1	How many is enough? Determining optimal count totals for						
2	ecological and palaeoecological studies of testate amoebae.						
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17							

18 **Running title:** Optimising testate amoebae counts in palaeoecology

1 Abstract

2 Testate amoebae are increasingly used in ecological and palaeoecological studies of 3 wetlands. To characterise the amoeba community. a certain number of individuals 4 need to be counted under the microscope. To date, most studies have aimed for 150 5 individuals, but that sample size is not based on adequate evidence. When testate 6 amoeba concentrations are low, it can be difficult or impossible to reach this total. 7 The impacts of lower count totals have never been seriously scrutinised. We 8 investigated the impact of count size on number of taxa identified, quantitative 9 inferences of environmental variables and the strength of the links between amoebae 10 and environmental data in the context of predicting depth to water table. Low counts 11 were simulated by random selection of individuals from four existing datasets. 12 Results show progressively diminishing returns by all criteria as count size increases 13 from low numbers to counts of 150. A higher count is required to identify all taxa 14 than to adequately characterise the community for transfer function inference. We 15 suggest that in most cases, it will be a more efficient use of time to count a greater 16 number of samples to a lower count. While a count of 50 individuals may be 17 sufficient for some samples from some sites we recommend that counts of 100 18 individuals should be sufficient for most samples. Counts need only be increased to 19 150 or more where the aim is to identify relatively minor, but still potentially 20 ecologically relevant community changes. This approach will help reduce lack of 21 replication and low resolution, which are common limitations in testate amoeba-based 22 palaeoecological and ecological studies. 23

1 Introduction

2 Testate amoebae are a group of unicellular eukaryotic micro-organisms characterised 3 by a decay-resistant shell (Meisterfeld 2000a, b). Testate amoeba analysis has been 4 used in a variety of ecological and palaeoecological applications (Foissner 1999; 5 Mitchell et al. 2008b) and has been intensively applied in palaeoecological studies 6 from peatlands with the aim to reconstruct mire palaeohydrology and thereby, at least 7 for ombrotrophic bogs, past climatic change (Charman 2001). By analysing testate 8 amoeba community changes down the length of a core and interpreting the results 9 with a transfer function model, it is possible to quantitatively reconstruct changing 10 mire surface wetness (Booth et al. 2004; Lamentowicz et al. 2008; Sillasoo et al. 11 2007; Woodland et al. 1998) as well as sea-levels (Gehrels et al. 2002; Roe et al. 12 2002) through the Holocene. Testate amoebae are also used over shorter time periods 13 to monitor the success of peatland restoration (Buttler et al. 1996; Davis and 14 Wilkinson 2004; Jauhiainen 2002; Laggoun-Défarge et al. 2008; Vickery and 15 Charman 2004) and the impact of pollution in aquatic or terrestrial ecosystems (Balik 16 1991; Kandeler et al. 1992; Kauppila et al. 2006; Kumar and Patterson 2000; Nguyen-17 Viet et al. 2007; Nguyen-Viet et al. 2004; Patterson et al. 1996; Patterson and Kumar 18 2002; Reinhardt et al. 1998; Scott et al. 2001; Török 2001).

To adequately characterise the community composition of any microfossil used in palaeoecology, a minimum number of individuals needs to be counted under the microscope. For different biological proxies, the minimal count required ranges between 50 and over 300, largely dictated by the number of taxa typically identified in a sample. Lytie and Wahl (2005) explored the effect of pollen sample size on vegetation reconstruction and observed that with counts as low as 150 grains, the vegetation reconstruction was accurate within >90%. Other studies suggest that

1 counts of 200 (Barkeley 1934) to 250 grains (Hill 1996) are sufficient. However, the 2 recommended sum is generally between 300 and 500 grains, and in order to obtain a 3 precise estimate for rare taxa, counts of 500 to 1000 grains are preferred (Birks and 4 Birks 1980). Heiri and Lotter (2001) investigated the impact of low counts on transfer 5 function inference for Chironomids and concluded that counts as low as 50 specimens 6 provided useful results. For charcoal, the minimum count (including exotic marker 7 grains) was estimated at 200 (Finsinger and Tinner 2005). For diatoms using the 8 random settling method, a total count of 300 has been shown to yield reasonable 9 estimates for the most abundant taxa, and increasing the count to 400 did not greatly 10 improve the accuracy (Boden 1991), although the reliability of this approach has been 11 criticized (Meng 1994). Totals of 500-1000 valves are frequently used (Lowe and 12 Walker 1997). In the case of foraminifera, the required minimum is 300-400 13 individuals (Lowe and Walker 1997). 14 For testate amoeba analysis, studies have usually aimed to count either 150 or, less commonly, 100, 200, or 300 individuals (Bobrov 2005; Charman et al. 2000; 15

16 Tolonen 1986; Warner 1990). Whether such counts are actually sufficient has not

17 been rigorously tested. Mitchell et al. (2000) looked at the relationship between

18 testate amoeba count and number of taxa and found a plateau was reached after

19 around 100 individuals. Warner (1990) recorded the number of taxa with increasing

20 count for 8 samples and found a plateau was reached after around 60 individuals.

