

Article

TiO² Modification with Transition Metallic Species (Cr, Co, Ni, and Cu) for Photocatalytic Abatement of Acetic Acid in Liquid Phase and Propene in Gas Phase

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Abstract: The commercial P25 titania has been modified with transition metallic species (Cr, Co, Ni, and Cu), added by impregnation with aqueous solutions of the corresponding nitrates. The preparation procedure also includes a heat treatment (500 ℃) in argon to decompose the nitrates, remove impurities and to strengthen the metal–TiO₂ interaction. The catalysts have been thoroughly characterized using N_2 adsorption, scanning electron microscopy (SEM), X-ray diffraction (XRD), UV-visible diffuse-reflectance spectroscopy (UV-vis DRS) and X-ray photoelectron spectroscopy (XPS), and have been tested in the aqueous phase decomposition of acetic acid and in the gas phase oxidation of propene, using an irradiation source of 365 nm in both cases. The photocatalytic activity of the four metal-containing catalysts varies with the nature of the metallic species and follows a similar trend in the two tested reactions. The effect of the nature of the added metallic species is mainly based on the electrochemical properties of the supported species, being Cu/P25 (the sample that contains copper) the best performing catalyst. In the photodecomposition of acetic acid, all the metal-containing samples are more active than bare P25, while in the gas phase oxidation of propene, bare P25 is more active. This has been explained considering that the rate-determining steps are different in gas and liquid media.

Keywords: photocatalysis; TiO₂; P25; transition metals; acetic acid; propene.

1. Introduction

With the increasing concern in the use of renewable resources and in finding better ways to use solar energy, solar light-activated photocatalysis has become an attractive tool. In particular, heterogeneous photocatalysis is a process of great potential for pollutants abatement in gas and liquid phases. This method has considerable advantages over some existing technologies: it destroys pollutants rather than transferring them to another phase; it usually leads to complete mineralization of organic pollutants into $CO₂$, H₂O and innocuous mineral salts; it operates at ambient conditions, with any type of substrate, without complex processing requirements and it can be easily implemented in both aqueous and gaseous applications [\[1](#page-14-0)[,2\]](#page-15-0).

In this work, heterogeneous photocatalysis, a promising advanced oxidation technology for removing contaminants at trace levels [\[3\]](#page-15-1), is selected to address the problem of two reactions of environmental interest: acetic acid decomposition and propene oxidation. In wastewater, several bacteria produce acetic acid (HAc) as a primary fermentation product $[4]$. However, due to its high chemical stability and/or low biodegradability [\[5\]](#page-15-3), HAc is a recalcitrant compound quite

difficult to decompose using conventional methods. Catalytic photodegradation is an interesting technique used for the degradation of recalcitrant contaminants, such as acetic acid, because of its relatively low cost and because it can be performed in mild operation conditions [\[6\]](#page-15-4). Furthermore, the photodecomposition of HAc generates gaseous products of energetic interest (mainly CH_4 and CO_2 , and H_2 in minor concentration) [\[7\]](#page-15-5). In this way, the transformation of water pollutants into syngas components and/or fuel compounds would be desirable from both the ecological and economical points of view. On the other hand, propene is a common air pollutant belonging to the category of volatile organic compounds (VOCs) that is involved in photochemical smog [\[8\]](#page-15-6) and is one of the major sources of indoor air pollution, as it is an important component of tobacco smoke [\[9\]](#page-15-7). Hence, its photo-oxidation is very interesting from an environmental point of view. However, the photocatalytic conversion of the two mentioned substrates is still far from practical implementation because of the lack of efficient and cheap catalysts [\[10\]](#page-15-8).

TiO² has emerged as one of the most promising semiconductor photocatalysts, as it is stable also in aqueous media, tolerant to both acidic and alkaline solutions, it is recyclable and relatively simple to prepare, and shows good ability for air and water purification [\[11\]](#page-15-9). However, a limitation of TiO² is that the photogenerated e−/h⁺ pairs undergo a rapid recombination, therefore decreasing the photocatalytic efficiency of the system. Besides, $TiO₂$ has a relatively wide band gap (3.0 eV for rutile and 3.2 eV for anatase phase [\[12\]](#page-15-10)), that lies in the UV radiation region and, as a result, only 5% of sunlight photons can be harvested to activate TiO₂ [\[13\]](#page-15-11).

