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### A HYDROGEOLOGICAL LANDSCAPE FRAMEWORK TO IDENTIFY PEATLAND WILDFIRE SMOULDERING HOTSPOTS

Hokanson, Kelly Jean; Moore, Paul A.; Lukenbach, Max; Devito, Kevin J.; Kettridge, Nicholas; Petrone, Rich; Mendoza, Carl; Waddington, James Michael

DOI:

[10.1002/eco.1942](https://doi.org/10.1002/eco.1942)

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*Document Version*

Peer reviewed version

*Citation for published version (Harvard):*

Hokanson, KJ, Moore, PA, Lukenbach, M, Devito, KJ, Kettridge, N, Petrone, R, Mendoza, C & Waddington, JM 2018, 'A HYDROGEOLOGICAL LANDSCAPE FRAMEWORK TO IDENTIFY PEATLAND WILDFIRE SMOULDERING HOTSPOTS', *Ecohydrology*. <https://doi.org/10.1002/eco.1942>

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1                   **A HYDROGEOLOGICAL LANDSCAPE FRAMEWORK**  
2                   **TO IDENTIFY PEATLAND WILDFIRE SMOULDERING HOTSPOTS**

3  
4                   **K.J. HOKANSON<sup>1,2,3\*</sup>, P.A. MOORE<sup>1</sup>, M.C. LUKENBACH<sup>1,3</sup>, K.J. DEVITO<sup>2</sup>,**  
5                   **N. KETTRIDGE<sup>4</sup>, R.M. PETRONE<sup>5</sup>, C.A. MENDOZA<sup>3</sup>, J.M. WADDINGTON<sup>1</sup>**

6  
7                   <sup>1</sup>School of Geography and Earth Sciences, McMaster University, Hamilton, ON, L8S  
8                   4K1, Canada.

9                   <sup>2</sup>Department of Biological Sciences, University of Alberta, Edmonton, AB, T6G 2E9  
10                  Canada.

11                  <sup>3</sup>Department of Earth and Atmospheric Sciences, University of Alberta, Edmonton, AB,  
12                  T6G 2E3, Canada

13                  <sup>4</sup>School of Geography, Earth and Environmental Sciences, University of Birmingham,  
14                  Edgbaston, Birmingham, B15 2TT, UK.

15                  <sup>5</sup>Department of Geography and Environmental Management, University of Waterloo,  
16                  Waterloo, ON, N2L 3G1, Canada.

17  
18                  \*Corresponding author: Kelly J Hokanson, email: hokanson@ualberta.ca

19  
20                  Keywords: peatland, wildfire, carbon, boreal, organic soil, hydrogeology,

21  
22                  Running Head: A hydrogeological framework to identify peatland smouldering hotspots

23  
24

25 **ABSTRACT**

26 Northern peatlands are important global carbon stores, but there is concern these boreal  
27 peat reserves are at risk due to increased fire frequency and severity as predicted by  
28 climate change models. In a sub-humid climate, hydrogeological position is an important  
29 control on peatland hydrology and wildfire vulnerability. Consequently, we hypothesized  
30 that in a coarse-textured glaciofluvial outwash, isolated peatlands lacking the moderating  
31 effect of large-scale groundwater flow would have greater water-table (WT) variability  
32 and would also be more vulnerable to deep WT drawdown and wildfire during dry  
33 climate cycles. A holistic approach was taken to evaluate three well accepted factors that  
34 are associated with smouldering in boreal peatlands: hollow microform coverage,  
35 peatland margin morphometry, and gravimetric water content. Using a combination of  
36 field measurements (bulk density, humification, WT position, hummock-hollow  
37 distribution, and margin width) and modelling (1-D vertical unsaturated flow coupled  
38 with a simple peat-fuel energy balance equation) we assessed the vulnerability of peat to  
39 smouldering. We found that a peatland in the regionally intermediate topographic  
40 position is the most vulnerable to smouldering due to the interaction of variable  
41 connectivity to large-scale groundwater flow and the absence of mineral stratigraphy for  
42 limiting WT declines during dry conditions. Our findings represent a novel assessment  
43 framework and tool for fire managers by providing *a priori* knowledge of potential peat  
44 smouldering hotspot locations in the landscape to efficiently allocate resources and  
45 reduce emergency response time to smouldering events.

## 46 INTRODUCTION

47 Peatland ecosystems cover 25 – 30% of boreal regions and represent a long-term sink of  
48 atmospheric CO<sub>2</sub>, storing ~ 220 – 550 Pg C (Yu, 2011). Wildfire is the largest  
49 disturbance affecting these ecosystems, accounting for >97% of all disturbances (by area)  
50 (Turetsky *et al.*, 2002). While peatlands are generally resilient to wildfire disturbance  
51 (Thompson and Waddington, 2013), northern peat fires can emit considerable amounts of  
52 CO<sub>2</sub> (*e.g.*, Turetsky *et al.*, 2002) and harmful smoke pollution (Shaposhnikov *et al.*,  
53 2014). Moreover, because the size of large (> 140,000 ha) wildfires has been shown to  
54 increase positively with peatland abundance (Turetsky *et al.*, 2004), northern peat fires  
55 also represent a challenging and costly fire management issue. These smouldering peat  
56 fires are especially challenging in sub-humid boreal regions, such as Western Canada,  
57 where the fire return interval is less than 100–120 years (Turetsky *et al.*, 2004) and the  
58 propensity for drier peat is common (Waddington *et al.*, 2015). Moreover, there is  
59 concern that peat burn severity and associated wildfire management costs will increase  
60 due to warmer and drier conditions with climate change (Turetsky *et al.*, 2004). As such,  
61 there is an urgent and growing need to identify potential hotspots for peat smouldering on  
62 the landscape to increase the efficacy of wildfire management and mitigation strategies.  
63 Here we present a landscape framework that combines moss ecohydrology, peatland  
64 hydrology, and regional hydrogeology to identify potential peat-smouldering hotspots in  
65 the Utikuma region of Alberta's Boreal Plains (BP) where peat fires are common (*e.g.*,  
66 Benschoter *et al.*, 2015; Lukenbach *et al.*, 2015).

67

68 Our hydrogeological landscape approach provides a framework for current and future  
69 research in this region, which has demonstrated that peat burn severity is higher in peat

70 profiles with low gravimetric water contents (GWC) (Rein *et al.*, 2008) and/or high peat  
71 dry bulk density ( $\rho_b$ ) (Benscoter *et al.*, 2011) and is a function of: i) *Sphagnum fuscum*  
72 (*Schimp.*) *H.Klinggr.* hummock cover (*e.g.*, Benscoter *et al.*, 2015), ii) peatland margin  
73 cover (Lukenbach *et al.*, 2015) and iii) groundwater connectivity (Hokanson *et al.*, 2016).  
74 Briefly, *S. fuscum* hummocks, which have high water retention and low  $\rho_b$ , often  
75 experience low burn severity, and in many cases are resistant to ignition (Benscoter *et al.*,  
76 2011). In the BP, margin peat is often denser and drier than peat in the central portion of  
77 the peatland due to more persistently low and/or fluctuating water-tables (see Lukenbach  
78 *et al.*, 2015 for details). Hokanson *et al.* (2016) also identified that peatlands with high  
79 groundwater connectivity had low burn severity owing to persistently higher GWC.  
80 Furthermore, Devito *et al.* (2012) illustrated the type of mineral sediment and relation to  
81 regional water-tables considerably influence location and connectedness of peatlands.  
82 Without the moderating effect of regional groundwater flow, isolated peatlands have  
83 greater WT variability, and are more vulnerable to deep WT drawdown during dry  
84 climate cycles. As such, the topographic position of a peatland in a coarse-textured HRA  
85 plays a large role in determining the hydrophysical properties of margin peat and the  
86 distribution of *S. fuscum* hummocks and therefore its vulnerability to combustion  
87 (Hokanson *et al.*, 2016).

