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Differential parasitism of seed-feeding *Cydia* (Lepidoptera: Tortricidae) by native and alien wasp species relative to elevation in subalpine *Sophora* (Fabaceae) forests on Mauna Kea, Hawaii

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Abstract

Alien parasitic wasps, including accidental introductions and purposefully released biological control agents, have been implicated in the decline of native Hawaiian Lepidoptera. Understanding the potential impacts of alien wasps requires knowledge of ecological parameters that influence parasitism rates for species in their new environment. *Sophora* seed-feeding *Cydia* spp. (Lepidoptera: Tortricidae) were surveyed for larval parasitoids to determine how native and alien wasps are partitioned over an elevation gradient (2200–2800 m) on Hawaii Island, Hawaii. Parasitism rate of native *Euderus metallicus* (Eulophidae) increased with increased elevation, while parasitism rate by immigrant *Calliephialtes grapholithae* (Ichneumonidae) decreased. Parasitism by *Pristomerus hawaiiensis* (Ichneumonidae), origins uncertain, also decreased with increased elevation. Two other species, *Diadegma blackburni* (Ichneumonidae), origins uncertain, and *Brasema cushmani* (Eupelmidae), a purposefully introduced biological control agent for pepper weevil, did not vary significantly with elevation. Results are contrasted with a previous study of this system with implications for the conservation of an endangered bird species that feed on *Cydia* larvae. Interpretation of results is hindered by lack of knowledge of autecology of moths and wasps, origins, phylogeny, systematics, competitive ability, and physiological limitations of each wasp species. These factors should be incorporated into risk analysis for biological control introductions and invasive species programs.

Introduction

Parasitism of native Hawaiian insects by alien wasps has been of growing concern in the last few

decades (Howarth 1983b; Gagne and Howarth 1985; Follet and Duan 1999; Follet et al. 2000; Asquith and Miramontes 2001; Henneman and Memmott 2001; Alyokhim and Messing 2003).

Alien wasps have been implicated in the decline or extinction of 16 lepidopteran species in Hawaii (Gagne and Howarth 1985), with focus being drawn to wasps intentionally introduced as biological control agents (Howarth 1983b, 1991, 2000, 2001; Asquith and Miramontes 2001; Henneman and Memmott 2001). Of particular concern is the extent to which alien species can disrupt native foodwebs, either directly by depletion of particular prey species, or indirectly through competition, or apparent competition that alters predator-prey relationships (Lewis et al. 2002; Morris and Lewis 2002; Pearson and Callaway 2003). As a result, endemic prey species can become rare or extinct (Tothill et al. 1930; Gagne and Howarth 1985; Howarth 1991, 2001).

Hawaii is home to 1149 known species of Lepidoptera, 957 of which are endemic (Zimmerman 1958a, b, 1978; Nishida 2002). Although there are no Lepidoptera families or subfamilies considered endemic to Hawaii, 35 of 65 genera containing endemic species are considered endemic genera, suggesting that radiations of Hawaii species are relatively recent (Zimmerman 1948; Gagne and Howarth 1985; Nishida 2002). Some notable radiations include at least 350 species of *Hypomocoma* (Cosmopterygidae) on lichens, mosses, ferns, and many genera of flowering plants; 62 species of *Eudonia* (Crambidae) on native plants; 49 species of *Carposina* (Carposinidae), mostly fruit and leaf miners; 40 species of *Thyrocopa* (Oecophoridae), including flightless species; 22 species of *Eupithecia* (Geometridae), most of which the larvae are predacious on other insects; and several species of cave-obligate *Shrankia* (Noctuidae) (Swezey 1954; Zimmerman 1958a, b, 1978; Montgomery 1982; Howarth 1983a; Nishida 2002).

In contrast, of 78 species of Ichneumonidae and 79 species of Braconidae (Hymenoptera), only 39 and 3 species are endemic, while 8 and 39 species are purposeful introductions, respectively (Nishida 2002). Over the 110 plus years of Hawaii's biological control program, at least 86 hymenopteran and tachinid (Diptera) parasitoid species, 12 carabid (Coleoptera) species, common mynah (*Acridotheres tristis* (Linnaeus)) and cane toad (*Bufo marinus* (Linnaeus)) have been released against lepidopteran pests (Funasaki et al. 1988). The list of accidental introductions is even greater (66 species of Ichneumonidae and Braconidae

alone). Until recently, however, the extent to which these alien species have infiltrated native habitats and affected native species has not been well documented.