The addition of transition metallic species to $TiO₂$ catalysts has shown to be useful to reduce the photogenerated e⁻/h⁺ recombination rate and, also, to improve the photoresponse to the visible light $[14-16]$ $[14-16]$. The band gap and electrochemical properties of TiO₂ are known to be modified when transition metallic ions replace Ti (IV) centers (substitutional doping), occupy interstitial sites (interstitial doping) or form aggregates on the $TiO₂$ surface [\[17\]](#page-15-14). Substitutional doping can occur in titanium dioxide when the difference between the atomic radii of the dopant and of Ti (IV) is less than 15% [\[18\]](#page-15-15). The introduction of dopants in the $TiO₂$ lattice causes local distortion of the crystal structure, thus altering the crystallinity degree and phase transformation. Doping (substitutional or interstitial) can also generate trap levels, which modify the band gap and allow the capture of the photo-induced e[−] in the conduction band (CB) of $TiO₂$, leaving the $h⁺$ in the valence band (VB) of $TiO₂$. This hinders the e⁻/h⁺ pairs recombination and, thus, promotes the photocatalytic efficiency [\[14](#page-15-12)[,16](#page-15-13)[,19,](#page-15-16)[20\]](#page-15-17). Additionally, metallic aggregates not chemically bonded to $TiO₂$ can act as electrons scavengers, also preventing the recombination of h^+ /e[−] pairs. On the other hand, when a metal with a certain work function (energy required for moving an electron from a Fermi level to the local vacuum level [\[21\]](#page-15-18)) is put in contact with a semiconductor of lower work function value, it provides a Schottky barrier that facilitates the transfer of electrons from the semiconductor to the metal. Thus, the metal serves as an electron trap which prevents electron migration to the semiconductor, avoiding recombination [\[22\]](#page-15-19). The ability of metal ions to act as effective traps is related, among others, with their electrochemical properties, the metal concentration and the intensity of the incident light [\[22–](#page-15-19)[24\]](#page-16-0). Considering all this, and with the purpose of enhancing the activity, the added transition metal ions should have a work function higher than that of $TiO₂$.

Although noble metals (such as Pt, Pd or Au) have proved to be effective to enhance the photocatalytic activity of $TiO₂$ [\[25\]](#page-16-1), they are unsuitable for large-scale commercial use due to their limited availability and high cost. In contrast, earth-abundant non-noble transition metal-based materials are promising alternatives to enhance the $TiO₂$ performance. For example, transition metals, such as Cu, Co, Ni, Cr, Fe, Mn and V [\[14,](#page-15-12)[16,](#page-15-13)[26\]](#page-16-2) have been reported to decrease the TiO₂ band gap and to extend its photo-response to the visible region.

The use of non-noble transition metals/TiO₂ systems for the photocatalytic degradation of organic pollutants in the liquid and gas phase has been described by several authors [\[14](#page-15-12)[,16](#page-15-13)[,19,](#page-15-16)[20,](#page-15-17)[27\]](#page-16-3). The incorporation of metallic species to titania has been often performed during the $TiO₂$ synthesis [\[14,](#page-15-12)[16,](#page-15-13)[26,](#page-16-2)[28\]](#page-16-4), and relatively few studies report the incorporation of metallic species by

impregnation over previously synthesized TiO₂ [\[19,](#page-15-16)[27\]](#page-16-3). However, the conclusions about the role of the incorporated metal are sometimes controversial. For example, Di Paola, et al. [\[27\]](#page-16-3) studied the 4-nitrophenol photodegradation in aqueous suspension and found a decrease of the $TiO₂$ photoactivity when it contains Co, Cr, Cu, Fe, Mo and V ions added by impregnation. In contrast, Tayade et al. [\[19\]](#page-15-16) found that the photocatalytic activity of Fe, Ni and Ag containing $TiO₂$ in the degradation of acetophenone and nitrobenzene in aqueous solution was higher than that of bare $TiO₂$, while the presence of Co or Cu led to a lower photoactivity.

All these considerations prompted us to prepare a series of $M/TiO₂$ photocatalysts (M = Cr, Co, Ni and Cu) by impregnation of the inexpensive, versatile and commercial P25 titania (from Degussa). These new materials were characterized by N_2 adsorption isotherms, scanning electron microscopy (SEM), X-ray diffraction (XRD), UV-visible diffuse-reflectance spectroscopy (UV-vis DRS) and X-ray photoelectron spectroscopy (XPS) analysis, and their photocatalytic activity was investigated in the photo-degradation of aqueous acetic acid into $CO₂$, $CH₄$ and $H₂$ and in the gas phase photo-oxidation of propene. The overall performances were discussed in terms of activity and selectivity as a function of the properties of the photocatalysts.

2. Materials and Methods

2.1. Preparation of M/P25 Samples

Degussa Aeroxide P25 TiO₂ has been used as a raw material because of its good activity in many photocatalytic reactions. Cr, Co, Ni, and Cu were incorporated to P25 by impregnation as follows: P25 (2 g) was put in contact with an aqueous solution (5 mL) of Cr(NO₃)₃.9H₂O, Co(NO₃)₂.6H₂O, Ni(NO₃)₂·6H₂O or Cu(NO₃)₂·3H₂O of the appropriate concentration to obtain 1 wt.% metal loading. The mixture was stirred (2 h) and sonicated (30 min) and, then, the solvent excess was removed (80 $^{\circ}$ C, 24 h).

All the samples were heat-treated in argon atmosphere (90 mL/min, $5\degree C/\text{min}$, $500\degree C$, 2 h) to decompose the nitrates, remove impurities and to strengthen the metal– $TiO₂$ interaction. The Ar atmosphere was selected according to the previous work [\[29\]](#page-16-5). The used nomenclature is M/P25-Ar $(M = Cr, Co, Ni or Cu)$. P25 was wetted and heat treated in the same conditions as $M/P25$ -Ar samples to use it as reference; this sample is named P25-Ar.

