

# Rothamsted Repository Download

## A - Papers appearing in refereed journals

Jensen, J. L., Schjonning, P., Watts, C. W., Christensen, B. T. and Munkholm, L. J. 2018. Soil Water Retention: Uni-Modal Models of Pore-Size Distribution Neglect Impacts of Soil Management. *Soil Science Society of America Journal*. 83 (1), pp. 18-26.

The publisher's version can be accessed at:

- <https://dx.doi.org/10.2136/sssaj2018.06.0238>
- <https://dl.sciencesocieties.org/publications/sssaj/articles/0/0/sssaj2018.06.0238>

The output can be accessed at: <https://repository.rothamsted.ac.uk/item/847y0>.

© 17 January 2019, Please contact [library@rothamsted.ac.uk](mailto:library@rothamsted.ac.uk) for copyright queries.

1 **Modelling Soil Water Retention: A Uni-modal Pore Size Distribution Does Not Reflect**  
2 **Reality**

3 Johannes L. Jensen<sup>\*a</sup>, Per Schjønning<sup>a</sup>, Christopher W. Watts<sup>b</sup>, Bent T. Christensen<sup>a</sup>, Lars J.  
4 Munkholm<sup>a</sup>

5 <sup>a</sup> *Department of Agroecology, Aarhus University, Tjele, Denmark*

6 <sup>2</sup> *Department of Sustainable Agriculture Sciences, Rothamsted Research, Harpenden, United*  
7 *Kingdom*

8 *\* Corresponding author: tel: +45 26 36 08 47, [jlj@agro.au.dk](mailto:jlj@agro.au.dk)*

9

10 **ACKNOWLEDGEMENTS**

11 The study was supported by the Green Development and Demonstration Programme (GUDP)  
12 of the Ministry of Environment and Food of Denmark (projects OptiPlant and OptiTill). The  
13 Rothamsted Long-term Experiments National Capability (grant code BBS/E/C00J0300) is  
14 supported by the UK Biotechnology and Biological Sciences Research Council (BBSRC) and  
15 the Lawes Agricultural Trust.

16 **SUPPLEMENTAL MATERIAL**

17 Supplemental material is available with the online version of this article. The supplemental  
18 document presents parameter estimates of the van Genuchten and double-exponential model  
19 for the 16 Danish top- and subsoils of the Jakobsen data set and the 16 plots at Highfield.

20 **Modelling Soil Water Retention: A Uni-modal Pore Size Distribution Does Not Reflect**  
21 **Reality**

22 **Abstract (It is 1683 characters incl. spaces ☐ should be max 1500)**

23 Reliable models for describing soil water retention are vital when simulating plant production  
24 and environmental effects in agroecosystems. Models for describing soil water retention often  
25 imply a uni-modal pore size distribution (PSD) such as the one suggested by van Genuchten  
26 (vanG). However, clear deviations from a uni-modal PSD has been documented. The double-  
27 exponential (Dex) model imply a bi-modal PSD and may better reflect reality. We evaluated  
28 how the vanG and Dex models fit to water retention data for top- and subsoil differing in  
29 texture, soil with contrasting management systems (Highfield), and a soil with different tillage  
30 practices (Flakkebjerg). Soils were subjected to matric potentials from -10 hPa to -1.5 MPa  
31 using conventional methods. The bi-modal Dex model provided a better fit to water retention  
32 data for relatively sandy top- and subsoil, and the contrasting treatments from Highfield and  
33 Flakkebjerg, than the uni-modal vanG model. Neither of the models worked well for highly  
34 sorted soils. Topsoil was less well described than subsoil when using the vanG model due to a  
35 more pronounced bi-modality of the PSD caused by increased soil organic carbon (SOC)  
36 content and tillage. The root mean square error of the vanG fit increased with an increase in  
37 SOC when going from the bare fallow to the grass treatment at Highfield. At the same time,  
38 the tillage intensity decreased, but the effect of SOC seemed to outweigh the lack of tillage.  
39 These observations were reflected in a more distinct bi-modality of the PSD for better  
40 structured soils. Consequently, we suggest that uni-modal models are too simplistic for  
41 describing management effects on PSD.

42 **Abbreviations:** AIC, Akaike's information criterion; C, capacity;  $d_2$ , dominating pore size of  
43 the structural peak; Dex, double-exponential; I, intensity; PSD, pore size distribution; Q,  
44 quantity; RMSE, root mean squared error; SOC, soil organic carbon; SOM, soil organic  
45 matter;  $V_1$ , textural void ratio;  $V_2$ , structural void ratio; vanG, van Genuchten; VIF, variance  
46 inflation factor.

47 **Core Ideas (3-5; 85 characters with spaces)**

- 48 • A uni- and a bi-modal soil water retention model were evaluated
- 49 • The bi-modal double-exponential model provided a better fit
- 50 • The uni-modal model fit was affected by texture, soil organic carbon and tillage
- 51 • Uni-modal models cannot accurately describe the pore size distribution

52

53 **Keywords (at least 2):** Pore size distribution; Soil water retention.

54 Soil water and air is crucial for plant growth, microbial activity and percolation (Rabot et al.,  
55 2018). Consequently, quantitative knowledge of the pore size distribution (PSD) is essential.  
56 As a result, reliable models for describing soil water retention are vital when simulating water  
57 and solute movement, plant growth, and microbial driven processes such as carbon turnover  
58 and denitrification.

59 Soil water can be described in terms of quantity (Q, e.g. volumetric water content) and  
60 intensity (I, e.g. matric potential). Consequently, Q at any I is the volumetric water content  
61 retained in all pore size classes smaller than that defined by I. The ratio of Q and I at any  
62 given I is then the capacity (C), providing a measure of the volume of pores for a given tube-  
63 equivalent pore size. Numerous models for describing soil water retention has been suggested,  
64 and the most widely used are uni-modal analytical expressions (Cornelis et al., 2005) such as  
65 the one proposed by Brooks and Corey (1964).

66 van Genuchten (1980) suggested the most widely used model for describing soil water  
67 retention (>10000 citations; Web of Science, May 2018). The fitting parameters of the van  
68 Genuchten (vanG) model are often used to indirectly determine the unsaturated hydraulic  
69 conductivity of a soil (Mualem, 1986), and many pedotransfer functions have been developed  
70 for predicting the vanG parameters based on basic soil properties (Cornelis et al., 2001;  
71 Minasny et al., 1999; Patil and Singh, 2016). The unsaturated hydraulic conductivity  
72 predicted by the vanG parameters are often used as input in simulation models such as Daisy  
73 (Hansen et al., 2012) and HYDRUS (Šimunek et al., 2012). The simulations are used to  
74 predict plant production and environmental effects and may be used as basis for political  
75 decision-making.

76 Also, the vanG model implies a uni-modal PSD. Uni-modal models implicitly assumes a  
77 maximum C at one specific I (corresponding to a specific pore size). However, presenting the  
78 size distribution of pores by frequency, by for example numerical differentiation of the Q/I  
79 curve, has documented clear and important deviations from a uni-modal PSD (Eden et al.,  
80 2011; Pulido-Moncada et al., 2018; Schjønning, 1992). This calls for a more flexible  
81 description of the pore system than that obtained by uni-modal expressions. Several non-uni-  
82 modal models have been proposed (e.g, Poulsen et al., 2006; Ross and Smettem, 1993).  
83 Dexter et al. (2008) proposed a double-exponential model (Dex), which describes bi-modal  
84 PSD, i.e. a size distribution of pores with two peaks. The two peaks are denoted textural and  
85 structural pore spaces, and the division is based on a mechanistic understanding of soil  
86 structure (Dexter, 1988; Dexter et al., 2008). Thus, the Dex model may be one way of trying  
87 to allow for a better description of reality. Further, the Dex model has the additional  
88 advantage in providing parameters with physical meaning when compared to the vanG model.  
89 The objective of this study was to evaluate how the uni-modal vanG and bi-modal Dex model  
90 fit to water retention data for 1) top- and subsoil with different texture, 2) soil with contrasting  
91 long-term crop rotations, and 3) a soil subject to different tillage practices.

## 92 **MATERIALS AND METHODS**

### 93 **The Jakobsen data set**

94 A PhD-study focused on the hydraulic properties of 16 contrasting soils distributed  
95 throughout Denmark, each with sampling done in the top- (~0.10 m) and subsoil (~ 0.50 m)  
96 layers (Jakobsen, 1989). The topsoil ranged from loamy sand to silt loam with the main part  
97 being sandy loam or loam soils. The Jyndevad and Tylstrup soils were extremely sorted with  
98 51.2 g 100 g<sup>-1</sup> minerals in the 200-500 µm fraction and 51.8 g 100 g<sup>-1</sup> minerals in the 63-125

99  $\mu\text{m}$  fraction, respectively. The soil textural composition of top- and subsoil can be seen in  
 100 Table 1 and 2, respectively.

