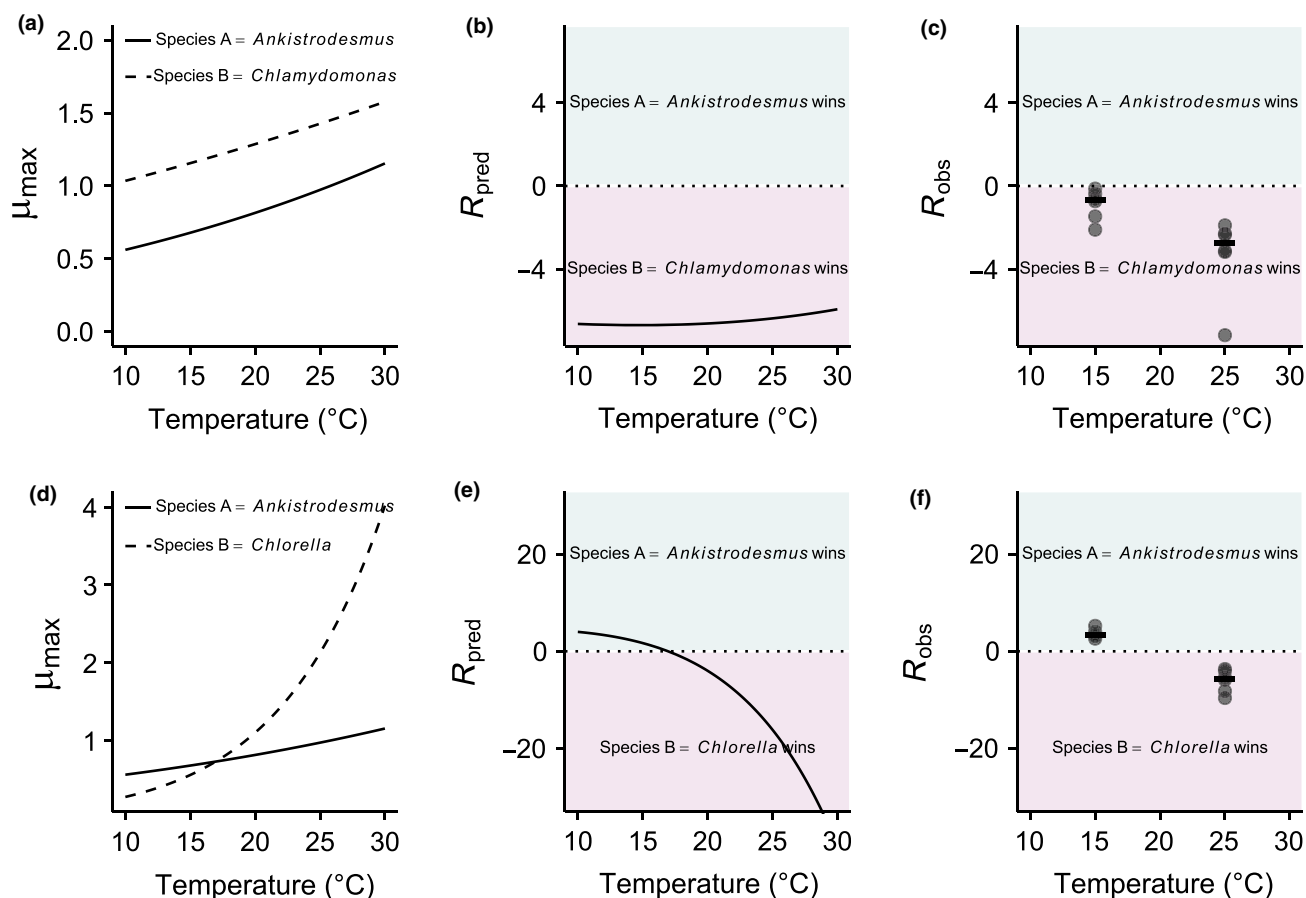
Observed reversals are qualified when the median  $R$  of a pair of species across six replicates changes sign with temperature. They are compared to reversals predicted by the model. We counted the number of times the model correctly predicted that a specific pair of species would reverse the sign of their competitive advantage.

and  $E_{ii}$ . Our findings also suggest that a greater understanding of the variation in metabolic traits at local to global scales is urgently needed if we are to predict how the structure and functioning of planktonic ecosystems will be affected by climate change (Litchman & Klausmeier 2008; Litchman *et al.* 2010).

Despite the good agreement between our model and the median experimental outcomes, the results should be

interpreted with some caution because the experimental competitive coefficients were often variable among the six replicates in each pairwise interaction (Fig. S3B). Such variability might reflect natural intrapopulation variability in traits not captured by the model, which is parameterised by the average trait values for each species. It could also be driven by experimental precision in quantifying the competitive advantage in small volume, high-throughput batch-culture experiments. Future work will be needed to verify these results in smaller-scale experiments using high-precision chemostat methods. Nevertheless, the competitive advantages were generally highly predictable, particularly when excluding interactions involving *Raphidocelis*, suggesting that the model's assumptions are nonetheless appropriate for the other five species. The poor predictability of interactions involving *Raphidocelis* warrants further attention. Our ability to discriminate and quantify this species when in competition using the linear discriminant algorithm was poor (Table S3A), and the confidence intervals around the TPCs of  $\mu_{\max}$  and  $K_S$  were also wide (Fig. 1, Tables S4A–D), which likely impaired the performance of the model. Other factors not accounted for in the model, such as direct interspecific interference (e.g. through the production of toxins), might be more important in this species' interactions. Indeed, total polyculture yields involving *Raphidocelis* were substantially lower than expectations based on the weighted average of the monoculture yields (Table S8A, Loreau & Hector 2001), indicating strong negative interactions, consistent with interspecific interference.

Our experiments and theory explored the short-term dynamics of two species colonising virgin environment when both are locally rare. The model can, however, also be extended to explore scenarios where a rare species (or genotype) invades a resident that is at population dynamics equilibrium (see Section S1), scenarios which are central to resource competition theory (Tilman 1981), modern coexistence theory (Chesson 2000) and adaptive dynamics (Dieckmann & Law 1996; Dieckmann 2003). Tilman (1981) proposed that the outcome of competition is determined by the species with the lowest  $R^*$  (in our notation,  $S^*$ ), that is, the species with the lowest equilibrium resource requirements. The  $R^*$  could, for this purpose, be derived from our model with the explicit temperature-dependent parameters we use here ( $\mu_{\max}$ ,  $K_S$ ), leading to predictions for the effects of differences in metabolic traits on invasion under a range of warming and nutrient manipulation scenarios. Tilman *et al.*'s  $R^*$  concept also extends to the adaptive dynamics framework, where the difference in  $R^*$  between a resident and a competing genotype is equivalent to the 'invasion fitness' criterion (e.g. see Section 4 in Dieckmann 2003). As with resource competition theory, for a competing genotype to successfully invade, its  $R^*$  needs to be lower than that of the resident. Differences in the temperature dependence of species' metabolism (or those of residents and mutants) would therefore be expected to lead to trade-offs in invasion fitness, comparable to those we have observed in the context of temperature-driven reversals in competitive advantage owing to species' differences in the activation energy and normalisation constant of maximum growth rate and the half-saturation constant (see Fig. 3).



**Figure 3** Predicting reversals in competitive advantage from mismatches in metabolic traits. (a–c) Competition between *Ankistrodesmus* and *Chlamydomonas*, (d–f) competition between *Ankistrodesmus* and *Chlorella*. (a and d) Represent the temperature dependence of  $\mu_{\max}$  derived from the Boltzmann-Arrhenius models. In (a),  $\mu_{\max}$  is always higher for *Chlamydomonas*, while in (d), *Ankistrodesmus* has a higher  $\mu_{\max}$  at low temperatures, but a lower  $\mu_{\max}$  at high temperatures. This translates into different shapes of predicted  $R_{\infty}$  with temperature, with a reversal of competitive advantage with temperature in the *Ankistrodesmus-Chlorella* competition (e) while there is no reversal in the *Ankistrodesmus-Chlamydomonas* competition (b). These theoretical predictions are in line with the experimental observations (c, f;  $N = 6$  replicates per temperature plus medians as segments).

A key assumption of our model is that populations are initially rare and cells grow exponentially with a constant growth rate and negligible change in nutrient concentrations over time. This assumption was violated in several of the experimental conditions at day 14, which were the data used to test the theory. The median time to equilibrium density during the single-species nutrient-gradient experiments at 15 and 25 °C was 11 and 9 days, respectively, at 0.1  $\mu\text{mol PO}_4^{3+} \text{L}^{-1}$ , 11 and 7 days at 1  $\mu\text{mol PO}_4^{3+} \text{L}^{-1}$ , and 15 and 9 days at 30  $\mu\text{mol PO}_4^{3+} \text{L}^{-1}$ . At high temperatures and low nutrient concentrations many species were no longer in the exponential phase of growth. We assessed the impact that this violation in the model's assumptions might have on the model-data comparisons by quantifying the correlation between the competitive advantages derived at day 5 (when all species were still in exponential growth under all conditions) with those used to test the model at day 14. The observed competitive advantages between species pairs were correlated between day 5 and day 14 (Pearson  $r = 0.67$  [95% CI: 0.56, 0.75]) and between day 14 and day 23 ( $r = 0.54$  [95% CI: 0.45, 0.62]). Furthermore, the performance of the model in predicting the competitive advantage was also

consistent between time points, with the model correctly predicting 66% of interactions after 5 days, 72% after 14 days and 68% after 23 days (Section S9). These results demonstrate that the competitive advantage at day 14 carries the signature of exponential growth because the initial competitive advantage results in an exponentially higher abundance of the competitively superior species (Supporting Information equations (15) and (20)). That is, the advantage persists into the phase of the two-species community assembly where the populations are no longer growing exponentially (effectively a stationary phase because of nutrient depletion). Whether this advantage persists at population equilibrium when nutrient supply is constant needs to be investigated in future work. In particular, we note that long-term equilibrium abundance, often called 'carrying capacity' or  $K$  in classical ecological theory, is thought to correlate with population growth rate, with evidence for both positive (Mallet 2012) and negative (Savage *et al.* 2004) associations.

Overall, our study shows that temperature-driven shifts in competitive advantage among phytoplankton can be predicted from basic information on the metabolic traits governing the thermal responses of growth and resource acquisition. These



results emphasise the potential for using metabolic traits to predict how directional environmental change (e.g. climatic warming) as well as environmental fluctuations influence the ecological dynamics of phytoplankton communities. Extending our theoretical and empirical work beyond pairwise interactions to complex multi-species communities will require further work in two main areas. First, the theory will need to be extended to understand how differences in metabolic traits play out in the context of indirect interactions in multi-species trophic interaction networks (Wootton 1994; Menge 1995; Montoya *et al.* 2009). Second, a more comprehensive understanding of metabolic trait variation at local and regional scales will be needed to expand the pairwise models to a trait-based meta-community framework for the effects of climate change on community dynamics.

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#### AUTHORSHIP

GYD and SP conceived the study, EB and GYD designed the experiments, EB and ES performed the experiments, BGC and SP developed the theory, EB and BGC analysed the data, EB wrote the first draft and all authors contributed to writing.