Woodland et al. (1998) showed that cumulative species diversity reached a plateau
after around 100 individuals.

23 Several studies have found it impossible, or at least impractical, to reach the 24 usual total. Low test concentrations have been encountered in situations such as 25 palaeoecological sequences with highly humified peats (Beyens and Chardez 1987:

Charman et al. 2001) and surface and sub-surface samples of fen peats (Jauhiainen
2002; Payne and Pates 2008). There are probably three factors contributing to these
problems: 1) low abundance of amoebae in the sediments on deposition, 2)
decomposition of tests over time and 3) a reduction in apparent test concentration due
to a high abundance of organic or inorganic material in the same size range as the
tests, which is not removed in sample preparation.

7 Low test concentrations present particular problems in palaeoecological 8 studies where an adequate count may not be achievable for parts of the sequence. This leads to the question of whether some data is better than none: is it better to have 9 10 portions of a sequence with very low counts or to leave a gap in the palaeoecological 11 record? Some palaeoecological studies have presented records with very low counts 12 and have made palaeoenvironmental inferences based on these results, albeit with 13 caveats clearly stated (e.g. Beyens and Chardez 1987). However no study has been 14 undertaken to assess the relationship between number of testate amoeba shells 15 counted and data quality.

This study attempts to determine the minimum count levels required for ecological and palaeoecological studies of testate amoebae by examining the impact of count size on the number of taxa identified in three contrasting sites, by simulating low counts using a random-selection approach with four previously established modern training sets, and then applying the same approach to two palaeoecological data sets.

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1 Methods

2 General approach and aims

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4 To determine the test count required to characterise a sample, we use four criteria: 1) 5 The number of taxa. As count size increases, the number of taxa identified should 6 increase and then stabilise as the total complement is reached. 2) Transfer function 7 inferences of environmental variables. As count increases, transfer function inferred 8 values and associated standard errors should stabilise. 3) The strength of the 9 relationship between the amoeba community and environmental variables. As the 10 amoeba community becomes better characterised with increasing count size, the 11 strength of the relationship with key environmental controls should improve. 4) The 12 inferred pattern of wet and dry periods in palaeoecological records: As count size increases, the pattern should become more and more similar to the full model. 13 14 15 Impact of count on number of taxa identified 16 17 Previous studies have looked at the impact of count size on the number of taxa 18 identified (Warner 1990; Woodland et al 1998). However these studies only 19 considered a limited number of samples from a limited range of sites, and it is useful 20 to examine this with additional data. We present new data from the Ispani-2 peatland, 21 a percolation bog in western Georgia (Connor et al. 2007; Joosten et al. 22 2003)(10samples), the Sterling peatland, a kettle-hole mire in southern Alaska (Payne 23 et al. 2006)(10 samples) and Moidach More a blanket bog in eastern Scotland (150 24 samples). The sites represent a variety of peatland types and cover a large geographic 25 range. Samples were extracted from the surface of the Ispani-2 and Moidach More

1	sites and just below the surface (10-20cm) in the Sterling site. The normal total of 150
2	tests was counted for each sample and the number of tests required to identify each
3	additional taxon recorded.

Impact of count on species-environment relationships and predicted water table depth

7 To assess the impact of count size on amoeba community recorded, low-counts were 8 simulated using four data sets from previous studies of peatland testate amoebae: 1) 9 the data of Payne & Mitchell (2007) from the Elatia mires in northern Greece, 2) the 10 data of Payne et al. (2008) from the Sürmene Ağaçbaşı Yaylası peatland in north-11 eastern Turkey, 3) the data of Mitchell et al. (1999) from peatlands of the Jura 12 Mountains (France and Switzerland) and 4) the data of Payne et al. (2006) from 13 peatlands in southern Alaska (Table 1). Samples in these studies were counted to at 14 least the usual total of 150 and recorded as percentages. Species complement varied 15 from 4 to 27 taxa. Much of the count is represented by a few taxa, and a single taxon 16 typically accounts for at least 25% of the total. Counts of 10, 20, 30..140 were 17 simulated by randomly selecting individuals from the full data set, and each 18 individual was able to be selected more than once. The transfer function models 19 calculated in the original studies were used to infer depth to water table based on 20 these simulated data sets. Standard errors were calculated using bootstrapping (1000 21 cycles). The procedure was repeated nine additional times with random re-selection of 22 individuals on each occasion, and the bias from the predicted values of the original 23 model were calculated.

