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1 Assessment of lignocellulosic nut wastes as an absorbent for gaseous formaldehyde

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10 11 Abstract

12 Indoor air quality is of growing concern with a current focus on formaldehyde emissions and
13 sick building syndrome (SBS). One of the main approaches to reduce indoor pollutant
14 concentrations has been to reduce formaldehyde use and emissions from products. Another
15 approach is the potential of materials to act as scavengers to actively sequester formaldehyde
16 from the indoor atmosphere. This paper evaluates the use of the shells of various types of
17 nuts, which are an abundant agricultural waste material. Nut shells were exposed to gaseous
18 formaldehyde using a Dynamic Vapour Sorption system and their nitrogen content
19 determined using the Kjeldahl method. It was found that formaldehyde absorption increased
20 with increasing nitrogen content and that walnut shell, peanut shell and sunflower seed shell
21 could absorb significantly higher quantities of formaldehyde gas than a sheep wool control.

22
23 **Key words:** Nut waste, Dynamic Vapour Sorption, Formaldehyde, Kjeldahl, Absorption

24 25 1. Introduction

26 Indoor air quality and the effects of airborne contamination on human health, has been of
27 growing concern in recent years (Mitchell et al., 2007; Salthammer et al., 2003; Takeda et al.,
28 2009). It was reported that a significant proportion of the population suffer from eye and
29 respiratory discomfort, headaches and feeling of lethargy linked to poor indoor air quality
30 (Haghighat and De Bellis, 1998). This situation is now referred to as sick building syndrome
31 (SBS) (Zhang and Xu, 2003). Formaldehyde (CH₂O) has been the focus of many
32 investigations as it contributes to poor indoor air quality. Formaldehyde occurs naturally in
33 the environment and is present and reversibly bound in all biological material (Trézl et al.,
34 1997) and is used in many industrial products emit formaldehyde from textiles to
35 disinfectants. A major source of formaldehyde is in pressed wood products, used in
36 construction and furnishings (Hun et al., 2010; Kim et al., 2010). Current guidelines stipulate
37 a limit of 0.1 mg/m³ in interior air to avoid adverse health effects (WHO, 2010). Historically
38 there has been considerable research into the reductions of formaldehyde emissions from their
39 original source, namely replacing formaldehyde based resins with bio-based resins (Jiang et
40 al., 2002; Pratelli et al., 2013). Another method is to actively modify a product to sequester
41 VOCs, for example using cost effective lignocellulosic scavengers (Kim, 2009).

42 Edible nuts are grown and cultivated in a variety of climates around the world on different
43 scales. This enormous production of nuts every year generates a considerable amount of
44 lignocellulosic waste. Table 1 summarises the cultivation, annual seed and waste production

45 and uses of 6 globally popular edible nuts. All of the mentioned wastes have demonstrated the
 46 potential to be used as an activated carbon for absorbing pollutants: walnut can be used as
 47 absorbent of copper ions (Kim et al., 2001), pistachio nut can remove organic compounds
 48 from air and water (mo Nor et al., 2013; Tavakoli Foroushani et al., 2016), coconut can
 49 remove methylene blue in aqueous solutions (Tan et al., 2008), sunflower seed shell (el-
 50 Halwany, 2013) and peanut shell can act as absorbents of CO₂ (Deng et al., 2015). This paper
 51 aims to evaluate and describe the potential of using these 6 promising agricultural wastes, in
 52 their natural, solid state for the adsorption of formaldehyde from the atmosphere to improve
 53 indoor air quality.