Recent studies demonstrate the toll that alien parasitoids exact on native Hawaiian Lepidoptera. For example, Asquith and Miramontes (2001) collected 16 species of Lepidoptera-specific parasitoids (Ichneumonidae, Braconidae: Hymenoptera) using Malaise traps in a native, montane forest on Kauai Island. Of the 2017 specimens collected, 10% (8 species) were Hawaiian natives, 40% (6 species) were accidental introductions, and 50% (2 species) were purposefully introduced biological control agents. Likewise, in a study of wet forest leaf-feeding Lepidoptera on Kauai Island, Henneman and Memmott (2001) found that 83% of parasitoids emerging from endemic Hawaiian caterpillars were originally introduced as biological control agents, 14% were accidental introductions, and only 3% were native to Hawaii. Similarly, 76% of wasps parasitizing seed-feeding caterpillars at high elevations on Hawaii Island are possibly of non-Hawaiian origins (Brenner et al. 2002).

There is little debate that alien species can have profound impacts on the native environments and food webs they invade. And the argument, that even well-intentioned biological control efforts can have devastating non-target impacts when proper risk analysis is not undertaken, is firmly established (Howarth 1991, 2000, 2001; Onstad and McManus 1996; Lockwood et al. 2001; Louda et al. 2003b; Carruthers 2004). It is now time to take the next step: to provide data on the ecological parameters that affect the relative impacts of both purposefully and accidentally introduced alien species in order to better evaluate risks of future control efforts and alien invasions (Arnett and Louda 2002; Louda et al. 2003a). In this study, we examine how native and alien parasitic wasps that parasitize seed-feeding Lepidoptera sort along an elevation gradient of *Sophora* forest in Hawaii. We hypothesize that elevation (a proxy for temperature and humidity) is an important ecological parameter for determining the relative impacts of competing parasitoid species.

Sophora chrysophylla (Salisb.) Seem (Fabaceae) is an ecologically important tree that dominates the high elevation forests of Mauna Kea on the Island of Hawaii. *Sophora* trees represent over

98% of the forest canopy from 2300 to 2800 m elevation (Hess et al. 1999), and provide nesting and foraging habitat for several endemic bird species (Scott et al. 1986; Hess et al. 2001). The seeds of *Sophora* are the primary food resource of palila, *Loxioides bailleui* Oustalet (Fringillidae: Drepanidinae), an endangered forest bird endemic to Mauna Kea (Van Riper 1980; Pratt et al. 1997; Banko et al. 2002). *Sophora* seeds are also the only known host for several species of endemic Hawaiian *Cydia* (Lepidoptera: Tortricidae) caterpillars (Zimmerman 1978). Furthermore, *Cydia* larvae found in *Sophora* seeds are the most important insect prey for developing palila nestlings (Banko et al. 2002), being present in 98% of fecal samples examined (USGS unpublished data).

Several species of wasps are historically known to parasitize Hawaiian *Cydia* larvae: the endemic *Eupelmus pelodes* Perkins (Eupelmidae); accidental alien *Trathala flavoorbitalis* (Cameron) (Ichneumonidae); and *Pristomerus hawaiiensis* Perkins (Ichneumonidae), origin uncertain (Perkins 1913; Swezey 1954; Zimmerman 1978). In addition to *P. hawaiiensis*, Brenner et al. (2002) reared the alien *Calliephialtes grapholithae* (Cresson) (Ichneumonidae); the native *Euderus metallicus* (Ashmead) (Eulophidae); *Diadegma blackburni* (Cameron) (Ichneumonidae), origin uncertain; and three specimens of the purposely introduced wasp *Braconia cushmani* (Crawford) (Eupelmidae) from *Sophora* seeds containing *Cydia* larvae.

In their 1996–1997 study, Brenner et al. (2002) found that parasitism of *Cydia* caterpillars ranged from 20 to 94% with overall parasitism rate decreasing with increased elevation. Because endangered palila feed on *Cydia*, a finer degree of resolution was needed to evaluate the threat posed by parasitoids. In this study we expand on the results of Brenner et al. (2002) to better understand the relationship between parasitism by different wasp species and elevation.

Methods

Site description

Sample sites were located from 2200 to 2800 m within the *Sophora* dominated forests of Mauna Kea volcano, Hawaii Island, Hawaii (Figure 1). Rainfall averages 511 mm/year at 2260 m eleva-

tion on the western slope of Mauna Kea (58 year average) (Juvik et al. 1993). Annual average temperature at 2600 m is 11 °C with mean daily maximum ranging from 15 to 17 °C and mean daily minimum ranging from 4 to 9 °C (Juvik et al. 1993). During the current study, rainfall and temperature varied throughout the study area with the general trend of temperature decreasing with increased elevation, and rainfall following a gradient decreasing from east to west (USGS unpublished data). Vegetation at the highest elevations is dominated by sparse *Sophora chrysophylla* trees. Several species of grasses (Poaceae), particularly *Dactylis glomerata* Linnaeus, *Trisetum glomeratum* (Kunth) Trin., *Poa* spp., and *Bromus* spp. become abundant along the mid-elevation sites. At the lowest elevations *Sophora* is codominant with *Myoporum sandwicense* Gray (Myoporaceae) among scattered patches of *Chamaesyce olowaluana* (Sherff) Croizat & Degener (Euphorbiaceae), *Coprosma montana* Hillebrand (Rubiaceae), *Dodonaea viscosa* Jacquin (Sapindaceae), *Dubautia* spp. (Asteraceae), and *Santalum paniculatum* Hooker & Arnott (Santalaceae) (Hess et al. 1999, USGS unpublished data).

