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1 **Climate, host phylogeny and the connectivity of host communities govern**  
2 **regional parasite assembly**

3

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37 **ABSTRACT**

38 **Aim** Identifying barriers that govern parasite community assembly and parasite  
39 invasion risk is critical to understand how shifting host ranges impact disease  
40 emergence. We studied regional variation in the phylogenetic compositions of  
41 bird species and their blood parasites (*Plasmodium* and *Haemoproteus* spp.) to  
42 identify barriers that shape parasite community assembly.

43 **Location** Australasia and Oceania

44 **Methods** We used a dataset of parasite infections from >10,000 host individuals  
45 sampled across 29 bioregions. Hierarchical models and matrix regressions were  
46 used to assess the relative influences of interspecies (host community  
47 connectivity and local phylogenetic distinctiveness), climate and geographic  
48 barriers on parasite local distinctiveness and composition.

49 **Results** Parasites were more locally distinct (co-occurred with distantly related  
50 parasites) when infecting locally distinct hosts, but less distinct (co-occurred  
51 with closely related parasites) in areas with increased host diversity and  
52 community connectivity (a proxy for parasite dispersal potential). Turnover and  
53 the phylogenetic symmetry of parasite communities were jointly driven by host  
54 turnover, climate similarity and geographic distance.

55 **Main conclusions** Interspecies barriers linked to host phylogeny and dispersal  
56 shape parasite assembly, perhaps by limiting parasite establishment or local  
57 diversification. Infecting hosts that co-occur with few related species decreases a  
58 parasite's likelihood of encountering related competitors, perhaps increasing  
59 invasion potential but decreasing diversification opportunity. While climate  
60 partially constrains parasite distributions, future host range expansions that  
61 spread distinct parasites and diminish barriers to host shifting will likely be key  
62 drivers of parasite invasions.

63

64 **Key words:** community assembly, host shifting, host specificity, interspecies  
65 barriers, parasite invasion, *Plasmodium*

66

67 **INTRODUCTION**

68 Regional variation in community composition is a central property in nature  
69 (Wallace, 1876; Kraft *et al.*, 2007). With increasing environmental destabilisation  
70 and biotic homogenisation, predicting how ecosystems will function following  
71 disturbance relies on identifying processes that govern **community assembly**  
72 (Ricklefs, 1987; Barnagaud *et al.*, 2014; see Table 1 for bold term definitions).  
73 Understanding parasite community assembly is crucial, as changes to parasite  
74 composition or the frequency of host-parasite interactions can alter risks of  
75 parasite invasions and emerging disease (Brooks & Hoberg, 2007; Hoberg &  
76 Brooks, 2008; Lafferty, 2009; Agosta *et al.*, 2010; Adlard *et al.*, 2015).

77       A strong incentive exists to identify barriers to species establishment and  
78 determine how these barriers modulate invasion risk (Hoberg, 2010; Kelly *et al.*,  
79 2009; Springborn *et al.*, 2015). For parasites, **geographic barriers** (such as  
80 distance or mountain ranges) are known to constrain species' distributions  
81 (Brooks & Ferrao, 2005; Lafferty, 2009; Warburton *et al.*, 2016; Krasnov *et al.*,  
82 2016). In addition, **environmental barriers** (such as temperature and  
83 precipitation) drive development or transmission rates for many parasites,  
84 especially vector-borne parasites such as those causing malaria and lyme disease  
85 (Githeko *et al.*, 2000; Epstein, 2001; Patz *et al.*, 2005). However, parasite  
86 distributions are also linked to host life histories and distributions (Poulin *et al.*,  
87 2011; Olsson-Pons *et al.*, 2015; Fecchio *et al.*, 2017). Such **interspecies barriers**  
88 are increasingly recognised to govern local assembly (HilleRisLambers *et al.*,  
89 2012; Wisz *et al.*, 2013; Mayfield & Stouffer, 2017). Predicting how parasite  
90 composition may change in the future relies on defining a consistent framework  
91 to identify patterns that improve knowledge of assembly and elucidate

92 underlying mechanisms acting as barriers. Such patterns may be driven by a  
93 hierarchical process, where parasites must first break through geographic  
94 and/or environmental barriers to initially colonise a new range (Brooks &  
95 Hoberg, 2007; Agosta *et al.*, 2010). Following colonisation, assembly may be  
96 limited by interspecies barriers that govern parasite spread and diversification  
97 (Fig. 1). This process, termed ‘ecological fitting’ (Janzen, 1985), suggests many  
98 parasites are capable of infecting a broader range of hosts than is currently  
99 realised, with changes to host and/or parasite distributions producing new  
100 associations that may be limited by host phylogenetic relationships (Brooks &  
101 Ferrao, 2005; Radtke *et al.*, 2007; Araujo *et al.*, 2015).

102 For parasites that rely on host dispersal to colonise new areas, regions  
103 comprising a diversity of host species whose ranges overlap with other potential  
104 hosts (i.e. high distributional connectivity to other regions; **host community**  
105 **connectivity**) should support broader parasite diversity due to increased niche  
106 space (Hector *et al.*, 2001) and a higher likelihood for parasites to break  
107 geographic and/or environmental barriers (Fig. 1). However biotic barriers  
108 could still limit parasite invasions in phylogenetically diverse systems,  
109 particularly if invasion success is positively related to the invader’s **local**  
110 **phylogenetic distinctiveness** (i.e. more locally distinct invaders are less likely  
111 to be limited by related competitors; HilleRisLambers *et al.*, 2012). Yet while host  
112 community connectivity can overcome geographic dispersal barriers, few studies  
113 recognise this aspect as a potential driver of parasite assembly (but see Buckee  
114 *et al.*, 2007).

115 Parasites are often restricted to hosts with phylogenetically conserved  
116 ecological or physiological traits (Janzen, 1968; Rohde, 1980; Streicker *et al.*,

117 2010; Schulze-Lefert & Panstruga, 2011), a phenomenon that has powerful  
118 consequences for species interactions and ecosystem functioning (Ehrlich &  
119 Raven, 1964; Hoberg & Brooks, 2008). As parasites with high **host specificity**  
120 may be unable to shift hosts, the local availability of suitable hosts can present an  
121 invasion barrier following initial dispersal, especially if parasites are adapted to  
122 hosts that do not commonly co-occur with closely related species (Brooks, 1979;  
123 Ewen *et al.*, 2012; Clark & Clegg, 2015; Ellis *et al.*, 2015; Mata *et al.*, 2015; Fig. 1).