These experiments assume that a count of 150 tests is sufficient to characterise the amoeba community, To test the impact of higher totals on transfer function

1	performance, a sub-set of samples was separately analysed. For some samples from
2	the Alaska and Jura studies, amoeba concentrations were high and over 250 tests were
3	counted. Counts of 150, 160,, 240 were simulated for these samples using the same
4	repeated random selection approach as above. Transfer function inferences using the
5	simulated data sets were compared to those of the full model for these samples.
6	Ordination was used to test the impact of count on the strength of relationship
7	between species and environmental data. The percent variance explained by the depth
8	to water table (DWT) data was tested using redundancy analysis (RDA) with each of
9	the simulated low count data-sets in CANOCO ver.4.53 (ter Braak and Šmilauer
10	1997-2004). As these analyses are labour-intensive only one cycle of data-selection
11	was used for this experiment.
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13	Impact of count on palaeoecological interpretation
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13 14 15	Impact of count on palaeoecological interpretation Finally, the same random data-selection approach was used to simulate low counts for
13 14 15 16	Impact of count on palaeoecological interpretation Finally, the same random data-selection approach was used to simulate low counts for two palaeoecological data sets from the Praz-Rodet peatland in Switzerland (Mitchell
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 13 14 15 16 17 18 19 20 21 22 	Impact of count on palaeoecological interpretation Finally, the same random data-selection approach was used to simulate low counts for two palaeoecological data sets from the Praz-Rodet peatland in Switzerland (Mitchell et al. 2001) and a small sub-Arctic mire in eastern Alaska, USA (termed site DLB; Payne & Blackford, unpublished data). Water table changes were inferred for each of the simulated data sets using the transfer functions of Mitchell et al. (1999, 2001) and Payne et al. (2006). Although low counts were simulated for the palaeoecological data, the full data sets were used for the transfer functions. The impact of lower palaeoecological counts was evaluated by comparing the pattern of inferred low and
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reselection, and standard deviations were calculated to assess the repeatability of
 predictions.

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- 4

5 **Results**

6 Impact of count size on number of taxa identified

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8 Fig. 1 shows average number of taxa encountered with various count sizes. With the 9 Ispani-2 and Sterling samples, plateaux in the number of taxa are reached with counts 10 of around 100 and 120 individuals, respectively. With the Moidach More samples 11 there is a continued but slowing increase in taxa complement through to 150 12 individuals. In 55 of these 150 samples, new taxa are still encountered with counts over 130. It is probable that 150 individuals is insufficient to identify all taxa in these 13 14 samples. There is no simple relationship between the total number of taxa in a sample 15 and the count required to identify all taxa. The Sterling samples are more diverse than 16 the Moidach More samples, but a count of 130 individuals would identify all taxa in 17 the Sterling samples and miss taxa in more than a third of the Moidach More samples. 18 The count required to identify all taxa depends not just on the number of taxa but the 19 relative abundances of those taxa.

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Impact of count size on species-environment relationships and predicted water tabledepth

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Fig. 2 shows box plots of transfer-function inferred water table depth with various
count totals for four representative samples from the Sürmene Ağaçbaşi Yaylasi site

1 using ten cycles of random selection of individuals. Increasing the count generally 2 reduces the scatter of predictions and brings them closer to the full-model predicted 3 result, but not necessarily the measured value. Clearer results are obtained when using 4 all samples and considering mean bias from the full-model predicted values (Fig. 3). With all data sets, increasing count gives predictions that are closer to the full model. 5 6 This improvement is most rapid in the range of 10-50 individuals. Counts over 50 and 7 up to 100 produce a slight further improvement, this is most noticeable with the 8 outliers in the Turkey data. Increasing the count from 100 to 140 appears to make 9 little further difference. Fig 4 shows the results if this is extended to counts of up to 10 240 using just those samples from the Jura and Alaska data sets with counts \geq 250. As 11 with the full data, the greatest difference is obtained by increasing the count to around 12 50 with less pronounced improvement beyond that. It is notable that increasing the 13 count beyond 150 seems to make a continued slight improvement in performance in 14 both data sets. It should, however, be noted that the data sets used to obtain this result 15 are small, with 12 and 18 samples for the Alaska and Jura data, respectively. 16 Increasing count reduces the standard error of predictions (Fig. 5). The count 17 required before values stabilise varies between data sets from 40 with the Alaska data 18 to perhaps 100 with the Turkey data. Increasing the count also increases the percent 19 variance explained in the redundancy analyses (Fig. 6). As these results are based on a 20 single cycle of random selection, there is more apparent 'noise' than with the results 21 discussed above. Values stabilise after counts of 100 in all sites. The increase is most 22 rapid with the Turkey data and least rapid with the Greece data. 23