Sampling

Sampling took place from December 1998 to December 2001 in March, June, September, and December along transects that follow an elevation gradient from tree line (circa 2900 m) to 2100 m that were established on Mauna Kea in 1983 to monitor forest bird populations. Four transects spaced at least 500 m apart were chosen for each of the north, south, and west facing slopes. *Sophora* forest is truncated on the eastern slope of Mauna Kea by pasture lands and therefore was not included in this study. Sample stations were located at 2200, 2350, 2500, 2650, and 2800 m elevation (Figure 1). For four days of each sampling month, we searched for seedpods that showed signs of *Cydia* caterpillars (≤ 1 mm silk plug) from within 50 m of each sample station. Trees sampled during each sampling bout were chosen by randomly selecting a distance (within 50 m) and direction (at 22.5° increments) from the center of each sampling station. Low *Sophora* tree densities at some stations, particularly at the highest and lowest elevations, meant that the same

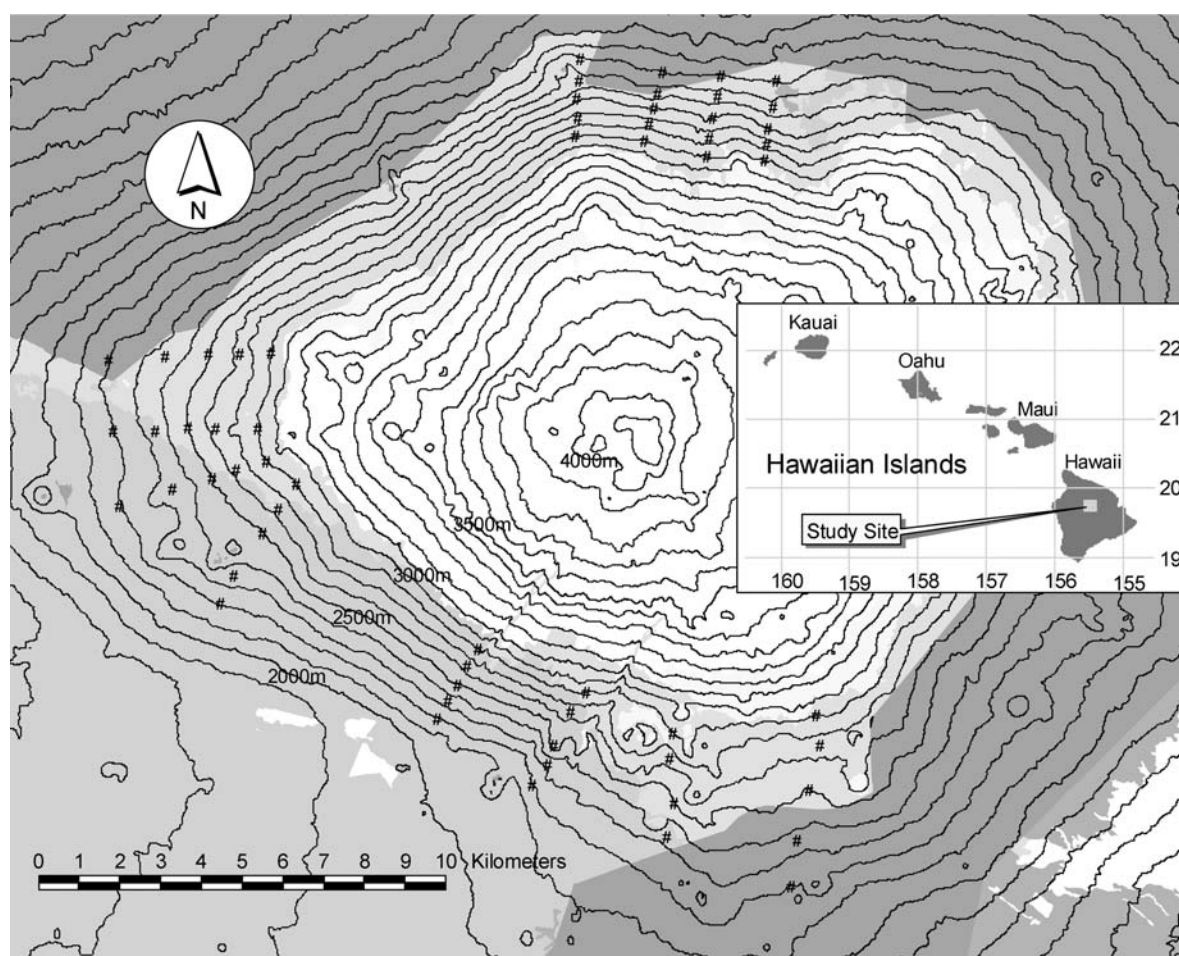


Figure 1. Location of study site on Mauna Kea volcano, Hawaii Island, Hawaii. Symbols (●) represent sites at which samples were collected. Contour lines are shown at 100 m intervals. Dark gray represents pasture lands with scattered *Sophora*, medium gray represents *Sophora* mixed with *Myoporum* woodland, and light gray represents *Sophora* dominated woodland. Elevations above woodlands and a narrow band of shrubland are barren subalpine stone desert with occasional herbs and grasses. The inset shows the study site in the context of the Hawaiian Archipelago.

trees were likely to be sampled during successive sampling bouts.

Ten seedpods collected from each of the five elevations on each of the three aspects of the mountain were returned to the laboratory for rearing of *Cydia* larvae (i.e. 10 pods \times 5 elevations \times 3 aspects = 150 seedpods for each sampling bout). Following Brenner et al. (2002), each seedpod was placed individually in 240 ml clear plastic containers fitted with screen lids (to promote air circulation and prevent molding). Containers were checked daily for the emergence of moths or wasps. After 3 months, seedpods showing no activity were dissected to determine caterpillar

occupancy. Live caterpillars were returned to their seedpods to finish development. During periods when seedpod production was low, less than 10 seedpods were selected. Not all seedpods collected contained caterpillars. Therefore, sample sizes varied from site to site for each sampling bout.

Analyses

Cydia parasitism rate was calculated as the percent of wasps that emerged from all caterpillars that survived to produce either a moth or a wasp. Parasitism rate was calculated for each wasp