88

89 To assess our hydrogeological landscape framework we examined a large topographic  
90 gradient, ranging from a low-lying flow-through peatland (*i.e.*, high groundwater  
91 connectivity) to a completely perched peatland (*i.e.*, no groundwater connectivity), and  
92 examined the primary hydrophysical controls on peatland burn severity and carbon loss:  
93 *S.fuscum* hummock cover, peatland margin cover,  $\rho_b$ , and GWC. We hypothesized that

94 the potential for smouldering hotspots would increase with decreasing connection from  
95 groundwater due to a decrease in higher WT buffering, an increase in percent margin  
96 cover and a decrease in the percent cover of *S.fuscum* hummocks. That is, low lying flow-  
97 through peatlands would be least vulnerable to deep smouldering due to higher WT  
98 buffering from a strong connection to the regional groundwater flow, with increasing  
99 vulnerability as the spatio-temporal connection to regional groundwater decreases.

## 100 **METHODS**

### 101 **Study sites**

102 This study was located at the Utikuma Region Study Area (URSA) located 370 km north  
103 of Edmonton, Alberta in the BP region of western Canada (Devito *et al.*, 2016). Annual  
104 potential ET often exceeds annual precipitation (517 mm and 481 mm respectively;  
105 Bothe and Abraham, 1993). Three URSA peatlands (Figure 1) were selected along a  
106 topographic gradient in the coarse-textured HRA (Figure 2d). Using historical (2003 –  
107 2014) hydrological data (Smerdon *et al.*, 2005; Devito, *et al.*, 2016; Lukenbach *et al.*,  
108 2017) each site is described below.

109

110 A low-lying flow-through kettle-hole peatland (site FT) is located on a regional  
111 topographic low in the URSA lake 208 catchment (Figure 1). Site FT is 0.8 ha and  
112 intersects a large-scale groundwater flow system connecting several ~450-900 ha lakes  
113 (Figure 1). These large groundwater-fed lakes moderate the water-table position in FT,  
114 minimizing extreme water-table fluctuations at the peatland margin and middle during  
115 periods of drought. Water-table fluctuations at FT range from 0.21 m below to 0.02 m

116 above the peat surface in the middle of the peatland, while margin water-table positions  
117 range from 0.32 m below to level with the peat surface (Table 1; Figure 2a).

118

119 A 0.68 ha peatland occupies an intermediate topographic position located in the URSA  
120 lake 16 catchment (Figure 1) and is ephemerally perched (site EP) due to transient  
121 connection to the regional water table. Site EP is located slightly above (~2.6 m) a  
122 regional groundwater flow system composed of a ‘staircase’ of lakes with an average  
123 horizontal gradient of  $0.002 \text{ m m}^{-1}$  (Smerdon *et al.*, 2005). The WT in the middle of the  
124 peatland ranges from 0.43 m to 0.93 m below the peat surface, while the margin  
125 experiences similar to greater long-term fluctuations, ranging from 0.45 m to 0.88 m  
126 below the surface (Figure 2b).

127

128 The peatland in the highest topographic position is a 1.56 ha perched peatland (site P)  
129 located in URSA lake 19 catchment (Figure 1). Site P has a laterally unconfined WT,  
130 confined vertically by layers of low permeability substrates overlying unsaturated coarse-  
131 textured sediments approximately 12 m above the regional WT. As such, P receives  
132 water solely from atmospheric inputs and has no connection to regional flow systems.  
133 The margin at P experiences large water-table fluctuations over time, ranging from 0.75  
134 m below to 0.005 m above the peat surface (Table 1; Figure 2c) while the middle of P  
135 experiences minimal water-table fluctuations, ranging from 0.41 m below to level with  
136 the peat surface (Table 1; Figure 2c).

137

138 *Study approach*

139 We mapped the coverage of margins and hummocks at each of the peatlands and  
140 undertook detailed transects to determine the peat properties at the margin and middle of  
141 each peatland. Using peat water retention data from previous work (Moore et al., 2015),  
142 we parameterize the Peat Smouldering and Ignition model (PSI) (see Thompson *et al.*,  
143 2015; Lukenbach *et al.*, 2015) to evaluate smouldering potential at each site. Details of  
144 the research design and methods are presented below.

145

#### 146 *Peatland mapping*

147 The margin zone at each site was classified using lack of peatland microtopography as an  
148 indicator of transitional plant community (see Lukenbach *et al.*, 2015) and mapped to  
149 determine the percent margin cover. The relative cover of hummocks and hollows at each  
150 site was determined by establishing two perpendicular 50 m transects in the middle of  
151 each peatland. At one meter intervals, hummock-hollow microtopography was identified  
152 1 m on either side of each transect (*i.e.*, 200 measurements per site). Peatland perimeter  
153 length was measured using DGPS points at roughly 2 m intervals. The peatland perimeter  
154 was defined by the location of a rapid transition in surface-ground cover from moss to  
155 bare soil and leaf litter, and lack of peat moss in the upper soil profile.

156

#### 157 *Peat properties*

158 A 20 m transect was established at each site perpendicular to the peatland margin  
159 extending from the outer edge of the peatland towards the middle of the peatland. Every 2  
160 m we described the presence or absence and type of surface peatland microform  
161 (hummock/hollow), measured organic soil depth (by coring), and determined vertical  
162 profiles of peat humification at 0.05 m intervals from the surface to mineral soil. The



163 degree of humification was determined using the von Post (VP) method (von Post and  
164 Granlund, 1926), which uses a categorical scale, from 1-10.

165

166 Peat cores (cross-sectional dimensions of 0.05 m x 0.05 m, depth 0.52 m) were extracted  
167 from both the margin and middle (hollow microforms only) of each peatland using a box  
168 corer to determine  $\rho_b$  (see Table 2 for sample sizes). Each monolith was sub-sampled  
169 vertically in the field at 0.04 m intervals using a serrated blade, and subsequently  
170 transported to a lab for analysis using standard methods. Peat humification was also  
171 determined on a random subset of monolith samples in order to develop a linear model  
172 for  $\rho_b$  using VP ( $F_{8,621}=189.7$   $p \ll 0.01$ ; Adjusted  $R^2$ : 0.706). Given the challenge and  
173 disturbance associated with extensive peat core extraction, this allowed us to estimate  $\rho_b$   
174 for our simulated water content profiles at depths greater than 0.52 m.

175

#### 176 *Simulated peat water content profiles*

177 Water content profiles were simulated by solving Richard's equation (Celia *et al.* 1990)  
178 for peat profiles with different specified pressure head ( $\psi$ ) boundary conditions based on  
179 water-table depth (WTD). Both wet and dry scenarios were simulated for each site (FT,  
180 EP, and P), and location (margin and middle). Zero water pressure was specified for the  
181 lower boundary condition based on the upper and lower quartile (Table 1) of measured  
182 WTDs for each site-location combination. Initial  $\psi$  was set equal to the height above WT  
183 except for the surface boundary condition. The surface boundary  $\psi$  was calculated as a  
184 function of WTD as follows (adapted from Lukenbach *et al.*, 2015):

$$185 \quad \begin{aligned} \psi &= -(WTD + 0.02 \cdot (WTD - 0.4)) & WTD > 0.4 \text{ m} \\ \psi &= -WTD & WTD \leq 0.4 \text{ m} \end{aligned} \quad (1)$$