#### DATA ACCESSIBILITY STATEMENT

Data is available on Zenodo, <https://doi.org/10.5281/zenodo.1161963>

#### REFERENCES

- Ahlgren, G. (1987). Temperature functions in biology and their application to algal growth constants. *Oikos*, 49, 177–190.
- Aksnes, D.L. & Egge, J.K. (1991). A theoretical model for nutrient uptake in phytoplankton. *Mar. Ecol. Prog. Ser.*, 70, 65–72.
- Amarasekare, P. (2015). Effects of temperature on consumer–resource interactions. *J. Anim. Ecol.*, 84, 665–679.
- Angilletta, M.J. (2009). *Thermal Adaptation: A Theoretical and Empirical Synthesis*. Oxford University Press, Oxford.
- Behrenfeld, M.J., Boss, E., Siegel, D.A. & Shea, D.M. (2005). Carbon-based ocean productivity and phytoplankton physiology from space. *Global Biogeochem. Cycles*, 19, GB1006.
- Bellard, C., Bertelsmeier, C., Leadley, P., Thuiller, W. & Courchamp, F. (2012). Impacts of climate change on the future of biodiversity. *Ecol. Lett.*, 15, 365–377.
- Bestion, E. & Cote, J. (2018). Species responses to climate change: integrating individual-based ecology into community and ecosystem studies. In: *The Encyclopedia of the Anthropocene* (eds DellaSala, D.A. & Goldstein, M.I.). Elsevier, Oxford, pp. 139–147.
- Boyd, P.W., Rynearson, T.A., Armstrong, E.A., Fu, F., Hayashi, K., Hu, Z. *et al.* (2013). Marine phytoplankton temperature versus growth responses from polar to tropical waters – outcome of a scientific community-wide study. *PLoS ONE*, 8, e63091.
- Brown, J.H., Gillooly, J.F., Allen, A.P., Savage, V.M. & West, G.B. (2004). Toward a metabolic theory of ecology. *Ecology*, 85, 1771–1789.
- Buchanan, R.L., Whiting, R.C. & Damert, W.C. (1997). When is simple good enough: a comparison of the Gompertz, Baranyi, and three-phase linear models for fitting bacterial growth curves. *Food Microbiol.*, 14, 313–326.
- Bulgakov, N.G. & Levich, A.P. (1999). The nitrogen : phosphorus ratio as a factor regulating phytoplankton community structure : nutrient ratios. *Archiv für Hydrobiologie*, 146, 3–22.
- Cahill, A.E., Aiello-Lammens, M.E., Fisher-Reid, M.C., Hua, X., Karanewsky, C.J., Ryu, H.Y. *et al.* (2013). How does climate change cause extinction? *Proc. R. Soc. B*, 280, 20121890.
- Carter, O.G. & Lathwell, D.J. (1967). Effects of temperature on orthophosphate absorption by excised corn roots. *Plant Physiol.*, 42, 1407–1412.
- Chesson, P. (2000). Mechanisms of maintenance of species diversity. *Annu. Rev. Ecol. Syst.*, 31, 343–366.
- Cohen, J.M., Venesky, M.D., Sauer, E.L., Civitello, D.J., McMahon, T.A., Roznik, E.A. *et al.* (2017). The thermal mismatch hypothesis explains host susceptibility to an emerging infectious disease. *Ecol. Lett.*, 20, 184–193.
- Dell, A.I., Pawar, S. & Savage, V.M. (2011). Systematic variation in the temperature dependence of physiological and ecological traits. *PNAS*, 108, 10591–10596.
- Dell, A.I., Pawar, S. & Savage, V.M. (2014). Temperature dependence of trophic interactions are driven by asymmetry of species responses and foraging strategy. *J. Anim. Ecol.*, 83, 70–84.
- Dieckmann, U. & Law, R. (1996). The dynamical theory of coevolution: a derivation from stochastic ecological processes. *J. Math. Biology*, 34, 579–612.
- Diekmann, O. (2003). A beginners guide to adaptive dynamics. *Banach Center Publications*, 63, 47–86.
- Downing, J.A., Watson, S.B. & McCauley, E. (2001). Predicting Cyanobacteria dominance in lakes. *Can. J. Fish Aquat. Sci.*, 58, 1905–1908.
- Dunn, R.R., Harris, N.C., Colwell, R.K., Koh, L.P. & Sodhi, N.S. (2009). The sixth mass coextinction: are most endangered species parasites and mutualists? *Proc. R. Soc. B*, 276, 3037–3045.
- Edwards, K.F. (2016). Community trait structure in phytoplankton: seasonal dynamics from a method for sparse trait data. *Ecology*, 97, 3441–3451.
- Elzhov, T.V., Mullen, K.M., Spiess, A.-N. & Bolker, B. (2010). R interface to the Levenberg-Marquardt nonlinear least-squares algorithm found in MINPACK.
- Eppley, R.W. & Thomas, W.H. (1969). Comparison of half-saturation constants for growth and nitrate uptake of marine phytoplankton. *J. Phycol.*, 5, 375–379.
- Field, C.B. & Barros, V.R. & Intergovernmental Panel on Climate Change (Eds.). (2014). *Climate Change 2014: Impacts, Adaptation, and Vulnerability: Working Group II Contribution to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change*. Cambridge University Press, New York, NY.
- Follows, M.J., Dutkiewicz, S., Grant, S. & Chisholm, S.W. (2007). Emergent biogeography of microbial communities in a model ocean. *Science*, 315, 1843–1846.
- Fritschie, K.J., Cardinale, B.J., Alexandrou, M.A. & Oakley, T.H. (2014). Evolutionary history and the strength of species interactions: testing the phylogenetic limiting similarity hypothesis. *Ecology*, 95, 1407–1417.
- Geider, R.J., MacIntyre, H.L. & Kana, T.M. (1998). A dynamic regulatory model of phytoplankton acclimation to light, nutrients, and temperature. *Limnol. Oceanogr.*, 43, 679–694.
- Gilbert, B., Tunney, T.D., McCann, K.S., DeLong, J.P., Vasseur, D.A., Savage, V. *et al.* (2014). A bioenergetic framework for the temperature dependence of trophic interactions. *Ecol. Lett.*, 17, 902–914.
- Gilman, S.E., Urban, M.C., Tewksbury, J., Gilchrist, G.W. & Holt, R.D. (2010). A framework for community interactions under climate change. *Trends Ecol. Evol.*, 25, 325–331.
- Goldman, J.C. & Carpenter, E.J. (1974). A kinetic approach to the effect of temperature on algal growth. *Limnol. Oceanogr.*, 19, 756–766.

- Grover, J.P. & Chrzanowski, T.H. (2006). Seasonal dynamics of phytoplankton in two warm temperate reservoirs: association of taxonomic composition with temperature. *J. Plankton Res.*, 28, 1–17.
- Kilham, S.S., Kreeger, D.A., Lynn, S.G., Goulden, C.E. & Herrera, L. (1998). COMBO: a defined freshwater culture medium for algae and zooplankton. *Hydrobiologia*, 377, 147–159.
- Kingsolver, J.G. (2009). The Well-Tempered Biologist. *Am. Nat.*, 174, 755–768.
- Kordas, R.L., Harley, C.D.G. & O'Connor, M.I. (2011). Community ecology in a warming world: the influence of temperature on interspecific interactions in marine systems. *Journal of Experimental Marine Biology and Ecology, Global change in marine ecosystems*, 400, 218–226.
- Litchman, E. & Klausmeier, C.A. (2008). Trait-based community ecology of phytoplankton. *Annu. Rev. Ecol. Evol. Syst.*, 39, 615–639.
- Litchman, E., dePinto, P. T., Klausmeier, C.A., Thomas, M.K. & Yoshiyama, K. (2010). Linking traits to species diversity and community structure in phytoplankton. *Hydrobiologia*, 653, 15–28.
- Loreau, M. & Hector, A. (2001). Partitioning selection and complementarity in biodiversity experiments. *Nature*, 412, 72–76.
- Magozzi, S. & Calosi, P. (2015). Integrating metabolic performance, thermal tolerance, and plasticity enables for more accurate predictions on species vulnerability to acute and chronic effects of global warming. *Glob. Change Biol.*, 21, 181–194.
- Mallet, J. (2012). The struggle for existence: how the notion of carrying capacity, K, obscures the links between demography, Darwinian evolution, and speciation. *Evol. Ecol. Res.*, 14, 627–665.
- Martin, T.L. & Huey, R.B. (2008). Why “Suboptimal” is optimal: Jensen’s inequality and ectotherm thermal preferences. *Am. Nat.*, 171, E102–E118.
- Menge, B.A. (1995). Indirect effects in marine rocky intertidal interaction webs: patterns and importance. *Ecol. Monogr.*, 65, 21–74.
- Monod, J. (1949). The growth of bacterial cultures. *Annual Reviews in Microbiology*, 3, 371–394.
- Montoya, J., Woodward, G., Emmerson, M.C. & Solé, R.V. (2009). Press perturbations and indirect effects in real food webs. *Ecology*, 90, 2426–2433.
- O'Connor, M.I., Gilbert, B. & Brown, C.J. (2011). Theoretical predictions for how temperature affects the dynamics of interacting herbivores and plants. *Am. Nat.*, 178, 626–638.
- Pacifici, M., Foden, W.B., Visconti, P., Watson, J.E.M., Butchart, S.H.M., Kovacs, K.M. *et al.* (2015). Assessing species vulnerability to climate change. *Nature Clim. Change*, 5, 215–224.
- Padfield, D., Yvon-Durocher, G., Buckling, A., Jennings, S. & Yvon-Durocher, G. (2016). Rapid evolution of metabolic traits explains thermal adaptation in phytoplankton. *Ecol. Lett.*, 19, 133–142.
- Pawar, S., Dell, A.I. & Savage, V.M. (2015). Chapter 1 - from metabolic constraints on individuals to the dynamics of ecosystems. In *Aquatic Functional Biodiversity*. (ed Jacob, A.B.W.). Academic Press, San Diego, pp. 3–36.
- Pawar, S., Dell, A.I., Savage, V.M. & Knies, J.L. (2016). Real versus artificial variation in the thermal sensitivity of biological traits. *Am. Nat.*, 187, E41–E52.
- R Core Team. (2014). *R: A Language and Environment for Statistical Computing*. R Foundation for Statistical Computing, Vienna, Austria.
- Reuman, D.C., Holt, R.D. & Yvon-Durocher, G. (2014). A metabolic perspective on competition and body size reductions with warming. *J. Anim. Ecol.*, 83, 59–69.
- Savage, V.M., Gillooly, J.F., Brown, J.H., West, G.B. & Charnov, E.L. (2004). Effects of body size and temperature on population growth. *Am. Nat.*, 163, 429–441.
- Senft, W.H., Hunchberger, R.A. & Roberts, K.E. (1981). Temperature dependence of growth and phosphorus uptake in two species of volvocales, Chlorophyta. *J. Phycol.*, 17, 323–329.
- Sterner, R.W. & Grover, J.P. (1998). Algal growth in warm temperate reservoirs: kinetic examination of nitrogen, temperature, light, and other nutrients. *Water Res.*, 32, 3539–3548.
- Thomas, M.K., Kremer, C.T. & Litchman, E. (2016). Environment and evolutionary history determine the global biogeography of phytoplankton temperature traits. *Glob. Ecol. Biogeogr.*, 25, 75–86.
- Thomas, M.K., Aranguren-Gassis, M., Kremer, C.T., Gould, M.R., Anderson, K., Klausmeier, C.A. *et al.* (2017). Temperature–nutrient interactions exacerbate sensitivity to warming in phytoplankton. *Glob. Change Biol.*, 23, 3269–3280.
- Tilman, D. (1981). Tests of resource competition theory using four species of lake michigan algae. *Ecology*, 62, 802–815.
- Tylianakis, J.M., Didham, R.K., Bascompte, J. & Wardle, D.A. (2008). Global change and species interactions in terrestrial ecosystems. *Ecol. Lett.*, 11, 1351–1363.
- Uszko, W., Diehl, S., Englund, G. & Amarasekare, P. (2017). Effects of warming on predator–prey interactions – a resource-based approach and a theoretical synthesis. *Ecol. Lett.*, 20, 513–523.
- Wootton, J.T. (1994). The nature and consequences of indirect effects in ecological communities. *Annu. Rev. Ecol. Syst.*, 25, 443–466.

## SUPPORTING INFORMATION

Additional Supporting Information may be found online in the supporting information tab for this article.

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# Supporting Information for

## Metabolic traits predict the effects of warming on phytoplankton competition

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## S1. Theory

Our objective is to quantify how interspecific differences in metabolic traits affect the competitive advantage of either of a pair of competing phytoplankton species when both species are rare and colonizing (co-invading) a virgin environment. For this, we start with the well-established model of two phytoplankton populations competing for a single limiting nutrient ( $S$ ) in a chemostat-type environment (Tilman 1977, 1981):

$$\frac{1}{N_a} \frac{dN_a}{dt} = (\mu_a - D) = \frac{\mu_{\max,a} S}{K_{S,a} + S} - D \quad (11a)$$

$$\frac{1}{N_b} \frac{dN_b}{dt} = (\mu_b - D) = \frac{\mu_{\max,b} S}{K_{S,b} + S} - D \quad (11b)$$

$$\frac{dS}{dt} = D(S_0 - S) - \sum_{i=1}^2 \frac{\alpha_i \mu_{\max,i} S}{K_{S,i} + S} N_i. \quad (11c)$$

Here,  $N_i$  is the  $i$ -th species density ( $\text{cells} \cdot \text{mL}^{-1}$ ),  $\mu_i$  is its realised growth rate ( $\text{d}^{-1}$ ),  $\mu_{\max,i}$  is its maximum growth rate in nutrient saturated conditions ( $\text{d}^{-1}$ ),  $K_{S,i}$  is the half-saturation constant ( $\mu\text{mol} \cdot \text{L}^{-1}$ ) (the nutrient concentration at which realised growth is  $\mu_{\max}/2$ ; a measure of performance at low nutrient concentrations),  $S$  is the nutrient concentration ( $\mu\text{mol} \cdot \text{L}^{-1}$ ),  $D$  is dilution rate, and  $S_0$  is the inflow concentration of nutrients. The constant  $\alpha_i$  converts units of nutrient to phytoplankton cell units ( $1000 \cdot \mu\text{mol} \cdot \text{cell}^{-1}$ ); that is, it is the inverse of the number of phytoplankton cells produced per unit of resource.

