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Impacts of a reduction in seawater pH mimicking ocean acidification on the structure and diversity of mycoplankton communities

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Short title: Ocean acidification impacts fungal community structure and diversity

1 **Abstract**

2 Increases in atmospheric carbon dioxide (CO₂) change ocean chemistry, as dissolved CO₂
3 leads to a reduction in the seawater pH. Many marine taxa have been shown to be affected by
4 ocean acidification, while information on marine fungi is lacking. Here, we analyze the effect
5 of pH on mycoplankton communities. The pH of microcosms was adjusted to a value
6 mimicking the predicted ocean acidification in the near future. Fungal communities were
7 analyzed using a double-marker gene approach, allowing a more detailed analysis of their
8 response using 454 pyrosequencing. Mycoplankton communities in microcosms with *in situ*
9 and adjusted water pH values differed significantly in terms of structure and diversity. The
10 differences were mainly based on abundance shifts among the dominant taxa rather than the
11 exclusion of fungal groups. A sensitivity to lower pH values was reported for several groups
12 across the fungal kingdom and was not phylogenetically conserved. Some of the fungal
13 species that dominated the communities of microcosms with a lower pH were known
14 pathogenic fungi. With the increasing awareness of the significant role fungi play in marine
15 systems, including performing a diverse range of symbiotic activities, our results highlight the
16 importance of including fungi in further research projects studying and modeling biotic
17 responses to the predicted ocean acidification.

18
19
20 Keywords: ocean acidification, marine fungi, phylogenetic signals, 18S rRNA gene sequence,
21 ITS, microcosm, double-marker gene approach

22 1. Introduction

23 Oceans cover approximately 72% of the Earth's surface and are home to an uncountable number of
24 organisms. However, over the last two centuries, environmental conditions in ocean systems have
25 been changing due to an enormous increase in atmospheric carbon dioxide (CO₂) caused by
26 anthropogenic activities, such as the burning of fossil fuels (Sabine et al. 2004). The oceans are
27 known to act as a net sink for atmospheric CO₂. Dissolution of CO₂ in water results in the formation
28 of carboxylic acid and therefore seawater acidification (Caldeira & Wickett 2003). This mechanism,
29 called ocean acidification, has led to a drop in seawater pH by 0.1 units since the industrial revolution
30 (Sabine et al. 2004). Future scenarios predict a further decrease by 0.4 units within the current
31 century (Caldeira & Wickett 2005).

32 Many organism groups seem to be sensitive to such a drop in pH, as reported for foraminifera
33 (Webster et al. 2013), animals (Harvell et al. 2002, Hoegh-Guldberg et al. 2007), and bacteria (Witt
34 et al. 2011, Krause et al. 2012). In contrast to other marine microorganisms, the importance of fungi
35 for marine ecosystem functioning has long been underappreciated, which is why they have been
36 highly understudied. Only in recent years, the use of next-generation sequencing (NGS) techniques
37 has shed some light on fungal diversity (Richards et al. 2015, Rama et al. 2016, Taylor & Cunliffe
38 2016) and functioning (Gutierrez et al. 2011, Orsi et al. 2013, Orsi et al. 2015). Fungal diversity
39 assessments are generally conducted using internal transcribed spacer (ITS) regions as marker genes
40 (Schoch et al. 2012). However, the ITS is not the best marker for all fungal groups, and many
41 environmental sequences can often only be classified to the phylum or kingdom level (Nilsson et al.
42 2016). To achieve a better resolution of fungal communities, Arfi *et al.* (2012) proposed to include a
43 second phylogenetic marker (18S rDNA). High-resolution studies on marine fungal communities and
44 their responses to ocean acidification are currently lacking, although fungi can be affected by changes
45 in pH in diverse ways; one example is the uptake of nutrients and organic compounds driven by the

46 electrochemical gradient of H⁺ ions across the plasma membrane (Bowman & Bowman 1986). The
47 activity and kinetics of excreted fungal exoenzymes are also mediated by the medium pH (Pritsch et
48 al. 2004, Courty et al. 2005). A further but indirect effect is the availability and accessibility of
49 substrates, such as the range and size of dissolved and particulate organic matter (Verdugo et al.
50 2004).

51 In a previous study, we tested the pH sensitivity of marine fungal communities in microcosms using
52 ITS fingerprinting (automated ribosomal intergenic spacer analysis; ARISA) (Krause et al. 2013).
53 The seawater pH was altered according to one of the possible near-future scenarios (Blackford &
54 Gilbert 2007). Interestingly, the structure of the marine fungal communities showed a significant pH-
55 dependent response. However, details on the taxonomic composition of the communities and the taxa
56 involved in these structural changes could not be provided due to the inherent limitations of the
57 applied fingerprinting technique. Thus, the purpose of the present study was to investigate the
58 underlying dynamics of the fungal community composition. We tested the modified double-marker
59 gene approach of Arfi *et al.* (2012) by applying a phylogenetic-based analysis using 18S rDNA in
60 addition to a hierarchical cluster-based approach for the ITS. We aimed to identify which parameters
61 of the communities change, how they change, whether specific fungal species benefit from changes
62 in pH, and whether the observed sensitivity of fungi toward lower pH values is clade-specific. We
63 hypothesize that the observed shift in the community structure is based on the response of a few
64 fungal species and that pH sensitivity is phylogenetically conserved.

65 2. Materials and Methods

66 2.1. Establishment of the microcosm experiments and sampling scheme

67 Surface seawater samples were collected at the Helgoland Roads station in the North Sea (Germany)
68 on the 3rd of May, 2012. The sampling site is located off the island of Helgoland in the southern
69 North Sea (German Bight). Due to strong tidal currents, surface water samples are representative of
70 the entire water body (Wiltshire et al. 2008). For the year 2100, a concentration of 1000 ppm of
71 atmospheric CO₂ has been predicted (IPCC 2014), which would result in a mean surface seawater pH
72 of 7.67 in the southern North Sea according to the model of Blackford & Gilbert (2007). Thus, after
73 the sampled seawater was poured into 1.6 l glass jars, the microcosms were incubated either at the
74 current *in situ* seawater pH (8.26) or at an adjusted pH of 7.67. The seawater pH was adjusted with 2
75 M HCl, which is a validated approach for acidifying microcosm-based experiments (Krause et al.
76 2012). For each pH treatment, 20 replicates were prepared and left in the dark over a maximum
77 period of 4 weeks at the *in situ* temperature on the day of sampling (8°C). The microcosms were
78 mixed daily by manual inversion. Every week, the biomass in 5 replicates for each pH treatment was
79 collected by filtration onto sterile nitrocellulose filters (0.45 µm pore size, 47 mm diameter, gray with
80 a grid; Sartorius, Göttingen, Germany). Parameters of the carbonate system of every microcosm,
81 including the pH (Suppl. Table S1), were calculated based on a previously described method (Krause
82 et al. 2013) and can be accessed at <https://doi.pangaea.de/10.1594/PANGAEA.831726> (Pangaea ID
83 10.3354/ame01622). Biomass filters were stored at -20°C until further treatment.

84 2.2. DNA extraction, PCR and pyrosequencing

85 DNA was extracted as described by Krause *et al.* (2013). The PCR on the ITS2 region was conducted
86 in a reaction of 50 µl containing 100 ng of template DNA, 5 µl Taq buffer (10×), 2.5 µM MgCl₂, 250
87 µM of each dNTP, 2.5 U Taq polymerase (5 Prime, Germany), 0.15 µM of the forward fungal-
88 specific primer gITS7 (5'GTGARTCATCGARTCTTTG) (Ihrmark et al. 2012) and 0.25 µM of the

89 tagged eukaryotic ITS4 primer (5'TCCTCCGCTTATTGATATGC) (White et al. 1990). Cycling
90 conditions followed Krause *et al.* (Krause et al. 2013). A second amplification round was conducted
91 in 2 µl of a 1:10 dilution of the first PCR products and using the primers described above. The 18S
92 rDNA was amplified in reactions of 50 µl containing 50 ng of the template DNA, 5 µl Taq buffer
93 (10×) (5 Prime), 1.25 µM MgCl₂, 200 µM of each dNTP, 0.3 µM each of the eukaryotic primer NS1
94 (5'GTAGTCATATGCTTGTCTC) and the fungal-specific primer NS8
95 (5'TCCGCAGGTTACCTACGGA) (White et al. 1990) and 1.25 U Taq polymerase. Cycling
96 conditions started with a denaturation step of 94°C for 15 min, followed by 30 cycles of 94°C for 1
97 min, 50°C for 1 min and 68°C for 2 min, with a final extension at 68°C for 10 min. For the second
98 round, the fungal-specific primer pair nu-SSU-0817 (5'TTAGCATGGAATAATRRAATAGGA)/nu-
99 SSU-1536 (5'ATTGCAATGCYCTATCCCCA) (Borneman & Hartin 2000) was used, with the first
100 primer carrying the barcode. Amplifications were conducted in reactions of 50 µl, each containing 2
101 µl of the first PCR products, 5 µl Taq buffer (10×), 1.25 mM MgCl₂, 250 µM of each dNTP, 0.48
102 µM of each primer and 2.5 U Taq polymerase (5 Prime). Cycling conditions were the same as for the
103 first PCR but with an annealing temperature of 68°C. The PCR steps included a positive control,
104 which was a seawater sample taken at the time of sampling. After purification using the PeqGold Gel
105 Extraction kit (PeqLab, Erlangen, Germany), the PCR products were sequenced on a Roche 454 GS-
106 FLX Titanium platform (LGC, Berlin, Germany) as single reads. Sequencing data are stored under
107 the study accession number SRP065054 and can be downloaded from the Sequence Read Archive
108 (SRA) of the National Center for Biotechnology Information (NCBI) website
109 (<http://www.ncbi.nlm.nih.gov/sra>).