species at each elevation for each month and year for a total of 63 data points. Samples were pooled across aspects for analysis of elevation, and across elevation for analyses of aspect. Analysis of seasonal trends, interactions of parameters, and variance within and among sites will be published separately. We used two-sample *t* tests to compare the means from the three aspects. We used analysis of variance to determine if parasitism rates were significantly different for each elevation and linear regression to determine how much of the variation is explained by elevation. Although log-transformed data showed more even variances around means, there was no difference in significance of trends found in non-transformed data. Therefore, non-transformed data are presented here for greater ease of interpretation.

Results

Of 1122 caterpillars reared over a three year period (1999–2001), 443 (39.5%) were parasitized and 679 (60.5%) emerged as adult *Cydia* moths (Table 1). An additional 763 caterpillars that were collected died as a result of desiccation, pseudoparasitism (Jones et al. 1986), inadequate food within the seed they occupied, or other undetermined causes. These larvae were not included in analyses. There was no trend for overall parasitism for the three aspects (Figure 2) (*t* test, $p > 0.10$ for all pairwise comparisons), or over the five elevations (Figure 3) ($R^2 = 0.01$, $F = 0.6$, $p = 0.46$).

Of 1122 caterpillars, 16% were parasitized by *Euderus metallicus* (Ashmead) (Eulophidae). Eight percent were parasitized by *Diadegma blackburni* (Cameron), 8% by *Pristomerus hawaiiensis* Perkins, and 3% by *Calliephialtes grapholithae* (Cresson) (Ichneumonidae). Additionally, 3% were parasitized by *Brasema cushmani* (Crawford) (Eupelmidae) (Table 1). Except for parasitism by *E. metallicus* being greater on the western slope than the northern slope, there were no obvious relationships between parasitoid species and aspect (Figure 2), and none were statistically significant (*t* test, $p > 0.10$). However, parasitism rates for some wasp species varied across elevations. Percent parasitism by *P. hawaiiensis* (Figure 4e) decreased significantly from 23% at 2200 m to 2% at 2800 m ($R^2 = 0.24$, $F = 19.6$, $p < 0.001$). Likewise, percent parasitism by *C. grapholithae* (Figure 4d) decreased significantly ($R^2 = 0.21$, $F = 16.4$, $p < 0.001$) with increased elevation from 8% at 2200 m to 1% at 2800 m. Conversely, *E. metallicus* (Figure 4a) increased significantly from 5% at 2200 m to 24% at 2800 m ($R^2 = 0.16$, $F = 11.3$, $p = 0.001$). The two other parasitoid species, *B. cushmani* and *D. blackburni*, (Figure 4b, c, respectively) did not vary significantly ($R^2 < 0.03$, $F < 2.0$, $p > 0.15$) with elevation.