124         While ecological fitting (governed at least partly by parasite host  
125 specificity and host evolutionary history) and host dispersal potential are clearly  
126 important mechanisms impacting parasite establishment and diversification,  
127 identifying their roles in natural host-parasite systems is challenging. We  
128 develop a framework to identify relative influences of barriers to regional  
129 parasite community assembly, and apply this framework to naturally-occurring  
130 parasite infections from Australasian bird communities. Haemosporidians  
131 (genera *Plasmodium* and *Haemoproteus*) are vector-borne blood parasites that  
132 display a range of host specificities (Križanauskienė *et al.*, 2006). Due to limited  
133 vector dispersal (Ejiri *et al.*, 2011), avian hosts are the primary vehicles by which  
134 these parasites disperse (Pérez-Tris & Bensch, 2005). Avian haemosporidians  
135 have been introduced to numerous bioregions, sometimes with devastating  
136 effects on native birds, raising questions about how interspecies and geographic  
137 barriers regulate parasite assembly and invasion potential (van Riper III *et al.*,  
138 1986; Hellgren *et al.*, 2014).

139         We assess barriers that may govern parasite local coexistence at the  
140 species level by estimating effects of host community connectivity and  
141 interspecies barriers (host phylogeny and parasite host specificity) on parasite

142 local phylogenetic distinctiveness. We then address barriers at the community  
143 level by (1) exploring effects of host **phylogenetic turnover**, environmental  
144 variation and geographic distance on parasite turnover and (2) testing if host  
145 connectivity or environmental variation influence parasite **phylogenetic**  
146 **community skewness**. We expect that increased host community connectivity  
147 reduces barriers to parasite establishment, leading to phylogenetically  
148 homogenised parasite communities. If host phylogeny acts as a relatively strong  
149 interspecies barrier to parasite assembly, we expect that distinct hosts carry  
150 distinct parasites and that between-region host turnover predicts parasite  
151 turnover. We also expect host-specialist parasites to be more locally distinct than  
152 generalists, as specialists may have less opportunity to diversify through host  
153 range expansions. Alternatively, if higher diversities of host specialists are able  
154 to co-occur through extensive niche packing (Ricklefs, 2010), then we expect  
155 specialists to be less distinct than generalists.

156

## 157 **METHODS**

### 158 **Host-parasite occurrence data and avian community connectivity**

159 We surveyed published literature and queried the MalAvi database  
160 (<http://mbio-serv2.mbioekol.lu.se/Malavi/>; accessed September 2016; Bensch  
161 *et al.*, 2009) to compile data from >10,000 sampled host individuals (from 297  
162 avian species) across 83 sites, ranging across latitudes -50.77 to 14.27 and  
163 longitudes -159.78 to 178.07 (Fig. 2). In all cases, parasite lineages were  
164 identified using PCR targeting the cytochrome-*b* (*cyt-b*) gene (Hellgren *et al.*,  
165 2004; Waldenström *et al.*, 2004). Evidence indicates lineages differing by as little  
166 as one base pair may be reproductively isolated (Bensch *et al.* 2004). We thus

167 regard each unique sequence as a parasite 'species'. Low numbers of recovered  
168 parasites at some sites meant we could not assess within-site composition. We  
169 thus grouped sites into 29 regions. Australian mainland sites were grouped by  
170 climate zone using the Bureau of Meteorology's Köppen classification, which  
171 defines zones using temperature, precipitation and vegetation data  
172 ([http://www.bom.gov.au/jsp/ncc/climate\\_averages/climate-classifications/](http://www.bom.gov.au/jsp/ncc/climate_averages/climate-classifications/);  
173 accessed November 2016). Papua New Guinea mainland sites were grouped  
174 based on elevation (highlands, mean altitude = 2500m; and lowlands, mean  
175 altitude = 60m). Island sites were either grouped by island (if at least three  
176 parasite species were recovered) or into regions representing nearby islands in  
177 an archipelago (Fig. 2; Supplementary Dataset 1).

178 We downloaded range maps for all avian species occurring in the study  
179 area (N = 3,024 species) from BirdLife International and NatureServe  
180 (<http://www.birdlife.org/datazone>; accessed October 2016). For each region, we  
181 obtained lists of occurring avian species (defined as the 'total' assemblage) by  
182 recording all species whose ranges overlapped 111 km buffers (1° at the  
183 equator) around sites. Bird range sizes were calculated as the total area of range  
184 polygons. Range sizes varied from 1km<sup>2</sup> (island endemics) to 28,000km<sup>2</sup> (wide  
185 ranging seabirds).

186 Avian community connectivity was calculated as an inverse Simpson  
187 diversity index (Simpson 1949) using species' range sizes as weights (instead of  
188 using species abundances). Here increased species richness, larger species range  
189 sizes and more even range size distributions all lead to increased collective  
190 mobility of a local host assemblage. Two connectivity indices were created, one  
191 using sampled hosts (***Sampled.Con<sub>H</sub>***) and a second using total assemblages (all

192 occurring avian species; **Total.Con<sub>H</sub>**). We included *Total.Con<sub>H</sub>* because many  
193 haemosporidians infect a diversity of avian species (Ewen *et al.* 2012; Olsson-  
194 Pons *et al.* 2015), suggesting unsampled but present host species impact parasite  
195 assembly. This will be especially relevant for generalist parasites, whereas  
196 sampled hosts should be representative for specialised parasites that are  
197 unlikely to occur in unsampled host species.

198

### 199 **Parasite and host phylogenetic reconstructions**

200 Parasite *cyt-b* sequences (205 *Haemoproteus* and 80 *Plasmodium* parasites) were  
201 used to reconstruct phylogenetic relationships in BEAST v1.8.1 (Drummond &  
202 Rambaut, 2007; See Fig. S1 in Supplementary Material). We identified the best  
203 evolutionary model (HKR+G) using maximum likelihood in MEGA v7.0 (Tamura  
204 *et al.*, 2007). We specified a Yule speciation prior and ran two chains of  
205 17,500,000 iterations, sampling every 100,000 and removing 2,500,000 samples  
206 as burn-in. Chains were examined visually for stationarity and convergence.