110 **2.3. Bioinformatics and OTU designation**

111 For quality control, ITS2 and 18S rDNA sequences without a valid primer or DNA tag sequence and
112 with average quality values of <25 Phred score, >2% ambiguous symbols or homopolymer chains of

113 >8 nt were discarded. After removing barcodes and primers from the remaining sequences, sequences
114 with a read length <200 nt were also discarded. In most sequence reads, quality values dropped
115 below a value of 25 at an nt position of ~ 420. Thus, all sequences were trimmed from the 3'-end to a
116 maximal length of 400 nt. Potential chimeric sequences were detected and removed using the
117 UCHIME program (Edgar et al. 2011). From this point on, the remaining ITS2 sequences were
118 subjected to hierarchical clustering analyses, while 18S rDNA sequences were further analyzed using
119 a phylogenetic-based approach.

120 **2.4. Fungal taxonomy**

121 Due to the primer attachment sites, the generated ITS reads contained sequence information from the
122 5.8S and 28S rDNA. As these fragments influence the sequence identification process, they were
123 removed from the generated sequence reads using the ITSx program (Bengtsson-Palme et al. 2013).
124 ITS2 sequences were clustered using a 97% sequence similarity threshold into operational taxonomic
125 units (OTUs) using the UCLUST algorithm (Edgar 2010). All these steps were processed using the
126 Quantitative Insights In Microbial Ecology (QIIME) pipeline (Caporaso et al. 2010). Sequences were
127 blasted using the FHiTINGS program (Dannemiller et al. 2014). Finally, OTUs that were represented
128 by fewer than five reads and observed only in a single sample were removed. 18S rDNA reads
129 passing the above-described quality control requirements were clustered and classified by mapping
130 the sequences to our existing maximum likelihood reference tree (for details on the sequences,
131 alignment and methods see Panzer *et al.* (2015). The effects of pH and time on specific fungal groups
132 were tested using generalized linear modeling (Zuur et al. 2009), as our data followed Poisson or
133 negative binomial distributions. Model selection was carried out using the Akaike information
134 criterion (AIC); models were ordered according to the AIC, and those with lowest scores were
135 compared using likelihood ratio tests (LR). The best model was then evaluated for significance of the
136 interaction term using LR. Tests were run in R (R Core Team 2015) using the “glm” and “glm.nb”

137 functions of the Mass library (Venables & Ripley 2002). The values from the “summary” table are
138 reported.

139 **2.5. Alpha diversity analyses based on the ITS2 dataset**

140 Diversity indices were calculated based on subsampled datasets using the PRIMER 6 software
141 (Plymouth Marine Laboratories, UK). Next, the effects of the factors “pH” and “incubation time” on
142 the diversity indices were analyzed. In the case for the Simpson and Shannon diversity data, the
143 statistical general least squares (GLS) model (Galecki & Burzykowski 2013) was chosen, which can
144 deal with data with heterogeneous variances. The details of the statistical procedure are given in Zuur
145 *et al.* (2009) and Galecki & Burzykowski (2013), and here we only describe them in brief. In the first
146 stage, we fitted models with different variance structure (i.e., different patterns of variance
147 heterogeneity) using restricted maximum likelihood (REML) and selected the model with the best
148 variance structure using the AIC and LR as explained in section 2.4. In the second step, we used the
149 model with the best variance structure to test for the effects of “pH” and “incubation time” on the
150 average diversity. In this step, models are fitted using the maximum likelihood method (ML) and are
151 then compared using the AIC, and significant effects are tested using LR. When significant effects
152 are found, the parameters in the model with the significant terms are refitted using REML in order to
153 express the appropriate estimates of variances and average trends. GLS analysis was carried out using
154 the package “nlme” (Pinheiro *et al.* 2015) implemented in R (R Core Team 2015). The effects of pH
155 and incubation time on OTU richness were evaluated using generalized linear modeling (Zuur *et al.*
156 2009) with Poisson and negative binomial error structure as in the case of species counts (see section
157 2.4).

158 **2.6. Beta diversity analyses based on the ITS2 dataset**

159 Beta diversity was analyzed based on subsampled datasets using permutational multivariate ANOVA
160 (PERMANOVA) of square root-transformed OTU numbers (as Bray-Curtis similarities). To
161 visualize the influence of the factors “pH” and “incubation time” on the fungal community structure,
162 we performed principal coordinate analyses (PCoAs). The similarity percentage (SIMPER) procedure
163 was used to identify OTUs that contributed most to the observed dissimilarity between the two pH
164 treatments. We examined only OTUs with a threshold of > 1% of the total contribution to
165 differences. All analyses were performed using the PRIMER 6 software.

166 **2.7. Phylogenetic diversity of the 18S rDNA dataset**

167 To evaluate the evolutionary relationships among taxa, we ran further phylogenetic metric-based
168 PCoAs using the 18S rDNA dataset. Phylogenetic separation patterns were identified using the
169 FastUniFrac suite implemented in the Galaxy platform (Giardine et al. 2005, Goecks et al. 2010).
170 Unweighted and weighted distances were calculated.

171 **2.8. Phylogenetic signals in the 18S rDNA dataset**

172 To test for possible genetic conservatism among fungi to acclimate to the factor “pH”, phylogenetic
173 signal analysis was performed using the software PHYLOCOM v 4.2 (Webb et al. 2008) with the
174 Comstruct function. Evolutionary conservatism was analyzed based on the net relatedness index
175 (NRI) and the nearest taxon index (NTI) using the 18S rDNA dataset. The NRI indicates how many
176 taxa in a community are dispersed (negative value) or clustered (positive value) over the whole
177 phylogenetic tree, while the NTI indicates dispersal/clustering at lower taxonomic levels. As a null
178 model, the “-m 2” function was chosen, while the number of randomizations was set to 9999. A p-
179 value of $p < 0.05$ against the random distributions was considered to indicate a significant
180 phylogenetic signal. In a last step, Nodesig analyses were run to identify the clades/taxa in the
181 phylogenetic tree that caused the phylogenetic signals.

182 **3. Results**

183 **3.1. Dataset description**

184 In total, 237,967 and 532,304 sequence reads were generated from 42 samples when targeting the
185 ITS2 and the 18S rDNA, respectively. For sample number 28, only five 18S rDNA reads were
186 generated, and it was thus excluded from further analysis of the 18S rDNA dataset. Sequence
187 trimming and processing resulted in 171,872 ITS2 and 404,281 18S rDNA sequence reads (Suppl.
188 Table S2). Of these, 116,056 ITS2 sequences and 126,051 18S rDNA sequences were of fungal
189 origin. The 18S rDNA sequences generated by co-amplification of the primer pair clustered mainly
190 within the groups of *Telonema*, Alveolata and Stramenopiles.

191 **3.2. Taxonomic composition of the samples**

192 The ITS2 dataset revealed Ascomycota to be the dominant phylum in all sample types, comprising
193 between 72 to 96% of the relative (rel.) sequence abundance, from which 38 – 78% could not be
194 further taxonomically assigned. Dothideomycetes and Eurotiomycetes were the most abundant
195 ascomycete classes under a pH of 8.26 and 7.67, respectively. Basidiomycota were found to be
196 present in all samples but with varying abundance between the two treatments. Up to 22% of the
197 sequences were assigned to Agaricomycetes as the dominant group in samples with a pH of 8.26. In
198 contrast, Microbotryomycetes was the most represented basidiomycete class in samples with a pH of
199 7.67. No basal fungal lineages were described in the ITS2 dataset, and up to 10% of the sequences
200 could not be assigned beyond the kingdom level (Fig. 1A). All abundant fungal groups in the present-
201 day pH samples with incubation times of up to two weeks were also detected in the positive control
202 (Suppl. Table 3).

