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## Diet management to effectively abate N<sub>2</sub>O emissions from surface applied pig slurry

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25 Benzoic acid was negatively correlated with N<sub>2</sub>O emission for slurries from OP diets,  
26 which had over double the hippuric acid content, and more than 1.8 times the benzoic  
27 acid content than the CM. However, this effect only occurred during the first week due  
28 to rapid degradation of this compound within soil. The possible toxic effect of benzoic  
29 acid did not appear to affect soil respiration, since a positive correlation was found.  
30 Results of a benzoic acid balance (considering both intake through feed and release  
31 through urine) indicated that the source of both acids were phenolic compounds  
32 (polyphenolic or lignin) present in the fibrous fraction. These results show that N<sub>2</sub>O  
33 emissions are more affected than CO<sub>2</sub> by to compounds within urine/faeces that can be  
34 manipulated indirectly through the diet.

35

36 Keywords: nitrous oxide emissions, carbon dioxide emissions, diet manipulation, diet  
37 fibre content

38

## 39 **1. Introduction**

40 Manures from intensive pig production systems are generally inefficiently recycled  
41 and can potentially lead to atmospheric pollution, from the entire manure management  
42 chain, including housing (IPCC, 2007). Nitrous oxide (N<sub>2</sub>O) is one of the main  
43 pollutants emitted after application of pig slurries to agricultural soils (Aguilera et al.,  
44 2013). This gas is a by-product of soil biochemical processes, mainly nitrification and  
45 denitrification (Firestone and Davidson, 1989). These processes are directly controlled  
46 by soil moisture, with low values of water-filled pore space (WFPS) favoring  
47 nitrification (WFPS<60%) and high values (WFPS>60%) being suitable for  
48 denitrification (Sanchez-Martin et al., 2010a). Emissions of N<sub>2</sub>O contribute  
49 considerably to the radiative forcing of the atmosphere, having a global warming

50 potential 298 times higher than that of CO<sub>2</sub> expressed on a weight basis (i.e. per kg)  
51 over a 100 year timescale (IPCC, 2007).

52 Ammonium and other labile N compounds of slurry (e.g. urea, creatinine) can  
53 influence the total N<sub>2</sub>O emissions when applied to agricultural soils (Whitehead et al.,  
54 1989). When slurries are applied to aerated soils, NH<sub>4</sub><sup>+</sup> is rapidly nitrified, producing  
55 large fluxes of N<sub>2</sub>O (Sanchez-Martin et al., 2010b). Additionally, the nitrate (NO<sub>3</sub><sup>-</sup>)  
56 obtained from nitrification in soils, together with degradable organic C compounds  
57 added with manures, often accelerates denitrification, especially under conditions of  
58 high soil WFPS (> 60%) (Cardenas et al., 2007).

59 Diet manipulation has been identified as a promising technique to modify urine  
60 and fecal composition (Hansen et al., 2014) and consequently this is proposed as an  
61 acceptable strategy to reduce environmental N pollution. Available information  
62 indicates that reducing crude protein (CP) in the diet of growing-finishing pigs can lead  
63 to an 8% reduction in N excretion for each percentage decrease in CP of the feed,  
64 without reduction of animal performance (Galassi et al., 2010). Moreover, the addition  
65 of fibrous feedstuffs to the diet has been related to a reduction of urea-N excretion  
66 (Jarret et al., 2011). At present, information regarding the influence of pig diet on slurry  
67 composition and its effect on N<sub>2</sub>O emission is scarce. Velthof et al. (2004) found that  
68 pig diet modified pig manure composition, but its effect on N<sub>2</sub>O was dependent on soil  
69 properties, especially soil organic matter content.

70 In laboratory experiments, it has been shown that minor constituents of ruminant  
71 urine, such as hippuric acid (HA) and benzoic acid (BA), can contribute to decreased  
72 N<sub>2</sub>O emissions (Kool et al., 2006a-b). van Groenigen et al. (2006) observed an effect of  
73 these aromatic compounds on denitrification, whereas, Bertram et al. (2009) found  
74 nitrification was also partially inhibited by these organic compounds. These findings

75 suggest an increase of hippuric and benzoic acid within ruminant urine could potentially  
76 be a N<sub>2</sub>O mitigation strategy, however, contradictory results have been obtained in the  
77 few field experiments conducted to date, with no inhibitory effects detected (Clough et  
78 al., 2009; Krol et al., 2015). Further research is required to determine whether  
79 manipulation of pig diet to enhance the production of these minor urinary compounds  
80 can reduce subsequent N<sub>2</sub>O emissions.

81         Within this context, two combined experiments were conducted, investigating  
82 the manipulation of pig diet, to assess the consequent effects on slurry composition, and  
83 the application of these slurries as N fertilizer to grassland mesocosms. The aims of the  
84 study were to: 1) evaluate the effect of the inclusion of fibrous by-products in pig's diet  
85 on hippuric and benzoic acid concentrations in the excreted urine/slurry, and 2)  
86 determine the possible effect of both acids on N<sub>2</sub>O emissions following application of  
87 the slurries to grassland soil. We predict that diets resulting in slurries with higher  
88 organic acids (hippuric and benzoic) concentration would result in lower soil N<sub>2</sub>O  
89 emissions following their application.

90

## 91 **2. Material and Methods**

### 92 *2.1 Selection of pig slurries.*

93         Thirty growing-finishing pigs, progeny of Danish Duroc × (Landrace × Large  
94 White), were fed five different diets, under controlled conditions, in an experimental  
95 farm in Castellón (Spain). The experimental conditions to obtain urine and faeces from  
96 each animal, and the collection period, are explained in detail in Beccaccia et al. (2015).  
97 The experimental feeds included a conventional diet, formulated to contain the most  
98 common ingredients used in commercial diets for growing-finishing pigs (wheat, barley

99 and soybean meal), and either orange pulp (OP) or carob meal (CM), at two dietary  
100 concentrations (75 or 150 g kg<sup>-1</sup>), in replacement of barley grain as a fibrous by-product.  
101 The diets were as follows: 1) pig slurry control (PSC), 2) 75 g kg<sup>-1</sup> of carob meal (CM-  
102 75), 3) 150 g kg<sup>-1</sup> of carob meal (CM-150), 4) 75 g kg<sup>-1</sup> orange pulp (OP-75), and 5)  
103 150 g kg<sup>-1</sup> of orange pulp (OP-150). These diets were designed to modify slurry  
104 composition by changing dietary fibre sources, but without an influence on pig  
105 performance or health (Beccaccia et al., 2015). In order to maintain constant neutral  
106 detergent fibre (NDF) dietary level, net energy, protein and essential amino acid levels  
107 between diets, lard, soybean meal and synthetic amino acids were added to the feeds  
108 including fibrous by-products. Essential nutrients were formulated according to the  
109 recommendations of FEDNA (2006).

110 Urine and faeces were collected and stored (covered) separately at -20°C in  
111 closed plastic bottles. Following the usual pig slurry's management, urine and faeces  
112 from each animal were mixed in the same proportions as they were excreted.  
113 Subsequently, slurries of four different animals fed the same diet were immediately  
114 mixed, in order to obtain the manure used in the mesocosms experiment. In order to  
115 consider the possible side effects of freeze-thaw on the constituents, pH, NH<sub>4</sub><sup>+</sup> and total  
116 N (Nt) were determined before application of the reconstituted slurry to soil. Standard  
117 electrodes were used to determine pH and titration of the liquid fraction and the  
118 Kjeldahl method to NH<sub>4</sub><sup>+</sup> and Nt, respectively. Total N content was used to calculate the  
119 amount of slurry to apply in the mesocosms experiment, in order to achieve equal N  
120 application rates.

121 Purine derivatives (allantoin, creatinine, uric acid, hippuric acid and benzoic  
122 acid) were analyzed directly in urine samples from the different pigs (n = 46) used for

123 each diet via high performance liquid chromatography (HPLC) on a Varian Pro Star 310  
124 HPLC system (Varian Inc., Palo Alto, CA), using a Phenomenex Luna® 5 µm SCX  
125 100Å column (250 × 4.6 mm) a variable wavelength detection set at 218 nm and a flow  
126 rate of 0.7 ml min<sup>-1</sup>. The method consisted of two mobile phases: mobile phase A  
127 (KH<sub>2</sub>PO<sub>4</sub>; 17 g L<sup>-1</sup>; adjusted to pH=4) and mobile phase B (40% methanol: 60% mobile  
128 phase A). The samples were centrifuged, and prepared for analysis in HPLC vials (1:10,  
129 urine: mobile phase A) and mixed with a vortex mixer before analysis.

130 Additionally, benzoic acid and total polyphenolic compounds were also  
131 determined in OP. Benzoic acid was extracted and determined by micellar  
132 electrophoresis as indicated in Ding et al. (2015), and polyphenolic compounds were  
133 determined using the Folin–Ciocalteu method (Obanda and Owuor, 1997).