101 **Table 1.** Soil textural composition and organic carbon (SOC) content in the  $\sim 0.10$  m layer of  
 102 the 16 Danish soils of the Jakobsen (1989) data set listed in order of increasing clay content.  
 103 The Rosin-Rammler parameters ( $\alpha$  and  $\beta$ ) were calculated by Eq. 6 and are based on the seven  
 104 particle size fractions listed in Jakobsen (1989).

Site	SOC	Clay, <2 $\mu\text{m}$	Silt, 2-20 $\mu\text{m}$	Silt, 20-63 $\mu\text{m}$	Sand, 63-2000 $\mu\text{m}$	$\alpha$	$\beta$
	(g 100 g <sup>-1</sup> minerals)					( $\mu\text{m}$ )	-
Hals	2.36	2.6	3.4	7.9	86.0	150	1.76
Tylstrup	1.30	3.7	4.9	17.2	74.2	88	3.58
Jyndevad	1.36	4.2	3.9	3.2	88.8	367	1.41
Borris	1.31	5.7	7.8	22.8	63.7	131	0.96
Hornum	1.86	5.8	8.4	13.3	72.5	180	0.93
Travsted	3.38	7.7	6.8	16.2	69.3	189	0.86
Foulum	1.49	7.9	10.1	15.6	66.4	176	0.75
Ødum	1.49	10.1	15.5	20.2	54.3	104	0.71
Årslev	1.36	10.6	14.9	21.1	53.4	95	0.79
Roskilde	1.43	10.8	17.3	19.3	52.7	93	0.74
Askov	1.55	11.0	12.6	16.5	59.9	124	0.77
Rønhave	1.24	14.5	15.6	27.5	42.4	65	0.78
Tystofte	1.18	14.7	16.4	19.5	49.4	75	0.73
Ø. Ulslev	1.38	15.8	15.5	16.5	52.2	102	0.58
Kalø	0.82	17.7	14.4	15.9	52.0	102	0.55
Højjer	1.73	18.6	15.4	39.9	26.0	42	1.00

105  
 106 **Table 2.** Soil textural composition and organic carbon (SOC) content in the  $\sim 0.50$  m layer of  
 107 the 16 Danish soils of the Jakobsen (1989) data set listed as in Table 1. The Rosin-Rammler  
 108 parameters ( $\alpha$  and  $\beta$ ) were calculated by Eq. 6 and are based on the seven particle size  
 109 fractions listed in Jakobsen (1989).

Site	SOC	Clay, <2 $\mu\text{m}$	Silt, 2-20 $\mu\text{m}$	Silt, 20-63 $\mu\text{m}$	Sand, 63-2000 $\mu\text{m}$	$\alpha$	$\beta$
	(g 100 g <sup>-1</sup> minerals)					( $\mu\text{m}$ )	-
Hals	0.17	2.0	0.5	1.0	96.5	190	3.31
Tylstrup	0.29	3.1	2.4	12.8	81.7	82	5.96
Jyndevad	0.35	3.5	1.9	1.0	93.6	359	2.25

Borris	0.29	11.2	7.3	14.9	66.6	136	0.90
Hornum	0.17	7.2	6.3	13.7	72.8	200	0.88
Travsted	0.35	10.8	6.7	10.8	71.7	194	0.84
Foulum	0.17	13.4	9.6	13.4	63.5	166	0.64
Ødum	0.17	16.5	12.6	16.4	54.4	106	0.60
Årslev	0.17	20.4	12.6	15.9	51.0	78	0.63
Roskilde	0.29	23.8	16.3	11.9	48.0	72	0.49
Askov	0.35	24.5	11.6	14.3	49.6	72	0.55
Rønhave	0.29	19.6	16.5	25.1	38.8	53	0.67
Tystofte	0.29	22.8	15.3	17.9	44.0	58	0.58
Ø. Ulslev	0.23	15.6	13.5	14.1	56.8	120	0.59
Kalø	0.29	26.8	12.4	14.3	46.6	77	0.43
Højer	0.24	7.9	6.4	35.6	50.1	69	3.08

110

111 The soils were all long-term arable and derived from the Weichsel glacial stage (glacial  
112 deposits: ten soils; glaciofluvial deposits: Jyndevad), the Saale glacial stage (glacial deposits:  
113 Borris and Travsted), the raised Holocene sea floor (Tylstrup and Hals), and one soil present-  
114 day marine marsh area (Højer).

115 At each site, topsoil was sampled in six plots of about one m<sup>2</sup>, whereas subsoil was sampled  
116 in one of these plots. In the topsoil, three 100 cm<sup>3</sup> soil cores (61 mm diameter, 34 mm height)  
117 were sampled in each plot providing 18 cores pr. site. In the subsoil, nine cores were sampled  
118 at each site.

### 119 **Rothamsted Highfield ley-arable experiment**

120 Data on soil texture, soil organic carbon (SOC) and pore characteristics for the Highfield  
121 long-term ley-arable experiment at Rothamsted Research, UK (51°80'N, 00°36'W) was  
122 recently published by Jensen et al. (2018) and Obour et al. (2018). In this study, we use these  
123 data for four treatments: Bare fallow maintained free of plants by regular tillage since 1959,  
124 Continuous arable rotation with winter cereals since 1948, Ley-arable rotation; a three-year  
125 grass-clover ley followed three years arable since 1948, and grass, grassland ploughed and  
126 reseeded to grass in 1948. The bare fallow treatment was cultivated three to five times per



127 year, arable once a year, ley-arable once in two years (in six-year cycle) and grass had not  
128 been cultivated since 1947. The arable, ley-arable and grass treatments were embedded in a  
129 randomized block design, whereas the bare fallow plots were not part of the original design  
130 and located at one end of the experiment. The soil type is silt loam and is classified as Aquic  
131 Paleudalf (USDA Soil Taxonomy System). The parent material includes a silty (loess-  
132 containing) deposit overlying and mixed with clay-with-flints (Avery and Catt, 1995).  
133 Soil was sampled in spring 2015. Six 100 cm<sup>3</sup> soil cores (61 mm diameter, 34 mm height)  
134 were extracted from the ~0.10 m soil layer of each of a total of four plots providing 24 cores  
135 per treatment. More details are given in Jensen et al. (2018) and Obour et al. (2018).

#### 136 **Flakkebjerg tillage experiment**

137 Previously published data on SOC and pore characteristics for the long-term experiment on  
138 conservation tillage at the Flakkebjerg experimental site in Denmark (55°19'N, 11°23'E) was  
139 used. Treatments kept under moldboard plowing to 0.20-m depth and direct drilling were  
140 compared after eleven years of different tillage practices. The treatments were embedded in a  
141 split-plot experiment with four replicates. The soil type is a sandy loam soil with 14.7 % clay  
142 (<2 µm), 13.7 % silt (2-20 µm), 42.6 % fine sand (20-200 µm) and 27% coarse sand (200-  
143 2000 µm). The soil is classified as Oxyaquic Agriudoll (USDA Soil Taxonomy System). The  
144 rotation included winter and spring crops (mainly cereals) with residues removed.

145 Soil was sampled in autumn 2013. Six 100 cm<sup>3</sup> soil cores (61 mm diameter, 34 mm height)  
146 were extracted from the 0.12-0.16 m soil layer of each plot providing 24 cores pr. treatment.  
147 Further details can be found in Abdollahi and Munkholm (2017).

#### 148 **Laboratory measurements**

149 Soil texture was determined on air dry bulk soil (<2-mm) with a combined hydrometer/sieve  
150 method (Gee and Or, 2002). Samples from Highfield were treated with hydrogen peroxide to  
151 remove soil organic matter (SOM), while this was not done for Flakkebjerg and the Jakobsen  
152 soils. The content of SOC was measured by dry combustion using a Thermo Flash 2000 NC  
153 Soil Analyser (Thermo Fisher Scientific, Waltham MA, USA) for Highfield and Flakkebjerg,  
154 and a LECO CNS-1000 analyzer (LECO Corporation, St. Joseph, MI) for the Jakobsen soils.

155 Before measuring soil water retention, the soil cores were placed on the top of a tension table  
156 to be saturated from beneath. For the Jakobsen data set, soil water retention was measured at -  
157 4 (Højer only), -10, -16, -50, -100, -160 and -500 hPa matric potential, and at -10, -30, -100, -  
158 300 and -1000 hPa matric potential for Highfield and Flakkebjerg using tension tables and  
159 pressure plates (Dane and Hopmans, 2002). The soil cores were oven-dried (105 °C for 24 h),  
160 and bulk density calculated. For Highfield, bulk density was corrected for stone weight and  
161 volume because the soil contained a significant amount of stones. Soil porosity was estimated  
162 from bulk density and particle density. Particle density was measured by the pycnometer  
163 method (Flint and Flint, 2002). For Highfield, particle density was measured on one plot from  
164 each treatment, and the particle density for the remaining plots were predicted from SOC by a  
165 linear regression model. For Flakkebjerg, a particle density of 2.65 g cm<sup>-3</sup> was used based on  
166 previous studies (Abdollahi and Munkholm, 2017).