All three ichneumonid wasps are solitary endoparasitoids that emerged one per *Cydia* larva. Likewise, the eupelmid *B. cushmani* emerged one per host larva. The eulophid *E. metallicus*, however is a gregarious ectoparasitoid having between 1–12 parasitoids emerge from each *Cydia* larva. A total

Table 1. Total number of *Cydia* moths and parasites reared from *Cydia* spp. larvae inhabiting *Sophora* pods collected from Mauna Kea (1999–2001)

Species:		<i>Euderus metallicus</i> ^a	<i>Brasema cushmani</i>	<i>Diadegma blackburni</i>	<i>Calliephialtes grapholithae</i>	<i>Pristomerus hawaiiensis</i>	<i>Cydia</i> spp.	Total
Overall	Total	182	34	97	37	93	679	1122
Aspect	N	50	10	24	17	36	247	384
	S	61	9	30	16	26	210	352
	W	71	15	43	4	31	222	386
Elev. (m)	2200	9	2	5	17	31	92	156
	2350	26	4	15	11	34	131	221
	2500	46	11	34	8	19	157	275
	2650	61	13	31	0	6	163	274
	2800	40	4	12	1	3	136	196

Numbers are summed for all sampling bouts for all three years. Numbers for aspect are pooled across elevations. Numbers for elevation are pooled across aspects. ^aNumbers represent the number of *Cydia* larvae parasitized by *E. metallicus*, rather than the number of *E. metallicus* that emerged because more than one wasp would often emerge from one moth larva. All other wasps emerged one per moth larva.

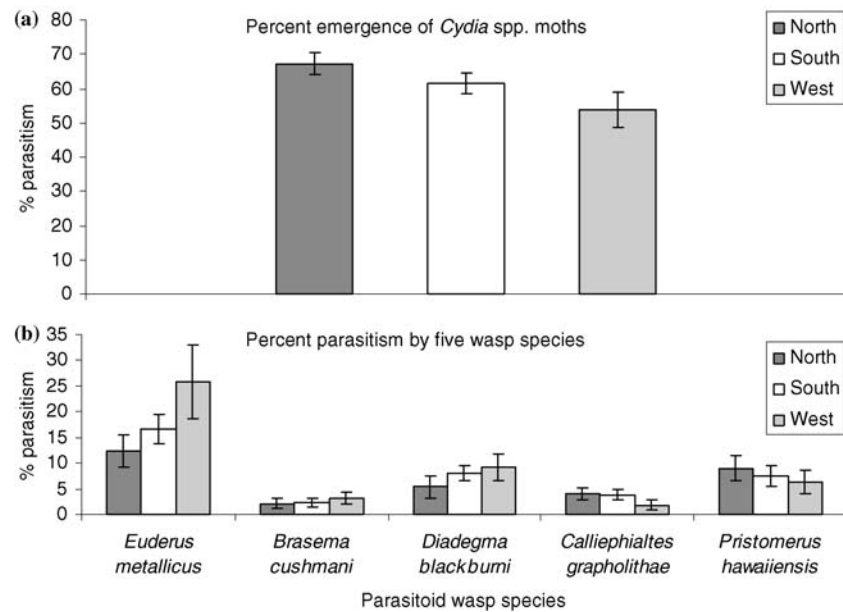


Figure 2. Mean percent emergence of *Cydia* moths or parasitoid wasps from *Cydia* larvae on Mauna Kea, Hawaii 1998–2001 relative to aspect (north, south, or west) of sample sites. (a) Mean percent emergence of *Cydia* moths. (b) Mean percent parasitism by five wasp species. Mean and standard error were calculated from 13 sampling bouts pooled across elevations. Note. Means for sampling bouts are different than dividing totals in Table 1 that are pooled across all sampling bouts. Pairwise *t* tests for aspect for each wasp were not statistically significant ($p > 0.10$).

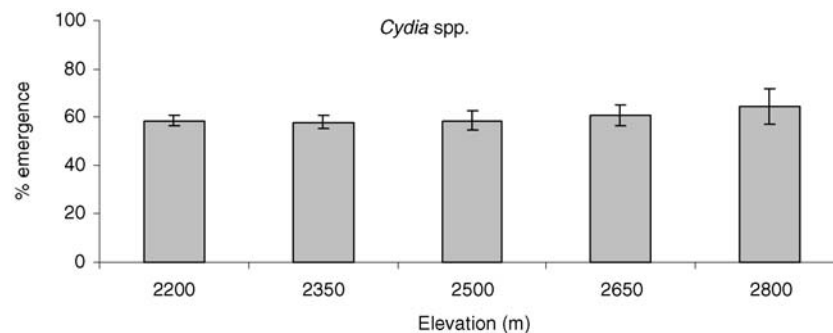


Figure 3. Mean percent emergence of *Cydia* moths (non-parasitized larvae) relative to elevation on Mauna Kea, Hawaii 1998–2001. Mean and standard error were calculated from 13 sampling bouts pooled across north, south, and west facing slopes for each elevation for each wasp species. Note. Means for sampling bouts are different than dividing totals in Table 1 that are pooled across all sampling bouts. There is no significant trend in emergence of *Cydia* over elevation ($R^2 = 0.01$).

of 1210 individuals of *E. metallicus* (664 males, 546 females) emerged from 182 *Cydia* larvae.