54

55 Table 1: 6 major edible nuts, their source and annual production

Nut	Sourced	Annual production	Waste
Almonds (<i>Prunus dulcis</i>)	Grown worldwide. North America, California greatest producer ⁴ (>637,000 tonnes/year) ²	2.09 million tonnes ¹	0.7-1.5 million tonnes waste per year and has little industrial value ¹
Walnut (<i>Juglans regia</i>)	17 major producers ³ . China largest producer (410,000 tonnes /year) ⁵ , North America the 2 nd (300,000 tonnes/year) ¹⁶ and Iran is the 3 rd (150,000 tonnes/year) ³	1.48 million tonnes ³	Multitudinous uses from dye in cosmetics, used in insecticides, fillers, asphalt, glues ⁴ and improving tyre grip ³
Pistachio (<i>Pistacia vera</i>)	Grown mainly in Iran, Turkey and North America. Iran alone producing (>250,000 tonnes/year) ^{7,8}	489,000 tonnes ⁶	Little industrial value, sent to landfill or burnt ¹⁹ and small use in mordant ⁴ and colouring and glues ²⁰
Coconut (<i>Cocos nucifera</i>)	Indonesia is the leading producer, followed by Philippines, India and Sri Lanka ¹⁶ . Malaysia alone requires 151,00ha of land for production ⁹	5.5 million tones ¹⁶	Husk used for rope and mats and core can be used as peat substitute ¹⁸ . 13.6 – 18.14 million tonnes husk waste per annum ¹⁷
Peanut (<i>Arachis hypogaea</i>)	Grown worldwide. China 1 st in production accounting for 40% of global production ¹⁰ (14.5 tonnes/year), followed by India (23%) ¹² .	32.22 million tonnes (including shell) ¹¹	Largely sold in shell or sent to landfill
Sunflower seeds (<i>Helianthus annuus</i>)	Grown worldwide. North American alone produces 1.72 million tonnes/year ¹⁵	27 million tonnes ¹³ (Almost exclusively cultivated for oil ¹⁴)	Small value, sent to landfill or used as low grade roughage for livestock ¹⁵ ,

56 Data derived from: (Pirayesh and Khazaeian, 2012)¹, (Jayasena, 2016)², (Malhotra, 2008)³, (Wickens
 57 G E, 1995)⁴, (Sze-Tao and Sathe, 2000)⁵, (Kahyaoglu, 2008)⁶, (Kashaninejad et al., 2006)⁷, (Razavi et
 58 al., 2007)⁸, (Tan et al., 2008)⁹, (Diop et al., 2004)¹⁰. (Zhang et al., 2012)¹¹, (Zhang et al., 2013)¹², (Li et
 59 al., 2011)¹³, (Hameed, 2008)¹⁴, (Kamireddy et al., 2014)¹⁵, (Anirudhan and Sreekumari, 2011)¹⁶, (van
 60 Dam et al., 2004)¹⁷, (Konduru et al., 1999)¹⁸, (Tavakoli Foroushani et al., 2016)¹⁹, (Fadavi et al.,
 61 2013)²⁰

62

63 It is known that formaldehyde is highly reactive to proteins (Mansour et al., 2016) and reacts
64 with the side chains of amino acids and amido groups of glucose (Curling et al., 2012). The
65 nitrogen (protein) content was therefore determined to assess correlations with formaldehyde
66 sorption. It is known that wool fibre will absorb formaldehyde (Curling et al., 2012) by
67 physisorption, (absorbed into micropores within its structure) and chemisorption (forms a
68 stable bond with the fibres). Wool fibre has therefore been used in this study as a comparative
69 control.

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71 **2. Materials and Methods**

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73 *2.1 Nut shell Waste and Wool*

74 The shell material was dry and oil free and crushed into small pieces (<3mm) and removing
75 any contaminating (non shell) material. Scoured wool fibre was also analysed as a control
76 material for formaldehyde absorption. Urea is a very common chemical added to materials
77 used to absorb free formaldehyde emitted from formaldehyde based products such as
78 particleboard. However the purpose of this study is to evaluate the potential of lignocellulosic
79 wastes used as a protein additive, to absorb ambient formaldehyde emitted from external
80 sources other than reducing a products' formaldehyde emissions. As such, urea is beyond the
81 scope of this study.

82

83 *2.2 Dynamic Vapour Sorption (DVS)*

84 Prior to the experiment, the nut shells and wool were conditioned at 23 ± 1 °C and $60 \pm 3\%$
85 RH until constant mass was obtained. Sorption analyses were performed using DVS system
86 (Surface Measurement Systems, London, UK) in accordance with the methodology described
87 by Curling *et al.* (2012). Three replicates were conducted for each sample.

88

89 *2.3 Nitrogen content*

90 To determine the nitrogen content of the waste nut shells, the Kjeldahl method was used.
91 Three replicates were completed for each nut shell and wool.

