

Manuscript Details

Manuscript number	HE_2016_1199
Title	Signal trends of microbial fuel cells fed with different food-industry residues
Article type	Full length article

Abstract

A microbial fuel cell (MFC) is an anaerobic bioreactor where soluble metabolites liberated by hydrolysis and fermentation of macromolecules are simultaneously available for anode respiring bacteria (ARB). ARB can be influenced by chemical imbalances in the liquid phase of the bioreactor. The objective of the work was to explore the trend of electric signals generated by MFCs, in relation to anaerobic biodegradation of four different solid food-industry residual substrates. Four sets of membraneless single-chamber MFCs were operated in batch mode, with solid waste substrates characterized by a different base component: i) mixed kitchen waste (fibers), ii) whey from dairy industries (sugar), iii) fisheries residues previously processed to recover oils (proteins), iv) pulp waste from citrus juice production (acidic). All the tested MFCs were able to produce an electric output with different trends, depending on the principal component of the solid substrate. MFC potential varied as function of the COD and the feeding cycle, as well as of the substrate. The pH variability during the fermentative process significantly affected the electric output. Citrus (acidic) pulp fed MFCs started to operate only when the pH raised up 6.5. MFCs fed with mix food wastes had a relatively stable electric signal; fish based waste caused spiking in the MFC signal and an averaging in the COD degradation trend. This phenomenon was attributed to a pH instability induced by proteins degradation forming ammonia. The fermentation process was strongly predominant with respect the electrochemical process in MFCs and the coulombic efficiency (CE) was low, ranging between 2 to 10 %. This result call for a deeper exploration of harvesting power from solid wastes and pointed also to the possibility of using a MFC to monitor important parameters of fermentation processes in biotech production plants.

Keywords	microbial fuel cells, solid food waste, citrus pulp, fish wastes, diary whey
Manuscript category	Fuel Cells & Applications
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Milan, 11 August 2016

Dear Scientific Committee of Journal of Hydrogen Energy,

It is my pleasure to submit to Your attention the revised version of the manuscript that was previously entitled: " Different solid-phase fuels in microbial fuel cells" and now entitled "***Signal trends of microbial fuel cells fed with different food-industry residues***" which I am submitting for exclusive consideration of publication as an original research article for the Special Issue of the conference EFC2016, held in Napoli (Italy) in December 2015.

We acknowledged all the suggestions and the requests of the Reviewer. We thank them, as they give us the opportunity of strongly improve the manuscript, indeed. At the end of this letter there is the punctual answer to each request and, in following, the

We also confirm, what declared in the previous cover letter:

The paper present, for the first time, the result of an experimentation with sets of membraneless single-chamber Microbial Fuel Cells (MFCs) operated with solid waste dried matrices, characterized by different principal components. All the tested MFCs were able to produce an electric output. Power varied as function of the substrate principal content (protein, sugar, acids, mix), the COD and the feeding cycle. The pH variability during the fermentative process significantly affected the electric output.

The result pointed to the possibility of using a MFC to monitor fermentation processes, such as the biogas production.

The publication is approved by all the authors and by the responsible authorities where the work was carried out.

I can confirm with the other authors that each of us have made substantial contributions to the work and that there are no conflict of interest, including any financial, personal or other relationships with other people or organizations within three years of beginning the submitted work that could inappropriately influence, or be perceived to influence, our work.

Hoping in a positive response, and thanking for Your consideration,
I look forward to the challenge of publishing with the journal.

Truly
Pierangela Cristiani



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ANSWERS TO REVIEWERS

We thank the reviewers for the comments that help us to strongly improve the manuscript and we apologies ourselves for the refuses and mistakes in the English of the original version.

We answered to all the requests of the Reviewers and we provided to strongly improve the manuscript, enlarging substantially the discussion, and strongly revising the English syntaxes.

We included the final version of the revised manuscript and a version of the revised manuscript with underlined in red all changed and insert words and phrases and in ~~red~~ all deleted worlds.

Comments from the editors and reviewers:

-Reviewer 1

- The manuscript describes MFC applications for solid-phase fuels. The manuscript must solve the following issues before the publications.

1. First of all, the authors make sure if the work is for power generation or just treatment of solid-phase matters. If the former one is the target for this work, the manuscript must include the detailed relationship between COD and coulombic efficiencies (CE). Also, thorough analysis of the CE data must be made since the CE determines the amount of COD that can be captured as electrical current by the end of each fed-batch cycle.

We thank the reviewer to give us the opportunity of better explain the objective of the work. Our objective was to explore and analyze microbial fuel cells output in function of different organics components of wastes that are usually digested in anaerobic bioreactors. The results indicated the possibility of exploiting the signal achieved, at least, for monitor the biodegradation process. This is a novelty and the most interesting aspect in our opinion.

Under this point of view, the low CE is not an issue, because the lowest is CE, the best for monitoring purpose. In fact, in this condition the MFCs don't influence the biodegradation process much, but it is influenced by it.

The original text was not sufficiently clear about this point, because we expressed this concept just at the end of the discussion and in the conclusion. Now we deeply modified the text, strongly improving the discussion part, and better address the objective in the Abstract and Conclusion.

We also improve the English syntaxes and the lay-out of the figure in order to make more visible all the details.

The abstract was modified adding the following sentence at the beginning:

“A microbial fuel cell (MFC) is an anaerobic bioreactor where soluble metabolites liberated by hydrolysis and fermentation of macromolecules are simultaneously available for anode respiring bacteria (ARB). ARB can be influenced by chemical imbalances in the liquid phase of the bioreactor. The objective of the work was to explore the trend of electric signals generated by MFCs, in relation to anaerobic biodegradation of four different solid food-industry residual substrates”

The introduction was also strongly improved with the following sentences (l. 59-82):

“Anaerobic bioconversion processes such as anaerobic digestion and dark fermentation rely on sequential microbial hydrolytic and fermentative processes that solubilize different substances in the liquid phase. Macromolecules are degraded to soluble molecules that become available to

secondary fermentations or anaerobic respirations, by other microbes forming part of complex consortia [3]. In many cases, the liberation of soluble metabolites might significantly change the chemical equilibria of the liquid and gaseous phases where microbes live [4]. This is often the cause of inhibition of more sensible microbial species and possible imbalances of the overall biodegradation process [5].

In anaerobic biodegradation, secondary metabolisms (e.g. acetogenesis, acetoclastic methanogenesis, hydrogenophilic methanogenesis, denitrification, sulphate reduction, etc.) rely on the availability of short-chain organic molecules, inorganic ions and soluble gasses, such as volatile fatty acids, di-hydrogen, hydrogen sulphide, ammonium, nitrate, carbon dioxide, bicarbonates, etc [6,7].

Similarly, in microbial electrochemical systems like microbial fuel cells (MFC), electroactive microbial species like anode respiring bacteria (ARB) rely on the same substances to transfer electrons and produce current [8]. Macromolecules, in regular MFCs, should be at least hydrolyzed and pre-fermented by fermentative microbial species before being available to ARB, which preferentially oxidize low-carbon carboxylates, as indicated in various literature contributions [8–11].

For this reason, MFCs can be used as a mirror-process for secondary biodegradation metabolisms. In anaerobic environments, the electrical signal produced by ARB activity can give real-time hints on the trend of the ongoing biodegradation mechanisms and biochemical conditions, such as availability of soluble low-carbon organics, availability of mineral nutrients, favorable chemical equilibria in the liquid medium like pH, electrical conductivity, etc. A widely recognized issue is the competition of methanogenic populations and ARB for the same organics [12,13]. This corroborates the assumption that ARB activity, measured as voltage generation, could be used as monitor of the interactions between fermentative and methanogenic microbial populations in anaerobic biodegradation environments.”

The discussion was improved in the Result and Discussion chapter, adding a sub-paragraphs for each different substrates used, in this way:

3.1 Electrical signal trend in CW-fed MFCs

where, other than a better revision of the language and English, the follow sentence has been add at the end:

This effect point to the possibility of monitoring on-line possible accumulations of that soluble metabolites inhibiting biodegradation processes. This would be particularly useful in high-solids anaerobic digestion plants [4], or other biodegradation processes at high organic loading rates [20].

3.2 Electrical signal trend in KW-fed MFCs

Where KW is “Kitchen Waste” that substitutes the name of waste previously named “mix from municipal wastes”, because more simple and adherent to the substrate used, as the preparation of the mix were made, in this case, starting from the single components specified in the Table 1, and not from a real wastes).

In the paragraph 3.2, this part has been included (l. 220-235):

“Interestingly, the maximum potential decreased cycle by cycle, in parallel with the peak of COD at the beginning of each cycle. The peak value of COD was over 3000 mg_{O2} L⁻¹ in the first cycle, and decreased progressively to less than 2000 mg_{O2} L⁻¹ in cycle 5.

This condition might be due to the accumulation of mineralized substances in the liquid phase, such as ammonium ions, and other nutrients, creating an inhibitory environment for microbial population and for ARB. The MFC signal followed the general trend of the primary phases of biodegradation (hydrolysis and fermentation). Progressively less organic matter was hydrolyzed from the solid phase (visible as decreasing peaks of COD). Additionally, each cycle lasted for longer time as compared to the first one. Biodegradation rate was visibly decreasing. All these effects were reflected in progressively decreasing peaks of the electrical signals.

The MFCs potential trends were, again, a mirror of an increase of limiting (or toxic) conditions for the overall microbial population in the bioreactor. In this case the electrical signals were indicating a progressive increase of inhibition conditions. This application of MFCs in bioreactors would be useful for monitoring the accumulation of potentially toxic metabolites of biodegradation (ammonia, H₂S, Na⁺, etc.) on long-term operations of bioreactors [5, 21].

3.3 Electrical signal trend in FW-fed MFCs

Here, the following sentence was mainly add (lines 261-267):

The peak values of COD were very similar at each cycle. Differently from CW and KW, peak of COD didn't correspond to peaks of cell potential, especially at cycles 2 and 3. This is likely to be due to the recalcitrants hydrolyzed from complex proteins of FW (Table 1). Contrarily, cell potentials lasted for longer time at relatively high values, as compared to COD values.

In this case, MFCs can be thought as monitors of the presence of long-term biodegradable fractions of the organic matter.

3.4 Electrical signal trend in CP-fed MFCs

Here, the following sentences was added, Line 282-289:

“In the first two cycles, acidic conditions (pH 4 – 5) were established in the bioreactor and the pH raised over 6.5 only after 40 days. Alkalinity was evidently insufficient to buffer the acidity of CP (rich in citric acid and other organic acids). COD decreased from 3400 mg_{O2}/l to 200 mg_{O2} L⁻¹ along over 45 days of operation of cycle 1 and nearly 50 days in cycle 2.

At cycle 3, a buffer medium (potassium bicarbonate, 5 g L⁻¹) was added to the bioreactor to equilibrate the pH, which remained stable in the range 7 – 7.8 along cycle 3. COD peaked to more than double (over 8000 mg_{O2} L⁻¹), due to more efficient microbial hydrolysis and fermentation of the solids accumulated in the bioreactor.”

And at the end (l. 320-324):

When stable pH was guaranteed by equilibrating CP acidity, cell potential trends followed exactly the trend of COD consumption in the liquid phase. These results highlighted another aspect

of the chemical equilibria in anaerobic biodegradation environments, which can be efficiently correlated to the trend of the electrical signal produced by ARB, living in the same environment. pH-related inhibition of microbial activity reflects very promptly in drops of MFCs cell potentials.

The following lines was deleted, because discussing about a phenomenon which data were not reported:

~~Additional CP-fed SMFCs were operated with 0.2 g COD_{substrate}/g VS_{inoculum}. An increase in potential was observed after 2 days of operation as already obtained with the other substrates (**results not shown here**). As already known from biogas and anaerobic digestion field [11], the results from CP-SMFCs highlighted that the substrate/inoculum ratio is another key limiting factor when organic solid wastes are used to fed microbial metabolism.~~

Also the discussion in chapter 3.5 Electrode polarization curves was improved at the end, from line 341:

This indicates, first of all, that the cathode was predominantly characterized by microbially-catalyzed reduction reactions. As reported in previous works for single-chamber MFCs, bio-cathodic instead of abiotic mechanisms drive oxygen reduction reactions [25]. This is an important aspect to consider, in the case that cell potential has to be used as indicator of microbial consortia activity in a bioreactor. To maximize the MFC system response to inhibitory effects, due to chemical imbalances in the liquid medium of biodegradation environments, single chamber MFCs might be the ideal solution. A double chamber architecture with abiotic cathode would be less sensible to inhibitory conditions for microbes.

FW biodegradation started showing instability at cycle 3. Error bars reported for SCOD measured in two MFCs indicate a condition of partial inhibition of the system. In biodegradation of protein-rich organic materials (see Table 1 for FW composition), it might typically be related to accumulations of ammonia, hydrogen sulfide or other toxic metabolites [26].

The considerable decrease of polarization currents from both bioanodes and biocathodes, already registered during cycle 2 (days 62 – 64, Figure 5), can be considered as an early-warning for inhibiting conditions for the whole bioreactor environment. Future experiments should more deeply focus on this aspect. Polarization of electrodes, as mirror of both anodic and cathodic microbial communities, could be studied as early-warning sensors for inhibiting conditions in anaerobic biodegradation environments.

3.6. Just monitoring or influencing the biodegradation process of COD removal?

This sub-chapter concludes the discussion (l. 372-380):

This indicates that biodegradation were negligibly influenced by the MFC process. In future applications of MFCs as sensors for monitoring biodegradation process, electrode surface/reactor volume ratio might be even scaled down, to monitor specific environments.

On the contrary, enlarging the electrodes and improving their surface/volume ratio using a different geometry, will allow to reach and enhance the performance of the MFC process in degrading solid phase organics with respect the current literature [14].

Finally, the optimization of the electrode surface finishing [29] and the use of different materials such as stainless steel [28] would address other possible needs for an useful application in both cases: monitoring or influencing the biodegradation process.

Finally, the Table 2 was simplified as below, reporting just three cycles, as the others are not performed for all the substrates:

Table 2 Coulombic efficiency and COD removal during fed-batch MFC operation.

	Sodium acetate (control)		Cheese whey		Kitchen waste		Fish waste		Citrus pulp	
	CE (%)	sCOD removal (%)	CE (%)	sCOD removal (%)	CE (%)	sCOD removal (%)	CE (%)	sCOD removal (%)	CE (%)	sCOD removal (%)
Cycle 1	22.4	98.1	0.76	95.1	3.87	84.37	7.77	88.69	3.13	93.52
Cycle 2	21.3	95.9	2.02	96.9	9.91	64.25	4.37	85.94	2.74	89.70
Cycle 3	23.2	97.4	1.87	95.8	9.40	61.40	3.53	89.20	-	-

2. Although the authors claim that the low CE was due to other non-optimized conditions in MFCs, the CE the authors provide in table 2 is significantly low compared to other studies and does not support that the COD removal is due to electron transfer reactions. Rather, the COD removal might be other anaerobic or aerobic processes. Then the treatment of the solid-phase matters might be made through simple microbial reactions not in the MFCs. The authors clearly mention about this.

Please see the answer above, valid also for this point.

3. The reviewer wonders if we can see this work as an actual solid-phase matters' treatment method. This is because all the matters were completely grinded and mixed with liquid phase solution. The actual raw wastewater does not look like this.

The referee is right and we thank you for underlying this important issue.

Although some other authors [14] already referred to “solid MFCs in similar cases, the definition of ‘solid’ MFCs was consequently removed from the text, and used just MFCs.

Grinded materials in suspension in a bulk sludge at around 20 g per liter of volatile solids is a typical concentration for optimized anaerobic digestion tests. In this test we worked at standard conditions for anaerobic biodegradation. To clarify this, the M&M section was deeply improved.

4. There are so many typos/errors in the manuscript and non-technical writing phrases/words. The authors need to thoroughly check the manuscript again.

We apologies us again for the mistakes, due to a quite hurry in submitting the manuscript within the deadline. We improved the manuscript with a deep revision of all the parts.

5. The figure 1, 2, & 3 must be more readable and clear so that the readers can understand. The labels are needed. Or some enlarged portion will be needed.

The figure 1, 2, 3 and 4 were completely revised and improved, to make them readable. Figure 6 (old version) should clarify a single cycle, now better visible, so it was removed in the new version of manuscript. The trends are now readable in the graphs of figures 1,2,3,4, split into two parts of 60 days each. Cycles were also indicated.

The new Figures are:

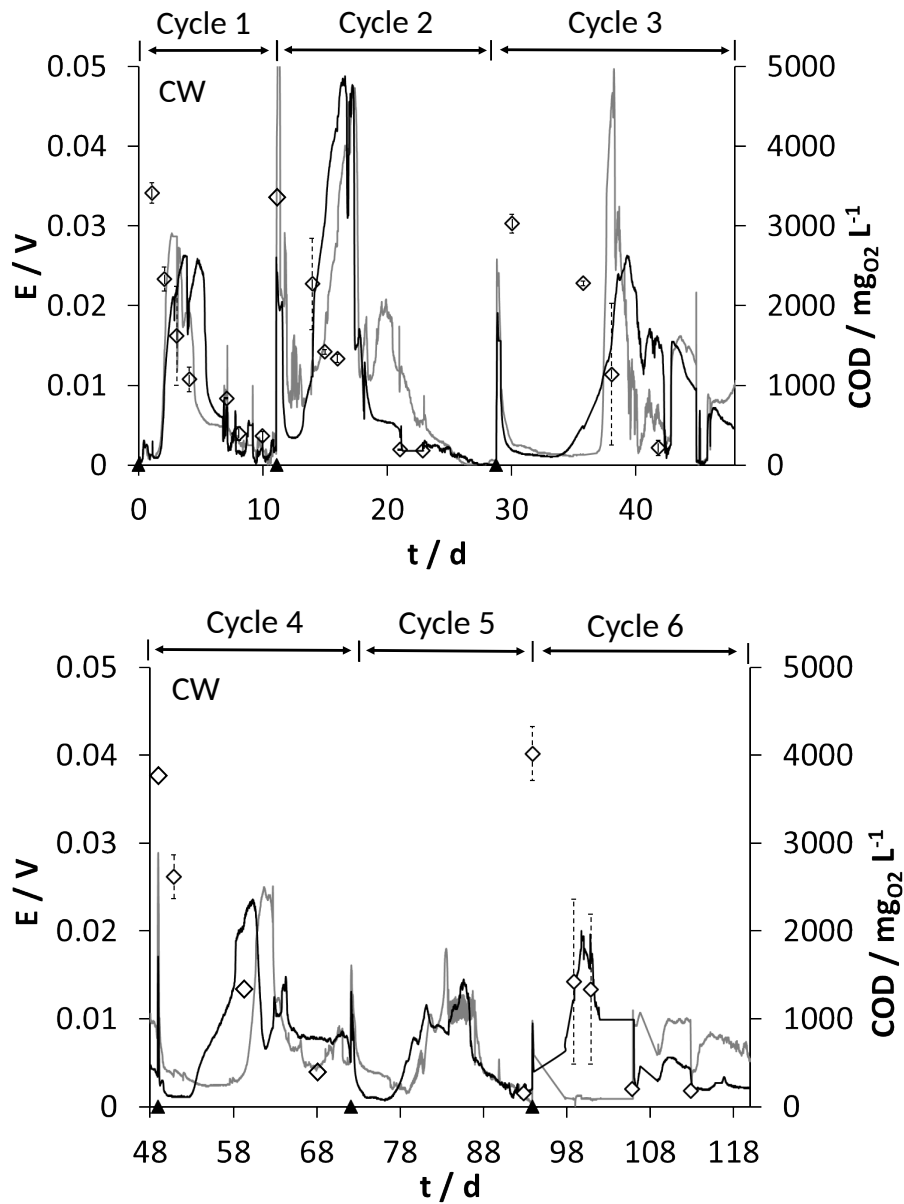


Fig. 1 Graphics of potential trends of two CW-fed MFCs (solid lines, left y-axis) and COD evolution (\diamond , right y-axis) over 125 days of operation **and six batch cycles**. Standard deviation of COD values is reported with vertical bars. Triangles on the x-axis indicate the day of feed addition.