186 where  $\psi$  for WTD>0.4 m reflects typical measured disequilibrium conditions in the near  
187 surface. Steady-state  $\psi$  profiles were iteratively solved using the finite-difference  
188 discretization of the mixed form of Richard's equation (Celia *et al.* 1990). Simulations  
189 were evaluated using 0.04 m thick layers, where a steady-state condition was defined by a  
190 maximum change in  $\psi$  of  $1 \times 10^{-5}$  m. Layer properties for upper 0.52 m were based on  
191 measured  $\rho_b$  profiles for each site-location combination, where 100 profiles per site-  
192 location were generated by randomly sampling from layer-specific distributions using the  
193 mean and standard deviation of measured  $\rho_b$  (Table 2). A similar approach was used to  
194 simulate peat layers below 0.52 m depth, but where  $\rho_b$  was derived from the linear model  
195 relating VP to  $\rho_b$  (see *Peat properties*). Error estimates on the linear model coefficients  
196 were used to account for the variance in  $\rho_b$  associated with a given value of VP.  
197 To parameterize saturated hydraulic conductivity ( $K_{\text{sat}}$ ), we opted to use the  $\rho_b$ -dependent  
198 equation presented in Boelter (1969). Uncertainty associated with our parameterization of  
199  $K_{\text{sat}}$  was not assessed in our analysis. Water retention and associated van Genuchten  
200 parameters were estimated from empirical relations between  $\psi$ ,  $\rho_b$ , and water content as  
201 presented in Moore *et al.* (2015):

$$\frac{\theta_{\psi}}{\phi} = \frac{(a \cdot \ln \psi + b)^{-1} \cdot \rho_b}{\sqrt{1 + ((a \cdot \ln \psi + b)^{-1} \cdot \rho_b)^2}}$$

202 where  $\theta_{\psi}$  is the volumetric water content at a specific  $\psi$ ,  $\phi$  is the porosity and  $a$  and  $b$  are  
203 fitted parameters. Empirical parameters were derived from water retention of peat  
204 samples from the URSA (Thompson and Waddington, 2013; Lukenbach *et al.*, 2015). To  
205 reduce the degrees of freedom, simulated profiles reflect water retention properties of  
206 hollow peat only, with corresponding  $a$  and  $b$  values of  $38.3 \pm 0.9$  and  $28.6 \pm 7.2$ ,  
207 respectively.

208

209 *Peat Smouldering and Ignition model*

210 We parameterized the PSI model to assess peat smouldering propagation potential by  
211 examining the ratio of the energy released by an overlying layer of peat ( $H_{comb}$ ) to the  
212 energy required to combust the layer of peat below ( $H_{ign}$ ).  $H_{comb}/H_{ign}$  ratios  $< 1$  have little  
213 potential to smoulder because there is not enough available energy from the combustion  
214 of the overlying layer to ignite the lower layer. The greater the  $H_{comb}/H_{ign}$  ratio the greater  
215 the potential for downward smouldering to progress. The PSI model does not attempt to  
216 model precise depths of burn, but has proven to be a useful approach to evaluating  
217 peatland vulnerability at the landscape scale (e.g., Lukenbach *et al.*, 2015).

218

219 *Statistical methods*

220 All statistical analyses were done using R (R Core Team, 2017). Linear model equations  
221 in text are presented in Wilkinson notation. To test the significance of site, location (i.e.  
222 middle, margin), and depth on measured  $\rho_b$  (i.e. samples taken to a maximum depth of  
223 0.52 m), we used a linear mixed-effects model (LMM) (R-package *lme4*). Location and  
224 depth were treated as fixed factors, and site as random. To test location as a factor, a  
225 dummy variable was created where margin=0, and middle=1. Overall model significance  
226 was assessed using analysis of variance (ANOVA) (R-function *anova*). Post-hoc tests  
227 were done using the *lsmeans* function (R-package *lsmeans*), based on Tukey-adjusted  
228 comparisons. A similar general linear model (GLM) approach was used for the simulated  
229 peat water content and  $H_{comb}/H_{ign}$  ratios, where WT scenario was included as an  
230 additional fixed factor, and site was treated as fixed as well. An ordinal logistic  
231 regression (R-package MASS: *polr*) was used to analyze the effects of site, location (i.e.

232 distance from upland), and depth on VP. Due to the need to study peatlands with  
233 extensive historic hydrogeological data, only one peatland was studied in each  
234 topographic position. We therefore interpret our statistical analysis with caution due to  
235 the clear pseudo-replication. Specifically, we look for differences in site (*i.e.* FT, EP, P)  
236 rather than topographic position.

## 237 **RESULTS**

### 238 *Peatland microtopography and morphometry*

239 Hummocks were the dominant microform at FT, while hollow microforms dominated  
240 both EP and P (Table 1). The margin width ranged from 2 to 10 m (Table 1), where FT  
241 had the narrowest margin and EP had the widest. The EP site had a slightly higher  
242 perimeter-to-area ratio of 0.034 m m<sup>-2</sup> compared to P and FT, with values of 0.028 and  
243 0.025 m m<sup>-2</sup>, respectively. Due to both a higher perimeter-to-area ratio and wide margin,  
244 EP had the greatest area classified as margin at 34%, compared 17% and 6% for FT and  
245 P, respectively (Table 1).

246

### 247 *Peat properties*

248 Both depth (Chi<sup>2</sup>=314.0; p<<0.01) and location (Chi<sup>2</sup>=38.2; p<<0.01) were found to be  
249 significant factors for explaining variance in  $\rho_b$  (H<sub>1</sub>:  $\rho_b = \text{Location} + \text{Depth} + (1|\text{Site})$ ),  
250 when compared to the null model using just site as a random factor (H<sub>0</sub>:  $\rho_b = 1 + (1|\text{Site})$ ).  
251 An ANOVA showed that nesting depth in location (H<sub>2</sub>: Location/Depth + (1|Site)) did  
252 not significantly improve the model of  $\rho_b$  (Chi<sup>2</sup>=0.3554; p=0.551) compared to H<sub>1</sub>,  
253 suggesting that the rate of change in  $\rho_b$  with depth is similar between middles and  
254 margins. The resulting linear model (H<sub>1</sub>) is

255  $\rho_b = 68 - 51 \cdot location + 356 \cdot depth$

256 where  $\rho_b$  is in  $\text{kg m}^{-3}$ , and *depth* is in m (the random site intercepts are omitted). The post-  
257 hoc test (*lsmeans*) showed that measured  $\rho_b$  (Table 2) was significantly different between  
258 the middle and margin (z-ratio=11.8;  $p \ll 0.01$ ) with a marginal mean difference of 51  $\text{kg}$   
259  $\text{m}^{-3}$  (margin>middle). Similarly, the LMM shows that there was a relatively large  
260 increase in  $\rho_b$  with depth, at  $356 \text{ kg m}^{-3} \text{ m}^{-1}$  based on measurements from the top 0.5 m of  
261 peat. Site differences in  $\rho_b$  account for 22% of overall variance, where all pairwise post-  
262 hoc tests show that  $\rho_b$  is significantly different between sites at a 0.05 significance level  
263 where  $FT < P < EP$ .

264

265 The humification profiles for FT (Figure 2a and 3) show that 98% of the first 0.4 m depth  
266 of the transect ranges from undecomposed (VP = 1) to slightly decomposed (VP = 4),  
267 which corresponds to an average  $\rho_b$  range of 26 to  $112 \text{ kg m}^{-3}$ . At EP and P, 44% and  
268 66% of the top 0.4 m were at or below a VP of 4, respectively. VP was modelled using an  
269 ordinal logistic regression, where the average classification accuracy was 70% based k-  
270 fold cross validation. Regression results show that site, depth ( $p \ll 0.01$ ), and the  
271 interaction of depth and distance along the transect ( $p=0.007$ ) had a significant effect on  
272 VP. While the odds ratio shows only a small likelihood of increasing VP with depth  
273 (1.03), this is on a per centimeter basis and thus becomes highly likely over the depth of a  
274 given peat profile. All else being equal, there is a significant likelihood of VP being  
275 lower at FT, and higher at EP compared to P, based on their respective odds ratios  
276 (FT=0.33; EP=1.86) (i.e  $FT < P < EP$ ). Finally, while VP tends to increase with depth, the  
277 interaction term suggests that for a given depth, there is a small likelihood of decreasing

278 VP (odds ratio = 0.98) as you move from the peatland edge to interior. Again, it should  
279 be noted that the reported likelihood is based on a one meter change in lateral position.

