

Removal of metallic elements

from real wastewater using zebra mussel bio-filtration process

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ABSTRACT

The metallic element pollution is a serious environmental problem but still unsolved since these contaminants are released mainly by human activity, reaching all the environmental compartments. Traditional wastewater treatment plants are very efficient in removing metallic elements only when their concentration is in the order of mg/L, but are not able to remove them until $\mu\text{g/L}$, as it would be needed to cope with the water quality standards in low flow receptors. Therefore, the aim of our study was to evaluate the potential removal of some recalcitrant metallic elements to the classical treatments, by the natural process of bio-filtration performed by the invasive zebra mussel (*Dreissena polymorpha*). For this purpose we built a pilot-plant at the Milano-Nosedo wastewater

24 treatment plant, where we placed about 40,000 *D. polymorpha* specimens appointed to the
25 wastewater bio-filtration. The metallic element removal due to zebra mussel activity was evaluated
26 in the treated wastewater with a plasma optical emission spectrometry (ICP-OES). Data obtained in
27 these experiments showed an encouraging metallic element removal due to *D. polymorpha* activity;
28 in particular, the total abatement (100%) of Cr after one day of bio-filtration exposure is
29 remarkable. Therefore, this study encourages further research related with the use of bivalves as a
30 new tool for the wastewater depuration process; in this regard, the contaminated mollusks used in
31 the bio-filtration could be incinerated or stored in special landfills, as is also the case of traditional
32 sewage sludge.

33
34 Keywords: zebra mussel; bio-filtration; wastewater treatment; metallic elements

35 ABBREVIATIONS: metallic elements (MEs); wastewater treatment plants (WWTPs)

36

37 1. INTRODUCTION

38 Metallic element (ME) pollution is a major global concern since these inorganic contaminants are
39 continuously released into the environment by human activities [1,2]. The ability of these
40 compounds to be accumulated in the organisms and to trig the onset of diseases and disorders
41 makes MEs very dangerous for many organisms, including humans, at very low concentrations [3].
42 In particular, the water pollution due to MEs is a serious and partially unsolved issue because the
43 removal needed to reach acceptable concentrations in the receiving waters (in the order of $\mu\text{g/L}$) is
44 well over the efficiency of wastewater treatment plants (WWTPs), normally reported as between 40
45 and 90% [4]. Because of this reason, alternative methods for the ME abatement have been identified
46 in order to be complementarily applied to traditional wastewater treatment processes. However,
47 most of these techniques, such as precipitation/neutralization, ion exchange, membrane separation,
48 reverse osmosis, electrodialysis and activated carbon adsorption [5,6,7] have high costs for the
49 regeneration of resins or activated carbons and/or for the disposal of chemical sludge or

50 concentrates [8]. Therefore, the attention of the scientific community need to be focused on the
51 development of natural methods which were more eco-sustainable and, possibly, less expensive. In
52 this regard, biosorption is a possible natural method for ME elimination; this term defines the
53 passive pollutant uptake from an aqueous solution by a dead or non-growing microbial biomass
54 [9,10]. Although this treatment has the advantage to not undergo inhibition due to the pollutants'
55 toxicity, the early biomass saturation by adsorbing contaminants represents an important limitation
56 for further exploitation of this process [7]. In addition to the biosorption, the bioaccumulation
57 process of many organic and inorganic contaminants by different aquatic microorganisms such as
58 fungi, algae, bacteria and yeast [11,12] may be considered. In particular, bioaccumulation due to
59 microorganisms living on aquatic macrophyte tissues is correlated with ME removal in constructed
60 wetlands. This methodology is certainly the most used natural system of wastewater treatment,
61 which couples accumulation in microbial biomass and in macrophytes such as *Phragmites australis*,
62 *Eichhornia crassipes* and *Lemna* spp. [13,14,15,16]. This alternative method, in addition to the
63 removal of MEs, also reduces organic matter and nutrients from wastewater [16]. Despite the
64 existence of these eco-friendly methodologies, in recent years, further studies have been conducted
65 in order to identify new methods for natural purification of waters from some recalcitrant pollutants.
66 In this regard, it is of great interest the research carried out by Ledda and co-workers [17] aimed at
67 assessing how small breeding of Mediterranean sponges *Ircinia variabilis* and *Agelas oroides* could
68 remove some contaminants from marine waters. In the same way, the use of other filtering
69 organisms can be interesting for the improvement of waters quality. In this context, the freshwater
70 bivalve *Dreissena polymorpha* has some characteristics that would make it suitable for the above
71 mentioned purpose: an enormous filtering capacity, ranging from 5 to 400 mL/bivalve/h [18,19], a
72 high population density, with more than 700,000 individuals/m² [20], and the ability to produce
73 faeces and pseudofaeces where many contaminants are adsorbed. In fact, these two *D. polymorpha*
74 waste products, being settleable [21], could easily remove from the water column the bounded
75 pollutants (as MEs). Moreover, taking into account the indirect ability of bivalves to bioaccumulate