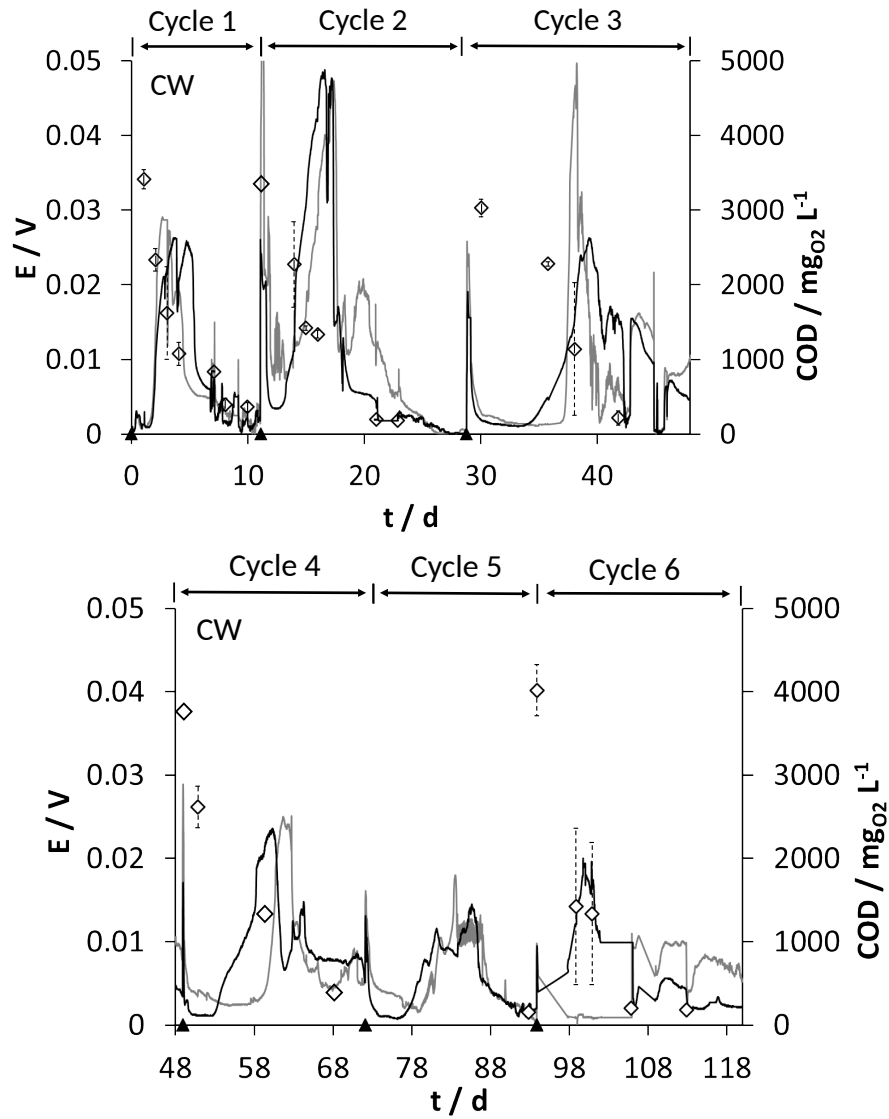


Fig. 1 Graphics of potential trends of two KW-fed MFCs (solid lines, left y-axis) and COD evolution (\diamond , right y-axis) over 125 days of operation **and six batch cycles**. Standard deviation of COD values is reported with vertical bars. Triangles on the x-axis indicate the day of feed addition.

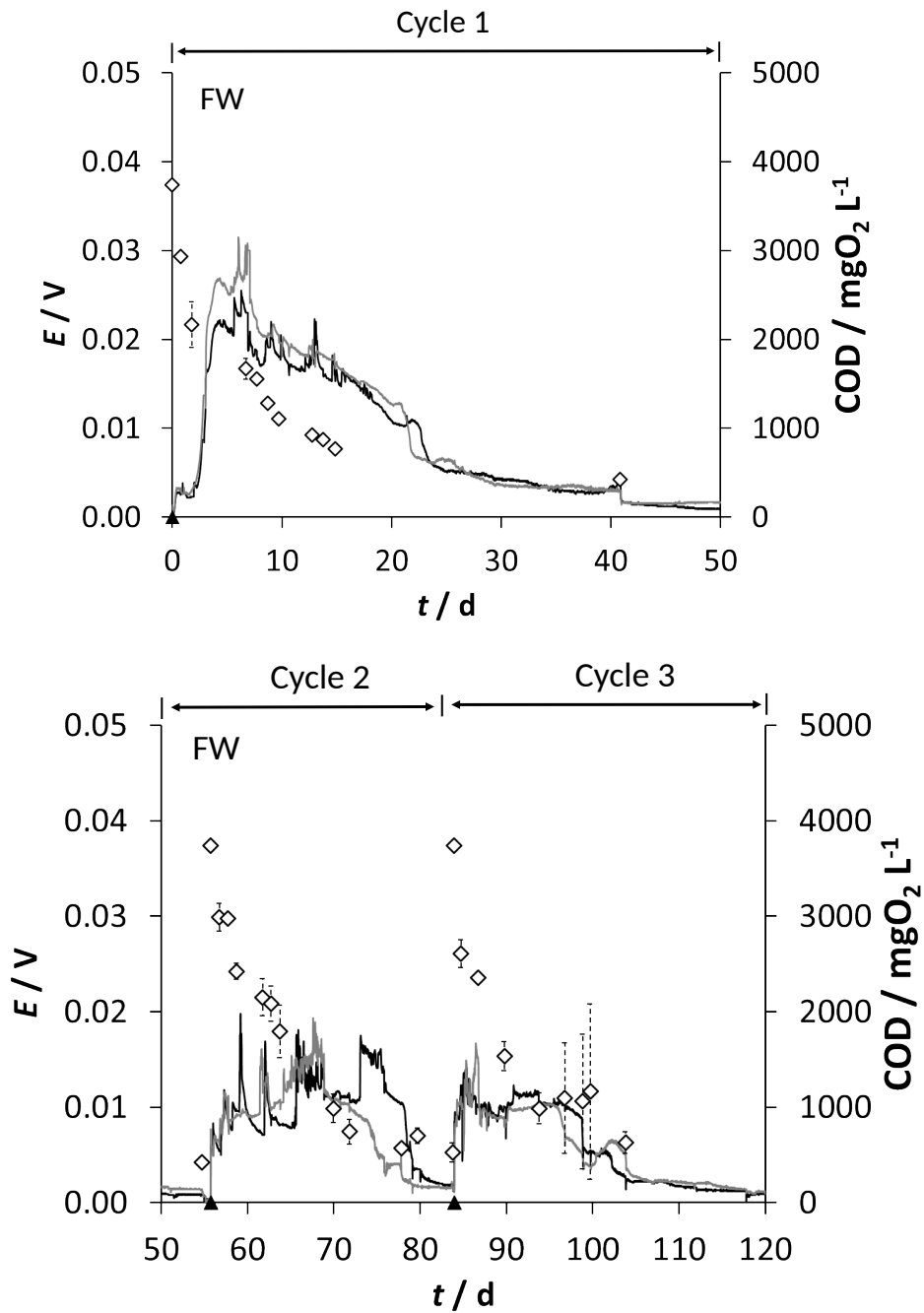


Fig. 2 Graphics of potential trends of two FW-fed MFCs (solid lines, left y-axis) and COD evolution (\diamond , right y-axis) over 125 days of operation and three batch cycles. Standard deviation of COD values is reported with vertical bars. Triangles on the x-axis indicate the day of feed addition.

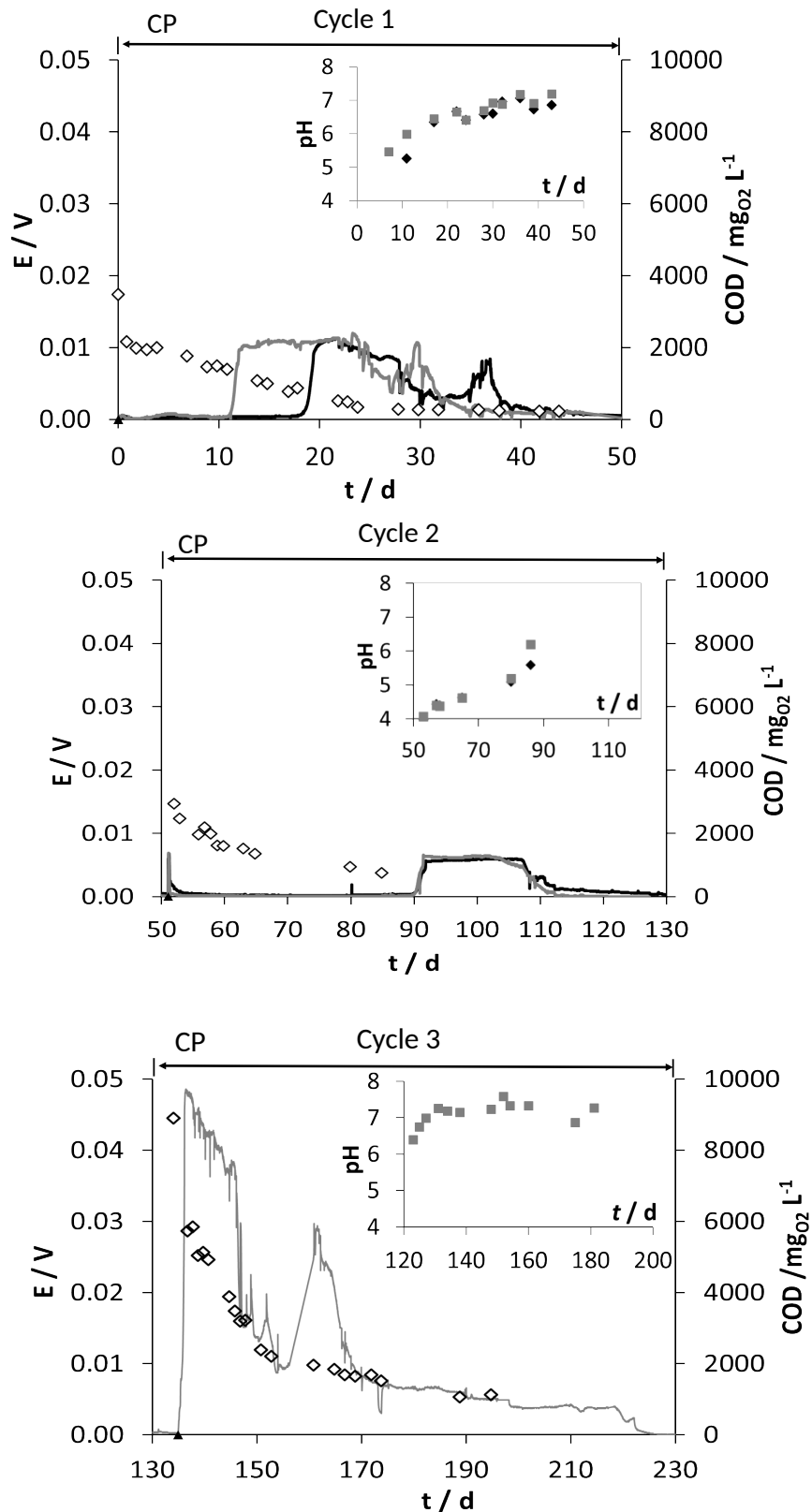


Fig. 4 Graphics of potential trends of two CP-fed MFCs (solid lines, left y-axis) and COD evolution (\diamond , right y-axis) over 230 days of operation and three batch cycles. Standard deviation of COD values is reported with vertical bars. Insets report the trend of pH over time.

-Reviewer

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In this manuscript, the authors investigated the power output of solid-phase microbial fuel cells with four solid organic substrates. The coulombic efficiency and COD removal were also obtained. In

my opinion, major revision is needed for publication in this journal based on the following comments:

(1) The title 'Different solid-phase fuels in microbial fuel cells' means the key point is the fuel, obviously, it can not accurately summarize the content of this manuscript.

The referee is completely right. We deeply modified the title and the abstract, accordingly. The results and discussion chapter was also revised, according to the main meaning of the work that was better underlined.

Please see the previous answers and the revised version with the track-changes underlined in red.

(2) In Fig. 1, there is no sCOD value in the 4th and 5th cycle, the same for the 4th cycle in Fig. 2. Why the test is discontinuous?

Some cycles were not monitored in deep. However, they are a minor part and the overall trend of the bioreactors are clear with the available data. We improved by analysing at least the beginning and the end of the cycle, where there were missing samples along the cycle.

(3) In Fig. 5, the insert figure can not show the pH value after day 90, the authors should present enough pH results. In addition, there is delay in producing power in the 1st and 2nd feeding cycle, however, why no delay exists in the 3rd cycle?

Thank you for the note about pH. The request was approached. Please see the revised figure 4. In addition, the range of pH in all other experiments was reported along the text. See lines 164, 183 and 210.

The other point of the referee was deeply explained in the discussion, in the section *3.4 Electrical signal trend in CP-fed MFCs*

(4) There are two 'Fig. 6' in the manuscript. In the first 'Fig. 6', no legend description for a, b, c and d. And there is no discussion in the text for the second 'Fig. 6'.

This was deeply improved. The "first figure 6" was removed, as it is included in Figures 1,2,3,4 with an improved view of the trends of Cell potentials.

Please see the new versions of all figures, also according to Referee 1.

(5) Some words in the manuscript are not the common expressions in microbial fuel cells, such as 'power intensity' and 'potential output'. Generally speaking, the cell performance is evaluated by 'power density'. In Fig.1-3 and Fig.5, I guess the left y-axis is the voltage of the fuel cell, not the potential output.

All these mistakes were due to hurry in the deadline for submission. We deeply improved the text and the terminology throughout the manuscript.

- (6) The manuscript needs careful editing paying attention to 'space' and spelling such as:
in lines 125, 126, 130 and 144 'SMFC s',
in line 218 'SMFCat',
in line 249 'SMFCoperation',
in line 257 'byporoducts'.

All these mistakes were due to hurry in the deadline for submission. We deeply improved the text and the terminology throughout the manuscript.

Here THE FULL revised MANUSCRIPT

Different solid-phase fuels in microbial fuel cells

Signal trends of microbial fuel cells fed with different food-industry residues

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ABSTRACT

A microbial fuel cell (MFC) is an anaerobic bioreactor where soluble metabolites liberated by hydrolysis and fermentation of macromolecules are simultaneously available for anode respiring bacteria (ARB). ARB can be influenced by chemical imbalances in the liquid phase of the bioreactor. The objective of the work was to explore the trend of electric signals generated by MFCs, in relation to anaerobic biodegradation of four different solid food-industry residual substrates. ~~Solid-phase microbial fuel cells (SMFCs) treat solid organic waste mixed in liquid phase as fuel to harvest power.~~ Four sets of membraneless single-chamber SMFCs were operated in batch mode, with solid waste substrates characterized by a different base component: i) mixed kitchen waste (fibers), ii) whey from dairy industries (sugar), iii) fisheries residues previously processed to recover oils (proteins), iv) pulp waste from citrus juice production (acidic).

All the tested SMFCs were able to produce an electric output with different trends, depending on the principal component of the solid substrate. ~~Power-intensity MFC potential~~ varied as function of the COD and the feeding cycle, as well as of the substrate. ~~The space occupied by the anode was less than 0.1 of the anode-chamber volume, consequently.~~

The pH variability during the fermentative process significantly affected the electric output. Citrus (acidic) pulp fed SMFCs started to operate only when the pH raised up 6.5. SMFCs fed with mix food wastes had a relatively stable electric signal; fish based waste caused spiking in the SMFC signal and an averaging in the COD degradation trend. This phenomenon was attributed to a pH instability induced by proteins degradation forming ammonia.

The fermentation process was strongly predominant with respect the electrochemical process in MFCs and the coulombic efficiency (CE) was low, ranging between 2 to 10 %. This result call for a deeper exploration of harvesting power from solid wastes and pointed also to the possibility of using a SMFC to monitor important parameters of fermentation processes in biotech production plants.

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Highlights

- Four different **component** of solid organic **wastes** were investigated ~~as fuel~~ in membraneless MFCs
- **Cell potential trends varied in function of different waste components in the bioreactors**
- ~~All the SMFCs fed with different fuels were able to produce an electric output~~
- **Cell potential trends** varied as function of the substrate, the COD and the feeding cycle
- The pH in the anodic chamber significantly affected the electric output
- **Results call for exploration of MFCs as sensors for fermentation/biodegradation processes**

Index Terms – microbial fuel cells, solid food waste, citrus pulp, fish wastes, dairy whey

51 1. Introduction

52 The new paradigm of circular economy claims new technological approach for energy and resource
53 recovery. Agro-food industry produces massive amounts of organic materials as secondary streams and
54 waste [1]. Microbes naturally evolve enzymes and pathways that can convert solid biomass-derived carbon
55 sources into valuable fuels and products, such as biomethane, biohydrogen, biodegradable polymers,
56 carboxylates [2]. Biological conversions **might** play a fundamental role in waste refinery chains and
57 especially of agricultural and food-industry residues. In this context, microbial electrochemical technologies
58 (METs) offer potential innovative approaches **in wastes treatment**.

59 **Anaerobic bioconversion processes such as anaerobic digestion and dark fermentation rely on sequential**
60 **microbial hydrolytic and fermentative processes that solubilize different substances in the liquid phase.**
61 **Macromolecules are degraded to soluble molecules that become available to secondary fermentations or**
62 **anaerobic respirations, by other microbes forming part of complex consortia [3]. In many cases, the**
63 **liberation of soluble metabolites might significantly change the chemical equilibria of the liquid and gaseous**
64 **phases where microbes live [4]. This is often the cause of inhibition of more sensible microbial species and**
65 **possible imbalances of the overall biodegradation process [5].**

66 **In anaerobic biodegradation, secondary metabolisms (e.g. acetogenesis, acetoclastic methanogenesis,**
67 **hydrogenophilic methanogenesis, denitrification, sulphate reduction, etc.) rely on the availability of short-**
68 **chain organic molecules, inorganic ions and soluble gasses, such as volatile fatty acids, di-hydrogen,**
69 **hydrogen sulphide, ammonium, nitrate, carbon dioxide, bicarbonates, etc [6,7].**

70 **Similarly, in microbial electrochemical systems like microbial fuel cells (MFC), electroactive microbial**
71 **species like anode respiring bacteria (ARB) rely on the same substances to transfer electrons and produce**
72 **current [8]. Macromolecules, in regular MFCs, should be at least hydrolyzed and pre-fermented by**
73 **fermentative microbial species before being available to ARB, which preferentially oxidize low-carbon**
74 **carboxylates, as indicated in various literature contributions [8–11].**

75 For this reason, MFCs can be used as a mirror-process for secondary biodegradation metabolisms. In
76 anaerobic environments, the electrical signal produced by ARB activity can give real-time hints on the trend
77 of the ongoing biodegradation mechanisms and biochemical conditions, such as availability of soluble low-
78 carbon organics, availability of mineral nutrients, favorable chemical equilibria in the liquid medium like pH,
79 electrical conductivity, etc. A widely recognized issue is the competition of methanogenic populations and
80 ARB for the same organics [12,13]. This corroborates the assumption that ARB activity, measured as voltage
81 generation, could be used as monitor of the interactions between fermentative and methanogenic microbial
82 populations in anaerobic biodegradation environments.

83 ~~SMFCs fed with solid organic waste are generally indicated as solid-phase SMFCs (SMFCs~~
84 Electricity harvesting ~~with complex biomass in solid-phase MFCs~~ was recently achieved by Mohan et al.
85 [14], who fed open-air cathode, single-chamber SMFCs with different types of canteen-based food waste.
86 The best performing configuration, where a proton exchange membrane separated the cathode from the
87 anodic chamber, achieved a peak power density of 170 mW m⁻² with open circuit potential (OCP) of 463
88 mV. Similar results with similar substrates were obtained by Goud et al [15], who tested increasing organic
89 loading rates (OLR) in bio-electrochemical reactors fed continuously. ~~Above~~ OLR of 1.39 kg COD/m³-day
90 both power density and OCP started decreasing, due to inhibiting concentrations of volatile fatty acids (>800
91 mg/L) and acidic pH conditions (pH=6).

92 Pretreatment of wastes from agro-food is achieved in several ways (e.g. to extract essential oils and proteins
93 from specific wastes, such as citrus pulp, residual fish) and MET can be though as downstream processing
94 for energy harvesting in Microbial Fuel Cells [14] or for further bio-processing (e.g. electrofermentation [12]
95), ~~however, very little is reported in literature about biodegradation pathways of complex organic matrices in~~
96 ~~the bulk medium. In particular, the relationships between the electric signal produced by ARB and the~~
97 ~~biodegradation process as-a-whole (hydrolysis of macromolecules, fermentative metabolisms, etc.).~~

98 Here, we ~~studied the electrical signals produced by MFCs during anaerobic biodegradation~~ of four different
99 types of agro-industrial residual materials of interest in Mediterranean agro-food sectors: citrus pulp, fishery
100 waste, cheese whey and ~~organic fraction of municipal solid kitchen~~ waste. ~~Voltage trends were monitored~~ on
101 long-term, ~~operation of SMFCs was monitored~~, over 100 days, with successive batch cycles, to evaluate the
102 response of the electrochemical system to the anaerobic biodegradation of the solid matrices.