The Monod equation's parameters  $\mu_{\max}$  and  $K_S$  are functional traits that depend on the species' physiology, and play an important role in shaping competitive dynamics in phytoplankton communities (Tilman 1981; Bulgakov & Levich 1999). Because the nutrients are not replenished in our colonisation experiments,  $D = 0$ , leaving

$$\frac{1}{N_a} \frac{dN_a}{dt} = \mu_a = \frac{\mu_{\max,a} S}{K_{S,a} + S} \quad (12a)$$

$$\frac{1}{N_b} \frac{dN_b}{dt} = \mu_b = \frac{\mu_{\max,b} S}{K_{S,b} + S} \quad (12b)$$

$$\frac{dS}{dt} = - \sum_{i=1}^2 \frac{\alpha_i \mu_{\max,i} S}{K_{S,i} + S} N_i. \quad (12c)$$

To calculate the competitive advantage during colonization we can assume that because the two populations are rare, cells initially grow exponentially with a constant growth rate and a negligible change in nutrient concentration over time:

$$\frac{1}{N_a} \frac{dN_a}{dt} = \mu_a = \frac{\mu_{\max,a} S}{K_{S,a} + S} \quad (13a)$$

$$\frac{1}{N_b} \frac{dN_b}{dt} = \mu_b = \frac{\mu_{\max,b} S}{K_{S,b} + S} \quad (13b)$$

$$\frac{dS}{dt} \approx 0. \quad (13c)$$

Then, to calculate competitive advantage when rare, we can solve eqns. 13:

$$N_a(t) = N_a(0)e^{\mu_a t} \quad (14a)$$

$$N_b(t) = N_b(0)e^{\mu_b t}, \quad (14b)$$

where  $t$  is time (in days). Assuming  $N_a(0) = N_b(0)$  (starting densities are equal, as in the experiments), we can define the competitive advantage ( $R$ ) of species  $a$  relative to species  $b$  by taking the log of the ratio of their abundances at time  $t$ :

$$R = \ln \frac{N_a(t)}{N_b(t)} = \mu_a - \mu_b = S \left( \frac{\mu_{\max,a}(T)}{K_{S,a}(T)+S} - \frac{\mu_{\max,b}(T)}{K_{S,b}(T)+S} \right) t. \quad (15)$$

We now incorporate the effects of temperature change on the parameters  $\mu_{\max}$  and  $K_S$  of eqn. 15 to predict the effects of warming on competitive advantage.

### ***Incorporating metabolic traits***

Maximum growth rate  $\mu_{\max}$  is tightly coupled to the rate of net photosynthesis (Geider *et al.* 1998) and consequently, its temperature dependence is expected to be exponential up to a peak value (the optimum temperature), followed by a steeper exponential decline (Angilletta 2009; Padfield *et al.* 2016; Schaum *et al.* 2017). The temperature range of the initial exponential increase up to the optimum is the ‘operational temperature range’ (OTR) — the range most likely to be encountered by the population (Martin & Huey 2008; Pawar *et al.* 2016), and it can be described by

$$\mu_{\max,i} = B'_{0,i} m_i^\beta \exp \left( -\frac{E_{\mu,i}}{k} \left( \frac{1}{T} - \frac{1}{T_{\text{ref}}} \right) \right), \quad (16)$$

where  $B'_{0,i}$  is a mass- and temperature-independent normalization constant, i.e., the value of  $\mu_{\max,i}$  at a reference temperature  $T_{\text{ref}}$  (in K),  $E_{\mu,i}$  is the activation energy (eV) that sets the relative rate of increase in  $\mu_{\max,i}$  with temperature,  $k$  is the Boltzmann constant (eV·K<sup>-1</sup>),  $T$  is temperature (K),  $m$  is cell mass (size), and  $\beta$  is the exponent of the scaling of growth rate with cell size (Eppley 1972; Kagami & Urabe 2001; Brown *et al.* 2004; DeLong *et al.* 2010). We define

$$B_{0,i} \equiv B'_{0,i} m_i^\beta \quad (17)$$

and therefore eqn. 16 becomes

$$\mu_{\max,i} = B_{0,i} \exp \left( -\frac{E_{\mu,i}}{k} \left( \frac{1}{T} - \frac{1}{T_{\text{ref}}} \right) \right). \quad (18)$$

Thus, interspecific differences in cell size  $m$  as well as the size scaling exponent  $\beta$  could contribute to differences in the species-specific normalization constants  $B_{0,i}$ , although the species used in the experiments were specifically chosen to have approximately similar cell sizes (Table S2A).

The shape of the relationship between  $K_S$  and temperature is less well known, with no comprehensive review on the subject. Several experimental studies found positive links between  $K_S$  and temperature in algae, plants and bacteria (Carter & Lathwell 1967; Shelef *et al.* 1970; Topiwala & Sinclair 1971; Thomas & Dodson 1974; Sawada *et al.* 1978; Mechling & Kilham 1982; Aksnes & Egge 1991; Sterner & Grover 1998), others found a hump-shaped relationship (Senft *et al.* 1981) or a negative relationship (Reay *et al.* 1999) while others found no evidence of temperature-dependence (Tilman *et al.* 1981; Ahlgren 1987). According to several theoretical studies,  $K_S$  is expected to increase with temperature (Goldman & Carpenter 1974; Aksnes & Egge 1991; Reuman *et al.* 2014). We assumed  $K_S$  to have a similar thermal response to  $\mu_{\max}$ , with the temperature dependence within the OTR of both  $\mu_{\max}$  and  $K_S$  following the Boltzmann-Arrhenius equation,

$$K_{S,i} = K_{0,i} \exp\left(-\frac{E_{K,i}}{k}\left(\frac{1}{T} - \frac{1}{T_{\text{ref}}}\right)\right), \quad (19)$$

where all parameters have the same meaning as in eqn. 16, and  $K_{0,i}$  has been redefined to be a mass-scaling dependent normalization constant ( $K_{0,i} \equiv K'_{0,i} m_i^\beta$ ). Our empirical results (see Figure 1 in the main text) support the use of the Boltzmann-Arrhenius function within the OTR. Comparing our empirical results to data on the same genus when available showed that our relationship was in accordance with previous experiments, with a positive temperature dependence in *Chlorella*, as found by Shelef *et al.* (1970), and no relationship with temperature in *Scenedesmus*, as found by Ahglren (1987). However, more empirical and theoretical work is needed to better understand the temperature-dependence of  $K_S$ .

### ***Effects of metabolic traits on the competitive advantage***

We can now substitute eqns. 18 and 19 into eqn. 15 to obtain the (relative) competitive advantage,  $R$ , of species  $a$  relative to species  $b$  in terms of differences in metabolic traits between the two species:

$$R = S \left( \frac{B_{0,a} \exp\left(-\frac{E_{\mu,a}}{k}\left(\frac{1}{T} - \frac{1}{T_{\text{ref}}}\right)\right)}{K_{0,a} \exp\left(-\frac{E_{K,a}}{k}\left(\frac{1}{T} - \frac{1}{T_{\text{ref}}}\right)\right) + S} - \frac{B_{0,b} \exp\left(-\frac{E_{\mu,b}}{k}\left(\frac{1}{T} - \frac{1}{T_{\text{ref}}}\right)\right)}{K_{0,b} \exp\left(-\frac{E_{K,b}}{k}\left(\frac{1}{T} - \frac{1}{T_{\text{ref}}}\right)\right) + S} \right) t. \quad (20)$$

Thus the value of  $R$  depends on the differences in the competing species' metabolism, that is, on the differences in the respective parameters that define the temperature dependence of  $\mu_{\text{max}}$  and  $K_S$  ( $B_0$ ,  $E_\mu$ ,  $K_0$ , and  $E_K$ ). When the parameters are equivalent in both species,  $R = 0$ , and both species are expected to be equally abundant at any time point  $t$ . When there are mismatches,  $R \neq 0$ , and the sign of  $R$  indicates which species has a competitive advantage: for  $R > 0$ , species  $a$  is expected to outnumber species  $b$  at time  $t$ , while the opposite is true for  $R < 0$ .

We can assess the relative importance of the metabolic traits characterising nutrient limited and resource saturated growth for predicting competitive advantage by comparing the full model for  $R$  (eq. 20) to a simplified version that assumes nutrient saturation (as  $S \rightarrow \infty$ ):

$$\begin{aligned} R_\infty &= \lim_{S \rightarrow \infty} R(S) = (\mu_{\text{max},a}(T) - \mu_{\text{max},b}(T)) t \\ &= \left( B_{0,a} \exp\left(-\frac{E_{\mu,a}}{k}\left(\frac{1}{T} - \frac{1}{T_{\text{ref}}}\right)\right) - B_{0,b} \exp\left(-\frac{E_{\mu,b}}{k}\left(\frac{1}{T} - \frac{1}{T_{\text{ref}}}\right)\right) \right) t. \end{aligned} \quad (21)$$

In this case, species  $a$  will grow faster than species  $b$  if  $R_\infty > 0$ , and therefore if

$$\ln \frac{B_{0,a}}{B_{0,b}} > \frac{E_{\mu,a} - E_{\mu,b}}{k} \left( \frac{1}{T} - \frac{1}{T_{\text{ref}}} \right). \quad (22)$$

Here, note that because the constants  $B_{0,i}$  include the effects of size (eqn. 17), part of the mismatch in normalisation constants is expected to come from differences in cell size. The trade-off between normalisation constants and activation energies here is explicit. At  $T = T_{\text{ref}}$ , the winner is entirely determined by the ratio in the normalisation constants (the right hand side of the inequality becomes zero). However, as  $T$  increases or decreases from  $T_{\text{ref}}$ , the relative importance of the activation energies increases, to the point that at a sufficiently large  $|T - T_{\text{ref}}|$ , the winner of the competition is entirely determined by the activation energy (see

Figure S1A below for an example). For narrower temperature ranges, such as those discussed in this study, the winner is determined by differences in both normalisation constants and activation energies.

A reversal in the competitive advantage  $R$  (a change in its sign) with temperature change is also possible, and can be determined numerically. For the nutrient saturated case, the temperature at which  $R_\infty = 0$  is given by

$$T_{\text{rev}} = \frac{E_{\mu,a} - E_{\mu,b}}{k \left[ \ln \frac{B_{0,a}}{B_{0,b}} + \frac{E_{\mu,a} - E_{\mu,b}}{k T_{\text{ref}}} \right]} \quad (23)$$

Here, if there is a reversal, the species that wins at the higher temperature depends only on the difference in activation energies; for example, assuming a reversal takes place, if  $E_{\mu,a} > E_{\mu,b}$ , species  $a$  is expected to outcompete species  $b$  for  $T > T_{\text{rev}}$ .

### ***Competitive advantage vs. competitive outcome***

In line with the empirical scenario of co-invasion and our experimental setup, the above theory investigates how the exponential growth phase during colonisation determines competitive advantage between species competing for a single limiting resource. However, in the long run, and once populations reach high enough population densities, density dependence and intraspecific competition might be expected to play an increasingly important role. In the experiments, we inoculated the same (small) number of cells for both species at the start of the colonisation experiment, and then use the density of each species after 14 days to test the theory. The relative abundances of each species after 14 days indicates which had a competitive advantage after colonizing an empty environment. Because the initial competitive advantage is expected to result in an exponentially higher abundance of the competitively superior species (SI eq. 15 and 20), we expect that the advantage at 14 days will persist at the end of the experiment even if the species are no longer growing exponentially. This positive association between population growth rate and the long-term competitive outcome is consistent with theory and data, which suggest that equilibrium densities reflect the balance between density independent growth and density dependent regulation and thus higher intrinsic rates of increase tend to lead to higher equilibrium densities and competitive advantage (Mallet 2012). For reference, the median times to equilibrium density in the growth rate experiments were 11 and 9 days for very low ( $0.1 \mu\text{mol}\cdot\text{L}^{-1}$  of phosphate) nutrient concentrations respectively for 15 and 25°C, 10.5 and 7 at low ( $1 \mu\text{mol}\cdot\text{L}^{-1}$  of phosphate) nutrient concentrations, and 14.5 and 9 days at high ( $30 \mu\text{mol}\cdot\text{L}^{-1}$  of phosphate) nutrient concentrations. Furthermore, we were also able to compare our theoretical predictions to results after 5 and 23 days of experiment, allowing us to check whether the assumption of the “carry-over signature” of competitive advantage beyond the exponential phase held true (Section S9).

### ***Extensions to adaptive dynamics, and Tilman’s $R^*$ theory***

The full model (eqns. (11)) can be used to study more scenarios, including invasion while rare (where one species is introduced while the other is at its equilibrium density), and to explore longer-term adaptive (competitive) dynamics. In the invasion-while-rare scenario, Tilman *et al.* (1981) show that the species with the lowest equilibrium requirements of nutrients ( $S^*$ ) will win, independent of the starting densities. In the model of eqn. (11), for instance,



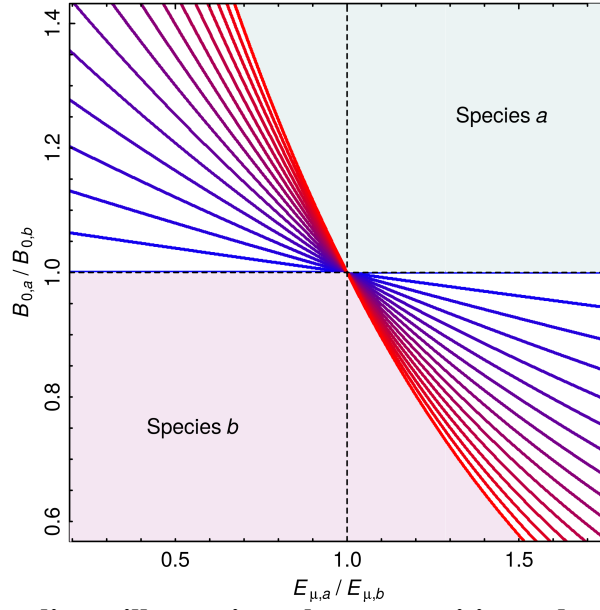
$$S_i^* = \frac{D K_{S,i}}{\mu_{\max,i} - D} \quad (24)$$

so if a (resident) species is assumed to be at its equilibrium density,  $N^*$ , and a new (invading) species is introduced while rare, as long as  $S_{\text{inv}}^* < S_{\text{res}}^*$ , the invasion will be successful. The same argument is made within the adaptive dynamics framework, where  $S^*$  effectively represents the invasion fitness. If a mutant with a lower  $S^*$  is introduced to a population, it will successfully invade. With the temperature dependence of  $\mu_{\max}$  and  $K_S$  made explicit using metabolic theory as we have done above, the conditions for invasion can be made explicitly temperature dependent and expressed in terms of mismatches between the resident and the invader population.