280

### 281 *Simulated peat water content profiles*

282 Simulated volumetric water content (VWC) shows that VWC increases rapidly with  
283 depth when the WT is near the surface (e.g. FT), and much slower when WT is deep (e.g.  
284 EP) (Table 1 and Fig. 4a-c). A global analysis of the effect of depth, site, WT scenario,  
285 location (middle/margin) (Table 3) show that all main factors have significant effects on  
286 simulated VWC. Overall, site and depth have the largest effects on VWC, but several  
287 significant two- and three-way interactions exist (Table 3). GWC, which is VWC  
288 normalized by  $\rho_b$ , shows less consistent depth dependent patterns compared to VWC.  
289 Because there is a relatively large increase in  $\rho_b$  with depth (Table 2), GWC tends to  
290 decrease with depth when WT is deep (e.g. EP). The margin locations under the dry  
291 scenario at EP (median GWC of 221%) and P (median GWC of 235%) exhibited the  
292 lowest simulated GWC profiles, ranging from  $350 \pm 91\%$  and  $293 \pm 82\%$  (EP and P,  
293 respectively) at the surface to  $166 \pm 78\%$  and  $149 \pm 62\%$  at depth (Figure 4). EP showed  
294 significantly drier simulated GWC on a site-basis ( $z\text{-ratio} \leq -7.11$ ,  $p < 0.0001$ ), except  
295 compared to the margin at P ( $z\text{-ratio} \geq -1.37$ ,  $p \geq 0.75$ ). Site P was similar to the  
296 intermediate site, EP, at the margin location, but more similar to FT at the middle  
297 location.

298

### 299 *Peat Smouldering and Ignition model*

300 Broadly, simulated  $H_{comb}/H_{ign}$  ratios tended to be low at FT, high at EP, and more  
301 location-dependent (middle v. margin) at P. With a median value of  $2.2 \pm 0.8$ , EP (dry,

302 margin) showed the highest  $H_{comb}/H_{ign}$  ratios, ranging from  $1.1\pm 0.3$  at the surface to  
303  $2.8\pm 0.6$  at depth (Figure 5). Conversely, FT (wet, middle) showed the lowest  $H_{comb}/H_{ign}$   
304 ratios, with a median value of  $0.27\pm 0.4$ , ranging from  $0.7\pm 0.8$  at the surface to  $0.26\pm 0.02$   
305 at depth (Figure 4). A global analysis of the effect of depth, site, WT scenario, location,  
306 and their interactions (Table 3) show that all main factors have significant effects on  
307 simulated  $H_{comb}/H_{ign}$  ratios. There are several significant two- and three-way interactions.  
308 Focusing on the categorical variables, Fig. 6 shows that the only strong two-way  
309 interaction is between site and location. This is due to P, where  $H_{comb}/H_{ign}$  is high in the  
310 margin and low in the middle which contrasts with EP where  $H_{comb}/H_{ign}$  in the  
311 middle/margin is relatively high, while for FT  $H_{comb}/H_{ign}$  is generally low in both  
312 locations. While  $H_{comb}/H_{ign}$  is generally higher under the dry WT scenario, the interaction  
313 with site and location is similar to the wet WT scenario (Fig. 6) where the three-way  
314 interaction is not significant (Table 3).

315

## 316 **DISCUSSION**

317 Previous literature (*e.g.*, Benschoter *et al.* 2011, Lukenbach *et al.*, 2015) has approached  
318 peatland vulnerability from a peat properties perspective, focusing on profile-scale  
319 controls on peat-smouldering dynamics, such as GWC. Although a prior study  
320 (Hokanson *et al.*, 2016) observed differences in burn severity between landscape  
321 positions and peatland physiognomy (*i.e.*, percent hollow, percent margin, GWC), no  
322 prior studies have compared entire peatlands and evaluated them for overall vulnerability  
323 to intense peat smouldering. Our holistic approach, evaluating peatland vulnerability  
324 using microform coverage, margin morphometry, and GWC distribution, shows that in a  
325 coarse-textured hydrogeological landscape, peatlands at intermediate positions (EP) are

326 most susceptible to deep smouldering during a wildfire. While it was hypothesized that  
327 the perched peatland (P) would be most vulnerable due to its complete isolation from  
328 larger groundwater flow systems, it was actually shown that it was less vulnerable than a  
329 peatland with intermittent groundwater connection. The peatland that intersected a large  
330 groundwater flow system (FT) was, by far, the least vulnerable. The presence of large-  
331 scale groundwater flow at a low-lying peatland (FT) fostered higher percent hummock  
332 coverage, relatively small margin area, and high GWC under all WT scenarios, thereby  
333 limiting its vulnerability to smouldering. In contrast, a peatland perched above the  
334 regional WT receiving only atmospheric inputs, and a peatland ephemerally connected to  
335 larger scale groundwater flow exhibited comparatively lower hummock cover, higher  
336 relative margin area, and lower GWC values.

337 The prediction of relatively low GWC profiles and high  $H_{comb}/H_{ign}$  ratios strongly suggest  
338 that the peatland in the intermediate topographic position (EP) is the most vulnerable to  
339 deep smouldering. It has the highest incidence of predicted  $H_{comb}/H_{ign}$  ratios exceeding  
340 1.0. While the margin at P has comparable  $H_{comb}/H_{ign}$  ratios under the dry scenario at  
341 some depths, it generally exhibited lower  $H_{comb}/H_{ign}$  ratios than the intermediate site, EP.

342

#### 343 *Peatland morphometry and physical properties*

344 Site had a clear influence on microtopographic distributions and peatland margin cover,  
345 where a broader survey of peatlands across topographic position would be needed to  
346 determine whether spatio-temporal patterns of groundwater connection have a strong  
347 influence on peatland microtopography in coarse-textured HRAs. Nevertheless, we  
348 propose that FT had the highest hummock coverage (60%), likely due to the stable and  
349 high WT, while EP, the intermediate site, and P, the most isolated site, showed lower



350 hummock coverage (40% and 45%, respectively). Given that previous studies have  
351 shown that hummock microforms are resistant to burning during a wildfire, whereas  
352 hollow microforms are more prone to deep burning (Benscoter *et al.*, 2015; Lukenbach *et*  
353 *al.*, 2015), FT exhibits lower vulnerability to burning.