203

204 In the 18S rDNA dataset, nearly all sequences could be assigned beyond the phylum level, leaving no
205 sequences with an ambiguous taxonomic assignment at higher levels. Ascomycota showed similar

206 abundance values (74 – 98%) as in the ITS2 dataset, with Eurotiomycetes and Leotiomycetes as
207 dominant groups in both sample types, followed by Dothideomycetes. The percentage of
208 Basidiomycota sequences was relatively small (1 – 7%) and did not exceed more than 2% of the rel.
209 sequence abundance in samples with a pH of 7.67. Using the 18S rDNA dataset, five basal fungal
210 lineages *sensu lato* were detected, namely, Chytridiomycota, Cryptomycota, LKM15,
211 Mucoromycotina and Neocallimastigomycota. In the *in situ* pH treatments, the fraction of sequences
212 assigned to basal fungal lineages declined over time but started with a value of 20% rel. sequence
213 abundance. The positive control was more similar to the present-day pH samples than to the near-
214 future pH samples but contained all abundant groups of both treatments (Suppl. Table 4).

215 Several fungal groups could not be tested for significant effects of pH and incubation time on their
216 occurrence, as their abundances were too low for reliable results. Thus, only five out of fifteen and
217 eleven out of twenty-seven groups of the ITS2 and 18S rDNA datasets, respectively, were included
218 in this analysis. Of those, the abundance of ten groups was significantly affected by the factor “pH”.
219 With the exceptions of Eurotiomycetes and ambiguous Ascomycota, the remaining eight groups,
220 namely, Cryptomycota, Chytridiomycetes, Microbotryomycetes, Agaricomycetes, Ascomycota
221 *incertae sedis*, Sordariomycetes, Saccharomycetes and Dothideomycetes, had consistently lower
222 abundances under the reduced pH of 7.67 (Table 1).

223 **3.3. ITS-based diversity analyses**

224 Alpha diversity: GLS analysis identified the interaction term of time and pH as significant for
225 Shannon diversity ($F_{3,31}=8.62$, $p=0.0003$). For Simpson diversity, the interaction term was marginally
226 significant ($F_{3,31}=2.66$, $p=0.065$), while time ($F_{3,34}=6.1$, $p=0.0022$) and pH ($F_{1,34}=10.59$, $p=0.0022$)
227 had significant effects. For richness, the best model in terms of residuals included the interaction

228 term (Table 2). The diversity of the abundant species declined at a pH of 7.67 (Suppl. Table S5,
229 Table 2). In contrast, the richness was higher at a pH of 8.26.

230 Beta diversity: Bray-Curtis similarity-based PCoA of the ITS2 dataset (Fig. 2) showed a clear
231 separation of samples according to the pH treatment, which is supported by the PERMANOVA
232 analysis (Table 3). However, neither the factor “incubation time” nor the interaction term was found
233 to be significant.

234 The SIMPER analyses indicated that the OTUs of 15 individual taxa were largely responsible for the
235 observed differences in fungal communities between the samples with different pH levels. The
236 average dissimilarity between the two tested groups was 81.9%. Fourteen of the taxa were part of the
237 core community of the microcosms with an adjusted pH of 7.67. In total, only seven taxa could be
238 identified to the species level, and of those, three were known potential pathogens (Table 4).

239 **3.4. Phylogenetic analysis**

240 Phylogenetic community structure: The unweighted UniFrac-based PCoA did not clearly separate the
241 samples with different pH treatments (Fig. 3A). Similar results were found when using the weighted
242 UniFrac as a metric, but the separation was slightly more pronounced than for the first analysis (Fig.
243 3B), while incubation time only showed an effect on samples from the first week. This indicates that
244 most of the samples are relatively similar from a phylogenetic perspective and that observed
245 community shifts must be explained based on shifts in taxa abundance, which is in accordance with
246 the non-phylogenetically based PCoA analyses.

247 Phylogenetic signal analysis: The Phylocom analysis showed significant phylogenetic signals in both
248 community types. The signals always indicated phylogenetic clustering with values of 2.2 for the
249 NRI and 1.2 for the NTI and solely with a value of 1.6 for the NRI in samples with a pH of 8.26 and
250 7.67, respectively (Fig. 4A). The Nodessig analysis indicated that the phylogenetic signals were

251 mainly caused by the same clades under both pH conditions. The exceptions were significant
252 sequence clustering within clades of Saccharomycetales and Pleosporales exclusively in samples
253 with an adjusted pH (Fig. 4B).

254 4. Discussion

255 4.1. Assessing the double-marker gene approach for community analysis

256 In our approach, we used two marker genes, namely, the ITS and 18S rDNA. The ITS resolves many
257 fungal taxa to the species or genus level but performs less well on basal fungal lineages and lacks
258 phylogenetic power (Schoch et al. 2012). Marine fungal communities are characterized by their
259 (larger) fraction of basal fungi (Panzer et al. 2015, Richards et al. 2015) and many undescribed fungal
260 taxa and groups (Manohar & Raghukumar 2013). Their classification depends on the respective
261 phylogenetic marker applied (Le Calvez et al. 2009, Jones et al. 2011, Nagahama et al. 2011). Indeed,
262 in our study, nearly all fungal sequences were classified to a lower taxonomic level with the 18S
263 rDNA, also revealing the presence of basal fungi. In contrast, the taxonomy of several fungal
264 sequences was only poorly resolved with the ITS, and no basal fungal lineages were detected (Fig.
265 1).

266 A phylogenetic-based community approach differs from a non-phylogenetic one with respect to the
267 method used to classify the sequence reads and the metrics used for statistics. The former
268 incorporates the phylogenetic relatedness among taxa into the analyses, which offers further insights
269 into the evolutionary and ecological drivers shaping the community (Horner-Devine & Bohannan
270 2006, Emerson et al. 2011). In contrast, the standard approach for non-phylogenetic markers, such as
271 the ITS, includes similarity threshold-based sequence clustering (e.g., (Li & Godzik 2006, Edgar
272 2010)), taxonomic OTU assignment by k-mer similarity search (e.g., (Altschul et al. 1990)) and -
273 most critically - statistics that consider all OTUs as equivalent independently of whether they are
274 phylogenetically highly divergent or not (Faith 1994, Martin 2002). In our approach, we combined
275 the individual advantages of the two marker genes rather than using them in an additive manner as
276 done by Arfi *et al.* (2012) and additionally minimized the primer bias (Hong et al. 2009). Finally, the
277 ITS, possessing good resolution to the species/genus level, was used for the alpha and beta diversity

278 analyses and for the identification of core species (by SIMPER). In contrast, the 18S rDNA was used
279 to resolve the phylogeny of undescribed taxa/groups and basal fungal lineages to reveal differences in
280 the phylogenetic structure among communities and to test for possible evolutionary conservatism of
281 pH sensitivity in fungal clades.

282 **4.2. Community responses**

283 The mycoplankton community of the microcosms showed a significant response to the pH treatment.
284 However, the communities in the two treatments were phylogenetically highly related, and the
285 observed effect was mainly based on a change in abundance among the dominant taxa rather than a
286 change in the general community composition or extinction of several fungal groups (Fig. 2). Thus,
287 the mycoplankton community can cope with a drop in pH within the tested range, but sensitivity to
288 the pH change differs highly among fungal taxa. The phylogenetic signal analysis indicated no clade-
289 specific sensitivity to one or the other pH (Fig. 4). A phylogenetic signal can occur when, for
290 example, closely related taxa share a character or suite of characters allowing them to be adaptive to
291 the given environmental conditions (Horner-Devine & Bohannan 2006). In the communities in the
292 two pH treatments, small significant clustering effects were observed, but phylogenetic signals were
293 nearly exclusively caused by the same clades in both pH treatments (Fig. 4). This observation points
294 to other factors than the pH causing the clustering effect. In our case, it can be assumed that the
295 habitat-specific conditions of the pelagial at Helgoland Roads are the structuring factors. Panzer *et al.*
296 (2015) recently demonstrated that aquatic fungal community structure is highly influenced by the
297 source habitat, resulting in habitat-specific phylogenetic signals. Thus, the sensitivity to a pH shift
298 observed in our study must be caused by different and (clade-) independent characters; it has, for
299 example, been shown that diverse transcriptional regulatory systems control the response to the
300 environmental pH (Penalva & Arst 2004, Selvig & Alspaugh 2011). In addition, fungi seem to react
301 quite differently to the environmental pH depending on their life stage; for example, the fungal

302 germination time and outgrowth can be affected by the environmental pH (Magan & Lacey 1984,
303 Porter et al. 1987). For pathogenic fungi with host-independent life stages (e.g., some Cryptomycota
304 or Chytridiomycetes species), responses to pH may depend on the life stage.

