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1 **Local human impacts decouple natural biophysical relationships on Pacific coral reefs**

2

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9

10 **Abstract:** Human impacts can homogenize and simplify ecosystems, favoring communities that
11 are no longer naturally coupled with (or reflective of) the background environmental regimes in
12 which they are found. Such a process of biophysical decoupling has been explored little in the
13 marine environment due to a lack of replication across the intact-to-degraded ecosystem
14 spectrum. Coral reefs lacking local human impacts provide critical baseline scenarios in which to
15 explore natural biophysical relationships, and provide a template against which to test for their
16 human-induced decoupling. Using 39 Pacific islands, 24 unpopulated (relatively free from local
17 human impacts) and 15 populated (with local human impacts present), spanning 45° of latitude
18 and 65° of longitude, we ask, what are “natural” biophysical relationships on coral reefs and do
19 we see evidence for their human-induced decoupling? Estimates of the percent cover of benthic
20 groups were related to multiple physical environmental drivers (sea surface temperature,
21 irradiance, chlorophyll-*a*, and wave energy) using mixed-effects models and island mean

22 condition as the unit of replication. Models across unpopulated islands had high explanatory
23 power, identifying key physical environmental drivers of variations in benthic cover in the
24 absence of local human impacts. These same models performed poorly and lost explanatory
25 power when fitted anew to populated (human impacted) islands; biophysical decoupling was
26 clearly evident. Furthermore, key biophysical relationships at populated islands (i.e. those
27 relationships driving benthic variation across space in conjunction with chronic human impact)
28 bore little resemblance to the baseline scenarios identified from unpopulated islands. Our results
29 highlight the ability of local human impacts to decouple biophysical relationships in the marine
30 environment and fundamentally restructure the natural rules of nature.

31

32 *Key words: biophysical decoupling; coral reef benthic regimes; crustose coralline algae; hard*
33 *coral; human impacts; human-environment interactions; macroalgae; mixed-effects models;*
34 *natural variation; physical drivers*

35

36 **Introduction**

37 Long-term environmental regimes, interspersed with anomalous disturbance events, play
38 a crucial role in determining both the rate and path of ecological succession (Odum 1969,
39 Connell and Slatyer 1977, Sousa 1984). When disturbances are rare, competitive exclusion
40 results in mature climax communities and the system can approach relative stability (Pianka
41 1970, Margalef 1975, Stearns 1977, Hughes and Connell 1999). In contrast, when disturbance
42 events are frequent or the system experiences continually harsh environmental conditions, the
43 abundance of stress-tolerant species with weedy life history strategies can increase, leading to

44 seemingly stochastic fluctuations of community structure (Stearns 1977, Sousa 1984, Hughes
45 and Connell 1999, Reznick et al. 2002). Across gradients in environmental regimes, therefore,
46 biological communities often change in discernable (and predictable) ways; adult survivorship
47 and community composition are often strongly coupled with and reflective of their surrounding
48 physical environment (Margalef 1975, Stearns 1977, Sousa 1984).

49 Superimposed over these naturally coupled biophysical relationships are the effects of
50 human activities. Humans can act to homogenize and simplify ecosystems (Odum 1969, Western
51 2001, Riegl et al. 2012), artificially favoring stress-tolerant species (Darling et al. 2013) and
52 forcing the system into an earlier successional state (Schulte et al. 2007, Sandin and Sala 2012).
53 Under chronic anthropogenic stress, the community becomes a product of human-induced
54 change (Graham et al. 1963, Odum 1969, Möllmann et al. 2009) and is no longer naturally
55 coupled with (or reflective of) the background environmental regime in which it is found (Curran
56 et al. 2004, Folke et al. 2011); we term this process *biophysical decoupling*. For example, in the
57 Borneo rainforest, human activities have shifted climax forest communities to expanses of early
58 successional plant communities fragmented by low diversity monoculture (Curran et al. 2004).
59 Under such human-induced change, wildfires associated with El Niño events, instead of being a
60 key natural driver of regenerative processes and ecological succession in the forests, are actually
61 detrimental to ecological succession. The natural biophysical relationships that once existed have
62 become decoupled and novel feedbacks have established (Curran et al. 2004, Folke et al. 2011).
63 Separating the independent and interacting relationships between naturally coupled biophysical
64 relationships versus those established or modified by local human impacts, however, remains
65 hard to test in many ecological systems. Human-induced change is so ubiquitous that unaltered
66 communities, ones that provide replication at the intact end of an intact-to-degraded spectrum,

67 often do not exist (Steffen et al. 2011). Testing for evidence of biophysical decoupling in
68 ecological systems, therefore, poses some practical challenges.