207 Avian phylogenies were gathered from Birdtree.org (<http://birdtree.org>;  
208 accessed September 2016), which contains a Bayesian posterior distribution of  
209 phylogenies for 9,993 avian species (Jetz *et al.*, 2012). We gathered 100 trees  
210 from the 'Ericsson All Species Trees' dataset for the 297 sampled host species,  
211 and another 100 trees for the 3,024 avian species occurring in the sample area.  
212 For all trees, branch lengths represented substitutions per site and were scaled  
213 (dividing branch lengths by the maximum) prior to analyses.

214

### 215 **Species level analyses**

216 *Host and parasite phylogenetic distinctiveness*

217 For sampled host species, local phylogenetic distinctiveness (***Sampled.Dis<sub>H</sub>***) was  
218 calculated as mean pairwise phylogenetic distance between a focal species and  
219 all other sampled host species in a region. This distance was divided by the mean  
220 of all pairwise distances in the region, resulting in region-specific distinctiveness  
221 (higher values indicating more distinct species). We calculated total host  
222 distinctiveness (***Total.Dis<sub>H</sub>***) using mean phylogenetic distance between a  
223 sampled host and all occurring avian species (sampled and unsampled) in a  
224 region. Parasite distinctiveness (***Dis<sub>P</sub>***) was calculated separately for each parasite  
225 genus.

226

#### 227 *Parasite host specificity*

228 Two indices described parasite host specificity. First, we built bipartite networks  
229 (using numbers of infected individuals for each host species) and calculated the  
230 ***d'*** specialisation index using Kullback-Leibler distances (Blüthgen *et al.*, 2006).  
231 Ranging from zero (no specialisation; i.e. using all available hosts) to one (perfect  
232 specialist), ***d'*** quantifies how strongly a parasite is 'specialised' compared to  
233 other parasites in terms of host range and interaction frequencies. We calculated  
234 phylospecificity for each parasite (***STD\****; Poulin & Mouillot, 2005), which  
235 accounts for the number of infected host species and their phylogenetic  
236 distances. Because ***STD\**** ranges from one (specialist) to greater than one, we  
237 used inverse ***STD\**** so both metrics could be interpreted in the same scale and  
238 direction. Parasite ***STD\**** and ***d'*** were uncorrelated (Pearson correlation;  $t = -1.41$ ,  
239  $p = 0.16$ ), suggesting they capture different aspects of parasite host specificity (***d'***  
240 capturing the level of host sharing by parasites and ***STD\**** capturing phylogenetic  
241 relationships of infected hosts).

242

243 *Influences of host community connectivity, host phylogeny and host specificity on*244 *parasite distinctiveness*

245 We tested whether interspecies barriers influenced parasite distinctiveness

246 ( $Dis_P$ ) with a hierarchical linear model, using 548 unique parasite\*host\*region247 combinations as data points (Supplementary Dataset 2). Because  $Dis_P$  indices

248 were non-negative and positively skewed, we log transformed values and

249 specified a Gaussian error distribution. Continuous predictors were the two host

250 distinctiveness metrics ( $Sampled.Dis_H$ ,  $Total.Dis_H$ ), the two host connectivity251 metrics ( $Sampled.Con_H$ ,  $Total.Con_H$ ), host geographic range and both parasite host252 specificity metrics ( $d'$ ,  $STD^*$ ). Because parasite genera showed different253 phylogenetic patterns (see Results) and  $Total.Dis_H$  explained a significant254 proportion of variance in  $Dis_P$  in preliminary analyses, we tested a255  $Total.Dis_H$ \*parasite genus interaction. To decompose variation among covariates

256 and account for underlying phylogeographic structure, host phylogeny and

257 sample region were included as random grouping terms, allowing inferences for

258 group-specific slopes whilst estimating between-group variation (Gelman &amp; Hill,

259 2007).

260 The model was fitted in a Bayesian framework using R package

261 MCMCglmm (Hadfield, 2010). We used a flat prior for residual variance and

262 parameter expansion (redundant multiplicative reparameterisation of the linear

263 model) for grouping terms, which reduces dependence among parameters and

264 improves mixing (Gelman, 2006). To account for phylogenetic uncertainty, we

265 ran separate models across 50 host trees (Guillerme &amp; Healy, 2014). Models

266 were run using two chains of 100,000 iterations with burn-in of 10,000 and

267 thinning interval of 300. Chains were inspected for mixing/convergence both  
268 visually and with the Gelman-Rubin diagnostic (Gelman & Rubin, 1992).  
269 Autocorrelations were calculated to ensure independence of coefficient  
270 estimates (all autocorrelations < 0.1).

271

## 272 **Community analyses**

### 273 *Interspecies and geographic barriers to parasite phylogenetic turnover*

274 To describe shifts in diversity among regions, parasite phylogenetic turnover  
275 ( $\beta_P$ ) was calculated (using binary occurrence data; Tsirogiannis & Sandel, 2015)  
276 between regions where three or more parasites occurred. Host turnover was  
277 calculated using either sampled hosts (***Sampled. $\beta_H$*** ) or total avian assemblages  
278 (***Total. $\beta_H$*** ). Distances between paired regions were calculated as beeline distance  
279 (km) between central points (mean latitude and longitude of regions). Regional  
280 climate dissimilarity was captured by three Gower's distance matrices (Gower,  
281 1971) to describe temperature and precipitation variation (both of which are  
282 thought to influence haemosporidian distributions; Sehgal *et al.*, 2010; Sehgal,  
283 2015). We used minimum temperature of the coldest month and mean  
284 temperature of the coldest quarter in a *min.temp* matrix, while a *max.temp*  
285 matrix included maximum temperature of the warmest month and mean  
286 temperature of the warmest quarter. Mean yearly precipitation and  
287 precipitations of the wettest and driest quarters were included in a *precip*  
288 matrix. For climate matrices, variables were sourced from ([www.worldclim.org](http://www.worldclim.org);  
289 accessed November 2016) and were continuous, unweighted and scaled by  
290 range (dividing by the maximum).

291 We tested if  $\beta_P$  was correlated with *Sampled. $\beta_H$* , *Total. $\beta_H$* , geographic

292 distance or climate dissimilarity matrices using multiple regressions on distance  
293 matrices (MRM; Goslee & Urban, 2007). Phylogenetic uncertainty was captured  
294 by repeating regressions over 1,000 iterations, where  $\beta$  values were re-  
295 calculated in each iteration using randomly sampled (with replacement) trees.  
296 To account for sampling variation that could bias turnover estimates (rare  
297 species may be more likely to be observed with larger sample sizes), we  
298 randomly removed subsets of species from well-sampled regions (>8 observed  
299 parasite species) prior to regression. We arbitrarily allowed the proportion of  
300 removed species to vary across a uniform distribution from zero to 30% in each  
301 iteration. Regression coefficients and  $R^2$  values were gathered from the 1,000  
302 iterations.