354

355 Contrary to our initial hypothesis, increasing isolation or disconnection from larger scale  
356 groundwater flow systems did not necessarily result in wider margins and greater margin  
357 cover. While the least isolated site, FT, had the lowest margin cover relative to P and EP,  
358 P had appreciably lower relative margin cover than EP. FT had, by far, the narrowest  
359 margin (2 m) resulting in a percentage of margin coverage of the total peatland of only  
360 5.5%. Due to the strong influence of the large-scale groundwater flow system on the WT  
361 at FT, similar WT dynamics occurred at the middle and margin of the site, making the  
362 margin peat subject to similar moisture conditions as the middle. Additionally, the  
363 overarching effect of large-scale groundwater flow on the hydrology of the site appears to  
364 have minimized the distance (*i.e.*, margin width) to observe processes associated with  
365 margin development/formation, which may explain the rapid transition (*i.e.*, narrow  
366 margin) from the peatland to the mineral upland at the site. At EP and P, the magnitude  
367 of WT fluctuations was much more dramatic, corresponding with wider peatland  
368 margins, (10 and 6 meters respectively) and greater margin cover (34% and 17%  
369 respectively). While the absolute elevations of the WT do not vary significantly between  
370 the margin and the middle at EP, the WT does decline into the mineral soil below the  
371 margin (Figure 2b), leaving the margin peat hydraulically disconnected and free to  
372 decompose and densify (Waddington *et al.*, 2015). In contrast, surface and near-surface  
373 peat in the middle of the peatlands still maintain capillary connections with deeper

374 saturated peat during low WT conditions, limiting decomposition (Figure 2a-c). The WT  
375 depths at the margin of P were appreciably deeper than those in the middle of the  
376 peatland due to the perched nature of the peatland on a fine-textured lens in a coarse-  
377 textured landscape. Therefore, the WT drops precipitously, corresponding to a narrow  
378 margin compared to that of EP.

379

### 380 *Simulated peat water content and Peat Smouldering and Ignition (PSI) model*

381 At all three sites,  $\rho_b$  was shown to be systematically higher at the margins than in the  
382 middle of the peatlands. This supports the findings of previous work (Lukenbach *et al.*,  
383 2015; Hokanson *et al.*, 2016). Bulk density accounts for the majority of the differences in  
384 GWC found between and within the sites (Figure 4), which compares well with previous  
385 studies (Benscoter *et al.*, 2011).

386

387 While some studies report GWC limits on smouldering as being between 93% and 145%  
388 (*e.g.*, Rein *et al.*, 2008), others report GWC limits ranging from 250% to 295%.  
389 Benscoter *et al.* (2011) observed smouldering of peat with GWC values of 295% and  
390 Davies *et al.* (2013) reported GWC values of over 252% in unburned reference cores  
391 while smouldering was occurring nearby in the same blanket bog. Furthermore,  
392 Benscoter *et al.* (2011) observed smouldering at depth at higher GWC limits than that  
393 required for surface ignition. Primarily, EP and P had GWC values fall within the range  
394 of previously reported values for smouldering peat. When both locations (middle,  
395 margin) and all associated depths are pooled, 54% of all simulated GWC values at EP  
396 under the dry scenario were <250%, while 41% of EP depths fell below 250% under the  
397 wet scenario. At P, 33% and 28% of depths fell below a GWC of 250% for dry and wet

398 scenarios, respectively. Only 3% of depths at FT under any scenario fell below a GWC of  
399 250%, which is unlikely to sustain peat smouldering.

400

401 Expectedly,  $H_{comb}/H_{ign}$  ratios followed GWC and  $\rho_b$  trends closely at all sites, wherein  
402 low GWC values and high  $\rho_b$  values resulted in high  $H_{comb}/H_{ign}$  ratios. It is important to  
403 note that, while  $H_{comb}/H_{ign}$  ratios are a function of GWC, the results are not directly  
404 equivalent since  $H_{comb}/H_{ign}$  ratios take into account the effect of peat layering (*i.e.*,  
405 changes in  $\rho_b$  with depth).  $H_{comb}/H_{ign}$  ratios equaling 1 translates into a fuel profile whose  
406 heat of combustion exactly equals the heat required to both drive off the water and ignite  
407 the fuel in the underlying layer, assuming no heat is lost by mechanisms such as radiative  
408 or convective heat loss. Downward heat efficiencies reported by previous studies range  
409 from 0.3 to 0.9 with a mean of 0.7 (*e.g.*, Frandsen, 1998). A downward efficiency of 0.7  
410 would require an  $H_{comb}/H_{ign}$  ratio of 1.4 for successful downward combustion between  
411 layers (Figure 5). The margins and middle at EP in the dry scenario meet this requirement  
412 at a majority of depths (76%). Site P, under dry conditions, only met this condition at  
413 45% of depths, FT exhibited  $H_{comb}/H_{ign}$  ratios over 1.4 only 4% of the time.

414

415 *von Post as a tool for rapid assessment of smouldering potential*

416 Humification transects (Figure 3) show generally low levels of decomposition (*i.e.*,  
417 density) at FT and P, compared to that of EP. Using variability in peatland  
418 margin/middle VP to broadly infer  $\rho_b$  and water retention capacity, future studies could  
419 assess the landscape-scale importance of margin peat properties on vulnerability to  
420 smouldering across the BP. Using information on spatial and depth-dependence of VP in

421 BP peatlands could also be used to develop a high-level assessment of peat smouldering  
422 risk for wildfire managers.

423

424 *Assessing peatland vulnerability to wildfire using a hydrogeological landscape approach*

425 While it was originally hypothesized that as hydrologic connectivity decreased,  
426 vulnerability to smouldering would increase, we show that the completely perched (*i.e.*  
427 disconnected from regional groundwater) peatland (P) had a more moderated WT, and  
428 therefore a smaller relative margin area and lower  $\rho_b$ , than the intermediate site (EP).  
429 These peatlands are hydraulically mounded, resulting in deep WTs at the margins, which  
430 causes densification and drying of the peat (Waddington *et al.*, 2015). Site P has no  
431 connection with the regional WT, and one would expect it to be the most vulnerable to  
432 wildfire, especially in times of drought. However, the site conditions at P under  
433 maximum and minimum WT orientations are such that only a very narrow portion of the  
434 peatland is exposed during dry conditions (Figure 2). The severe WT decline at the  
435 margin is due to the sharp lithological transition of the silt and clay underlying the  
436 peatland to the sandy silt and fractured clay surrounding the peatland. While P is  
437 permanently perched well above the regional WT, intermediate sites (*e.g.*, EP) do not  
438 require such unique hydrostratigraphy, because they are transiently connected with the  
439 regional WT during wet climate cycles. This ephemeral connection could result in peat  
440 accumulation, and during dry climate conditions, result in drying and densification of  
441 margin peat as it becomes disconnected from the larger groundwater system.

442

443 **CONCLUSION**

444 We suggest that hydrogeological setting and topographic position are major controlling  
445 factors for deep smouldering hotspots in the BP. Low-lying flow-through peatlands that  
446 intersect the regional water table (FT) are the least vulnerable to deep smouldering, while  
447 peatlands in intermediate landscape positions (EP) are most vulnerable. Having *a priori*  
448 knowledge of potential smouldering hotspot locations in the landscape is beneficial for  
449 fire managers, allowing them to efficiently allocate resources and reduce emergency  
450 response time to smouldering events. While our goal was not to precisely model depths  
451 of burn, this approach is valuable for evaluating a peatland's relative vulnerability to deep  
452 smouldering and is a sound method of identifying wildfire vulnerability of peatland types  
453 within a particular HRA.

454

455 **ACKNOWLEDGEMENTS**

456 This research was funded by a Natural Sciences Engineering Research Council Discovery  
457 Grant to JMW and a research grant from Syncrude to KJD, NK, RMP and JMW, and  
458 Natural Sciences and Engineering Research Council-Collaborative Research and  
459 Development grant (NSERC-CRDPI477235-14) of Canada with industry partners  
460 Syncrude Canada Ltd and Canadian Natural Resources Limited. We thank Reyna  
461 Matties, Sarah Irvine and Cameron Irvine for assistance in the field and Carolynn Forsyth  
462 for camp facilities at Artis Inn. Lidar data were acquired by Airborne Imaging, Inc. and  
463 licensed to the Government of Alberta. The DEM was derived by Laura Chasmer by  
464 classifying ground and non-ground returns. We recognize and thank Todd Redding, Joe  
465 Riddell and countless others for their data collection efforts as part of the long-term  
466 URSA project. Finally, we thank Matt Davies and an anonymous reviewer.