69 Coral reefs provide an interesting case study in which to examine the effects of human-
70 induced change to biophysical relationships in ecological communities as they provide example
71 of some of the most human-degraded (Pandolfi et al. 2003, Mora 2008, Hughes et al. 2010) but
72 also the most pristine (Sandin et al. 2008, Vroom et al. 2010, Williams et al. 2013) ecosystems
73 on our planet. In particular, remote islands harboring coral reefs relatively free from local human
74 impacts provide replication at the intact end of the spectrum, enabling biophysical relationships,
75 and their potential for human-induced decoupling, to be investigated. The foundational benthic
76 organisms that contribute to coral reef development and persistence are those that deposit
77 calcium carbonate, particularly hard (scleractinian) corals and crustose coralline algae (CCA).
78 The latter are critical for reef growth, acting to consolidate reef substrate (Littler and Littler
79 1984) and facilitate coral recruitment (Price 2010). When conditions are such to allow
80 community succession, CCA provide a mechanism for ecological resilience following
81 disturbance (Nyström et al. 2008), and the system may transition to a state where hard corals
82 hold competitive superiority against algal turfs and larger macroalgae (Grigg 1983, Littler and
83 Littler 1985, Barott et al. 2012). Macroalgae represent another community attractor on coral reefs
84 that directly compete with corals and CCA for space. Both calcifying and fleshy forms exist,
85 with the former more typical of functionally intact systems (Vroom and Braun 2010, Williams et
86 al. 2013) and the latter characteristic of more degraded, human-impacted coral reef environments
87 (Littler and Littler 1985, Hughes et al. 2010).

88 Here we use 39 Pacific coral reef islands and atolls (hereafter referred to as islands) in a
89 macroecological setting to test the hypothesis that local human impacts are capable of

90 decoupling natural biophysical relationships in the marine environment. These islands span 45°
91 of latitude and 65° of longitude, crossing multiple gradients in physical environmental drivers
92 (Gove et al. 2013) and human population density (Williams et al. 2011). We quantify changes in
93 the percent cover of three major benthic groups (hard coral, crustose coralline algae, and
94 macroalgae) to ask two questions: 1) Given the absence of local human populations (and thus
95 direct local human impacts), under which set of physical environmental drivers do individual
96 benthic groups predominate (i.e. what are “natural” biophysical relationships on coral reefs?),
97 and 2) Do we see evidence for human-induced decoupling of these natural relationships?

98 **Methods**

99 *Study regions*

100 Benthic surveys were conducted as part of the Coral Reef Ecosystem Division (CRED) of
101 the NOAA Pacific Island Fisheries Science Center’s (PIFSC) Pacific Reef Assessment and
102 Monitoring Program (RAMP). The 39 islands are located within four major geopolitical regions:
103 the Hawaiian Archipelago, the Mariana Archipelago, the Pacific Remote Island Areas (PRIAs),
104 and the islands of American Samoa (Fig. 1). With the exception of the unpopulated PRIAs, some
105 islands within each region harbor dense human populations and are directly impacted by human
106 activities (e.g. fishing and coastal development), such as Oahu in the Hawaiian Islands, Guam in
107 the Mariana Archipelago, and Tutuila in American Samoa. Conversely, islands that are
108 unpopulated and far from human population centers are relatively free of local human impacts
109 (Williams et al. 2011). In this study, we describe islands as either ‘unpopulated’ (relatively free
110 of local human impacts) or ‘populated’ (local human impacts are present). In this manner, 24
111 islands were classified as unpopulated and 15 as populated.

112 ***Benthic community surveys***

113 Inter-island variation in the percent cover of benthic groups was quantified using a
114 towed-diver survey method, a spatially expansive method that is effective at characterizing
115 benthic communities at a coarse taxonomic resolution (hard coral, crustose coralline algae, and
116 macroalgae) (Kenyon et al. 2006). Although the towed divers also recorded percent cover of
117 sand, rubble, and ‘other’, these categories were not investigated in detail here (raw data
118 presented in Supplementary material Appendix 1, Fig. A1). For each island, all tows over the
119 forereef habitat (depth range of 8 – 20 m) crossing consolidated hard (habitable) substrate were
120 averaged to give an island mean percent cover of each benthic group. We used surveys carried
121 out during the shortest interval of consecutive years that had the largest number of islands
122 surveyed (2008 and 2009) to maximize sample size while minimizing temporal variation (*sensu*
123 Mora 2008). A more detailed description of the towed-diver survey technique is given in
124 Supplementary material Appendix 2 and details pertaining to the location, timing, and relative
125 effort of surveys are given in Supplementary material Appendix 3, Table A1.