303

#### 304 *Barriers to parasite phylogenetic community skewness*

305 Host and parasite phylogenetic community skewness were calculated using  
306 pairwise phylogenetic distance distributions. A measure of symmetry, this index  
307 will be less than zero (right skewed) if communities are made up of relatively  
308 more closely than distantly related species (Schweiger *et al.*, 2008), suggesting  
309 future colonising parasites have a greater likelihood of being locally distinct.  
310 Thus, regions with right skewed communities may be more vulnerable to  
311 invasions by distantly related species if parasites are able to overcome  
312 environmental barriers and colonise. Skewness was calculated for regions where  
313 three or more parasites occurred.

314 We tested if parasite skewness was predicted by host connectivity  
315 (*Sampled.Con<sub>H</sub>*, *Total.Con<sub>H</sub>*) using linear regression with Gaussian error  
316 distribution. Mean annual precipitation and mean temperatures of the warmest

317 and coldest quarters were included as continuous covariates to account for  
318 possible climate influences, while sampled and total host skewness were  
319 included to account for influences of host phylogenetic symmetry. Parasite genus  
320 was included as a categorical covariate. The model was fitted using MCMCglmm  
321 with a flat prior for residual variance. We ran two chains of 100,000 iterations  
322 with burn-in of 10,000 and thinning interval of 300, following procedures above  
323 to examine convergence and estimate autocorrelations.

324         For all phylogenetic metrics (skewness, distinctiveness and *STD\**), we  
325 accounted for phylogenetic uncertainty by calculating median indices across  
326 1,000 randomly sampled host and parasite trees. Significance of model effects  
327 was determined by examining if 95% quantiles (for MRM models) or 95%  
328 credible intervals (CI; for Bayesian models) of regression coefficients did not  
329 overlap zero. Continuous predictors were scaled (centred and divided by one  
330 standard deviation), and variances explained were calculated following  
331 Nakagawa & Schielzeth (2013). Data was analysed in R v3.2.1 (R Core Team,  
332 2016; R: A language and environment for statistical computing). Data and R code  
333 are presented in Supplementary Data and the Dryad Digital Repository: (doi:  
334 XXXXX XXXXX).

335

## 336 **RESULTS**

### 337 **Host phylogeny, local distinctiveness and connectivity drive parasite**

#### 338 **distinctiveness**

339 Parasite distinctiveness (*Dis<sub>P</sub>*) was strongly related to host phylogeny (variance  
340 explained = 46.8 to 78.3%), with hosts from certain clades more likely to carry  
341 distinct parasites (Fig. 3). These included carriers of distinct *Haemoproteus* spp.

342 such as doves (Columbidae), kingfishers (Alcedinidae) and corvids such as  
 343 crows (Corvidae) and whistlers (Pachycephalidae; Fig. 3), all of which occupy a  
 344 range of regions yet rarely co-occur with sympatric sister species (Dutson, 2012;  
 345 Jønsson *et al.*, 2014). After accounting for the strong influence of host phylogeny,  
 346  $Dis_P$  was also positively predicted by local host total distinctiveness ( $Total.Dis_H$ ;  
 347 coefficient 95%CI = 0.04 to 0.12; variance explained = 2.48 to 6.38%; Fig. 3),  
 348 suggesting host relatedness to the local avian assemblage acts as an interspecies  
 349 barrier to parasite assembly. This relationship varied between parasite genera,  
 350 as increases in  $Total.Dis_H$  lead to a 1.95 times higher increase in  $Dis_P$  for  
 351 *Haemoproteus* than for *Plasmodium* parasites.

352  $Dis_P$  decreased with increasing total host connectivity ( $Total.Con_H$ ;  
 353 coefficient = 0.01 to 0.09; variance explained = 0.04 to 7.7%; See Fig. S2 in  
 354 Supplementary Material), indicating greater host diversity and collective  
 355 mobility increases a parasite's chance of encountering related parasites.  
 356  $Total.Con_H$  was highest in Malaysia (509 avian species;  $Total.Con_H = 83.60$ ) and  
 357 southeast Australia (468 avian species;  $Total.Con_H = 80.42$ ), moderate in Papua  
 358 New Guinea where many endemic avian species occur (mean species = 520.5;  
 359 mean  $Total.Con_H = 42.62$ ) and lowest in Vanuatu and New Caledonia (mean  
 360 species = 115 and 110; mean  $Total.Con_H = 32.3$  and  $31.6$ , respectively).  $Dis_P$  was  
 361 not influenced by  $Sampled.Con_H$ ,  $Sampled.Dis_H$  or individual host range  
 362 (coefficient CIs overlapped zero).

363 We observed considerable variation in host specificity for both parasite  
 364 genera, though neither specificity metric influenced  $Dis_P$  (coefficients overlapped  
 365 with zero). For both genera,  $STD^*$  (phylospecificity) ranged from 0.41 to 1 (mean  
 366 = 0.79 and 0.87 for *Plasmodium* and *Haemoproteus*, respectively), while  $d'$

367 (network specificity) ranged from 0 to 1 (means = 0.65 and 0.67). In total, fixed  
 368 effects ( $d'$ ,  $STD^*$ , host range size,  $Total.Con_H$ ,  $Sampled.Con_H$ ,  $Total.Dis_H$ ,  
 369  $Sampled.Dis_H$ ) explained 5.7 to 13.2% of variance in  $Dis_P$  while the full model  
 370 (including host phylogeny and region grouping terms) explained 69.8 to 88.9%.  
 371

### 372 **Host phylogeny and climate shape parasite community structure**

373 We found evidence that both environmental and interspecies barriers influence  
 374 parasite turnover. For *Plasmodium*,  $\beta_P$  was positively correlated with  $Sample.\beta_H$   
 375 (MRM coefficient = 1.01 to 1.86), indicating host phylogeny influences shifts in  
 376 parasite diversity. *Plasmodium*  $\beta_P$  also correlated positively with geographic  
 377 distance (0.56 to 1.21), but negatively with  $max.temp$  (-0.09 to -0.18). For  
 378 *Haemoproteus*,  $\beta_P$  correlated positively with both host turnover metrics  
 379 ( $Sampled.\beta_H$  coefficient = 0.30 to 0.61;  $Total.\beta_H$  = 0.58 to 1.13), and with  
 380 geographic distance and  $max.temp$  (0.04 to 1.37; 0.16 to 0.45, respectively), but  
 381 negatively with  $min.temp$  (-0.11 to -0.28). Variance explained by predictors  
 382 ranged from 47 to 57% for *Haemoproteus*  $\beta_P$  and from 4 to 11% for *Plasmodium*  
 383  $\beta_P$ .