76 many environmental contaminants, including MEs [22], we can point out the potential of *D.*
77 *polymorpha* to this purpose [23,24,25,26]. In this regard, a study conducted in 1983 by Piesik [27]
78 highlighted how *D. polymorpha* is able to remove nutrients from eutrophic waters and a subsequent
79 research confirmed the potential of *D. polymorpha* in the reduction of algal density [28]. In the last
80 two decades, several other studies have demonstrated the filtering capacity of this bivalve, whose
81 breeding could be developed for an alternative treatment of polluted freshwaters [25,29,30,31]. In
82 this regard, a recent study conducted by Binelli and co-workers [21] showed the ability of this
83 mollusk to remove different types of emerging contaminants, such as pharmaceuticals and drugs of
84 abuse, from wastewaters. Nevertheless, it is important to take into account that *D. polymorpha* is
85 considered an invasive alien species all over Europe and the United States, even if this mollusk was
86 present in Europe before the last glaciation [32] and was then bounded in some basins of Eastern
87 Europe in the post-glacial period until the 18th century [33]. The human activity has then favored
88 the distribution of *D. polymorpha* all over its original European areal; in Italy, for example, this
89 bivalve has first been found in 1973 [34] and its presence in the Italian inland waters has been
90 confirmed by subsequent studies [35,36,37]. Therefore, the idea of using this invasive species for
91 anthropic purposes (bio-filtration, human food, animal feed, fertilizer and biogas) [29] would be of
92 huge interest, especially in the economic sphere. On the basis of these above mentioned studies on
93 *D. polymorpha*, we assessed the efficiency of this bivalve as a new biological method as the last
94 step of wastewater treatment in a conventional WWTP. For this purpose, we built at the Milano-
95 Nosedo WWTP (Northern Italy) a pilot-plant in which 40,000 *D. polymorpha* specimens were
96 added in order to filtrate some types of wastewaters and we subsequently evaluated the abatement
97 of some MEs, such as Aluminum (Al), Chromium (Cr), Copper (Cu), Iron (Fe), Manganese (Mn),
98 Nickel (Ni) and Lead (Pb). This study is particularly innovative because, according to our
99 knowledge, for the first time, *D. polymorpha* has been used in a real civil WWTP for the removal of
100 some micropollutants. In fact, the few studies conducted using *D. polymorpha* as bio-filtering agent

101 mostly evaluated algal or organic matter removal, but not the abatement of emerging contaminants
102 (as previously reported in Binelli and co-workers) [21] or potentially toxic metals.

103

104 2. MATERIALS AND METHODS:

105 2.1 Pilot-plant construction and placement at the Milano-Nosedo WWTP

106 A scuba diver collected the bivalves from the Lake Maggiore and Lake Lugano, both located close
107 to the Italy-Switzerland border. Since it is well-known that *D. polymorpha* is a biofouling organism
108 [38], we placed approximately 40,000 specimens in an attachment tank in order to let them
109 naturally re-adhered to twenty Plexiglas[®] panels (size: 70x40 cm; Figure 1) via their *byssus* over a
110 period of two weeks. During this acclimatization period, the bivalves were kept in tap water and fed
111 with the green-blue alga *Spirulina* spp. The Plexiglas[®] panels were then placed into the pilot-plant
112 (Figure 1), a stainless steel tank with a volume of about 1000 L (L=154.0 cm, h=102.0 cm, w =80.5
113 cm), where were disposed following a zig-zag pathway (yellow line, Figure 2), in order to increase
114 both the surface and the contact time between the wastewater and each bivalve. In addition to the
115 steel tank, we installed a recirculation tank (Figure 1) with a volume of 200 L with a submerged
116 pump to allow a constant wastewater flow (3,500 L/h) into the pilot-plant. The recirculation tank
117 further increases the contact time between the wastewater and the filter-feeding bivalves placed into
118 the pilot-plant, as well as limits the efficiency of settling which would remove part of the
119 contaminants adsorbed on suspended solids. The pilot-plant can directly collect the effluent from
120 the canal placed between the sedimentation tanks and the sand filters of the Nosedo WWTP using a
121 submersible pump (0-5,000 L/h). The installation site of the pilot-plant allows to test a clarified
122 effluent and to avoid the risk that suspended solids can not only compromise the filtration capability
123 of bivalves but also cause the animal death due to gill occlusion. Moreover, the pilot-plant position
124 into the Nosedo WWTP guaranteed the lack of any possible accidental release of *D. polymorpha*
125 specimens into the surrounding environment because the sand filters and the following process of
126 disinfection with peracetic acid stop and kill any possible leaked organism.

127 2.2 Evaluation of *D. polymorpha* filtration ability

128 The preliminary tests designed to evaluate the filtering and purifying performance of *D.*
129 *polymorpha* have been described in detail by Binelli and co-workers [21]. In that study, the
130 following issues have been discussed: 1) the adaptation of *D. polymorpha* to wastewater; 2) the
131 estimation of *D. polymorpha* filtering efficiency; and 3) the analysis of *D. polymorpha* capacity in
132 the removal of a new class of environmental pollutants (PPCPs and illicit drugs).

133

134 2.3 Experimental design and samples collection

135 As previously described, an important point was the necessity to evaluate the removal efficiency of
136 *D. polymorpha* independently from any other settling process, which would remove the metals
137 adsorbed on suspended solids. The filtering action of *D. polymorpha* was first evaluated on the
138 effluent outflowing secondary settling. However, since this effluent had a very low COD (≈ 10
139 mg/L) and, consequently, a low suspended solid concentration (on which a relevant amount of MEs
140 is normally adsorbed) [39,40], the following tests were performed with three other different
141 wastewater mixtures, previously filtered through a 1 mm mesh bag filter to remove coarse matter.
142 This allowed us to evaluate the filtration efficiency of *D. polymorpha* on wastewater with polluting
143 load and different amounts of suspended particulate, also taking into account that this bivalve
144 selects particles for food with a diameter ranging between 15 and 40 μm [41]. The mixtures used in
145 the tests, in addition to 100% outlet, are the following: 25% inlet/75% outlet, 50% inlet/50% outlet
146 and 100% inlet (wastewater incoming at WWTP). The ME removal evidence from wastewater were
147 carried out through the measurement of their concentrations in the water samples taken from the
148 pilot-plant with bivalves inside; at the same time, control tests were conducted into the pilot-plant
149 without adhering animals. All tests were performed in triplicate. The ME removal progress was
150 monitored for 4 hours, by sampling the wastewaters every 30 min, which enabled to obtain the
151 removal slope for each MEs. We chose to evaluate the ME removal within 4 hours, taking into
152 account that the treated wastewaters remain in the Milano-Nosedo WWTP for about 24 h; thus, the