92 The shell materials were prepared by dry milling the shells into <5mm pieces and removing
93 any contaminating material. The material was then oven dried overnight in a 50°C oven.
94 Between 0.2g and 0.3g of the oven dried waste shell, weighed to four decimal places, and
95 were placed into digestion tubes to which two Kjeldahl peroxide tablets and 12ml of sulphuric
96 acid were added. The digestion tubes were then placed in a preheated (420°C) digester and
97 left to digest for 1 hour from time of first vapour sighting. Once digestion was complete the
98 cooled samples were transferred to the distilling unit. The distilled sample was removed for
99 titration. Hydrochloric acid (HCl) was titrated into the sample until it became neutral (clear)
100 with the volume of HCl recorded. The nitrogen content was calculated using equation 1:

101

$$102 \quad \% N = 14.01 \times ((t_s - t_b)/m) \times M_{sd} \quad \text{[Equation 1]}$$

103

104 Where: t_s ml of titration of sample, t_b ml of titration blank, m oven dry weight of sample and
105 M_{sd} molarity of standard HCl (0.01).

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3. Results and Discussion

Table 2 and fig 1 show the maximum formaldehyde absorption by the different shell wastes and wool fibre.

Table 2: Formaldehyde absorption by shell waste and wool fibre and their nitrogen content

Scavenger	Formaldehyde absorption (g kg ⁻¹)	SD	Nitrogen content (%)	SD
Wool	49.80	0.35	17.16	0.02
Walnut shell	90.19	0.91	1.12	0.22
Almond shell	64.86	0.67	0.26	0.11
Coconut husk fibre	49.29	0.52	0.31	0.00
Pistachio shell	31.70	0.49	0.10	0.01
Peanut shell	81.48	0.43	0.73	0.03
Sunflower seed shell	101.97	0.22	4.17	0.18

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Figure 1 shows the mass change of each waste shell and wool fibre, over 6 cycles (6 cycles was chosen based on previous experience). The graph reveals there is a rapid mass change in the first cycle and then generally a gradual increase, except for coconut husk fibre, pistachio shell and wool fibre, which appear to have reached a maximum absorption. The other four shell wastes did not reach equilibrium in the 6 cycles. Theoretical maximum absorption values were determined via regression of the absorption curves for the Almond (65.25 g kg⁻¹), Walnut (92.88 g kg⁻¹), Sunflower (117.313 g kg⁻¹) and Peanut (81.52 g kg⁻¹). The calculated values for almond and peanut are within the standard deviation of the observed values with only the walnut and sunflower giving theoretical values outside the standard deviation of the observed.

The nitrogen content was analysed to determine if there was a relationship between protein content and formaldehyde absorption. Table 2 also shows the Kjeldahl nitrogen content results of the waste shells and wool fibre. The higher nitrogen content of sunflower seed shell, walnut shell and peanut shell (4.17%, 1.12% and 0.73% respectively) correlates with their higher capacity to absorb formaldehyde (101.97 g kg⁻¹, 90.19 g kg⁻¹ and 81.48 g kg⁻¹ respectively). However, it appears the wool fibre values do not fit this relationship. Wool has a significantly higher nitrogen content 17.16%, as it is of a protein structure, but it absorbed significantly less formaldehyde (49.80 g kg⁻¹), than the top three shell waste scavengers. The Kjeldahl method measures total nitrogen and therefore may detect non protein nitrogen compounds within the wool.

The reactions between formaldehyde and other compounds and molecules is very complex, as formaldehyde has low specificity and will readily react with a number of compounds in different ways (Reddie and Nicholls, 1971). The reactions between wool and formaldehyde are very complex. Polyamides form the backbone of the wool proteins and are comprised of many functional groups, each with varying reactivity (Reddie and Nicholls, 1971). The wool keratin reacts with formaldehyde and formaldehyde irreversibly binds to asparagine amide groups of the wool (Alexander et al., 1951; Middlebrook, 1949).