384 Mainland communities such as Papua New Guinea and eastern Australia  
 385 showed low mean parasite turnover among paired regions (low average  
 386 pairwise  $\beta_P$  after accounting for geographic distance; Fig. 2; Supplementary  
 387 Dataset 3), suggesting these assemblages were less phylogenetically unique  
 388 within the study area. Parasite assemblages on Melanesian islands (New  
 389 Caledonia and Vanuatu) showed moderate mean turnover, while relatively  
 390 isolated and less well-sampled communities such as Christmas Island and  
 391 northwest Australia showed high turnover (Fig. 2). *Plasmodium* communities in

392 New Zealand and Micronesia, where many occurring parasites are known to be  
393 introduced (Beadell *et al.* 2006; Ewen *et al.*, 2012), showed high mean turnover  
394 (Fig. 2).

395 Parasite community skewness indices were predominantly negative  
396 (right-skewed; Fig. 4), with assemblages generally made up of more closely than  
397 distantly related parasites. Parasite skewness was not influenced by host  
398 community connectivity or host skewness, but was driven by mean temperature  
399 of the coldest quarter (coefficient = 0.02 to 2.98; variance explained = 0.2 to  
400 10.6%), with colder regions harbouring more negatively skewed communities  
401 (Fig. 4). Parasite skewness also differed between genera (coefficient = -0.91 to -  
402 0.02; variance explained = 7.5 to 27.10%), with *Plasmodium* more negatively  
403 skewed than *Haemoproteus* communities (Fig. 4). Interestingly, *Haemoproteus*  
404 communities in Papua New Guinea were positively skewed, while those in  
405 eastern Australian were negatively skewed (Fig. 4), suggesting neighbouring  
406 parasite assemblages with low phylogenetic turnover (Fig. 2) can vary  
407 substantially in community structure.

408

## 409 **DISCUSSION**

410 We illustrate a framework for identifying relative influences of interspecies,  
411 environmental and geographic barriers to parasite community assembly. Using  
412 this framework, we show that host phylogeny is a key driver of local parasite  
413 assembly, while climate and the regional connectivity of host assemblages play  
414 lesser but nonetheless important roles. Moreover, host phylogeny and  
415 geographic distance were more important than environmental barriers in  
416 shaping parasite turnover, indicating alterations to host movement and

417 community composition may strongly affect parasite dispersal and invasion  
418 potential across biogeographic scales.

419

#### 420 **Barriers to parasite community assembly and their roles in parasite spread**

421 Host phylogeny was an important driver of parasite distinctiveness and species  
422 turnover, supporting suggestions that host identity drives shifts in  
423 haemosporidian diversity and implicating host evolutionary history as a  
424 determinant of regional parasite assembly (Scordato & Kardish, 2014; Fecchio *et*  
425 *al.*, 2017). Phylogenetic signals are a proxy for physical (i.e physiological,  
426 morphological, biochemical) and ecological traits, where closely related species  
427 resemble each other more than random pairs, indicating conserved attributes  
428 likely play a role in modulating interspecies barriers to regional parasite  
429 assembly (Huang *et al.* 2014). Yet an important consideration here is that we do  
430 not know which shared host traits influence blood parasite assembly patterns.  
431 Determining underlying interspecies barriers to parasite composition will  
432 require additional interdisciplinary work, combining data on host traits with  
433 methods that can decompose phylogenetic and ecological similarity to improve  
434 inference (Cadotte *et al.*, 2013; Clark & Clegg, 2017).

435 Future host range shifts may considerably impact parasite spread and  
436 disease emergence, both by breaking down existing barriers to host shifting and  
437 by increasing parasite dispersal (Atkinson & LaPointe 2009; Young *et al.*, 2017).  
438 Here, a positive relationship between host and parasite distinctiveness indicates  
439 that diminishing phylogeographic barriers (where host range shifts may alter  
440 local host distinctiveness) could present more opportunities for parasites to shift  
441 between related hosts. Yet a strong host phylogenetic signal, where distinct

442 parasites are more strongly associated with certain host clades, suggests  
443 alterations to host species' distributions may have different effects on parasite  
444 spread depending on host evolutionary history. For instance, we identified  
445 multiple host clades as prominent carriers of distinct parasites, including non-  
446 passerines (kingfishers and doves) as well as certain passerine groups (crows  
447 and whistlers), indicating that future range shifts for these host groups could  
448 lead to novel parasite introductions. Our work therefore corroborates a large  
449 body of literature to show that interactions between ecological fitting and  
450 shifting geographic distributions will have powerful influences on parasite  
451 assembly and emergence potential (Brooks & Hoberg, 2007; Hoberg & Brooks,  
452 2008; Hoberg, 2010; Agosta *et al.*, 2010; Araujo *et al.*, 2015). However, a  
453 significant influence of host community connectivity suggests that parasite  
454 distinctiveness is not only driven by host phylogeny, but also by forces that limit  
455 host diversity and distributional overlap (i.e. competitive exclusion or dispersal  
456 barriers; Ricklefs, 2010; Ewen *et al.*, 2012). This finding generates exciting new  
457 avenues for studying parasite assembly, particularly since few studies relate the  
458 connectivity of host communities to parasite dispersal opportunity (but see  
459 Buckee *et al.*, 2007).

460       Our findings that environmental effects influence parasite turnover and  
461 community skewness agree with previous studies to suggest that even if  
462 dispersal barriers break down, climate and perhaps other environmental  
463 conditions may constrain parasite distributions (Kutz *et al.*, 2014; Sehgal, 2015;  
464 Clark *et al.*, 2016a,b). Indeed, regional temperature similarity impacted shifts in  
465 diversity for both parasite genera, albeit with different directional relationships.  
466 One possible explanation could be that haemosporidians are subject to

467 influences of external temperature changes on ectothermic vectors (Paaijmans *et*  
468 *al.*, 2010), and *Plasmodium* and *Haemoproteus* parasites are transmitted by  
469 different arthropods (mosquitoes from family Culicidae and midges from family  
470 Ceratopogonidae, respectively; Santiago-Alarcon *et al.*, 2012). However, little is  
471 known about the particular vector species transmitting avian haemosporidians  
472 in the South Pacific (but see Ishtiaq *et al.*, 2008), and so drawing conclusions  
473 from these different patterns remains challenging. Intriguingly, regions with  
474 colder temperatures harboured more closely related communities for both  
475 parasite genera, perhaps indicating minimum temperatures act as a strong filter  
476 for haemosporidian diversity, a finding that warrants future study. Regardless of  
477 the biological mechanism, accounting for interspecies interactions and  
478 environmental conditions can improve predictions of species distributions  
479 following climate shifts (Choler *et al.*, 2001; Wells *et al.*, 2014; Mayfield &  
480 Stouffer, 2017).