153 selected time seemed to be a fair compromise in view of integrating the conventional treatment with
154 limited dimensional requirements. To check the practicability of such assumptions, we carried out
155 further final tests in single for a period of 24 h, taking only two samples, one at the beginning and
156 one at the end of the tests. The tests were conducted with an initial flow rate corresponding to 3,500
157 L/h, which would imply 18 minute contact time, recirculating the effluent in the pilot-plant 84 times
158 to obtain an overall 24 h contact time. After each test, the entire pilot-plant was washed with tap
159 water, to avoid memory-effects related to the previous tests. For this reason, to minimize this
160 problem, as well as to decrease the bivalve stress, the test schedule started with the most diluted
161 waste (100% outlet) and gradually increased its concentration until 100% inlet. We monitored the
162 wastewater temperature both at the beginning and at the end of each tests in order to take into
163 account its possible interference with the filtration activity of zebra mussels. The wastewater
164 temperature within the pilot-plant during the spring season ranged from 14 to 24 °C, comparable
165 with the optimal values for *D. polymorpha* filtration activity (10-20 °C) [42]; we can thus exclude
166 any negative interference of temperature on the filtration-removal process. Samples were taken
167 from the pilot-plant at the selected times by the use of a 250 mL plastic bottles, acidified with 1% of
168 HNO₃ and stored at 4 °C at dark until analysis.

169

170 2.4 Evaluation of ME abatement

171 We evaluated the removal of some MEs relatively abundant in civil wastewaters: Aluminum (Al),
172 Chromium (Cr), Copper (Cu), Iron (Fe), Manganese (Mn), Nickel (Ni) and Lead (Pb). The samples,
173 taken from the pilot-plant, were treated according to the CNR IRSA 3010 method. Briefly, an
174 aliquot of each sample was transferred into a flask and heated up to 100 °C to remove turbidity.
175 After cooling, samples were brought back to the starting volume with distilled water. Samples were
176 analyzed in a plasma optical emission spectrometer (ICP-OES; OPTIMA 2100 DV, Perkin Elmer;
177 detection limits for each ME: Al 0.5 µg/L; Fe 0.2 µg/L; Mn 0.1 µg/L; Ni 0.5 µg/L; Pb 1.0 µg/L; Cu
178 0.5 µg/L; Cr 0.2 µg/L) equipped with ultrasonic nebulizer (CETAC Ultrasonic Nebulizer, model

179 U5000AT +). The ME concentrations were quantified by a calibration curve at two points, starting
180 from appropriate dilutions of mixed certificate standard (AccuStandard MES 16-1).

181

182 2.5 Statistical analyses

183 Data normality and homoscedasticity were verified using the Shapiro-Wilk and Levene's tests,
184 respectively. We performed a statistical comparison (SPSS 21 IBM software package) between tests
185 carried out with and without mussels in the pilot-plant, where the dependent variable is the ME
186 concentration in the wastewater and the fixed factors are the treatment and the exposure time. For
187 all these cases, we conducted the comparison using the two-way analysis of variance (two-way
188 ANOVA; * $p < 0.05$; ** $p < 0.01$).

189

190 3. RESULTS AND DISCUSSION

191 3.1 Evaluation of *D. polymorpha* filtering ability in the ME removal

192 The results obtained from the tests carried out with a 25% inlet/75% outlet mixture (Figure 3A, B,
193 C, D, E, F, G) showed a good removal performance by *D. polymorpha* due to the bio-filtration
194 effect, probably because of the suitable concentration of the suspended matter. The removals
195 obtained for each ME tested through the filtering activity of bivalves were always greater than those
196 related to the natural sedimentation evaluated in controls. In fact, for the majority of the analyzed
197 MEs, the contribution of the zebra mussel filtration was evident, since the difference between the
198 removal percentage with and without bivalves in the pilot-plant was statistically significant: Al
199 ($F=36.809$, $p < 0.01$); Fe ($F=62.686$, $p < 0.01$); Mn ($F=125.452$, $p < 0.01$); Ni ($F= 5.695$, $p < 0.05$); Pb
200 ($F=16.645$, $p < 0.01$); Cu ($F=6.220$, $p < 0.05$). In detail, observing the trends reported in Figure 3, it
201 has to be highlighted that the differences between the removal percentages measured at the end of
202 the tests reached the 30% for Fe and Pb, while for Al, Ni and Mn the removal was about 20-25%
203 higher than controls. Thus, in only 4 h, zebra mussels have been able to significantly decrease levels
204 of most of the tested MEs, even if the removal of Cu was only 8% higher than natural