2.2. Characterization

Porosity (textural properties) of the photocatalysts has been charaterized by physical adsorption of gases. N² adsorption–desorption at −196 ◦C (after degasification (250 ◦C, 4 h)) was measured in a Quantachrome Autosorb-6B equipment (Quantachrome Instruments, Boynton Beach, FL, USA). The specific BET surface area (S_{BET}) and the micropore volume (V_{N2}) were determined by applying, respectively, the Brunauer–Emmett–Teller (BET) or Dubinin–Radushkevich equations to the N_2 adsorption isotherms. These determinations have an associated error of 2%. Mesopore volume (V_{meso}) was estimated by the difference in the volume of N₂ adsorbed as liquid at $P/P_0 = 0.9$ and $P/P_0 = 0.2$. The total pore volume (V_T) was determined by the nitrogen adsorption volume at a P/P₀ (relative pressure) of 0.99. For such calculation, the density of liquid nitrogen at −196 °C was used. The average pore size (\emptyset) was calculated from the nitrogen adsorption isotherms by the Barret-Joyner-Halenda (BJH) method [\[30–](#page-16-6)[32\]](#page-16-7).

SEM characterization was performed using the JSM-840 JEOL equipment (JEOL Ltd., Tokyo, Japan) and the scanning electron microscopy with energy dispersive X-ray (SEM-EDX) mapping was conducted with a Hitachi S3000N microscope (Hitachi, Tokyo, Japan) and the Bruker XFlash 3001 X-ray energy dispersive spectrometer (Bruker, Billerica, MA, USA).

XRD analysis using Cu K α (1.54 Å) radiation was performed in a SEIFERT 2002 equipment (Rich.Seifert & Co., Ahrensburg, Germany), in the angular 2θ range 6–80◦ and with a scanning velocity

of 2 \degree /min. The experimental errors are within \pm 0.1 \degree . The average crystallite size, referred to as crystal size, was calculated by Scherrer's equation (Equation (1)) [\[33\]](#page-16-8):

$$
B = \frac{K\lambda}{\beta\cos\theta} \tag{1}
$$

where B is the average crystallite size (nm); K is a constant, with a chosen value of 0.9; λ is the wavelength of the radiation source; $β$ is the full width at half maximum intensity (FWHM) and $θ$ is the Bragg angle of the maximum intensity peak. The amount of crystalline anatase (A_{Crvs}) and rutile (R_{Crvst}) has been calculated using the equations proposed by Jensen et al. [\[34\]](#page-16-9):

$$
A_{\text{Cryst}} = \frac{\frac{A_{\text{Anatase}(101)}}{A_{\text{CaF}_2(220)}} \times 100}{1.25}
$$
 (2)

$$
R_{\text{Cryst}} = \frac{\frac{A_{\text{Rutile}(110)}}{A_{\text{CaF}_2(220)}} \times 100}{0.90} \tag{3}
$$

where $A_{\text{Anatase (101)}}$, $A_{\text{Rutile(110)}}$ and $A_{\text{CaF2(220)}}$ are the peak areas determined in XRD spectra measured for a $50/50$ (w/w) $\text{TiO}_2/\text{CaF}_2$ mixture, and 1.25 and 0.90 are the ratios between areas of the same peaks when a 100% anatase or 100% rutile crystalline TiO₂ sample (respectively) is used. The percentage of crystalline (W_{Cryst}) and amorphous (W_{Am}) phases present in the photocatalysts were determined as indicated in Equations (4) and (5), respectively.

$$
W_{\text{Cryst}} = A_{\text{Cryst}} + R_{\text{Cryst}} \tag{4}
$$

$$
W_{Am} = 100 - W_{\text{Cryst}} \tag{5}
$$

Diffuse reflectance spectra (DRS) were determined by means of a Jasco V-670 UV-vis spectrophotometer (JASCO, Inc., Pfungstadt, Germany), using BaSO₄ as the reference sample. The spectra were recorded at room temperature in air, in the 200–800 nm range. The absorption edge wavelength was estimated from the intercept at zero absorbance of the high slope portion of each spectrum in the 200–800 nm range (absorbance method). The band gap energy was calculated using three different methods: absorbance, indirect and direct ones. In the absorbance method, the band gap can be calculated using Equation (6), where E_g is the band gap energy (eV) and λ is the edge wavelength (nm) [\[35\]](#page-16-10).

$$
E_g = \frac{1239.8}{\lambda} \tag{6}
$$

In the direct and indirect methods, the band gap values are obtained from the Kubelka-Munk function of the reflectance, plotting, respectively, $(F(R)$ hυ)² vs. hυ or $(F(R)$ hυ)^{0.5} vs. hυ [\[36\]](#page-16-11).