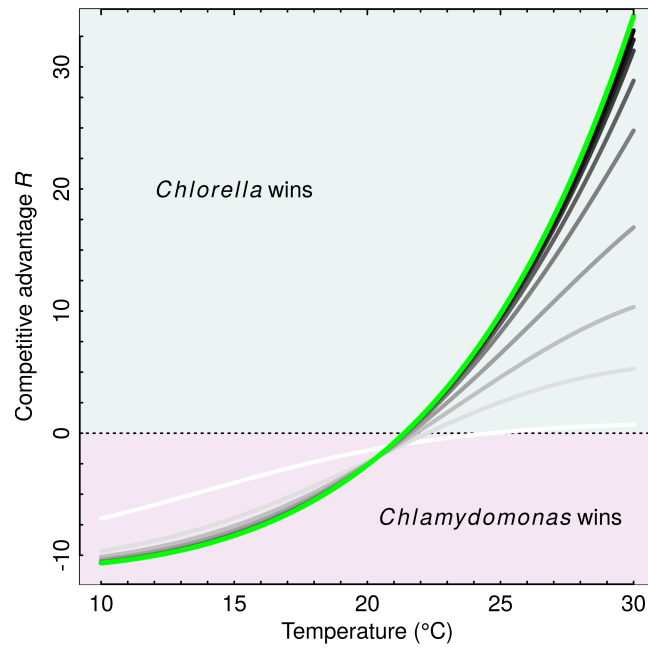
## References

- Ahlgren, G. (1987). Temperature Functions in Biology and Their Application to Algal Growth Constants. *Oikos*, 49, 177–190.
- Aksnes, D.L. & Egge, J.K. (1991). A theoretical model for nutrient uptake in phytoplankton. *Marine Ecology Progress Series*, 70, 65–72.
- Angilletta, M.J. (2009). *Thermal Adaptation: A Theoretical and Empirical Synthesis*. Oxford University Press.
- Brown, J.H., Gillooly, J.F., Allen, A.P., Savage, V.M. & West, G.B. (2004). Toward a metabolic theory of ecology. *Ecology*, 85, 1771–1789.
- Bulgakov, N.G. & Levich, A.P. (1999). The nitrogen : Phosphorus ratio as a factor regulating phytoplankton community structure : Nutrient ratios. *Archiv für Hydrobiologie*, 146, 3–22.
- Carter, O.G. & Lathwell, D.J. (1967). Effects of Temperature on Orthophosphate Absorption by Excised Corn Roots. *Plant Physiol.*, 42, 1407–1412.
- DeLong, J.P., Okie, J.G., Moses, M.E., Sibly, R.M. & Brown, J.H. (2010). Shifts in metabolic scaling, production, and efficiency across major evolutionary transitions of life. *PNAS*, 107, 12941–12945.
- Eppley, R.W. (1972). Temperature and phytoplankton growth in the sea. *Fishery Bulletin*, 70, 1063–1085.
- Geider, R.J., MacIntyre, H.L. & Kana, T.M. (1998). A dynamic regulatory model of phytoplankton acclimation to light, nutrients, and temperature. *Limnology and Oceanography*, 43, 679–694.
- Goldman, J.C. & Carpenter, E.J. (1974). A kinetic approach to the effect of temperature on algal growth. *Limnology and Oceanography*, 19, 756–766.
- Kagami, M. & Urabe, J. (2001). Phytoplankton growth rate as a function of cell size: an experimental test in Lake Biwa. *Limnology*, 2, 111–117.
- Mallet, J. (2012). The struggle for existence: how the notion of carrying capacity, K, obscures the links between demography, Darwinian evolution, and speciation. *Evol Ecol Res*, 14, 627–665.
- Martin, T.L. & Huey, R.B. (2008). Why “Suboptimal” Is Optimal: Jensen’s Inequality and Ectotherm Thermal Preferences. *The American Naturalist*, 171, E102–E118.
- Mechling, J.A. & Kilham, S.S. (1982). Temperature Effects on Silicon Limited Growth of the Lake Michigan Diatom *Stephanodiscus Minutus* (bacillariophyceae)1. *Journal of Phycology*, 18, 199–205.
- Padfield, D., Yvon-Durocher, G., Buckling, A., Jennings, S. & Yvon-Durocher, G. (2016). Rapid evolution of metabolic traits explains thermal adaptation in phytoplankton. *Ecol Lett*, 19, 133–142.
- Pawar, S., Dell, A.I., Savage, V.M. & Knies, J.L. (2016). Real versus Artificial Variation in the Thermal Sensitivity of Biological Traits. *The American Naturalist*, 187, E41–E52.

- Reay, D.S., Nedwell, D.B., Priddle, J. & Ellis-Evans, J.C. (1999). Temperature Dependence of Inorganic Nitrogen Uptake: Reduced Affinity for Nitrate at Suboptimal Temperatures in Both Algae and Bacteria. *Appl. Environ. Microbiol.*, 65, 2577–2584.
- Reuman, D.C., Holt, R.D. & Yvon-Durocher, G. (2014). A metabolic perspective on competition and body size reductions with warming. *J Anim Ecol*, 83, 59–69.
- Sawada, T., Chohji, T. & Kuno, S. (1978). Kinetic Analysis of Unbalanced Bacterial Growth in Temperature Shift. In: *Chemical Reaction Engineering—Houston* (eds. Weekman, V.W. & Luss, D.). AMERICAN CHEMICAL SOCIETY, WASHINGTON, D. C., pp. 163–172.
- Schaum, C.-E., Barton, S., Bestion, E., Buckling, A., Garcia-Carreras, B., Lopez, P., *et al.* (2017). Adaptation of phytoplankton to a decade of experimental warming linked to increased photosynthesis. *Nature Ecology & Evolution*, 1, 0094.
- Senft, W.H., Hunchberger, R.A. & Roberts, K.E. (1981). Temperature Dependence of Growth and Phosphorus Uptake in Two Species of Volvox (volvocales, Chlorophyta). *Journal of Phycology*, 17, 323–329.
- Shelef, G., Oswald, J. & Golueke, C. (1970). Assaying algal growth with respect to nitrate concentration by a continuous flow turbidostat. *Proc. Int. Conf. Water. Pollut. Res.*, 3–25.
- Sterner, R.W. & Grover, J.P. (1998). Algal growth in warm temperate reservoirs: kinetic examination of nitrogen, temperature, light, and other nutrients. *Water Research*, 32, 3539–3548.
- Thomas, W.H. & Dodson, A.N. (1974). Effect of interactions between temperature and nitrate supply on the cell-division rates of two marine phytoflagellates. *Mar. Biol.*, 24, 213–217.
- Tilman, D. (1977). Resource Competition between Plankton Algae: An Experimental and Theoretical Approach. *Ecology*, 58, 338–348.
- Tilman, D. (1981). Tests of Resource Competition Theory Using Four Species of Lake Michigan Algae. *Ecology*, 62, 802–815.
- Tilman, D., Mattson, M. & Langer, S. (1981). Competition and nutrient kinetics along a temperature gradient: An experimental test of a mechanistic approach to niche theory. *Limnol. Oceanogr.*, 26, 1020–1033.
- Topiwala, H. & Sinclair, C.G. (1971). Temperature relationship in continuous culture. *Biotechnol. Bioeng.*, 13, 795–813.



**Figure S1A. Contour lines illustrating the competitive advantage for a range of parameter combinations, assuming nutrient saturation ( $R_\infty$ ).** The colour of the lines correspond to different temperatures, ranging from 15°C for the blue line, to 30°C for the red line. For example, for  $E_{\mu,a}/E_{\mu,b} = 1$  and  $B_{0,a}/B_{0,b} = 0.8$ , species  $b$  grows faster than species  $a$ , but for  $E_{\mu,a}/E_{\mu,b} = 0.5$  and  $B_{0,a}/B_{0,b} = 1.2$ , which species grows faster depends on the temperature. Here,  $B_{0,b} = 1$ ,  $E_{\mu,b} = 0.55$ , and  $T_{\text{ref}} = 15^\circ\text{C}$ . Therefore, at  $T = 15^\circ\text{C}$ , which species wins is determined by  $B_{0,a}/B_{0,b}$  (the blue line is horizontal and insensitive to the ratio in activation energies), while as temperatures move further away from  $T_{\text{ref}}$ , the ratio of activation energies becomes increasingly important in determining the competitive advantage. As temperature increases beyond the range shown here, the lines become increasingly vertical, and as a result, insensitive to the ratio of normalization constants.



**Figure S1B. Example of a reversal in the competitive advantage,  $R$ , across a temperature range.** The green line is for nutrient saturated conditions ( $R_{\infty}$ ), and the grayscale lines are for different nutrient concentrations, ranging from  $S = 0.1 \mu\text{mol}\cdot\text{L}^{-1}$  for the light gray line, to  $50 \mu\text{mol}\cdot\text{L}^{-1}$  for the black line. The example uses parameters for *Chlorella* and *Chlamydomonas*, where  $R > 0$  means *Chlorella* has a competitive advantage over *Chlamydomonas*.

## S2. Experimental design

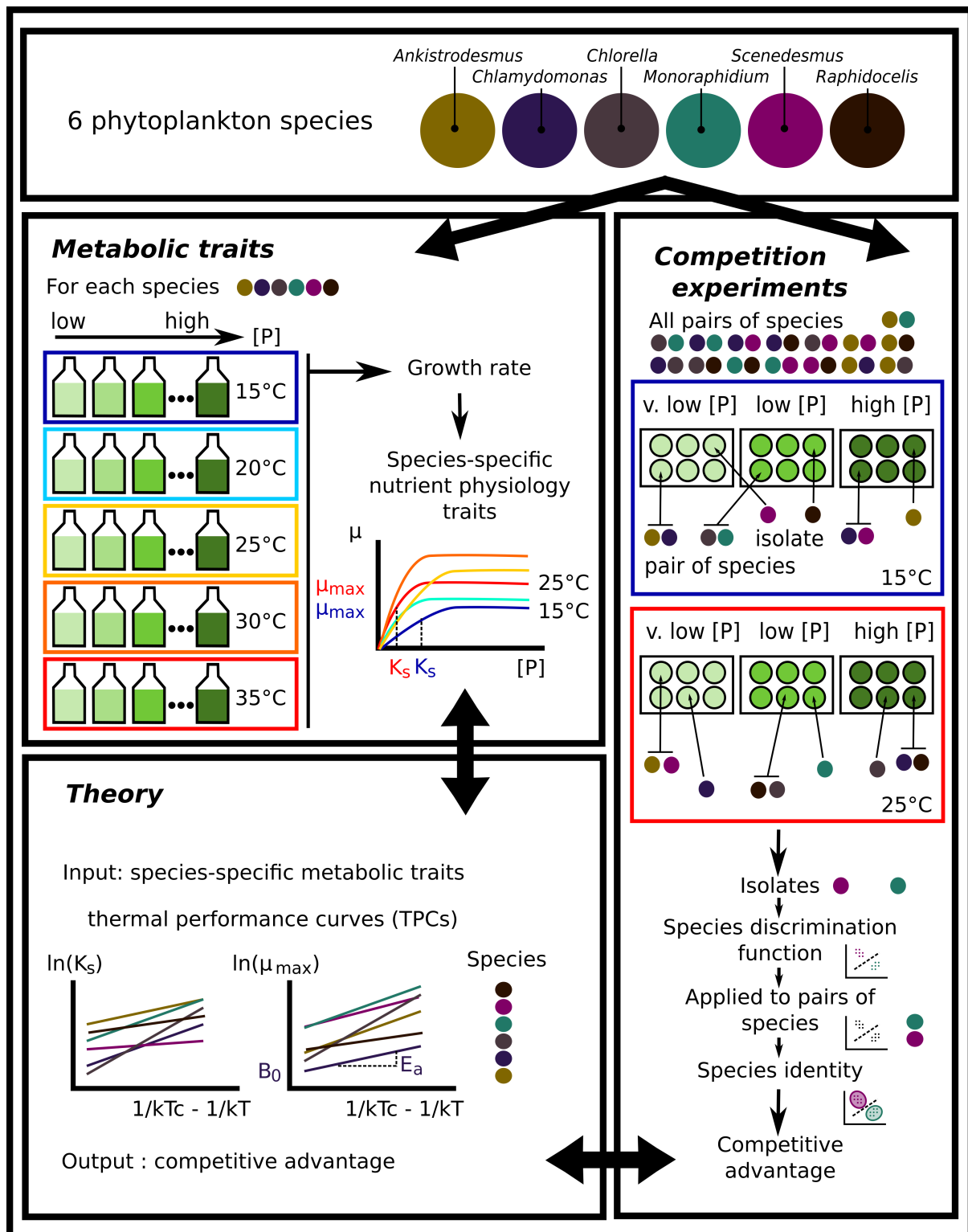


Figure S2A. Flow chart of the experimental design

**Table S2A. Detailed information about the six species.**

The species were ordered from the Culture Collection of Algae and Protozoa ([www.ccap.ac.uk](http://www.ccap.ac.uk)). Cell diameters are calculated from microscopy pictures as the average of the longest and shortest diameter of the cell over 30 cells.