205 sedimentation. On the other hand, the time selection to conduct the tests is crucial for the possible
206 engineering of the process that cannot be longer than few hours, since the entire cycle of the
207 wastewater treatment ends in about 24 h. Tests carried out by adding 50% of inlet to the WWTP
208 outlet (Figure 4A, B, C, D, E, F, G) showed a lower difference compared to control than the
209 previous tests, probably due to an excessive presence of suspended particulate matter that
210 determined a stress condition to the animals, which may require a longer time than 4 h to acclimate
211 and begin the filtering process. Moreover, we cannot exclude the possible presence of toxic
212 compounds into the inlet of WWTP that could have led to a further decrease in the filtration activity.
213 Despite these possible interfering processes, we found statistically significant difference between
214 tests carried out with bivalves in the pilot-plant and their respective controls for Al (F=68.587
215 $p<0.01$), Mn (F=38.710, $p<0.01$), Pb (F=26.183, $p<0.01$), Cu (F=22.861, $p<0.01$) and Cr (F=4.729,
216 $p<0.01$). In this regard, at the end of the test the removal was around 20-25%, comparable to the
217 results obtained for the mixture 25% inlet/75% outlet for Al, Mn, Pb and Cu, whilst for the other
218 tested metals it decreased dramatically. The fluctuating values obtained for Ni could be due to the
219 low concentration of this metal in the analyzed wastewater ($<10 \mu\text{g/L}$), taking into account the huge
220 variability of pollutant load in the inlet wastewaters. The role of the initial concentration of metallic
221 elements into the considered mixtures (Table 1), which depends on the WWTP inlet, must always
222 be considered when drawing conclusions in terms of percent removal: if these are very low, small
223 variations (which could also be partly due to analytical reasons) assume relevant percent weight. In
224 both the considered tests (25% inlet/75% outlet and 50% inlet/50% outlet mixtures) negative values
225 of sedimentation, comprised in a range of -5 and -10%, are observable; these values are likely to be
226 related to the coefficient of variation of the method used to perform the wastewater ME
227 quantification. These data do not appear to be random, because, except for the fluctuating values of
228 Ni (Fig. 4E), Mn and Pb showed null sedimentation values in both tests performed (Fig. 3D,F and
229 4D,F). This result can be reasonably related to the chemical speciation phenomenon because these
230 metals can probably be dissolved in water and not bounded to the particulate. Therefore, the

231 observed Mn and Pb removal process carried out by *D. polymorpha* could mainly be related to
232 bioaccumulation. Further studies are needed in order to deepen the knowledge about some of the
233 above-mentioned aspects, as also suggested by Camusso and co-workers [43]. In this regard, the
234 wastewater pH value, which influences the metal speciation, is kept constant in WWTPs and,
235 therefore, should not compromise the *D. polymorpha* purification activity. Finally, with regard to
236 the test with 100% inlet, there has been a serious decline in the bivalves' performance related to a
237 high mortality of the animals (data not shown). This result further confirms how an excessive
238 suspended particulate matter amount and the possible presence of toxic substances into the WWTP
239 can decrease the bivalves filtering capacity and even compromise their health status. However, this
240 aspect does not limit the possible engineering of this method, since it would be sufficient to control
241 the particulate matter of the wastewater, as suggested by Binelli and co-workers [21]. Moreover,
242 despite the suspended matter concentration represents a limiting factor of *D. polymorpha* filtering
243 capacity, it should be noted that the specimens used in this study are the same used in the
244 pharmaceuticals and illicit drugs removal process, described by Binelli and co-workers [21].
245 Despite an exposure to multiple pollutants, the bivalves' purifying ability is stable during the whole
246 experimental trial, representing a sure advantage in the use of this very resistant organism.
247 Furthermore, the data shown refer to the ME removal within the firsts 4 h of wastewaters exposure
248 to *D. polymorpha*, and that the bivalves' performance can be improved with increasing contact time
249 between mollusk and wastewater, as described below.

250

251 3.2 Time influence on the ME removal by *D. polymorpha*

252 Data obtained by the above-mentioned tests suggested that the contact time between wastewater and
253 the filter-feeding bivalves was probably one of most crucial parameters, affecting the extent of ME
254 removal from wastewater. As previously mentioned, although the increase of contact time could be
255 almost impossible at full scale, we decided to carry out tests 24 h long. On the basis of the results
256 obtained at 4 h, the 24 h tests were performed only on 25% inlet/75% outlet and 50% inlet/50%

257 outlet mixtures. For most of the MEs, the removal due to mussel filtration was about 70% with the
258 25% inlet/75% outlet mixture (Figure 5A). The natural sedimentation, at the same time, removed
259 50% of Cr and Fe and, surprisingly, only 10-25% of Cu, Mn and Pb (Figure 5A). Thus, zebra
260 mussels' filtration is able to increase the removal of Pb and Mn by about 60% with respect to the
261 settling effect in blanks. Notably, Cr removal appeared very interesting because of its high toxicity
262 for aquatic organisms [44,45]; in fact, *D. polymorpha* completely removed it in 24 h, while the
263 blank removal was only 50%. Therefore, contact time seems to affect significantly the extent of ME
264 removal by the filter-feeding bivalves, considering that at the end of the firsts 4 h the mean removal
265 was 20% higher with *D. polymorpha* than in the blank tests. This was also confirmed in the test
266 performed with the 50% inlet/50% outlet mixture (Figure 5B), where the ME removal due to *D.*
267 *polymorpha* was always over 70%. In particular, for Cu, Mn and Pb the net removal due to *D.*
268 *polymorpha* (calculated as the difference from the blank removal) was 50%, 70% and 60%,
269 respectively. At the same time, the high removal observed for Ni contradicts the results obtained in
270 the experimental data set. The 24 h tests, although only performed in single and therefore needing
271 further confirmation, provide first evidence that better ME removal performances may be obtained
272 by increasing the contact time between the bivalves and the feed. Further, the obtained data may
273 indicate that the bivalve could need a period of acclimation to the wastewater, especially if
274 characterized by a considerable amount of suspended particulate material, before starting the
275 filtration process.

276

277 4. FUTURE PERSPECTIVES

278 Due to the scarcity of scientific data regarding the use of *D. polymorpha* in the wastewater
279 treatment context, we faced many technical and logistical problems during our research, not
280 foreseeable during the experimental design drafting; in fact, the best performances of bio-filtration
281 were obtained with prolonged exposure times (24 h) and with moderate amounts of particulates.
282 Therefore, the ability of *D. polymorpha* to remove certain types of pollutants from pretreated

283 wastewater could suggest, in a possible future research or in an engineered scenario, the placement
284 of this filter-feeding bivalve as the last step of conventional WWTPs or to include it in other natural
285 systems, such as constructed wetlands or lagooning, where the hydraulic retention time is of one or
286 more days, and thus a longer contact time between wastewater and the bivalves is allowed.
287 Furthermore, in future studies, it would be interesting to investigate the ME removal mechanisms
288 and to monitorate the fate and presence of MEs in the bivalve soft tissues, shells, feces and
289 pseudofaeces.