481         Determining which species are likely to be introduced and become  
482 invasive are prominent ecological questions (Wiens, 2011; Springborn *et al.*,  
483 2015). Our results suggest that parasites introduced to regions with low host  
484 community connectivity, high host turnover and low minimum temperatures  
485 may be more likely to invade the community. These patterns highlight that New  
486 Zealand, which showed high rates of host and parasite turnover and contained  
487 distantly related (phylogenetically left skewed) *Plasmodium* communities, may  
488 be particularly vulnerable to invasions. Distinct invaders can have key  
489 competitive advantages and a greater chance of becoming invasive  
490 (HilleRisLambers *et al.*, 2012), as has been the case in the Galapagos where the  
491 invasive fly, *Philornis downsi*, parasitizes a diversity of endemic bird species

492 (Fessl *et al.* 2006). Indeed, invasive avian malaria parasites have already been  
493 recorded infecting a diversity of native New Zealand birds, with evidence  
494 suggesting that introduced birds play key roles in driving parasite spread (Ewen  
495 *et al.*, 2012; Schoener *et al.*, 2013). Parasites introduced to highly connected host  
496 regions, on the other hand, may be more likely to experience competition with  
497 closely related parasites, perhaps curbing invasion potential. Under this  
498 consideration, areas such as eastern Australia and mainland Papua New Guinea  
499 may be less vulnerable to parasite invasions (though not immune; see Clark *et*  
500 *al.*, 2015), as these regions contain a relatively balanced phylogenetic diversity of  
501 parasites and experience high host community connectivity.

502

### 503 **Accounting for unsampled host species in parasite assembly studies**

504 Our study raises a critical point for assessing parasite composition, as measures  
505 of host relationships were more important in driving parasite assembly when  
506 considering the total host assemblage rather than only sampled hosts. A host's  
507 distinctiveness with respect to the entire avian community positively predicted  
508 parasite distinctiveness, while considering only sampled hosts had no influence  
509 on parasite distinctiveness. Phylogenetic turnover of the total avian assemblage  
510 was also a stronger predictor of *Haemoproteus* turnover than was sampled host  
511 turnover. These findings imply that variation in unsampled but locally present  
512 host species are important for driving parasite establishment. Inferences beyond  
513 those obtained from sampled hosts are clearly needed, a process which is rarely  
514 considered in host-parasite interactions (but see Wells *et al.*, 2012), despite  
515 being a well-known problem in the sample survey literature (Little, 2004).

516

## 517 **Caveats and conclusions**

518 There are several ways in which our study framework can be improved. First, we  
519 did not consider individual sites in our study as our data was limited by small  
520 sample sizes for many sites. Inclusion of site-specific species and climate data  
521 could be used as an additional source of information to examine possible impacts  
522 of sampling bias on regional community inferences. Second, consideration of  
523 sampling distribution across regions may have an impact on community  
524 turnover estimates, as regions such as Christmas Island and Micronesia had a  
525 relatively high turnover that could have been influenced by low overall sample  
526 sizes and large geographic distances to many other study regions. Future studies  
527 that sample smaller and more regular geographic intervals could help to address  
528 this drawback. Finally, our phylogenetic metrics relied only on binary species  
529 occurrences (present or absent), and may be improved with better consideration  
530 of species' relative abundances, since host abundance plays a role in host  
531 reservoir potential and cross-species parasite transmission (Kilpatrick *et al.*,  
532 2006). Unfortunately, such data for host abundance were not available and  
533 would require additional field survey efforts.

534 In summary, our study agrees with previous work to suggest that in  
535 addition to identifying environmental barriers, considering host phylogenetic  
536 relationships and dispersal abilities is key to understanding regional parasite  
537 assembly (Brooks & Ferrao, 2005; Agosta *et al.*, 2010; Wells *et al.*, 2015; Sehgal,  
538 2015). Moreover, we show that accounting for the overall connectivity of the  
539 host community, rather than solely focussing on individual host species'  
540 dispersal potentials, may be crucial to predicting future parasite invasions. With  
541 the pervasive need to understand how interspecies interactions shape species

542 distributions (Wisz *et al.*, 2013), our study represents an important step towards  
543 predicting how parasite assemblages will be shaped following future global  
544 change.

545

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562

563 **SUPPLEMENTARY MATERIAL**

564 **Fig. S1:** Phylogenetic relationships of *Haemoproteus* and *Plasmodium*  
565 cytochrome-*b* sequences.

566 **Fig. S2:** Relationship between regional total host community connectivity  
567 (*Total.Con<sub>H</sub>*) and parasite local phylogenetic distinctiveness (*Dis<sub>P</sub>*).

568 **Dataset 1:** Sample locations, host species sample sizes and parasite infection  
569 prevalence across regions.

570 **Dataset 2:** Raw data used to analyse parasite local phylogenetic distinctiveness  
571 (*Dis<sub>P</sub>*)

572 **Dataset 3:** Turnover estimates and avian species richness metrics across  
573 regions.

574

575 **DATA ACCESSIBILITY**

576 Newly reported parasite sequences will be uploaded to GenBank and the MalAvi  
577 avian malaria database upon acceptance. R code and raw datasets will be  
578 uploaded as supplements and to the Dryad digital repository upon acceptance.

579

580 **BIOSKETCH**

581 Nicholas Clark is a disease ecologist interested in evolutionary ecology and the  
582 biogeography of wildlife pathogens. His research interests concern topics in  
583 computational phylogenetics, biodiversity and host-pathogen interactions, as  
584 well as large-scale patterns in community assembly.

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## 868 TABLES

869

870 **Table 1:** Glossary of definitions for proposed community assembly barriers and  
871 metrics used in analyses.

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**Community assembly:** the establishment and maintenance of local communities through arrival of potential colonists from external species pools.