103

104 **2. Materials and methods**

105 **2.1 SMFC configuration and setup**

106 Four sets of membraneless single-chamber SMFCs were operated in duplicate and in parallel over more than
107 100 days. The total volume of ~~each MFC~~ was 125 ml and the design ~~was previously~~ reported [16]. Anodes
108 were made of 3×5 cm rolled carbon cloth sheet (Saati C1, Legnano, Italy), electrically connected to a copper
109 wire. Three layers of non-conductive high-viscosity epoxy resin (Mapei Epojet) were applied to ensure
110 insulation at the connection between copper and carbon cloth. Cathodes were made of 5×5cm carbon cloth
111 sheets modified by the addition of a Gas Diffusion Layer (GDL) on the air side. The GDL composition has

112 been described in [17] and the PTFE content is 80%_{w/w} with respect to carbon powder. The geometric
 113 cathodic surface area exposed to the solution was 3 cm². Anode and cathode were then connected through an
 114 external circuit with a resistance of 100 Ω.

115 All SMFCs were operated at mesophilic temperature of 35±1 °C in batch mode without pH adjustment.
 116 SMFCs were inoculated with 90 ml of anaerobic mesophilic sludge obtained from a municipal wastewater
 117 treatment plant (Cremona, Italy). The volatile solids (VS) content in the sludge was 15 g_{VS}kg⁻¹. This
 118 concentration is typically used in standard batch-like anaerobic digestion tests [18]. The sludge was not
 119 subjected to any pretreatment. A concentrated solution of nutrients was added at the beginning of the
 120 experimentation. The stock solution of nutrients contained (in g/L): KH₂PO₄ (0.27), Na₂HPO₄·12H₂O (1.12),
 121 NH₄Cl (0.53), CaCl₂·2H₂O (0.075), MgCl₂·6H₂O (0.10), FeCl₂·4H₂O (0.02). Analytical grade reagents and
 122 double distilled water were used.

123 Two SMFCs were fed with each organic substrate in the form of dried powder (1 mm particle size): i) cheese
 124 whey powder (CW); ii) ~~dried organic fraction municipal solid kitchen~~ waste (KW); iii) ~~dried~~ fish waste
 125 (FW); iv) ~~dried~~ citrus pulp (CP). The macromolecular composition of the four ~~solid-waste~~ substrates was
 126 reported in Table 1. CW was a commercial by-product from dairy industries (Cremona, Italy) used as animal
 127 feed. ~~OFMSWKW~~ was a mixture of animal and vegetal food waste prepared in lab according to the
 128 following recipe: 30 g egg shells; 30 g dried bread, 50 g corn flour, 100g grated cheese, 75 g cracker, 10 g
 129 coffee grounds, 130g apple peel, 300g green salad, 145 g orange peel, 85 g zucchini peel, 68 g banana peel,
 130 56 g carrots, 30 pumpkin skin, 20 g kiwi peel, 30 g fennels, 16 g potato peel. Food mixture was grinded with
 131 a kitchen blender, homogenized and finally dried at 105 °C. FW was obtained from fish after an enzymatic
 132 pretreatment to remove oils (no alcohols used). CP was obtained from citrus juice production plant (Catania,
 133 Italy).

134

135 **Table 1 Main macromolecular constituents of the four solid wastes**

Substrate	Carbohydrates (% of DM)	Fibers (% of DM)	Fats (% of DM)	Proteins (% of DM)	Ashes (% of DM)
CW	70 (lactose)	-	-	12	8.5
KW	53.4	19.2	9.6	14.3	3.5
FW	0.3	-	3.8	51.2	20.1
CP	8.5	43.1 (33.2 cellulose, 9.9 hemicellulose)	3.1	26.9 (19.7 pectin)	18.4

136

137 The amounts of inoculum and organic substrates introduced in each SMFCs were determined on the basis of
 138 preliminary analytics determination (volatile solids and total solids). The organic substrate to inoculum ratio
 139 was 0.35 g sCOD_{substrate}/g VS_{inoculum}. A new dose of feed was added when negligible potential values were
 140 obtained and soluble Chemical Oxygen Demand (COD) fell down to a constant value.

141

142 2.2 Tests

143 2.2.1 Data acquisition, electrochemical experiments and calculations

144 The potential difference across the 100 Ω resistance (R) was acquired every 10 minutes, **via** a multichannel
145 Data Logger (Graphtech midi Logger GL820). The generated current (I) was calculated by the equation $I =$
146 V/R , where I is the current flowing through the external resistance. The total charge flowed into the electrical
147 circuit at the end of each batch cycle was calculated by integrating the current over time. Coulombic
148 efficiency (CE) was then evaluated on the basis of degraded **soluble** COD.

149 Quasi-steady stationary polarization curves were recorded *in situ* on anodes and cathodes. Experiments were
150 performed with a classical three-electrode configuration, using a Compactstat IVIUM potentiostat connected
151 to a personal computer. Anodes and cathodes were used as working electrode, a platinum wire as counter
152 electrode and an Ag/AgCl (3M) electrode as reference. All the potentials throughout the text are referred to
153 the Ag/AgCl (3M) electrode. For polarizations on the cathode, a Luggin capillary was adopted to minimize
154 the ohmic drop into the solution. Before each experiment, **SMFC** s were allowed to equilibrate at the open
155 circuit potential (o.c.p.) for at least 30 minutes. Potential was then moved at a scan rate of 10 mV/min from
156 the o.c.p. to 0.1 mV for polarization on anodes, and from the o.c.p. to -0.5 for polarization on cathodes.

157

158 2.2.2 Chemical **oxygen demand analysis** characterizations

159 The soluble Chemical Oxygen Demand was periodically measured by a spectrophotometric method. A
160 portion of solution sampled from each **SMFC** was centrifuged for 15 minutes at 6000 rpm, carefully added to
161 HT-COD cuvette test (Hach Lange GmbH), and digested at 175°C for 15 min (Lange HT 200 S). Upon
162 cooling, the COD value was read by an UV- spectrophotometer (Lange DR 3900).

163

164 **3. Results and discussion**

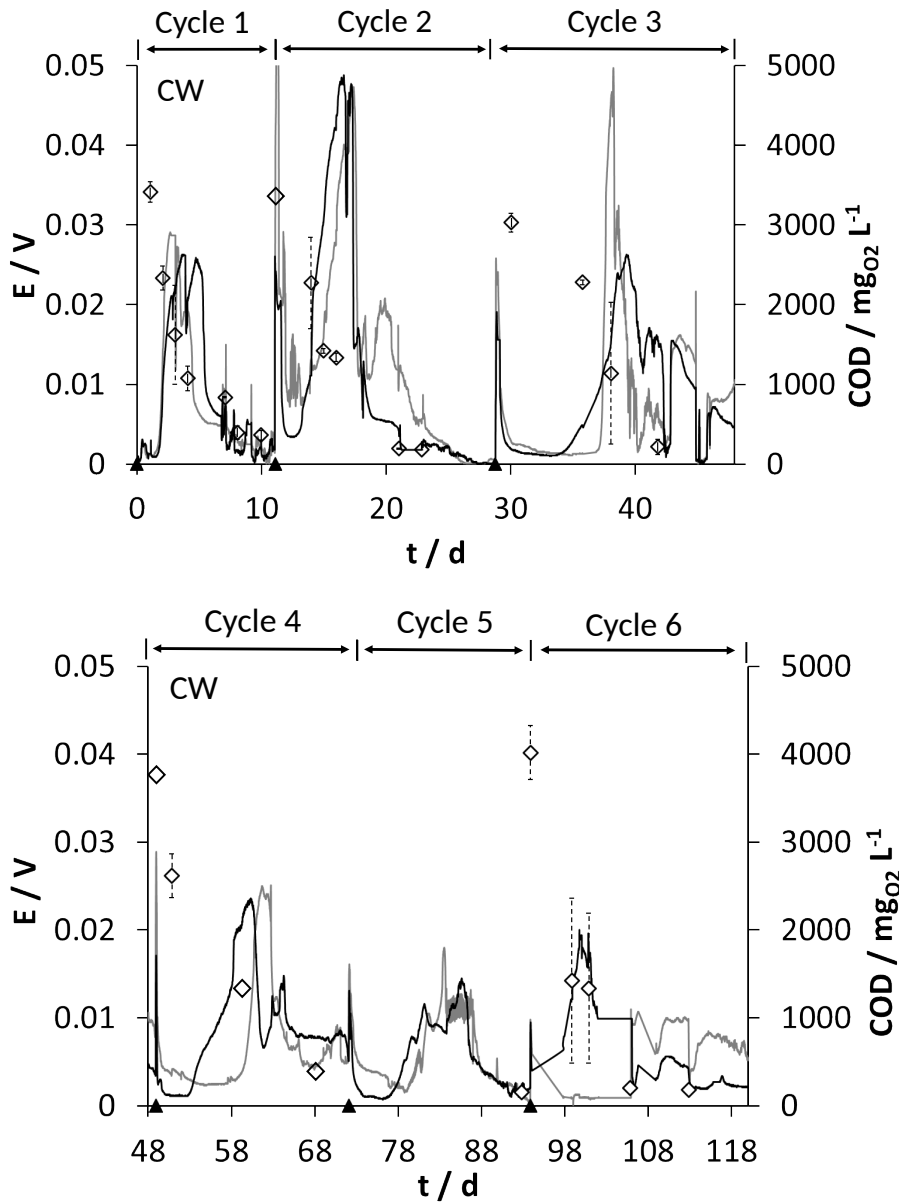
165 *3.1 Electrical signal trend in CW-fed MFCs*

166 **Error! Reference source not found.** reports the evolution of potential trend for two **SMFC** s fed by CW
167 along with the degradation of the soluble COD. **pH was stably in the range 6.5 – 7.5, in all cycles.** The
168 acquisition of **cell** potential over all the time and the measurements of **sCOD** provided an indication of the
169 productivity of each **SMFC** and of the rate of organic substrate degradation.

170 The **SMFC** produced power within 2 days, along with the establishment of anaerobic conditions **insidein** the
171 **SMFC-sanodic chamber** and the colonization of anode and cathode by biofilms. **SMFC** produced a peak of
172 potential at days 4-5 and then potential rapidly decreased down to a negligible value. COD continuously
173 decreased from the first day. **The first eyeCycle 1** of CW degradation **gotwas** completed within 11 days,
174 reaching 95% of COD removal. When a new dose of CW was added at day 11, **SMFC** immediately produced
175 a **sudden**-spike of **cell** potential followed by dramatic drop and a new broader peak with a maximum of 50
176 mV at day 16. After the **deceayingdecay** of **cell** potential to a negligible values and the decrease of COD, a
177 third dose of CW was added. Six cycles of feeding were operated over 125 days. **pH varied in the range 6.7 –**
178 **7.5.**

179 The same trend in **cell** potential-**output** was observed for all cycles, with the generation of a spike just after
180 the addition of a dose of feed followed by the development of a broader peak after a lag time. The duration

181 of the eyelecycles became longer from the third cycle on (about 21 days) and the maximum generatedcell
 182 potential decreased from the fourth cycle on. The rates of soluble COD removal, however, seem quite similar
 183 in each cycle. These two aspects pointed to a progressive deactivation of the whole-SMFCs, over long-term
 184 operation, due to electrode scaling, as documented for wastewater operated with the same SMFC in previous
 185 works [19] as consequence of the alkalinity generated at cathode [20].



186

187

188 **Fig. 1 Graphics of potential trends of two CW-fed SMFCs (solid lines, left y-axis) and COD evolution (◇, right y-axis) over 125 days of operation and six batch cycles. Standard deviation of COD values is reported with vertical**
 189 **bars. Triangles on the x-axis indicate the day of feed addition.**
 190
 191

192 The initial spike of cell potential, found at each cycle, can be attributed to a rapid increase of easily
 193 degradable molecules in the liquid phase, as a consequence of acidogenic fermentation of lactose. Lactose is
 194 easily hydrolyzed to glucose and galactose; sugars fermentation to short chain fatty acids in the bulk
 195 anaerobic medium happens at high rates [21]. The sudden drop of cell potential is likely due to a temporary

196 inhibitory effect of soluble metabolites (e.g. volatile fatty acids) on ARB activity, with a detrimental effect
197 on the electrochemical signal of SMFCs.

198 This effect point to the possibility of monitoring on-line possible accumulations of that soluble metabolites
199 inhibiting biodegradation processes. This would be particularly useful in high-solids anaerobic digestion
200 plants [4], or other biodegradation processes at high organic loading rates [22].

201

202 *3.2 Electrical signal trend in KW-fed MFCs*

203 Fig. 3 Fig. 3 reports the evolution of cell potential over time for two SMFCs fed by OFMSWKW along with
204 the ~~degradation of~~ the soluble COD ~~degradation~~. pH was stably in the range 6.5 – 7.5, in all cycles.

205 Just after the first day of operation, an increase in ~~the cell~~ potential ~~output~~ was observed. The potential
206 reached a maximum of 40 mV at day 2 and then stepwise decreased. COD values continuously decayed from
207 the first day ~~to a constant value of about~~ of each cycle to around 500 mg_{O₂}/L. Further degradation beyond
208 that it was not further degraded. This value was never achieved. At the first cycle, 84% of removal initial
209 COD could be ~~reached~~ removed within 11 days. The second cycle was operated from day 11, when a new
210 dose of OFSMWKW was added into the SMFCs. An initial spike of potential was observed, immediately
211 followed by two broader shoulders. The duration of the second cycle was 17 days.

212 Five cycles of feeding were operated over ~~125~~120 days. The two MFCs gave similar cell potential
213 ~~output signals~~ during all ~~the~~ cycles, even though the duration of the cycles became longer and the resolution
214 between the two shoulders of potentials trends became less defined.

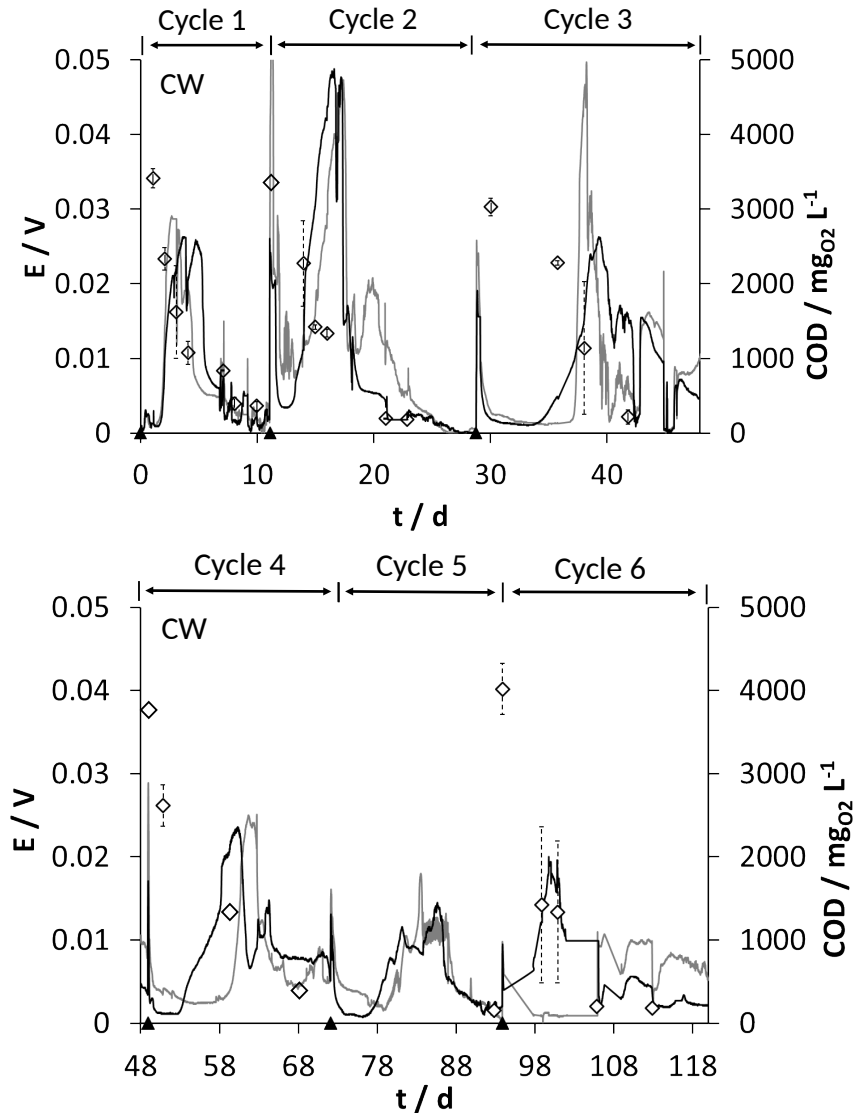
215 The evolution of potential over time is significantly different for KW than CW, even if the duration of COD
216 removal is quite similar. ~~In the case of CW, the produced potential is characterized by sudden increases and~~
217 ~~rapid drops. On the other hand, when SMFCs are fed by OFMSW~~ In this case, the potential remained higher
218 and more stable for longer time. This aspect highlights that the accumulation of VFA is not the main issue in
219 the degradation process, since the composition of KW is more complex than CW.

220 It is possible to deduce that the macromolecular degradation, producing still complex organics other than
221 volatile acids, might prevent the inhibition of the bacterial activity and had an effect on the stabilization of
222 the electrochemical signal of the whole MFCs.

223 ~~Interestingly, the maximum potential decreased cycle by cycle, in parallel with the peak of COD at the~~
224 ~~beginning of each cycle. The peak value of COD was over 3000 mg_{O₂} L⁻¹ in the first cycle, and decreased~~
225 ~~progressively to less than 2000 mg_{O₂} L⁻¹ in cycle 5.~~

226 This condition might be due to the accumulation of mineralized substances in the liquid phase, such as
227 ammonium ions, and other nutrients, creating an inhibitory environment for microbial population and for
228 ARB. The MFC signal followed the general trend of the primary phases of biodegradation (hydrolysis and
229 fermentation). Progressively less organic matter was hydrolyzed from the solid phase (visible as decreasing
230 peaks of COD). Additionally, each cycle lasted for longer time as compared to the first one. Biodegradation
231 rate was visibly decreasing. All these effects were reflected in progressively decreasing peaks of the
232 electrical signals.

233 The MFCs potential trends were. Again, a mirror of an increase of limiting or toxic conditions the overall
 234 microbial population in the bioreactor. The MFCs potential trends were, again, a mirror of an increase of
 235 limiting (or toxic) conditions for the overall microbial population in the bioreactor. In this case the electrical
 236 signals were indicating a progressive increase of inhibition conditions. This application of MFCs in
 237 bioreactors would be useful for monitoring the accumulation of potentially toxic metabolites of
 238 biodegradation (ammonia, H₂S, Na⁺, etc.) on long-term operations of bioreactors [5, 23].