167 Water retention at -1.5 MPa was determined on <2-mm air-dry soil for each site and depth for  
168 the Jakobsen soils and at plot level for Highfield. For the Jakobsen and Highfield soils a  
169 pressure plate system and a WP4-T Dewpoint Potentiometer, respectively, was used (Scanlon  
170 et al., 2002). For Flakkebjerg, water retention at -1.5 MPa was predicted based on clay and  
171 SOC content using Eq. 1 in Hansen (1976).

172 Pore water suction was assumed to relate to an average pore size by the approximate relation:

$$173 \quad d = -3000/h \quad [1]$$

174 where  $d$  is the tube-equivalent pore diameter ( $\mu\text{m}$ ) and  $h$  is the soil matric potential (hPa).

### 175 **Soil water retention models**

176 The water retention data was fitted to the van Genuchten (1980) model (termed vanG):

$$177 \quad \theta = (\theta_{\text{sat}} - \theta_{\text{res}}) \left[ 1 + (\alpha h)^n \right]^{-m} + \theta_{\text{res}} \quad [2]$$

178 where  $\theta_{\text{sat}}$  and  $\theta_{\text{res}}$  are the water contents at saturation and the residual water content,  
179 respectively,  $h$  is the soil matric potential,  $\alpha$  is a scaling factor for  $h$  and  $n$  and  $m$  are  
180 parameters that control the shape of the curve. The widely used Mualem (1976) restriction ( $m$   
181  $= 1-1/n$ ) was used to prevent over-parametrization (Dexter et al., 2008) and unstable results  
182 (van Genuchten et al., 1991). The Mualem restriction is also recommended, when only  
183 measured values in the wet range are used (van Genuchten et al., 1991). The PSD predicted  
184 by the vanG model was obtained by differentiating Eq. 2 with respect to matric potential:

$$185 \quad \frac{d\theta}{d(\log_{10} h)} = (\theta_{\text{sat}} - \theta_{\text{res}}) (\alpha n (\alpha h)^{n-1} (-m) (1 + (\alpha h)^n)^{-m-1}) h \ln 10 \quad [3]$$

186 The double-exponential model proposed by Dexter et al. (2008) was fitted to water retention  
187 data (termed Dex):

$$188 \quad \theta = C + A_1 e^{(-h/h_1)} + A_2 e^{(-h/h_2)} \quad [4]$$

189 where  $C$  is the residual water content,  $A_1$  and  $A_2$  are the amount of textural and structural pore  
190 space, respectively, and  $h_1$  and  $h_2$  are the characteristic pore water suctions at which the

191 textural and structural pore spaces empty, respectively. The PSD predicted by the Dex model  
192 was obtained by differentiating Eq. 4 with respect to matric potential:

$$193 \quad \frac{d\theta}{d(\log_{10}h)} = -\frac{A_1}{h_1} e^{(-h/h_1)} h \ln 10 - \frac{A_2}{h_2} e^{(-h/h_2)} h \ln 10 \quad [5]$$

194 The parameters of the vanG model were obtained using the curve-fitting program RETC (van  
195 Genuchten et al., 1991), which is based on a nonlinear least-squares optimization approach.  
196 Similarly, the parameters of the Dex model were obtained by nonlinear regression analysis to  
197 achieve the smallest residual sum of squares.

### 198 **Calculations and statistics**

199 The Rosin-Rammler equation (Eq. 2 in Rosin and Rammler (1933)) was fitted to the seven  
200 chemically dispersed particle size fraction listed in Jakobsen (1989), i.e. <2, 2-20, 20-63, 63-  
201 125, 125-200, 200-500 and 500-2000  $\mu\text{m}$ , for each soil. It can be written as:

$$202 \quad P(X < x) = 1 - \exp\left(-\left(\frac{x}{\alpha}\right)^\beta\right) \quad [6]$$

203 where  $P(X < x)$  is the fraction by weight of particles less than size  $x$ ,  $\alpha$  indicates the coarseness  
204 of particles and  $\beta$  indicates the spread of particle sizes. Eq. 6 described the particle size  
205 distribution of the soils well, with coefficients of determination ( $R^2$ ) from 0.95 to 1.00.

206 For the statistical analysis, the R-project software package Version 3.4.0 (R Foundation for  
207 Statistical Computing) was used. Treatment effects for Highfield was analyzed as described in  
208 Jensen et al. (2018). The key indices of goodness of fit were Akaike's information criterion  
209 (AIC), which was used to compare models with different number of parameters (Akaike,  
210 1973), and the root mean squared error (RMSE):

211 
$$\text{RMSE} = \sqrt{\frac{1}{N} \sum (\theta_{\text{meas}} - \theta_{\text{fitted}})^2} \quad [7]$$

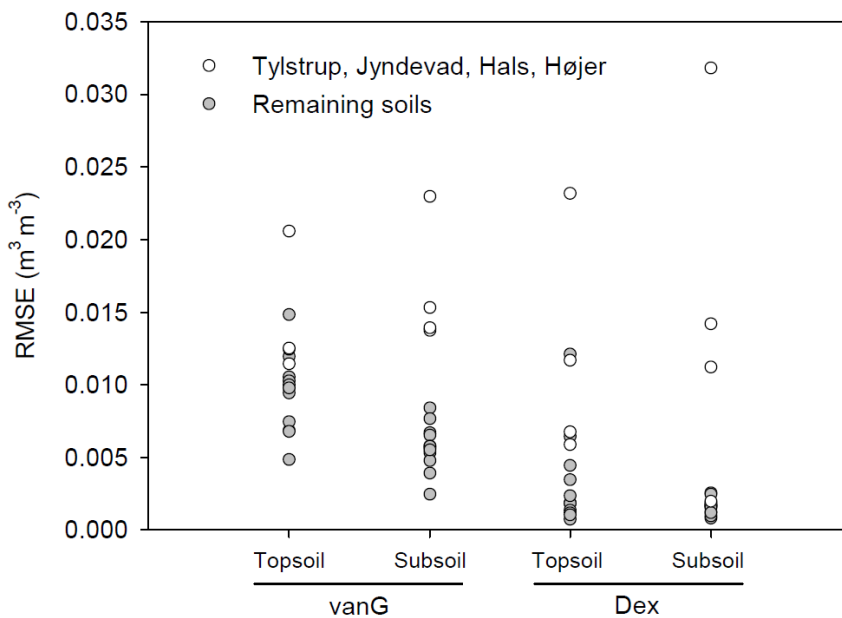
212 where  $N$  is the number of matric potentials. Multiple linear regression was used to identify  
213 how structural void ratio ( $V_2$ ) related to SOC, soil texture and void ratio. Structural void ratio  
214 was calculated as follows:  $V_2 = A_2 / (1-P)$ , where  $A_2$  is the Dex model estimate of structural  
215 pore space, and  $P$  is porosity. Likewise, textural void ratio ( $V_1$ ) was calculated. The variance  
216 inflation factor (VIF) was calculated when more than one predictor was used in the  
217 regression. The VIF expresses the degree of multicollinearity among the predictors. Upper  
218 value limits of VIF for non-erroneous conclusions from multiple regressions has been set to  
219 five (Rogerson, 2001) or ten (Kutner et al., 2004).

## 220 **RESULTS**

### 221 **Jakobsen data set**

222 The soils differed in their textural composition and SOC content (Tables 1 and 2). In the  
223 topsoil, clay ranged from 2.6 to 18.6 g 100 g<sup>-1</sup> minerals and SOC from 0.82 to 3.38 g 100 g<sup>-1</sup>  
224 minerals, whereas the range in the subsoil was from 2.0 to 26.8 g clay 100 g<sup>-1</sup> minerals and  
225 from 0.17 to 0.35 g SOC 100 g<sup>-1</sup> minerals. The  $\alpha$ -parameter for topsoil ranged from 42 to 367  
226  $\mu\text{m}$  and for subsoil from 53 to 359  $\mu\text{m}$ . The Jyndevad soil, however, stand out being very  
227 coarse textured ( $\alpha_{\text{topsoil}}=367 \mu\text{m}$ ,  $\alpha_{\text{subsoil}}=359 \mu\text{m}$ ), and the range changed to 42 to 200  $\mu\text{m}$  if  
228 omitting Jyndevad. The  $\beta$ -parameter describes the spread of particle sizes, with large values  
229 indicating that the soil is well sorted (a narrow range of particle sizes), and small values  
230 indicating that the soil is graded with an evenly distributed mass of particles in all size  
231 classes. The  $\beta$ -parameter for topsoil ranged from 0.55 to 3.58 and for subsoil from 0.43 to

232 5.96. The twelve glacial till soils, however, had a narrow range from 0.43 to 0.96, whereas  
 233 Hals, Tylstrup, Jyndevad and Højer were highly sorted with  $\beta > 1$  (Tables 1 and 2).  
 234 Mean AIC and RMSE values, when using the vanG and Dex models were calculated for the  
 235 top- and subsoil. For topsoil, values of AIC were -58.6 and -70.6 and RMSE were 0.011 and  
 236 0.005  $\text{m}^3 \text{m}^{-3}$  using vanG and Dex, respectively. For subsoil, values of AIC were -63.2 and -  
 237 75.1 and RMSE were 0.008 and 0.005  $\text{m}^3 \text{m}^{-3}$ . However, the four highly sorted soils ( $\beta > 1$ ) had  
 238 relatively poor goodness of fit measures both when using the vanG and Dex models (Fig. 1,  
 239 Tables S1 and S2).