141 It is well reported that formaldehyde will react and bind with amino groups and result in the
142 formation of a methylol derivative (Alexander et al., 1951; Levy and Silberman, 1937;
143 Puchtler and Meloan, 1984; Reddie and Nicholls, 1971). Other crosslinks are formed between
144 amine and amide, amine and phenol and amine and indole groups (Alexander et al., 1951).
145 Lignocellulosic wastes composition contain a wide variety of functional groups (Altun and
146 Pehlivan, 2012; Miretzky and Cirelli, 2010; Okuda et al., 2003; Reddie and Nicholls, 1971;
147 Zitouni et al., 2000). The predominant amino acids found in the lignocellulose material varies
148 with species; walnut contains lysine, in almonds cysteine and methionine and peanut
149 threonine and methionine (Venkatachalam and Sathe, 2006). These differences in the type,
150 composition and quantity of the functional groups may be key factors in determining the
151 ability of a material to absorb and bind formaldehyde. Determination of the different types of
152 functional groups on these waste nut shells may help to explain the differences observed in
153 the quantity of formaldehyde absorbed by the shells and wool. Physical factors may also play
154 an important role as there may be differences due to access via diffusion into the materials
155 and due to different quantities of active nitrogen sites.

156

157 **4. Conclusions**

158 The purpose of this study was to determine if low cost and unutilised waste nut shell could be
159 used in their natural state to absorb formaldehyde. The study reveals that all the 6 shell types
160 can absorb formaldehyde, with pistachio nut shell absorbing the least and sunflower seed shell
161 absorbing the greatest amount. The Kjeldahl results revealed that the amount of formaldehyde
162 absorbed increased as nitrogen content within the shells increased. To conclude, sunflower
163 seed shell, peanut shell, almond and walnut shell biowaste could be better utilised as organic
164 scavengers to absorb formaldehyde from the atmosphere and improve indoor air quality.

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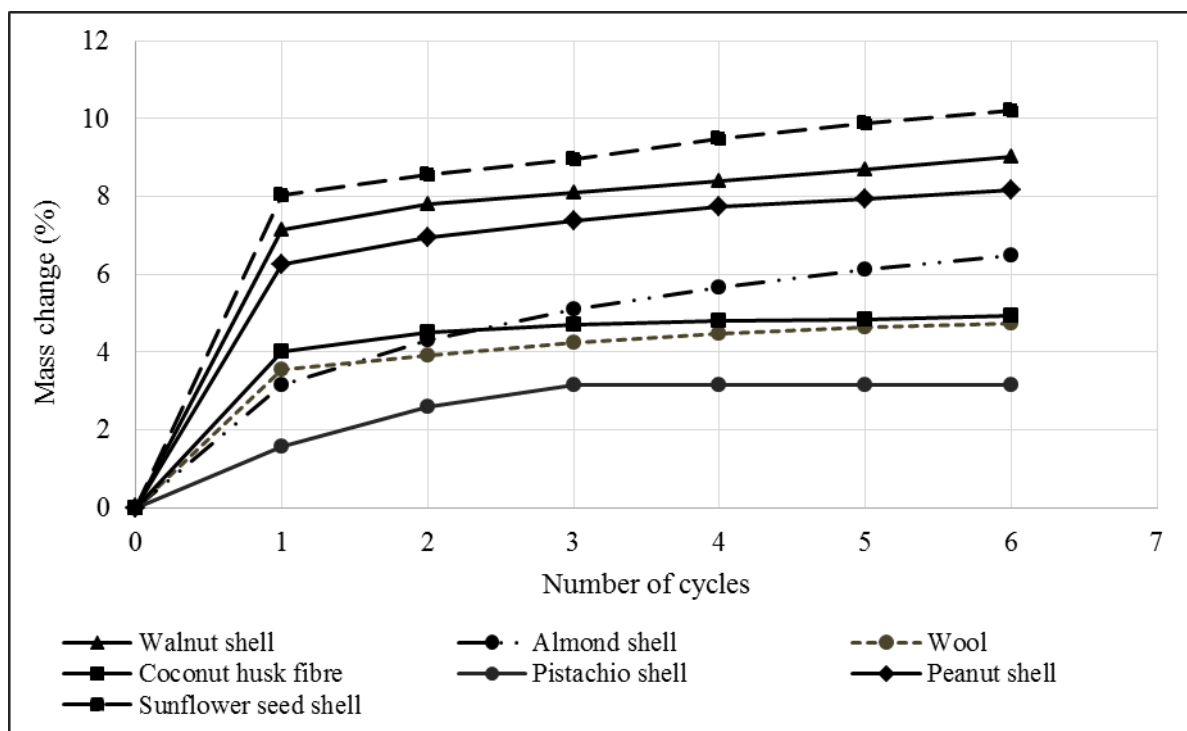
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305 Figure 1: Shell waste and mass change over 6 cycles.

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