XPS spectra were obtained using a K-Alpha spectrophotometer (Thermo-Scientific, Waltham, MA, USA) with a high-resolution monochromator and the following specifications: Al anode (1486.6 eV) X-ray source, 5×10^{-9} mbar analysis chamber pressure and detection in constant energy mode with pass energy of 200 eV for the survey spectrum, and of 50 eV for the sweep in each individual region. Data analysis was performed with the peak-fit software of the spectrometer (Thermo Scientific Avantage Software, Waltham, MA, USA) using Gaussian functions with 20% Lorentzian component, and binding energy values were adjusted to the C1s transition (284.6 eV). This process has an associated error of \pm 0.2 eV [\[37\]](#page-16-12).

2.3. Photocatalytic Activity Measurements

Details of the photocatalytic activity measurements in the two studied reactions are presented next.

2.3.1. Photocatalytic Decomposition of Acetic Acid

Photocatalytic tests were performed in a cylindrical quartz reactor (Heraeus, type UV-RS-2, Hanau, Germany) with a medium pressure mercury vapor lamp (TQ-150, $\lambda_{\text{max}} = 365$ nm, Hanau, Germany) located in the center of the vessel and surrounded by a water-circulating jacket. Gas inlet and outlet are in the upper part of the reactor. One molar solution of acetic acid (0.35 L) and the photocatalyst (0.35 g) were introduced in the reactor. A stream of He (60 mL/min) was bubbled through the reactor for 2 h to purge lines and remove dissolved oxygen and, then, the cooling system and the UV lamp were turned on. The process was conducted for 12 h under continuous magnetic stirring. Some experiments were repeated twice to check reproducibility. A blank experiment (without catalyst) was also performed in the same conditions to measure the photolysis effect. Analysis of the outlet gas stream was performed by mass spectrometry (Balzers, Thermostar GSD 301 01, Asslar, Germany). Calibrated cylinders of 2000, 2000 and 500 ppmv in helium of CH_4 , CO_2 and H_2 , respectively, were used to quantify the products. The photocatalytic activity was expressed as indicated in Equation (7).

$$
Production of X = \frac{mmol X}{mol HAc}
$$
 (7)

where X is CH_4 , CO_2 or H_2 .

2.3.2. Photocatalytic Oxidation of Propene

The experimental system used to perform the activity tests was designed in our laboratory and consists of a quartz reactor (AFORA, Fisher Scientific, Hampton, NH, USA) and a 365 nm Philips UV-A lamp placed parallel to the quartz reactor at about 1 cm. The assembly quartz reactor lamp is surrounded by a cylinder covered by tinfoil. A scheme of this system is shown elsewhere [\[8\]](#page-15-6).

The photocatalyst (0.11 g) was placed on a quartz wool plug inside the reactor and then, after purging with helium, a stream of 100 ppmv propene in air was fed to the reactor at 25 ◦C, being the outlet gas continuously analyzed by mass spectrometry (MS, Balzers, Thermostar GSD 301 01, Asslar, Germany). Once the propene concentration is stable (after about 3 h), the lamp is switched on and kept working until a stationary propene signal is achieved (usually 3 h). All the tests were carried out with 30 and 60 mL/min propene flow. A blank experiment (without catalyst) was also performed in the same conditions to measure the photolysis effect. The experiments were repeated at least twice to check reproducibility. Propene conversion was calculated using Equation (8):

$$
\text{Propene conversion } (\%) = \frac{C_{\text{initialC}_3\text{H}_6} - C_{\text{stationary}}_{\text{C}_3\text{H}_6}}{C_{\text{initialC}_3\text{H}_6}} \times 100 \tag{8}
$$

where $C_{initialC3H6}$ is the initial propene concentration, 100 ppmv, and $C_{stationaryC3H6}$ is the stationary propene concentration reached after a certain irradiation time.

The produced $CO₂$ was quantified using a calibrated cylinder (300 ppmv $CO₂$ in helium). Water evolution was also followed by MS and a mass scan was carried out to determine if additional oxidation compounds were present in the outlet stream. An example of the type of information obtained in this type of experiment has been shown in a previous publication [\[38\]](#page-16-13).

3. Results

3.1. Textural and Morphological Properties

Figure [1](#page-5-0) shows the N_2 adsorption–desorption isotherms of $M/P25$ -Ar and P25-Ar samples. It can be observed that the adsorption isotherms are type IV, according to the IUPAC classification [\[39\]](#page-16-14), with a hysteresis cycle typical of mesoporous materials. The incorporation of metallic species results in a slightly lower adsorption capacity than P25-Ar.

Figure 1. N₂ adsorption–desorption isotherms at −196 °C of the photocatalysts.

The parameters calculated from the adsorption isotherms (Fig[ure](#page-5-0) 1) are collected in Table 1. It The parameters calculated from the adsorption isotherms (Figure 1) are collected in Table [1.](#page-5-1) It can be observed that all the M/P25 samples have similar textural properties, with mesopore volume, total pore volume and average pore diameter (Ø) slightly lower than those of P25-Ar. It can be total pore volume and average pore diameter (Ø) slightly lower than those of P25-Ar. It can be mentioned that the surface area and micropore volume of Cr/P25-Ar are slightly higher than for the mentioned that the surface area and micropore volume of Cr/P25-Ar are slightly higher than for the rest of samples. rest of samples.