Discussion

The results of this study contrast those of Brenner et al. (2002) in several important ways. Although

overall parasitism rate in this study (40%) was comparable to that of the previous (39%), there was no trend for overall parasitism rate over elevation in the present study. The previous study demonstrated a highly significant trend for overall reduction of parasitism with increased elevation. This difference may be based partly on the areas surveyed in the two studies. The previous study

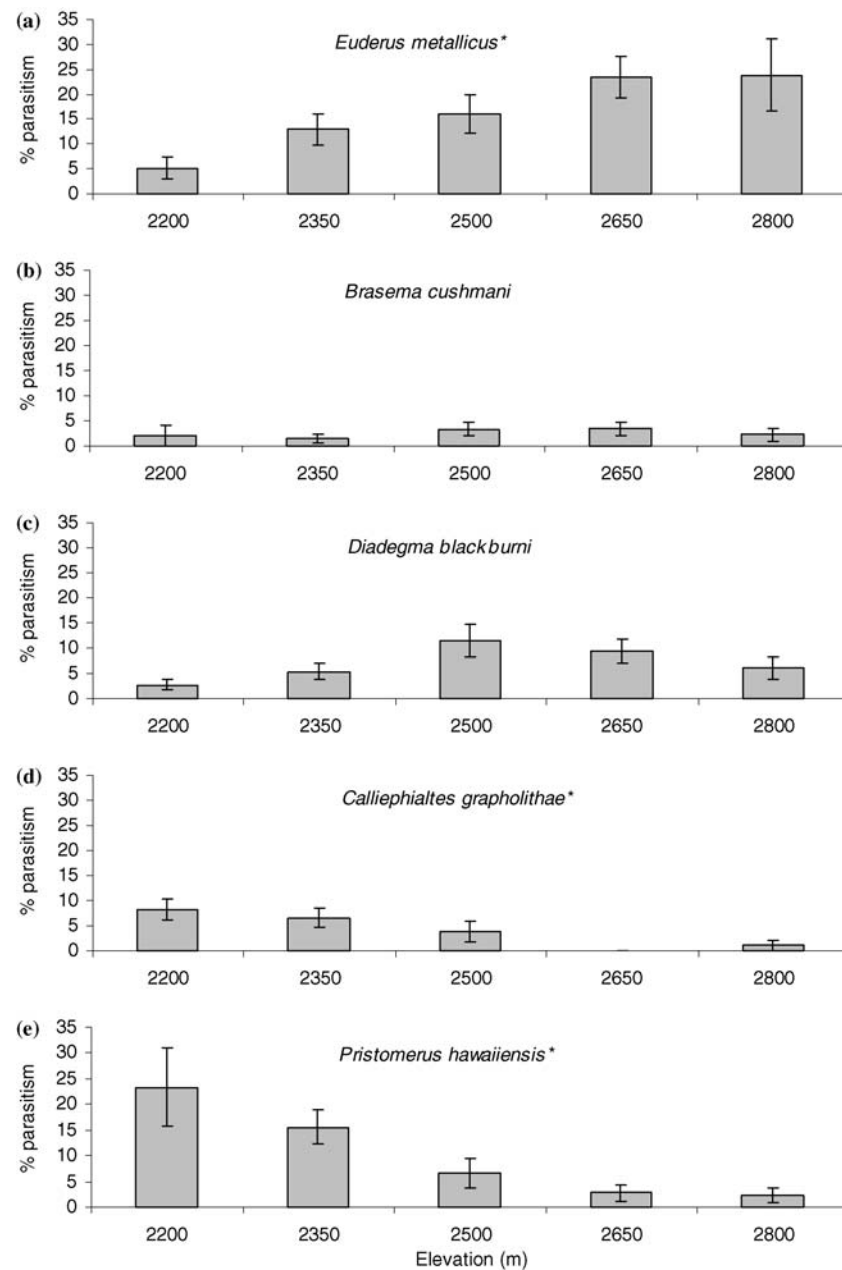


Figure 4. Mean percent parasitism of *Cydia* larvae by five wasp species on Mauna Kea, Hawaii 1998–2001. Mean and standard error were calculated from 13 sampling bouts pooled across north, south, and west facing slopes for each elevation for each wasp species. Note. Means for sampling bouts are different than dividing totals in Table 1 that are pooled across all sampling bouts. (*) after a name indicates a statistically significant elevational trend.

included two sites below 2000 m where parasitism reached 100% in some months, while the present study included a greater number of sites, all above 2100 m. One important conclusion to be drawn from the current results is that the apparent refu-

gium for *Cydia* larvae at higher elevations in the previous study is not corroborated by this study. Banko et al. (2002) have demonstrated that the endangered forest bird, *Loxioides bailleui*, whose principle food resources are *Sophora* seeds and

Cydia larvae, require a wide elevation gradient for adequate foraging. Lack of an upper elevation refugium, therefore, may have important consequences to the long-term viability of *Cydia* species as well as the birds that feed on them.