24 Impact of count size on palaeoecological inferences

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1 The palaeoecological reconstructions from Praz-Rodet and site DLB are shown in 2 Fig. 7a and Fig. 7b, respectively. With low counts, there is a great deal of high-3 amplitude variability. Increasing the counts generally gives a record with less noise. 4 Some features of the record are only resolved with a higher count such as the double peak at the base of the Praz-Rodet record, and the trough at 21-22 cm in the DLB 5 6 record. It is clearly the case that increasing the count size gives a less noisy TI-DWT 7 record, which agrees better with the results of the full model. However, it is perhaps 8 surprising that counting very few tests still shows many of the major trends. Even 9 with counts as low as 10 tests, the DLB record picks out the three distinct wet phases 10 at 26-29 cm, 53-56 cm and 61-71 cm. In the Praz-Rodet record, counts of ten tests 11 still show the distinctive peaks at 126, 166 and 312 cm, and troughs at 156 and 226 12 cm. Increasing count produces a slight decrease in boot-strapped errors, but this is 13 minor and makes little difference to the interpretation of the results. Inferences based 14 on a count of 10 individuals probably would not be very different from those based on 15 the full count. The correlation between the reduced and full data further shows that 16 counts as low as 10 already lead to a correlation coefficient of 0.6-0.75, and a 17 correlation of 0.9 is reached with counts of 50, with little further subsequent 18 improvement (Fig. 8). A plot of the mean standard deviation of the predictions shows 19 high standard deviation for lower counts, representing distinct differences between 20 individual cycles of data selection. As count size increases, the repeatability of 21 predictions improves. The improvement produced with higher counts declines after 22 around 40-50 individuals, but still continues to make a difference through to 140 23 individuals (Fig. 9).

24

25 **Discussion**

This is the first in-depth attempt to determine the minimum count sizes required for ecological and palaeoecological studies of testate amoebae. We addressed the impact of count size on species richness, community composition and the ecological and palaeoecological information that can be obtained from the community structure data. This study, therefore, has direct relevance for a range of ecological and palaeoecological applications of testate amoebae (Charman 2001; Mitchell et al. 2008a).

8 Increasing count total increases the number of taxa identified, improves the 9 precision of transfer function prediction and reduces the standard error of those 10 predictions. As count size increased, there were progressively diminishing returns to 11 further increasing the number of tests counted. The point at which counting more tests 12 fails to produce any further improvement varies depending on what criterion is used. 13 For standard errors, there appears to be little gain from counting more than 60 tests. 14 To characterise a sample for redundancy analysis, 100 individuals seems to be 15 enough. A total of 100 tests also appears to be enough to achieve precision in transfer 16 function predictions, although slight further benefit may be obtained by counting 17 more individuals. To identify all the taxa in a sample, a count of at least 100 is 18 generally required and counts of 150 or more may be needed in samples that are 19 species-rich, or which include very rare taxa.

A frequent limitation of testate-amoeba based studies is low sampling resolution and the lack of true replication. The sampling intensity of many ecological studies may be insufficient to characterise the full range of testate amoeba communities and environmental variables under consideration. In transfer function studies this may lead to 'no-analogue' problems when the data are used for environmental reconstruction. In most palaeoecological studies, only one core is

analysed, but it is possible that this core may not be truly representative of the site.
 Studies that examined several cores show that differences in peat accumulation
 patterns can be relatively important (Charman 2007). The sampling resolution of
 cores is also frequently inadequate. It is common that some palaeoecological "events"
 are documented by very few samples, and indeed often only a single sample.