240  
 241 **Fig. 1.** The root mean squared error (RMSE) value for the Danish top- and subsoil of the  
 242 Jakobsen data set using the van Genucthen (vanG) or double-exponential (Dex) model.  
 243 If the highly sorted soils are omitted in the calculation of the mean AIC and RMSE values, the  
 244 vanG model gives AIC values of -59.6 and -66.8 and RMSE values of 0.009 and 0.006  $\text{m}^3 \text{m}^{-3}$   
 245 in top- and subsoil, respectively. The Dex model gives AIC values of -75.1 and -80.9 and  
 246 RMSE values of 0.003 and 0.002  $\text{m}^3 \text{m}^{-3}$  in top- and subsoil, respectively. The lower AIC and

247 RMSE values obtained for the Dex compared to the vanG model indicate a better ability to  
248 describe data.

249 We tested the correlation between the structural void ratio ( $V_2$ ) and the variables  $\alpha$ ,  $\beta$ , void  
250 ratio, SOC and clay content. This was done for both top- and subsoils and with and without  
251 exclusion of the highly sorted soils ( $\beta > 1$ ). For topsoil samples  $V_2$  could be well predicted by  
252  $\log(\beta)$  and clay content:

$$253 \quad V_2 = 0.558^{***}(\pm 0.118) \times \log(\beta) - 0.011^*(\pm 0.005) \times \text{clay} + 0.424^{***}(\pm 0.048),$$
$$254 \quad s = 0.068, R^2 = 0.84 \text{ [8]}$$

255 Excluding the highly sorted samples from topsoil gave:

$$256 \quad V_2 = 0.878^{***}(\pm 0.143) \times \beta - 0.441^{**}(\pm 0.110), s = 0.057, R^2 = 0.79 \text{ [9]}$$

257 For subsoil samples  $V_2$  could be well predicted by  $\log(\beta)$  and  $\alpha$ :

$$258 \quad V_2 = 0.592^{***}(\pm 0.078) \times \log(\beta) - 0.001^{**}(\pm 0.0003) \times \alpha + 0.184^{**}(\pm 0.050),$$
$$259 \quad s = 0.100, R^2 = 0.85 \text{ [10]}$$

260 Excluding the highly sorted samples from subsoil gave:

$$261 \quad V_2 = 0.289^{***}(\pm 0.025) \times \beta, s = 0.057, R^2 = 0.55 \text{ [11]}$$

262 In Eq. 8-11, the numbers in parentheses are standard errors of estimate, and  $s$  is the standard  
263 deviation of the predicted value. When developing the four models, we tested for  
264 multicollinearity and interaction among the predictors, but only low VIF values and no  
265 significant interactions were found.

266 **Rothamsted Highfield ley-arable experiment**

267 The soils at Highfield ranged from 0.84 to 4.04 g SOC 100 g<sup>-1</sup> minerals and soil texture was in  
268 general not significantly different between treatments (Table 3).

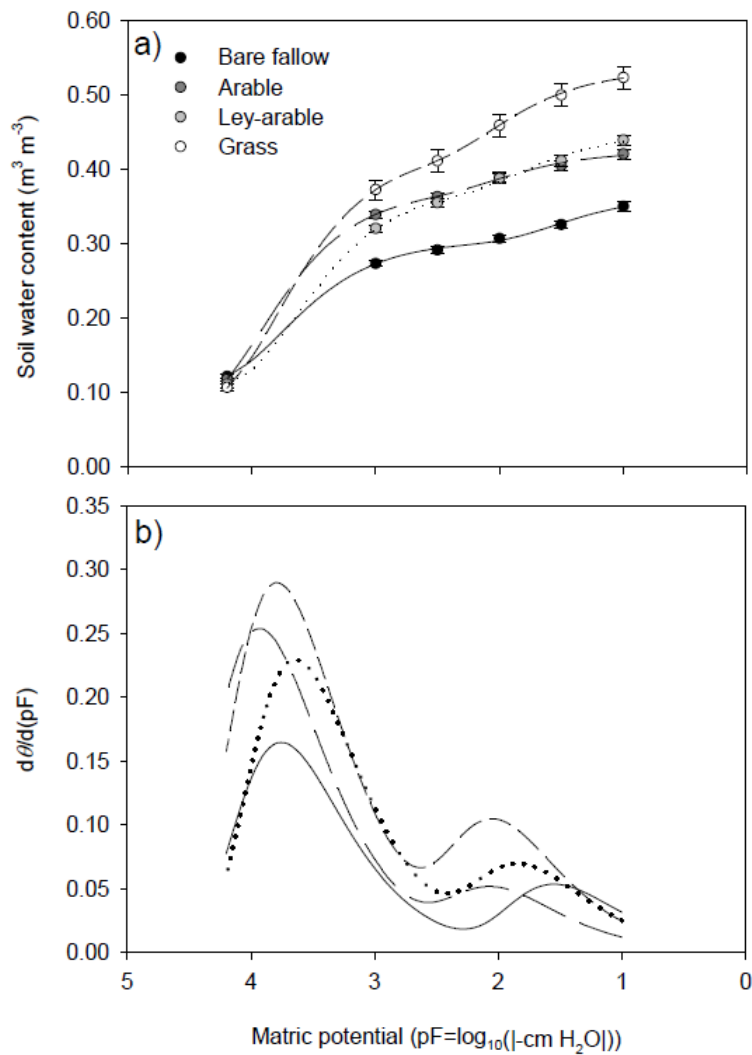
269 **Table 3.** Soil textural composition and organic carbon (SOC) content of the four treatments  
270 from Highfield. Within rows, letters denote statistical significance at  $P<0.05$  for the  
271 comparison of Arable, Ley-arable and Grass. An asterisk (\*) indicates if Bare fallow is  
272 significantly different from the other treatments based on a pairwise  $t$ -test. Data from Jensen  
273 et al. (2018).

Treatment	SOC	Clay, <2 $\mu\text{m}$	Silt, 2-20 $\mu\text{m}$	Silt, 20-63 $\mu\text{m}$	Sand, 63-2000 $\mu\text{m}$
	(g 100 g <sup>-1</sup> minerals)				
Bare fallow	0.90	27.0	24.9	33.5	14.6
Arable	1.73a*	26.4	26.3	31.8	15.5
Ley-arable	2.16a*	25.5	26.1	32.4	16.0
Grass	3.29b*	26.1	27.2*	31.9	14.8

274

275 Thus, the effect of contrasting long-term management could be investigated without  
276 confounding effects related to variations in soil type. The Dex model generally fitted the  
277 water retention data for the contrasting treatments well (Fig. 2a).



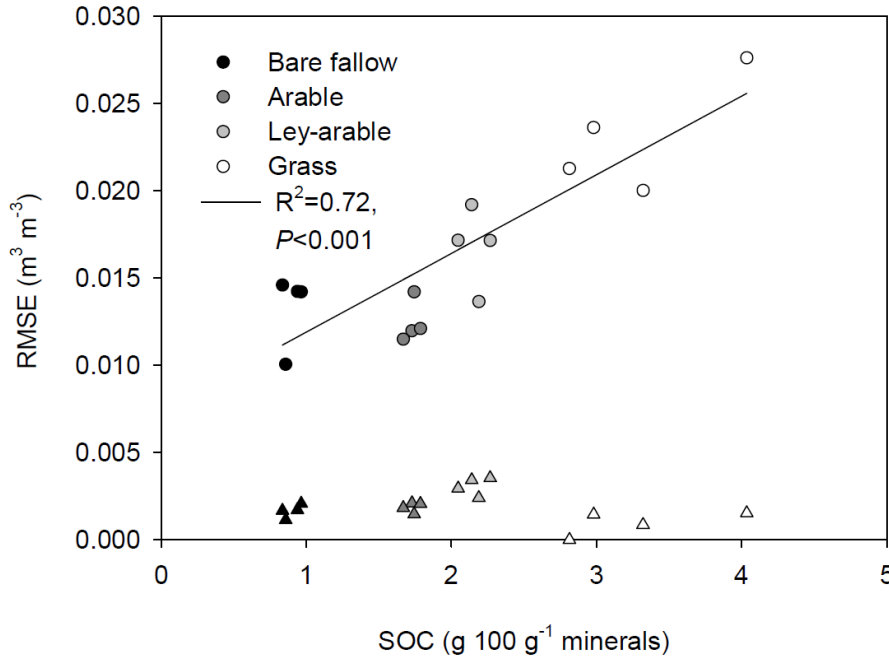


278

279 **Fig. 2. (a)** Measured volumetric water content for the four treatments at Highfield and fits of  
 280 the double-exponential (Dex) model as a function of matric potential. The standard error of  
 281 the mean are indicated ( $n=4$ ). **(b)** Pore size distribution ( $d\theta/d(pF)$ ) as a function of matric  
 282 potential for the four treatments. Eq. 5 was used to obtain the pore size distributions.