## 134 2.2 Mesocosms experiment

135 In an experimental greenhouse of the Technical University of Madrid farm, 36  
136 PVC cylindrical containers (26 cm diameter, 15 cm height) were used for the  
137 mesocosms experiment. Each container was filled with 6 kg of dry soil (7cm height).  
138 The soil was previously collected randomly from a 700 m<sup>2</sup> area, at the experimental  
139 field station ‘El Encín’ (40° 32’N, longitude 3° 17’W) from 0 - 25 cm soil depth. In the  
140 laboratory, soil was air-dried at 20 °C, sieved through a 2 mm mesh and repeatedly  
141 mixed to ensure homogeneity. Some physico-chemical properties of the top 0–25 cm of  
142 the soil layer, measured by standard methods of soil analysis (Burt, 2004) were: total  
143 organic C, 8.2 ± 0.4 g kg<sup>-1</sup>; pH<sub>H2O</sub>, 7.8 and CaCO<sub>3</sub>, 13.1 ± 0.3 g kg<sup>-1</sup>. According to Soil  
144 Survey Staff, 1992, the soil used was a *Calcic Haploxerept* with a clay loam texture  
145 (clay, 28%; silt, 17%; sand, 55%). At the beginning of the experiment, the soil mineral  
146 N content was 1.1 and 10.2 mg N kg soil<sup>-1</sup> for NH<sub>4</sub><sup>+</sup> and NO<sub>3</sub><sup>-</sup>, respectively. The

147 temperature was maintained between 10 and 15°C. Ryegrass was sown and the soil  
148 moisture content was brought to 60% water-holding capacity (WHC) during 15 days  
149 before applying treatments. The containers were maintained at this WHC level  
150 throughout the experiment, by replacing the weight loss with distilled water, on a daily  
151 basis.

152 Perennial ryegrass (*Lolium perenne L.*) seed was sown at a density of 6 g of seed  
153 per container. Two weeks after sowing, the seedlings were clipped to 3 cm above the  
154 soil surface, following which the pig slurries were applied (i.e., day 0) to the soil  
155 surface. The slurry was applied at an equal N application rate of 100 kg N ha<sup>-1</sup> (0.531 g  
156 N container<sup>-1</sup>). Due to the different N content of the slurries, different total amounts of  
157 slurry were applied to the containers, as follows: 55.2, 61.2, 67.9, 66.8 and 54.7 g of  
158 slurry per pot for PSC, OP-75, OP-150, CM-75 and CM-150, respectively. Additionally,  
159 water was added to the slurry application, in order to apply an equal total volume of 75  
160 mL container<sup>-1</sup> across all treatments. The amount of NH<sub>4</sub><sup>+</sup> added to different treatments  
161 were: 0.293, 0.274, 0.235, 0.250 and 0.238 g N container<sup>-1</sup> for the same treatments,  
162 respectively. The control treatment received the same water and seed density as the rest  
163 of the treatments but it did not receive any N application.

164 The experiment was arranged in a factorial randomized complete block design,  
165 with six containers for each treatment. Half of the pots were used for GHG emission  
166 measurements (non-destructive;  $n = 3$ ), and the other half were used in order to sample  
167 the soil (destructive;  $n = 3$ ).

168

169 *2.3 Soil sampling and analysis*



170 In the pots designed for soil sampling, two soil cores per container were taken  
171 for each sample date, using a 1 cm diameter soil auger (10 cm long). The hole produced  
172 by the auger was filled with sand in order to maintain the stability of the soil structure.  
173 Soil samples were analyzed for dissolved organic C (DOC), extractable mineral N  
174 ( $\text{NH}_4^+$  and  $\text{NO}_3^-$ ) and soil moisture. Soil DOC was determined by extracting 8 g soil  
175 with 50 mL of deionized water. Afterward, DOC was analyzed with a total organic  
176 carbon analyser (multi N/C 3100 Analytik Jena, Jena, Germany). From another 8 g of  
177 homogeneously mixed soil,  $\text{NH}_4^+$ -N and  $\text{NO}_3^-$ -N were extracted with 50 mL of KCl (1  
178 M) over a 1 hour period, filtered and measured by automated colorimetric determination  
179 using a flow injection analyzer (FIAS 400 Perkin Elmer, USA) provided with a UV-V  
180 spectrophotometer detector.

181 Water-filled pore space (WFPS) was estimated by dividing the volumetric water  
182 content by total soil porosity. Total soil porosity was calculated by measuring the bulk  
183 density of the soil according to the relationship: soil porosity =  $1 - (\text{soil bulk}$   
184  $\text{density}/2.65)$ , assuming a particle density of  $2.65 \text{ g cm}^{-3}$  (Danielson and Sutherland,  
185 1986). The bulk density, which was calculated from the volume of soil in the cores, was  
186  $1.29 \pm 0.1 \text{ Mg m}^{-3}$ . Soil samples were taken three days during the first two weeks after  
187 fertilizers applications. After the first week, samples were taken twice or once a week.

188

#### 189 *2.4 GHG sampling and analysis*

190 Emissions of GHG were measured using a closed dark static chamber approach,  
191 following the same sampling schedule as described in section 2.3. Each container was  
192 used as a chamber, closing it with a perfectly fitting lid for 40 minutes, resulting in a  
193 headspace of approx 8 L. The closure period was selected after testing, before the  
194 experiment, the linearity of the increasing gas concentrations of  $\text{N}_2\text{O}$  and  $\text{CO}_2$  inside the

195 chamber (Ábalos et al., 2014). Gas samples were taken at 0, 20 and 40 min through a  
196 three way valve, which was installed in the lid. Gas samples were taken using a 100 ml  
197 syringe and stored in 20 ml chromatography vials prior to analysis.

198 Concentrations of N<sub>2</sub>O, and CO<sub>2</sub> were quantified by gas chromatography, using  
199 a HP-6890 gas chromatograph (GC; Agilent Technologies, Barcelona, Spain) equipped  
200 with a Turbomatrix autoanalyzer (Perkin Elmer, Madrid, Spain). Gas samples were  
201 injected through HP Plot-Q capillary columns to a <sup>63</sup>Ni electron-capture detector (ECD)  
202 to analyze N<sub>2</sub>O concentrations and to a flame-ionization detector (FID) fitted with a  
203 methanizer for CO<sub>2</sub> concentrations. Helium was used as carrier gas and the oven was  
204 operated at a constant temperature of 35 °C. Greenhouse gas flux rates were calculated  
205 from the change in gas concentration in the headspace air during the sampling period,  
206 where total N<sub>2</sub>O-N and CO<sub>2</sub>-C fluxes per container were estimated by successive linear  
207 interpolations of the flux measurements.

208

### 209 *2.5 Biomass sampling and analysis*

210 Grass was cut (to a height of 3 cm) two weeks after the beginning of the  
211 experiment and at the end of the experimental period. The biomass obtained from these  
212 two cuts were weighed in order to obtain the yield, and total foliar C and N content  
213 were determined with an elemental analyzer (TruMac CN Leco, USA).

214

### 215 *2.6 Calculations and statistical analysis*

216 For the effect of diet on minor chemical constituents of pig urine, individual  
217 animals (*n*=6 per treatment) were the experimental unit for all the diet treatments  
218 studied. Data were analyzed as a completely randomized design with type of diet as  
219 main factor, by using PROC GLM of SAS (2008). The effects of type of diet on urine

220 composition were analyzed as a factorial arrangement by using orthogonal contrasts  
221 with source and level of inclusion of fibrous by products as main factors. Contrasts of  
222 each of the experimental treatments against the control diet were done by using a  
223 Dunnet test. For the effect of different slurries on soil N<sub>2</sub>O and CO<sub>2</sub> emission, each  
224 container ( $n=3$  per treatment) was the experimental unit. Differences between  
225 treatments at each sampling event and between the mean and cumulative emissions  
226 were evaluated using analysis of variance (ANOVA,  $P < 0.05$ ). The least significant  
227 difference (LSD) test was used for multiple comparisons between means. Prior to the  
228 statistical tests the data were analyzed to determine whether the assumptions of  
229 normality (Kolmogorov–Smirnov test) and equality of variance (Levene's test) were  
230 satisfied. Where needed to fulfill these assumptions, the data were log-transformed  
231 before analysis. Cumulative N<sub>2</sub>O and CO<sub>2</sub> fluxes were estimated by successive linear  
232 interpolation between weekly sampling dates to study the possible effect of benzoic acid  
233 on the emissions due to the rapid degradation of this compound. Correlations between  
234 total N<sub>2</sub>O and slurry chemical constituents, such as hippuric and benzoic acid or NH<sub>4</sub><sup>+</sup>  
235 content, were also performed at these periods during the experiment. Other correlation  
236 between N<sub>2</sub>O and CO<sub>2</sub> fluxes with soil parameters such as NH<sub>4</sub><sup>+</sup>-N, NO<sub>3</sub><sup>-</sup>-N, DOC were  
237 performed, with a 95% significant level.

238

### 239 **3. Results**

#### 240 *3.1 Composition and effect of diet on some urine chemical constituents*

241 The chemical composition of the experimental diets can be seen in Table 1, and  
242 are further described in detail in Beccaccia et al. (2015). Feeds containing CM generally  
243 had a higher concentration of acid detergent lignin (ADL) compared to PSC, OP-75 and

244 OP-150, and feeds containing OP had a greater soluble fibre content compared to PSC,  
245 CM-75 and CM-150.

246 The composition of the slurries from the contrasting dietary treatments can be  
247 seen in Table 2. The PSC diet resulted in the slurry with the highest pH and  $\text{NH}_4^+$   
248 content, whereas, the OP-150 diet resulted in a slurry with the lowest  $\text{NH}_4^+$  content (3.5  
249  $\text{mg N L}^{-1}$ ).