6 Given these limitations we suggest that, when counting time is limited, it is 7 likely to be more useful to count a greater number of samples to a lower count total 8 than to count fewer samples to the usual total of 150. While a count of 50 tests may be 9 sufficient for some samples from some sites, we advocate a count of 100 tests based 10 on our analyses that examined depth to water table. This total should yield most of the 11 ecological information provided by counts of 150 tests, but for a significantly reduced 12 expenditure in counting time. The time taken to count a sample is only one 13 component of the total time invested in data generation. Additional time is required 14 for fieldwork, sub-sampling and preparation. However, in most cases the greatest 15 proportion of time is spent counting, and time savings in this would allow 16 considerably more samples to be analysed. Higher counts will still be required in 17 some situations where it is important to identify very fine environmental changes, or 18 if the focus of the study is on diversity. It is likely that even counts of 150 individuals 19 will fail to identify some taxa in some samples. Higher count totals may have more 20 value in transfer function studies that will be used for multiple inferences than for 21 individual palaeoecological studies. For most studies the small amount of information 22 lost in using a lower count will not be important, and the gain in number of samples 23 will be a worthwhile trade-off.

In some studies low test concentrations may make it impossible to reach a count of 150, or even 50 tests in a reasonable time frame, particularly if the amount of

material is limited. Lower counts will inevitably reduce the accuracy and precision of transfer function inferences. However ,our palaeoecological results suggest that a meaningful palaeoecological signal may still predominate over random noise. Where counts are very low, results must be treated with considerable caution, but it is still possible for major changes to be shown.

6 Our findings are largely based on four modern training sets. Although these 7 training sets cover a large geographic range and considerable differences in amoebae 8 community structure, it is possible that our conclusions are not applicable to all such 9 samples. Our findings may also not be applicable to transfer functions that aim to 10 reconstruct different environmental variables, and particularly to testate amoebae 11 communities in other habitats that may have quite different diversity. The majority of 12 our experiments also assume that a count of 150 tests is in itself sufficient. If this is 13 not the case, then our conclusions may be erroneous.

14 Testate amoeba analysis and specifically its application for palaeohydrology, 15 is a relatively young technique compared to more established palaeoecological 16 methods such as palynology and diatom analysis. As the method becomes more 17 routinely applied it is necessary to address many of its underlying assumptions. Count 18 totals are one of these assumptions along with issues such as taphonomy, taxonomy 19 and the precise environmental controls on amoeba communities (Barber and Langdon 20 2007; Charman 2007; Charman et al. 2004; Mitchell et al. 2008a; Mitchell et al. 21 2008b; Payne 2007; Payne and Mitchell 2007; Swindles and Roe 2007). Testate 22 amoeba analysis is increasingly used in palaeoecology and ecology. Our results show 23 that most of the valuable data that lies in the structure of fossil testate amoeba 24 communities can be obtained at a reduced cost.

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9 **References**

10 Balik V (1991) The effect of the road traffic pollution on the communities of testate

11 amoebae (Rhizopoda, Testacea) in Warsaw (Poland). Acta Protozool 30: 5-11

12 Barber KE, Langdon PG (2007) What drives the peat-based palaeoclimate record? A

13 critical test using multi-proxy climate records from northern Britain. Quat Sci Rev 26:

14 3318-3327

15 Barkeley F (1934) The statistical theory of pollen analysis. 15: 283-289

16 Beyens L and Chardez D (1987) Evidence from testate amoebae for changes in some

- 17 local hydrological conditions between c. 5000 BP and c. 3800 BP on Edgeøya
- 18 (Svalbard). Polar Res 5: 165-169

19 Birks HJB and Birks HH (1980) Quaternary Palaeoecology. Edward Arnold, London

20 Bobrov AA (2005) Testate amoebas and the regularities of their distribution in soil.

- 21 Eurasian Soil Sci 38: 1001
- 22 Boden P (1991) Reproducibility in the Random Settling Method for Quantitative
- 23 Diatom Analysis. Micropaleontology 37: 313-319
- 24 Booth RK, Jackson ST, Gray CED (2004) Paleoecology and high-resolution
- 25 paleohydrology of a kettle peatland in upper Michigan. Quat Res 61: 1–13