283 Mean values of AIC, when using the vanG and Dex models were -43.8 and -69.1,  
 284 respectively. Similarly, mean values of RMSE were larger when using the vanG compared to  
 285 the Dex model with values of 0.016 and 0.002 m³ m⁻³, respectively. The RMSE when using  
 286 the vanG model increased from 0.010 to 0.028 m³ m⁻³ with an increase in SOC from 0.84 to  
 287 4.04 g 100 g⁻¹ minerals (Fig. 3), whereas no systematic error was observed when using the

288 Dex model ( $P=0.532$ ). Parameter estimates and goodness of fit measures for the 16 individual  
 289 plots can be seen in Table S3.



290  
 291 **Fig. 3.** The root mean squared error (RMSE) value as a function of soil organic carbon (SOC)  
 292 for the four treatments at Highfield using the van Genuthen (vanG) model (circle symbols)  
 293 and the double-exponential (Dex) model (triangle symbols).

294 Textural ( $V_1$ ) and structural void ratio ( $V_2$ ) increased with increasing SOC content and  
 295 decreasing tillage intensity ( $V_1$ :  $R^2=0.91$ ,  $P<0.001$ ,  $V_2$ :  $R^2=0.74$ ,  $P<0.001$ ). The dominating  
 296 pore size of the structural peak ( $d_2$ ) was estimated to 86  $\mu\text{m}$  for the bare fallow treatment,  
 297 whereas it was significantly lower for the arable and grass treatments, and in between for ley-  
 298 arable treatment (Table 4).

299 **Table 4.** Estimated parameters of the double-exponential model (Dex) of the four treatments  
 300 from Highfield. Within rows, letters denote statistical significance at  $P<0.05$  for the  
 301 comparison of Arable, Ley-arable and Grass. An asterisk (\*) indicates if Bare fallow is

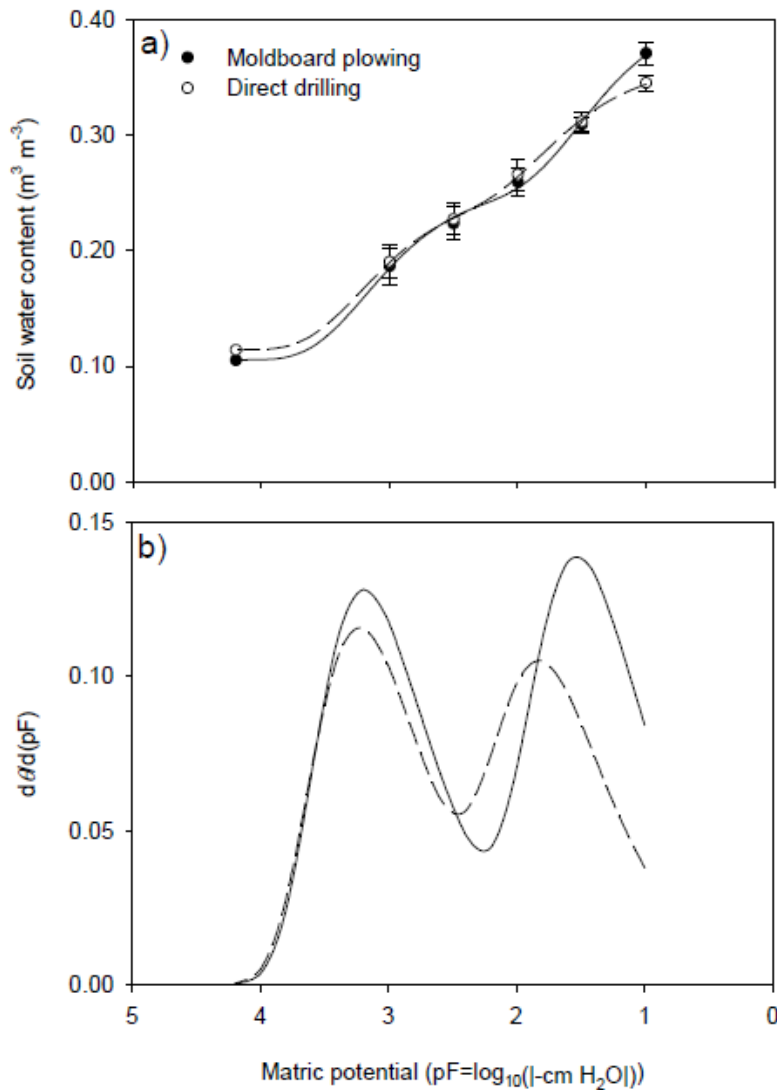
302 significantly different from the other treatments based on a pairwise *t*-test.  $d_1$  and  $d_2$  indicate  
 303 the dominating pore size of the textural and structural peak, respectively, and were estimated  
 304 by Eq. 1.

Treatment	Parameters of the Dex model						
	$C$ ( $\text{m}^3 \text{m}^{-3}$ )	$A_1$ ( $\text{m}^3 \text{m}^{-3}$ )	$h_1$ (hPa)	$d_1$ ( $\mu\text{m}$ )	$A_2$ ( $\text{m}^3 \text{m}^{-3}$ )	$h_2$ (hPa)	$d_2$ ( $\mu\text{m}$ )
Bare fallow	0.110	0.195	5729	0.5	0.061	35	86
Arable	0.068a*	0.305ab*	8707b*	0.3	0.051a	97b*	31
Ley-arable	0.104b	0.271a*	4379a	0.7	0.073a	63a	48
Grass	0.080ab*	0.345b*	6216a	0.5	0.110b*	102b*	29

305

### 306 **Flakkebjerg tillage experiment**

307 Moldboard plowing to 0.20-m depth and direct drilling had contents of 1.25 and 1.08 g SOC  
 308  $100 \text{ g}^{-1}$  minerals, respectively, in the 0.12-0.16 m layer. The Dex model fitted the two  
 309 treatments well (Fig. 4), and better compared to the vanG model as revealed by lower AIC  
 310 and RMSE values (Plowing:  $\text{AIC}_{\text{vanG}}=-53.8$  and  $\text{AIC}_{\text{Dex}}=-58.5$ ,  $\text{RMSE}_{\text{vanG}}=0.006 \text{ m}^3 \text{m}^{-3}$  and  
 311  $\text{RMSE}_{\text{Dex}}=0.003 \text{ m}^3 \text{m}^{-3}$ ; Direct drilling:  $\text{AIC}_{\text{vanG}}=-62.5$  and  $\text{AIC}_{\text{Dex}}=-69.3$ ,  $\text{RMSE}_{\text{vanG}}=0.003$   
 312  $\text{m}^3 \text{m}^{-3}$  and  $\text{RMSE}_{\text{Dex}}=0.001 \text{ m}^3 \text{m}^{-3}$ ).



313

314 **Fig. 4. (a)** Measured volumetric water content for moldboard plowing to 0.20-m depth and  
 315 direct drilling and fits of the double-exponential (Dex) model as a function of matric  
 316 potential. The standard error of the mean are indicated ( $n=4$ ), except for pF 4.2 which is  
 317 predicted based on Eq. 1 in Hansen (1976). **(b)** Pore size distribution ( $d\theta/d(pF)$ ) as a function  
 318 of matric potential for the two treatments. Eq. 5 was used to obtain the pore size distributions.  
 319 Structural void ratio ( $V_2$ ) for moldboard plowing and direct drilling was 0.30 and 0.19,  
 320 respectively. The dominating pore size of the structural peak ( $d_2$ ) was 50  $\mu\text{m}$  for direct drilling  
 321 and 94  $\mu\text{m}$  for moldboard plowing.