**Environmental barriers:** environmental differences between regions that may govern species' distributions, including variation in macroclimate, habitat and altitude.

**Geographic barriers:** physical barriers to between-region parasite dispersal, including geographic distance, mountain ranges, and water barriers.

**Host community connectivity:** the distributional overlap of host communities among regions, taking into account host species richness and host geographic range sizes. Here, *Sampled.Con<sub>H</sub>* describes host community connectivity while considering only sampled avian host species, and *Total.Con<sub>H</sub>* describes connectivity for all occurring avian species within a local assemblage.

**Host specificity:** the range and diversity of hosts a parasite is observed to infect. Here, *d'* describes parasite host specificity using host-parasite interaction networks, while *STD\** describes phylogenetic host specificity using host phylogenetic distances.

**Interspecies barriers:** for parasites, interspecies barriers relate to variation in host species attributes that prevent parasite spread and diversification. These may include host phylogenetic relatedness and ecological similarity (e.g. microhabitat use, nesting behaviour, and feeding behaviour).

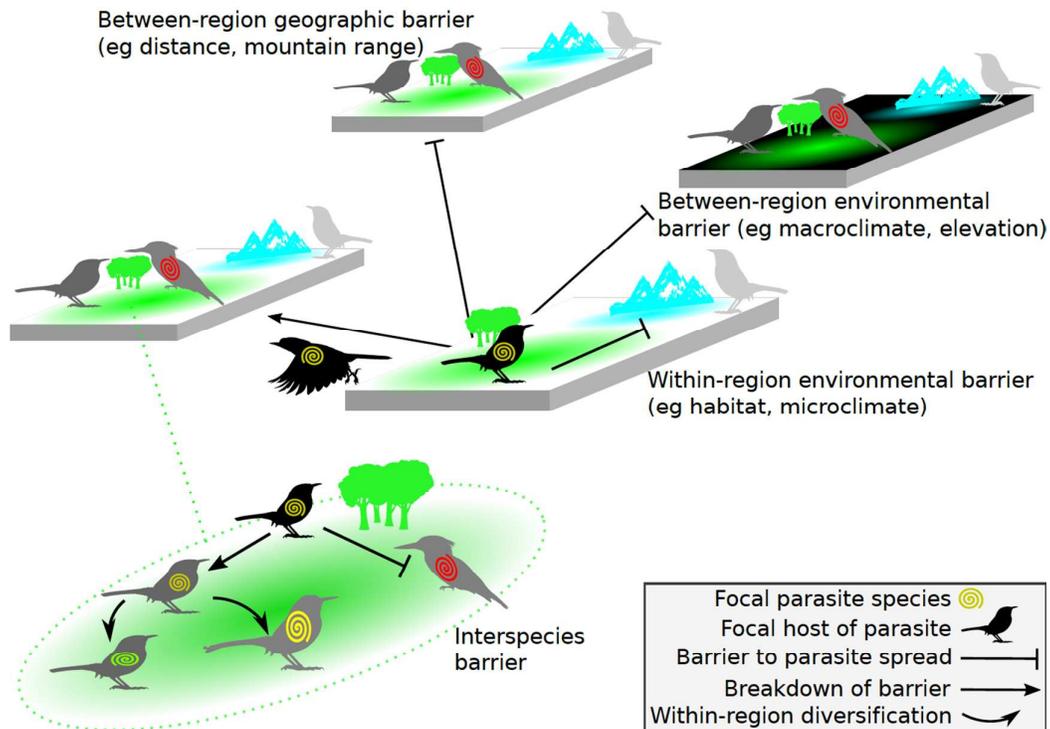
**Local phylogenetic distinctiveness:** the average pairwise phylogenetic distance between a focal taxon and co-occurring taxa within a local assemblage. Here, *Dis<sub>P</sub>* describes parasite species distinctiveness, *Sampled.Dis<sub>H</sub>* describes host species distinctiveness with respect to co-occurring sampled host species, and *Total.Dis<sub>H</sub>* describes host species distinctiveness with respect to all co-occurring sampled avian species.

**Phylogenetic community skewness:** a measure of the asymmetry of species' pairwise phylogenetic distances, where a left-skew indicates relatively more distantly than closely related species in a community, while a right-skew indicates the opposite.

**Phylogenetic turnover ( $\beta$ ):** shifts in phylogenetic diversity between communities. Here,  $\beta_P$  describes parasite phylogenetic turnover, *Sampled. $\beta_H$*  describes turnover of sampled host assemblages, and *Total. $\beta_H$*  describes turnover of total avian assemblages.

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875 **FIGURES**

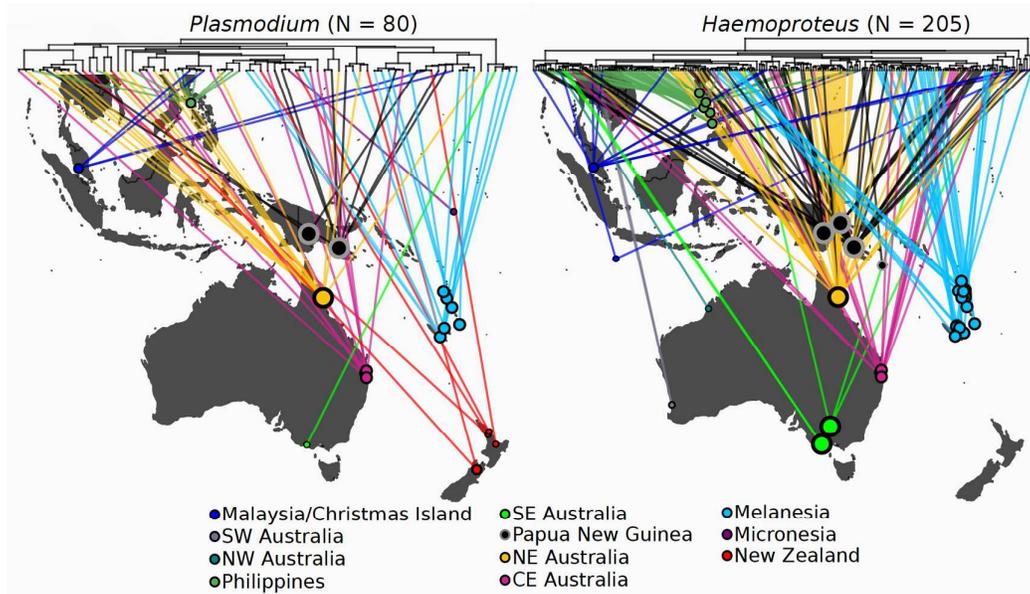
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878 **Fig. 1:** Schematic illustrating potential barriers to regional spread and  
 879 diversification for parasites that rely on host movement for dispersal. Plates  
 880 represent different bioregions, while zones (forest, mountain) within plates  
 881 represent different habitat types. At the bottom left is a sectional zoom of the  
 882 forested habitat in the left-hand plate, illustrating within-region parasite  
 883 diversification where closely related host species enable the breakdown of  
 884 interspecies barriers. Shown in black is the focal host of a given parasite species,  
 885 with ecologically or phylogenetically similar host species depicted as similar  
 886 shapes in varying shades of grey. A distantly related host species is depicted as a  
 887 different body shape. Concentric oval shapes represent parasites, with different  
 888 shapes and colours representing different parasite species.