126 ***Predictor variables***

127 We quantified four major physical environmental drivers known to structure coral reef
128 benthic communities: sea surface temperature (SST), irradiance, wave energy, and chlorophyll-*a*
129 (as a proxy for phytoplankton biomass) (Odum and Odum 1955, Glynn 1976, Done 1983, Brown
130 1997, Hoegh-Guldberg 1999). SST data were generated using Pathfinder v5.0, irradiance and
131 chlorophyll-*a* using the Moderate Resolution Imaging Spectroradiometer (MODIS) and wave
132 energy using the global, full spectral Wave Watch III wave model (see Appendix 2 for full
133 details on how these data were generated). Each of the satellite-derived (SST, irradiance,

134 chlorophyll-*a*) and modeled (wave energy) time series data sets were quality controlled and
135 proportionally scaled to the size of each island in order to control for variable island sizes across
136 our data set (*sensu* Gove et al. 2013). The following metrics were derived from the above time
137 series data sets: long-term climatological mean and standard deviation and positive anomalous
138 events, represented as both the frequency of positive anomalies (the annual average percentage
139 of time above the maximum climatological value) and the magnitude of positive anomalies (the
140 annual average magnitude of events above the maximum climatological value) (Gove et al.
141 2013). Only information up to the survey date for each individual island was included. Human
142 population density estimates were obtained from Williams et al. (2011). A summary of these
143 data, with their units and range across our study region, are provided in Supplementary material
144 Appendix 3, Table A2.

145 ***Statistical modeling***

146 We constructed a series of models to test our questions: 1) a best-fit model across
147 unpopulated islands – to quantify “natural” biophysical relationships in the absence of local
148 human impacts, 2) the unpopulated island best-fit model structure fitted anew to populated
149 islands – to test whether biophysical relationships remained unaltered in the presence of local
150 human impacts, 3) the unpopulated island best-fit model structure fitted anew to populated
151 islands with variations in human population density forcibly included as a predictor – to assess
152 for improved model performance, 4) a best-fit model across populated islands – to test whether
153 novel biophysical relationships can emerge on coral reefs under the influence of local human
154 impact, and 5) a best-fit model across all 39 islands – to quantify biophysical relationships when
155 no *a priori* groups across the islands were defined.

156 To build each of the models we used generalized additive mixed-effects models
157 (GAMM) (Wood 2012). We incorporated a random factor (island groups) to account for possible
158 spatial autocorrelation between islands. Groupings were identified using hierarchical clustering
159 based on pairwise Euclidean distances between each of our 39 islands and an inflection point in
160 the intra-island group variance identified (n = 12 groups total, see Supplementary material
161 Appendix 2 for more details on these methods). GAMMs were fitted using the *gamm4* and *lme4*
162 packages for R. Predictor variables were investigated for co-linearity (with a threshold
163 correlation for inclusion set at 0.75) and normalized to account for the large variation in their
164 numerical values (see Appendix 3, Supplementary Tables A2 and A3 for the list of final
165 predictor variables included, their ranges, and their co-lineation values). Non-linear smoothness
166 was determined using penalized cubic regression splines, with the number of knots spread evenly
167 throughout each covariate value and limited to four to reduce overfitting. We coded a
168 convenience wrapper R function (*all.subsets.gamm*) to fit each GAMM to all possible
169 combinations of the predictors to avoid stepwise selection procedures (see Appendix 4 for the
170 *all.subsets.gamm* R code). Candidate models were subsequently ranked based on AICc relative-
171 importance weights (w_i), with the most plausible model having the highest weight
172 (Wagenmakers and Farrell 2004). We report all models with >15 % of the model-based support
173 from the w_i results (MacNeil et al. 2009), or the top three where this resulted in a single model to
174 highlight the rapid drop in w_i . To assess the stability of the best-fit models, we sequentially
175 deleted single data points from the original response variable data set (delete-one jackknife) and
176 re-calculated the GAMM. We then calculated the percentage of times this resulted in the same
177 model structure. Finally, we calculated a measure of predictor variable relative importance
178 within each candidate model by calculating the sum of AICc model weights for each predictor

179 (i.e. the sum of model weights across all models containing each predictor). A more detailed
180 outline of our statistical modeling methods is given in Supplementary material Appendix 2.