Species name	Class	Order	Strain	Origin	Mean cell diameter ( $\mu\text{m}$ )
<i>Ankistrodesmus nannoselene</i> Skuja (1948)	Chlorophyceae	Sphaeropleales	CCAP 202/6A	Siggeforsajon, Sweden (1948)	2.8
<i>Chlamydomonas moewusii</i> Gerlof (1940)	Chlorophyceae	Chlamydomonadales	CCAP 11/5A	Freshwater	8.1
<i>Chlorella sorokiniana</i> Shihira& Krauss (1965)	Trebouxiophyceae	Chlorellales	CCAP 211/8K	Austin, Texas, USA (1953)	4.2
<i>Monoraphidium minutum</i> (Nägeli) Komarkova-Legnerova (1969)	Chlorophyceae	Sphaeropleales	CCAP 278/3	Texas, USA	4.7
<i>Scenedesmus obliquus</i> (Turpin) Kützing (1833)	Chlorophyceae	Sphaeropleales	CCAP 276/3B	Lund, Sweden (1939)	7.1
<i>Raphidocelis subcapitata</i> (formerly <i>Selenastrum capricornutum</i> ) Printz (1913)	Chlorophyceae	Sphaeropleales	CCAP 278/4	Akershus, Norway (1959)	5.8

**Table S2B. Phosphate concentration levels for each solution, in  $\mu\text{mol}\cdot\text{L}^{-1}$  and  $\mu\text{g}\cdot\text{L}^{-1}$ .**

We created 13 solutions of different phosphate concentrations ranging from  $0.01 \mu\text{mol}\cdot\text{L}^{-1}$  of phosphate to  $50 \mu\text{mol}\cdot\text{L}^{-1}$  of phosphate by mixing different amounts of COMBO medium without potassium phosphate dibasic (P- COMBO) and normal COMBO medium (P+ COMBO) in 40 mL tissue culture flasks. We used a modified version of the standard COMBO medium without animal trace solution in which we increased the fraction of carbonate by adding 10 mL of a stock solution of  $55.8 \text{ g}\cdot\text{L}^{-1}$  of sodium bicarbonate to maintain a DIC of more than  $6.6 \text{ mmol}\cdot\text{L}^{-1}$  in order to prevent carbon limitation, which maintained a C:N:P ratio of 132:20:1 in the P+ COMBO solution, above the Redfield ratio of 106:16:1.

Phosphate concentration ( $\mu\text{mol}\cdot\text{L}^{-1}$ )	50	40	30	20	10	8	6	4	2	1	0.5	0.1	0.01
Phosphate concentration ( $\mu\text{g}\cdot\text{L}^{-1}$ )	4750	3800	2850	1900	950	760	570	380	190	95	47.5	9.5	0.95
Amount of P+ COMBO (mL)	40	32	24	16	8	6.4	4.8	3.2	1.6	0.8	0.4	0.08	0.008
Amount of P- COMBO (mL)	0	8	16	24	32	33.6	35.2	36.8	38.4	39.2	39.6	40	40



### S3. Discrimination between species in the competition experiment

To investigate the joint effects of temperature and phosphate availability on competition, we competed all species in all pairwise combinations (15 pairs) at two temperatures (15 and 25°C; low temperature and a temperature close to the optimum for most species, Fig. 1) and three phosphate concentrations: one saturating [ $30 \mu\text{mol}\cdot\text{L}^{-1}$ ] and two limiting [ $1 \mu\text{mol}\cdot\text{L}^{-1}$  and  $0.1 \mu\text{mol}\cdot\text{L}^{-1}$ ] concentrations, chosen from the Monod curves, Fig. 1), with each replicated 6 times (Fig. S2A), amounting to 540 samples. Along with the pairwise competition trials, we grew all 6 species in monoculture at the two temperatures and three nutrient levels. This was to train the discrimination algorithm used to separate cells from different species in the competition trial. The monoculture trials were divided into two subsets, one to train the cell discrimination algorithm, which was replicated 3 times per temperature and nutrient levels, and a testing subset used to test the accuracy of the cell discrimination algorithm, which was replicated 6 times. This testing subset was also used to calculate total yield in monoculture to compare it to yield in biculture (see Section S8). The competition experiments were carried out in two batches, a first batch in June 2016 for the  $30 \mu\text{mol}\cdot\text{L}^{-1}$  and  $1 \mu\text{mol}\cdot\text{L}^{-1}$  P and a second batch in October 2017 for the very low nutrient concentration ( $0.1 \mu\text{mol}\cdot\text{L}^{-1}$  P). This second batch was added to further explore nutrient limited competition as the Monod curves indicated that  $1 \mu\text{mol}\cdot\text{L}^{-1}$  P was above the half-saturating constant for some species, particularly at low temperatures (see Table S4B). The competition experiments were carried out in 24 well plates filled with 2 mL of media and inoculated with  $100 \text{ cells}\cdot\text{mL}^{-1}$  of each species, ensuring that the increase in phosphate concentration due to the inoculum volume ( $1 \mu\text{L}$  of sample at  $2 \times 10^5 \text{ cells}\cdot\text{mL}^{-1}$ ) or due to potential storages of phosphate in the cells was minimal ( $0.025 \mu\text{mol}\cdot\text{L}^{-1}$  P). Plates were covered with AeraSeal™ breathable membrane, minimising evaporation and contamination but allowing gas exchange. The competition plates were incubated in the same way as described for the monoculture growth curves. At day 5, 14 and 23, a  $200 \mu\text{L}$  sample was taken and preserved as described in the metabolic traits section. Cell density was determined by flow cytometry on the slow flux setting ( $14 \mu\text{L}\cdot\text{min}$ ), counting  $20 \mu\text{L}$  per sample. A preliminary test measuring twice the same sample on 54 samples (6 species x 9 replicates) gave a mean variation between cell counts of 9%. We focus on the results from day 14 in the main results and for the description of the discrimination algorithm method; however, rerunning the analyses using day 5 or day 23 gave qualitatively similar results (see Supplementary Section S9 for results on these two other days).

FSC files returned by the flow cytometer were read with the Bioconductor ‘FlowCore’ package in R, returning side scatter (SSC), forward scatter (FSC), green fluorescence (FL1), orange fluorescence (FL2), red fluorescence (FL3), and blue fluorescence (FL4) values that could be used to define species morphology and pigment composition and thus discriminate between species in the pairwise competition assays. We first filtered the data to remove noise by removing every data point where either  $\ln(\text{FSC.H}) < 10.3$ ,  $\ln(\text{SSC.H}) < 3$  or  $\ln(\text{FL3.H}) < 1.5$ , which are below minimum values observed for life cells of all 6 species. The training dataset was used to determine discrimination functions between pairs of species. We used the data collected at day 14 to train the discrimination algorithm, except for the  $\text{P} = 0.1 \mu\text{mol}\cdot\text{L}^{-1}$  dataset where we pooled all of the data together to get a greater discrimination power as cell densities were very low under these conditions. We first removed outliers from this dataset by manually inspecting FSC.H by FL3.H clustering plots and choosing visual thresholds for these two values for each species. We then applied 3 different procedures to discriminate between pairs of species for each temperature and phosphate level: a linear discriminant analysis with the ‘lda’ function from the ‘MASS’ package, a random forest analysis with the ‘randomForest’ function from the ‘randomForest’ package, and a recursive partitioning and regression tree analysis with the

‘rpart’ function from the ‘rpart’ package. These analyses were performed using the natural logarithm of the 10 variables returned by the flow cytometer (that is FSC.H, FSC.A, SSC.H, SSC.A, FL1.H, FL1.A, FL2.H, FL2.A, FL3.H, FL3.A, FL4.H and FL4.A, .H standing for height and .A for area), on each of the 15 pairs of species for each combination of temperature and phosphate level, except for the  $P = 0.1 \mu\text{mol}\cdot\text{L}^{-1}$  dataset, where we pooled all temperatures together to get a greater discrimination power. These different discriminant functions were then applied to the testing dataset to test the accuracy of the predictions for the different discriminant methods. For each pair of species, we used the training set to create *in silico* competition experiments where 100% of the cells would pertain to one of the species. We applied the discrimination algorithm and calculated the percentage of times where a cell was wrongly attributed to the other species. We then chose the method that gave the maximum level of accuracy to apply to the competition dataset (Fig. S3A). The best method was the linear discriminant analysis, which gave 78% accuracy (Table S3A). However, we checked that the results were robust to the statistical method used to discriminate between species (Section S6 in SI).

After determining species identity for each sample, we computed cell density and calculated the competitive advantage  $R$  of species  $a$  relative to species  $b$  by taking the  $\ln$  ratio of their densities ( $\text{cells}\cdot\text{mL}^{-1}$ ) at time  $t$ , adding 1 to each species density for instances when one species became locally extinct (i.e., when density = 0). We also computed a binary competitive advantage where species  $a$  was competitively dominant when  $R > 0$  and vice versa. Because the efficacy of the discrimination algorithm depends on having a sufficient quantity of data with which to assign identities, we set a minimum threshold of  $N_{\text{tot}} = 500 \text{ cells}\cdot\text{mL}^{-1}$ . This led us to discard 171 replicates out of 540 for day 14. Furthermore, in comparisons with the model, we removed 9 replicates for which the observed  $R = 0$ , because the model necessarily predicts a non-zero  $R$  (traits characterising the TPCs for  $\mu_{\text{max}}$  and  $K_S$  were never identical for any species pair).

**Table S3A. Performance of the discrimination algorithms at day 14.**

LDA: linear discriminant analysis, Random Forest analysis, RPART: recursive partitioning and regression tree. Summaries by (a) species for all nutrient and thermal conditions, (b) pairs of species for all nutrient and thermal conditions, (c) phosphate and nutrient conditions for all pairs of species.

a

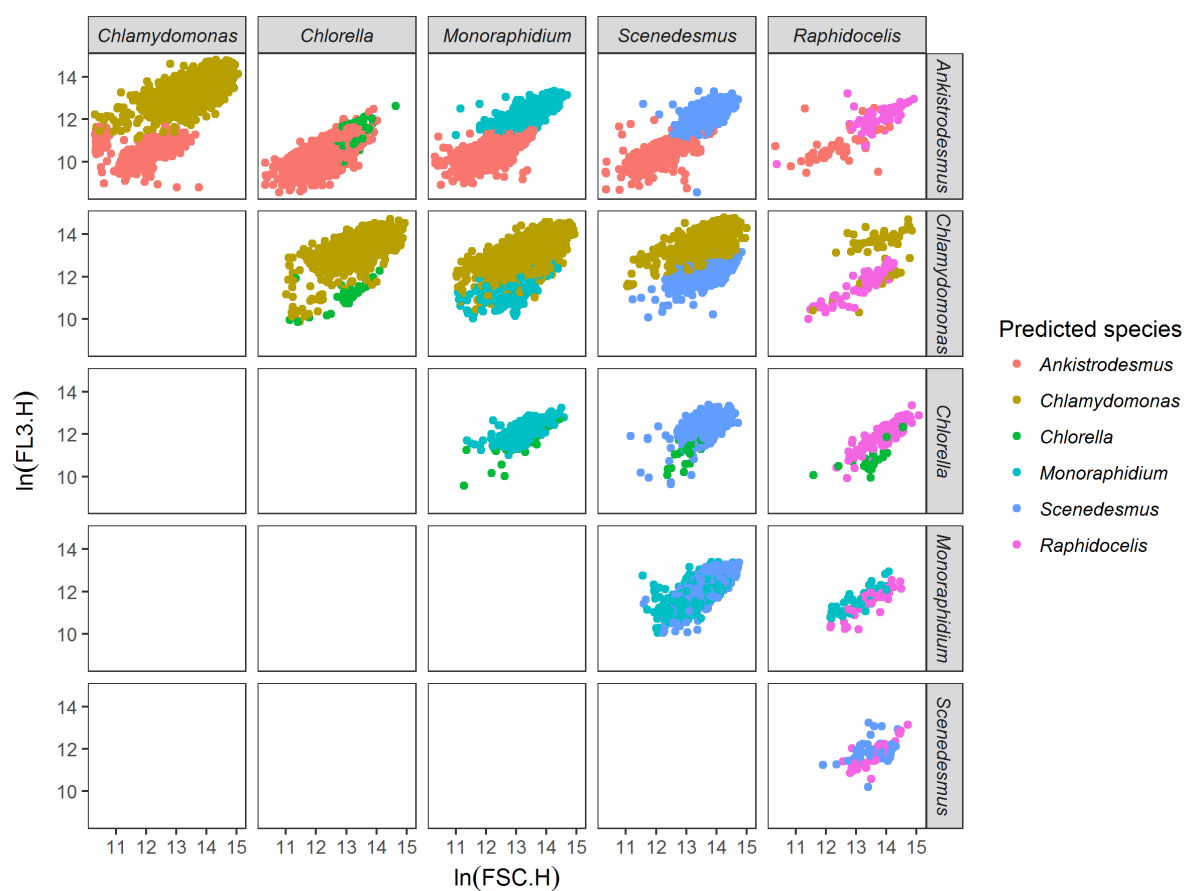
<b>Species</b>	<b>LDA</b>	<b>Randomforest</b>	<b>RPART</b>
<i>Ankistrodesmus</i>	0.84	0.78	0.67
<i>Chlamydomonas</i>	0.89	0.90	0.81
<i>Chlorella</i>	0.74	0.77	0.62
<i>Monoraphidium</i>	0.76	0.74	0.64
<i>Scenedesmus</i>	0.79	0.76	0.63
<i>Raphidocelis</i>	0.65	0.66	0.50
<b>Mean</b>	<b>0.78</b>	<b>0.77</b>	<b>0.65</b>