A second difference among the two studies is that in the present study parasitism by *Euderus metallicus* became predominant at higher elevations. The previous study indicated that parasitism by *E. metallicus* did not vary significantly with elevation. The three ichneumonid species showed similar patterns of parasitism in the two studies, with *Pristomerus hawaiiensis* and *Calliephialtes grapholithae* rates decreasing with increased elevation, although the increase in *Diadegma blackburni* rates at higher elevations was not significant in the present study. A third difference is the greater abundance of *Brasema cushmani* in the present study. Only three specimens of *B. cushmani* were recorded during the previous study, compared to 34 specimens (7% of parasitoids) recovered in the present study. This is more likely a result of more extensive and intensive sampling than a spread of this wasp in the last few years.

Given differential parasitism rates for each wasp species relative to elevation, how do the origins of each elucidate patterns of resource exploitation by native and alien species? Unfortunately, determining the origins of wasp species is problematic due to incomplete knowledge of the Hawaiian fauna, extensive commerce and travel through the Pacific region, clandestine introductions, and incomplete records of purposeful biological control introductions, particularly in the first half of the 20th century (Swezey 1931; Funasaki et al. 1988; Howarth 1991).

Of one species, *Brasema cushmani* (Eupelmidae), we can be sure, because it was first released in Hawaii in 1934 (Swezey 1939) to control pepper weevil (*Anthonomus eugenii* Cano). It has since been recorded from the moth, *Agonoxena argaulu* Meyrick, as well as six beetle species, one tephritid fruit fly (*Procecidochares utilis* Stone), and the mantis *Tenodora angustipennis* Saussure, all alien to Hawaii (Funasaki et al. 1988). Interestingly, four of the non-target hosts (fruit fly and three weevils) are themselves purposeful introductions against invasive weeds in Hawaii (Funasaki et al. 1988). Stein (1983) indicated that although this wasp has been recorded up to 1525 m elevation in Hawaii, it was more prevalent below 1000 m. He

also suggests, counter-intuitively, that in Hawaii, cooler temperatures at higher elevations may shorten the egg and larval stages while prolonging pupation; which can affect its competitive status relative to other larval parasitoids. In our study, *B. cushmani* did not show any significant trends relative to elevation, but maintained a low level of parasitism across the study area (2200–2800 m).

Calliephialtes grapholithae (Ichneumonidae) was first detected in Hawaii on Oahu in 1976 (Beardsley 1980), but may have arrived accidentally much earlier. It is the most prominent parasitoid of *Cryptophlebia illepida* (Butler) (Tortricidae), a native pest of macadamia nut in Hawaii (Jones 2002). Previous reviews of *C. illepida* parasitoids from macadamia nuts before 1957 include *Pristomerus hawaiiensis* and *Euderus metallicus*, but not *C. grapholithae* (Namba 1957). Nor do Zimmerman (1978) or Stein (1983) mention parasitism of Lepidoptera larvae by *C. grapholithae* in their reviews of parasitoids of native Hawaii caterpillars. This wasp may therefore still be in its early stages of establishment and expansion. In this study, *C. grapholithae* was most prominent at lower elevations and declined with increased elevation.

Euderus metallicus (Eulophidae) is most likely a native to Hawaii, although Nishida (2002) lists it as possibly an accidental introduction. It was first collected in Hawaii in 1894, just 4 years after Hawaii's officially sanctioned biological control program began (Ashmead 1901; Funasaki et al. 1988). This wasp is also known from Guam, Palau, Yap, and the Caroline Atolls (Yoshimoto 1965; Yoshimoto and Ishii 1965) and is probably widespread on Pacific islands. In Hawaii it has been recorded from over 25 species of caterpillars in eight microlepidoptera families, as well as an *Agromyza* species (Diptera: Agromyzidae) and *Heteramphus swezeyi* Perkins (Coleoptera: Curculionidae) (Yoshimoto 1965; Zimmerman 1978). There may be some confusion about this insect, however, because although Swezey (1915) indicates that this species feeds singly on host larvae, we reared up to 12 individuals from a single host larva. Swezey (1915) also claims to have reared a number of close relatives of this species, which are not recognized in a more recent review of Hawaiian Eulophidae (Yoshimoto 1965). In the present study, this wasp showed increased prevalence with increased elevation, indicating it is adapted to the highest elevations of *Sophora* forest on Hawaii.