305 **4.3. Relevance for food webs**

306 Microbial communities highly influence resource use, disease dynamics and the stability of
307 ecosystem processes (Duffy & Stachowicz 2006, Stachowicz & Byrnes 2006). For example, a pH-
308 dependent loss of fungal diversity in freshwater systems (Wood-Eggenschwiler & Bärlocher 1983,
309 Tolkkinen et al. 2015) can lead to a drastic reduction in fungal leaf breakdown (Baudoin et al. 2008).
310 So far, comparable data for marine fungi are lacking, but recent studies on marine mycoplankton
311 have elucidated their importance in the breakdown of organic matter (Gutierrez et al. 2011) and the
312 control of phytoplankton populations (Gleason et al. 2008, Gutierrez et al. 2016). Thus, zoosporic
313 fungi are part of the pelagic food web, forming a “mycoloop”; herein, they (i) act as a food source for
314 zooplankton grazers, (ii) decompose particulate organic matter, (iii) convert inorganic compounds
315 into organic compounds, and (iv) parasite/hyperparasite on diverse organisms (Kagami et al. 2007,
316 Kagami et al. 2014). In our study, players in the “mycoloop”, namely, Cryptomycota and
317 Chytridiomycetes, showed significant reductions under a reduced pH (Table 1). According to
318 dynamical model predictions, a reduction in zoosporic fungi in pelagic systems would affect the
319 overall food web structure: fewer fungal parasites lead to a higher abundance of large phytoplankton
320 species (>40 µm), which in turn lowers the numbers of smaller phytoplankton species due to resource
321 competition. As zooplankton species only graze on smaller phytoplankton, a reduction in zoosporic
322 fungi has an indirect and negative effect on zooplankton grazers (Miki et al. 2011). The involvement
323 of filamentous fungi in the pelagic food web has so far not been modeled due to missing data. In our
324 study, two of the most prominent groups, Eurotiomycetes and Dothideomycetes, both showed a
325 significant pH effect (Table 1). Gutierrez *et al.* (2011) suggest an important role in the degradation of

326 phytoplankton bloom-derived organic matter for these fungi. Thus, a pH-driven increase in
327 filamentous fungal biomass might have a negative impact on organisms competing for the same food
328 source.

329 Three of the seven key taxa (described to the genus level) that significantly benefited from the pH
330 reduction are pathogenic fungi that cause disease in marine algae (Almeida et al. 2010) and animals
331 (Wirth & Goldani 2012) (Table 4). Pathogenic fungal activities in marine environments may lead to
332 the breakdown of ecosystems (Raghukumar & Ravindran 2012) or cause significant economic losses
333 when commercially important organisms are infected (Gachon et al. 2010, Hatai 2012). Given that
334 pathogenicity is regulated by several factors (e.g., host physiology, species interactions and
335 environmental factors (Fuhrman 2009, Selvig & Alspaugh 2011)), more detailed studies are urgently
336 needed to examine the effects of ocean acidification on this important group.

337 **4.4. Conclusion**

338 Although the marine fungal communities in the two pH treatments were phylogenetically closely
339 related, their taxonomic composition and diversity differed significantly. The observed changes were
340 mainly caused by a restructuring of the abundant species. Sensitivity to the pH reduction was not
341 phylogenetically conserved. The reported pH sensitivity contrasts with the wide pH tolerance
342 observed in marine fungi under controlled conditions. We have only just begun to understand the
343 ecological role of marine fungi, but given the newly discovered role mycoplankton play in organic
344 matter breakdown and phytoplankton bloom dynamics, further studies are urgently needed to place
345 the observed pH-dependent community shift and possible consequences for the marine ecosystem
346 into context.

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352 **6. Author contributions**

353 Experimental design: EK, AW, GG. Sampling and setup of the experiment: EK. Data analyses: MR,
354 KP. Statistical analyses: GG, LG. Writing of the paper: MR.

355 **7. References**

- 356 Almeida C, Eguereva E, Kehraus S, Siering C, König GM (2010) Hydroxylated sclerosporin
357 derivatives from the marine-derived fungus *Cadophora malorum*. *J Nat Prod* 73:476-478
- 358 Altschul SF, Gish W, Miller W, Myers EW, Lipman DJ (1990) Basic local alignment search tool. *J*
359 *Mol Biol* 215:403-410
- 360 Arfi Y, Buee M, Marchand C, Levasseur A, Record E (2012) Multiple markers pyrosequencing
361 reveals highly diverse and host-specific fungal communities on the mangrove trees *Avicennia*
362 *marina* and *Rhizophora stylosa*. *FEMS Microbiol Ecol* 79:433-444
- 363 Baudoin JM, Guerold F, Felten V, Chauvet E, Wagner P, Rousselle P (2008) Elevated aluminium
364 concentration in acidified headwater streams lowers aquatic hyphomycete diversity and
365 impairs leaf-litter breakdown. *Microb Ecol* 56:260-269
- 366 Bengtsson-Palme J, Ryberg M, Hartmann M, Branco S, Wang Z, Godhe A, De Wit P, Sanchez-
367 Garcia M, Ebersberger I, de Sousa F, Amend AS, Jumpponen A, Unterseher M, Kristiansson
368 E, Abarenkov K, Bertrand YJK, Sanli K, Eriksson KM, Vik U, Veldre V, Nilsson RH (2013)
369 Improved software detection and extraction of ITS1 and ITS2 from ribosomal ITS sequences
370 of fungi and other eukaryotes for analysis of environmental sequencing data. *Methods Ecol*
371 *Evol* 4:914-919
- 372 Blackford JC, Gilbert FJ (2007) pH variability and CO₂ induced acidification in the North Sea. *J*
373 *Marine Syst* 64:229-241
- 374 Borneman J, Hartin RJ (2000) PCR primers that amplify fungal rRNA genes from environmental
375 samples. *AEM* 66:4356-4360
- 376 Bowman BJ, Bowman EJ (1986) H⁺-ATPases from mitochondria, plasma membranes, and vacuoles
377 of fungal cells. *J Membrane Biol* 94:83-97
- 378 Caldeira K, Wickett ME (2003) Oceanography: Anthropogenic carbon and ocean pH. *Nature*
379 425:365-365
- 380 Caldeira K, Wickett ME (2005) Ocean model predictions of chemistry changes from carbon dioxide
381 emissions to the atmosphere and ocean. *J Geophys Res-Oceans* 110:C09S04
- 382 Caporaso JG, Kuczynski J, Stombaugh J, Bittinger K, Bushman FD, Costello EK, Fierer N, Pena
383 AG, Goodrich JK, Gordon JI, Huttley GA, Kelley ST, Knights D, Koenig JE, Ley RE,
384 Lozupone CA, McDonald D, Muegge BD, Pirrung M, Reeder J, Sevinsky JR, Tumbaugh PJ,
385 Walters WA, Widmann J, Yatsunenkov T, Zaneveld J, Knight R (2010) QIIME allows analysis
386 of high-throughput community sequencing data. *Nat Methods* 7:335-336
- 387 Courty PE, Pritsch K, Schloter M, Hartmann A, Garbaye J (2005) Activity profiling of
388 ectomycorrhiza communities in two forest soils using multiple enzymatic tests. *New Phytol*
389 167:309-319

- 390 Dannemiller KC, Reeves D, Bibby K, Yamamoto N, Peccia J (2014) Fungal high-throughput
391 taxonomic identification tool for use with Next-Generation Sequencing (FHiTINGS). *J Basic*
392 *Microb* 54:315-321
- 393 Duffy JE, Stachowicz JJ (2006) Why biodiversity is important to oceanography: potential roles of
394 genetic, species, and trophic diversity in pelagic ecosystem processes. *Mar Ecol Prog Ser*
395 311:179-189
- 396 Edgar RC (2010) Search and clustering orders of magnitude faster than BLAST. *Bioinformatics*
397 26:2460-2461
- 398 Edgar RC, Haas BJ, Clemente JC, Quince C, Knight R (2011) UCHIME improves sensitivity and
399 speed of chimera detection. *Bioinformatics* 27:2194-2200
- 400 Emerson BC, Cicconardi F, Fanciulli PP, Shaw PJA (2011) Phylogeny, phylogeography,
401 phylobetadiversity and the molecular analysis of biological communities. *Philosophical*
402 *Transactions of the Royal Society B-Biological Sciences* 366:2391-2402
- 403 Faith DP (1994) Phylogenetic Pattern and the Quantification of Organismal Biodiversity. *Philos T*
404 *Roy Soc B* 345:45-58
- 405 Fuhrman JA (2009) Microbial community structure and its functional implications. *Nature* 459:193-
406 199
- 407 Gachon CMM, Sime-Ngando T, Strittmatter M, Chambouvet A, Kim GH (2010) Algal diseases:
408 spotlight on a black box. *Trends Plant Sci* 15:633-640
- 409 Galecki A, Burzykowski T (2013) Linear mixed-effects models using R, Vol. Springer, Berlin
410 Heidelberg
- 411 Giardine B, Riemer C, Hardison RC, Burhans R, Elnitski L, Shah P, Zhang Y, Blankenberg D, Albert
412 I, Taylor J, Miller W, Kent WJ, Nekrutenko A (2005) Galaxy: A platform for interactive
413 large-scale genome analysis. *Genome Res* 15:1451-1455
- 414 Gleason FH, Kagami M, Lefèvre E, Sime-Ngando T (2008) The ecology of chytrids in aquatic
415 ecosystems: roles in food web dynamics. *Fungal Biol Rev* 22:17-25
- 416 Goecks J, Nekrutenko A, Taylor J, Team G (2010) Galaxy: a comprehensive approach for supporting
417 accessible, reproducible, and transparent computational research in the life sciences. *Genome*
418 *Biol* 11
- 419 Gutierrez MH, Jara AM, Pantoja S (2016) Fungal parasites infect marine diatoms in the upwelling
420 ecosystem of the Humboldt Current System of central Chile. *Environ Microbiol*:In press
- 421 Gutierrez MH, Pantoja S, Tejos E, Quinones RA (2011) The role of fungi in processing marine
422 organic matter in the upwelling ecosystem off Chile. *Mar Biol* 158:205-219
- 423 Harvell CD, Mitchell CE, Ward JR, Altizer S, Dobson AP, Ostfeld RS, Samuel MD (2002) Ecology -
424 Climate warming and disease risks for terrestrial and marine biota. *Science* 296:2158-2162