239

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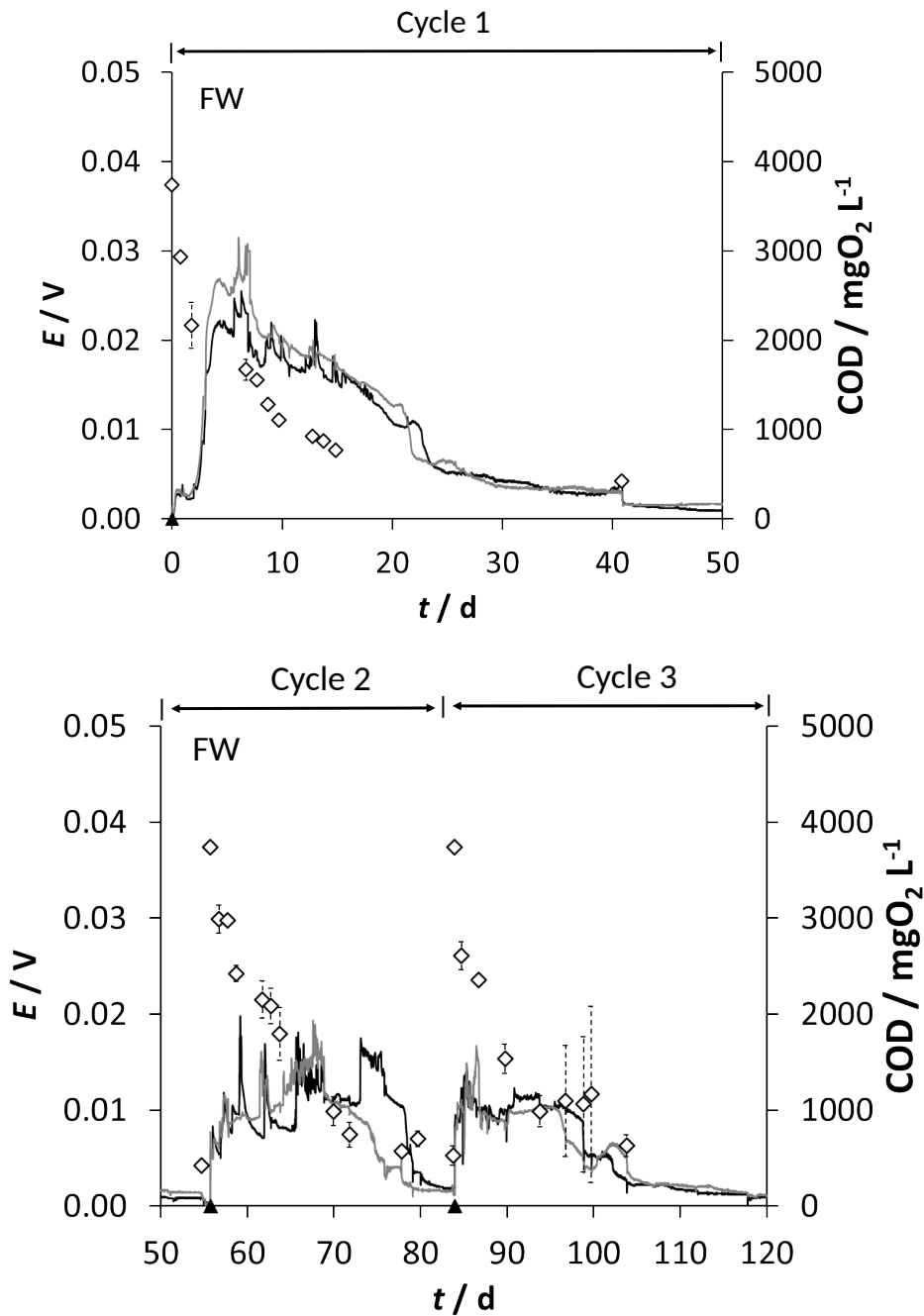
241 **Fig. 4 Graphics of potential trends of two KW-fed SMFCs (solid lines, left y-axis) and COD evolution (\diamond , right y-axis) over 125 days of operation and six batch cycles. Standard deviation of COD values is reported with vertical bars. Triangles on the x-axis indicate the day of feed addition.**

244

245 3.3 Electrical signal trend in FW-fed MFCs

246 Figure 3Fig. 5 reports the evolution of cell potential over time for two SMFCs fed by FW along with the
 247 degradation of the soluble COD degradation. pH was stably in the range 7.5 – 7.9 in all cycles. Both MFCs
 248 started to generate electric signal from day 2, with the cell potential that rapidly increased up to a maximum
 249 of 30 mV and then slowly decreased from day 7 to day 25. In At the same time, COD spiked to nearly 3700
 250 mgO₂ L⁻¹ and continuously decreased from 3700 to 400 mg(O₂)/l and no further decreased beyond this

251 value $\text{mgO}_2 \text{ L}^{-1}$. During the first cycle, the 88% of COD removal was achieved in around 40 days. Electric
 252 signal and COD varied with a really similar trend along the first cycle.
 253



254

255

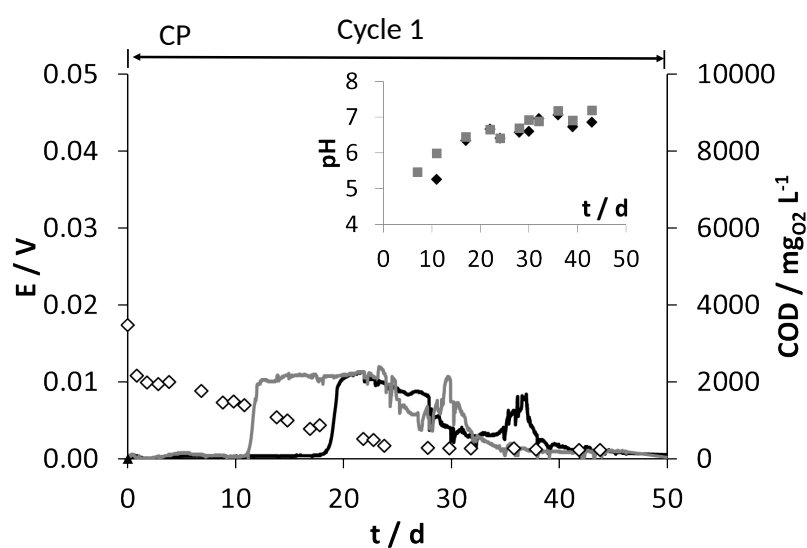
256 **Fig. 5 Graphics of potential trends of two FW-fed SMFCs (solid lines, left y-axis) and COD evolution (\diamond , right y-axis) over 125 days of operation and three batch cycles. Standard deviation of COD values is reported with**
 257 **vertical bars. Triangles on the x-axis indicate the day of feed addition.**
 258
 259

260 At day 55, a new dose of FW was added. Upon addition, the potential firstly rapidly increased up to 10 mV,
 261 then maintained in the range between 10 mV and 20 mV for at least 25 days. The first cycle of operation was
 262 the more efficient in term of electrochemical performance and COD degradation, and different behavior was
 263 observed in the following cycles. At day 84, a new dose of FW was added and evolution of potential and

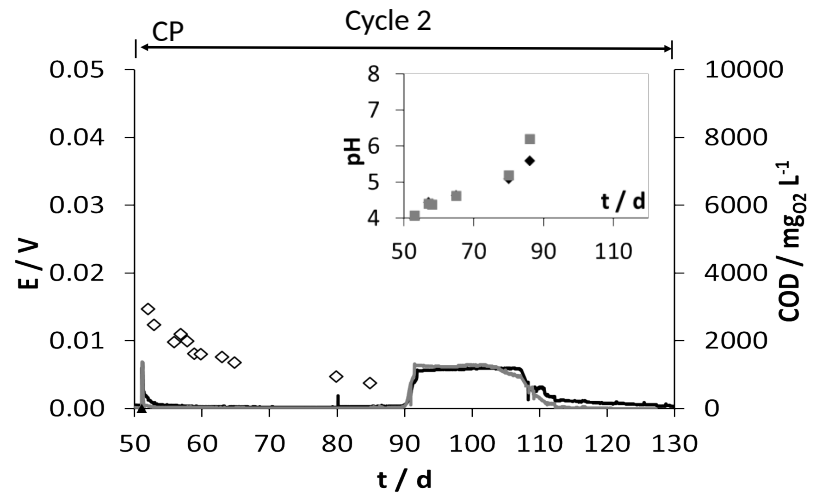
264 COD similar than in the second cycle were achieved. ~~In the first days after the feed addition, the COD was~~
 265 ~~removed very fast for each cycle, while in the last days of each cycle the~~
 266 In cycle 2 and 3, COD removal happened with nearly a double rate, ~~of degradation experienced a slowing~~
 267 ~~down,~~ in parallel with ~~an increasing residual COD concentration cycle by cycle.~~ cell potentials. The peak
 268 values of COD were very similar at each cycle. Differently from CW and KW, peak of COD didn't
 269 correspond to peaks of cell potential, especially at cycles 2 and 3. This is likely to be due to the recalcitrants
 270 hydrolyzed from complex proteins of FW (Table 1). Contrarily, cell potentials lasted for longer time at
 271 relatively high values, as compared to COD values.
 272 In this case, MFCs can be thought as monitors of the presence of long-term biodegradable fractions of the
 273 organic matter.

274 The production of uric acids and ammonia, could affected the pH stability on the electrodes, contrasting the
 275 effect of acidic fermentation. It was recently proved that urea is quickly oxidized at the anode, inducing an
 276 increase of the electric output in single chamber MFCs and pH increase over 9 due to ammonia [24]. A
 277 variability in the concentration trend of ammonia, in contrast with acidic components, including volatile fatty
 278 acids, could cause the signal instability in the SMFCs fed with fish waste. Furthermore, pH variability and
 279 the accumulation of less degradable byproducts in time could stressed the microbial communities on both the
 280 electrodes (bioanode and biocathode), globally lowering the SMFC performances and making also instable
 281 the COD degradation (see error bar in Fig. 3). In fact, the pH of those SMFCs, measured periodically close to
 282 the anode in the bulk solution, was in the range of 7-8, and never decrease below 7 since the first days, while
 283 in the case of CW and OFMSKW, with a lower protein content (Table 1) the pH was around 6.7 ± 0.2 during
 284 the whole experimentation time.

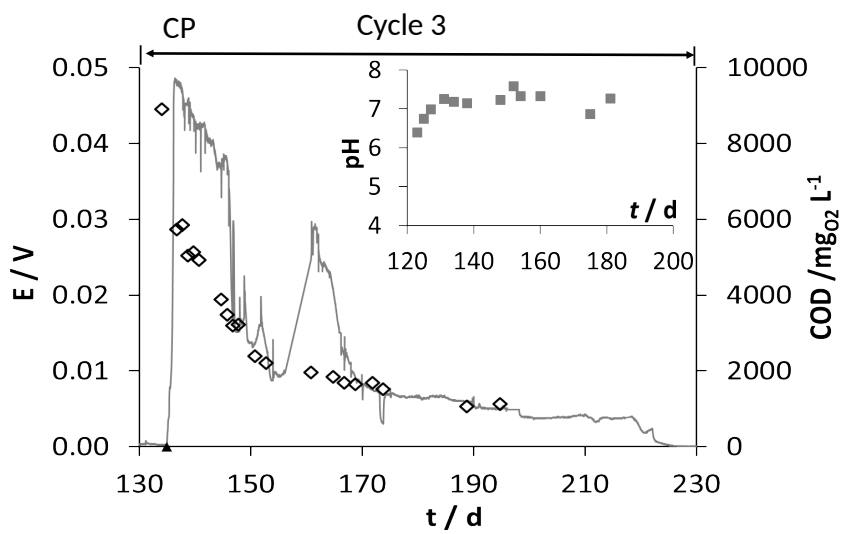
285
 286 *3.4 Electrical signal trend in CP-fed MFCs*
 287



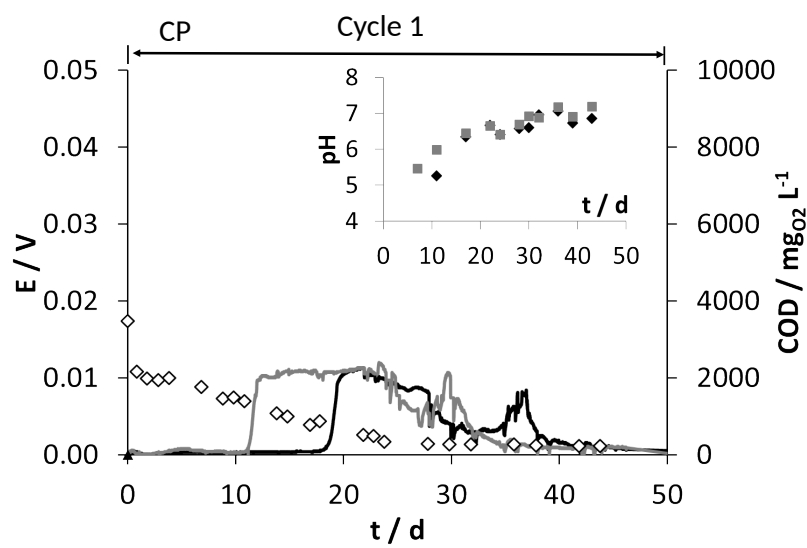
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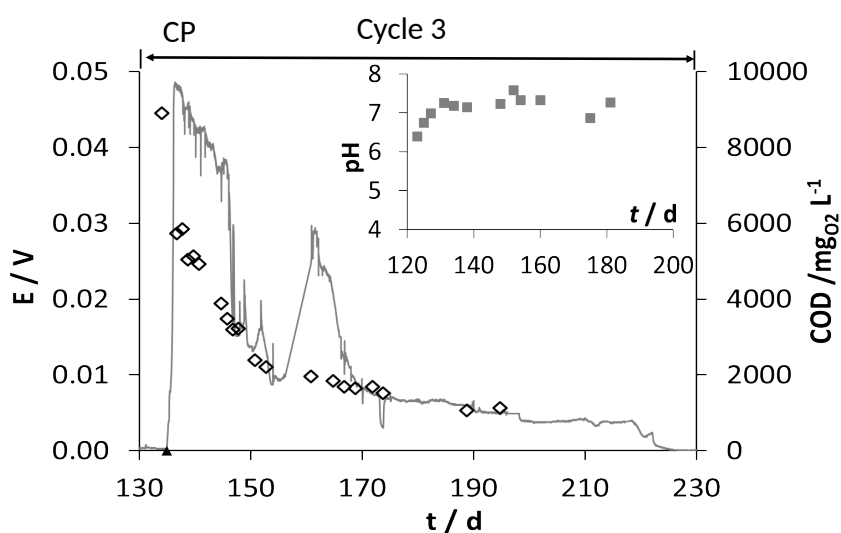
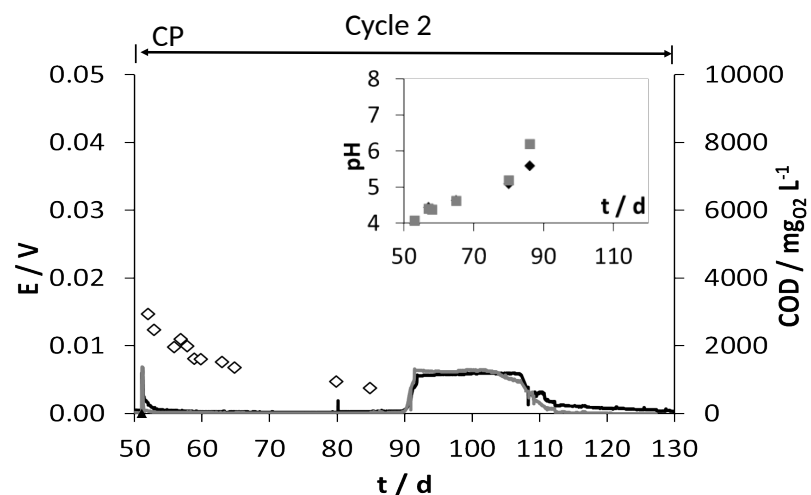
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291 Fig. Figure 4 reports the evolution of potential over time for two SMFCs fed by CP along with the degradation of
292 the soluble COD. In the inset of Figure 4



293



294

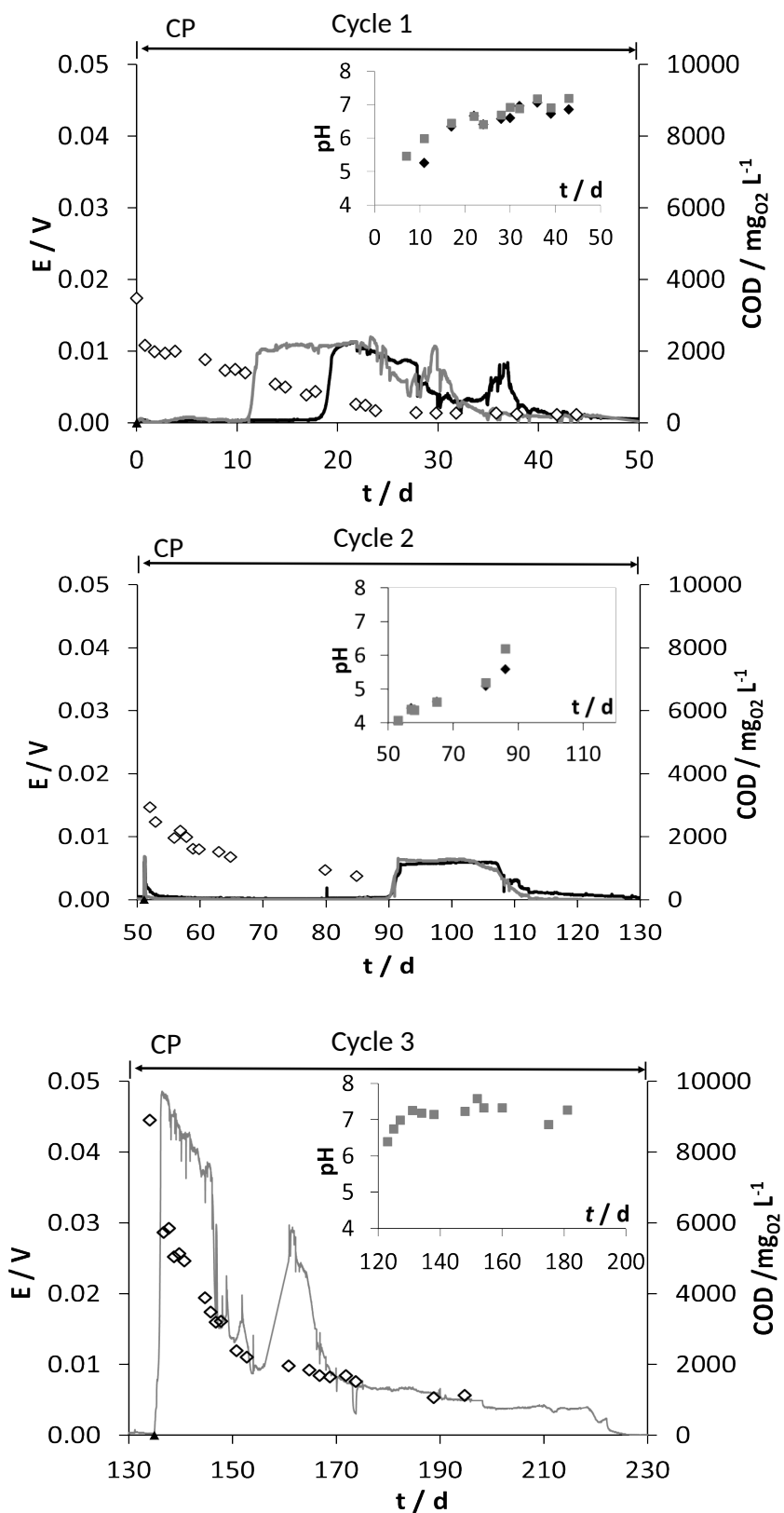
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296 Fig. the variation of pH in the bulk liquid phase of the MFCs bioreactor is shown. In the first two cycles,
 297 acidic conditions (pH 4 – 5) were established in the bioreactor and the pH raised over 6.5 only after 40 days.
 298 Alkalinity was evidently insufficient to buffer the acidity of CP (rich in citric acid and other organic acids).
 299 COD decreased from 3400 mg_{O2}/l to 200 mg_{O2} L⁻¹ along over 45 days of operation of cycle 1 and nearly 50
 300 days in cycle 2.

301 At cycle 3, a buffer medium (potassium bicarbonate, 5 g L⁻¹) was added to the bioreactor to equilibrate the
 302 pH, which remained stable in the range 7 – 7.8 along cycle 3. COD peaked to more than double (over 8000
 303 mg_{O2} L⁻¹), due to more efficient microbial hydrolysis and fermentation of the solids accumulated in the
 304 bioreactor.

305 The SMFC_s did not show any current generation until day 11 and 18. One of the SMFC_s started to produce
 306 power at day 11 and the other SMFC_s at day 18. The rapid increase of cell potential was associated with the
 307 establishment increase of a bulk pH of at least in the liquid phase over 6.5. After the rapid increase up to 10
 308 mV, the potential remained stable for about 10 days then started to slowly decay. Another peak of potential
 309 was observed, along with COD. This happened identically during the decaying. The delay in producing
 310 power can be certainly attributed to the low value of pH that cycle 2. pH acidic conditions inhibited in
 311 perfect parallel the overall biodegradation rate and the activity of exoelectrogenic bacteria. The absence of

312 electric signal from ARB was accomplished by an evidently slow-rate biodegradation, due to inhibited
313 microbial activity.
314



315

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317

318 **Fig. 4 Graphics of potential trends** of two CP-fed MFCs (solid lines, left y-axis) and COD evolution (\diamond , right y-axis) over 230 days of operation and three batch cycles. Standard deviation of COD values is reported with vertical bars. Insets report the trend of pH over time.
319
320

321

322 However, the absence of electric signal was not accomplished by absence of microbial activity inside the
323 MFC. As a matter of fact, COD decreased from 3400 mg_{O2}/l to 200 mg_{O2}/l during all the 45 days of
324 operation of the first cycle. When a new dose of CP was added at day 51, the bulk pH lowered below 4.5 As
325 long as the pH stayed lower than 6.5, apart from an initial spike, no electrical output was obtained from the
326 SMFCs. After day 90, the potential rapid increased up to 7 mV and fixed there for 20 days then dropped to
327 negligible values. Fermentative anaerobic degradation of the substrate took place, but electricity generation
328 was initially inhibited.

329 Since the electric signal is severely affected by the pH, the increase of the potential is a clear indication that
330 optimal condition of pH has been achieved inside the MFC. The kinetics of degradation is slightly lower in
331 the second cycle and this is likely due variation of not-buffered pH.