Photocatalyst	S_{BET} (m ² /g)	V_{N2} (cm ³ /g)	V_{meso} (cm ³ /g)	V_T (cm ³ /g)	\varnothing ¹ (nm)
$P25-Ar$	58	0.02	0.48	0.53	36
$Cr/P25-Ar$	73	0.03	0.37	0.43	23
$Co/P25-Ar$	62	0.02	0.31	0.35	23
$Ni/P25-Ar$	53	0.02	0.34	0.37	28
$Cu/P25-Ar$	63	0.02	0.33	0.38	24
¹ Average pore diameter.					

Table 1. Textural properties of the prepared materials. **Table 1.** Textural properties of the prepared materials.

samples and P25-Ar, nor among the M/P25-Ar samples. Figure S1 (in Supplementary Information) shows the SEM images of P25-Ar and Cu/P25-Ar samples as examples. The SEM-EDX mapping image of Cu/P25-Ar shows the uniform distribution of the Cu species on TiO $_2$ (see Figure S2). SEM analysis did not reveal noticeable morphological differences neither between the M/P25-Ar

image of Cu/P25-Ar shows the uniform distribution of the Cu species on TiO2 (see Figure S2). *3.2. X-ray Diffraction*

Figure 2 shows the XRD patterns obtained for P25-Ar and M/P25-Ar samples. All of them present the characteristic peaks of anatase (2 θ values of 25.3 \degree (101), 37.8 \degree (004), 48.0 \degree (200), 54.5 \degree (105), 55.0° (211), 62.7° (204), 70.4° (116), and 75.2° (220)) and rutile (2θ values of 27.5° (110), 36.1° (101) and 54.4 $°$ (211)) [\[40\]](#page-16-15). No characteristic peaks of transition metal oxides have been identified in the M/P25-Ar XRD patterns. This can be explained by the low metal loading, because the metal oxide particles are highly dispersed, and/or because the metallic ions have been partially introduced into the TiO₂ structure.

Figure 2. X-ray diffraction patterns of the photocatalysts. **Figure 2.** X-ray diffraction patterns of the photocatalysts.

Table [2](#page-6-1) summarizes the average crystallite size of anatase and rutile and, also, the contribution of crystalline and amorphous $TiO₂$ in each photocatalyst.

Photocatalyst	Average Crystal Size (nm)		Crystalline $TiO2$ (wt.%)		
	Anatase	Rutile	Anatase	Rutile	Amorphous TiO ₂ $(\%)$
$P25-Ar$	19	32	77		12
$Cr/P25$ -Ar	19	25	76	13	
$Co/P25-Ar$	17	30	74	14	12
$Ni/P25-Ar$	20	32	76		12
$Cu/P25-Ar$	19	31	76		

Table 2. Crystalline properties determined from XRD patterns. **Table 2.** Crystalline properties determined from XRD patterns.

modify in a relevant way the TiO₂ crystalline structure, the phases distribution or the crystal size of any of the phases. It can be mentioned that the Cr/P25-Ar sample presents the smallest rutile average crystal size, and this could be related to its slightly higher specific surface area (see Table [1\)](#page-5-1). crystal size, and this could be related to its slightly higher specific surface area (see Table 1). From data in Table [2](#page-6-1) it can be observed that the addition of transition metallic species does not

3.3. UV-Vis Diffuse Reflectance Spectroscopy

Figure [3](#page-7-0) compiles the UV-vis diffuse reflectance spectra (DRS) of M/P25-Ar and P25-Ar samples. It can be observed that bare P25-Ar has low absorption in the visible region (>400 nm), while the M/P25 samples show higher absorption in this region. Moreover, the absorption edge for the M/P25-Ar samples is shifted to the visible range (423-436 nm), which results in a decrease in their band gap energy values. The absorption edge of the studied photocatalysts increases as follows: P25-Ar < Ni/P25-Ar < Cu/P25-Ar < Cr/P25-Ar < Co/P25-Ar (see Table 3). The inset in Figure 3 presents a photograph of the prepared samples, showing that they are coloured, which reveals the presence of photograph of the prepared samples, showing that they are coloured, which reveals the presence of the presence

Figure 3. UP-vis diffuse reflection of the prepared photocatalysts.
 Figure 3. Critical and Cu) and catalysts. Inset: photograph of the prepared photocatalysts. **Figure 3.** UV-vis diffuse reflectance spectra for bare P25-Ar and for M/P25-Ar (M = Cr, Co, Ni and Cu)

considering the direct allowed transitions (direct method) and the indirect allowed transitions (indirect method); they are presented in Table [3.](#page-7-1) The data obtained by the different methods show a similar trend, being those calculated by the absorbance and indirect methods quite close, and lower than those calculated by the direct method. Furthermore, it can be observed that the difference in the E_g values of samples $M/P25-Ar$ are not significant, and lower than that of P25-Ar. Eg values of samples M/P25-Ar are not significant, and lower than that of P25-Ar. The band gap (E_g) values of the photocatalysts have been calculated by the absorbance method,

photocatalysts. **Table 3.** Absorption edge wavelength and calculated E_g values for P25-Ar and M/P25-Ar

Photocatalysts	Absorption Edge Wavelength (nm)	E_g ¹ (eV)	E_g^2 $\tilde{\textbf{eV}}$	E_g^3 (eV)
$P25-Ar$	403	3.08	3.53	3.11
$Cr/P25$ -Ar	436	2.85	3.40	2.86
$Co/P25-Ar$	436	2.84	3.40	2.88
$Ni/P25-Ar$	422	2.93	3.41	2.95
$Cu/P25-Ar$	426	2.91	3.39	2.92

¹ Absorbance method. ² Direct allowed transitions (direct method). ³ Indirect allowed transitions (indirect method).