- 425 Hatai K (2012) Diseases of fish and shellfish caused by marine fungi. In: Raghukumar C (ed)
426 Biology of marine fungi. Springer, Heidelberg
- 427 Hoegh-Guldberg O, Mumby PJ, Hooten AJ, Steneck RS, Greenfield P, Gomez E, Harvell CD, Sale
428 PF, Edwards AJ, Caldeira K, Knowlton N, Eakin CM, Iglesias-Prieto R, Muthiga N,
429 Bradbury RH, Dubi A, Hatziolos ME (2007) Coral reefs under rapid climate change and
430 ocean acidification. *Science* 318:1737-1742
- 431 Hong SH, Bunge J, Leslin C, Jeon S, Epstein SS (2009) Polymerase chain reaction primers miss half
432 of rRNA microbial diversity. *ISME J* 3:1365-1373
- 433 Horner-Devine MC, Bohannan BJM (2006) Phylogenetic clustering and overdispersion in bacterial
434 communities. *Ecology* 87:S100-S108
- 435 Ihrmark K, Bodeker ITM, Cruz-Martinez K, Friberg H, Kubartova A, Schenck J, Strid Y, Stenlid J,
436 Brandstrom-Durling M, Clemmensen KE, Lindahl BD (2012) New primers to amplify the
437 fungal ITS2 region - evaluation by 454-sequencing of artificial and natural communities.
438 *FEMS Microbiol Ecol* 82:666-677
- 439 IPCC (2014) Climate change 2014: Mitigation of Climate Change. In: Edenhofer O, Pichs-Madruga
440 R, Sokona Y, Farahani E, Kadner S, Seyboth K, Adler A, Baum I, Brunner S, Eickemeier P,
441 Kriemann B, Savolainen J, Schlömer S, von Stechow C, Zwickel T, Minx JC (eds)
442 Contribution of Working Group III to the Fifth Assessment Report of the Intergovernmental
443 Panel on Climate Change. Cambridge University Press, Cambridge, UK
- 444 Jones MDM, Forn I, Gadelha C, Egan MJ, Bass D, Massana R, Richards TA (2011) Discovery of
445 novel intermediate forms redefines the fungal tree of life. *Nature* 474:200-U234
- 446 Kagami M, de Bruin A, Ibelings BW, Van Donk E (2007) Parasitic chytrids: their effects on
447 phytoplankton communities and food-web dynamics. *Hydrobiologia* 578:113-129
- 448 Kagami M, Miki T, Takimoto G (2014) Mycoloop: chytrids in aquatic food webs. *Front Microbiol* 5
- 449 Krause E, Wichels A, Gimenez L, Gerdts G (2013) Marine fungi may benefit from ocean
450 acidification. *Aquat Microb Ecol* 69:59-67
- 451 Krause E, Wichels A, Gimenez L, Lunau M, Schilhabel MB, Gerdts G (2012) Small Changes in pH
452 have direct effects on marine bacterial community composition: A microcosm approach.
453 *PLoS ONE* 7:e47035
- 454 Le Calvez T, Burgaud G, Mahe S, Barbier G, Vandenkoornhuysen P (2009) Fungal diversity in Deep-
455 Sea hydrothermal ecosystems. *AEM* 75:6415-6421
- 456 Li WZ, Godzik A (2006) Cd-hit: a fast program for clustering and comparing large sets of protein or
457 nucleotide sequences. *Bioinformatics* 22:1658-1659
- 458 Magan N, Lacey J (1984) Effect of temperature and pH on water relations of field and storage fungi.
459 *T Brit Mycol Soc* 82:71-81

- 460 Manohar CS, Raghukumar C (2013) Fungal diversity from various marine habitats deduced through
461 culture-independent studies. *FEMS Microbiol Lett* 341:69-78
- 462 Martin AP (2002) Phylogenetic approaches for describing and comparing the diversity of microbial
463 communities. *AEM* 68:3673-3682
- 464 Miki T, Takimoto G, Kagami M (2011) Roles of parasitic fungi in aquatic food webs: a theoretical
465 approach. *Freshwater Biol* 56:1173-1183
- 466 Nagahama T, Takahashi E, Nagano Y, Abdel-Wahab MA, Miyazaki M (2011) Molecular evidence
467 that deep-branching fungi are major fungal components in deep-sea methane cold-seep
468 sediments. *Environ Microbiol* 13:2359-2370
- 469 Nilsson RH, Wurzbacher C, Bahram M, Coimbra VRM, Larsson E, Tedersoo L, Eriksson J, Ritter
470 CD, Svantesson S, Sanchez-Garcia M, Ryberg M, Kristiansson E, Abarenkov K (2016) Top
471 50 most wanted fungi. *Myckeys*:29-40
- 472 Orsi W, Biddle JF, Edgcomb V (2013) Deep sequencing of seafloor eukaryotic rRNA reveals
473 active fungi across marine subsurface provinces. *PLoS ONE* 8
- 474 Orsi WD, Richards TA, Santoro AE (2015) Cellular maintenance processes that potentially underpin
475 the survival of seafloor fungi over geological timescales. *Estuar Coast Shelf S* 164:A1-A9
- 476 Panzer K, Yilmaz P, Weiß M, Reich L, Richter M, Wiese J, Schmaljohann R, Labes A, Imhoff JF,
477 Glöckner FO, Reich M (2015) Identification of habitat-specific biomes of aquatic fungal
478 communities using a comprehensive nearly full-length 18S rRNA dataset enriched with
479 contextual data. *PLoS ONE* 10:e0134377
- 480 Penalva MA, Arst HN (2004) Recent advances in the characterization of ambient pH regulation of
481 gene expression in filamentous fungi and yeasts. *Annu Rev Microbiol* 58:425-451
- 482 Pinheiro J, Bates D, DebRoy S, Team RC (2015) *nlme*: Linear and nonlinear mixed effects models.
483 In: R package version 3.1-128, <http://CRAN.R-project.org/package=nlme>.
- 484 Porter WM, Robson AD, Abbott LK (1987) Field survey of the distribution of vesicular arbuscular
485 mycorrhizal fungi in relation to soil-pH. *J Appl Ecol* 24:659-662
- 486 Pritsch K, Raidl S, Marksteiner E, Blaschke H, Agerer R, Schloter M, Hartmann A (2004) A rapid
487 and highly sensitive method for measuring enzyme activities in single mycorrhizal tips using
488 4-methylumbelliferone-labelled fluorogenic substrates in a microplate system. *J Microbiol*
489 *Methods* 58:233-241
- 490 R Core Team (2015) R: A language and environment for statistical computing. In: R Foundation for
491 statistical computing, Vienna, Austria
- 492 Raghukumar C, Ravindran J (2012) Fungi and their role in corals and coral reef ecosystems. In:
493 Raghukumar C (ed) *Biology of marine fungi*. Springer, Heidelberg