250 Some minor N-containing urine compounds, such as creatinine or allantoin,  
251 were not affected by the source and level of dietary fibre by-product, however, the  
252 amount of hippuric and benzoic acid was related to the type of fibre included in diets  
253 (Table 3). In fact, the OP urines (OP-75 and OP-150) contained the highest amount ( $P$   
254  $<0.001$ ) of hippuric acid and tended ( $P = 0.072$ ) to contain a high content of benzoic  
255 acid than those obtained from CM diets. Urine from OP diets had more than double the  
256 hippuric acid, and more than 1.8 times the benzoic acid content than with CM-75 and  
257 CM-150. Hippuric acid concentration increased with the amount of OP (75  $\text{g kg}^{-1}$  or 150  
258  $\text{g kg}^{-1}$ ), but the same trend was not observed for benzoic acid. The concentration of uric  
259 acid ranged from 40 to 69  $\text{mg L}^{-1}$  for CM-150 and PSC respectively.

260 The OP contained 30  $\text{mg benzoic acid kg}^{-1}$  and 20  $\text{g galic acid kg}^{-1}$  (total  
261 polyphenol). Taken into account the percentage of this ingredient in feeds, the dietary  
262 treatment OP-75 contained 2.25  $\text{mg benzoic acid kg}^{-1}$  and 1.5  $\text{g galic acid kg}^{-1}$ , and the  
263 dietary treatment OP-150 contained 4.5  $\text{mg benzoic acid kg}^{-1}$  and 3.0  $\text{g galic acid kg}^{-1}$ .

264

265 *3.2 Nitrous oxide emissions*

266 Nitrous oxide emissions began to increase in the slurry treatments 3 days  
267 following application (Fig. 1). At 7 to 17 days following treatment application,  
268 emissions were higher than  $1 \text{ mg N}_2\text{O-N m}^{-2} \text{ d}^{-1}$  for all slurry treatments (Fig. 1).  
269 Although during some of the sampling days there were no significant differences  
270 between treatments, a clear and significant ( $P < 0.05$ ) effect was observed during the  
271 highest period of  $\text{N}_2\text{O}$  flux. The highest fluxes were always measured from the PSC  
272 treatment, with a maximum peak on day 15 ( $4.3 \text{ mg N}_2\text{O-N m}^{-2} \text{ d}^{-1}$ ). In contrast,  
273 application of slurries obtained from pigs fed with OP diets (OP-75 and OP-150)  
274 resulted in lower fluxes than CM diets (CM-75 and CM-150) (Fig. 1). Significant  
275 differences ( $P < 0.05$ ) between the two levels of fibre (75 or  $150 \text{ g kg}^{-1}$ ) in each pig's  
276 diet were only found on day 15.

277 At the end of the experimental period, the  $\text{N}_2\text{O}$  emissions from the soils  
278 amended with slurries obtained from fibrous by-products diets, OP and CM, were 65  
279 and 47% (respectively) lower than from soil amended with the slurry obtained through a  
280 conventional pig diet. Considering cumulative  $\text{N}_2\text{O}$  emission by periods of a week,  
281 significant differences at  $P < 0.05$  between OP and CM were found for the first two  
282 weeks (Fig. 2a). However, for the 3<sup>rd</sup> and 4<sup>th</sup> weeks, both treatments produced similar  
283 cumulative  $\text{N}_2\text{O}$  fluxes.

284 A strong positive correlation was found between daily  $\text{N}_2\text{O}$  fluxes and soil  
285 mineral N content, where  $r = 0.52$ ,  $P < 0.01$ ,  $n = 39$  for  $\text{NO}_3^-$  and  $r = 0.33$ ,  $P = 0.04$ ,  $n =$   
286  $36$  for  $\text{NH}_4^+$ . Daily mean flux of  $\text{N}_2\text{O}$  was also positive correlated with DOC ( $P < 0.01$ ,  
287  $r=0.63$ ,  $n=54$ ). During the first week, mean  $\text{N}_2\text{O}$  cumulative fluxes were negatively  
288 correlated ( $P < 0.05$ ) with the amount of benzoic acid applied in slurries ( $r = -0.92$ ,  $n =$   
289  $5$ ).

290

### 291 4.3 Carbon dioxide emissions

292 Carbon dioxide emissions were measured under dark conditions to provide an  
293 indication of plant and soil processes (plant respiration + plant root respiration + soil  
294 microbial respiration).

295 Different pig's diets did not influence daily CO<sub>2</sub> emissions, which only showed  
296 significant differences ( $P < 0.05$ ) between the control and the slurry treatments on days  
297 9, 11, and 15 (Fig. 3). Overall, CO<sub>2</sub> fluxes from the different treatments showed a  
298 similar emissions pattern to that found for N<sub>2</sub>O, with the highest fluxes ( $>1000$  mg CO<sub>2</sub>-  
299 C m<sup>-2</sup> d<sup>-1</sup>) were found between 7 and 20 days following slurry application.

300 At the end of the experimental period, cumulative CO<sub>2</sub> fluxes from all slurry  
301 treatments were not significantly different ( $P > 0.05$ ) from the control. During the first  
302 week, however, significantly ( $P < 0.05$ ) higher CO<sub>2</sub> losses were found from OP-75, OP-  
303 150 and CM-150 treatments than from CM-75 or PSC (Fig. 2b). In the second week,  
304 treatments produced by fibrous by-products shown similar emissions than from  
305 conventional diet, PSC.

306 Mean daily CO<sub>2</sub> fluxes during the first week were positive correlated with  
307 benzoic acid ( $r = 0.95$ ;  $P < 0.05$ ;  $n = 5$ ) and hippuric acid ( $r = 0.93$ ;  $P < 0.01$ ;  $n=5$ )  
308 added with slurry. The cumulative CO<sub>2</sub> fluxes during the first week were also positively  
309 correlated with mean soil DOC ( $r = 0.88$ ;  $P < 0.05$ ;  $n = 5$ ).

### 310 4.5 Soil mineral N and DOC

311 Soil NH<sub>4</sub><sup>+</sup>-N concentration increased significantly with the addition of slurry  
312 (Fig. 4a). Generally, treatments from diets rich in fibrous by-products decreased the

313 mineral N within in the slurry, and consequently that extracted from the soil, but neither  
314 OP nor CM showed lower soil  $\text{NH}_4^+$  concentrations than PSC (Fig. 4a). Soil  $\text{NH}_4^+$   
315 content was higher than  $20 \text{ mg N kg}^{-1}$  during the first two weeks of the meso-cosm  
316 incubation, after which it returned to that of the control soil. The mean  $\text{NH}_4^+$   
317 concentration in soil was positively correlated with hippuric acid ( $r = 0.90$ ,  $P < 0.05$ ,  $n$   
318 = 5), but not with benzoic acid. The maximum soil  $\text{NO}_3^-$  content ( $106.3 \text{ mg N kg}^{-1}$ )  
319 appeared 20 days after slurry application (Fig. 4b), due to nitrification of soil  $\text{NH}_4^+$ .  
320 During the last two weeks of the experiment, the amount of soil  $\text{NO}_3^-$  was also higher  
321 than  $20 \text{ mg N kg}^{-1}$  but no significant ( $P < 0.05$ ) differences between dietary's treatments  
322 were found.

323 Soil DOC concentration ranged from  $73$  to  $194 \text{ mg C kg}^{-1}$  (Fig. 5a). The  
324 maximum concentration ( $358 \text{ mg C kg}^{-1}$ ) was found in OP-150 at the beginning of the  
325 experiment. Although there were no significant differences between treatments in the  
326 daily measurements, the trend was that the OP treatments (OP-75 and OP-150), showed  
327 the highest soil DOC content during the first 15 days (Fig. 5a). This is consistent with  
328 the average of the soil DOC during the experiment, where the highest concentration  
329 were from treatments rich in soluble fibre,  $122$  y  $149 \text{ mg C kg}^{-1}$  for OP-75 and OP-150,  
330 respectively (Fig. 5b).

#### 331 *4.6 Harvest yield*

332 In general, there was no significant effect of diet on the grass yield (Table 4).  
333 The grass biomass obtained in two cuts was higher in conventional than in the fibrous  
334 by-products-diets. The control showed a 27% less biomass compared with the diets  
335 treatments which indicated that the addition of a labile N source through pig slurry  
336 application had a clear effect on the yield, although independent of the type of diet.

337           The total N analysed in the grass biomass was significantly lower ( $P < 0.05$ ) in  
338 the control than in the rest of the slurries treatments (Table 4). The total N obtained in  
339 the grass from different pig's diet treatments ranged from 2.7 to 3.2 %, but there were  
340 no significant differences ( $P > 0.05$ ) between them.