181 **Results**

182 *Spatial variation in benthic cover*

183 Across our study system, mean percent cover of hard coral equaled 18.8 % (upper CI
184 =15.7, lower CI = 22.1), crustose coralline algae 12.5 % (9.6, 15.8), and macroalgae 15.0 %
185 (12.1, 18.1), but there was considerable variation across a variety of hierarchical scales (Fig. 2).
186 In summary, hard coral cover was higher at unpopulated islands (22.5 %) than at populated ones
187 (12.9 %) (see Fig. 2 for associated CIs). Across regions, hard coral cover was highest in the
188 Pacific Remote Island Areas (PRIAs) (30.7 %) and lowest in the Hawaiian Archipelago (13.4
189 %). Within any single region harboring both unpopulated and populated islands, only within the
190 Mariana Archipelago was hard coral cover higher at unpopulated islands (20.1 %) than at
191 populated ones (10.6 %) (Fig. 2).

192 The overall mean percent cover of crustose coralline algae (CCA) did not differ with
193 island status (unpopulated *versus* populated islands) (Fig. 2). Across regions, CCA cover was
194 highest in American Samoa (31.2 %) and the PRIAs (16.2 %) and lowest in the Hawaiian (9.5
195 %) and Mariana Archipelago (6.5 %) (see Fig. 2 for associated CIs). Within any single region
196 harboring both unpopulated and populated islands, CCA cover did not differ with island status.
197 Similarly, the overall mean percent cover of macroalgae did not differ with island status or
198 across regions (Fig. 2); however, macroalgae cover was higher at unpopulated (23.4 %) than
199 populated (10.8 %) islands within the Hawaiian Archipelago and higher at populated (25.5 %)
200 than unpopulated (8.6 %) islands within the Mariana Archipelago. In summary, the complex

201 variation in benthic cover across regions and island status (unpopulated *versus* populated islands)
202 provided motivation for an island-level model-fitting approach.

203 ***Biophysical relationships across Pacific coral reefs***

204 *Hard coral:* Across unpopulated islands, mean sea surface temperature (SST), the
205 magnitude of wave anomalies, and mean chlorophyll-*a* formed the best-fit model, explaining
206 72.3 % of the variation in hard coral cover (Table 1). Hard coral cover was higher at unpopulated
207 islands experiencing a higher mean SST ($> 28^{\circ}\text{C}$), wave anomaly events of a lower magnitude ($<$
208 30 kW m^{-1} above the maximum climatological value), and higher levels of mean chlorophyll-*a* ($>$
209 0.15 mg m^{-3}) (Fig. 3a). Within our confidence set of models, hard coral cover was also higher at
210 unpopulated islands where mean wave energy was lower ($< 25 \text{ kW m}^{-1}$) and where wave
211 anomalies were rare ($< 2 \%$ of the time); however, these predictors had low relative importance
212 scores (Table 1, Supplementary material Appendix 1, Fig. A2).

213 The unpopulated island best-fit model structure performed poorly when fitted anew to
214 populated islands, explaining only 14.7 % of the variation in hard coral cover (Table 1); $p > 0.05$
215 for all the smooth terms (Fig. 3b). The addition of human population density as a predictor across
216 populated islands did not improve model performance (Table 1) and no significant relationship
217 was seen with hard coral cover (Fig. 3c). The populated islands best-fit model identified mean
218 irradiance as the strongest predictor, explaining 25.7 % of the variation in hard coral cover
219 (Table 1). Hard coral cover was generally higher at populated islands experiencing a lower mean
220 irradiance ($< 42 \text{ E m}^{-2} \text{ d}^{-1}$); above this value the relationship appeared to level off but was
221 associated with increased error (Fig. 3d). This single model had strong relative plausibility, as
222 shown by the high Akaike weight (Table 1). The best-fit model for hard coral cover across all 39

223 of our study islands, with no *a priori* grouping of the data, performed more poorly than each of
224 the unpopulated island and populated island models (Supplementary material Appendix 3, Table
225 A4).