## 322 **DISCUSSION**

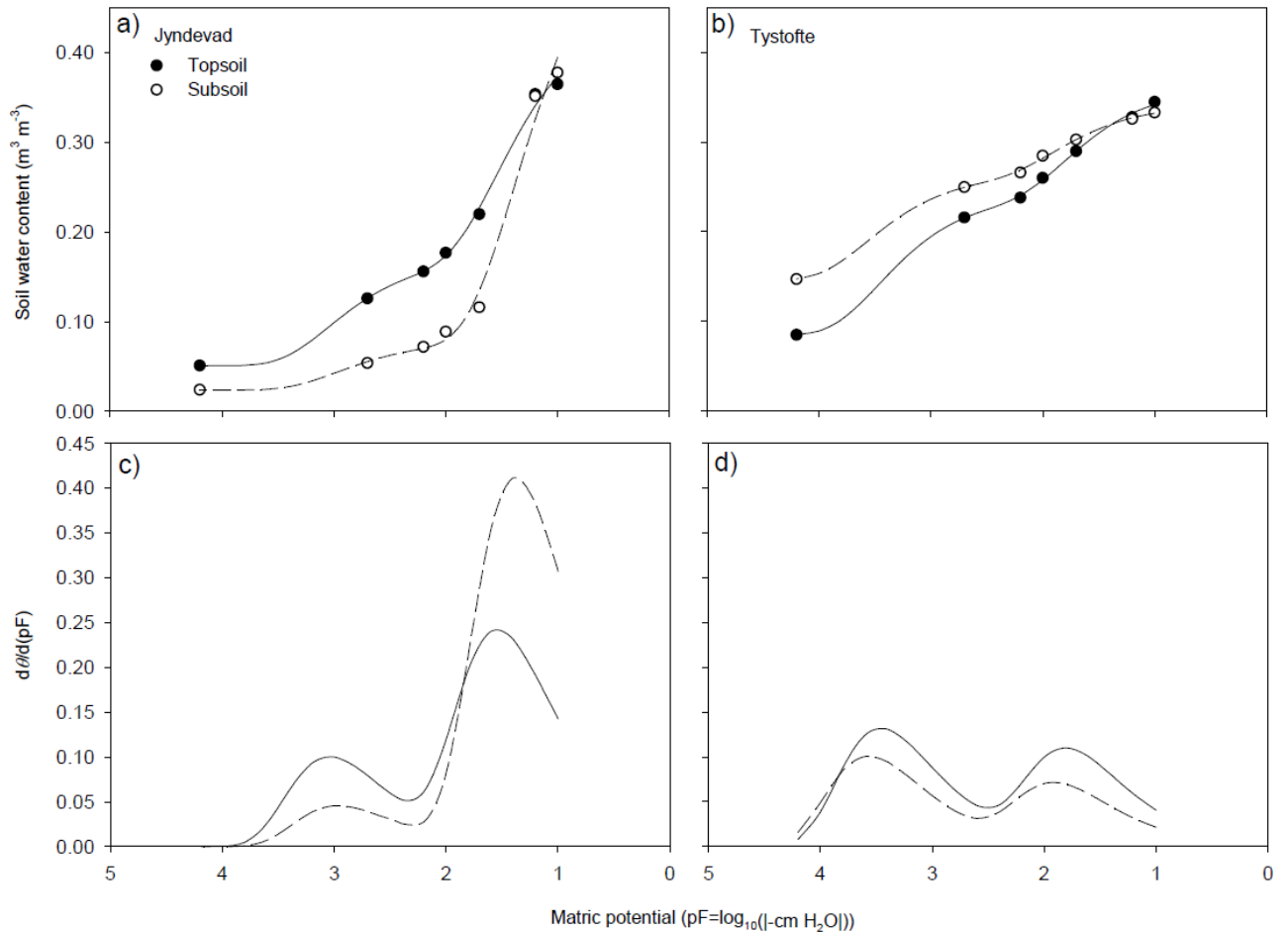
### 323 **Model fit**

324 The Dex model provided a better fit to soil water retention data than the vanG model for the  
325 Jakobsen glacial till top- and subsoils. The same was the case for the contrasting treatments  
326 from Highfield and Flakkebjerg. Thus, the PSD for these soils was better described with a bi-  
327 rather than a uni-modal model. Also, Schjønning (1992) observed that the vanG model was  
328 not able to describe a general pattern of a double-peak PSD for glacial till soils. Similarly,  
329 Dexter et al. (2008) and Berisso et al. (2012) found that the Dex model fitted their data better  
330 than the vanG model. Dexter et al. (2008) based their analysis on 42 Polish soils (26 topsoils,  
331 six samples from 0.30-0.35 m depth and ten subsoils) ranging from 2 to 25 g clay 100 g<sup>-1</sup>,  
332 whereas the study by Berisso et al. (2012) focused on a sandy clay loam ranging from 19 to  
333 27 g clay 100 g<sup>-1</sup>. Our study included soils ranging in clay content from 2.0 to 30.0 g 100 g<sup>-1</sup>  
334 minerals substantiating that the Dex model is superior for soils <30 g clay 100 g<sup>-1</sup> minerals. In  
335 summary, uni-modal models seem too simplistic for describing the size distribution of pores  
336 in most soils with less than 30 g clay 100 g<sup>-1</sup> minerals.

337 Neither of the models worked well for highly sorted soils ( $\beta > 1$ ). This finding calls for  
338 alternative water retention models for soils with a narrow distribution of pore sizes. Dexter et  
339 al. (2008) mentioned the problems associated with the use of the Dex model for uniform  
340 sands. However, we emphasize that the Dex as well as the vanG model cannot describe highly  
341 sorted soils well regardless of the dominating particle size.

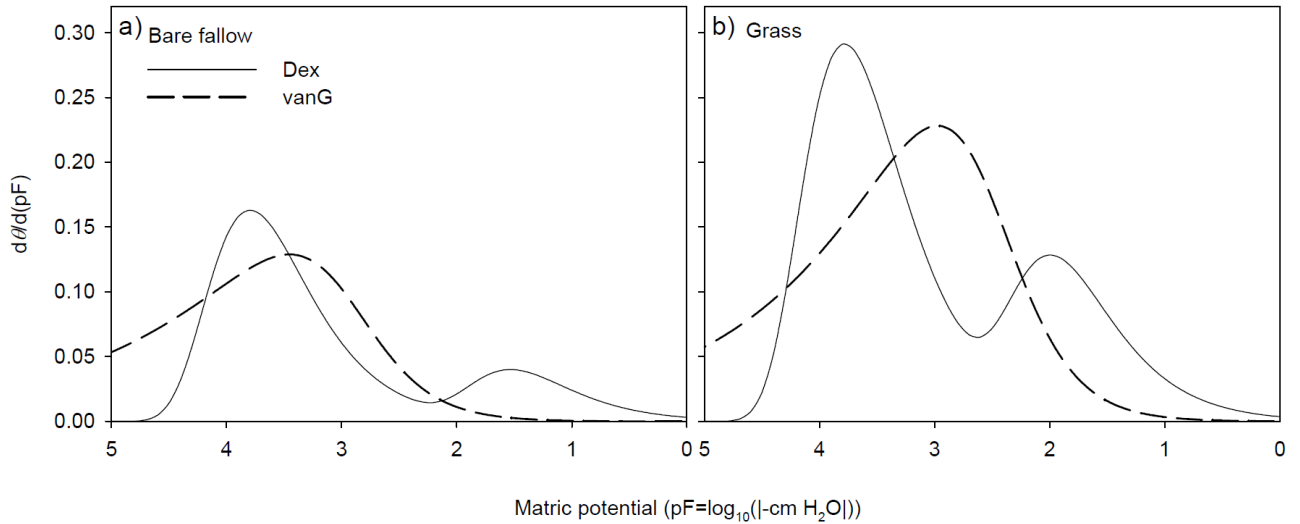
### 342 **Pitfalls using uni-modal models – effects of soil organic carbon and tillage**

343 The topsoil of the Jakobsen data set were less well described with the vanG model than the  
 344 subsoil (Fig. 1). This could be ascribed to a more distinct bi-modal PSD for topsoil, which can  
 345 be interpreted as a combination of larger SOC content and the presence of tillage (Fig. 5).



346  
 347 **Fig. 5.** (a, b) Measured volumetric water content for Jyndeved and Tystofte top- and subsoils  
 348 of the Jakobsen data set and fits of the double-exponential (Dex) model as a function of  
 349 matric potential. (c, d) Pore size distribution ( $d\theta/d(pF)$ ) as a function of matric potential for  
 350 Jyndeved and Tystofte top- and subsoils. Eq. 5 was used to obtain the pore size distributions.  
 351 Tillage increases the amount of large structural pores, and the effect of structure forming  
 352 agents in subsoil are much reduced, which limits structural pore space at depth. Similarly, the  
 353 systematic increase in RMSE with increasing SOC content for Highfield (Fig. 3) could be

354 ascribed to a more pronounced bi-modal behaviour (Fig. 2b), most clearly seen for the grass  
355 treatment (Fig. 6).



356

357 **Fig. 6.** Pore size distribution ( $d□/d(pF)$ ) as a function of matric potential for (a) the bare  
358 fallow and (b) the grass treatment at Highfield either obtained by Eq. 5 for the double-  
359 exponential (Dex) model (solid line) or Eq. 3 for the van Genuchten (vanG) model (dashed  
360 line).

361 Soil organic carbon content may increase the textural pore space especially in soils with less  
362 than 19 g clay 100 g<sup>-1</sup> (Rawls et al., 2003) due to its absorptive capacity for water. The  
363 structural pore space is mainly affected by SOC through improved aggregation (Bronick and  
364 Lal, 2005). Both  $V_1$  and  $V_2$  were positively affected by SOC at Highfield. However, the  
365 estimate of the mean size of structural voids ( $d_2$ ) decreased with SOC. For Flakkebjerg, where  
366 plowing was compared to direct drilling, both  $V_2$  and  $d_2$  increased with tillage intensity. The  
367 limited effect of tillage on  $V_2$  when going from the grass to the bare fallow treatment at  
368 Highfield suggests that the improving effect of SOC on soil structure outweighed any effect  
369 of tillage for these long-term treatments. Interestingly,  $d_2$  was larger for the bare fallow than

370 the grass treatment indicating that large pores were introduced by tillage as seen in  
371 Flakkebjerg.