290

291 5. CONCLUSIONS

292 This work, according to our knowledge, represents one of the very few studies concerning the
293 possibility to use bivalves in the wastewater treatment processes. The results appear to be very
294 encouraging, considering that the use of non-native species, such as *D. polymorpha*, for
295 anthropogenic purposes, could have interesting economic implications and represents an important
296 starting point for the alien species exploitation. In this regard, the prevention strategies regarding
297 the non-native and invasive species introduction determine complex social and ethical implications;
298 furthermore, while the procedures on how to respond to invasions have been delineated, their
299 application is still severely limited. Therefore, in the exclusive case of *D. polymorpha*, it may be
300 advantageous to exploit the potential of this bivalve, now almost present in all the Europe inland
301 waters. This will not certainly be an easy process; in fact, being *D. polymorpha* considered a serious
302 threat for the aquatic environment and a dangerous fouling agent of many industrial structures
303 [46,47], is poorly perceived by the scientific community as a valid filtering factor, despite the
304 presence of encouraging results in the depuration context [28,48,25,49,30,21]. In this regard, the
305 construction of appropriate facilities for bio-filtration, followed by further downstream treatment
306 aimed to contain bivalves accidentally leaked from the plant (such as the peracetic acid treatment
307 and sand filters) would avoid the problem related to fouling. The ideal condition would be to use
308 native bivalves, such as unionids; however these mollusks, besides being affected by a serious

309 population decline [50], have a parasite larval stage that would be disadvantageous for the
310 engineering of the bio-filtration process. Once contaminated by the filtration process, the bivalves
311 may then be dehydrated and stored in dedicated landfills or incinerated, as it is currently the case for
312 sewage sludges.

313

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319

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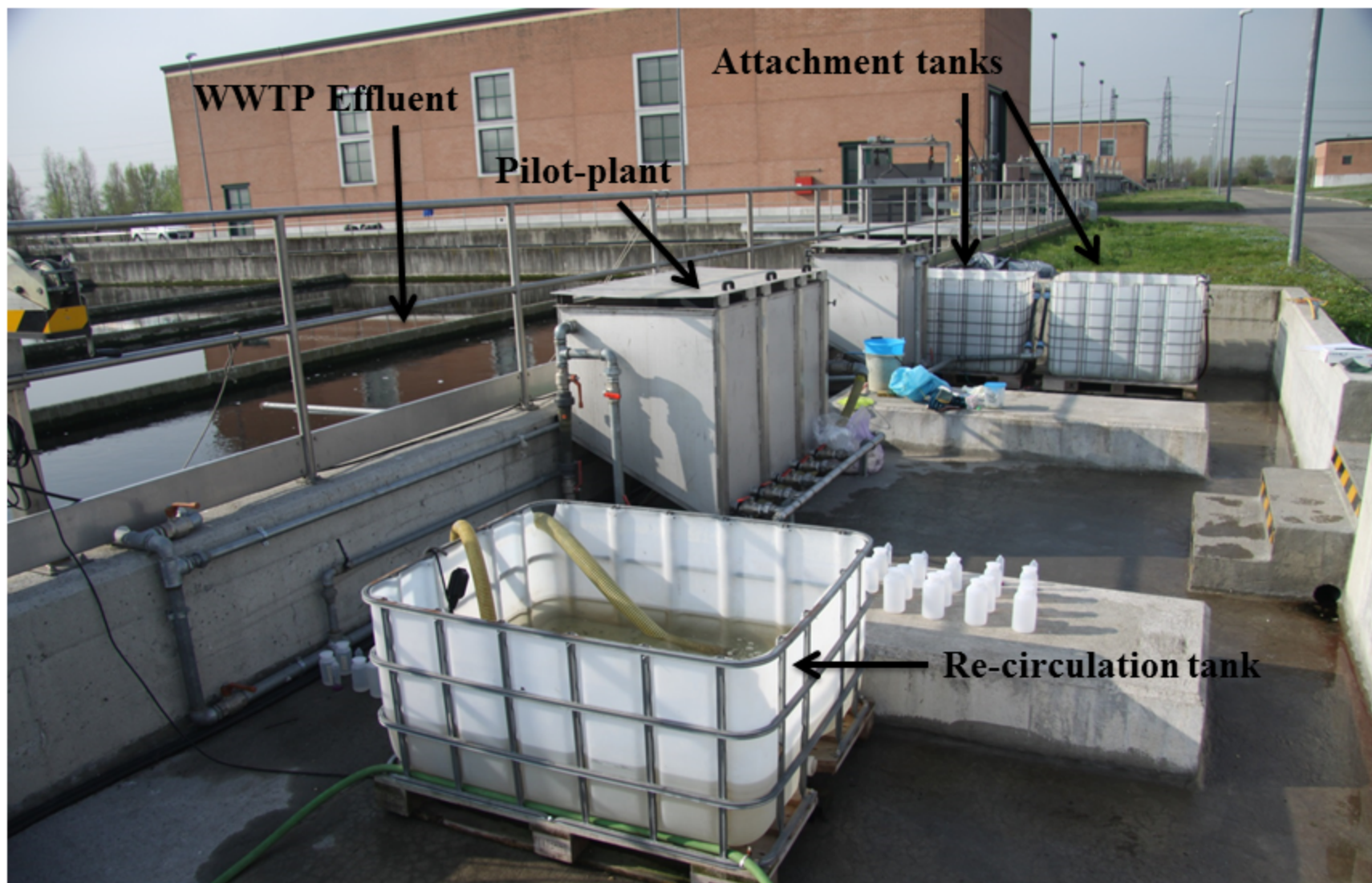


Fig. 1: Structure of the pilot-plant located at the Milano-Nosedo WWTP.