341

342

## 343 **5. Discussion**

### 344 *5.1 Slurry composition and gas emissions*

345           Manipulation of pig diet has modified the composition of urine and faeces  
346 (Philippe et al., 2011), which can directly affect the emissions of  $N_2O$  from agricultural  
347 soils after manure is applied to soil as fertilizer. In this experiment, we demonstrated  
348 that slurries from pig fed with diets based on OP and CM produced lower emission of  
349  $N_2O$  than slurry produced from a conventional diet. Although the amount of total N  
350 added with slurries was the same in all containers, the emission factor over the  
351 experimental period (37 days) ranged from 0.32% CM-75 to 0.19% for OP-150,  
352 compared with 0.56% for the conventional slurry.

353           Slurry contains multiple compounds that could individually affect production or  
354 consumption of  $N_2O$  in soil after its application. One of the most important components  
355 is  $NH_4^+$  within the slurry, as it is the substrate for the process of nitrification, which is  
356 one of most important process involved in  $N_2O$  emission in Mediterranean cropping  
357 systems (Sahrawat and Keeny, 1986). The positive correlation found between the  $NH_4^+$   
358 content in the upper part of soil and  $N_2O$  fluxes are in consistent with that. However the  
359  $NH_4^+$  added with slurry was not enough to explain the important differences in fluxes,  
360 because the OP-150 and OP-75, which received 20% and 7% less  $NH_4^+$  than PSC,



361 respectively, resulted in <50% of total N<sub>2</sub>O in comparison to that of PSC. Probably  
362 others minor slurry compounds could have an important effect on N<sub>2</sub>O emission. In fact,  
363 an interesting finding of this experiment was that only benzoic acid added with slurries  
364 explained differences in N<sub>2</sub>O fluxes. So, a negative correlation between benzoic acid  
365 added with pig slurry and total N<sub>2</sub>O emission was detected during the first 15 days after  
366 slurry application. Some laboratory studies reported that N<sub>2</sub>O losses can be mitigated by  
367 >50% by increasing hippuric acid concentration in cattle slurries (van Groenigen et al.,  
368 2006), but this effect has only been observed when in the presence of benzoic acid  
369 (Kool et al, 2006b), since this compound is a recognized antimicrobial agent (Marwan  
370 and Nagel, 1986) and it is the sub-product of the hippuric acid degradation. Our results  
371 partly corroborate the results from these studies, as the treatments which had the highest  
372 amounts of both acids (OP-75 and OP-150) produced the lowest N<sub>2</sub>O emissions.  
373 However, the higher concentration of hippuric acid of OP-150 in comparison to that of  
374 OP-75 produced a similar total N<sub>2</sub>O emission in both treatments. Neither, concentration  
375 of hippuric acid in slurries explained differences in fluxes between CM and PSC  
376 treatments, because N<sub>2</sub>O emission were lower in the CM treatments (CM-75 and CM-  
377 150) than in PSC, despite the higher urine hippuric acid concentration from the PSC  
378 diet.

379         There are contradictory results in the literature regarding the effect of hippuric  
380 and benzoic acid on N<sub>2</sub>O emissions, as studies carried out under field conditions did not  
381 find significant differences in N<sub>2</sub>O emission from cattle urine with different  
382 concentrations of both acids (Clough et al., 2009; Krol et al., 2015). The first study  
383 argued that the low WFPS (< 35%) and the high pH (> 6.4) of the soil during the  
384 experimental period were not appropriate conditions to promote the inhibitory effect of  
385 hippuric and benzoic acid on N<sub>2</sub>O emissions. In the case of Krol et al. (2015), the

386 authors suggest that both organic acids lose their power to mitigate N<sub>2</sub>O emissions as a  
387 consequence of other effects such as leaching, and plant root activity, which are not  
388 realistically represented in laboratory incubations that do not contain plants, or have  
389 sufficient soil depth. Therefore, this study is the first to show a mitigation effect, likely  
390 to be associated with benzoic acid concentration within slurry, where both soil  
391 microorganism and plants can compete for the available N. However, the possible  
392 inhibitory effect on processes producing N<sub>2</sub>O emission was only maintained for a short  
393 period of time in this soil, as deduced from the negative correlation during the first two  
394 weeks between cumulative N<sub>2</sub>O emission and benzoic acid. According to Clough et al.  
395 (2009), high soil pH could promote a rapid degradation of benzoic acid into its  
396 conjugated base (benzoate) diminishing its inhibitory effect. In our experimental  
397 conditions, with high pH both in slurries and soil, the conditions for degradation were  
398 very favourable. Benzoic acid could be metabolized in agricultural soils by some  
399 microorganisms (*Pseudomonas* and *Burkholderia* species), as demonstrated Pumphrey  
400 and Madsen (2008), reducing its concentration after addition. In this study, slurry was  
401 applied on the soil surface following the most common practices in the field. Therefore,  
402 C and N compounds added with slurry were concentrated on the upper part of the soil  
403 where the WFPS ranged from 60-70% on most of the sampling dates. This situation  
404 may have favoured N<sub>2</sub>O production via denitrification, although, the rapid decrease in  
405 soil NH<sub>4</sub><sup>+</sup> concentration during the first two weeks of the study indicates that  
406 nitrification was also taking place (Bateman and Baggs, 2005). The inhibitory effect of  
407 benzoic acid seems to affect the denitrification process (van Groenigen et al. 2006,  
408 Bertram et al. 2009), although nitrification could also be affected (Bertram et al. 2009).  
409 Taking into account that both nitrification and denitrification processes contributed to  
410 the overall N<sub>2</sub>O emissions from the mesocosms. We speculate that in this small soil

411 volume, the amount of benzoic acid (and hippuric acid) added with treatments, such as  
412 OP-75 and OP-150, could have been enough to partially inhibit N<sub>2</sub>O production.  
413 However, this needs to be checked with additional experiments.

414 An interesting result was the significant correlation between added hippuric acid  
415 with slurries and mean NH<sub>4</sub><sup>+</sup> concentration in soil, which could indicate that  
416 nitrification rate was retarded by hippuric acid, and therefore the amount of N<sub>2</sub>O  
417 emission from this process. Also Bertram et al. (2009) suggested that hippuric acid  
418 reduced the activity of nitrifiers, however additional experiments are needed to confirm  
419 this. Additionally, the higher DOC concentrations and the higher CO<sub>2</sub> soil respiration  
420 observed for OP treatments after slurry application could have contributed to increased  
421 electron demand for denitrifiers, favouring the consumption of N<sub>2</sub>O and consequently  
422 the reduction of N<sub>2</sub>O/N<sub>2</sub> ratio. This effect, which has been observed when a source of  
423 labile C was added to soil (Cardenas et al., 2007), also contributes to a reduction in total  
424 N<sub>2</sub>O emissions.

425 Hippuric acid as well as benzoic acid could have been used by soil  
426 microorganisms, enhancing soil respiration rate as indicated by the positive correlation  
427 found between cumulative CO<sub>2</sub> and added hippuric acid during the first two weeks after  
428 slurry application. Based on this finding, it is possible to indicate that both acids did not  
429 have a general inhibitory effect of microbial activity, at least at these concentrations,  
430 although as discussed before, benzoic acid added with slurry or derived from hippuric  
431 acid degradation in soil could have affected denitrification and nitrification activity.

432

### 433 *5.2 Diets and N<sub>2</sub>O emission in soil*

434 Our results demonstrated that by manipulating pig diet, it is possible to modify  
435 urine and faeces composition and subsequently reduce N<sub>2</sub>O emissions following soil

436 application of the slurry. As slurries with the highest amount of benzoic acid produced  
437 the lowest N<sub>2</sub>O emission, strategies based on increasing concentration of this compound  
438 in slurry (as well hippuric acid) could be considered as a potential option to mitigate  
439 N<sub>2</sub>O emissions from slurry applications.

440           Organic acids and their salts, such as benzoic acid, are used in monogastric  
441 animal nutrition as alternatives to antibiotic growth promoters (Hansen et al., 2007).  
442 benzoic acid is absorbed in the small intestine, and metabolized in liver producing  
443 hippuric acid, which is subsequently excreted in the urine (Bridges et al., 1970).  
444 Murphy et al. (2011) observed a linear increase of N retained/intake when pig diet was  
445 supplemented with benzoic acid in the range 0 to 30 g kg<sup>-1</sup>. This was attributed to a  
446 reduction in the total aerobic bacteria in the ileum, thus increasing digestibility. Also the  
447 lowering of pH in the gastrointestinal microbiota improves N absorption (Sauer et al.,  
448 2009). This effect could have contributed to lower NH<sub>4</sub><sup>+</sup> in the OP-150 treatment.

449           Another possible cause of the reduction of the amount of NH<sub>4</sub><sup>+</sup> in slurry, and its  
450 subsequent effect on fluxes, was through the effect of increasing soluble fibre in the  
451 diets (Beccaccia et al., 2015). These authors found that increasing soluble fibre,  
452 through incorporating OP within feeds, reduced the total N excreted via urine and the  
453 urine:faeces N ratio was reduced. This effect was caused as consequence of a  
454 consumption of N by microorganisms, which transformed soluble N, such as urea of  
455 urine, into organic N. This last fraction (organic N) which is mainly included in faeces,  
456 is normally mineralized more slowly in soil than soluble N, reducing the risk of  
457 volatilization in the following days after application. In fact, Beccaccia et al. (2015)  
458 found an important reduction of NH<sub>3</sub> emissions from diets with high percentage on fibre  
459 and lignin such as OP.