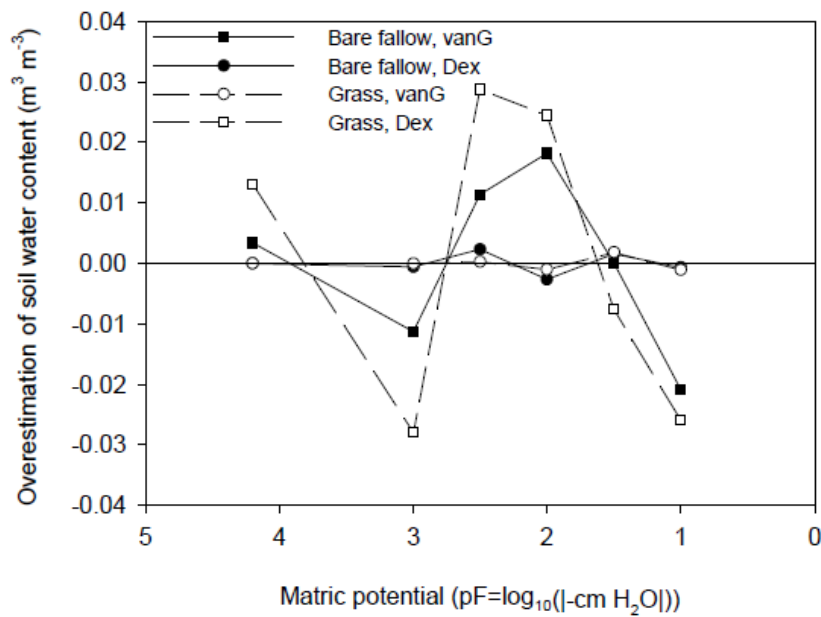
### 372 **Structural void ratio**

373 The structural void ratio ( $V_2$ ) is an important parameter for soil functioning such as air  
374 exchange and water uptake by plants. We used  $V_2$  rather than volumetric water content in  
375 order to allow a comparison across soils with varying bulk density. From the previous section,  
376 we found that SOC content and tillage intensity were important drivers for  $V_2$ . In addition,  
377 soil texture affects  $V_2$  through a positive relation to  $\beta$  (Eq. 8 to 11) indicating that the more  
378 sorted soils have larger  $V_2$  than graded soils. This is in agreement with Ehlers and Claupein  
379 (1994), who reported that graded coarse textured soils readily compact to high densities.  
380 Similarly, Schjønning and Thomsen (2013) found that graded soils low in SOC showed a  
381 hard-setting behavior. A low  $V_2$  may reduce soil aeration and potentially affect root growth  
382 and air exchange processes in a negative direction. Therefore, specific management strategies  
383 should be targeted on graded soil low in SOC, e.g. SOM promoting management.

### 384 **Implications**

385 Predicting water retention using the vanG model induced a larger error in top- than subsoil.  
386 Consequently, modelling whole profile water flow may be systematically biased down  
387 through the soil profile. For Highfield, the vanG model overestimated the pore volume in the  
388 size range 10-30  $\mu\text{m}$  (pF 2.5-2) and underestimated the pore volume at pF 3 and 1. The error  
389 was more pronounced for the more structured grass than bare fallow soil (Fig. 7).





390

391 **Fig. 7.** Overestimation of soil water content (fitted-measured values) for the bare fallow  
 392 (black symbols) and grass (white symbols) treatment at Highfield when fitted to the van  
 393 Genuchten (vanG) model (square symbols) and the double-exponential (Dex) model (circle  
 394 symbols) as a function of matric potential.

395 Introducing systematic errors depending on management (i.e. SOC and tillage) may have  
 396 severe impacts when modelling key soil processes. We recommend using more flexible  
 397 models such as the Dex model to describe the Q/I relation since it was better able to take into  
 398 account the real variation in the distribution of pore sizes across the entire range of I. At the  
 399 same time, we discourage uncritical use of uni-modal models such as the vanG model.

## 400 CONCLUSIONS

401 The bi-modal double-exponential (Dex) model provided a better fit to soil water retention  
 402 data for relatively sandy Danish glacial till top- and subsoils, a silt loam and a sandy loam soil  
 403 than the uni-modal van Genuchten (vanG) model. However, neither of the models worked  
 404 well for highly sorted soils. Topsoil was less well described than subsoil when using the vanG

405 model due to a more pronounced bi-modality of the size distribution of pores caused by  
406 increased SOC content and tillage. Similarly, RMSE of the vanG fit increased with an  
407 increase in SOC when going from the bare fallow to the grass treatment at Highfield. At the  
408 same time, the tillage intensity decreased, but for these long-term treatments, the effect of  
409 SOC seemed to outweigh the lack of tillage. These observations were reflected in a more  
410 pronounced bi-modality of the size distribution of pores for better structured soils.  
411 Consequently, we suggest that uni-modal models are too simplistic for describing  
412 management effects on PSD. Structural void ratio ( $V_2$ ) estimated by the Dex model increased  
413 with SOC content for Highfield and with tillage intensity for Flakkebjerg, whereas the degree  
414 at which the soil is sorted (spread of particle sizes) was the primary driver affecting  $V_2$  for  
415 Danish top- and subsoil samples.

416 **REFERENCES**

- 417 Abdollahi L., Munkholm L.J. (2017) Eleven Years' Effect of Conservation Practices for  
418 Temperate Sandy Loams: II. Soil Pore Characteristics. Soil Science Society of  
419 America Journal 81:392-403. DOI: 10.2136/sssaj2016.07.0221.
- 420 Akaike H. (1973) Information theory and an extension of the maximum likelihood principle,  
421 in: B. N. Petrov and F. Cáski (Eds.), Second International Symposium in Information  
422 Theory, Akadémiai Kiadó, Budapest. pp. 267-281.
- 423 Avery B.W., Catt J.A. (1995) The soils at Rothamsted. Lawes Agricultural Trust:1-44.
- 424 Berisso F.E., Schjønning P., Keller T., Lamandé M., Etana A., de Jonge L.W., Iversen B.V.,  
425 Arvidsson J., Forkman J. (2012) Persistent effects of subsoil compaction on pore size  
426 distribution and gas transport in a loamy soil. Soil and Tillage Research 122:42-51.  
427 DOI: <https://doi.org/10.1016/j.still.2012.02.005>.
- 428 Bronick C.J., Lal R. (2005) Soil structure and management: a review. Geoderma 124:3-22.  
429 DOI: <http://dx.doi.org/10.1016/j.geoderma.2004.03.005>.
- 430 Brooks R.H., Corey A.T. (1964) Hydraulic properties of porous media. Hydrological paper 3.  
431 Civil Engineering Department, Colorado State University, Fort Collins.
- 432 Cornelis W.M., Khlosi M., Hartmann R., Van Meirvenne M., De Vos B. (2005) Comparison  
433 of Unimodal Analytical Expressions for the Soil-Water Retention Curve. Soil Science  
434 Society of America Journal 69:1902-1911. DOI: 10.2136/sssaj2004.0238.
- 435 Cornelis W.M., Ronsyn J., Van Meirvenne M., Hartmann R. (2001) Evaluation of  
436 Pedotransfer Functions for Predicting the Soil Moisture Retention Curve. Soil Science  
437 Society of America Journal 65:638-648. DOI: 10.2136/sssaj2001.653638x.

438 Dane J.H., Hopmans J.W. (2002) Water Retention and Storage, in: J. H. Dane and G. C. Topp  
439 (Eds.), *Methods of Soil Analysis. Part 4 - Physical methods*, Soil Science Society of  
440 America, Inc. Madison, Wisconsin, USA. pp. 671-720.

441 Dexter A.R. (1988) Advances in characterization of soil structure. *Soil and Tillage Research*  
442 11:199-238. DOI: [http://dx.doi.org/10.1016/0167-1987\(88\)90002-5](http://dx.doi.org/10.1016/0167-1987(88)90002-5).

443 Dexter A.R., Czyż E.A., Richard G., Reszkowska A. (2008) A user-friendly water retention  
444 function that takes account of the textural and structural pore spaces in soil. *Geoderma*  
445 143:243-253. DOI: <http://dx.doi.org/10.1016/j.geoderma.2007.11.010>.

446 Eden M., Schjønning P., Moldrup P., De Jonge L.W. (2011) Compaction and rotovation  
447 effects on soil pore characteristics of a loamy sand soil with contrasting organic matter  
448 content. *Soil Use and Management* 27:340-349. DOI: 10.1111/j.1475-  
449 2743.2011.00344.x.

450 Ehlers W., Claupein W. (1994) Approaches toward conservation tillage in Germany, in: M. R.  
451 Carter (Ed.), *Conservation Tillage in Temperate Agroecosystems*, Lewis Publishers,  
452 Boca Raton, Ann Arbor, London, Tokyo. pp. 141-155.