b

<b>Pair of species</b>	<b>LDA</b>	<b>Randomforest</b>	<b>RPART</b>
<i>Ankistrodesmus-Chlamydomonas</i>	0.96	0.98	0.93
<i>Ankistrodesmus-Chlorella</i>	0.91	0.65	0.52
<i>Ankistrodesmus-Monoraphidium</i>	0.84	0.76	0.72
<i>Ankistrodesmus-Scenedesmus</i>	0.90	0.89	0.73
<i>Ankistrodesmus-Raphidocelis</i>	0.67	0.61	0.46
<i>Chlamydomonas-Chlorella</i>	0.91	0.94	0.83
<i>Chlamydomonas-Monoraphidium</i>	0.93	0.94	0.86
<i>Chlamydomonas-Scenedesmus</i>	0.82	0.85	0.71
<i>Chlamydomonas-Raphidocelis</i>	0.80	0.81	0.76
<i>Chlorella-Monoraphidium</i>	0.62	0.72	0.59
<i>Chlorella-Scenedesmus</i>	0.80	0.81	0.66
<i>Chlorella-Raphidocelis</i>	0.57	0.72	0.49
<i>Monoraphidium-Scenedesmus</i>	0.82	0.69	0.63
<i>Monoraphidium-Raphidocelis</i>	0.57	0.62	0.37
<i>Scenedesmus-Raphidocelis</i>	0.60	0.54	0.44
<b>Mean</b>	<b>0.78</b>	<b>0.77</b>	<b>0.65</b>

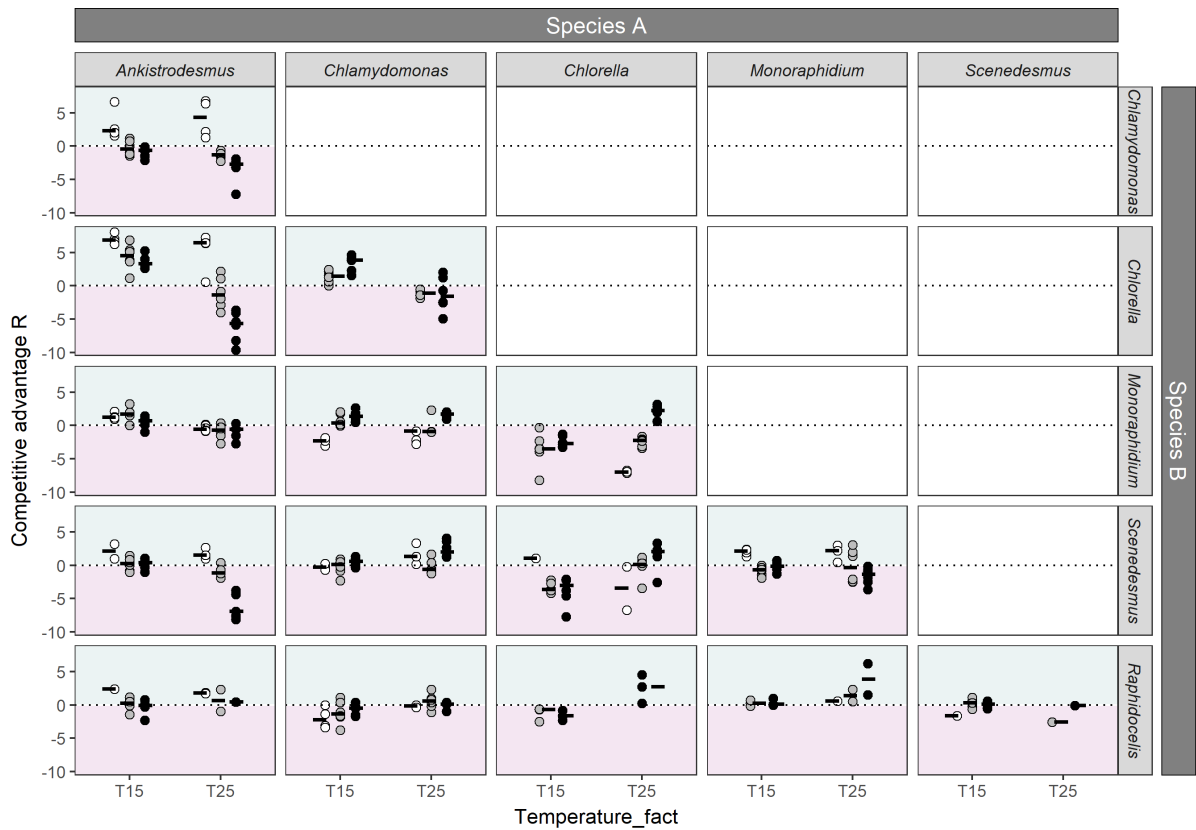
c

Temperature	Nutrient	LDA	Random forest	RPART
15	0.1	0.62	0.62	0.58
15	1	0.79	0.68	0.64
15	30	0.85	0.8	0.76
25	0.1	0.63	0.71	0.62
25	1	0.7	0.69	0.68
25	30	0.64	0.66	0.62
<b>Mean</b>		<b>0.71</b>	<b>0.69</b>	<b>0.65</b>



**Figure S3A. Example of discrimination between species among pairs of species.**

We here show species grown at 15°C in saturating nutrient conditions ( $P = 30 \mu\text{mol}\cdot\text{L}^{-1}$ ) after 14 days of experiment. Each dot represents a cell, here mapped on FSC.H (size proxy) and FL3.H (chlorophyll a proxy) characteristics from the flow cytometer. Colours represent the species predicted by the discrimination algorithm. The discrimination algorithm is a linear discriminant analysis trained with flow cytometer data (FSC.H, FSC.A, SSC.H, SSC.A, FL1.H, FL1.A, FL2.H, FL2.A, FL3.H, FL3.A, FL4.H, and FL4.A) from the species grown in isolates at the same temperature and nutrient conditions. For example, *Chlamydomonas* has a competitive advantage over *Chlorella* in these nutrient and temperature conditions (there are more *Chlamydomonas* cells).



**Figure S3B. Competition outcomes at day 14.**

For each pair of species, the competitive advantage  $R$ . Circle colour represents the nutrient conditions of the trial; white circles: very low nutrient concentration ( $0.1 \mu\text{mol}\cdot\text{L}^{-1}$  of phosphate); grey circles: low nutrient concentration ( $1 \mu\text{mol}\cdot\text{L}^{-1}$  of phosphate); black circles: saturated nutrient solution ( $30 \mu\text{mol}\cdot\text{L}^{-1}$  of phosphate). Points represent the values of each of the 6 replicates per condition. Note that when the total cell density did not reach a threshold value of  $500 \text{ cells}\cdot\text{mL}^{-1}$ , the replicates were discarded (see Methods), thus for some of the very low nutrient concentration cases no replicates were kept for a given pair. The segment represents the median of the replicates. The dotted lines represent the situation where there is no competitive advantage between the species ( $N_A = N_B$ ). The area above the line shows an advantage for species A (turquoise colour), while area below the line shows an advantage for species B (pink colour). We can see for instance that for the *Ankistrodesmus-Chlorella* pair of species, *Ankistrodesmus* dominates at low temperatures for all nutrient conditions while *Chlorella* dominates at high temperatures, particularly at high nutrient conditions.

#### S4. Temperature dependence of the Monod model parameters

**Table S4A. Metabolic traits for each alga.**

Normalization constants ( $B_0$  and  $K_0$  resp. for  $\mu_{\max}$  and  $K_S$ ) and activation energies ( $E_\mu$  and  $E_K$  resp. for  $\mu_{\max}$  and  $K_S$ ) derived from a Boltzmann-Arrhenius model fit on ln scales using nonlinear least squares to the values of  $\mu_{\max}$  and  $K_S$  for all replicates, for temperatures between 15 and 25°C, and using a reference temperature  $T_{\text{ref}} = 15^\circ\text{C}$  (estimates  $\pm$  SE). Note that for some replicates, the Monod model gave  $K_S = 0$ . Because the Boltzmann-Arrhenius model was fit on ln scales and to avoid infinite values when applying the logarithm to these values, these were set to the minimum quantity of nutrients in the experiment, that is  $K_S = 0.001$ .

Species	$K_S$		$\mu_{\max}$	
	$\ln B_0$	$E_K$	$\ln B_0$	$E_\mu$
<i>Ankistrodesmus</i>	$-6.49 \pm 0.51$	$3.26 \pm 0.59$	$-0.39 \pm 0.04$	$0.27 \pm 0.05$
<i>Chlamydomonas</i>	$-2.47 \pm 0.63$	$0.96 \pm 0.72$	$0.15 \pm 0.07$	$0.16 \pm 0.08$
<i>Chlorella</i>	$-2.71 \pm 0.19$	$1.49 \pm 0.22$	$-0.58 \pm 0.07$	$0.99 \pm 0.08$
<i>Monoraphidium</i>	$-3.44 \pm 0.73$	$1.47 \pm 0.83$	$-0.54 \pm 0.09$	$0.59 \pm 0.10$
<i>Scenedesmus</i>	$-1.30 \pm 0.46$	$0.00 \pm 0.52$	$0.22 \pm 0.07$	$0.00 \pm 0.08$
<i>Raphidocelis</i>	$-1.89 \pm 0.52$	$2.30 \pm 0.60$	$-0.50 \pm 0.17$	$0.90 \pm 0.19$

**Table S4B: Half-saturation constants ( $K_S$ ) and degree of nutrient saturation.**

Percentage of  $\mu_{\max}$  at the low and very low experimental nutrient concentrations for the competition experiment calculated from values in Table S4A. For each species, this indicates whether species are close to nutrient saturation at the experimental temperature and phosphate concentration chosen for the competition experiment.

Species	$K_S$		Growth at 1 $\mu\text{mol}\cdot\text{L}^{-1}$ as % of $\mu_{\max}$		Growth at 0.1 $\mu\text{mol}\cdot\text{L}^{-1}$ as % of $\mu_{\max}$	
	15°C	25°C	15°C	25°C	15°C	25°C
<i>Ankistrodesmus</i>	0.002	0.124	100%	89%	99%	44%
<i>Chlamydomonas</i>	0.085	0.309	91%	76%	54%	24%
<i>Chlorella</i>	0.067	0.500	94%	67%	60%	16%
<i>Monoraphidium</i>	0.032	0.233	97%	81%	76%	30%
<i>Scenedesmus</i>	0.274	0.274	79%	79%	27%	27%
<i>Raphidocelis</i>	0.151	3.386	87%	23%	40%	3%



**Table S4C. Results from the GAMs of  $\ln(\mu_{\max})$  as a function of temperature**  
 For each species. See Fig. 1 for the representation of the GAMs.

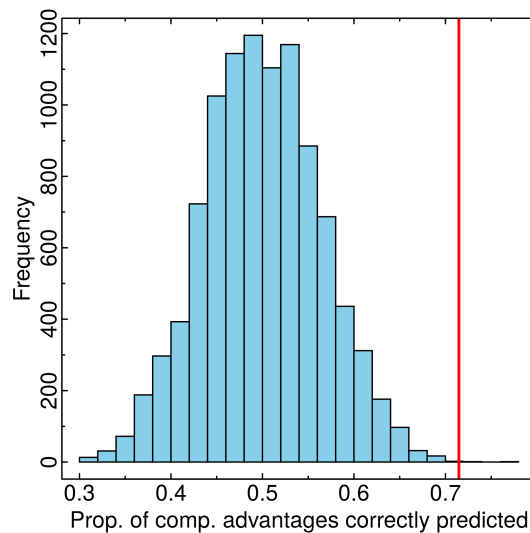
Species	edf	F	p-value	R <sup>2</sup>
<i>Ankistrodesmus</i>	2	8.33	0.005**	0.51
<i>Chlamydomonas</i>	2	3.96	0.048*	0.30
<i>Chlorella</i>	2	113.6	>0.001***	0.94
<i>Monoraphidium</i>	2	70.4	>0.001***	0.91
<i>Scenedesmus</i>	2	0.34	0.716	-0.10
<i>Raphidocelis</i>	2	9.60	0.003**	0.56

**Table S4D. Results from the GAMs of  $\ln(K_s)$  as a function of temperature**  
 For each species. See Fig. 1 for the representation of the GAMs.