- 494 Rama T, Davey M, Norden J, Halvorsen R, Błaalid R, Mathiassen G, Alsos I, Kauserud H (2016)
 495 Fungi sailing the Arctic Ocean: Speciose communities in North Atlantic driftwood as
 496 revealed by high-throughput amplicon sequencing. *Microb Ecol* 72:295-304
- 497 Richards TA, Leonard G, Mahé F, del Campo J, Romac S, Jones MDM, Maguire F, Dunthorn M, De
 498 Vargas C, Massana R, Chambouvet A (2015) Molecular diversity and distribution of marine
 499 fungi across 130 European environmental samples. *Proc R Soc Lond [Biol]* 282
- 500 Sabine CL, Feely RA, Gruber N, Key RM, Lee K, Bullister JL, Wanninkhof R, Wong CS, Wallace
 501 DWR, Tilbrook B, Millero FJ, Peng TH, Kozyr A, Ono T, Rios AF (2004) The oceanic sink
 502 for anthropogenic CO₂. *Science* 305:367-371
- 503 Schoch CL, Seifert KA, Huhndorf S, Robert V, Spouge JL, Levesque CA, Chen W, Bolchacova E,
 504 Voigt K, Crous PW, Miller AN, Wingfield MJ, Aime MC, An KD, Bai FY, Barreto RW,
 505 Begerow D, Bergeron MJ, Blackwell M, Boekhout T, Bogale M, Boonyuen N, Burgaz AR,
 506 Buyck B, Cai L, Cai Q, Cardinali G, Chaverri P, Coppins BJ, Crespo A, Cubas P, Cummings
 507 C, Damm U, de Beer ZW, de Hoog GS, Del-Prado R, B D, Dieguez-Uribeondo J, Divakar
 508 PK, Douglas B, Duenas M, Duong TA, Eberhardt U, Edwards JE, Elshahed MS, Fliegerova
 509 K, Furtado M, Garcia MA, Ge ZW, Griffiths GW, Griffiths K, Groenewald JZ, Groenewald
 510 M, Grube M, Gryzenhout M, Guo LD, Hagen F, Hambleton S, Hamelin RC, Hansen K,
 511 Harrold P, Heller G, Herrera G, Hirayama K, Hirooka Y, Ho HM, Hoffmann K, Hofstetter V,
 512 Hognabba F, Hollingsworth PM, Hong SB, Hosaka K, Houbraken J, Hughes K, Huhtinen S,
 513 Hyde KD, James T, Johnson EM, Johnson JE, Johnston PR, Jones EB, Kelly LJ, Kirk PM,
 514 Knapp DG, Koljalg U, GM K, Kurtzman CP, Landvik S, Leavitt SD, Liggenstoffer AS,
 515 Liimatainen K, Lombard L, Luangsa-Ard JJ, Lumbsch HT, Maganti H, Maharachchikumbura
 516 SS, Martin MP, May TW, McTaggart AR, Methven AS, Meyer W, Moncalvo JM,
 517 Mongkolsamrit S, Nagy LG, Nilsson RH, Niskanen T, Nyilasi I, Okada G, Okane I, Olariaga
 518 I, Otte J, Papp T, Park D, Petkovits T, Pino-Bodas R, Quaedvlieg W, Raja HA, Redecker D, T
 519 R, Ruibal C, Sarmiento-Ramirez JM, Schmitt I, Schussler A, Shearer C, Sotome K, Stefani
 520 FO, Stenroos S, Stielow B, Stockinger H, Suetrong S, Suh SO, Sung GH, Suzuki M, Tanaka
 521 K, Tedersoo L, Telleria MT, Tretter E, Untereiner WA, Urbina H, Vagvolgyi C, Vialle A, Vu
 522 TD, Walther G, Wang QM, Wang Y, Weir BS, Weiss M, White MM, Xu J, Yahr R, Yang
 523 ZL, Yurkov A, Zamora JC, Zhang N, Zhuang WY, Schindel D, Consortium FB (2012)
 524 Nuclear ribosomal internal transcribed spacer (ITS) region as a universal DNA barcode
 525 marker for Fungi. *Proc Natl Acad Sci U S A* 109:6241-6246
- 526 Selvig K, Alspaugh JA (2011) pH response pathways in fungi: Adapting to host-derived and
 527 environmental signals. *Mycobiology* 39:249-256
- 528 Stachowicz JJ, Byrnes JE (2006) Species diversity, invasion success, and ecosystem functioning:
 529 Disentangling the influence of resource competition, facilitation, and extrinsic factors. *Mar*
 530 *Ecol Prog Ser* 311:251-262
- 531 Taylor JD, Cunliffe M (2016) Multi-year assessment of coastal planktonic fungi reveals
 532 environmental drivers of diversity and abundance. *ISME J* 10:2118-2128
- 533 Tolkkinen M, Mykra H, Annala M, Markkola AM, Vuori KM, Muotka T (2015) Multi-stressor
 534 impacts on fungal diversity and ecosystem functions in streams: natural vs. anthropogenic
 535 stress. *Ecology* 96:672-683

- 536 Venables WN, Ripley BD (2002) Modern applied statistics with S., Vol. Springer-Verlag, New York
- 537 Verdugo P, Alldredge AL, Azam F, Kirchman DL, Passow U, Santschi PH (2004) The oceanic gel
538 phase: a bridge in the DOM-POM continuum. Mar Chem 92:67-85
- 539 Webb CO, Ackerly DD, Kembel SW (2008) Phylocom: software for the analysis of phylogenetic
540 community structure and trait evolution. Bioinformatics 24:2098-2100
- 541 Webster NS, Negri AP, Flores F, Humphrey C, Soo R, Botte ES, Vogel N, Uthicke S (2013) Near-
542 future ocean acidification causes differences in microbial associations within diverse coral
543 reef taxa. Env Microbiol Rep 5:243-251
- 544 White TJ, Bruns T, Lee S, Taylor JW (1990) Amplification and direct sequencing of fungal
545 ribosomal RNA genes for phylogenetics. In: Innis MA, Gelfand DH, Sninsky JJ, White TJ
546 (eds) PCR Protocols: A guide to methods and applications. Academic Press, Inc., New York,
547 USA
- 548 Wiltshire KH, Malzahn AM, Wirtz K, Greve W, Janisch S, Mangelsdorf P, Manly BFJ, Boersma M
549 (2008) Resilience of North Sea phytoplankton spring bloom dynamics: An analysis of long-
550 term data at Helgoland Roads. Limnol Oceanogr 53:1294-1302
- 551 Wirth F, Goldani LZ (2012) Epidemiology of *Rhodotorula*: An emerging pathogen. Interdiscip
552 Perspect Infect Dis:465717
- 553 Witt V, Wild C, Anthony KRN, Diaz-Pulido G, Uthicke S (2011) Effects of ocean acidification on
554 microbial community composition of, and oxygen fluxes through, biofilms from the Great
555 Barrier Reef. Environ Microbiol 13:2976-2989
- 556 Wood-Eggenschwiler S, Bärlocher F (1983) Aquatic hyphomycetes in sixteen streams in France,
557 Germany and Switzerland. T Brit Mycol Soc 81:371-379
- 558 Zuur A, Ieno EN, Walker N, Saveliev AA, Smith GM (2009) Mixed effects models and extensions in
559 ecology with R, Vol. Springer, Heidelberg
- 560
- 561

562 **Table 1: Generalized modeling framework test of the effects of pH and week on the 13 most abundant fungal groups.** For several
563 groups, the test was not applicable due to low abundance (n/a). Displayed are the average abundances of the tested fungal groups at a pH of
564 8.26 and 7.67 as observed in the two different datasets, ITS and 18S rDNA. The p-value indicates whether the pH effect had a significant
565 effect on the occurrence of the fungal group and whether this was dependent on incubation time. Finally, the trend in abundance within the
566 two pH treatments is given. Av. Abund., average abundance; SD, standard deviation; P, p-value; L, pH 7.67; H, pH 8.26; ns, not significant.

Fungal group	Av. Abund. \pm (SD) at pH 8.26 (ITS)	Av. Abund. \pm (SD) at pH 7.67 (ITS)	p (ITS)	Av. Abund. \pm (SD) at pH 8.26 (18S rDNA)	Av. Abund. \pm (SD) at pH 7.67 (18S rDNA)	P (18S rDNA)	Trend in abund. over the two pH treatments
Cryptomycota	0	0	n/a	350 \pm 25.1	23 \pm 1.1	4.3x10 ⁻¹⁴	L<H
Chytridiomycetes	0	0	n/a	184 \pm 5.1	61 \pm 2.8	1.47x10 ⁻⁰⁸	L<H
Microbotryomycetes	16 \pm 1.3	58 \pm 2.5	10 ⁻⁴	44 \pm 2.4	35 \pm 1.4	8.8x10 ⁻³ (for week 4) 4.2x10 ⁻² (for week 1)	L>H L<H
Agaricomycetes	92 \pm 3.6	1 \pm 0	10 ⁻⁸	66 \pm 5.7	14 \pm 1.3	8x10 ⁻⁴	L<H
Ambiguous Ascomycota	371 \pm 8.1	582 \pm 4.9	2x10 ⁻⁴	16 \pm 1.2	0	n/a	L>H
Ascomycota <i>Incertae Sedis</i>	2 \pm 0	0	n/a	42 \pm 2.1	16 \pm 1	2x10 ⁻³	L<H
Sordariomycetes	15 \pm 1.2	4 \pm 0	n/a	146 \pm 14	80 \pm 1.9	0.01 (for week 3)	L<H
Saccharomycetes	29 \pm 3.9	0	n/a	62 \pm 6.2	4 \pm 0	3.57x10 ⁻⁰⁵	L<H
Leotiomycetes	42 \pm 5.9	2 \pm 0	n/a	2163 \pm 13.3	2174 \pm 19.4	0.95	ns
Lecanoromycetes	5 \pm 0.5	0	n/a	33 \pm 1.2	38 \pm 1.3	0.41	ns
Dothideomycetes	133 \pm 10.2	7 \pm 0.8	n/a	360 \pm 14.8	201 \pm 3.5	0.0	L<H
Eurotiomycetes	1 \pm 0	133 \pm 2.7	n/a	1520 \pm 26.5	201 \pm 3.5	2.25x10 ⁻⁰⁵	L>H

567

568 **Table 2: Results of generalized linear model analysis testing the effects of incubation time and pH on the richness.** Test included the
 569 examination of negative binomial and Poisson error structure. Likelihood ratio tests (LR) and associated (Chi squared-based) p-values
 570 compare the top two models, testing for the significance of the interaction term. The LR uses the degrees of freedom of the model with the
 571 interaction (df=31) and the one without the interaction (df=34) to calculate the difference (df=3) and then the LR(3)=15.46. AIC: Akaike
 572 information criterion; LR: likelihood ratio test; p: p-value (Chi squared-based).