467

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Table 1: Historical water-table depth (WTD) range and site characteristics at the flow through (Site FT), ephemeraally perched (Site EP), and perched (Site P) peatlands in the URSA coarse-grained hydrological response area.

Variable	FT	EP	P
<b>Peatland margin WTD (m)*</b>			
Historic range	0.32 to 0.00	0.88 to 0.45	0.75 to -0.01
2013-2014 median	-0.04	0.72	2.82
2013-2014 quartiles	(-0.11, 0.05)	(0.52, 0.85)	(2.24, 3.32)
<b>Peatland middle WTD (m)*</b>			
Historic range	0.21 to -0.02	0.92 to 0.43	0.41 to 0.00
2013-2014 median	0.04	0.74	0.15
2013-2014 quartiles	(0,0.11)	(0.5, 0.89)	(0.05, 0.25)
Peatland area (ha)	0.79	0.68	1.56
Peatland perimeter (m)	709	707	906
Average margin width (m)	2	10	6
Margin area (ha)	0.04	0.23	0.26
Margin cover (%)	5.5	34.4	16.6
Average von Post	2.2	6.6	4.8
Hummock cover (%)	64	40	45

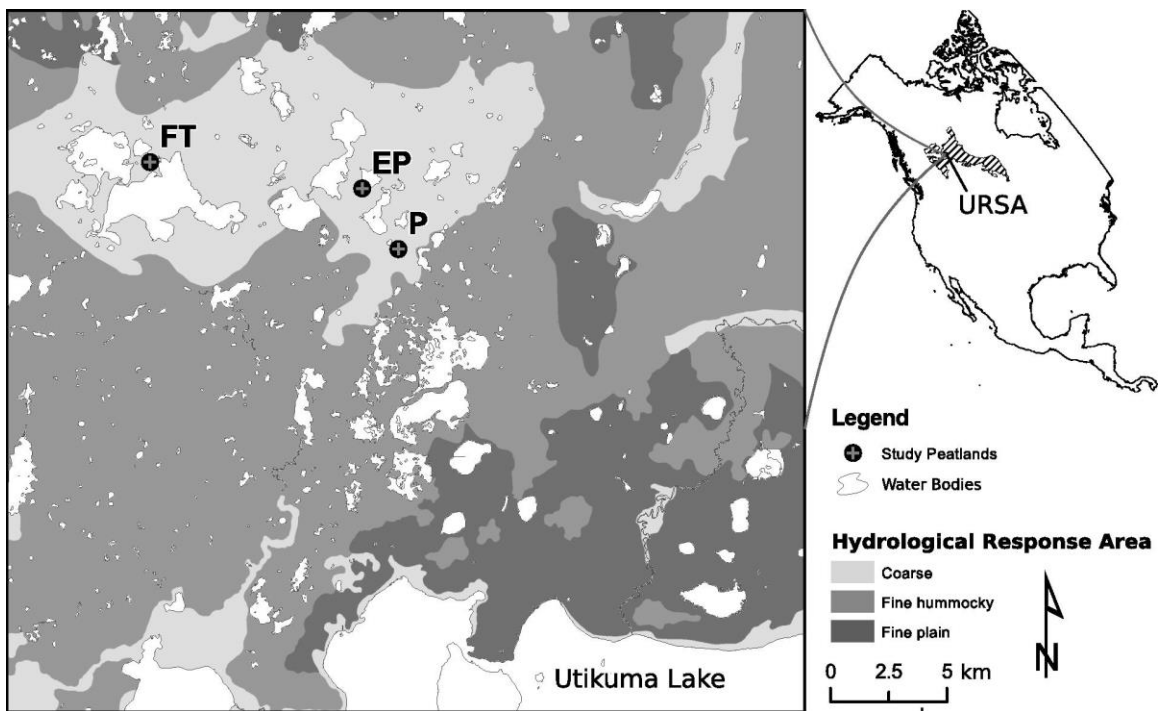
\*long term WTD data collection part of long term URSA study (Devito *et al.*, 2016)

569 Table 2: Average bulk density ( $\rho_b$ ; kg m<sup>-3</sup>) profiles, with standard error of the mean in parentheses, at the flow through (Site FT),  
 570 ephemerally perched (Site EP), and perched (Site P) peatlands in the URSA coarse-grained hydrological response area. Number of  
 571 measurements denoted by  $n$ .  
 572

Depth (m)	Site FT				Site EP				Site P			
	Margin		Middle		Margin		Middle		Margin		Middle	
	$n$	$\rho_b$	$n$	$\rho_b$	$n$	$\rho_b$	$n$	$\rho_b$	$n$	$\rho_b$	$n$	$\rho_b$
0.02	12	40 (7)	6	27 (4)	16	79 (5)	7	26 (6)	6	74 (8)	3	27 (2)
0.06	12	53 (9)	6	42 (10)	15	76 (6)	7	36 (4)	6	71 (7)	3	40 (6)
0.10	9	41 (8)	6	46 (11)	16	85 (8)	7	48 (7)	6	87 (9)	3	50 (6)
0.14	9	49 (5)	6	48 (9)	15	117 (8)	7	44 (4)	6	100 (14)	3	62 (10)
0.18	9	67 (9)	6	48 (13)	15	169 (20)	7	51 (13)	6	133 (17)	3	78 (10)
0.22	6	70 (16)	6	57 (10)	14	156 (25)	7	80 (23)	6	147 (18)	3	90 (6)
0.26	3	112 (39)	4	61 (21)	13	176 (19)	7	103 (19)	5	130 (5)	3	94 (8)
0.30	2	221 (48)	4	64 (19)	13	217 (20)	7	129 (23)	5	146 (6)	3	110 (8)
0.34	1	231 (--)	4	75 (29)	12	252 (20)	7	150 (21)	5	165 (8)	3	107 (2)
0.38	1	289 (--)	3	76 (28)	11	272 (34)	7	172 (24)	5	167 (7)	3	112 (1)
0.42			4	92 (42)	9	240 (23)	7	199 (32)	4	169 (18)	3	105 (5)
0.46					9	277 (43)	4	295 (24)	4	182 (32)	3	115 (2)
0.50					6	343 (56)	3	299 (30)	2	297 (46)	1	130 (--)
0.54					1	301 (--)						

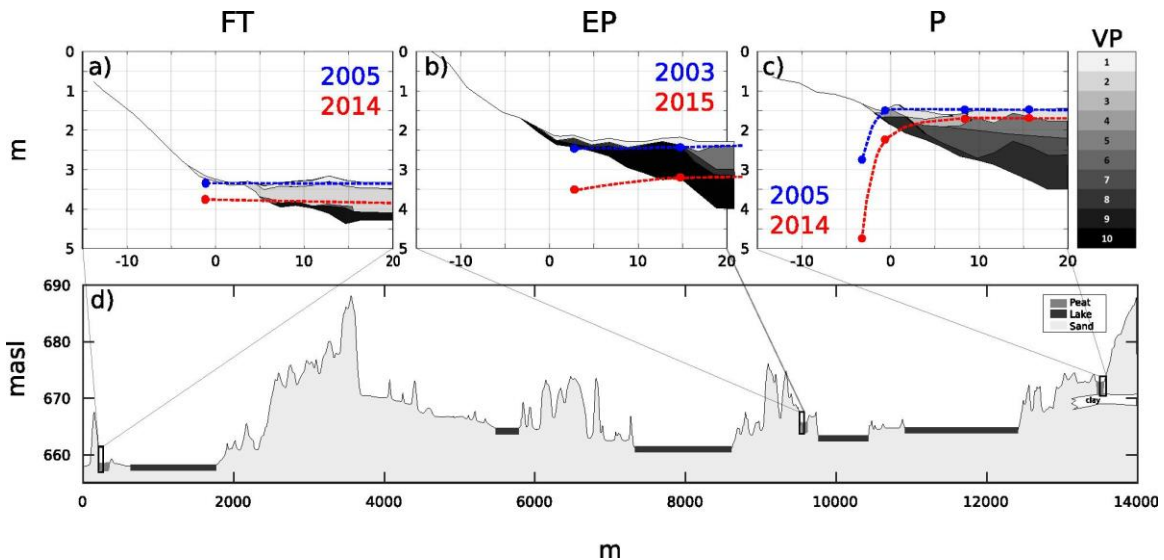
Table 3: ANOVA results for simulated volumetric water content (VWC), and  $H_{comb}/H_{ign}$  ratios