As reported in the literature [\[19,](#page-15-16)[20,](#page-15-17)[27\]](#page-16-3), the presence of transition metallic species, even in samples prepared by impregnation, might introduce new intra band gap states in the TiO₂ structure, which to two phenomena: (i) these new energy levels would be located below the lower conduction band edge of TiO₂, decreasing the TiO₂ band gap and shifting the absorption band to the visible part of the spectrum; and (ii) these energy levels would also act as e^- traps, increasing the h⁺ lifetime and decreasing the e^{-}/h^{+} recombination rate $[1/1]$ decreasing the e[−]/h⁺ recombination rate [\[14\]](#page-15-12). could enhance their photoefficiency in the visible region. This behaviour can be mainly attributed

3.4. XPS Materials **2018**, *11*, x FOR PEER REVIEW 9 of 19

P25-Ar and the M/P25-Ar catalysts were analysed by XPS in order to study their surface P25-Ar and the M/P25-Ar catalysts were analysed by XPS in order to study their surface composition and the oxidation state of the metal cations. Figure 4a,b show, respectively, the obtained composition and the oxidation state of the metal cations. Figure [4a,](#page-8-0)b show, respectively, the obtained Ti 2 $p_{3/2}$ and O 1s XPS spectra. The binding energy of Ti 2 $p_{3/2}$, about 458.6 eV (Figure [4a](#page-8-0)), indicates the presence of titanium in an octahedral anatase network [41], which is consistent with Ti (IV). In the case presence of titanium in an octahedral anatase networ[k \[4](#page-16-16)1], which is consistent with Ti (IV). In the of the Cr/P25-Ar sample, the B.E. is slightly lower than in the rest of the samples, which according to the literature can be attributted to the presence of some Ti (III) o[n th](#page-16-17)e surface $[42]$.

The O 1s spectra (Figure 4b) show two contributions, one at about 529.8 eV, ascribed to the lattice The O 1s spectra (Figure [4b](#page-8-0)) show two contributions, one at about 529.8 eV, ascribed to the lattice O⁻² in crystalline TiO₂ [4[3\],](#page-16-18) and a second one at 531.5 eV, that can be due to O⁻ ions located in defect sites related to grain boundaries $[43]$. It can be observed that the intensity of the peak at 531.5 eV is slighlty higher in Cr/P25-Ar, which is in agreement with the larger amount of oxygen vacancies slighlty higher in Cr/P25-Ar, which is in agreement with the larger amount of oxygen vacancies proposed before. The differences between the Cr/P25-Ar sample and the rest of M/P25-Ar catalysts proposed before. The differences between the Cr/P25-Ar sample and the rest of M/P25-Ar catalysts revealed by XPS are consistent with the previoulsy commented XRD and BET results. There is not a revealed by XPS are consistent with the previoulsy commented XRD and BET results. There is not a signal due to oxygen bonded to metallic species, which should appear at about 530 eV [41], probably signal due to oxygen bonded to metallic species, which should appear at about 530 eV [[41\],](#page-16-16) probably due to the low metallic content present in the catalysts. due to the low metallic content present in the catalysts.

Figure 4. Deconvoluted XPS spectra for: (a) Ti $2p_{3/2}$ and (b) O 1s in P25-Ar and in the photocatalysts. M/P25-Ar photocatalysts.

Figure [5 s](#page-9-0)hows the Cr $2p_{3/2}$, Co $2p_{3/2}$, Ni $2p_{3/2}$ and Cu $2p_{3/2}$ deconvoluted spectra obtained.

Figure 5. XPS deconvoluted spectra of: (**a**) Cr 2p_{3/2} in Cr/P25-Ar, (**b**) Co 2p_{3/2} in Co/P25-Ar, (**c**) Ni in Ni/P25-Ar and (**d**) Cu 2p3/2 in Cu/P25-Ar. 2p3/2 in Ni/P25-Ar and (**d**) Cu 2p3/2 in Cu/P25-Ar.