Diadegma blackburni and *Pristomerus hawaiiensis* (Ichneumonidae) are more problematic. Although first collected in Hawaii in 1883 (Ashmead 1901), *D. blackburni*, listed by Nishida (2002) as an accidental introduction, was also recorded from Oregon in 1897 (Carlson 1979). In the 103 years since its original description based on specimens from Hawaii, *D. blackburni* has gone through several name changes and may not represent the same species that was found in Oregon (Ashmead 1901; Zimmerman 1978; Carlson 1979). It may be that *D. blackburni* in Hawaii is an endemic species. Until a revision of this genus is completed, however, the place of origin cannot be assumed. In our study, *D. blackburni* did not vary significantly with elevation, although there was a notable increase in parasitism by this wasp from 2500 to 2650 m.

Pristomerus hawaiiensis was first described from Hawaii in 1910 (Perkins 1910). Although the only record of this wasp outside of Hawaii is a purposeful introduction from Hawaii to California for control of pink bollworm (*Pectinophora gossypiella*: Gelechiidae) (Legner 1979), Fullaway and Krauss (1945) and Stein (1983) suggest it is an oriental immigrant without providing the reasoning behind their assumption. In contrast, Nishida (2002) lists this species as possibly endemic to Hawaii. Both *P. hawaiiensis* and *D. blackburni* are common and have a broad host range in Hawaii (over 25 species in seven families for *P. hawaiiensis*, and over 10 species in six families for *D. blackburni* (Zimmerman 1978)), particularly in concealed-feeding Lepidoptera (i.e. inside fruits, seeds, or twigs). This can be interpreted as a long history of adaptation in the Hawaiian Islands in the absence of competitors, or as a recent introduction of a weedy species. Zimmerman (1978) also suggests that *P. hawaiiensis* may have been confused with one or more other species for some parasitism records. Therefore we cannot definitely denote *P. hawaiiensis* as native or alien. In the current study, *P. hawaiiensis* decreased with increased elevation.

Differential parasitism of these five wasps over a 600 m elevation gradient suggests partitioning of the environment. The decrease in parasitism rate by *C. grapholithae*, a known alien, and increase by *E. metallicus*, presumed native, with increased elevation suggests competition between natives and aliens is forcing native wasps upslope. However, *E. metallicus* is a prominent parasitoid

of Lepidoptera in many native forests in Hawaii and can be found at elevations below 900 m (Swezey 1954; Zimmerman 1978; Oboyski, unpublished data). Therefore, controlled competition experiments will be a more direct method to determine the importance of competition in this system.

Alternatively, decrease of parasitism rates with increased elevation may be an indication of physiological limitations. Cold tolerance experiments of *C. grapholithae* parasitizing *Cydia caryana* (Fitch) demonstrated that 6 week exposure to 4 °C resulted in 50% reduction in emergence of parasitoids (Yonce et al. 1996). In the present study elevation correlates with mean temperature, temperature variance, and relative humidity, with a winter average daily minimum of approximately 4 °C at 2600 m (Juvik et al. 1993). Decreased parasitism rates by *C. grapholithae* may reflect physiological limitations at higher elevations in our study area. Therefore, controlled experiments manipulating temperature and humidity will be important to understand how physiological limitations impact the distribution and competitive status of native and alien species.

The lack of knowledge of species identities and their places of origin may be the greatest impediment to addressing concerns of conservation in this system. Whether the current assemblage of interacting species represents millions of years of coevolution or a recent perturbation to the system will make a difference in the expected outcome of current dynamics. Swezey (1954) witnessed the decline of *Omiodes* (Pyralidae) caterpillars on wild banana, a result of parasitism by *Trathala flavororbitalis* (Cameron) and *Casinarina infesta* (Cresson) (Ichneumonidae) within a 30 year period following their documented arrival in Hawaii in 1910 and 1921, respectively. Likewise, Illingworth (1916) observed the extinction of a local population of *Omiodes blackburni* (Butler) in Palolo Valley, Oahu, via predation of their eggs and larvae by the immigrant ant *Pheidole megacephala* (Fabricius). Such direct, although anecdotal, observations of decline in native species are rare, or at least not well documented, but are especially instructive regarding the impacts of alien species. For three species in the current study (*D. blackburni*, *P. hawaiiensis*, and *E. metallicus*) it is unclear how long they have been associated with native Hawaiian Lepidoptera. Only extensive

long-term surveys can help predict the fate of these interactions until phylogenetic and biogeographic histories of these species are completed.

Although differential parasitism rates by five wasp species in this study indicate some sorting along an elevation gradient, interpretation of results is hindered by a lack of knowledge of place of origin, identity, competitive abilities, and physiological limitations for each species. Increased biotic surveys in lesser understood biotic regions around the world will provide more certainty to biogeographic patterns of invasive species. Phylogenetic studies of proposed biological control agents can provide more certainty to the identity of species and insights into the evolutionary and physiological tendencies of control agents. Finally, it will be impossible to predict the potential impacts of incipient alien species or biological control introductions without first conducting the proper experiments to determine competitive abilities within guilds and physiological tolerances and limitations of competing species.

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