460 Analyzing the component of diets, and considering a mean of 2.4 L urine pig<sup>-1</sup>  
461 day<sup>-1</sup> excreted (in this experiment), benzoic or hippuric acid excreted in urine for OP-  
462 150 diet was 0.35 g BA and 3.96 g HA pig<sup>-1</sup> day<sup>-1</sup>, respectively. As the mean  
463 consumption of feed per pig per day was 2.5 kg in this experiment, only 11.25 mg  
464 benzoic acid was included in the feed for the OP-150 treatment. Therefore, other  
465 compounds (e.g. polyphenolic compounds) are necessary as a source of excreted  
466 benzoic acid. So, polyphenolic compounds included in OP, and ingested by the pig was  
467 5.5 g galic acid pig<sup>-1</sup> day<sup>-1</sup> (OP-150), which was enough to generate the additional  
468 benzoic acid excreted in the urine.

469 However, more studies are needed to understand how degradation of  
470 polyphenolic compounds in the intestine generates benzoic acid in pig excreta.

471 To date, there is scarce literature (Petersen et al., 2013; Eriksen et al., 2014)  
472 regarding the use of slurries/urines with manipulated hippuric or benzoic acid content,  
473 and most of these have been conducted by increasing their concentration through direct  
474 addition of these compounds into the slurry (Fangueiro et al., 2015). However, this  
475 practice could be difficult to achieve due to the low solubility of both compounds (Krol  
476 et al., 2015). The manipulation of diet, as demonstrated in this experiment, provides an  
477 available strategy for increasing these compounds within the excreta (Dijkstra et al.,  
478 2013).

479 Meat producing countries, such as Spain, which is the 4<sup>th</sup> largest producer of  
480 pork in the world, need to develop strategies for sustainable pig meat production in  
481 order to decrease the release of N pollutants to the environment. Modifying pig diet  
482 using sub-products rich in fibre, such as OP or CM, is a potentially economically viable  
483 strategy to reduce N<sub>2</sub>O since fibre rich feed ingredients are often cheaper than usual

484 ones and can reduce the competition with food (cereals) for human nutrition. However,  
485 these products are many times locally produced and seasonally available and the  
486  $N_2O$  emissions, only represents 1-5% of the total N applied from the slurries  
487 which will not result in increased the nitrogen used efficiency (NUE). If this effect were  
488 combined with a reduction in nitrate leaching, then it could become important as an  
489 increase the N use efficiency of the meat sector butore studies should be necessary to  
490 achieve these challenges.

491

## 492 **6. Conclusion**

493 Changes in dietary fibre composition, as a consequence of including fibrous by-  
494 products, had an important effect on the concentration of urine and faeces compounds.  
495 Benzoic and hippuric acid concentrations in urine were related to the type of fibrous by-  
496 product in the diet, being higher for OP than for CM or barley grain. Results of a  
497 benzoic acid balance considering both intake through feed and release through urine  
498 indicated that the source of this acid and its precursor (i.e. hippuric acid) should be  
499 phenolic compounds (other than benzoic acid), probably associated with the  
500 polyphenolic or lignin content in the fibrous fraction.

501 The composition of slurry also had an important effect on  $N_2O$ . Emission of this  
502 gas was correlated with the benzoic acid added with urine, but not directly with hippuric  
503 acid concentrations. Under denitrification favoring condition (WFPS close to 70%), the  
504 inhibitory effect was only observed for 15 days following slurry application, probably  
505 because of the degradation of this compound in soil within that period.

506 In contrast, microorganisms increased soil  $CO_2$  emissions in these first two  
507 weeks from OP or CM treatments. This could indicate that there were not toxic effects  
508 of benzoic acid at this relatively low concentration on soil respiration.

509 Further knowledge is required on which compounds within urine and faeces  
510 have a natural inhibitory effect on denitrification or nitrification. Improving knowledge  
511 within this area will contribute to the range of approaches that can be used to mitigate  
512 greenhouse gas emission from livestock systems. These results show the potential of  
513 alternative feeding strategies for the reduction of environmental problems associated  
514 with agriculture, including the external dependency of raw material imports for feeding  
515 animals in Spain

516

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### 529 **References**

530 Abalos, D., De Deyn, G.B., Kuyper, T.W., van Groenigen, J.W., 2014. Plant species  
531 identity surpasses species richness as a key driver of N<sub>2</sub>O emissions from grassland.  
532 Glob. Change. Biol. 20, 265–275.

533 Aguilera, E., Lassaletta, L., Sanz-Cobena, A., Garnier, J., Vallejo, A., 2013. The poten-  
534 tial of organic fertilizers and water management to reduce N<sub>2</sub>O emissions in  
535 Mediterranean climate cropping systems. *Agric. Ecosyst. Environ.* 164, 32–52.

536 Bateman, E.J., Baggs, E.M., 2005. Contributions of nitrification and denitrification to  
537 N<sub>2</sub>O emissions from soils at different water-filled pore space. *Biol. Fertil. Soils.* 41,  
538 379–388.

539 Beccaccia, A., Calvet, S., Cerisuelo, A., Ferrer, P., García-Rebollar, P., De Blas, C.,  
540 2015. Effects of nutrition on digestion efficiency and gaseous emissions from slurry  
541 in growing-finishing pigs. I. Influence of the inclusion of two levels of orange pulp  
542 and carob meal in isofibrous diets. *Anim. Feed Sci. Technol.* 208, 158–169.

543 Bertram, J.E., Clough, T.J., Sherlock, R.R., Condrón, L.M., O’Callaghan, M., Wells, N.,  
544 Ray, J.L., 2009. Hippuric acid and benzoic acid inhibition of urine derived N<sub>2</sub>O  
545 emissions from soil. *Glob. Change Biol.*, 15, 2067–2077.

546 Bridges, J.W., French, M.R., Smith, R.L., Williams, R.T., 1970. The fate of benzoic  
547 acid in various species. *Biochem. J.* 118, 47–51.

548 Burt, R., 2004. (Eds.) *Soil Survey Laboratory Methods Manual*. Soil Survey  
549 Investigations Report No 42, Version 4.0, USDA, Washington, DC.

550 Cardenas, L.M., Chadwick, D., Scholefield, D., Fychan, R., Marley, C.L., Jones, R.,  
551 Bol, R., Well, R., Vallejo, A., 2007. The effect of diet manipulation on nitrous oxide  
552 and methane emissions from manure application to incubated grassland soils. *Atmos.*  
553 *Environ.*, 41, 7096–7107.

554 Clough, T.J., Kelliher, F.M., 2005. Dairy farm effluent effects on urine patch nitrous  
555 oxide and carbon dioxide emissions. *J. Environ. Qual.* 34, 979–986.

556 Clough, T.J., Ray, J.L., Buckthought, L.E., Calder, J., Baird, D., O’Callaghan, M.,  
557 Sherlock, R.R., Condrón, L.M., 2009. The mitigation potential of hippuric acid on



558 N<sub>2</sub>O emissions from urine patches: an in situ determination of its effect. *Soil Biol.*  
559 *Biochem.* 41 (10), 2222–2229.

560 Danielson, R.E., Sutherland, P.L., Klute, A., 1986. Porosity. *Methods of soil analysis.*  
561 Part 1. *Phys. Mineral. Methods* 443–461.

562 Dijkstra, J., Oenema, O., van Groenigen, J.W., Spek, J.W., van Vuuren, A.M.,  
563 Bannink, A., 2013. Diet effects on urine composition of cattle and N<sub>2</sub>O emissions.  
564 *Animal.* 7, 292–302.

565 Ding, X-J., Xie, N., Zhao, S., Wu, Y-C., Li, J., Wang, Z., 2015. Simultaneous  
566 determination of ten preservatives in ten kind of foods by micellar electrokinetic  
567 chromatography. *Food Chem.* 181, 207-217.

568 Eriksen, J., Nørgaard, J.V., Poulsen, H.D., Poulsen, H.V., Jensen, B.B., Petersen, S.O.,  
569 2014. Effects of Acidifying Pig Diets on Emissions of Ammonia, Methane, and  
570 Sulfur from Slurry during Storage. *J. Environ. Qual.* 43, 2086-2095.

571 Fanguero D., Hjorth, M., Gioelli, F., 2015. Acidification of animal slurry a review. *J.*  
572 *Environ. Manag.* 149, 46-56.

573 FEDNA., 2006 In: de Blas, C., Gasa, J., Mateos, G.G. (Eds), *Necesidades nutricionales*  
574 *para ganado porcino: normas FEDNA.* Fundación Española para el Desarrollo de la  
575 *Nutrición Animal.* Madrid. Spain, pp.55

576 Firestone, M.K., Davidson, E.A., 1989. Microbial basis of NO and N<sub>2</sub>O production and  
577 consumption in soil. In: Andreae, M.O., Schimel, D.S. (Eds.), *Exchange of Trace*  
578 *Gases between Terrestrial Ecosystems and the Atmosphere.* John Wiley & Sons Ltd,  
579 Berlin, Germany, pp. 7–21.