226 *Crustose coralline algae (CCA)*: Across unpopulated islands, the frequency and
227 magnitude of chlorophyll-*a* anomalies formed the best-fit model, explaining 80.2 % of the
228 variation in CCA cover (Table 1). CCA cover was higher at unpopulated islands with frequent
229 chlorophyll-*a* anomalies (> 10 % of the time) of a high magnitude (> 0.02 mg m⁻³ above the
230 maximum climatological value) (Fig. 4a). Within our confidence set of models, CCA cover was
231 also higher at unpopulated islands experiencing higher mean wave energy (> 40 kW m⁻¹), but
232 where wave anomaly events were of a low magnitude (< 20 kW m⁻¹ above the maximum
233 climatological value); however, these predictors had low overall relative importance scores
234 (Supplementary material Appendix 1, Fig. A3).

235 The unpopulated island best-fit model structure performed poorly when fitted anew to
236 populated islands, explaining only 11.7 % of the variation in CCA cover (Table 1). The
237 relationships with the frequency and magnitude of chlorophyll-*a* anomalies decoupled ($p > 0.05$
238 for the smooth term) and reversed in direction, respectively (Fig. 4b). CCA cover was lower at
239 populated islands experiencing chlorophyll-*a* anomalies of a high magnitude (> 0.015 mg m⁻³
240 above the maximum climatological value); however, this relationship was associated with
241 increased error in the smooth term at very high magnitudes (Fig. 4b). The addition of human
242 population density as a predictor improved model performance across populated islands and
243 suggested an increase in CCA cover at lower population densities; however, the relationship was
244 weak (Fig. 4c) and increased the overall variation explained by only 6.9 % (Table 1). The
245 populated islands best-fit model identified the magnitude of irradiance anomalies as the strongest

246 predictor, explaining 87.0 % of the variation in CCA cover (Table 1). CCA cover was higher at
247 populated islands where irradiance anomalies were of a high magnitude ($> 3.4 \text{ E m}^{-2} \text{ d}^{-1}$ above
248 the maximum climatological value) (Fig. 4d), and this single model heavily dominated in terms
249 of relative plausibility as shown by the high Akaike weight (Table 1). The best-fit model for
250 CCA cover across all 39 of our study islands, with no *a priori* grouping of the data, performed
251 more poorly than each of the unpopulated island and populated island models (Supplementary
252 material Appendix 3, Table A4).

253 *Macroalgae:* Across unpopulated islands, the frequency of chlorophyll-*a* anomalies
254 formed the best-fit model, explaining 61.2 % of the variation in macroalgae cover (Table 1).
255 Macroalgae cover was higher at unpopulated islands experiencing lower frequencies of
256 chlorophyll-*a* anomalies (< 1 % of the time) (Fig. 5a). Within our confidence set of models,
257 relationships with mean SST and the frequency of wave anomalies also featured (Table 1), with
258 macroalgae cover higher at unpopulated islands experiencing a lower mean SST (< 23.5 °C) and
259 lower frequencies of wave anomalies (< 1 % of the time) (Supplementary material Appendix 1,
260 Fig. A4).

261 The unpopulated island best-fit model structure performed poorly when fitted anew to
262 populated islands, explaining only 18.1 % of the variation in macroalgae cover (Table 1). The
263 overall negative effect of increasing chlorophyll-*a* anomaly frequencies on macroalgae cover still
264 held at populated islands, but was associated with increased error in the relationship (Fig. 5b).
265 Variations in human population density across populated islands had no significant effect on
266 model performance (Table 1) and showed no significant relationship with macroalgae cover (Fig.
267 5c). The populated islands best-fit model for macroalgae identified mean wave energy as the
268 strongest predictor, explaining 87.7 % of the variation in macroalgae cover (Table 1).

269 Macroalgae cover was lower at populated islands experiencing a higher mean wave energy (> 20
270 kW m^{-1}) (Fig. 5d), and this single model dominated in terms of relative plausibility as shown by
271 the high Akaike weight (Table 1). The best-fit model for macroalgae cover across all 39 of our
272 study islands, with no *a priori* grouping of the data, performed more poorly than each of the
273 unpopulated island and populated island models (Supplementary material Appendix 3, Table
274 A4).