1	Buttler A,	Warner BG,	Grosvernier	P, Matthey	Y (19	96) Vertical	patterns	of testate
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- 2 amoebae (Protozoa: Rhizopoda) and peat forming vegetation on cutover bogs in the
- 3 Jura, Switzerland. New Phytol 134: 371-382
- 4 Charman DJ (2001) Biostratigraphic and palaeoenvironmental applications of testate
- 5 amoebae. Quat Sci Rev 20: 1753-1764
- 6 Charman DJ (2007) Summer water deficit variability controls on peatland water-table
- 7 changes: implications for Holocene palaeoclimate reconstructions. Holocene 17: 2178 227
- 9 Charman DJ, Brown AD, Hendon D, Karofeld E (2004) Testing the relationship
- 10 between Holocene peatland palaeoclimate reconstructions and instrumental data at
- 11 two European sites. Quaternary Sci Rev 23: 137–143
- 12 Charman DJ, Caseldine C, Baker A, Gearey B, Hatton J, Proctor C (2001)
- 13 Paleohydrological records from peat profiles and speleothems in Sutherland,
- 14 northwest Scotland. Quat Res 55: 223-234
- 15 Charman DJ, Hendon D, Woodland WA (2000) The identification of testate amoebae
- 16 (Protozoa: Rhizopoda) in peats. Quaternary Research Association, London, 147 pp
- 17 Connor SE, Thomas I, Kvavadze EV (2007) A 5600-yr history of changing
- 18 vegetation, sea levels and human impacts from the Black Sea coast of Georgia. 17:
- 19 25-36
- 20 Davis SR, Wilkinson DM (2004) The conservation management value of testate
- 21 amoebae as 'restoration' indicators: speculations based on two damaged raised mires
- in northwest England. Holocene 14: 135–143
- 23 Finsinger W, Tinner W (2005) Minimum count sums for charcoal-concentration
- estimates in pollen slides: accuracy and potential errors. Holocene 15: 293-297

- 1 Foissner W (1999) Soil protozoa as bioindicators: pros and cons, methods, diversity,
- 2 representative examples. Agric Ecosyst Environ 74: 95-112
- 3 Gehrels WR, Roe HM, Charman DJ (2002) Foraminifera, testate amoebae and
- 4 diatoms as sea-level indicators in UK saltmarshes: a quantitative multiproxy approach
- 5 (vol 16, pg 201, 2001). 17: 285-285
- 6 Heiri O, Lotter AF (2001) Effect of low count sums on quantitative environmental
- 7 reconstructions: an example using subfossil chironomids. J Paleolimn 26: 343-350
- 8 Hill TR (1996) Statistical determination of sample size and contemporary pollen
- 9 counts, Natal Drakensberg, South Africa. Grana 35: 119-124
- 10 Jauhiainen S (2002) Testacean amoebae in different types of mire following drainage
- 11 and subsequent restoration. Europ J Protistol 38: 59-72
- 12 Joosten H, Kaffke A, Matchutadze I (2003) The mires of the Kolkheti lowlands
- 13 (Georgia). 2003: 19-23
- 14 Kandeler E, Luftenegger G, Schwarz S (1992) Soil microbial processes and testacea
- 15 (Protozoa) as indicators of heavy-metal pollution. Z Pflanzen Bodenk 155: 319-322
- 16 Kauppila T, Kihlman S, Makinen J (2006) Distribution of arcellaceans (testate
- 17 amoebae) in the sediments of a mine water impacted bay of Lake Retunen, Finland.
- 18 Water Air Soil Pollut 172: 337
- 19 Kumar A, Patterson RT (2000) Arcellaceans (thecamoebians): new tools for
- 20 monitoring long- and short-term changes in lake bottom acidity. Environ Geol 39:
- 21 689-697
- 22 Laggoun-Défarge F, Mitchell EAD, Gilbert D, Disnar J.-R, Comont L, Warner B,
- 23 Buttler A (2008) Cutover peatland regeneration assessment using organic matter and
- 24 microbial indicators (bacteria and testate amoebae). J Appl Ecol 45: 716-727

1	Lamentowicz M, Obremska M, Mitchell EAD (2008) Autogenic succession, land-use
2	change, and climatic influences on the Holocene development of a kettle hole mire in
3	Northern Poland. Rev Palaeobot Palynology: in press
4	Lowe JJ, Walker MJC (1997) Reconstructing Quaternary Environments. Prentice
5	Hall, 446 pp
6	Lytle DE, Wahl ER (2005) Palaeoenvironmental reconstructions using the modern
7	analogue technique: the effects of sample size and decision rules. Holocene 15: 554-
8	566
9	Meisterfeld R (2000a) Order Arcellinida Kent, 1880. In: JJ Lee, GF Leedale, P
10	Bradbury (eds.), The illustrated guide to the protozoa. Society of protozoologists,
11	Lawrence, Kansas, USA, pp. 827-860
12	Meisterfeld R (2000b) Testate amoebae with filopodia. In: JJ Lee, GF Leedale, P
13	Bradbury (eds.), The illustrated guide to the protozoa. Society of protozoologists,
14	Lawrence, Kansas, USA, pp. 1054-1084
15	Meng L (1994) How Accurate Is the Random Settling Method for Quantitative
16	Diatom Analysis - a Test Using Lycopodium Spore Tablets. Micropaleontology 40:
17	261-266
18	Mitchell EAD, Borcard D, Buttler AJ, Grosvernier P, Gilbert D, Gobat JM (2000)
19	Horizontal distribution patterns of testate amoebae (Protozoa) in a Sphagnum
20	magellanicum carpet. Microb Ecol 39: 290-300
21	Mitchell EAD, Buttler AJ, Warner BG, Gobat JM (1999) Ecology of testate amoebae
22	(Protozoa : Rhizopoda) in Sphagnum peatlands in the Jura mountains, Switzerland
23	and France. Ecoscience 6: 565-576