332 ~~Additional CP-fed SMFCs were operated with 0.2 g COD_{substrate}/gVS_{inoculum}. An increase in potential was~~
333 ~~observed after 2 days of operation as already obtained with the other substrates (results not shown here). As~~
334 ~~already known from biogas and anaerobic digestion field [11], the results from CP-SMFCs highlighted that~~
335 ~~the substrate/inoculum ratio is another key limiting factor when organic solid wastes are used to feed~~
336 ~~microbial metabolism.~~

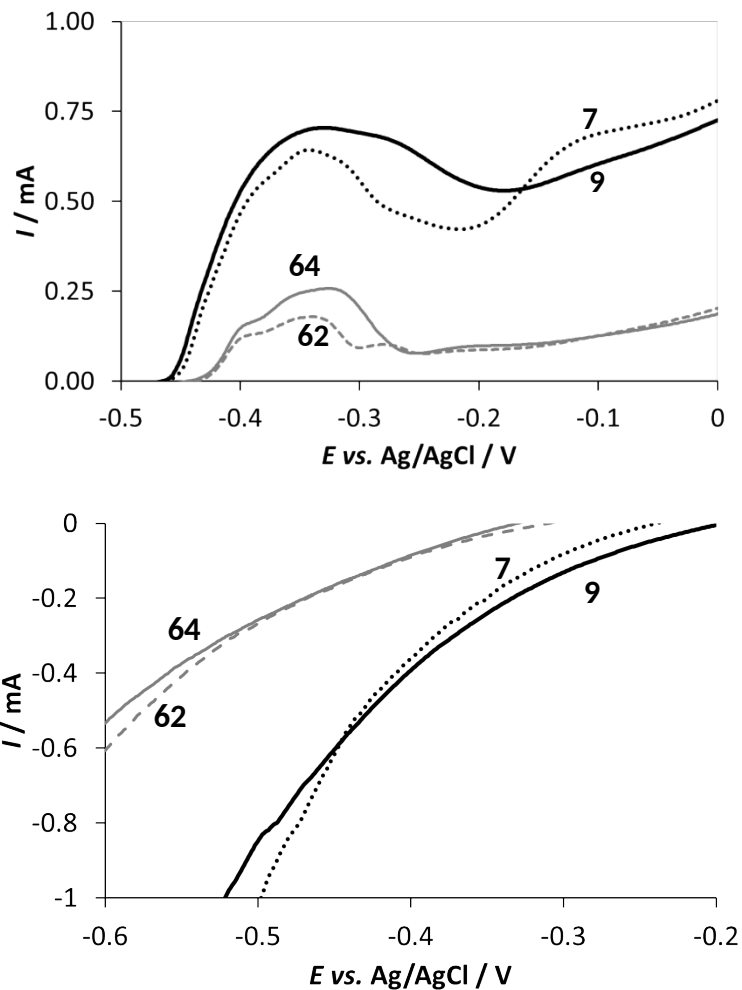
337 When stable pH was guaranteed by equilibrating CP acidity, cell potential trends followed exactly the trend
338 of COD consumption in the liquid phase. These results highlighted another aspect of the chemical equilibria
339 in anaerobic biodegradation environments, which can be efficiently correlated to the trend of the electrical
340 signal produced by ARB, living in the same environment. pH-related inhibition of microbial activity reflects
341 very promptly in drops of MFCs cell potentials.

342

343 3.5 Electrode polarization curves

344 Polarization curves recorded on anodes and cathodes of a FW-fed MFC at different days of operation are
345 reported in Fig. . Polarizations recorded at day 7 and 9 refer to cycle 1, polarization at day 62 and 64 refers to
346 cycle 2. Anodes exhibited a peak around -0.35 V, which can be related to exogenous redox mediator and
347 which position did not change significantly with the time. This potential range can be typically related to the
348 oxidation of short-chain carboxylates, like volatile fatty acids [25]. The current delivered by the anode is
349 about 0.5 mA lower in cycle 2. Similarly, polarization of cathodes exhibited considerably higher currents in
350 cycle 1 than in the second one cycle 2.

351



352

353

354

355 **Fig. 5 Polarization curves of anodes (a) and cathodes (b) recorded at different days in a FW-fed MFC. Numbers**
 356 **near the curves indicate the day of experiments.**

357

358 This indicates, first of all, that the cathode was predominantly characterized by microbially-catalyzed
 359 reduction reactions. As reported in previous works for single-chamber MFCs, bio-cathodic instead of abiotic
 360 mechanisms drive oxygen reduction reactions [16, 26]. This is an important aspect to consider, in the case
 361 that cell potential has to be used as indicator of microbial consortia activity in a bioreactor. To maximize the
 362 MFC system response to inhibitory effects, due to chemical imbalances in the liquid medium of
 363 biodegradation environments, single chamber MFCs might be the ideal solution. A double chamber
 364 architecture with abiotic cathode would be less sensible to inhibitory conditions for microbes.

365 FW biodegradation started showing instability at cycle 3. Error bars reported for SCOD measured in two
 366 MFCs indicate a condition of partial inhibition of the system. In biodegradation of protein-rich organic
 367 materials (see Table 1 for FW composition), it might typically be related to accumulations of ammonia,
 368 hydrogen sulfide or other toxic metabolites [24, 27].

369 The considerable decrease of polarization currents from both bioanodes and biocathodes, already registered
 370 during cycle 2 (days 62 – 64, Figure 5), can be considered as an early-warning for inhibiting conditions for
 371 the whole bioreactor environment. Future experiments should more deeply focus on this aspect. Polarization

372 of electrodes, as mirror of both anodic and cathodic microbial communities, could be studied as early-
 373 warning sensors for inhibiting conditions in anaerobic biodegradation environments.

374

375 3.6. Just monitoring or influencing the biodegradation process of COD removal?

376 ~~The~~ kinetics of the various cycles of each ~~waste-organic substrates~~ is reported in Table 2 **-Error! Reference**
 377 **source not found.** Overall there was a more rapid kinetic of the first days and the first cycle. This points to
 378 a component of degradation due to an aerobic metabolism which preceded the formation of anaerobic
 379 biofilm on the anode and on the cathode. The trend of kinetics is slower in subsequent cycles for all
 380 matrices, while coming practically going to 100% in the case of the milk and near 90% of COD in all the
 381 other cases. As expected, the ~~most-quickquickest~~ degradable substrate was ~~milk~~CW and the less one was
 382 food wastes (the most complex) ~~,that~~KW. For KW, COD removal was never ~~overcame~~higher than 85% due
 383 to the significant presence of ~~oils recalcitrant fractions (e.g. fats and fibers). lignocellulose compounds,~~
 384 ~~which byproducts accumulated cycle by cycle.~~

385 For sake of comparison, in Table 2 the COD removal and CE are listed for the four different substrates.

386 The results showed that the COD removal ranged from 61.4% to 98.77% and CE ranged from 0.76 to 9.9%
 387 respectively. The so low CE achieved is consequence of the un-optimized anode electrode surface/volume
 388 ratio, as the anode occupied a marginal part of the cell volume. The maximum COD removal (98.77) was
 389 obtained for the CW substrate (cycle 6) and the maximum CE (9.91) was determined for the KW substrate
 390 (cycle 2).

391 This indicates that biodegradation were negligibly influenced by the MFC process. In future applications of
 392 MFCs as sensors for monitoring biodegradation process, electrode surface/reactor volume ratio might be
 393 even scaled down, to monitor specific environments.

394 On the contrary, enlarging the electrodes and improving their surface/volume ratio using a different
 395 geometry, will allow to reach and enhance the performance of the MFC process in degrading solid phase
 396 organics with respect the current literature [14].

397 Finally, the optimization of the electrode surface finishing [29] and the use of different materials such as
 398 stainless steel [28] would address other possible needs for an useful application in both cases: monitoring or
 399 influencing the biodegradation process.

400

401

402

403 Table 2 Coulombic efficiency and COD removal during fed-batch MFC operation.

	Sodium acetate (control)		Cheese whey		Kitchen waste		Fish waste		Citrus pulp	
	CE (%)	sCOD removal (%)	CE (%)	sCOD removal (%)	CE (%)	sCOD removal (%)	CE (%)	sCOD removal (%)	CE (%)	sCOD removal (%)
Cycle 1	22.4	98.1	0.76	95.1	3.87	84.37	7.77	88.69	3.13	93.52
Cycle 2	21.3	95.9	2.02	96.9	9.91	64.25	4.37	85.94	2.74	89.70
Cycle 3	23.2	97.4	1.87	95.8	9.40	61.40	3.53	89.20	-	-

404

405 **4. Conclusions**

406 Four sets of membraneless single-chamber Microbial Fuel Cells were operated in duplicate and in parallel
407 over more than 100 days, inoculated with anaerobic sludge of a biogas production plant and cyclically fed
408 with the following different organic substrates: i) organic fraction of **Kitchen** waste (KW), ii) **Cheese** whey
409 from dairy industries (CW), iii) residues of fish previously processed to recover oils (FW), iv) pulp waste
410 from citrus juice production (CP).

411 ~~All the tested SMFCs were able to produce an electric signal that varied in intensity as function of the~~
412 ~~substrate principal component, the COD concentration and the feeding cycle. Nevertheless, the pH mostly~~
413 ~~seems affect the electric signal.~~

414 ~~This suggest the possibility of using SMFC output as sensor to control the pH and other parameters in~~
415 ~~several industrial anaerobic fermentation processes, such as biogas plants.~~

416 All MFCs were able to produce electric signals that varied in intensity as function of the chemical equilibria
417 in the liquid phase of the bioreactor and the biodegradation rates achieved.

418 Sudden upcoming inhibiting conditions for microbial community in the MFC bioreactor corresponded to
419 sudden drops in cell potentials. Progressive decrease of biodegradation efficiency in successive batch cycles
420 corresponded to diminishing peaks of cell potentials. pH drops below 6.5 inhibited both biodegradation and
421 anodic exoelectrogenic activity. The presence of recalcitrant fractions gave a delay between soluble COD
422 degradation trend and the electrical signal over time. Both anodic and cathodic polarization curves, gave
423 lower currents corresponding to incoming inhibiting conditions in the bioreactors.

424 Biotic mechanisms driving cathodic reduction reactions can help in having more sensible responses from
425 MFCs coupled to bioreactors. A deeper study should follow these preliminary indications in considering
426 MFCs as sensor to monitor and control anaerobic biodegradation processes.

427

428 **References**

- 429 [1] Fava F, Totaro G, Diels L, Reis M, Duarte J, Carioca OB, et al. Biowaste biorefinery in Europe:
430 opportunities and research & development needs. *N Biotechnol* 2013;32.
431 doi:10.1016/j.nbt.2013.11.003.
- 432 [2] Agler MT, Wrenn B a., Zinder SH, Angenent LT. Waste to bioproduct conversion with undefined
433 mixed cultures: The carboxylate platform. *Trends Biotechnol* 2011;29:70–8.
434 doi:10.1016/j.tibtech.2010.11.006.
- 435 [3] Manzini E, Scaglia B, Schievano A, Adani F. Dark fermentation effectiveness as a key step for waste
436 biomass refineries: influence of organic matter macromolecular composition and bioavailability. *Int J*
437 *Energy Res* 2015;31:n/a – n/a. doi:10.1002/er.3347.
- 438 [4] Schievano A, D'Imporzano G, Malagutti L, Fragali E, Ruboni G, Adani F. Evaluating inhibition
439 conditions in high-solids anaerobic digestion of organic fraction of municipal solid waste. *Bioresour*
440 *Technol* 2010;101:5728–32.

- 441 [5] Chen Y, Cheng JJ, Creamer KS. Inhibition of anaerobic digestion process: A review. *Bioresour*
442 *Technol* 2008;99:4044–64. doi:10.1016/j.biortech.2007.01.057.
- 443 [6] Drake HL, Küsel K, Matthies C. Acetogenic Prokaryotes. *The Prokaryotes*, Berlin, Heidelberg:
444 Springer Berlin Heidelberg; 2013, p. 3–60. doi:10.1007/978-3-642-30141-4_61.
- 445 [7] Thauer RK, Hedderich R, Fischer R. Reactions and Enzymes Involved in Methanogenesis from CO₂
446 and H₂. *Methanogenesis*, Boston, MA: Springer US; 1993, p. 209–52. doi:10.1007/978-1-4615-2391-
447 8_5.
- 448 [8] Logan BE. *Nature Reviews Microbiology* 7, 375-381 (May 2009) | doi:10.1038/nrmicro2113.
- 449 [9] Rismani-Yazdi H, Christy AD, Carver SM, Yu Z, Dehority BA, Tuovinen OH. Effect of external
450 resistance on bacterial diversity and metabolism in cellulose-fed microbial fuel cells. *Bioresour*
451 *Technol* 2011;102:278–83. doi:10.1016/j.biortech.2010.05.012.
- 452 [10] Fornero JJ, Rosenbaum M, Angenent LT. Electric Power Generation from Municipal, Food, and
453 Animal Wastewaters Using Microbial Fuel Cells. *Electroanalysis* 2010;22:832–43.
454 doi:10.1002/elan.200980011.
- 455 [11] Clauwaert P, Rabaey K, Aelterman P, De Schamphelaire L, Pham TH, Boeckx P, et al. Biological
456 denitrification in microbial fuel cells. *Environ Sci Technol* 2007;41:3354–60. doi:10.1021/es062580r.
- 457 [12] Rago L, Ruiz Y, Baeza JA, Guisasola A, Cortés P. Microbial community analysis in a long-term
458 membrane-less microbial electrolysis cell with hydrogen and methane production.
459 *Bioelectrochemistry* 2015;106:359–68. doi:10.1016/j.bioelechem.2015.06.003.
- 460 [13] Chae K-J, Choi M-J, Kim K-Y, Ajayi FF, Park W, Kim C-W, et al. Methanogenesis control by
461 employing various environmental stress conditions in two-chambered microbial fuel cells. *Bioresour*
462 *Technol* 2010;101:5350–7. doi:10.1016/j.biortech.2010.02.035.
- 463 [14] Mohan SV, Chandrasekhar K. Solid phase microbial fuel cell (SMFC) for harnessing bioelectricity
464 from composite food waste fermentation: Influence of electrode assembly and buffering capacity.
465 *Bioresour Technol* 2011;102:7077–85. doi:10.1016/j.biortech.2011.04.039.
- 466 [15] Goud RK, Babu PS, Mohan SV. Canteen based composite food waste as potential anodic fuel for
467 bioelectricity generation in single chambered microbial fuel cell (MFC): Bio-electrochemical
468 evaluation under increasing substrate loading condition. *Int J Hydrogen Energy* 2011;36:6210–8.
469 doi:10.1016/j.ijhydene.2011.02.056.
- 470 [16] Guerrini E, Cristiani P, Trasatti SPM. Relation of anodic and cathodic performance to pH variations
471 in membraneless microbial fuel cells. *Int J Hydrogen Energy* 2013;38:345–53.
472 doi:10.1016/j.ijhydene.2012.10.001.
- 473 [17] Guerrini E, Grattieri M, Faggianelli A, Cristiani P, Trasatti S. PTFE effect on the electrocatalysis of
474 the oxygen reduction reaction in membraneless microbial fuel cells. *Bioelectrochemistry* 2015.
475 doi:10.1016/j.bioelechem.2015.05.008.
- 476 [18] Schievano A, Pognani M, D’Imporzano G, Adani F. Predicting anaerobic biogasification potential of
477 ingestates and digestates of a full-scale biogas plant using chemical and biological parameters.

- 478 Bioresour Technol 2008;99:8112–7.
- 479 [19] M. Santini, M. Guilizzoni, M. Lorenzi, P. Atanassov, E. Marsili, S. Fest-Santini, P. Cristiani, C.
480 Santoro. Three-Dimensional X-ray Micro Computed Tomography Of Carbonates And Biofilm On
481 Operated Cathode In Single Chamber Microbial Fuel Cell. *Biointerphases*, 10 (3):031009 (2015);
482 <http://dx.doi.org/10.1116/1.4930239>
- 483 [20] Gajda I, Greenman J, Melhuish C, Santoro C, Li B, Cristiani P, et al. Water formation at the cathode
484 and sodium recovery using Microbial Fuel Cells (MFCs). *Sustain Energy Technol Assessments*
485 2014;7:187–94.
- 486 [21] Frascari D, Cappelletti M, Mendes JDS, Alberini A, Scimonelli F, Manfreda C, et al. A kinetic study
487 of biohydrogen production from glucose, molasses and cheese whey by suspended and attached cells
488 of *Thermotoga neapolitana*. *Bioresour Technol* 2013;147:553–61.
489 [doi:10.1016/j.biortech.2013.08.047](https://doi.org/10.1016/j.biortech.2013.08.047).
- 490 [22] Davila-Vazquez G, Cota-Navarro CB, Rosales-Colunga LM, de León-Rodríguez A, Razo-Flores E.
491 Continuous biohydrogen production using cheese whey: Improving the hydrogen production rate. *Int*
492 *J Hydrogen Energy* 2009;34:4296–304. [doi:10.1016/j.ijhydene.2009.02.063](https://doi.org/10.1016/j.ijhydene.2009.02.063).
- 493 [23] Hartmann H, Ahring BK. Anaerobic digestion of the organic fraction of municipal solid waste:
494 Influence of co-digestion with manure. *Water Res* 2005;39:1543–52.
495 [doi:10.1016/j.watres.2005.02.001](https://doi.org/10.1016/j.watres.2005.02.001).
- 496 [24] C. Santoro, I. Ieropoulos, J. Greenman, P. Cristiani, T. Vadas, A. Mackay, B. Li. Current Generation
497 in Membraneless Single Chamber Microbial Fuel Cells (MFCs) Treating Urine. *Journal of Power*
498 *Sources* 238 (2013) 190-196. <http://dx.doi.org/10.1016/j.jpowsour.2013.03.095>
- 499 [25] Rabaey K, Rozendal R a. Microbial electrosynthesis - revisiting the electrical route for microbial
500 production. *Nat Rev Microbiol* 2010;8:706–16. [doi:10.1038/nrmicro2422](https://doi.org/10.1038/nrmicro2422).
- 501 [26] Guerrini E, Grattieri M, Trasatti SP, Bestetti M, Cristiani P. Performance explorations of single
502 chamber microbial fuel cells by using various microelectrodes applied to biocathodes. *Int J Hydrogen*
503 *Energy* 2014;39:21837–46. [doi:10.1016/j.ijhydene.2014.06.132](https://doi.org/10.1016/j.ijhydene.2014.06.132).
- 504 [27] Kim DH, Oh SE. Continuous high-solids anaerobic co-digestion of organic solid wastes under
505 mesophilic conditions. *Waste Manag* 2011;31:1943–8. [doi:10.1016/j.wasman.2011.05.007](https://doi.org/10.1016/j.wasman.2011.05.007).
- 506 [28] E. Guerrini, P. Cristiani, M. Grattieri, C. Santoro, B. Li, S. Trasatti. Electrochemical behavior of
507 stainless steel anodes in microbial fuel cells. *J. Electrochem. Soc.* 2014 161(3): H62-H67 2014.
508 <http://dx.doi.org/10.1149/2.096401jes> . *Journal of Electrochemical Society*. volume 161, issue 3,
509 H62-H67.
- 510 [29] G. Papaharalabos, J. Greenman, C. Melhuish, I. Ieropoulos, P. Cristiani, C. Santoro, B. Li. Increased
511 power output from micro porous layer (MPL) cathode microbial fuel cells (MFC). *International*
512 *Journal of Hydrogen Energy* 38 (2013) 11552-11558.
513 <http://dx.doi.org/10.1016/j.ijhydene.2013.05.138>
- 514

Signal trends of microbial fuel cells fed with different food-industry residues

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ABSTRACT

A microbial fuel cell (MFC) is an anaerobic bioreactor where soluble metabolites liberated by hydrolysis and fermentation of macromolecules are simultaneously available for anode respiring bacteria (ARB). ARB can be influenced by chemical imbalances in the liquid phase of the bioreactor. The objective of the work was to explore the trend of electric signals generated by MFCs, in relation to anaerobic biodegradation of four different solid food-industry residual substrates. Four sets of membraneless single-chamber MFCs were operated in batch mode, with solid waste substrates characterized by a different base component: i) mixed kitchen waste (fibers), ii) whey from dairy industries (sugar), iii) fisheries residues previously processed to recover oils (proteins), iv) pulp waste from citrus juice production (acidic).

All the tested MFCs were able to produce an electric output with different trends, depending on the principal component of the solid substrate. MFC potential varied as function of the COD and the feeding cycle, as well as of the substrate.

The pH variability during the fermentative process significantly affected the electric output. Citrus (acidic) pulp fed MFCs started to operate only when the pH raised up 6.5. MFCs fed with mix food wastes had a relatively stable electric signal; fish based waste caused spiking in the MFC signal and an averaging in the COD degradation trend. This phenomenon was attributed to a pH instability induced by proteins degradation forming ammonia.

The fermentation process was strongly predominant with respect the electrochemical process in MFCs and the coulombic efficiency (CE) was low, ranging between 2 to 10 %. This result call for a deeper exploration of harvesting power from solid wastes and pointed also to the possibility of using a MFC to monitor important parameters of fermentation processes in biotech production plants.

39 1. Introduction

40 The new paradigm of circular economy claims new technological approach for energy and resource
41 recovery. Agro-food industry produces massive amounts of organic materials as secondary streams and
42 waste [1]. Microbes naturally evolve enzymes and pathways that can convert solid biomass-derived carbon
43 sources into valuable fuels and products, such as biomethane, biohydrogen, biodegradable polymers,
44 carboxylates [2]. Biological conversions play a fundamental role in waste refinery chains and especially of
45 agricultural and food-industry residues. In this context, microbial electrochemical technologies (METs) offer
46 potential innovative approaches in wastes treatment.