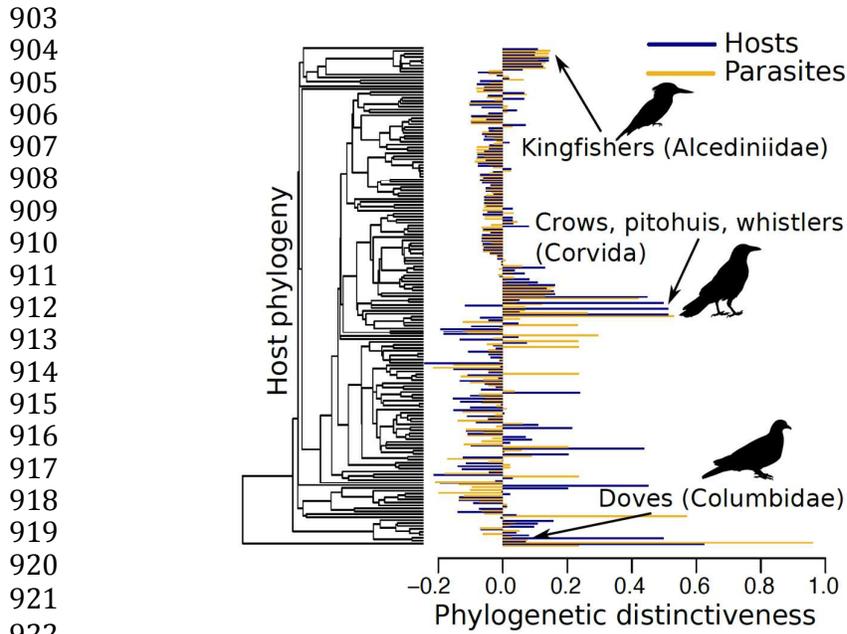
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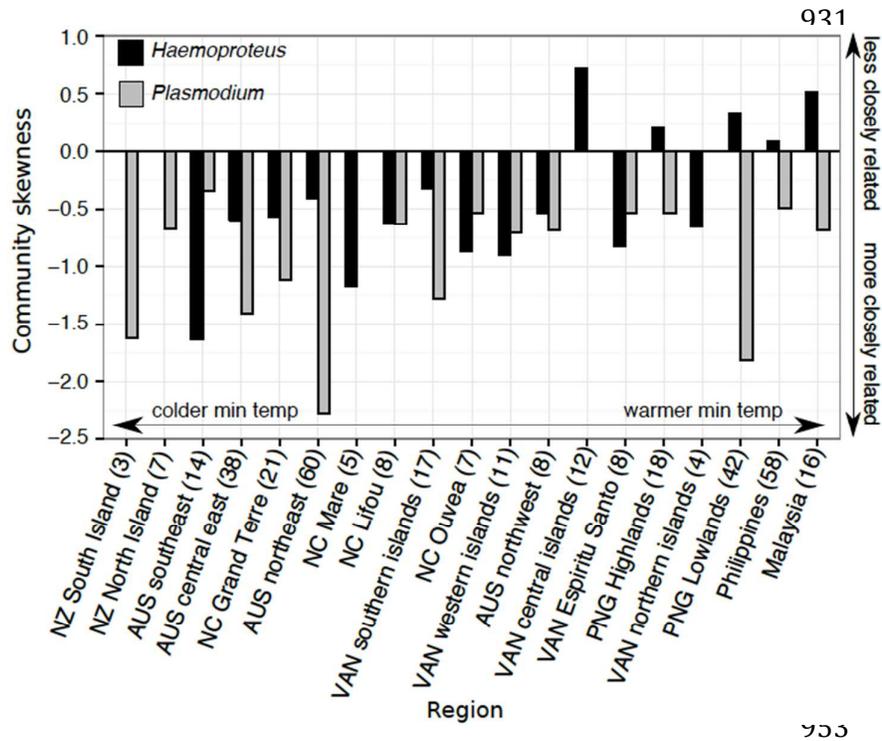


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 891 **Fig. 2:** Distribution of parasites across the study area. Lines connect phylogenetic  
 892 parasite lineages to the region where they were most frequently observed. Circle  
 893 sizes are inversely proportional to mean phylogenetic turnover ( $\beta_P$ ) between the  
 894 region and remaining regions, accounting for geographic distance. Hence, larger  
 895 circles show communities with lower mean turnover to surrounding regions,  
 896 which can be thought of as having more 'connected' parasite communities. Lines  
 897 and circles are coloured according to region, with closely situated regions  
 898 grouped to improve clarity.

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923 **Fig. 3:** Distribution of local phylogenetic distinctiveness for hosts ( $Total.Dis_H$ )  
924 and their parasites ( $Dis_P$ ) across the host phylogeny. Distinctiveness represents  
925 mean phylogenetic distance between the focal species and all co-occurring  
926 species within a region. Values are scaled so values > zero indicate taxa that are  
927 more distinct, while those < zero indicate less distinct taxa.  
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954 **Fig. 4:** Parasite phylogenetic community skewness across regions. Skewness >  
 955 zero indicates co-occurring parasites are relatively distantly related (left skewed  
 956 pairwise distance distribution), while < zero indicates parasites are relatively  
 957 closely related (right skewed distance distribution). Regions are ordered based  
 958 on mean temperature of the coldest quarter, with numbers in parentheses  
 959 indicating the number of parasites recovered in each region. NZ, New Zealand;  
 960 AUS, Australia; NC, New Caledonia; VAN, Vanuatu; PNG, Papua New Guinea.  
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