	<b>Factor</b>	<b>Sum of Square</b>	<b>d.f.</b>	<b>F-stat</b>	<b>p-value</b>
<b>VWC</b>	Depth	2.43	1	114.9	4.7E-19
	Site	4.86	2	114.7	3.3E-28
	WT Scenario	0.21	1	9.7	2.2E-03
	Location (middle v. margin)	0.21	1	9.8	2.2E-03
	Depth • Site	0.58	2	13.7	4.7E-06
	Depth • WT Scenario	0.03	1	1.4	0.25
	Depth • Location	0.36	1	17.1	6.9E-05
	Site • WT Scenario	0.07	2	1.7	0.18
	Site • Location	1.65	2	38.9	1.2E-13
	WT Scenario • Location	0.09	1	4.3	0.04
	Depth • Site • WT Scenario	0.06	2	1.5	0.23
	Depth • Site • Location	0.38	2	9.1	2.2E-04
	Depth • WT Scenario • Location	0.05	1	2.5	0.12
	Site • WT Scenario • Location	0.09	2	2.1	0.13
	Depth • Site • WT Scenario • Location	0.04	2	1.1	0.35
	<i>Factor = Error / Sum Sq = 2.46 / d.f. = 116</i>				
<b><math>H_{comb}/H_{ign}</math></b>	Depth	10.60	1	197.4	2.2E-26
	Site	29.98	2	279.2	4.7E-44
	WT Scenario	1.07	1	19.9	2.0E-05
	Location (middle v. margin)	9.81	1	182.8	3.3E-25
	Depth • Site	4.57	2	42.6	1.9E-14
	Depth • WT Scenario	0.03	1	0.51	0.47
	Depth • Location	2.58	1	48.1	2.9E-10
	Site • WT Scenario	0.15	2	1.4	0.26
	Site • Location	5.31	2	49.5	4.3E-16
	WT Scenario • Location	0.18	1	3.3	0.07
	Depth • Site • WT Scenario.	0.29	2	2.7	0.07
	Depth • Site • Location	1.64	2	15.2	1.4E-06
	Depth • WT • Location	0.10	1	1.8	0.18
	Site • WT Scenario • Location	0.10	2	0.9	0.42
	Depth • Site • WT Scenario • Location	0.16	2	1.5	0.24
	<i>Factor = Error / Sum Sq = 5.96 / d.f. = 111</i>				



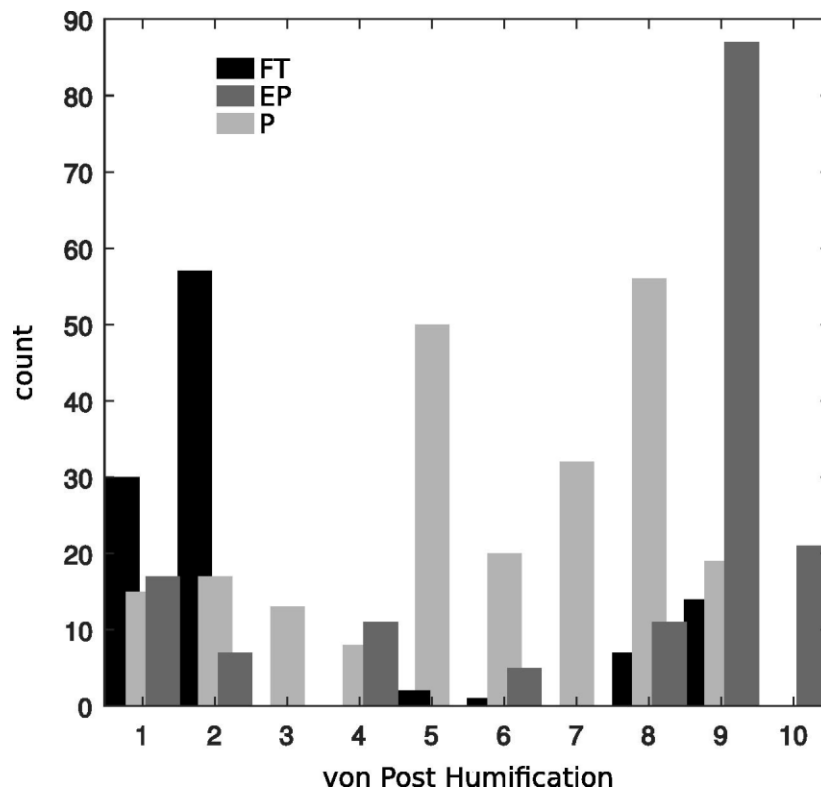
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576 Figure 1: Location of flow-through (Site FT), ephemerally perched (Site EP) and perched  
 577 (Site P) peatlands, and hydrography relative to the geology in the URSA (adapted from  
 578 Fenton et al., 2013). Inset shows the URSA's relative position in the Boreal Plains and  
 579 North America.



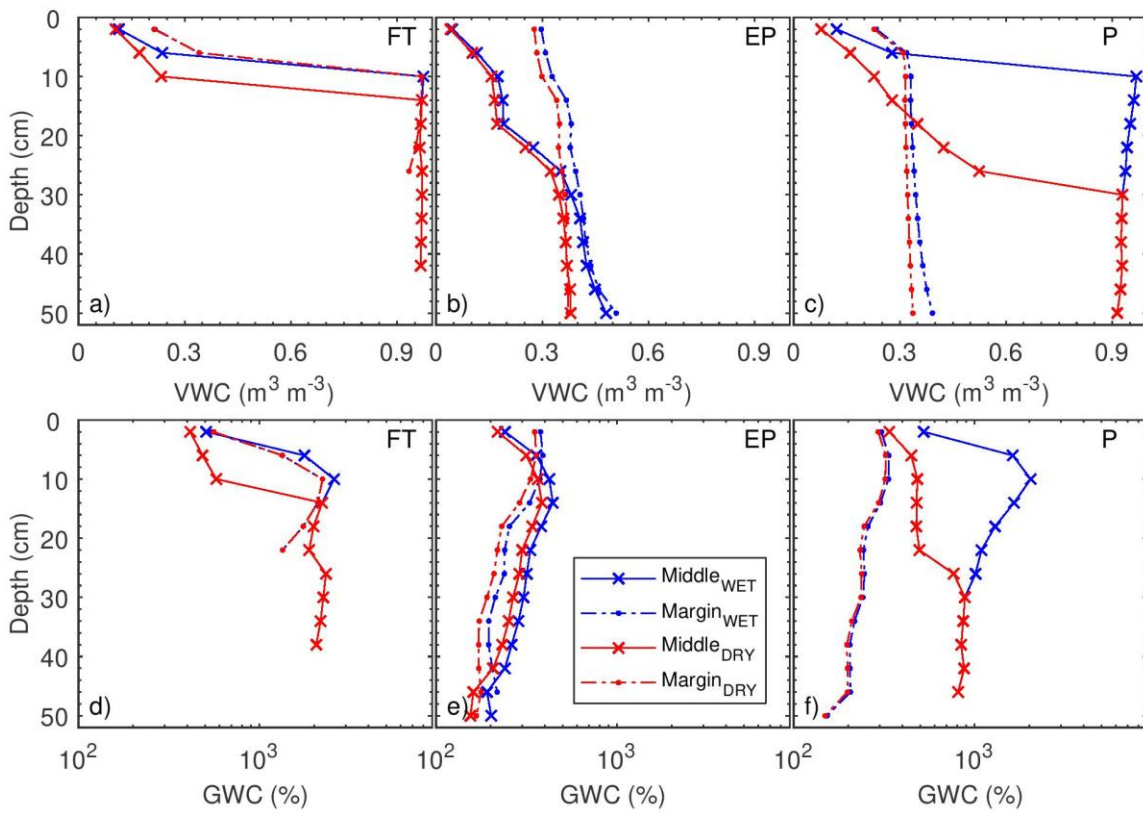
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581 Figure 2: Cross section profiles along 20 m transects at the margins of (a) Site FT, (b) Site  
 582 EP and (c) Site P. Historic high (blue) and low (red) water-table configurations are shown  
 583 for each site. von Post depth-profiles (numerical scalebar) are also shown for each site.  
 584 A cross section of the coarse textured outwash at the URSA (d) shows the relative  
 585 topographic position of each site. Vertical exaggeration is  $\sim 2.5$  times.