275 **Discussion**

276 Using 39 coral reef islands across the Pacific, we show that local human impacts
277 fundamentally alter natural biophysical relationships in the marine environment. Relationships
278 across unpopulated islands (i.e. natural biophysical relationships in the absence of local human
279 impacts) were easily discernable and had high explanatory power. Here variations in physical
280 drivers were linked in a predictable manner to variations in the percent cover of each of the three
281 benthic groups (hard coral, crustose coralline algae, and macroalgae). In contrast, these same
282 model structures performed poorly when fitted anew to populated (human impacted) islands,
283 with many of the biophysical relationships predominating at unpopulated islands becoming
284 decoupled or altering in the direction of their relationship; we term this process *biophysical*
285 *decoupling* (Fig. 6). To our knowledge, our study is among the first to take a macroecological
286 approach to demonstrate a clear human-induced decoupling of natural biophysical relationships
287 in the marine environment.

288 In the absence of local human impacts, the principal physical drivers of variations in
289 benthic groups were largely consistent with natural history gleaned from decades of coral reef
290 science. For example, hard coral cover was higher at unpopulated islands experiencing a higher

291 mean sea surface temperature; where mean temperatures were lower, corals decreased in cover
292 and competitive macroalgae increased in cover. Such latitudinal limits to coral reef development
293 have been previously noted (Johannes et al. 1983, Lough and Barnes 2000) and are thought to
294 reflect variations in energy influx into the system dictating community diversity and succession
295 (Fraser and Currie 1996). Hard coral cover also declined under conditions of high wave energy,
296 particularly anomalous, high-magnitude events. This is a common biophysical relationship on
297 Pacific coral reefs (Dollar 1982, Grigg 1983, Page-Albins et al. 2012, Williams et al. 2013, Gove
298 et al. *In Press*) and often results from colony dislodgement and abrasive damage (Madin and
299 Connolly 2006) and the inability of coral larvae to settle under high hydrodynamic disturbance
300 (Abelson and Denny 1997). In contrast, crustose coralline algae (CCA) increased in cover at
301 unpopulated islands experiencing higher mean wave energy (i.e. higher levels of disturbance) as
302 predicted by the relative dominance model proposed by Littler and Littler (1985). Similarly to
303 hard corals, however, CCA cover was lower at unpopulated islands experiencing a higher
304 frequency of large-magnitude wave anomalies. Although physically adapted to high wave energy
305 environments (Sheppard 1980), CCA are still vulnerable to abrasion at very high levels of wave
306 energy and can give way to earlier successional states, such as algal turfs (Williams et al. 2013,
307 Gove et al. *In Press*). As well as increasing in cover in warmer waters, hard coral cover also
308 increased under conditions of higher mean chlorophyll-*a* (indicative of more nutrient-rich
309 waters). Under oligotrophic conditions, increased nutrient supply to corals may act to promote
310 persistence and resilience following disturbance (Grottoli et al. 2006, Connolly et al. 2012). CCA
311 also positively responded to increased chlorophyll-*a*, increasing in cover particularly where there
312 were frequent anomalies of a large-magnitude. These paralleled relationships between hard coral
313 and CCA with increases in chlorophyll-*a* likely reflect the positive effects CCA can have on hard

314 coral cover (Price 2010) and the negative effects CCA can have on competitive macroalgae
315 (Vermeij et al. 2011). In the absence of local human impacts, it would appear background
316 increases in chlorophyll-*a* drive macroalgae cover down, likely reflecting a loss of competitive
317 dominance with hard corals and CCA for space.

318 Across populated islands, all the natural biophysical relationships identified from
319 unpopulated islands decoupled, lost explanatory power, or became fundamentally altered in the
320 direction of their relationship (Fig. 6). Hard coral cover no longer showed a significant
321 relationship with either mean temperature or mean chlorophyll-*a* and the relationship between
322 CCA cover and the frequency and magnitude of chlorophyll-*a* anomalies decoupled and
323 reversed, respectively. On occasion, novel biophysical relationships even appeared to develop.
324 For example, hard coral cover across populated islands decreased as mean irradiance increased,
325 perhaps reflecting a stress-reinforcing factor reducing the persistence of corals in an already
326 chronically stressed system (Lesser and Farrell 2004). In contrast, CCA cover increased as the
327 frequency of irradiance anomalies increased, particularly anomalies of a large magnitude. CCA
328 possess a tremendous ability to photoacclimate (Bulleri 2006) and thus often thrive under well-lit
329 conditions (Sheppard 1980, Williams et al. 2013). Incoming irradiance may therefore be a key
330 physical driver allowing CCA to persist in more chronically stressed reef systems. Finally, across
331 populated islands, the relationship between macroalgae cover and mean chlorophyll-*a*
332 substantially weakened; macroalgae cover was instead primarily related to wave energy, with
333 cover higher at populated islands experiencing a lower mean wave energy. Macroalgae, like
334 corals, are vulnerable to dislodgement (Dethier et al. 1991, Engelen et al. 2005) and it therefore
335 follows that increased wave energy can reduce their overall cover on coral reefs (Gove et al. *In*
336 *Press*). Also, it is interesting to note that our findings do not follow a common expectation that