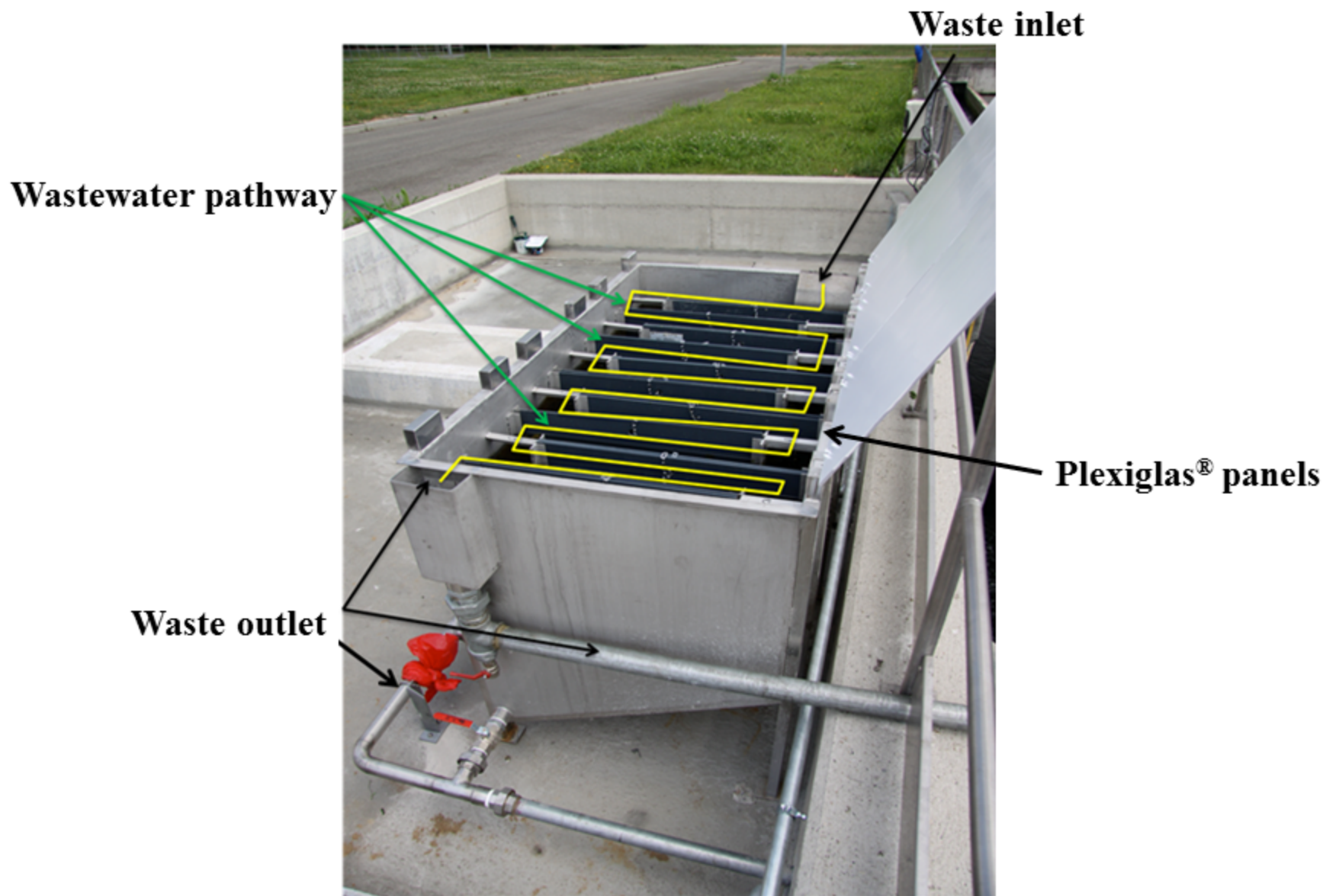


Fig.2: Plexiglas® panels placed into the pilot-plant. The yellow line indicates the zig-zag flow pathway of wastewater within the pilot-plant.