580 Galassi, G., Colombini, S., Malagutti, L., Crovetto, G.M., Rapetti, L., 2010. Effects of  
581 high fibre and low protein diets on performance, digestibility, nitrogen excretion and  
582 ammonia emission in the heavy pig. *Anim. Feed Sci. Tech.* 161, 140–148.

583 Hansen, M.J., Chwalibog, A., Tauson, A.H., 2007. Influence of different fibre sources  
584 in diets for growing pigs on chemical composition of faeces and slurry and ammonia  
585 emission from slurry. *Anim.Feed Sci. Tech.*, 134, 326–336.

586 Hansen, M.J., Nørgaard, J.V., Adamsen, A.P.S., Poulsen, H.D., 2014. Effect of reduced  
587 crude protein on ammonia, methane, and chemical odorants emitted from pig houses.  
588 *Livest. Sci.*, 169, 118-124.

589 IPCC., 2007. Climate change 2007: the Physical Science basis. Working Group I  
590 Report. In: Solomon, S., Qin, D., Manning, M., Chen, Z., Marquis, M., Averyt, K.B.,  
591 Tignor, M., Miller, H.L., (Eds.). Cambridge University Press, New York, NY, USA,  
592 pp. 996.

593 IPCC, 2007. Climate change. In: Synthesis Report of the Fourth Assessment Report of  
594 IPCC, (Chapter 3) pp. 49.

595 Jarret, G., Martinez, J., Dourmad, J.Y., 2011. Effect of biofuel co-products in pig diets  
596 on the excretory patterns of N and C on the subsequent ammonia and methane  
597 emissions from pig effluent. *Anim.* 5, 622–631.

598 Kool, D.M., Hoffland, E., Abrahamse, S.P.A., van Groenigen, J.W., 2006a. What  
599 artificial urine composition is adequate for simulating soil N<sub>2</sub>O fluxes and mineral N  
600 dynamics? *Soil Biol. Biochem*, 38, 1757–1763.

601 Kool, D.M., Hoffland, E., Hummelink, E.W.J., van Groenigen, J.W., 2006b. Increased  
602 hippuric acid content of urine can reduce soil N<sub>2</sub>O fluxes. *Soil Biol. Biochem.* 38,  
603 1021–1027.

604 Krol, D.J., Forrester, P.J., Lanigan, G.J., Richards, K.G., 2015. In situ N<sub>2</sub>O emissions  
605 are not mitigated by hippuric and benzoic acids under denitrifying conditions. *Sci.*  
606 *Total Environ.* 511, 362–368.

607 Marwan, A.G., Nagel, C.W., 1986. Quantitative determination of infinite inhibition  
608 concentrations of antimicrobial agents. *Appl. Environ. Microb.* 51, 559–561.

609 Murphy, D.P., O’Doherty, J.V., Boland, T.M., O’Shea, C.J., Callan, J.J., Pierce, K.M.,  
610 Lynch, M.B., 2011. The effect of benzoic acid concentration on nitrogen metabolism,  
611 manure ammonia and odour emissions in finishing pigs. *Anim. Feed Sci. Tech.* 163,  
612 194–199.

613 Obanda, M., Owuor, P.O., 1997. Flavanol composition and caffeine content of green leaf  
614 as quality potential indicators of Kenyan black teas. *J. Sci. Food Agri.* 74, 209–215.

615 Petersen, S.O., Dorno, N., Lindholm, S., Feilberg, A., Eriksen, J., 2013. Emissions of  
616 CH<sub>4</sub>, N<sub>2</sub>O, NH<sub>3</sub> and odorants from pig slurry during winter and summer storage.  
617 *Nutr. Cycling Agroecosyst.* 95, 103–113.

618 Philippe, F.X., Cabaraux, J.F., Nicks, B., 2011. Ammonia emissions from pig houses:  
619 Influencing factors and mitigation techniques. *Agric. Ecosyst. Environ.* 141, 245–  
620 260.

621 Pumphrey, G.M., Madsen, E.L., 2008. Field-based stable isotope probing reveals the  
622 identities of benzoic acid-metabolizing microorganisms and their in situ growth in  
623 agricultural soil. *Appl. Environ. Microb.* 74, 4111–4118.

624 Sahrawat, K.L., Keeny, D.R., 1986. Nitrous oxide emission from soils. In: *Advances in*  
625 *soil Science* 4. Springer, New York, Berlin Heidelberg, pp. 103–148.

626 Sanchez-Martín, L., Mejjide, A., Garcia-Torres, L., Vallejo, A., 2010b. Combination of  
627 drip irrigation and organic fertilizer for mitigating emissions of nitrogen oxides in  
628 semiarid climate. *Agr. Ecosyst. Environ.* 137, 99–107.

629 Sanchez-Martín, L., Sanz-cobena, A., Mejjide, A., Quemada, M., Vallejo, A., 2010a.  
630 The importance of the fallow period for N<sub>2</sub>O and CH<sub>4</sub> fluxes and nitrate leaching in a  
631 Mediterranean irrigated agroecosystem. *Eur. J. Soil Sci.* 61, 710–720.

632 Sauer, W., Cervantes, M., Yanez, J., Araiza, B., Murdoch, G., Morales, A., Zijlstra,  
633 R.T., 2009. Effect of dietary inclusion of benzoic acid on mineral balance in growing  
634 pigs. *Livest. Sci.* 122, 162–168.

635 van Groenigen, J.W., Palermo, V., Kool, D.M., Kuikman, P.J., 2006. Inhibition of  
636 denitrification and N<sub>2</sub>O emission by urine-derived benzoic and hippuric acid. *Soil*  
637 *Biol. Biochem.* 38, 2499–2502.

638 Velthof, G.L., Nelemans, J.A., Oenema, O., Kuikmana, P.J., 2004. Gaseous Nitrogen  
639 and Carbon Losses from Pig Manure Derived from Different Diets. *J. Environ. Qual.*  
640 34, 698-706

641 Whitehead, D.C., Lockyer, D.R., Raistrick, N., 1989. Volatilization of ammonia from  
642 urea applied to soil: influence of hippuric acid and other constituents of livestock  
643 urine. *Soil Biol. Biochem.* 21, 803–808.

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669 **Table 1.** Chemical composition of experimental diets (g kg<sup>-1</sup>, as fed basis) (Beccaccia et  
670 al., 2015). Control pig slurry (PSC), 75 g kg<sup>-1</sup> orange pulp (OP-75), 150 g kg<sup>-1</sup> orange  
671 pulp (OP-150), 75 g kg<sup>-1</sup> carob meal (CM-75) and 150 g kg<sup>-1</sup> carob meal (CM-150).

	<b>PSC</b>	<b>OP-75</b>	<b>OP-150</b>	<b>CM-75</b>	<b>CM-150</b>
Dry matter	912	902	903	895	899
Crude protein	158	156	154	153	157
NDICP <sup>a</sup>	21.4	26.7	21.2	23.8	20.8
Total dietary fibre	194	214	234	200	212
NDF <sup>b</sup>	154	165	158	164	161
ADF <sup>c</sup>	45.6	52.3	56.3	61.0	75.4
ADL <sup>d</sup>	8.0	9.0	10.7	18.9	33.9
Soluble fibre	61.2	75.9	97.3	60.1	71.4

672 <sup>a</sup>Neutral detergent insoluble crude protein.

673 <sup>b</sup> Neutral detergent fibre

674 <sup>c</sup> Acid detergent fibre without residual ash;

675 <sup>d</sup> Acid detergent lignin

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680 **Table 2.** Chemical characteristics of pig slurries (faeces+urine) from different dietary  
681 treatments<sup>a</sup>. Control pig slurry (PSC), 75 g kg<sup>-1</sup> orange pulp (OP-75), 150 g kg<sup>-1</sup> orange pulp  
682 (OP-150), 75 g kg<sup>-1</sup> carob meal (CM-75) and 150 g kg<sup>-1</sup> carob meal (CM-150).

	<b>PSC</b>	<b>OP-75</b>	<b>OP-150</b>	<b>CM-75</b>	<b>CM-150</b>
	Slurry <sup>a</sup>				
Total N (g kg <sup>-1</sup> )	9.64	8.68	7.82	7.95	9.7
NH <sub>4</sub> <sup>+</sup> (g kg <sup>-1</sup> )	5.32	4.48	3.46	3.75	4.35
pH	8.89	7.93	8.08	8.38	8.20

683 <sup>a</sup> Slurry samples were obtained mixing individual excretas (faeces+urine) of 6 pigs by dietary  
684 treatment.

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686

687 **Table 3.** Effect of source (S) and level of inclusion (L) of fibrous by products on the  
688 concentration of minor components in pig urines. Control pig slurry (PSC), 75 g kg<sup>-1</sup> orange  
689 pulp (OP-75), 150 g kg<sup>-1</sup> orange pulp (OP-150), 75 g kg<sup>-1</sup> carob meal (CM-75) and 150 g kg<sup>-1</sup>  
690 carob meal (CM-150).