337 the cover of macroalgae should be positively related to factors associated with human presence
338 on coral reefs (Littler and Littler 1985, Hughes 1994, Mora 2008). Only within the Mariana
339 Archipelago did this expectation hold true and the opposite trend was actually observed within
340 the Hawaiian Archipelago. Across the Hawaiian Archipelago, macroalgae cover was 2.5 times
341 higher across the unpopulated and highly protected Northwestern Hawaiian Islands than in the
342 populated Main Hawaiian Islands. Such disparities have been previously noted within the
343 Hawaiian Archipelago (Vermeij et al. 2010, Vroom and Braun 2010, Jouffray et al. In Press) and
344 likely reflect the broad functional roles different types of macroalgae have on coral reefs (Vroom
345 2011, Williams et al. 2013).

346 *Future work to further our understanding of biophysical decoupling*

347 We hypothesize that the decoupling of natural biophysical relationships across populated
348 (human impacted) islands reflects, in part, human-induced ecological homogenization (Riegl et
349 al. 2012). In other words, while human impacts may not always result in wholesale changes to
350 benthic cover (as was largely the case in our study), the dominant species within each group may
351 shift as a result of human-induced selective pressure. For example, the relative dominance of
352 fleshy *versus* calcified macroalgae often differs across gradients of human impact (Vroom et al.
353 2006, Vroom 2011) and coral communities are flexible and can reorganize in response to
354 selective pressure, each species differing in their degree of stress-tolerance (Darling et al. 2013).
355 Local human impacts may be artificially selecting for more stress-tolerant, weedy species within
356 each benthic group whose abundances are dictated more by stochastic recruitment processes and
357 less by adult survivorship across gradients in natural physical drivers (Margalef 1969, Pianka
358 1970, Reznick et al. 2002, Darling et al. 2013). The result is an overall decoupling of natural
359 biophysical relationships within the system. To test this working hypothesis would require data

360 across similar spatial scales but at a finer taxonomic resolution and likely reveal more intricate
361 ways in which human impacts disrupt and decouple natural biophysical relationships on coral
362 reefs. Furthermore, we require information pertaining to the types of human impacts occurring
363 across our large spatial gradient. Our binary split of ‘unpopulated’ and ‘populated’ islands as a
364 proxy for local human impacts, while necessary for the current analysis, is likely reducing model
365 performance. For example, in a meta-analysis by Cinner et al. (2013), distance of the island to
366 local markets was a better predictor of total reef fish biomass than variations in local human
367 population density. Quantifying the types of human impacts and their severity across our
368 gradient would likely improve future interpretation of the spatial patterns in benthic cover we
369 observed. Also, our analysis did not examine for the effects of intra-island spatial gradients in
370 physical drivers on biophysical relationships. Wave energy, for example, often varies around an
371 island, leading to complex intra-island patterns of benthic cover (Williams et al. 2013, Gove et
372 al. *In Press*). Human impacts to biophysical relationships may not, therefore, be uniform around
373 entire islands and future work should focus on determining the range of spatial scales at which
374 human-induced biophysical decoupling is evident. Finally, changes in reef fish biomass are often
375 more sensitive to local human impacts than changes in benthic cover on coral reefs (McClanahan
376 et al. 2011). Future work should therefore focus on whether a similar phenomenon of biophysical
377 decoupling occurs within the reef fish community across similar spatial scales on coral reefs.

378 **Conclusion**

379 Using 39 Pacific coral reef islands, we show that considerable spatial variation exists in
380 the cover of three benthic groups (hard coral, crustose coralline algae, and macroalgae), even in
381 the absence of local human impacts, as a result of broad spatial gradients in key physical
382 environmental drivers. We further show that local human impacts appear capable of disrupting

383 and fundamentally altering these natural biophysical relationships, a process we term *biophysical*
384 *decoupling*. At populated islands, chronic human impacts do not simply result in wholesale
385 changes in benthic cover, instead they likely alter communities in such a way that they are no
386 longer reflective of (or reactive to) the background physical regimes in which they reside. These
387 findings demonstrate the ability of local human impacts to fundamentally disrupt the natural
388 rules of nature in the marine environment.