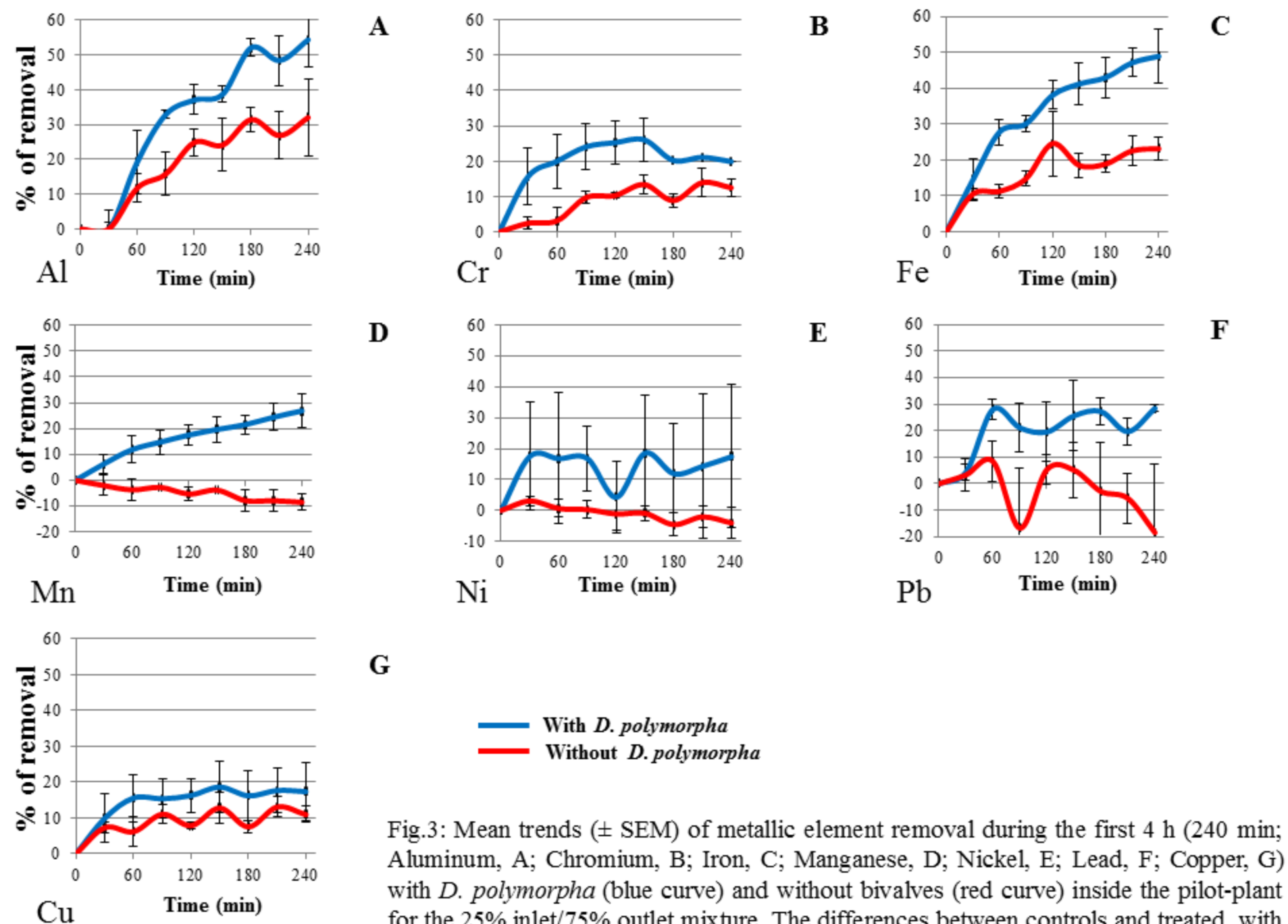


Fig.3: Mean trends (\pm SEM) of metallic element removal during the first 4 h (240 min; Aluminum, A; Chromium, B; Iron, C; Manganese, D; Nickel, E; Lead, F; Copper, G) with *D. polymorpha* (blue curve) and without bivalves (red curve) inside the pilot-plant for the 25% inlet/75% outlet mixture. The differences between controls and treated, with the exception of Chromium, were statistically significant (Two-way ANOVA).