	PSC	OP-75	OP-150	CM-75	CM-150	SEM <sup>1</sup>	S	L	SxL
Allantoin (mg L <sup>-1</sup> urine)	679	575	653	552	505	141	0.551	0.912	0.661
Creatinine (mg L <sup>-1</sup> urine)	1888	1821	1744	1636	1690	342	0.730	0.973	0.851
Uric acid (mg L <sup>-1</sup> urine)	69.2	52.1	64.8	54.4	40.5	439	0.422	0.967	0.331
Hippuric acid <sup>2</sup> (mg L <sup>-1</sup> urine)	848	1231	1651	552	699	189	<0.001	0.146	0.479
Benzoic acid (mg L <sup>-1</sup> urine)	61.4	147	149	81.1	73.8	37.3	0.072	0.939	0.906

691 <sup>1</sup>Standard error of means (n=6)

692 <sup>2</sup>Contrast PSC vs OP-150 (*P*<0.05)

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711 **Table 4.** Harvest yield and total N in the ryegrass. Control  
 712 pig slurry (PSC), 75 g kg<sup>-1</sup> orange pulp (OP-75), 150 g kg<sup>-1</sup>  
 713 orange pulp (OP-150), 75 g kg<sup>-1</sup> carob meal (CM-75) and 150  
 714 g kg<sup>-1</sup> carob meal (CM-150).

	Harvest Yield	N	
	(g DM m <sup>-2</sup> )	(%)	
Control	14.33 ± 0.9	1.71 ± 0.0	715
PSC	19.37 ± 1.0	3.16 ± 0.2	716
CM-75	19.77 ± 0.5	2.95 ± 0.1	717
CM-150	19.57 ± 0.4	2.86 ± 0.2	718
OP-75	20.45 ± 0.5	2.92 ± 0.1	719
OP-150	20.54 ± 0.9	2.78 ± 0.1	720

721 Values are the mean of three replicates ± standard deviation.

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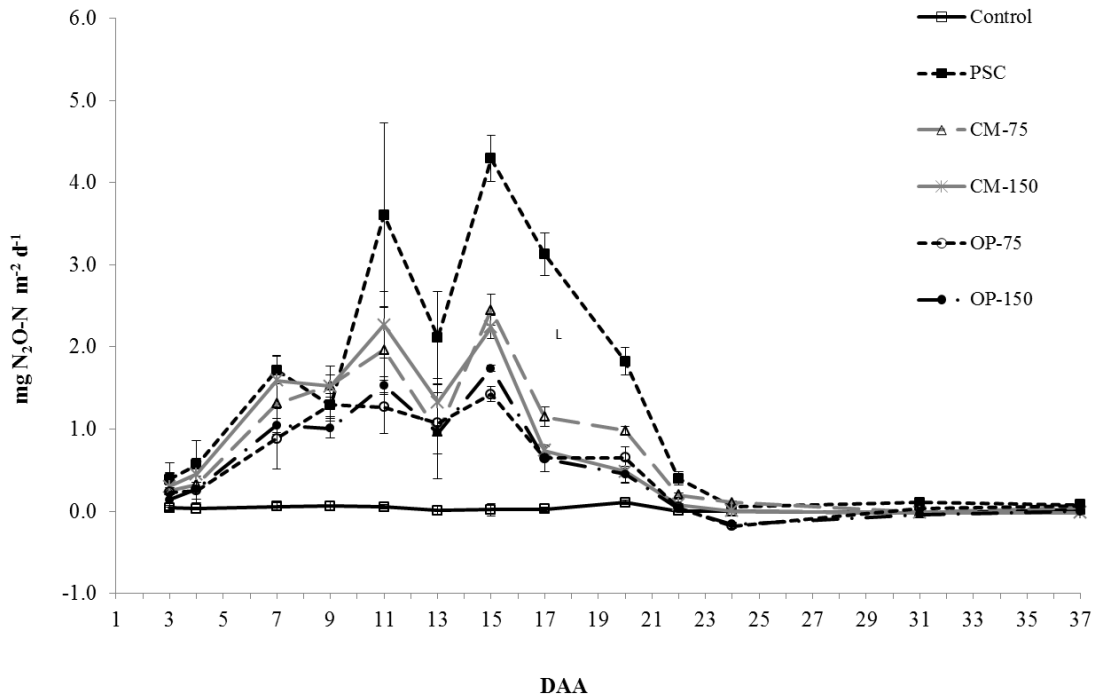
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738 **Figures**

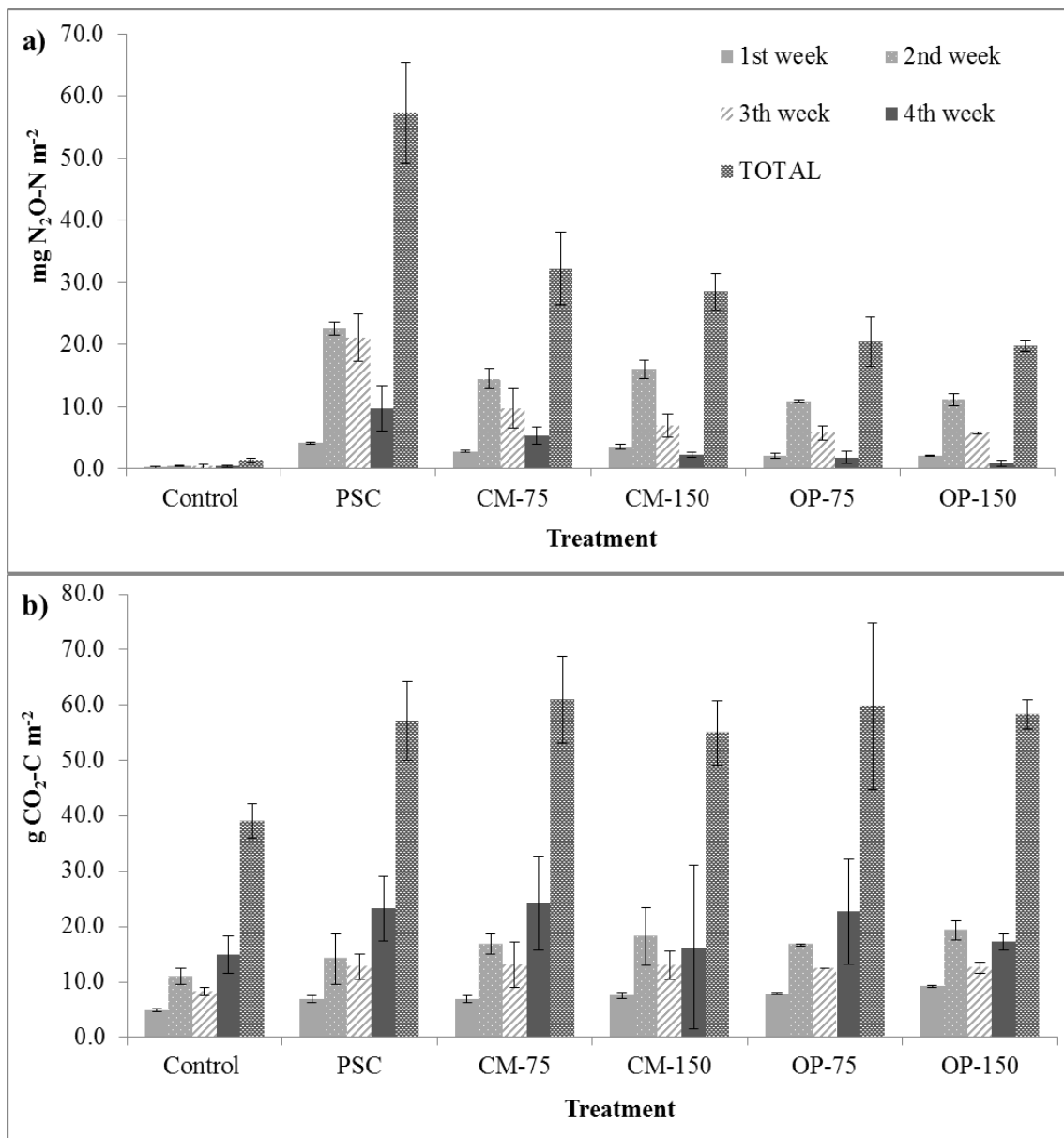
739 **Fig. 1** Daily soil N<sub>2</sub>O emissions (mg N<sub>2</sub>O-N m<sup>-2</sup> d<sup>-1</sup>) from different slurry treatments  
740 following application (DAA). Control pig slurry (PSC), 75 g kg<sup>-1</sup> orange pulp (OP-75),  
741 150 g kg<sup>-1</sup> orange pulp (OP-150), 75 g kg<sup>-1</sup> carob meal (CM-75) and 150 g kg<sup>-1</sup> carob  
742 meal (CM-150). Vertical bars indicate standard errors for each sampling date.