47 Anaerobic bioconversion processes such as anaerobic digestion and dark fermentation rely on sequential
48 microbial hydrolytic and fermentative processes that solubilize different substances in the liquid phase.
49 Macromolecules are degraded to soluble molecules that become available to secondary fermentations or
50 anaerobic respirations, by other microbes forming part of complex consortia [3]. In many cases, the
51 liberation of soluble metabolites might significantly change the chemical equilibria of the liquid and gaseous
52 phases where microbes live [4]. This is often the cause of inhibition of more sensible microbial species and
53 possible imbalances of the overall biodegradation process [5].

54 In anaerobic biodegradation, secondary metabolisms (e.g. acetogenesis, acetoclastic methanogenesis,
55 hydrogenophilic methanogenesis, denitrification, sulphate reduction, etc.) rely on the availability of short-
56 chain organic molecules, inorganic ions and soluble gasses, such as volatile fatty acids, di-hydrogen,
57 hydrogen sulphide, ammonium, nitrate, carbon dioxide, bicarbonates, etc [6,7].

58 Similarly, in microbial electrochemical systems like microbial fuel cells (MFC), electroactive microbial
59 species like anode respiring bacteria (ARB) rely on the same substances to transfer electrons and produce
60 current [8]. Macromolecules, in regular MFCs, should be at least hydrolyzed and pre-fermented by
61 fermentative microbial species before being available to ARB, which preferentially oxidize low-carbon
62 carboxylates, as indicated in various literature contributions [8–11].

63 For this reason, MFCs can be used as a mirror-process for secondary biodegradation metabolisms. In
64 anaerobic environments, the electrical signal produced by ARB activity can give real-time hints on the trend
65 of the ongoing biodegradation mechanisms and biochemical conditions, such as availability of soluble low-
66 carbon organics, availability of mineral nutrients, favorable chemical equilibria in the liquid medium like pH,
67 electrical conductivity, etc. A widely recognized issue is the competition of methanogenic populations and
68 ARB for the same organics [12,13]. This corroborates the assumption that ARB activity, measured as voltage
69 generation, could be used as monitor of the interactions between fermentative and methanogenic microbial
70 populations in anaerobic biodegradation environments.

71 Electricity harvesting with complex biomass was recently achieved by Mohan et al. [14], who fed open-air
72 cathode, single-chamber MFCs with different types of canteen-based food waste. The best performing
73 configuration, where a proton exchange membrane separated the cathode from the anodic chamber, achieved
74 a peak power density of 170 mW m⁻² with open circuit potential (OCP) of 463 mV. Similar results with
75 similar substrates were obtained by Goud et al [15], who tested increasing organic loading rates (OLR) in bio-

76 electrochemical reactors fed continuously. OLR of 1.39 kg COD/m³-day both power density and OCP started
77 decreasing, due to inhibiting concentrations of volatile fatty acids (>800 mg/L) and acidic pH conditions
78 (pH=6).

79 Pretreatment of wastes from agro-food is achieved in several ways (e.g. to extract essential oils and proteins
80 from specific wastes, such as citrus pulp, residual fish) and MET can be thought as downstream processing
81 for energy harvesting in Microbial Fuel Cells [14] or for further bio-processing (e.g. electrofermentation
82 [12]), however, very little is reported in literature about biodegradation pathways of complex organic
83 matrices in the bulk medium. In particular, the relationships between the electric signal produced by ARB
84 and the biodegradation process as-a-whole (hydrolysis of macromolecules, fermentative metabolisms, etc.).
85 Here, we studied the electrical signals produced by MFCs during anaerobic biodegradation of four different
86 types of agro-industrial residual materials of interest in Mediterranean agro-food sectors: citrus pulp, fishery
87 waste, cheese whey and kitchen waste. Voltage trends were monitored on long-term, over 100 days, with
88 successive batch cycles, to evaluate the response of the electrochemical system to the anaerobic
89 biodegradation of the solid matrices.

90

91 **2. Materials and methods**

92 **2.1 MFC configuration and setup**

93 Four sets of membraneless single-chamber MFCs were operated in duplicate and in parallel over more than
94 100 days. The total volume of each MFC was 125 ml and the design was previously reported [16]. Anodes
95 were made of 3×5 cm rolled carbon cloth sheet (Saati C1, Legnano, Italy), electrically connected to a copper
96 wire. Three layers of non-conductive high-viscosity epoxy resin (Mapei Epojet) were applied to ensure
97 insulation at the connection between copper and carbon cloth. Cathodes were made of 5×5cm carbon cloth
98 sheets modified by the addition of a Gas Diffusion Layer (GDL) on the air side. The GDL composition has
99 been described in [17] and the PTFE content is 80%_{w/w} with respect to carbon powder. The geometric
100 cathodic surface area exposed to the solution was 3 cm². Anode and cathode were then connected through an
101 external circuit with a resistance of 100 Ω.

102 All MFCs were operated at mesophilic temperature of 35±1 °C in batch mode without pH adjustment. MFCs
103 were inoculated with 90 ml of anaerobic mesophilic sludge obtained from a municipal wastewater treatment
104 plant (Cremona, Italy). The volatile solids (VS) content in the sludge was 15 g_{VS}kg⁻¹. This concentration is
105 typically used in standard batch-like anaerobic digestion tests [18]. The sludge was not subjected to any
106 pretreatment. A concentrated solution of nutrients was added at the beginning of the experimentation. The
107 stock solution of nutrients contained (in g/L): KH₂PO₄ (0.27), Na₂HPO₄·12H₂O (1.12), NH₄Cl (0.53),
108 CaCl₂·2H₂O (0.075), MgCl₂·6H₂O (0.10), FeCl₂·4H₂O (0.02). Analytical grade reagents and double distilled
109 water were used.

110 Two MFCs were fed with each organic substrate in the form of dried powder (1 mm particle size): i) cheese
111 whey powder (CW); ii) kitchen waste (KW); iii) fish waste (FW); iv) citrus pulp (CP). The macromolecular
112 composition of the four substrates was reported in Table 1. CW was a commercial by-product from dairy

113 industries (Cremona, Italy) used as animal feed. KW was a mixture of animal and vegetal food waste
 114 prepared in lab according to the following recipe: 30 g egg shells; 30 g dried bread, 50 g corn flour, 100g
 115 grated cheese, 75 g cracker, 10 g coffee grounds, 130g apple peel, 300g green salad, 145 g orange peel, 85 g
 116 zucchini peel, 68 g banana peel, 56 g carrots, 30 pumpkin skin, 20 g kiwi peel, 30 g fennels, 16 g potato peel.
 117 Food mixture was grinded with a kitchen blender, homogenized and finally dried at 105 °C. FW was
 118 obtained from fish after an enzymatic pretreatment to remove oils (no alcohols used). CP was obtained from
 119 citrus juice production plant (Catania, Italy).

120

121 **Table 1 Main macromolecular constituents of the four solid wastes**

Substrate	Carbohydrates (% of DM)	Fibers (% of DM)	Fats (% of DM)	Proteins (% of DM)	Ashes (% of DM)
CW	70 (lactose)	-	-	12	8.5
KW	53.4	19.2	9.6	14.3	3.5
FW	0.3	-	3.8	51.2	20.1
CP	8.5	43.1 (33.2 cellulose, 9.9 hemicellulose)	3.1	26.9 (19.7 pectin)	18.4

122

123 The amounts of inoculum and organic substrates introduced in each MFCs were determined on the basis of
 124 preliminary analytics determination (volatile solids and total solids). The organic substrate to inoculum ratio
 125 was 0.35 g sCOD_{substrate}/g VS_{inoculum}. A new dose of feed was added when negligible potential values were
 126 obtained and soluble Chemical Oxygen Demand (COD) fell down to a constant value.

127

128 2.2 Tests

129 2.2.1 Data acquisition, electrochemical experiments and calculations

130 The potential difference across the 100 Ω resistance (*R*) was acquired every 10 minutes, via a multichannel
 131 Data Logger (Graphtech midi Logger GL820). The generated current (*I*) was calculated by the equation $I =$
 132 V/R , where *I* is the current flowing through the external resistance. The total charge flowed into the electrical
 133 circuit at the end of each batch cycle was calculated by integrating the current over time. Coulombic
 134 efficiency (CE) was then evaluated on the basis of degraded soluble COD.

135 Quasi-steady stationary polarization curves were recorded *in situ* on anodes and cathodes. Experiments were
 136 performed with a classical three-electrode configuration, using a Compactstat IVIUM potentiostat connected
 137 to a personal computer. Anodes and cathodes were used as working electrode, a platinum wire as counter
 138 electrode and an Ag/AgCl (3M) electrode as reference. All the potentials throughout the text are referred to
 139 the Ag/AgCl (3M) electrode. For polarizations on the cathode, a Luggin capillary was adopted to minimize
 140 the ohmic drop into the solution. Before each experiment, MFC s were allowed to equilibrate at the open
 141 circuit potential (o.c.p.) for at least 30 minutes. Potential was then moved at a scan rate of 10 mV/min from
 142 the o.c.p. to 0.1 mV for polarization on anodes, and from the o.c.p. to -0.5 for polarization on cathodes.

143

144 2.2.2 Chemical characterizations

145 The soluble Chemical Oxygen Demand was periodically measured by a spectrophotometric method. A
146 portion of solution sampled from each MFC was centrifuged for 15 minutes at 6000 rpm, carefully added to
147 HT-COD cuvette test (Hach Lange GmbH), and digested at 175°C for 15 min (Lange HT 200 S). Upon
148 cooling, the COD value was read by an UV- spectrophotometer (Lange DR 3900).

149

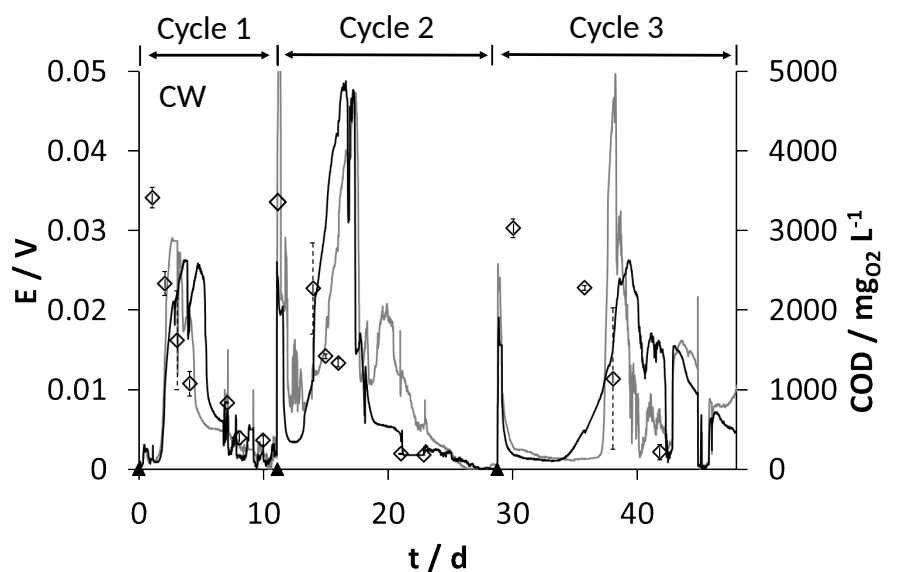
150 3. Results and discussion

151 3.1 Electrical signal trend in CW-fed MFCs

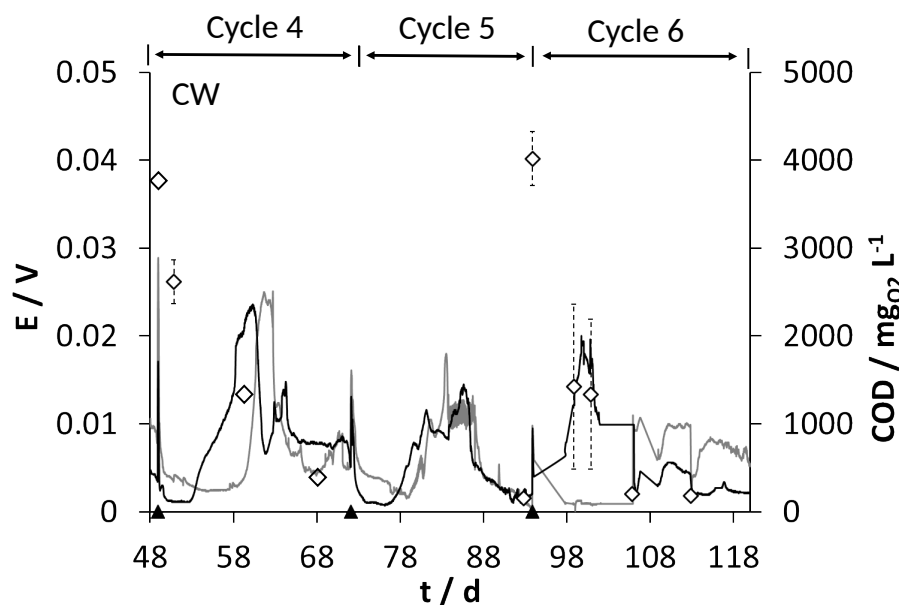
152 **Error! Reference source not found.** reports the evolution of potential trend for two MFC s fed by CW
153 along with the degradation of the soluble COD. pH was stably in the range 6.5 – 7.5, in all cycles. The
154 acquisition of cell potential over all the time and the measurements of sCOD provided an indication of the
155 productivity of each MFC and of the rate of organic substrate degradation.

156 The MFC produced power within 2 days, along with the establishment of anaerobic conditions in the anodic
157 chamber and the colonization of anode and cathode by biofilms. MFC produced a peak of potential at days 4-
158 5 and then potential rapidly decreased down to a negligible value. COD continuously decreased from the first
159 day. Cycle 1 of CW degradation was completed within 11 days, reaching 95% of COD removal. When a new
160 dose of CW was added at day 11, MFC immediately produced a spike of cell potential followed by dramatic
161 drop and a new broader peak with a maximum of 50 mV at day 16. After the decay of cell potential to a
162 negligible values and the decrease of COD, a third dose of CW was added. Six cycles of feeding were
163 operated over 125 days. pH varied in the range 6.7 – 7.5.

164 The same trend in cell potential was observed for all cycles, with the generation of a spike just after the
165 addition of a dose of feed followed by the development of a broader peak after a lag time. The duration of
166 the cycles became longer from the third cycle on (about 21 days) and the maximum cell potential decreased
167 from the fourth cycle on. The rates of soluble COD removal, however, seem quite similar in each cycle.
168 These two aspects pointed to a progressive deactivation of the MFCs, over long-term operation, due to
169 electrode scaling, as documented for wastewater operated with the same MFC in previous works [19] as
170 consequence of the alkalinity generated at cathode [20].



171



172

173 **Fig. 1** Graphics of potential trends of two CW-fed MFCs (solid lines, left y-axis) and COD evolution (\diamond , right y-axis) over 125 days of operation and six batch cycles. Standard deviation of COD values is reported with vertical
 174 bars. Triangles on the x-axis indicate the day of feed addition.
 175
 176

177 The initial spike of cell potential, found at each cycle, can be attributed to a rapid increase of easily
 178 degradable molecules in the liquid phase, as a consequence of acidogenic fermentation of lactose. Lactose is
 179 easily hydrolyzed to glucose and galactose; sugars fermentation to short chain fatty acids in the bulk
 180 anaerobic medium happens at high rates [21]. The sudden drop of cell potential is likely due to a temporary
 181 inhibitory effect of soluble metabolites (e.g. volatile fatty acids) on ARB activity, with a detrimental effect
 182 on the electrochemical signal of MFCs.

183 This effect point to the possibility of monitoring on-line possible accumulations of that soluble metabolites
 184 inhibiting biodegradation processes. This would be particularly useful in high-solids anaerobic digestion
 185 plants [4], or other biodegradation processes at high organic loading rates [22].
 186

187 3.2 Electrical signal trend in KW-fed MFCs

188 Fig. 1 Fig. 1 reports the evolution of cell potential over time for two MFCs fed by KW along with the soluble
189 COD degradation. pH was stably in the range 6.5 – 7.5, in all cycles.

190 Just after the first day of operation, an increase in cell potential was observed. The potential reached a
191 maximum of 40 mV at day 2 and then stepwise decreased. COD values continuously decayed from the first
192 day of each cycle to around 500 mg_{O₂} L⁻¹. Further degradation beyond this value was never achieved. At the
193 first cycle, 84% of initial COD could be removed within 11 days. The second cycle was operated from day
194 11, when a new dose of KW was added into the MFCs. An initial spike of potential was observed,
195 immediately followed by two broader shoulders. The duration of the second cycle was 17 days.

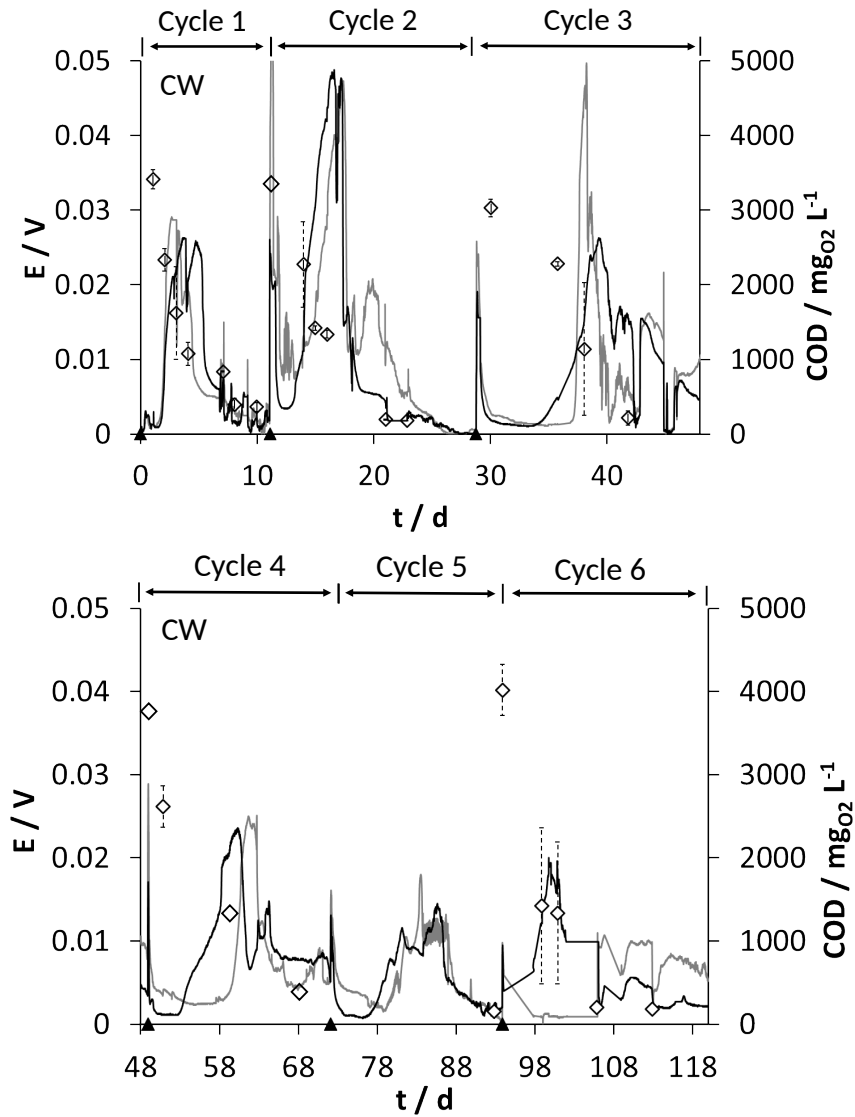
196 Five cycles of feeding were operated over 120 days. The two MFCs gave similar cell potential signals during
197 all cycles, even though the duration of the cycles became longer and the resolution between the two
198 shoulders of potentials trends became less defined. The evolution of potential over time is significantly
199 different for KW than CW, even if the duration of COD removal is quite similar. In this case, the potential
200 remained higher and more stable for longer time. This aspect highlights that the accumulation of VFA is not
201 the main issue in the degradation process, since the composition of KW is more complex than CW.

202 It is possible to deduce that the macromolecular degradation, producing still complex organics other than
203 volatile acids, might prevent the inhibition of the bacterial activity and had an effect on the stabilization of
204 the electrochemical signal of the whole MFCs.

205 Interestingly, the maximum potential decreased cycle by cycle, in parallel with the peak of COD at the
206 beginning of each cycle. The peak value of COD was over 3000 mg_{O₂} L⁻¹ in the first cycle, and decreased
207 progressively to less than 2000 mg_{O₂} L⁻¹ in cycle 5.