Fixed structure	Error structure	AIC	LR	p
Time + pH + TxpH	Negative binomial	341	15.46	0.0014
Time + pH	Negative binomial	351		
Time + pH + TxpH	Poisson	401		
Time + pH	Poisson	467		

573

574 **Table 3: PERMANOVA main test of fungal community composition based on ITS2-based Bray-Curtis dissimilarities of operational**
 575 **taxonomic units.** Displayed are tests for the factors “pH” and “incubation time”, their interaction and the partitioning of multivariate
 576 variation. p-values were obtained using type III sums of squares and 9999 permutations under the full model. *, statistical significance (p <
 577 0.05); d.f.: degrees of freedom; SS: sums of squares; Sq. root: square root of the component of variation attributable to that factor in the
 578 model in units of Bray-Curtis dissimilarity.

Sources of variation	d.f.	SS	Pseudo-F	Sq.root
pH*	1	28308	15,177	36,925
Time	3	8301	14,835	96,348
pH × time	3	7791	13,924	12,275
Residuals	31	57821	43,188	
Total	38	1025200000		

579

580 **Table 4: SIMPER analysis of the ITS2 dataset.** Operational taxonomic units (OTUs) contributing most to the observed community
581 dissimilarity among the two pH sample types. Only OTUs with more than 1% contribution to the observed community dissimilarities are
582 listed. Av. Abund., average abundance; Av. Diss, average dissimilarity.

Species	Av. Abund. at pH 8.26	Av. Abund. at pH 7.67	Av. Diss.	COD (%)	Phylum	Subphylum	Class	Order
<i>Cyphellophora reptans</i>	0.49	32.78	13.39	16.33	Ascomycota	Pezizomycotina	Eurotiomycetes	Chaetothyriales
<i>Pezizomycotina</i> sp.	7.85	22.94	6.6	8.06	Ascomycota	Pezizomycotina	Incertae sedis	Incertae sedis
<i>Sporobolomyces symmetricus</i>	0.38	7.56	2.85	3.48	Basidiomycota	Pucciniomycotina	Cystobasidiomycetes	Sporidiobolales
<i>Cadophora malorum</i>	1.16	6.87	2.84	3.46	Ascomycota	Pezizomycotina	Leotiomycetes	Helotiales
<i>Sakaguchia dacryoidea</i>	1.04	4.87	2.14	2.6	Basidiomycota	Pucciniomycotina	Cystobasidiomycetes	Incertae sedis
<i>Pezizomycotina</i> sp.	2.13	6.27	1.82	2.22	Ascomycota	Pezizomycotina	Incertae sedis	Incertae sedis
<i>Eurotiomycetes</i> sp.	0.11	3.63	1.47	1.8	Ascomycota	Pezizomycotina	Eurotiomycetes	Incertae sedis
<i>Pezizomycotina</i> sp.	1.37	4.35	1.29	1.58	Ascomycota	Pezizomycotina	Incertae sedis	Incertae sedis
<i>Neophaeosphaeria quadriseptata</i>	2.28	1.2	1.26	1.54	Ascomycota	Pezizomycotina	Dothideomycetes	Pleosporales
<i>Pezizomycotina</i> sp.	1.15	3.95	1.25	1.53	Ascomycota	Pezizomycotina	Incertae sedis	Incertae sedis
<i>Pezizomycotina</i> sp.	1.22	3.89	1.21	1.48	Ascomycota	Pezizomycotina	Incertae sedis	Incertae sedis
<i>Pezizomycotina</i> sp.	0.77	2.82	0.91	1.11	Ascomycota	Pezizomycotina	Incertae sedis	Incertae sedis
<i>Eurotiomycetes</i> sp.	0	2.2	0.9	1.1	Ascomycota	Pezizomycotina	Eurotiomycetes	Incertae sedis
<i>Acremonium butyri</i>	0.76	2.57	0.88	1.07	Ascomycota	Pezizomycotina	Sordariomycetes	Hypocreales

<i>Rhodotorula aurantiaca</i>	0.21	2.3	0.85	1.04	<i>Basidiomycota</i>	<i>Pucciniomycotina</i>	<i>Cystobasidiomycetes</i>	Incertae sedis
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583

584 **Figure legends**

585 **Figure 1: Taxonomic assemblage of fungal communities in the two tested pH sample types over**
586 **four weeks of incubation.** Taxonomic composition based on the (A) ITS and (B) 18S rDNA dataset.

587 Fungal groups in bold showed a significant pH response: *, reduced abundance at a pH of 7.67; \$,
588 increased abundance at a pH of 7.67; #, not all time points showed consistent effects. For more detail,
589 see Table 2. 1-4, weeks of incubation. Bottom to dashed lines: *Ascomycota*; dashed to dotted lines:
590 *Basidiomycota*; dashed lines to top: basal fungal lineages *sensu lato*.

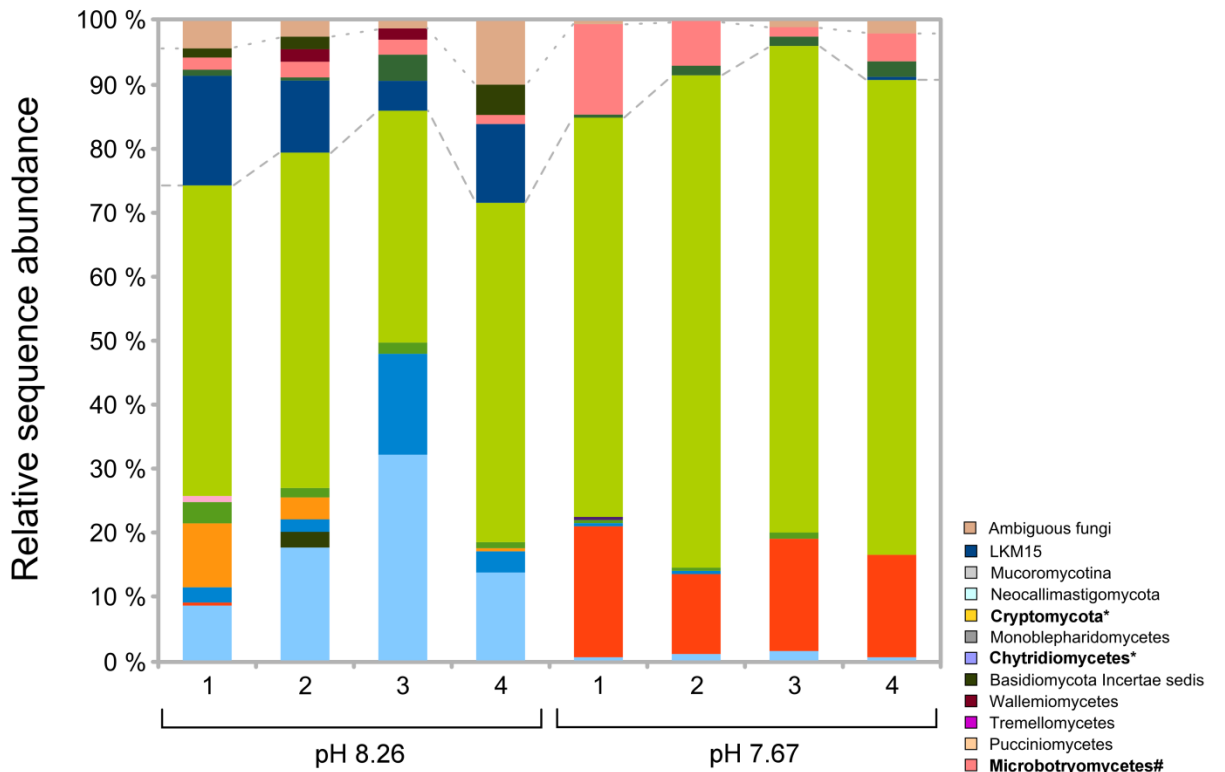
591 **Figure 2: Principal coordinate analysis (PCoA) of the ITS2 dataset.** The test is based on square
592 root-transformed data using Bray-Curtis dissimilarities of operational taxonomic unit counts.