1	Mitchell EAD, Charman DJ, Warner BG (2008a) Testate amoebae analysis in							
2	ecological and paleoecological studies of wetlands: past, present and future.							
3	Biodiversity and Conservation. 17: 2115-2137							
4	Mitchell EAD, Payne RJ, Lamentowicz M (2008b) Potential implications of							
5	differential preservation of testate amoebae shells for paleoenvironmental							
6	reconstruction in peatlands. J Paleolimnol 40: 603-618							
7	Mitchell EAD, van der Knaap WO, van Leeuwen JFN, Buttler A, Warner BG, Gobat							
8	JM (2001) The palaeoecological history of the Praz-Rodet bog (Swiss Jura) based on							
9	pollen, plant macrofossils and testate amoebae (Protozoa). Holocene 11: 65-80							
10	Nguyen-Viet H, Bernard N, Mitchell EAD, Cortet J, Badot PM, Gilbert D (2007)							
11	Relationship Between Testate Amoeba (Protist) Communities and Atmospheric							
12	Heavy Metals Accumulated in Barbula indica (Bryophyta) in Vietnam. Microb Ecol							
13	53: 53-65							
14	Nguyen-Viet H, Gilbert D, Bernard N, Mitchell EAD, Badot P-M (2004) Relationship							
15	between atmospheric pollution characterized by NO2 concentrations and testate							
16	amoebae abundance and diversity. Acta Protozool 43: 233-239							
17	Patterson RT, Barker T, Burbidge SM (1996) Arcellaceans (thecamoebians) as							
18	proxies of arsenic and mercury contamination in northeastern Ontario lakes. J							
19	Foraminifer Res 26: 172-183							
20	Patterson RT, Kumar A (2002) A review of current testate rhizopod (thecamoebian)							
21	research in Canada. Paleogeogr Paleoclimatol Paleoecol 180: 225-251							
~~								
22	Payne R (2007) Laboratory experiments on testate amoebae preservation in peats:							
22 23	Payne R (2007) Laboratory experiments on testate amoebae preservation in peats: implications for palaeoecology and future studies. Acta Protozool 46: 325-332							

1	Payne R, Mitchell EAD (2007) Testate amoebae-environment relationships and a
2	hydrological transfer function from mires in the Central Rhodope Mountains, Greece.
3	Protist 158: 159-171
4	Payne R, Pates J (2008) Vertical stratification of testate amoebae in the Elatia Mires,
5	northern Greece: Palaeoecological evidence for a wetland response to recent climatic
6	change, or autogenic processes? Wetlands Ecol Manage in press DOI
7	10.1007/s11273-008-9112-8
8	Payne RJ, Charman DJ, Eastwood W (2008) Testate amoebae as palaeoclimatic
9	proxies in Sürmene Ağaçbaşi Yaylasi peatland (Northeast Turkey). Wetlands 28: 311-
10	323
11	Payne RJ, Kishaba K, Blackford JJ, Mitchell EAD (2006) Ecology of testate amoebae
12	(Protista) in south-central Alaska peatlands: building transfer-function models for
13	palaeoenvironmental studies. Holocene 16: 403-414
14	Reinhardt EG, Dalby AP, Kumar A, Patterson RT (1998) Arcellaceans as pollution
15	indicators in mine tailing contaminated lakes near Cobalt, Ontario, Canada.
16	Micropaleontology 44: 131-148
17	Roe HM, Charman DJ, Gehrels WR (2002) Fossil testate amoebae in coastal deposits
18	in the UK: implications for studies of sea-level change. J Quaternary Sci 17: 411-429
19	Scott DB, Medioli FS, Schafer CT (2001) Monitoring in Coastial Environments Using
20	Foraminifera and Thecoamoebian Indicators. Cambridge Unversity Press, Cambridge
21	Sillasoo U, Mauquoy D, Blundell A, Charman D, Blaauw M, Daniell JRG, Toms P,
22	Newberry J, Chambers FM, Karofeld E (2007) Peat multi-proxy data from
23	Mannikjarve bog as indicators of late Holocene climate changes in Estonia. Boreas
24	36: 20-37