Species	edf	F	p-value	R <sup>2</sup>
<i>Ankistrodesmus</i>	2	31.6	>0.001***	0.81
<i>Chlamydomonas</i>	2	4.39	0.037*	0.33
<i>Chlorella</i>	2	27.5	>0.001***	0.79
<i>Monoraphidium</i>	2	6.21	0.014*	0.43
<i>Scenedesmus</i>	2	1.49	0.265	0.06
<i>Raphidocelis</i>	2	12.28	0.001**	0.62

### S5. Significance of competitive advantage predicted by the model.

To quantify the significance of the theory's ability to predict competitive advantage, we ran the analysis 10,000 times, sampling the values of  $B_0$ ,  $E_\mu$ ,  $K_0$ , and  $E_K$  independently, with replacement, from the pool of available values. The analysis produced 10,000 sets of predictions, and therefore 10,000 proportions of competitive advantages correctly predicted (e.g., Fig. S5A). The proportion of runs that correctly predicted a greater number of competitive advantages than the real parameter values are then given as the  $P$  values in Table 1. Therefore,  $P=0.05$  means that 500 out of 10,000 random parameter combinations correctly predicted a greater proportion of competitive advantages.



**Figure S5A. Histogram of proportions of competitive advantages correctly predicted for 10,000 random parameter combinations.**

The real parameters correctly predicted the competitive advantage in 72% of the competitions (red line), and 2 of the 10,000 random parameter combinations produced a greater predictive power ( $>72\%$  of correct predictions; runs to the right of the red line).

### S6. Robustness of the results to different statistical methods

We used three different methods of discrimination to determine the number of cells from each species, a linear discriminant analysis, a random forest analysis and a recursive partitioning and regression tree (rpart, see Section S3 in SI). Because the linear discriminant analysis was found to have the best predictive power overall (Table S3A), we used this method throughout the manuscript. However, we tested whether our results were robust to the method of species discrimination by comparing results from the competition model to predictions using the random forest analysis and the rpart discrimination method (Table S6A and S6B). The results were similar, with a lower predictive power of each variable and of the model due to the lower discrimination power of the two methods, but no significant discrepancies between species and temperature and nutrient conditions.

#### Table S6A. Proportion of competitive advantages correctly predicted by theory using the random forest discrimination method at day 14.

Analogous to Table 1 in the main text, but using the random forest discrimination method.

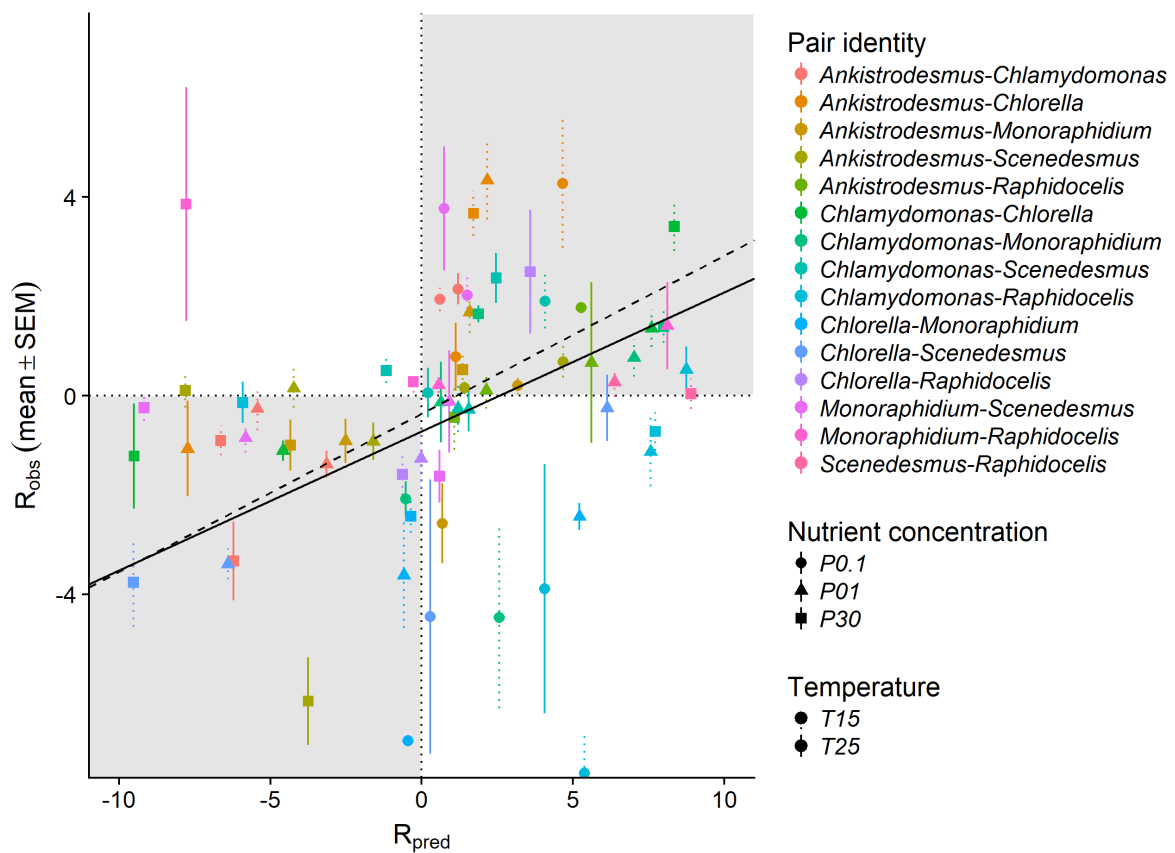
	$R_\infty$		$R$		$N$
<i>Full dataset</i>					
	0.60	(0.014)	0.70	(0.000)	365
<i>By temperature</i>					
$T = 15^\circ\text{C}$	0.66	(0.054)	0.72	(0.003)	192
$T = 25^\circ\text{C}$	0.54	(0.138)	0.68	(0.005)	173
<i>By nutrient</i>					
$[\text{P}] = 0.1 \mu\text{mol}\cdot\text{L}^{-1}$	0.33	(0.786)	0.78	(0.051)	69
$[\text{P}] = 1 \mu\text{mol}\cdot\text{L}^{-1}$	0.59	(0.136)	0.62	(0.055)	151
$[\text{P}] = 30 \mu\text{mol}\cdot\text{L}^{-1}$	0.74	(0.005)	0.74	(0.005)	145
<i>By species</i>					
<i>Ankistrodesmus</i>	0.64	(0.015)	0.80	(0.000)	137
<i>Chlamydomonas</i>	0.59	(0.012)	0.66	(0.014)	140
<i>Chlorella</i>	0.75	(0.026)	0.83	(0.003)	119
<i>Monoraphidium</i>	0.55	(0.151)	0.69	(0.010)	134
<i>Scenedesmus</i>	0.57	(0.079)	0.62	(0.031)	126
<i>Raphidocelis</i>	0.46	(0.752)	0.54	(0.239)	74

**Table S6B. Proportion of competitive advantages correctly predicted by theory using the rpart discrimination method at day 14.**

Analogous to Table 1 in the main text, but using the rpart discrimination method.

	$R_\infty$		$R$		$N$
<i>Full dataset</i>					
	0.60	(0.026)	0.68	(0.001)	367
<i>By temperature</i>					
$T = 15^\circ\text{C}$	0.64	(0.083)	0.71	(0.012)	193
$T = 25^\circ\text{C}$	0.55	(0.188)	0.66	(0.017)	174
<i>By nutrient</i>					
$[\text{P}] = 0.1 \mu\text{mol}\cdot\text{L}^{-1}$	0.37	(0.741)	0.73	(0.092)	71
$[\text{P}] = 1 \mu\text{mol}\cdot\text{L}^{-1}$	0.58	(0.176)	0.62	(0.073)	150
$[\text{P}] = 30 \mu\text{mol}\cdot\text{L}^{-1}$	0.73	(0.014)	0.73	(0.014)	146
<i>By species</i>					
<i>Ankistrodesmus</i>	0.65	(0.022)	0.77	(0.000)	137
<i>Chlamydomonas</i>	0.56	(0.081)	0.63	(0.047)	140
<i>Chlorella</i>	0.76	(0.020)	0.84	(0.003)	119
<i>Monoraphidium</i>	0.56	(0.131)	0.67	(0.023)	135
<i>Scenedesmus</i>	0.56	(0.124)	0.60	(0.055)	126
<i>Raphidocelis</i>	0.45	(0.598)	0.55	(0.287)	77

## S7. Quantitative relationship between theoretical and experimental outcomes



**Figure S7A. Correlation between the observed and predicted competitive advantage at day 14.**

Different species pairs are in different colours, circles are for very low nutrient concentration, triangle for low nutrient concentration and squares for high nutrient concentrations, and the type of the standard error line stands for the temperature (dotted for low temperature, solid for high temperature). Most of the binary experimental outcomes (sign of observed  $R$ ) fall in the same region (grey rectangles) as the binary theoretical outcomes (sign of predicted  $R$ ). The full line represents the results of a linear mixed model of observed  $R$  as a function of predicted  $R$  as a fixed effect plus pair ID, temperature and nutrients as random intercepts on the whole dataset, while the dashed line represents the results from the same model but excluding pairs involving *Raphidocelis* (see Table S7A and Table S7B for details about the model).

**Table S7A. Results from the linear mixed model investigating observed  $R$  as a function of predicted  $R$  at day 14.**

Model includes predicted  $R$  as a fixed effect plus pair ID, temperature and nutrients as random intercepts with lmer function from lme4 package ( $R_{obs} \sim R_{pred} + (1|temperature) + (1|nutrient) + (1|species\ pair)$ ).  $N = 369$ .

<b>Factor</b>	<b>Estimate</b>	<b>SE</b>	<b>t-value</b>	<b><math>\chi^2</math> statistics</b>	<b><math>R^2</math></b>
<i>Fixed effect</i>					<i>marginal <math>R^2</math></i>
$R_{pred}$	0.29	0.02	13.04	$\chi^2 = 150, p < 2e^{-16}$	0.29
<i>Random effect</i>	<b>Variance</b>				<i>conditional <math>R^2</math></i>
Temperature	0.08				0.54
Nutrient	0.10				
Pair identity	2.04				
Residual	4.21				

**Table S7B. Results from the linear mixed model investigating observed  $R$  as a function of predicted  $R$  at day 14 excluding pairs involving *Raphidocelis*.**

Model includes predicted  $R$  as a fixed effect plus pair ID, temperature and nutrients as random intercepts with lmer function from lme4 package ( $R_{obs} \sim R_{pred} + (1|temperature) + (1|nutrient) + (1|species\ pair)$ ).  $N = 292$ .

<b>Factor</b>	<b>Estimate</b>	<b>SE</b>	<b>t-value</b>	<b><math>\chi^2</math> statistics</b>	<b><math>R^2</math></b>
<i>Fixed effect</i>					<i>marginal <math>R^2</math></i>
$R_{pred}$	0.34	0.02	14.33	$\chi^2 = 205, p < 2e^{-16}$	0.34
<i>Random effect</i>	<b>Variance</b>				<i>conditional <math>R^2</math></i>
Temperature	0.43				0.62
Nutrient	0.08				
Pair identity	2.65				
Residual	4.09				

**Table S7C. Link between observed and predicted  $R$  at day 14 by species.**

Results from a mixed effect model of  $R_{obs} \sim R_{pred} + (1|temperature) + (1|nutrient) + (1|species\ pair)$  for each subset of competitions.

<b>Species</b>	<b>Fixed <math>R_{pred}</math> effect estimate</b>	<b>SD</b>	<b>t-value</b>	<b>Marginal <math>R^2</math></b>	<b>Conditional <math>R^2</math></b>	<b>N</b>
<i>Ankistrodesmus</i>	0.35	0.05	6.94	0.27	0.63	138
<i>Chlamydomonas</i>	0.26	0.04	6.91	0.29	0.53	141
<i>Chlorella</i>	0.38	0.03	12.8	0.41	0.73	120
<i>Monoraphidium</i>	0.24	0.04	6.15	0.16	0.53	136
<i>Scenedesmus</i>	0.21	0.04	5.34	0.21	0.40	126
<i>Raphidocelis</i>	0.02	0.04	0.52	0.01	0.43	77



## S8. Nature of species interactions

We used our experimental data to investigate the nature of the interactions between our species pairs. In its strictest definition, interspecific competition involves any mechanism whereby the fitness (e.g. per-capita rate of increase or population density) of a given species is reduced by the presence of another, for instance because the other species uses more resources. We calculated a relative density (RD) index for each species in each pairwise interaction, according to Fritschie *et al.* (2014). Relative density of each species,  $i$  was the ratio of species  $i$ 's population density in its biculture:monoculture ratio. Ratios below 1 indicate competitive interactions because the density in bi-culture is less than the species was able to achieve when in monoculture – i.e. it incurs a fitness cost due to interspecific competition. Conversely, ratios over 1 indicate facilitation because the focal species achieves a greater density when in the presence of another taxon than it was able to reach when alone. We found that 78.8% of pairs fell into the mutual competition scenario (where  $RD_i < 1$ ), while 17.5 % of pairs fell into an intermediate scenario where one species facilitated while the other species did not, and 3.7 % of pairs fell into a full facilitation scenario (Fig. S8A). Thus, interactions in our experiment are mainly competitive *sensu stricto*. Note that RD is a property of individual species, and any two species grown in biculture may have very different values of RD due to asymmetry in interaction strength.