208 This condition might be due to the accumulation of mineralized substances in the liquid phase, such as
209 ammonium ions, and other nutrients, creating an inhibitory environment for microbial population and for
210 ARB. The MFC signal followed the general trend of the primary phases of biodegradation (hydrolysis and
211 fermentation). Progressively less organic matter was hydrolyzed from the solid phase (visible as decreasing
212 peaks of COD). Additionally, each cycle lasted for longer time as compared to the first one. Biodegradation
213 rate was visibly decreasing. All these effects were reflected in progressively decreasing peaks of the
214 electrical signals.

215 The MFCs potential trends were. Again, a mirror of an increase of limiting or toxic conditions the overall
216 microbial population in the bioreactor. The MFCs potential trends were, again, a mirror of an increase of
217 limiting (or toxic) conditions for the overall microbial population in the bioreactor. In this case the electrical
218 signals were indicating a progressive increase of inhibition conditions. This application of MFCs in
219 bioreactors would be useful for monitoring the accumulation of potentially toxic metabolites of
220 biodegradation (ammonia, H₂S, Na⁺, etc.) on long-term operations of bioreactors [5, 23].



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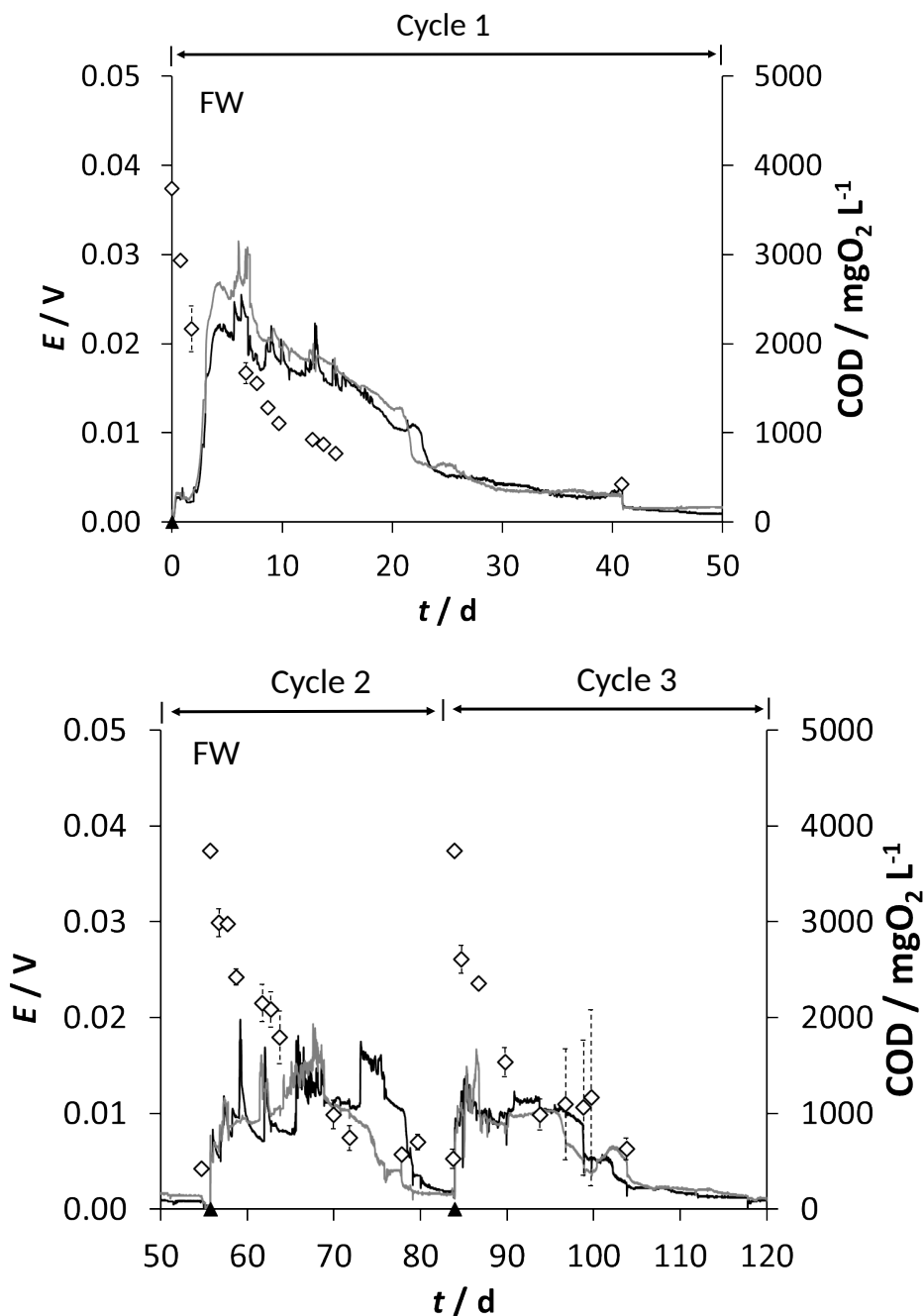
223

224 **Fig. 2** Graphics of potential trends of two KW-fed MFCs (solid lines, left y-axis) and COD evolution (\diamond , right y-axis) over 125 days of operation and six batch cycles. Standard deviation of COD values is reported with vertical
 225 bars. Triangles on the x-axis indicate the day of feed addition.
 226
 227

228 *3.3 Electrical signal trend in FW-fed MFCs*

229 Figure 3 Fig. 3 reports the evolution of cell potential over time for two MFCs fed by FW along with the
 230 soluble COD degradation. pH was stably in the range 7.5 – 7.9 in all cycles. Both MFCs started to generate
 231 electric signal from day 2, with cell potential that rapidly increased up to a maximum of 30 mV and then
 232 slowly decreased from day 7 to day 25. At the same time, COD spiked to nearly 3700 mgO₂ L⁻¹ and
 233 continuously decreased to 400 mgO₂ L⁻¹. During the first cycle, 88% of COD removal was achieved in
 234 around 40 days. Electric signal and COD varied with a really similar trend along the first cycle.
 235

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239 **Fig. 3** Graphics of potential trends of two FW-fed MFCs (solid lines, left y-axis) and COD evolution (◇, right y-
 240 axis) over 125 days of operation and three batch cycles. Standard deviation of COD values is reported with
 241 vertical bars. Triangles on the x-axis indicate the day of feed addition.
 242

243 At day 55, a new dose of FW was added. Upon addition, the potential firstly rapidly increased up to 10 mV,
 244 then maintained in the range between 10 mV and 20 mV for at least 25 days. The first cycle of operation was
 245 the more efficient in term of electrochemical performance and COD degradation, and different behavior was
 246 observed in the following cycles. At day 84, a new dose of FW was added and evolution of potential and
 247 COD similar than in the second cycle were achieved.

248 In cycle 2 and 3, COD removal happened with nearly a double rate,, in parallel with cell potentials. The peak
 249 values of COD were very similar at each cycle. Differently from CW and KW, peak of COD didn't

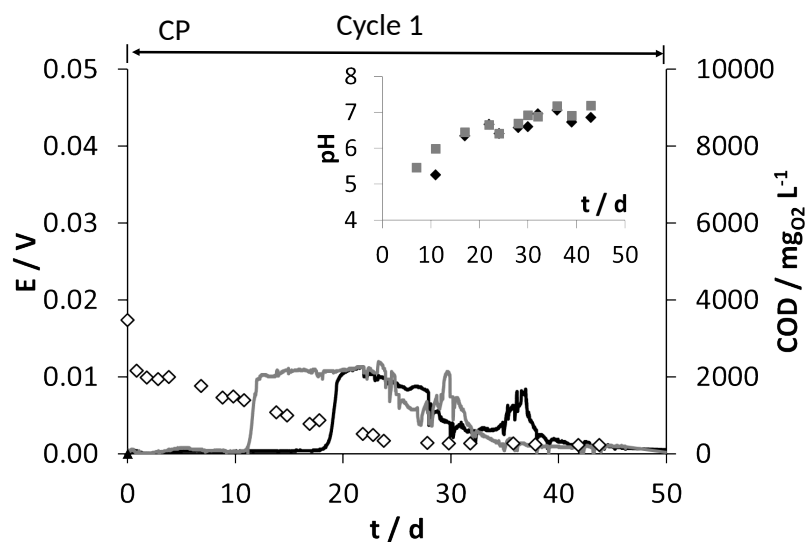
250 correspond to peaks of cell potential, especially at cycles 2 and 3. This is likely to be due to the recalcitrants
251 hydrolyzed from complex proteins of FW (Table 1). Contrarily, cell potentials lasted for longer time at
252 relatively high values, as compared to COD values.

253 In this case, MFCs can be thought as monitors of the presence of long-term biodegradable fractions of the
254 organic matter.

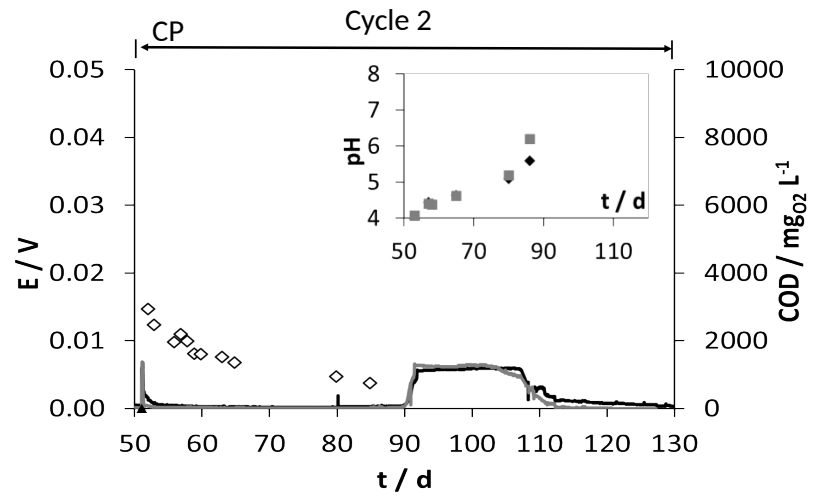
255 The production of uric acids and ammonia, could affected the pH stability on the electrodes, contrasting the
256 effect of acidic fermentation. It was recently proved that urea is quickly oxidized at the anode, inducing an
257 increase of the electric output in single chamber MFCs and pH increase over 9 due to ammonia [24]. A
258 variability in the concentration trend of ammonia, in contrast with acidic components, including volatile fatty
259 acids, could cause the signal instability in the MFCs fed with fish waste. Furthermore, pH variability and the
260 accumulation of less degradable byproducts in time could stressed the microbial communities on both the
261 electrodes (bioanode and biocathode), globally lowering the MFC performances and making also instable the
262 COD degradation (see error bar in Fig. 3). In fact, the pH of those MFCs, measured periodically close to the
263 anode in the bulk solution, was in the range of 7-8, and never decrease below 7 since the first days, while in
264 the case of CW and KW, with a lower protein content (Table 1) the pH was around 6.7 ± 0.2 during the
265 whole experimentation time.

266 3.4 Electrical signal trend in CP-fed MFCs

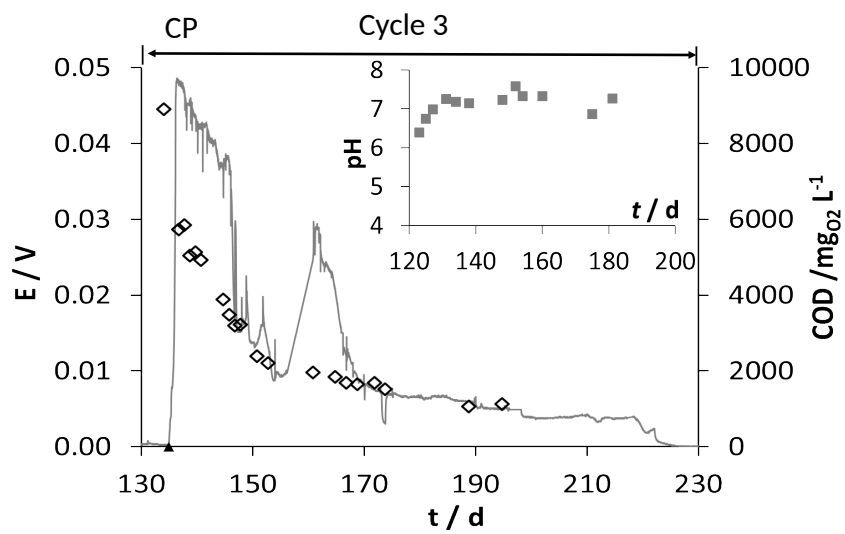
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278 **Fig. Figure 4 reports the evolution of potential over time for two MFCs fed by CP along with the degradation of**
 279 **the soluble COD. In the inset of Figure 4**

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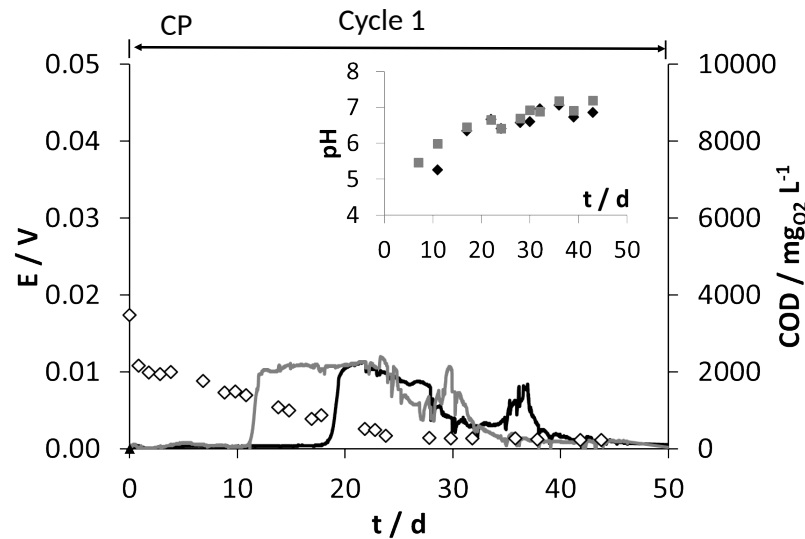
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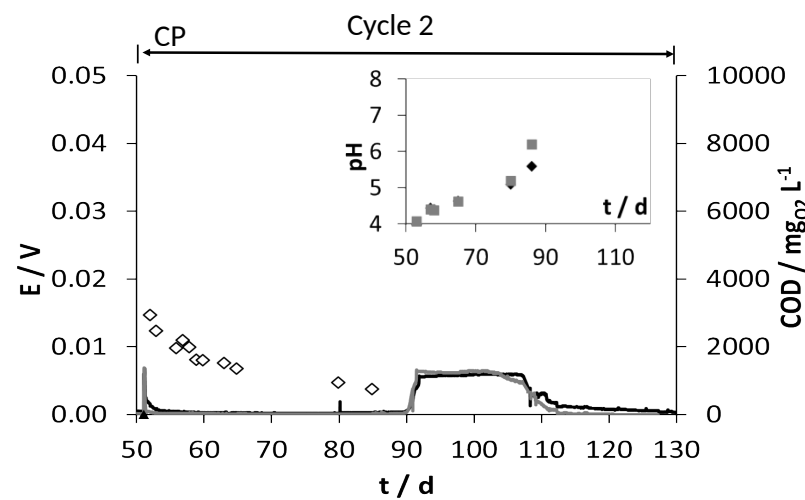
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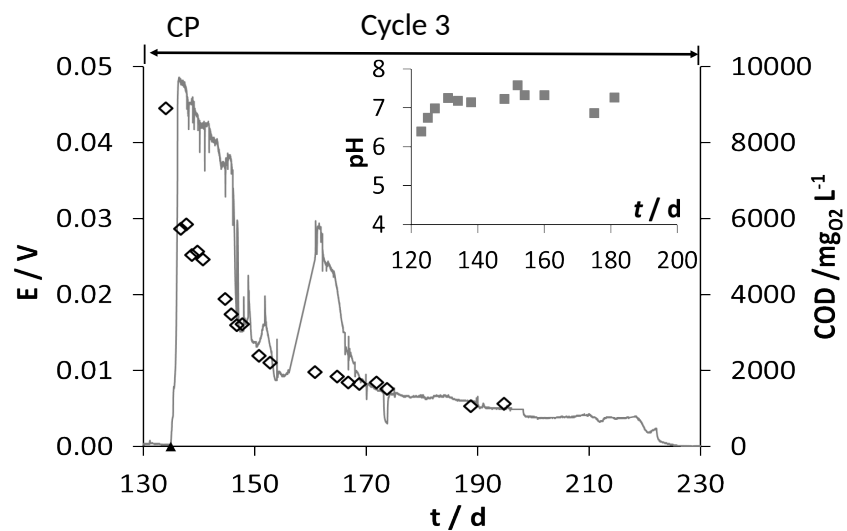
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290 Fig. the variation of pH in the liquid phase of the MFCs bioreactor is shown. In the first two cycles, acidic
291 conditions (pH 4 – 5) were established in the bioreactor and the pH raised over 6.5 only after 40 days.
292 Alkalinity was evidently insufficient to buffer the acidity of CP (rich in citric acid and other organic acids).



293 COD decreased from 3400 mg_{O2}/l to 200 mg_{O2} L⁻¹ along over 45 days of operation of cycle 1 and nearly 50
294 days in cycle 2.

295 At cycle 3, a buffer medium (potassium bicarbonate, 5 g L⁻¹) was added to the bioreactor to equilibrate the
296 pH, which remained stable in the range 7 – 7.8 along cycle 3. COD peaked to more than double (over 8000
297 mg_{O2} L⁻¹), due to more efficient microbial hydrolysis and fermentation of the solids accumulated in the
298 bioreactor.

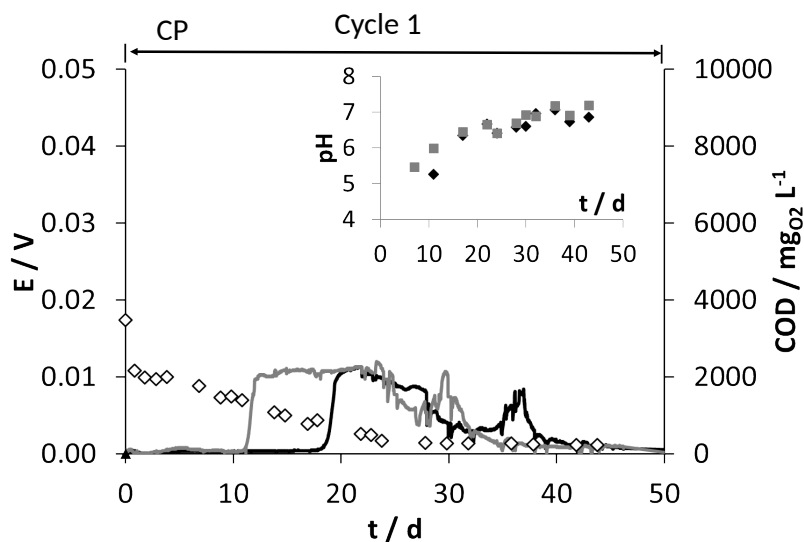
299 The MFCs did not show any current generation until day 11 and 18. One of the MFCs started to produce
300 power at day 11 and the other MFCs at day 18. The rapid increase of cell potential was associated with the
301 increase of pH in the liquid phase over 6.5. After the rapid increase up to 10 mV, the potential remained
302 stable for about 10 days then started to slowly decay, along with COD. This happened identically during
303 cycle 2. pH acidic conditions inhibited in perfect parallel the overall biodegradation rate and the activity of
304 exoelectrogenic bacteria. The absence of electric signal from ARB was accomplished by an evidently slow-
305 rate biodegradation, due to inhibited microbial activity.