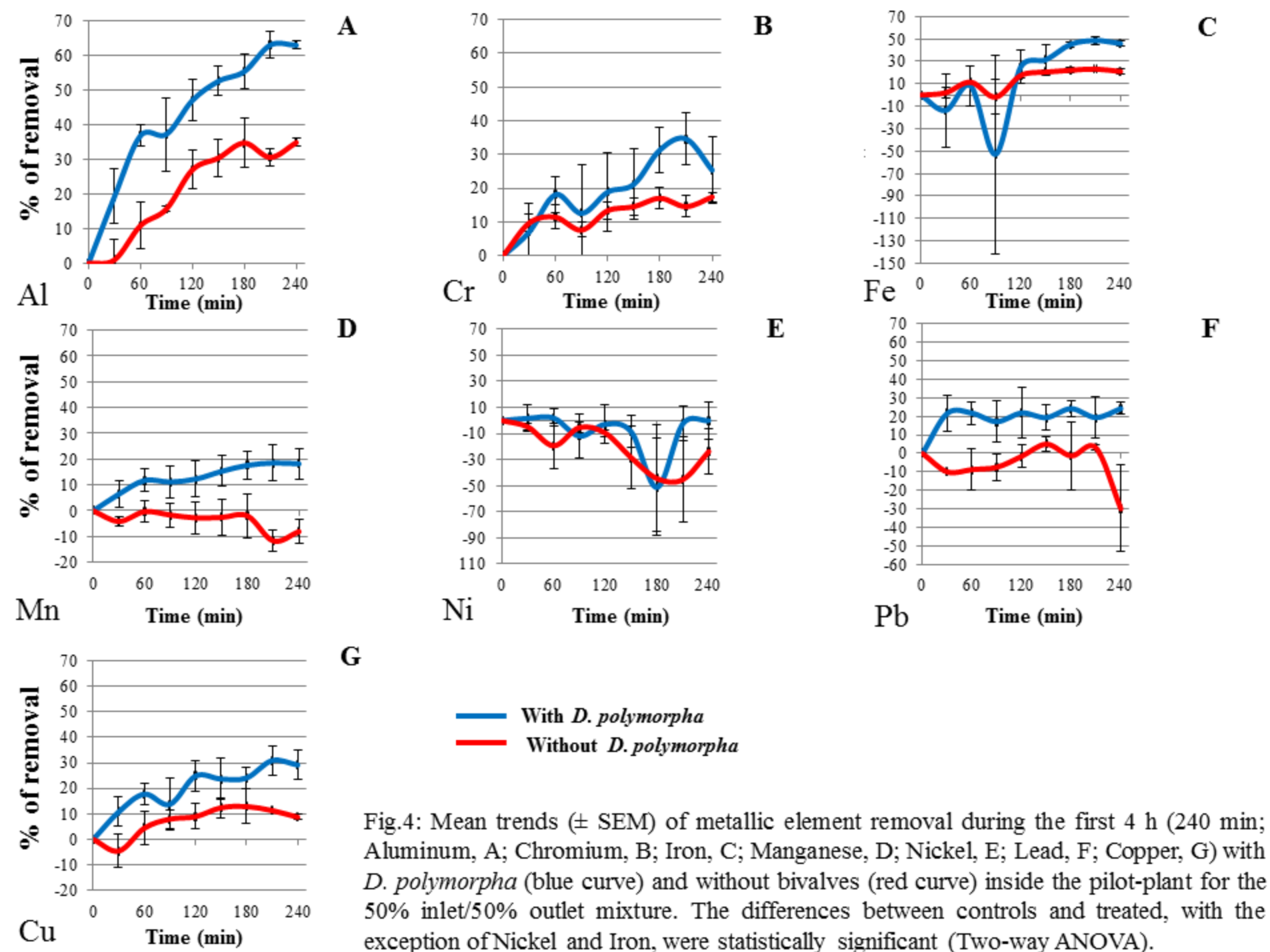
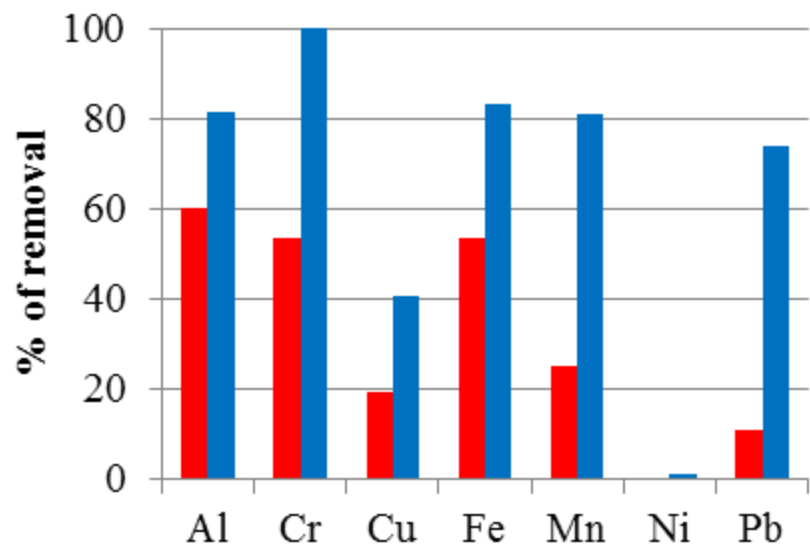
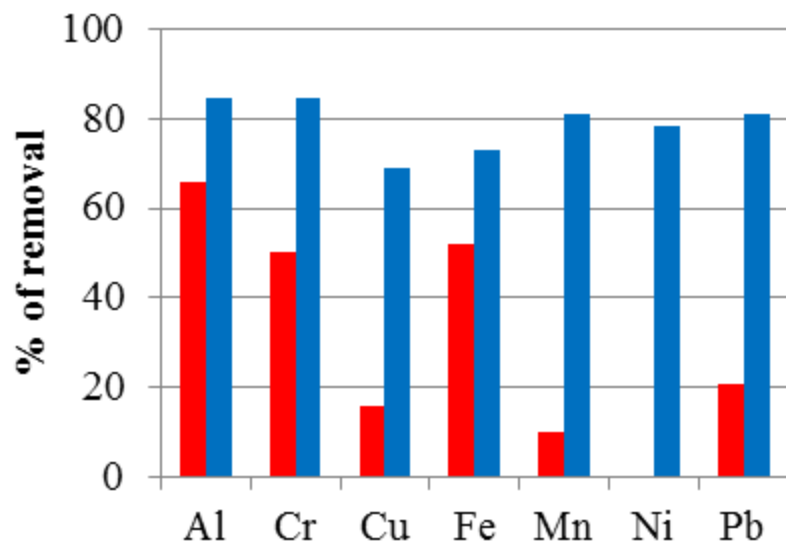


Fig.4: Mean trends (\pm SEM) of metallic element removal during the first 4 h (240 min; Aluminum, A; Chromium, B; Iron, C; Manganese, D; Nickel, E; Lead, F; Copper, G) with *D. polymorpha* (blue curve) and without bivalves (red curve) inside the pilot-plant for the 50% inlet/50% outlet mixture. The differences between controls and treated, with the exception of Nickel and Iron, were statistically significant (Two-way ANOVA).



A



B

— With *D. polymorpha*
— Without *D. polymorpha*

Fig.5: Removal percentage of metallic elements from wastewater after 24 h in the 25% inlet/75% outlet (A) and 50% inlet/50% outlet (B) mixtures.

| test without <i>D. polymorpha</i> | | | | | | | | test with <i>D. polymorpha</i> | | | | | | |
|-----------------------------------|-----|-------|------|-----|-----|------|--|--------------------------------|-----|-------|------|-----|-----|------|
| 25% IN/ 75% OUT | | | | | | | | | | | | | | |
| Al | Cr | Fe | Mn | Ni | Pb | Cu | | Al | Cr | Fe | Mn | Ni | Pb | Cu |
| 56.6 | 2.5 | 802.4 | 41.9 | 5.2 | 2.7 | 11.7 | | 67.5 | 1.5 | 249.2 | 15.1 | 3.3 | 3.0 | 11.7 |
| 50% IN/ 50% OUT | | | | | | | | | | | | | | |
| Al | Cr | Fe | Mn | Ni | Pb | Cu | | Al | Cr | Fe | Mn | Ni | Pb | Cu |
| 88.2 | 3.4 | 469.3 | 25.8 | 4.5 | 3.4 | 11.6 | | 103.2 | 2.1 | 272.7 | 17.4 | 2.2 | 3.4 | 13.6 |

Table 1: Initial concentrations ($\mu\text{g/L}$) of metallic elements detected in 25% IN/ 75% OUT and 50% IN/ 50% OUT mixtures into the pilot-plant at the beginning of the removal tests without and with *D. polymorpha*. The data related to the initial concentration of COD and total suspended solids into the two considered mixtures are shown in Binelli et al., 2014.