The XPS spectra of Cr 2 $p_{3/2}$ in Cr/P25-Ar (Figure [5a](#page-9-0)) can be deconvoluted into two contributions: at 576.6 eV, assigned to Cr (III) species, and at 578.6 eV, which is attributed to Cr (VI) [44]. The Co at 576.6 eV, assigned to Cr (III) species, and at 578.6 eV, which is attributed to Cr (VI) [\[44\]](#page-16-19). The Co $2p_{3/2}$ XPS spectra (Figure [5b](#page-9-0)) shows a main asymmetric peak located at 780.7 eV and a satellite peak positioned at 786.5 eV. These patterns reveal that $Co/P25$ -Ar sample only contains Co (II) [\[37\]](#page-16-12). The Ni 2 $p_{3/2}$ spectra (Figure [5c](#page-9-0)) shows a main peak at 855.6 eV, ascribed to Ni (II) in an oxygen environment [45], a minor peak at 857.3 eV that corresponds to Ni (III) species [46], and a satellite environment [\[45\]](#page-16-20), a minor peak at 857.3 eV that corresponds to Ni (III) species [\[46\]](#page-17-0), and a satellite peak, at 861.3 eV, that supports the presence of divalent nickel [\[45\]](#page-16-20). Finally, the Cu 2 $\rm p_{3/2}$ XPS spectrum (Figure [5d](#page-9-0)) shows two contributions at 932.4 eV and 934.0 eV, that can be assigned, respectively, to (Figure 5d) shows two contributions at 932.4 eV and 934.0 eV, that can be assigned, respectively, to Cu (I) and Cu (II) $[43,47]$ $[43,47]$. The Cu (II) state also leads to the shake-up satellite peaks at 940.8 and eV [47]. 943.6 eV [\[47\]](#page-17-1).

Table [4](#page-10-0) shows the binding energy values and the identified oxidation states for the metallic Table 4 shows the binding energy values and the identified oxidation states for the metallic species, with the estimated proportions, in each sample. It can be mentioned that, in some cases, the species, with the estimated proportions, in each sample. It can be mentioned that, in some cases, the metallic species show oxidation states different from those of the metallic ions in the salts used for metallic species show oxidation states different from those of the metallic ions in the salts used for impregnation, which can be explained by the interaction of the metallic precursor with TiO_2 , being copper species the most affected by this interaction. copper species the most affected by this interaction.

Photocatalyst	Binding Energy (eV)			Indentified Metal	Proportion $\frac{1}{\sqrt{2}}$
	$Ti 2p_{3/2}$	O _{1s}	$M 2p_{3/2}$ Oxidation States		
$P25-Ar$	458.6	529.8			-
$Cr/P25$ -Ar			576.6	Cr (III)	74
	458.4	529.7	578.6	Cr (VI)	26
$Co/P25-Ar$	458.6	529.9	780.7	Co (II)	100
$Ni/P25-Ar$			855.7	Ni (II)	66
	458.6	529.8	857.3	Ni (III)	34
$Cu/P25-Ar$	458.7		932.4	Cu (I)	86
		529.9	934.1	Cu (II)	14

Table 4. Binding energies and predominant oxidation states of the metal transition ions, determined from YPS spectra from XPS spectra. **Proportion1 Ti 2p3/2 O 1s M 2p3/2 (%)**

 1 Data obtained from the characteristic XPS peak areas for each oxidation state of the metallic species.

The metal content on the catalyst's surface has been calculated from XPS data and a value around 3 wt.% was obtained for all the M/P25-Ar samples. Considering that the nominal (and very likely actual) metal loading is 1 wt.%, this reveals that the samples show a certain surface enrichment in metallic species on the more external surface of the catalyst. *3*.*5*. *Photocatalytic Activity*

3.5. Photocatalytic Activity

The results obtained in the study of the acetic acid photodecomposition in the liquid phase and the photooxidation of propene at low concentration in the gas phase are presented and discussed next.

3.5.1. Photocatalytic Decomposition of Acetic Acid 3.5.1. Photocatalytic Decomposition of Acetic Acid

Figure [6 p](#page-10-1)lots the amount of CH_4 , CO_2 and H_2 (in mmol per mol of acetic acid (HAc)) generated during 12 h in the different catalytic tests performed. Because the H_2 production is very low, the comments on the obtained results will be based on the production of CH_4 and CO_2 . It can be observed that the produced amounts of CH_4 and CO_2 by photolysis are small compared to those produced in the presence of the photocatalysts. the presence of the photocatalysts.

Figure 6. CH₄, CO₂ and H₂ produced after 12 h with the prepared photocatalysts and in the absence of a photocatalyst.

Methane formation during the photocatalytic decomposition of HAc is supposed to follow the so-called photo-Kolbe reaction (Equation (9)),

$$
CH_3COOH + h^+ \rightarrow CH_3^- + CO_2 + H^+ \tag{9}
$$

the main products of which, CH_4 and CO_2 [\[7](#page-15-5)[,48](#page-17-2)[–50\]](#page-17-3), are ideally produced in CH_4/CO_2 molar ratio equal to 1. However, it is interesting to note that in the obtained results, the amount of $CO₂$ produced is higher than the amount of CH_4 , being the CH_4/CO_2 ratio lower than one for all the tested catalysts. This means that some other reactions, besides the photo-Kolbe one, have to be taken into consideration. Mozia, et al. [\[48\]](#page-17-2) suggested that the excess of $CO₂$ could be originated by the oxidation of acetic acid with photogenerated O_2 , as shown in Equation (10):

$$
CH_3COOH + 2 O_2 \rightarrow 2 CO_2 + 2 H_2O \tag{10}
$$

Thus, it could be assumed that both the photo-Kolbe (Equation (9)) and the oxidation reaction (Equation (10)) occur simultaneously in the experiments carried out. If both reactions would occur up to the same extent, the theorethical CH_4/CO_2 ratio would be 0.30. However, the measured $CH₄/CO₂$ ratio ranges between 0.47 and 0.77, meaning that with the tested catalysts and conditions, the photo-Kolbe reaction is predominant, prevaling acetic acid degradation over its oxidation.