586

587 Figure 3: Histogram of von Post humification indices observed at site FT, EP, and P. Each  
 588 count represents a cored sample from the 20 m transect, where VP samples were taken  
 589 every 0.05m vertically and 1m horizontally.

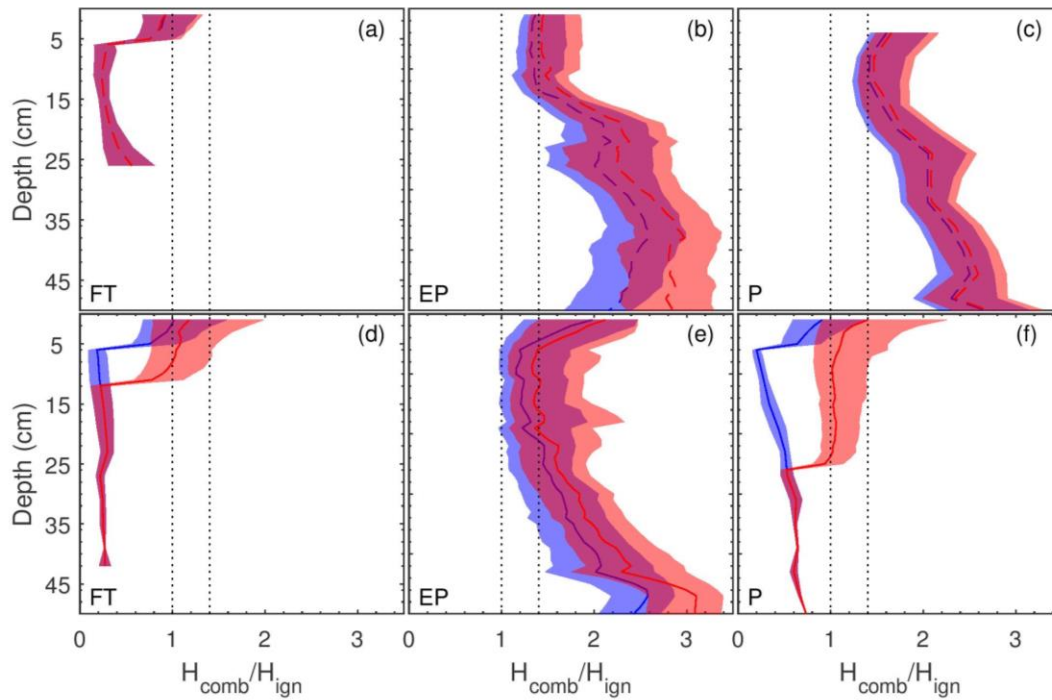


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591 Figure 4: Simulated volumetric water content (a-c) and gravimetric water content (d-f)  
 592 for middle (solid line with 'x' marker) and margin (dashed line with dot marker)  
 593 locations at three peatlands (FT, EP, and P sites) in a coarse-grained hydrological  
 594 response area

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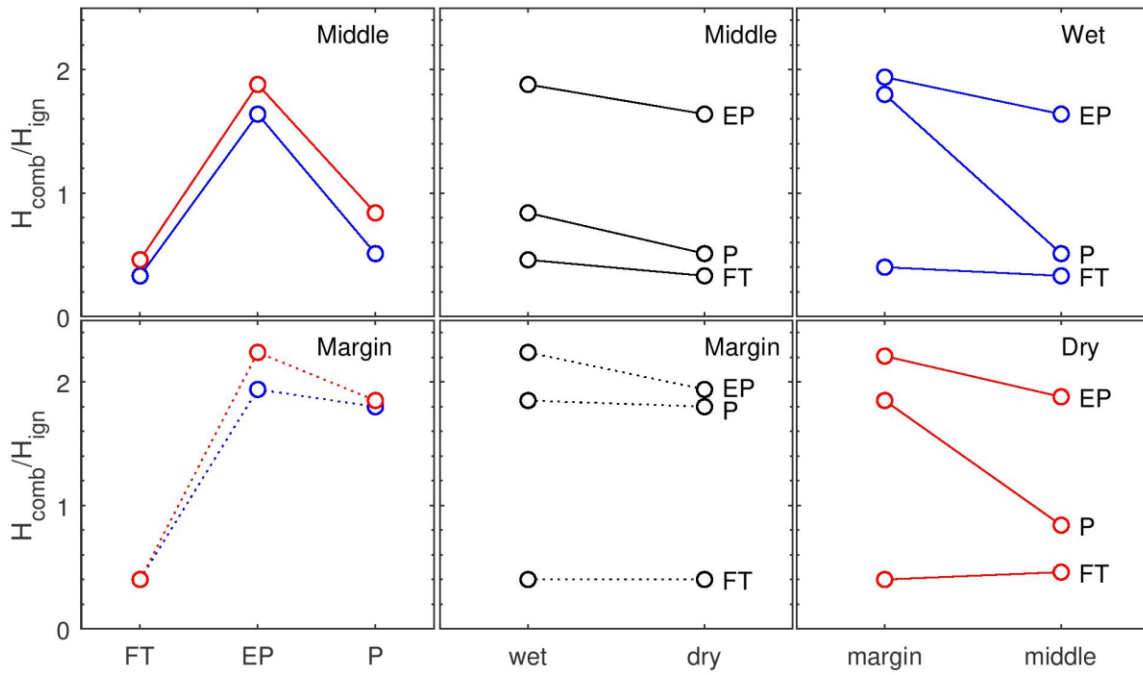




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597 Figure 5: Simulated ratio for heat of combustion over heat of ignition ( $H_{comb}/H_{ign}$ ) for  
 598 margin (a-c) and middle (d-f) locations. Red and blue lines indicate the median  
 599 simulated  $H_{comb}/H_{ign}$  ratio for dry and wet WT scenarios, respectively. Shaded areas  
 600 represent the range in simulated data from the 5th to 95th percentile. Dotted vertical  
 601 lines at 1.0 and 1.4 indicate the ratio of heat of combustion to heat of ignition required  
 602 to sustain smouldering at downward heat efficiencies of 1.0 and 0.7, respectively.

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605 Figure 6: Interaction plot for general linear model of  $H_{comb}/H_{ign}$  showing all two-way  
 606 interactions between site (FT, EP, P), WT scenarios (wet and dry) and location (margin  
 607 and middle). Comparison of top and bottom panels are meant to show three-way  
 608 interactions.

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