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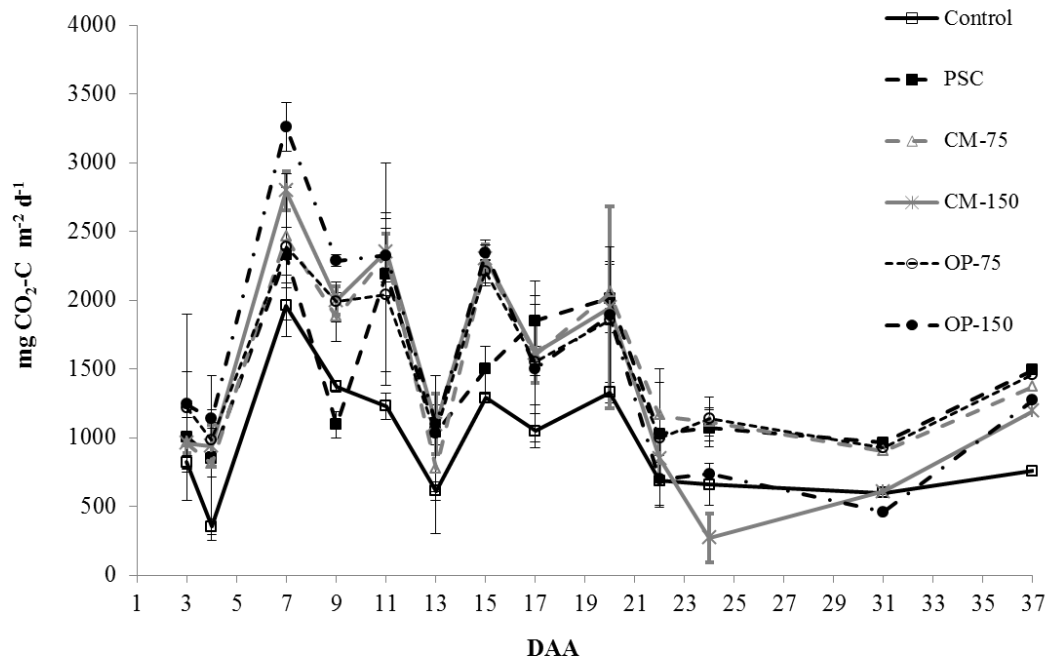
745 **Fig. 2** Cumulative (a) N<sub>2</sub>O and (b) CO<sub>2</sub> emissions per week from soil amended with  
 746 different pig slurries. Control pig slurry (PSC), 75 g kg<sup>-1</sup> orange pulp (OP-75), 150 g kg<sup>-1</sup>  
 747 orange pulp (OP-150), 75 g kg<sup>-1</sup> carob meal (CM-75) and 150 g kg<sup>-1</sup> carob meal (CM-  
 748 150). Vertical bars indicate standard errors for each sampling date. Values are the mean  
 749 of three replicates ± standard deviation. Different letters indicate significant differences  
 750 at (P < 0.05) between treatments in the same week.



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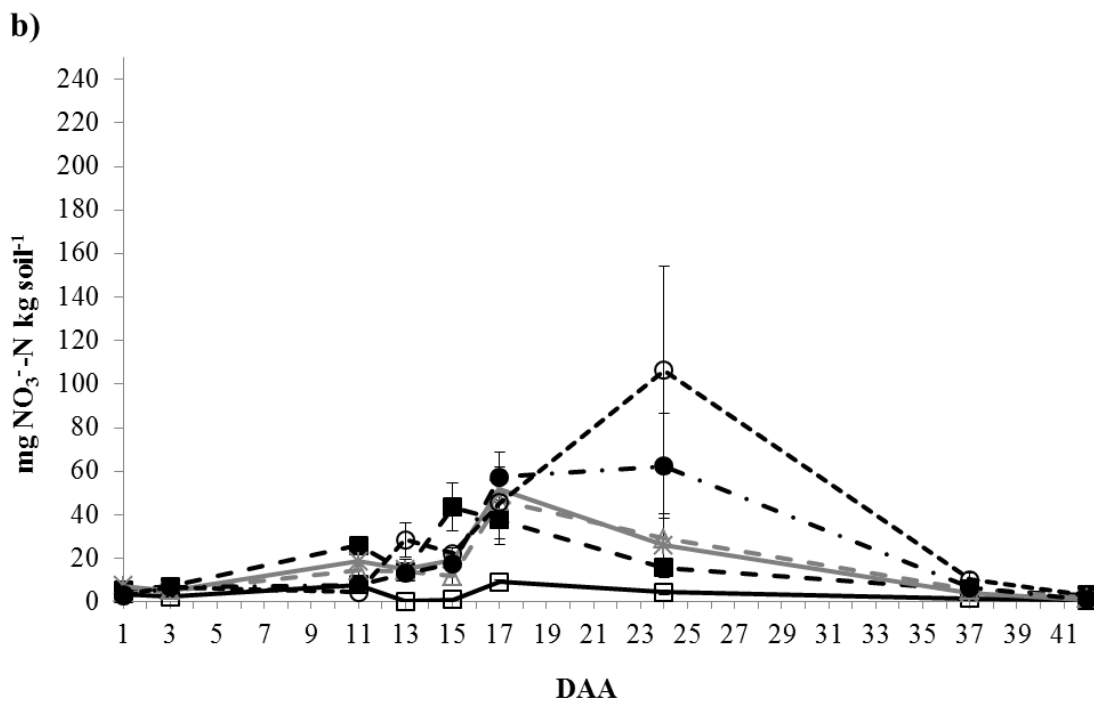
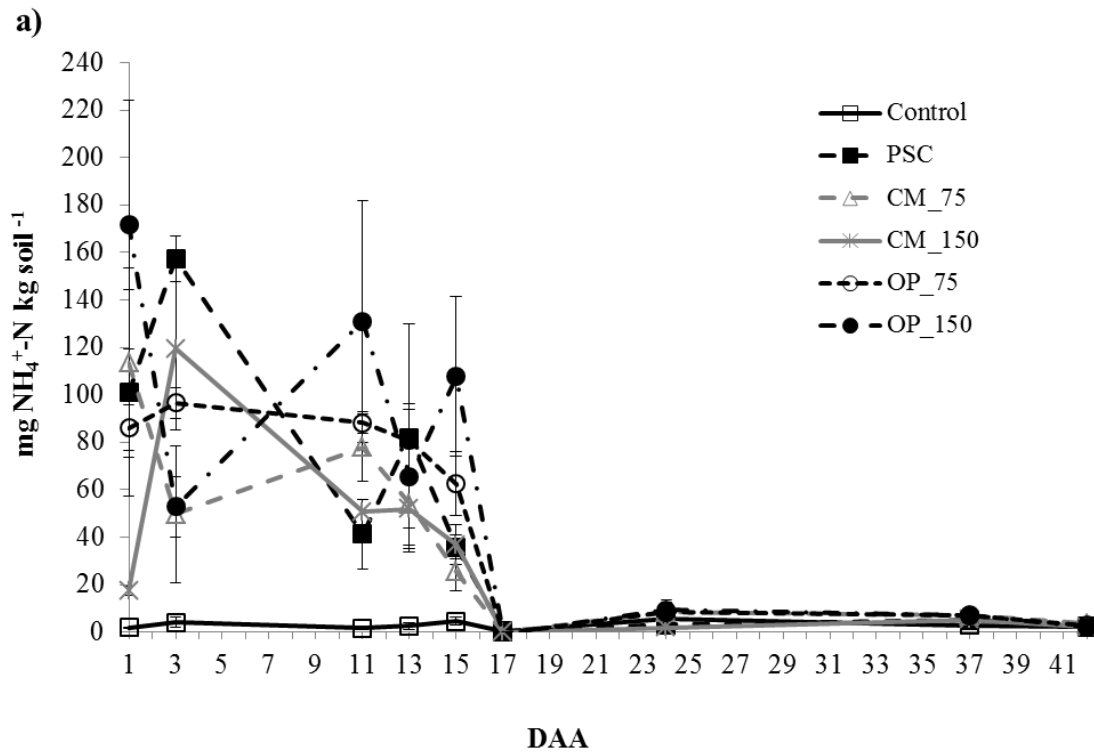
753 **Fig. 3** Daily soil CO<sub>2</sub> emissions (mg CO<sub>2</sub>-C m<sup>-2</sup> d<sup>-1</sup>) from different slurry treatments  
 754 following application (DAA). Control pig slurry (PSC), 75 g kg<sup>-1</sup> orange pulp (OP-75),  
 755 150 g kg<sup>-1</sup> orange pulp (OP-150), 75 g kg<sup>-1</sup> carob meal (CM-75) and 150 g kg<sup>-1</sup> carob  
 756 meal (CM-150). Vertical bars indicate standard errors for each sampling date.



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759 **Fig. 4** (a) Soil  $\text{NH}_4^+\text{-N}$  ( $\text{mg NH}_4^+\text{-N kg}^{-1}$ ) and (b)  $\text{NO}_3^-\text{-N}$  ( $\text{mg NO}_3^-\text{-N kg}^{-1}$ ) for different  
 760 slurry treatments following application (DAA). Vertical bars indicate standard errors for  
 761 each sampling date.



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763 **Fig. 5** (a) Soil C ( $\text{mg C kg}^{-1}$ ) for different slurry treatments following application  
 764 (DAA) during the experimental period and (b) mean soil C ( $\text{mg C kg}^{-1}$ ) during the  
 765 experimental period for each treatment. Vertical bars indicate standard errors.