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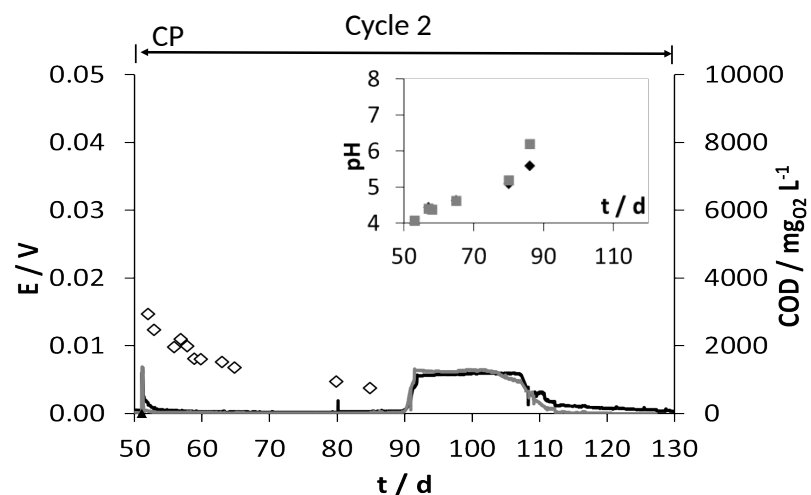
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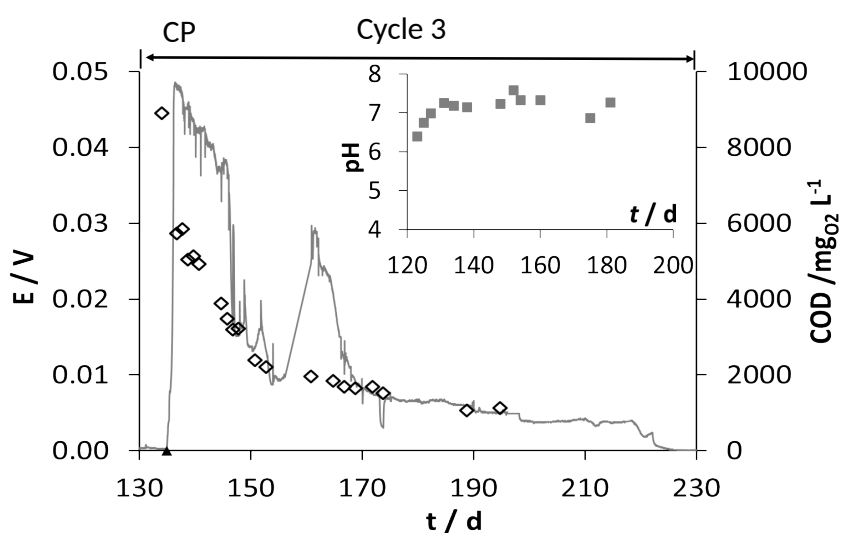
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317 **Fig. 4** Graphics of potential trends of two CP-fed MFCs (solid lines, left y-axis) and COD evolution (\diamond , right y-axis) over 230 days of operation and three batch cycles. Standard deviation of COD values is reported with vertical bars. Insets report the trend of pH over time.

320 However, the absence of electric signal was not accomplished by absence of microbial activity inside the
 321 MFC. As a matter of fact, COD decreased from 3400 $\text{mg}_{\text{O}_2}/\text{l}$ to 200 $\text{mg}_{\text{O}_2}/\text{l}$ during all the 45 days of
 322 operation of the first cycle. When a new dose of CP was added at day 51, the bulk pH lowered below 4.5. As
 323 long as the pH stayed lower than 6.5, apart from an initial spike, no electrical output was obtained from the
 324 MFCs. After day 90, the potential rapidly increased up to 7 mV and fixed there for 20 days then dropped to
 325 negligible values. Fermentative anaerobic degradation of the substrate took place, but electricity generation
 326 was initially inhibited.

327 Since the electric signal is severely affected by the pH, the increase of the potential is a clear indication that
 328 optimal condition of pH has been achieved inside the MFC. The kinetics of degradation is slightly lower in
 329 the second cycle and this is likely due to variation of non-buffered pH.

330 When stable pH was guaranteed by equilibrating CP acidity, cell potential trends followed exactly the trend
 331 of COD consumption in the liquid phase. These results highlighted another aspect of the chemical equilibria
 332 in anaerobic biodegradation environments, which can be efficiently correlated to the trend of the electrical

333 signal produced by ARB, living in the same environment. pH-related inhibition of microbial activity reflects
334 very promptly in drops of MFCs cell potentials.

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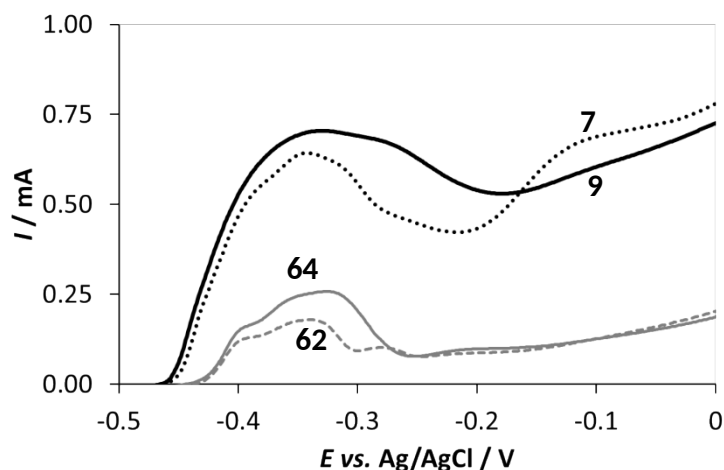
336 3.5 Electrode polarization curves

337 Polarization curves recorded on anodes and cathodes of a FW-fed MFC at different days of operation are
338 reported in Fig. . Polarizations recorded at day 7 and 9 refer to cycle 1, polarization at day 62 and 64 refers to
339 cycle 2. Anodes exhibited a peak around -0.35 V, which can be related to exogenous redox mediator and
340 which position did not change significantly with the time. This potential range can be typically related to the
341 oxidation of short-chain carboxylates, like volatile fatty acids [25]. The current delivered by the anode is
342 about 0.5 mA lower in cycle 2. Similarly, polarization of cathodes exhibited considerably higher currents in
343 cycle 1 than in cycle 2.

344 This indicates, first of all, that the cathode was predominantly characterized by microbially-catalyzed
345 reduction reactions. As reported in previous works for single-chamber MFCs, bio-cathodic instead of abiotic
346 mechanisms drive oxygen reduction reactions [16, 26]. This is an important aspect to consider, in the case
347 that cell potential has to be used as indicator of microbial consortia activity in a bioreactor. To maximize the
348 MFC system response to inhibitory effects, due to chemical imbalances in the liquid medium of
349 biodegradation environments, single chamber MFCs might be the ideal solution. A double chamber
350 architecture with abiotic cathode would be less sensible to inhibitory conditions for microbes.

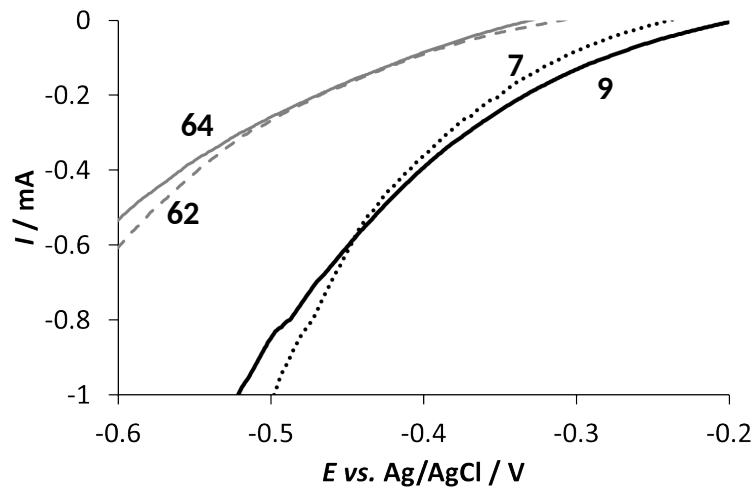
351 FW biodegradation started showing instability at cycle 3. Error bars reported for SCOD measured in two
352 MFCs indicate a condition of partial inhibition of the system. In biodegradation of protein-rich organic
353 materials (see Table 1 for FW composition), it might typically be related to accumulations of ammonia,
354 hydrogen sulfide or other toxic metabolites [24, 27].

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360 **Fig. 5 Polarization curves of anodes (a) and cathodes (b) recorded at different days in a FW-fed MFC. Numbers**
 361 **near the curves indicate the day of experiments.**

362 The considerable decrease of polarization currents from both bioanodes and biocathodes, already registered
 363 during cycle 2 (days 62 – 64, Figure 5), can be considered as an early-warning for inhibiting conditions for
 364 the whole bioreactor environment. Future experiments should more deeply focus on this aspect. Polarization
 365 of electrodes, as mirror of both anodic and cathodic microbial communities, could be studied as early-
 366 warning sensors for inhibiting conditions in anaerobic biodegradation environments.

367

368 *3.6. Just monitoring or influencing the biodegradation process of COD removal?*

369 kinetics of the various cycles of each organic substrates is reported in Table 2 **Error! Reference source not**
 370 **found..** Overall there was a more rapid kinetic of the first days and the first cycle. This points to a
 371 component of degradation due to an aerobic metabolism which preceded the formation of anaerobic biofilm
 372 on the anode and on the cathode.. The trend of kinetics is slower in subsequent cycles for all matrices, while
 373 coming practically going to 100% in the case of the milk and near 90% of COD in all the other cases. As
 374 expected, the quickest degradable substrate was CW and the less one was food wastes (the most complex)
 375 KW. For KW, COD removal was never higher than 85% due to the significant presence of recalcitrant
 376 fractions (e.g. fats and fibers).

377 For sake of comparison, in Table 2 the COD removal and CE are listed for the four different substrates.

378 The results showed that the COD removal ranged from 61.4% to 98.77% and CE ranged from 0.76 to 9.9%
 379 respectively. The so low CE achieved is consequence of the un-optimized anode electrode surface/volume
 380 ratio, as the anode occupied a marginal part of the cell volume. The maximum COD removal (98.77) was
 381 obtained for the CW substrate (cycle 6) and the maximum CE (9.91) was determined for the KW substrate
 382 (cycle 2).

383 This indicates that biodegradation were negligibly influenced by the MFC process. In future applications of
 384 MFCs as sensors for monitoring biodegradation process, electrode surface/reactor volume ratio might be
 385 even scaled down, to monitor specific environments.

386 On the contrary, enlarging the electrodes and improving their surface/volume ratio using a different
 387 geometry, will allow to reach and enhance the performance of the MFC process in degrading solid phase
 388 organics with respect the current literature [14].
 389 Finally, the optimization of the electrode surface finishing [29] and the use of different materials such as
 390 stainless steel [28] would address other possible needs for an useful application in both cases: monitoring or
 391 influencing the biodegradation process.

392

393 Table 2 Coulombic efficiency and COD removal during fed-batch MFC operation.

	Sodium acetate (control)		Cheese whey		Kitchen waste		Fish waste		Citrus pulp	
	CE (%)	sCOD removal (%)	CE (%)	sCOD removal (%)	CE (%)	sCOD removal (%)	CE (%)	sCOD removal (%)	CE (%)	sCOD removal (%)
Cycle 1	22.4	98.1	0.76	95.1	3.87	84.37	7.77	88.69	3.13	93.52
Cycle 2	21.3	95.9	2.02	96.9	9.91	64.25	4.37	85.94	2.74	89.70
Cycle 3	23.2	97.4	1.87	95.8	9.40	61.40	3.53	89.20	-	-

394

395 4. Conclusions

396 Four sets of membraneless single-chamber Microbial Fuel Cells were operated in duplicate and in parallel
 397 over more than 100 days, inoculated with anaerobic sludge of a biogas production plant and cyclically fed
 398 with the following different organic substrates: i) organic fraction of Kitchen waste (KW), ii) Cheese whey
 399 from dairy industries (CW), iii) residues of fish previously processed to recover oils (FW), iv) pulp waste
 400 from citrus juice production (CP).

401 All MFCs were able to produce electric signals that varied in intensity as function of the chemical equilibria
 402 in the liquid phase of the bioreactor and the biodegradation rates achieved.

403 Sudden upcoming inhibiting conditions for microbial community in the MFC bioreactor corresponded to
 404 sudden drops in cell potentials. Progressive decrease of biodegradation efficiency in successive batch cycles
 405 corresponded to diminishing peaks of cell potentials. pH drops below 6.5 inhibited both biodegradation and
 406 anodic exoelectrogenic activity. The presence of recalcitrant fractions gave a delay between soluble COD
 407 degradation trend and the electrical signal over time. Both anodic and cathodic polarization curves, gave
 408 lower currents corresponding to incoming inhibiting conditions in the bioreactors.

409 Biotic mechanisms driving cathodic reduction reactions can help in having more sensible responses from
 410 MFCs coupled to bioreactors. A deeper study should follow these preliminary indications in considering
 411 MFCs as sensor to monitor and control anaerobic biodegradation processes.

412

413 References

- 414 [1] Fava F, Totaro G, Diels L, Reis M, Duarte J, Carioca OB, et al. Biowaste biorefinery in Europe:
 415 opportunities and research & development needs. N Biotechnol 2013;32.
 416 doi:10.1016/j.nbt.2013.11.003.
- 417 [2] Agler MT, Wrenn B a., Zinder SH, Angenent LT. Waste to bioproduct conversion with undefined

- 418 mixed cultures: The carboxylate platform. *Trends Biotechnol* 2011;29:70–8.
419 doi:10.1016/j.tibtech.2010.11.006.
- 420 [3] Manzini E, Scaglia B, Schievano A, Adani F. Dark fermentation effectiveness as a key step for waste
421 biomass refineries: influence of organic matter macromolecular composition and bioavailability. *Int J*
422 *Energy Res* 2015;31:n/a – n/a. doi:10.1002/er.3347.
- 423 [4] Schievano A, D’Imporzano G, Malagutti L, Fragali E, Ruboni G, Adani F. Evaluating inhibition
424 conditions in high-solids anaerobic digestion of organic fraction of municipal solid waste. *Bioresour*
425 *Technol* 2010;101:5728–32.
- 426 [5] Chen Y, Cheng JJ, Creamer KS. Inhibition of anaerobic digestion process: A review. *Bioresour*
427 *Technol* 2008;99:4044–64. doi:10.1016/j.biortech.2007.01.057.
- 428 [6] Drake HL, Küsel K, Matthies C. Acetogenic Prokaryotes. *The Prokaryotes*, Berlin, Heidelberg:
429 Springer Berlin Heidelberg; 2013, p. 3–60. doi:10.1007/978-3-642-30141-4_61.
- 430 [7] Thauer RK, Hedderich R, Fischer R. Reactions and Enzymes Involved in Methanogenesis from CO₂
431 and H₂. *Methanogenesis*, Boston, MA: Springer US; 1993, p. 209–52. doi:10.1007/978-1-4615-2391-
432 8_5.
- 433 [8] Logan BE. *Nature Reviews Microbiology* 7, 375-381 (May 2009) | doi:10.1038/nrmicro2113.
- 434 [9] Rismani-Yazdi H, Christy AD, Carver SM, Yu Z, Dehority BA, Tuovinen OH. Effect of external
435 resistance on bacterial diversity and metabolism in cellulose-fed microbial fuel cells. *Bioresour*
436 *Technol* 2011;102:278–83. doi:10.1016/j.biortech.2010.05.012.
- 437 [10] Fornero JJ, Rosenbaum M, Angenent LT. Electric Power Generation from Municipal, Food, and
438 Animal Wastewaters Using Microbial Fuel Cells. *Electroanalysis* 2010;22:832–43.
439 doi:10.1002/elan.200980011.
- 440 [11] Clauwaert P, Rabaey K, Aelterman P, De Schampelaire L, Pham TH, Boeckx P, et al. Biological
441 denitrification in microbial fuel cells. *Environ Sci Technol* 2007;41:3354–60. doi:10.1021/es062580r.
- 442 [12] Rago L, Ruiz Y, Baeza JA, Guisasola A, Cortés P. Microbial community analysis in a long-term
443 membrane-less microbial electrolysis cell with hydrogen and methane production.
444 *Bioelectrochemistry* 2015;106:359–68. doi:10.1016/j.bioelechem.2015.06.003.
- 445 [13] Chae K-J, Choi M-J, Kim K-Y, Ajayi FF, Park W, Kim C-W, et al. Methanogenesis control by
446 employing various environmental stress conditions in two-chambered microbial fuel cells. *Bioresour*
447 *Technol* 2010;101:5350–7. doi:10.1016/j.biortech.2010.02.035.
- 448 [14] Mohan SV, Chandrasekhar K. Solid phase microbial fuel cell (SMFC) for harnessing bioelectricity
449 from composite food waste fermentation: Influence of electrode assembly and buffering capacity.
450 *Bioresour Technol* 2011;102:7077–85. doi:10.1016/j.biortech.2011.04.039.
- 451 [15] Goud RK, Babu PS, Mohan SV. Canteen based composite food waste as potential anodic fuel for
452 bioelectricity generation in single chambered microbial fuel cell (MFC): Bio-electrochemical
453 evaluation under increasing substrate loading condition. *Int J Hydrogen Energy* 2011;36:6210–8.
454 doi:10.1016/j.ijhydene.2011.02.056.

- 455 [16] Guerrini E, Cristiani P, Trasatti SPM. Relation of anodic and cathodic performance to pH variations
456 in membraneless microbial fuel cells. *Int J Hydrogen Energy* 2013;38:345–53.
457 doi:10.1016/j.ijhydene.2012.10.001.
- 458 [17] Guerrini E, Grattieri M, Faggianelli A, Cristiani P, Trasatti S. PTFE effect on the electrocatalysis of
459 the oxygen reduction reaction in membraneless microbial fuel cells. *Bioelectrochemistry* 2015.
460 doi:10.1016/j.bioelechem.2015.05.008.
- 461 [18] Schievano A, Pognani M, D'Imporzano G, Adani F. Predicting anaerobic biogasification potential of
462 ingestates and digestates of a full-scale biogas plant using chemical and biological parameters.
463 *Bioresour Technol* 2008;99:8112–7.
- 464 [19] M. Santini, M. Guilizzoni, M. Lorenzi, P. Atanassov, E. Marsili, S. Fest-Santini, P. Cristiani, C.
465 Santoro. Three-Dimensional X-ray Micro Computed Tomography Of Carbonates And Biofilm On
466 Operated Cathode In Single Chamber Microbial Fuel Cell. *Biointerphases*, 10 (3):031009 (2015);
467 <http://dx.doi.org/10.1116/1.4930239>
- 468 [20] Gajda I, Greenman J, Melhuish C, Santoro C, Li B, Cristiani P, et al. Water formation at the cathode
469 and sodium recovery using Microbial Fuel Cells (MFCs). *Sustain Energy Technol Assessments*
470 2014;7:187–94.
- 471 [21] Frascari D, Cappelletti M, Mendes JDS, Alberini A, Scimonelli F, Manfreda C, et al. A kinetic study
472 of biohydrogen production from glucose, molasses and cheese whey by suspended and attached cells
473 of *Thermotoga neapolitana*. *Bioresour Technol* 2013;147:553–61.
474 doi:10.1016/j.biortech.2013.08.047.
- 475 [22] Davila-Vazquez G, Cota-Navarro CB, Rosales-Colunga LM, de León-Rodríguez A, Razo-Flores E.
476 Continuous biohydrogen production using cheese whey: Improving the hydrogen production rate. *Int*
477 *J Hydrogen Energy* 2009;34:4296–304. doi:10.1016/j.ijhydene.2009.02.063.
- 478 [23] Hartmann H, Ahring BK. Anaerobic digestion of the organic fraction of municipal solid waste:
479 Influence of co-digestion with manure. *Water Res* 2005;39:1543–52.
480 doi:10.1016/j.watres.2005.02.001.
- 481 [24] C. Santoro, I. Ieropoulos, J. Greenman, P. Cristiani, T. Vadas, A. Mackay, B. Li. Current Generation
482 in Membraneless Single Chamber Microbial Fuel Cells (MFCs) Treating Urine. *Journal of Power*
483 *Sources* 238 (2013) 190-196. <http://dx.doi.org/10.1016/j.jpowsour.2013.03.095>
- 484 [25] .Rabaey K, Rozendal R a. Microbial electrosynthesis - revisiting the electrical route for microbial
485 production. *Nat Rev Microbiol* 2010;8:706–16. doi:10.1038/nrmicro2422.
- 486 [26] Guerrini E, Grattieri M, Trasatti SP, Bestetti M, Cristiani P. Performance explorations of single
487 chamber microbial fuel cells by using various microelectrodes applied to biocathodes. *Int J Hydrogen*
488 *Energy* 2014;39:21837–46. doi:10.1016/j.ijhydene.2014.06.132.
- 489 [27] Kim DH, Oh SE. Continuous high-solids anaerobic co-digestion of organic solid wastes under
490 mesophilic conditions. *Waste Manag* 2011;31:1943–8. doi:10.1016/j.wasman.2011.05.007.

- 491 [28] E. Guerrini, P. Cristiani, M. Grattieri, C. Santoro, B. Li, S. Trasatti. Electrochemical behavior of
492 stainless steel anodes in microbial fuel cells. *J. Electrochem. Soc.* 2014 161(3): H62-H67 2014.
493 <http://dx.doi.org/10.1149/2.096401jes> . *Journal of Electrochemical Society.* volume 161, issue 3,
494 H62-H67.
- 495 [29] G. Papaharalabos, J. Greenman, C. Melhuish, I. Ieropoulos, P. Cristiani, C. Santoro, B. Li. Increased
496 power output from micro porous layer (MPL) cathode microbial fuel cells (MFC). *International*
497 *Journal of Hydrogen Energy* 38 (2013) 11552-11558.
498 <http://dx.doi.org/10.1016/j.ijhydene.2013.05.138>
499

Highlights

- Four different component of solid organic wastes were investigated in membranelessMFCs
- Cell potential trends varied in function of different waste components in the bioreactors
- Cell potential trends varied as function of the substrate, the COD and the feeding cycle
- The pH in the anodic chamber significantly affected the electric output
- Results call for exploration of MFCs as sensors for fermentation/biodegradation processes