We also computed another metric of interaction strength, the deviation from expected total yield  $\Delta Y$ , which is a property of the community. To do so, we computed the total cell density of the two species grown in competition and the total cell density of each species grown in isolation. We calculated a deviation from expected yield  $\Delta Y$  according to Loreau & Hector (2001), as

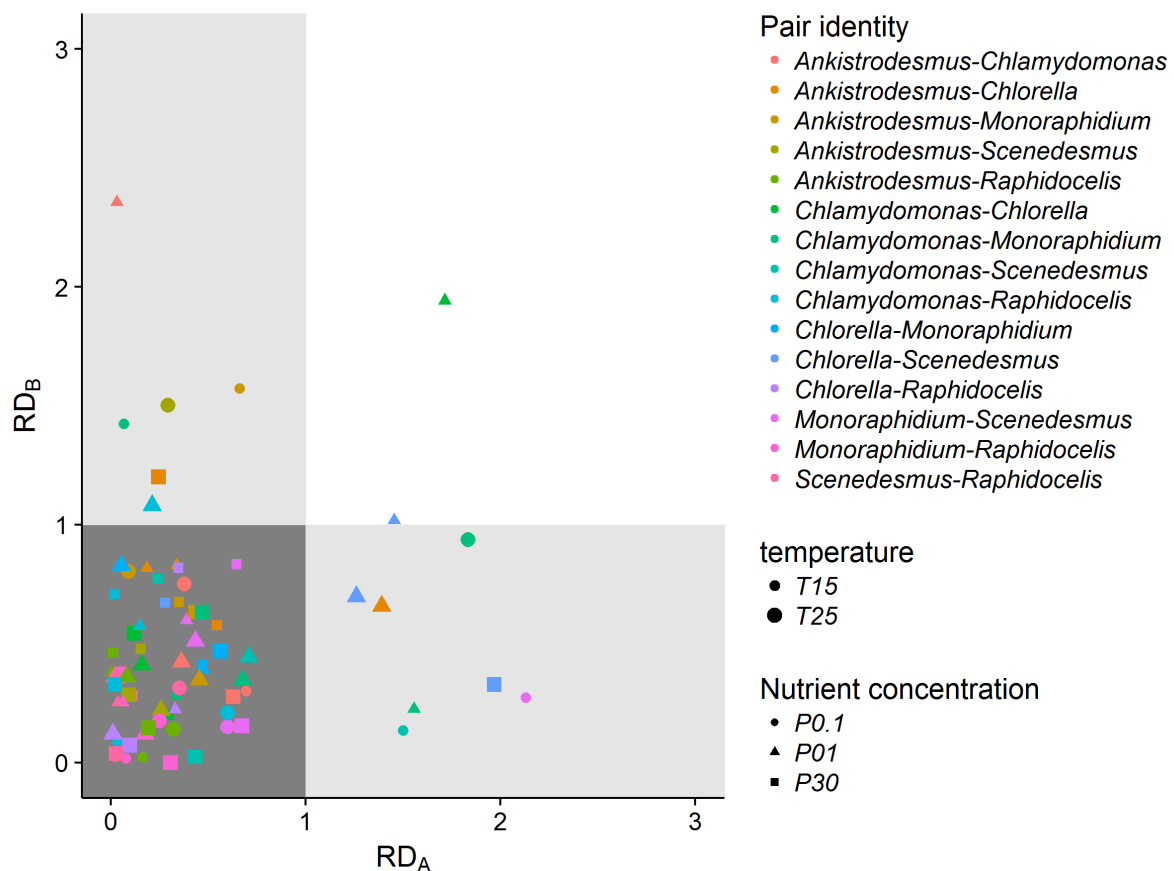
$$\Delta Y = Y_o - Y_e = \sum_i RY_{oi} - \sum_i RY_{ei},$$

where  $Y_o$  is the observed yield of the two-species mixture at day 14 (in cells·mL<sup>-1</sup>),  $Y_e$  is the expected yield of the two-species mixture, and  $RY_{oi}$  and  $RY_{ei}$  are the observed and expected relative yields of species  $i$  in the mixture. The expected relative yield of species  $i$  in the mixture are equal to half of the yield observed in monoculture (as they theoretically have access to half of the nutrients in a two-species mixture). We studied whether the deviation from expected yield varied with species identity (Table S8A). Positive deviations indicate complementarity effects (e.g., niche partitioning or facilitation) while negative deviations indicate competitive interactions diminishing total biomass. In line with the RD calculations the vast majority interactions were negative, indicating strong resource competition characterised the interactions among these 6 species of algae. Interactions involving *Raphidocelis* were strongly negative, while interactions involving *Scenedesmus* were less negative and there was no distinguishable negative interaction for *Chlamydomonas* (Table S8A).

### References

Fritschie, K.J., Cardinale, B.J., Alexandrou, M.A. & Oakley, T.H. (2014). Evolutionary history and the strength of species interactions: testing the phylogenetic limiting similarity hypothesis. *Ecology*, 95, 1407–1417.

Loreau, M. & Hector, A. (2001). Partitioning selection and complementarity in biodiversity experiments. *Nature*, 412, 72–76



**Figure S8A. Distribution of algal communities across an interaction gradient.**

Joint distribution of species relative densities (each data point is the mean across replicates of a single biculture at a specific temperature and nutrient level). Relative density (RD) is the mean density of the focal species in competition divided by its mean density when cultivated in isolation. The background colour indicates a gradient of competition strength, dark grey indicates that both species experienced stronger interspecific versus intraspecific competition ( $RD_i < 1$ ,  $N = 63$ ), while the white background indicates that both species were facilitated ( $RD_i > 1$ ,  $N = 3$ ). A small subset of interactions fell in interaction scenarios (light grey) where one species was facilitated while the other experienced interspecific competition ( $RD_i < 1$ ,  $RD_j > 1$ ,  $N = 14$ ). Note that interactions involving *Chlorella* at the very low ( $P = 0.1$ ) nutrient concentration were removed as the isolates for this species fell below the threshold of 500 cells  $\text{mL}^{-1}$  (see Methods).

**Table S8A. Deviation from the expected yield per species at day 14.**Values from two tailed t-test of  $\log_{10}(\Delta Y)$ .

<b>Species</b>	<b>mean</b>	<b>Confidence interval</b>	<b>t-value</b>	<b>df</b>	<b>pvalue</b>
<i>Ankistrodesmus</i>	-0.32	[-0.42,-0.23]	-6.99	179	>0.001***
<i>Chlamydomonas</i>	-0.04	[-0.13,0.04]	-1.08	177	0.281
<i>Chlorella</i>	-0.13	[-0.23,-0.04]	-2.76	169	0.006**
<i>Monoraphidium</i>	-0.11	[-0.18,-0.04]	-3.13	174	0.002**
<i>Scenedesmus</i>	-0.10	[-0.19,-0.001]	-2.00	174	0.046*
<i>Raphidocelis</i>	-0.57	[-0.67,-0.47]	-11.5	169	>0.001***

## S9. Competitive advantage at day 5 and day 23

In addition to our main results at day 14 of the competition experiment, we also measured competitive advantage at day 5 and day 23. We trained linear discrimination algorithms on isolate data collected at day 5 for day 5 and day 14 for day 23 respectively, except for the very low phosphate concentration for which we trained the discrimination algorithm on the whole dataset to have a better discrimination function given the low density of cells. We chose to train the data using isolate data collected at day 14 for day 23 as the lower noise in the training dataset gave better discrimination results. The discrimination algorithms give 78 % accuracy in discriminating between species at day 5, and 70 % at day 23.

We compared the competitive advantage at day 5 and day 23 to the results of the theory (Table S9A and S9B). Because of a technical problem, we lost results from 63 samples out of 90 from 25°C and  $\text{PO}_4^{3-} = 1 \mu\text{mol L}^{-1}$  at day 5. Thus comparisons for this phosphate level for day 5 are to be taken with caution due to low sample size. We note that the results are similar between days, with an overall agreement between theory and experiment of 56 % ( $R_\infty$ ) and 66 % ( $R$ ) at day 5 and of 63 ( $R_\infty$ ) and 68 % ( $R$ ) at day 23. Further, measured competitive advantage was correlated across days (Pearson  $r = 0.67$  [0.56, 0.75],  $t = 10.52$ ,  $df = 137$ ,  $p = 2e^{-16}$  and  $r = 0.53$  [0.45, 0.62],  $t = 11.0$ ,  $df = 293$ ,  $p > 2e^{-16}$  respectively for correlation between day 5 and 14 and for correlation between day 14 and 23). This suggests that the competitive advantage at 14 day did indeed carry the signature from the exponential growth phase at day 5 where no species were at carrying capacity, and that this carry-over effect was continued over longer time periods, at a time where all species were at carrying capacity (median time to carrying capacity during the growth rate experiment at 15 and 25°C: 11 and 9 days respectively at very low nutrient concentrations (0.1  $\mu\text{mol}\cdot\text{L}^{-1}$  of phosphate), 10.5 and 7 days at low nutrient concentrations (1  $\mu\text{mol}\cdot\text{L}^{-1}$  of phosphate), and 14.5 and 9 at high nutrient concentrations (30  $\mu\text{mol}\cdot\text{L}^{-1}$  of phosphate)). It is noteworthy that the predictive power of the model is lower at day 23 than at day 14, likely due to the lower accuracy of the discrimination algorithm.

**Table S9A. Proportion of competitive advantages correctly predicted by theory at day 5 using the linear discrimination algorithm.**

Analogous to Table 1 in the main text, but for day 5.

	$R_\infty$		$R$		$N$
<i>Full dataset</i>					
	0.56	(0.137)	0.66	(0.007)	192
<i>By temperature</i>					
$T = 15^\circ\text{C}$	0.50	(0.458)	0.64	(0.113)	58
$T = 25^\circ\text{C}$	0.59	(0.112)	0.67	(0.025)	134
<i>By nutrient</i>					
$[\text{P}] = 0.1 \mu\text{mol}\cdot\text{L}^{-1}$	0.36	(0.820)	0.58	(0.262)	84
$[\text{P}] = 1 \mu\text{mol}\cdot\text{L}^{-1}$	0.65	(0.182)	0.65	(0.178)	23
$[\text{P}] = 30 \mu\text{mol}\cdot\text{L}^{-1}$	0.74	(0.019)	0.74	(0.016)	85
<i>By species</i>					
<i>Ankistrodesmus</i>	0.65	(0.022)	0.86	(0.000)	74
<i>Chlamydomonas</i>	0.53	(0.230)	0.65	(0.052)	57
<i>Chlorella</i>	0.62	(0.054)	0.75	(0.024)	63
<i>Monoraphidium</i>	0.53	(0.252)	0.71	(0.001)	72
<i>Scenedesmus</i>	0.49	(0.547)	0.57	(0.140)	68
<i>Raphidocelis</i>	0.56	(0.407)	0.32	(0.855)	50

**Table S9B. Proportion of competitive advantages correctly predicted by theory at day 23 using the linear discrimination algorithm.**

Analogous to Table 1 in the main text, but for day 23.

	$R_\infty$		$R$		$N$
<i>Full dataset</i>					
	0.63	(0.000)	0.68	(0.000)	339
<i>By temperature</i>					
$T = 15^\circ\text{C}$	0.69	(0.001)	0.66	(0.016)	170
$T = 25^\circ\text{C}$	0.57	(0.047)	0.69	(0.001)	169
<i>By nutrient</i>					
$[\text{P}] = 0.1 \mu\text{mol}\cdot\text{L}^{-1}$	0.38	(0.640)	0.92	(0.002)	26
$[\text{P}] = 1 \mu\text{mol}\cdot\text{L}^{-1}$	0.54	(0.251)	0.55	(0.249)	159
$[\text{P}] = 30 \mu\text{mol}\cdot\text{L}^{-1}$	0.77	(0.005)	0.77	(0.005)	154
<i>By species</i>					
<i>Ankistrodesmus</i>	0.63	(0.020)	0.74	(0.000)	123
<i>Chlamydomonas</i>	0.60	(0.022)	0.61	(0.045)	121
<i>Chlorella</i>	0.69	(0.000)	0.73	(0.000)	115
<i>Monoraphidium</i>	0.61	(0.125)	0.66	(0.038)	119
<i>Scenedesmus</i>	0.66	(0.009)	0.66	(0.023)	112
<i>Raphidocelis</i>	0.61	(0.016)	0.67	(0.002)	88