1	Swindles GT, Roe HM (2007) Examining the dissolution characteristics of testate
2	amoebae (Protozoa: Rhizopoda) in low pH conditions: Implications for peatland
3	palaeoclimate studies. Paleogeogr Paleoclimatol Paleoecol 252: 486-496
4	ter Braak CJF, Šmilauer P (1997-2004) CANOCO ver.4.53 Biometris- Plant
5	Research International, Wageningen, The Netherlands
6	Tolonen K (1986) Rhizopod analysis. In: BE Berglund (ed.), Handbook of Holocene
7	Palaeoecology and Palaeohydrology. John Wiley and Sons, Chichester, pp. 645-666
8	Török JK 2001. Diversity of testacean Protozoa in River Tsiza after a cyanide and
9	metal pollution at the beginning of year 2000: Preliminary study. Opusc Zool
10	Budapest XXXIII: 91-98
11	Vickery E, Charman DJ (2004) Biomonitoring of peatland restoration using testate
12	amoebae. In: JTA Verhoeven, E Dorland and M Coemans (eds.), 7th INTECOL
13	International Wetlands Conference, Utrecht, NL, 25-30 July 2004, p. 342
14	Warner BG (1990) Testate Amoebae (Protozoa). In: BG Warner (ed.), Methods in
15	Quaternary Ecology. Geoscience Canada, St. John's, Newfoundland, pp. 65-74
16	Woodland WA, Charman DJ, Sims PC (1998) Quantitative estimates of water tables
17	and soil moisture in Holocene peatlands from testate amoebae. Holocene 8: 261-273
18	

Tables & Figures

2

1

3 Figure 1. Impact of increasing test count on number of taxa recorded in samples from





Figure 2. Box plots showing impact of increasing simulated test count on transfer
function predicted depth to water table (DWT) for four samples spanning the
hydrological gradient in the Sürmene Ağaçbaşi Yaylasi site. Results based on ten
cycles of random re-selection of individuals. Solid horizontal line shows full model
predicted value, dashed line shows measured value. Box plots show median (central
line), first and third quartiles (grey box), tenth and ninetieth percentiles ('whiskers')
and fifth and ninety-fifth percentiles (dots).





2 Figure 3. Mean bias from full model predicted DWT for simulated low-counts with

3 all samples from four data-sets, based on ten cycles of random re-selection of

4 individuals. Box plots as for Fig. 2.



Figure 4. Mean bias from full model predicted DWT with simulated low-counts
extended to 240 tests using 12 samples from the Alaska training set and 18 samples
from the Jura training set in which at least 250 tests were counted. Box plots as for
Fig. 2.



- 1
- 2 Figure 5. Boot-strapped error estimates (mean of ten cycles of random selection) for
- 3 all samples from four data sets. Box plots as for Fig. 2.



- 6 Figure 6. Impact of increasing simulated count on % variance in amoebae species data
- 7 explained by DWT. All relationships are significant at P<0.001.



2 Figures 7a & b. Testate amoebae inferred depth to water table (TI-DWT) with

3 simulated counts of 10, 20 and 50 tests for palaeoecological sequences from the Praz-

4 Rodet peatland, Switzerland (7a) and site DLB, Alaska (7b).



- 1 Figure 8. Relationship between count and r^2 of the regression between the inferred
- 2 DWT from the full and the reduced data sets from the Praz-Rodet peatland,



3 Switzerland (8a) and site DLB, Alaska (8b).

- 4
- 5 Figure 9. Mean standard deviation of predictions based on ten cycles of random data
- 6 selection for the Praz-Rodet peatland, Switzerland (9a) and site DLB, Alaska (9b).



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 Table 1.Attributes of the training sets employed in this study showing number of samples, structure of transfer

 function model and jack-knifed root mean square error of prediction (RMSEP), Maximum Bias arid R

Location	n	Model structure	RMSEP _{jack} [cm]	Max Biaş _{ack} [cm]	R^2_{jack}	Reference
Alaska	91	WA-PLS (2 component)	9.7	14	0.55	Payne et al. (2006)
Greece	57	ML	1.8	2	0.82	Payne & Mitchell (2007)
Jura	48	WA-PLS (2 component)	8.0	21	0.62	Mitchell et al. (1999)
Turkey	42	ML	7.1	21	0.81	Payne et al. (2008)