453 Flint A.L., Flint L.E. (2002) Particle density, in: J. H. Dane and G. C. Topp (Eds.), *Methods*  
454 *of Soil Analysis. Part 4 - Physical methods*, Soil Science Society of America, Inc.  
455 Madison, Wisconsin, USA. pp. 229-240.

456 Gee G.W., Or D. (2002) Particle-size analysis, in: J. H. Dane and G. C. Topp (Eds.), *Methods*  
457 *of Soil Analysis. Part 4 - Physical methods*, Soil Science Society of America, Inc.  
458 Madison, Wisconsin, USA. pp. 255-294.

459 Hansen L. (1976) Soil types at the Danish State Experimental Stations (in Danish with  
460 English summary). *Tidsskrift for Planteavl* 80:742-758.

461 Hansen S., Abrahamsen P., T. Petersen C., Styczen M. (2012) Daisy: Model Use, Calibration,  
462 and Validation. Transactions of the ASABE 55:1317. DOI:  
463 <https://doi.org/10.13031/2013.42244>.

464 Jakobsen O.H. (1989) Unsaturated Hydraulic Conductivity for Some Danish soils (in Danish  
465 with English summary). Report No. S2030 From the Danish Institute of Plant and Soil  
466 Science, Copenhagen. [http://web.agrsci.dk/pub/S\\_beretning\\_2030\\_1989.pdf](http://web.agrsci.dk/pub/S_beretning_2030_1989.pdf).  
467 Tidsskrift for Planteavl's Specialserie:1-60.

468 Jensen J.L., Schjønning P., Watts C.W., Christensen B.T., Peltre C., Munkholm L. (2018)  
469 Relating soil C and organic matter fractions to soil structural stability. Geoderma  
470 (revised version submitted).

471 Kutner M.H., Nachtsheim C., Neter J. (2004) Applied Linear Regression Models McGraw-  
472 Hill, New York.

473 Minasny B., McBratney A.B., Bristow K.L. (1999) Comparison of different approaches to the  
474 development of pedotransfer functions for water-retention curves. Geoderma 93:225-  
475 253. DOI: [https://doi.org/10.1016/S0016-7061\(99\)00061-0](https://doi.org/10.1016/S0016-7061(99)00061-0).

476 Mualem Y. (1976) A new model for predicting the hydraulic conductivity of unsaturated  
477 porous media. Water Resources Research 12:513-522. DOI:  
478 10.1029/WR012i003p00513.

479 Mualem Y. (1986) Hydraulic Conductivity of Unsaturated Soils: Prediction and Formulas, in:  
480 A. Klute (Ed.), Methods of Soil Analysis: Part 1—Physical and Mineralogical  
481 Methods, Soil Science Society of America, American Society of Agronomy, Madison,  
482 WI. pp. 799-823.

483 Obour P.B., Jensen J.L., Lamandé M., Watts C.W., Munkholm L. (2018) Soil organic matter  
484 widens the range of water contents for tillage. Soil and Tillage Research 182:57-65.

485 Patil N.G., Singh S.K. (2016) Pedotransfer Functions for Estimating Soil Hydraulic  
486 Properties: A Review. *Pedosphere* 26:417-430. DOI: [https://doi.org/10.1016/S1002-](https://doi.org/10.1016/S1002-0160(15)60054-6)  
487 [0160\(15\)60054-6](https://doi.org/10.1016/S1002-0160(15)60054-6).

488 Poulsen T.G., Moldrup P., Yoshikawa S., Komatsu T. (2006) Bimodal Probability Law Model  
489 for Unified Description of Water Retention, Air and Water Permeability, and Gas  
490 Diffusivity in Variably Saturated Soil. *Vadose Zone Journal* 5:1119-1128. DOI:  
491 10.2136/vzj2005.0146.

492 Pulido-Moncada M., Munkholm L.J., Schjønning P. (2018) Wheel load, repeated wheeling,  
493 and traction effects on subsoil compaction. *Soil and Tillage Research* (submitted).

494 Rabot E., Wiesmeier M., Schlüter S., Vogel H.J. (2018) Soil structure as an indicator of soil  
495 functions: A review. *Geoderma* 314:122-137. DOI:  
496 <https://doi.org/10.1016/j.geoderma.2017.11.009>.

497 Rawls W.J., Pachepsky Y.A., Ritchie J.C., Sobecki T.M., Bloodworth H. (2003) Effect of soil  
498 organic carbon on soil water retention. *Geoderma* 116:61-76. DOI:  
499 [https://doi.org/10.1016/S0016-7061\(03\)00094-6](https://doi.org/10.1016/S0016-7061(03)00094-6).

500 Rogerson P.A. (2001) *Statistical Methods for Geography* SAGE Publications.

501 Rosin P., Rammler E. (1933) Laws governing the fineness of powdered coal. *Journal of the*  
502 *Institute of Fuel* 7:29-36.

503 Ross P.J., Smettem K.R.J. (1993) Describing Soil Hydraulic Properties with Sums of Simple  
504 Functions. *Soil Science Society of America Journal* 57:26-29. DOI:  
505 10.2136/sssaj1993.03615995005700010006x.

506 Scanlon B.R., Andraski B.J., Bilskie J. (2002) Miscellaneous Methods for Measuring Matric  
507 or Water Potential, in: J. H. Dane and G. C. Topp (Eds.), *Methods of Soil Analysis*.

508 Part 4 - Physical methods, Soil Science Society of America, Inc. Madison, Wisconsin,  
509 USA. pp. 643-670.

510 Schjønning P. (1992) Size Distribution of Dispersed and Aggregated Particles and of Soil  
511 Pores in 12 Danish Soils. *Acta Agriculturae Scandinavica, Section B — Soil & Plant*  
512 *Science* 42:26-33. DOI: 10.1080/09064719209410196.

513 Schjønning P., Thomsen I.K. (2013) Shallow tillage effects on soil properties for temperate-  
514 region hard-setting soils. *Soil and Tillage Research* 132:12-20. DOI:  
515 <http://dx.doi.org/10.1016/j.still.2013.04.006>.

516 Šimunek J., Th. van Genuchten M., Šejna M. (2012) HYDRUS: Model Use, Calibration, and  
517 Validation. *Transactions of the ASABE* 55:1263. DOI:  
518 <https://doi.org/10.13031/2013.42239>.

519 van Genuchten M.T. (1980) A Closed-form Equation for Predicting the Hydraulic  
520 Conductivity of Unsaturated Soils. *Soil Science Society of America Journal* 44:892-  
521 898. DOI: 10.2136/sssaj1980.03615995004400050002x.

522 van Genuchten M.T., Leij F.J., Yates S.R. (1991) The RETC code for quantifying the  
523 hydraulic functions of unsaturated soils.

524

525

526 **Figure captions**

527 **Fig. 1.** The root mean squared error (RMSE) value for the Danish top- and subsoil of the  
528 Jakobsen data set using the van Genuchten (vanG) or double-exponential (Dex) model.

529 **Fig. 2. (a)** Measured volumetric water content for the four treatments at Highfield and fits of  
530 the double-exponential (Dex) model as a function of matric potential. The standard error of  
531 the mean are indicated ( $n=4$ ). **(b)** Pore size distribution ( $d\theta/d(pF)$ ) as a function of matric  
532 potential for the four treatments. Eq. 5 was used to obtain the pore size distributions.

533 **Fig. 3.** The root mean squared error (RMSE) value as a function of soil organic carbon (SOC)  
534 for the four treatments at Highfield using the van Genuchten (vanG) model (circle symbols)  
535 and the double-exponential (Dex) model (triangle symbols).

536 **Fig. 4. (a)** Measured volumetric water content for moldboard plowing to 0.20-m depth and  
537 direct drilling and fits of the double-exponential (Dex) model as a function of matric  
538 potential. The standard error of the mean are indicated ( $n=4$ ), except for pF 4.2 which is  
539 predicted based on Eq. 1 in Hansen (1976). **(b)** Pore size distribution ( $d\theta/d(pF)$ ) as a function  
540 of matric potential for the two treatments. Eq. 5 was used to obtain the pore size distributions.

541 **Fig. 5. (a, b)** Measured volumetric water content for Jyndevad and Tystofte top- and subsoils  
542 of the Jakobsen data set and fits of the double-exponential (Dex) model as a function of  
543 matric potential. **(c, d)** Pore size distribution ( $d\theta/d(pF)$ ) as a function of matric potential for  
544 Jyndevad and Tystofte top- and subsoils. Eq. 5 was used to obtain the pore size distributions.

545 **Fig. 6.** Pore size distribution ( $d\theta/d(pF)$ ) as a function of matric potential for **(a)** the bare  
546 fallow and **(b)** the grass treatment at Highfield either obtained by Eq. 5 for the double-  
547 exponential (Dex) model (solid line) or Eq. 3 for the van Genuchten (vanG) model (dashed  
548 line).



549 **Fig. 7.** Overestimation of soil water content (fitted-measured values) for the bare fallow  
550 (black symbols) and grass (white symbols) treatment at Highfield when fitted to the van  
551 Genuchten (vanG) model (square symbols) and the double-exponential (Dex) model (circle  
552 symbols) as a function of matric potential.