593 **Figure 3: Ordination analysis of phylogenetic distances revealing phylogenetic relatedness**
594 **among communities.** PCoAs are based on (A) unweighted and (B) weighted UniFrac metrics using
595 the 18S rDNA data.

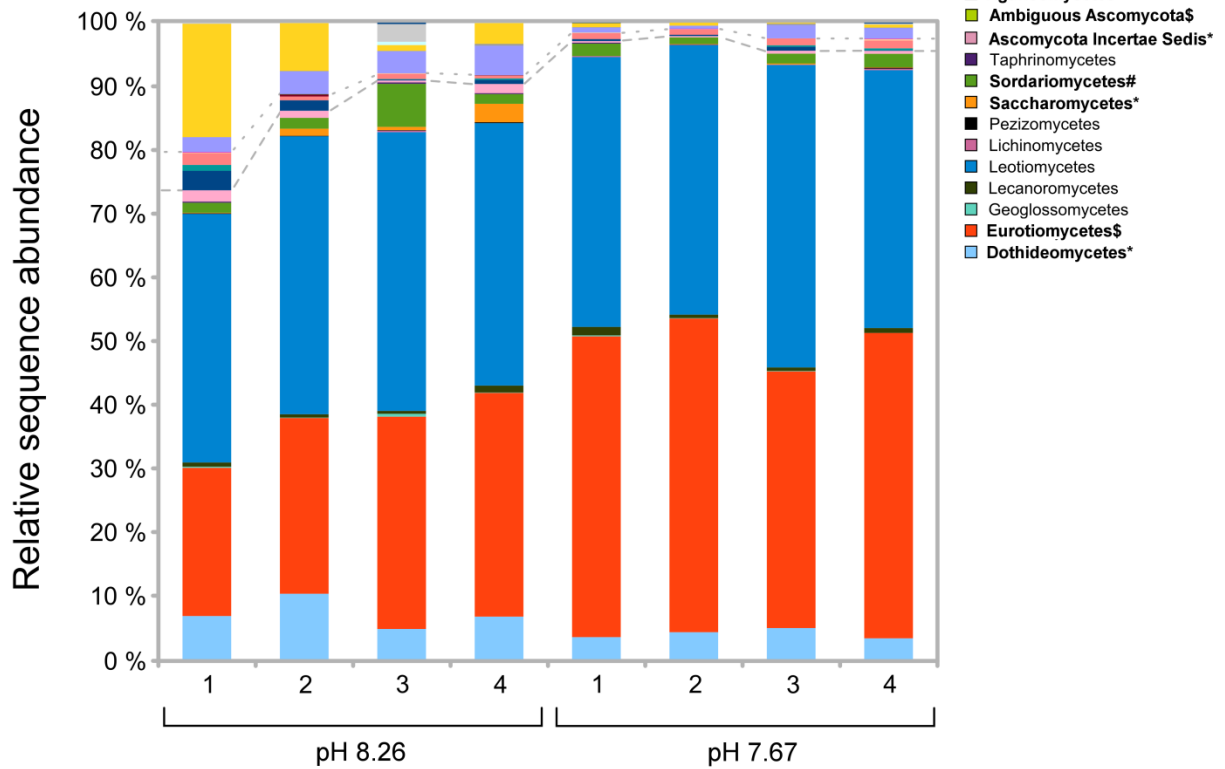
596 **Figure 4: Identification of phylogenetic signals via Phylocom analysis testing for clade-specific**
597 **adaptation to a pH of 8.26 and 7.67.** (A) Comstruct analysis revealed significant phylogenetic
598 clustering in samples for both pH treatments. (B) Nodesig analysis identified the fungal groups that
599 caused the observed phylogenetic signals. The schematic phylogenetic tree depicts only the fungal
600 clades that significantly contributed to the observed phylogenetic clustering. Identified significant
601 clustering was assigned to class/subphylum (left) and order levels (right). NRI, net relatedness index;
602 NTI, nearest taxon index; blue, clustering effect present in both pH sample types; red, clustering
603 effect observed only in the sample type with a pH of 8.26; green, effect only observed in sample type
604 with a pH of 7.67.

605 **Figure 1:**

A)



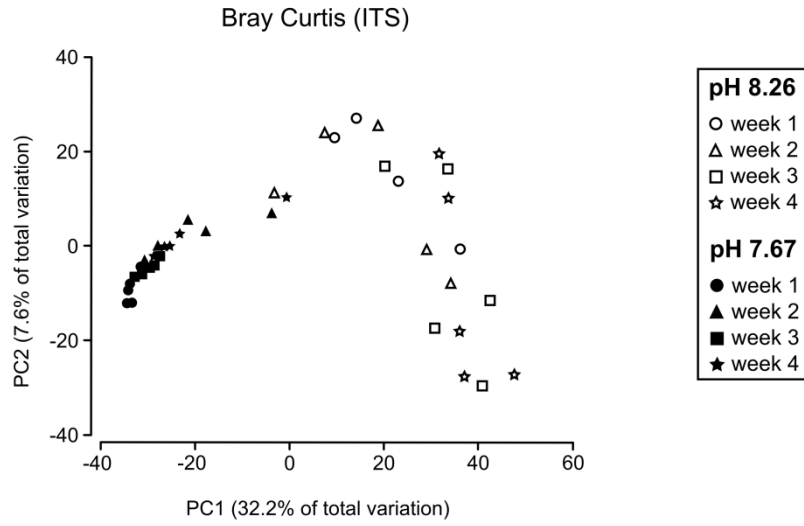
B)



606

607 **Figure 2:**

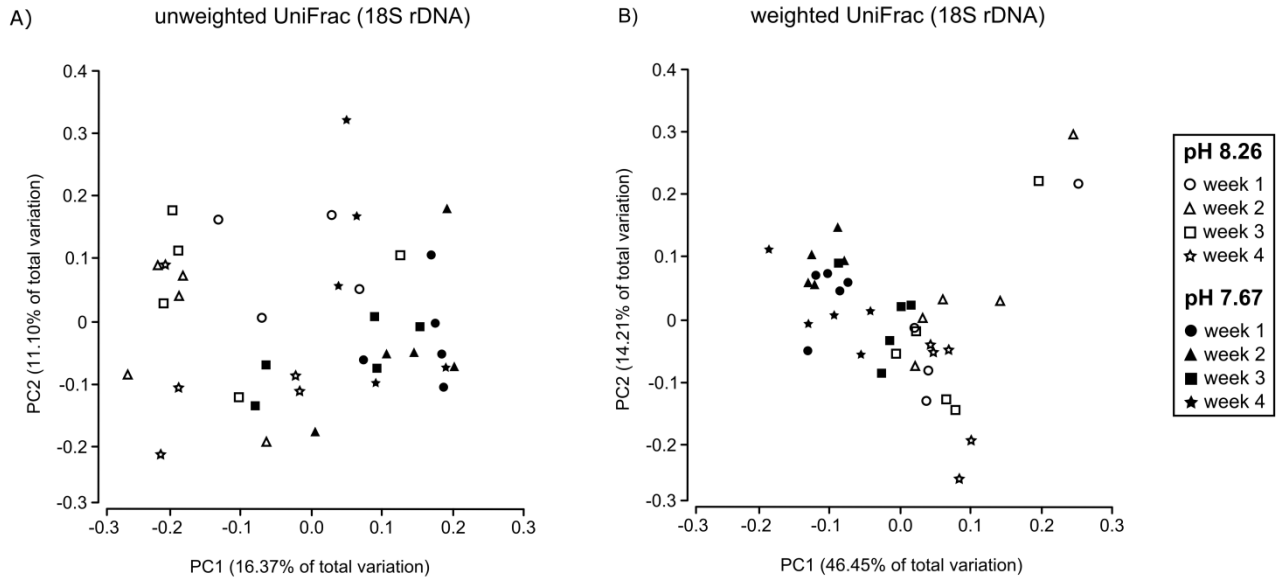
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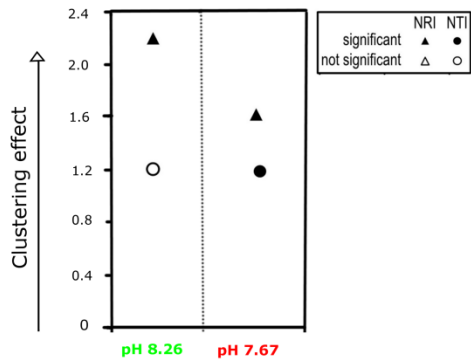
610 **Figure 3:**

611



612

A)



B)