389

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397

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534

535 **Supplementary Material**

536 **Appendix 1.** Supplementary partial residual plots from the generalized additive mixed-effects
537 models (GAMM).

538 **Appendix 2.** Supplementary methods for benthic community surveys, satellite-derived and
539 modeled predictor variable data, and statistical modeling procedures.

540 **Appendix 3.** Supplementary tables, including location and timing of all benthic surveys,
541 predictor variable codes and units, predictor variable correlation values, and summary outputs of
542 GAMMs across all 39 islands.

543 **Appendix 4.** R code for our *all.subsets.gamm* convenience wrapper function.

544

545 **Figure legends**

546 Figure 1. Location of the 39 U.S.-affiliated coral reef islands surveyed within four geopolitical
547 regions in the Pacific Ocean: Hawaiian Archipelago (Northwestern Hawaiian Islands and the
548 Main Hawaiian Islands), Mariana Archipelago, the Pacific Remote Island Areas (PRIAs), and
549 the islands of American Samoa. Islands are classified as unpopulated (filled circles) and
550 populated (open circles). Island codes are defined in full in Appendix 3 (Table A1).

551 Figure 2. Island mean (+ 95 % confidence interval) percent cover of three benthic groups: hard
552 coral (a), crustose coralline algae (b), and macroalgae (c) across 39 U.S.-affiliated coral reef
553 islands spanning four geopolitical regions in the Pacific Ocean (left panels). Islands are classified
554 as unpopulated (U, grey shading) or populated (P, no shading). Bootstrapped means and 95 %
555 confidence intervals (generated using 10,000 random iterations of the data with replacement) are
556 shown across a series of hierarchical levels (right panels): *geopolitical region*, *island status*
557 (unpopulated *versus* populated islands) within each region, and *island status* across all 39 islands
558 (ALL, far right). Samoa, American Samoa; PRIAs, Pacific Remote Island Areas.

559 Figure 3. Generalized additive mixed-effects models (GAMM) showing the influence of
560 predictor variables on hard coral cover variation across unpopulated and populated islands in the

561 U.S. Pacific. a) best-fit model across unpopulated islands ($n = 24$), b) unpopulated island best-fit
562 model structure fitted anew to populated islands ($n = 15$), c) unpopulated island best-fit model
563 structure fitted anew to populated islands with the forced inclusion of variations in human
564 population density as a predictor, and d) best-fit model across populated islands. Data points
565 represent distribution of standardized partial residuals (SPR). The y-axes are on the scale of the
566 linear predictor of the model, i.e. the units are link (response units) with the smooth terms
567 centered to ensure model identifiability (sum to zero over the covariate values). The number
568 center-top within each plot represents the relative importance of each predictor variable (sum of
569 AICc model weights across all models containing each predictor). The deciles of the distribution
570 of the predictors are indicated by tick marks along the bottom of the plots. Black lines represent
571 model-fitted splines of the estimated smoothing functions bounded by 95 % Bayesian credible
572 intervals (solid grey shading).

573 Figure 4. Generalized additive mixed-effects models (GAMM) showing the influence of
574 predictor variables on crustose coralline algae cover variation across unpopulated and populated
575 islands in the U.S. Pacific. See Fig. 3 legend for details.

576 Figure 5. Generalized additive mixed-effects models (GAMM) showing the influence of
577 predictor variables on macroalgae cover variation across unpopulated and populated islands in
578 the U.S. Pacific. See Fig. 3 legend for details.

579 Figure 6. *Biophysical decoupling* on Pacific coral reefs – the natural biophysical relationships
580 predominating at unpopulated islands decouple (i.e. flat-line or dramatically weaken) or alter in
581 the direction of their relationship at populated (human impacted) islands. The percentages
582 indicate the overall spatial variation in benthic groups explained by the physical environmental

583 drivers: hard coral (top), crustose coralline algae (middle), and macroalgae (bottom). SPR,
584 standardized partial residuals. Temp, mean sea-surface temperature; WAV_mag, mean
585 magnitude of wave anomalies; Chl-*a*, mean chlorophyll-*a*; CAV, frequency of chlorophyll-*a*
586 anomalies; CAV_mag, mean magnitude of chlorophyll-*a* anomalies. Imagery obtained from the
587 University of Maryland Center for Environmental Science (ian.umces.edu/imagelibrary/).