Figure [6](#page-10-1) shows that M/P25-Ar photocatalysts produce more CH_4 and CO_2 than P25-Ar, confirming a clearly positive effect of the metal. Data of Figure [6](#page-10-1) also indicate that the photocatalytic activity is significantly influenced by the nature of the added metal. Concerning the methane output, the photocatalytic efficiency of the M/P25-Ar samples follows the trend: $Cu/P25-Ar >> Ni/P25-Ar >$ $Cr/P25-Ar \approx Co/P25-Ar$. The acetic acid conversion varies between 0.01% and 0.13% for all of them, which corresponds, respectively, to 0.10 and 1.28 mmol of acetic acid converted per gram of photocatalyst in 12 h.

3.5.2. Propene Oxidation

Figure [7](#page-12-0) shows the propene conversion values obtained with the five prepared photocatalysts and in the blank experiment, tested at 30 and 60 mL/min. The mass scan carried out to determine oxidation compounds in the outlet stream reveals that $CO₂$ is the only oxidation product. Also, quantification of the produced $CO₂$ has allowed to perform a carbon balance which confirms that total mineralization of propene takes place, according to the following reaction (Equation (11)), and in agreement with the literature [\[8](#page-15-6)[,51\]](#page-17-4).

$$
2 C_3 H_6 + 9 O_2 \rightarrow 6 CO_2 + 6 H_2 O \tag{11}
$$

Photodegradation of propene by photolysis is negligible compared with the photocatalytic oxidation. The activity order of the investigated photocatalysts follows the same trend in both flow-rates tested (P25-Ar > Cu/P25-Ar > Ni/P25-Ar > Cr/P25-Ar \approx Co/P25-Ar), which can be considered as a proof of reproducibility. As expected, propene conversion is higher when the used flow is 30 mL/min.

Figure 7. Propene conversion (at 30 and 60 mL/min) obtained without catalyst, and with P25-Ar and **Figure 7.** Propene conversion (at 30 and 60 mL/min) obtained without catalyst, and with P25-Ar and M/P25-Ar photocatalysts. M/P25-Ar photocatalysts.

4. Discussion 4. Discussion

As mentioned above, the photocatalytic activity of the M/P25-Ar samples in the two studied As mentioned above, the photocatalytic activity of the M/P25-Ar samples in the two studied reactions varies with M, and the samples can be ordered according to their activity following a similar reactions of the samples can be ordered according to their activity following a similar trend in both reaction media: $Cu/P25-Ar > Ni/P25-Ar > Cr/P25-Ar \approx Co/P25-Ar$. Because of the similarity in the textural properties (Table 1), crystallinity and TiO2 phase distribution (Table 2) of the similarity in the textural properties (Table [1\)](#page-5-1), crystallinity and TiO² phase distribution (Table [2\)](#page-6-1) of the investigated photocatalysts, it can be concluded that the differences in activity found are not related investigated photocatalysts, it can be concluded that the differences in activity found are not related with the mentioned properties. Thus, it must be considered that in both reactions the photocatalytic with the mentioned properties. Thus, it must be considered that in both reactions the photocatalytic activity depends on the nature of the incorporated metallic species. As the metals have different activity depends on the nature of the incorporated metallic species. As the metals have different electrochemical properties, their ability to influence the photocatalytic behaviour of titania can be electrochemical properties, their ability to influence the photocatalytic behaviour of titania can be different. The ionic radius and the work function are relevant parameters that would determine their intervalsed the material control to the intervalsed the material control to the material control to the material control behaviour. Table 5 lists the ionic radii [52] and the work functions [53] of each M species identified behaviour. Table [5](#page-13-0) lists the ionic radii [\[52\]](#page-17-5) and the work functions [\[53\]](#page-17-6) of each M species identified by XPS in the M/P25-Ar samples.

The ionic radius of Cr (III) is similar to that of Ti (IV), and this would let Cr (III) replace Ti (IV) $\overline{T} \odot 1$ iii $\overline{1}$ is similar to that of Ti (IV), and this would let Cr (III) replace Ti (IV) in the TiO₂ lattice. Such a substitution in M/TiO₂ samples, has been previously reported in samples $\frac{1}{100}$ also prepared by impregnation and heat-treated at about 500 °C, [\[27\]](#page-16-3). Moreover, the ionic radius of G (\overline{G}) and \overline{G} (\overline{G}) and \overline{G} (\overline{G}) and \overline{G}) and \overline{G} of Cr (VI) is smaller than that of Ti (IV), so Cr (VI) species could also be incorporated in the TiO₂ structure. The incorporation of Cr species in the TiO₂ lattice might introduce some defects and oxygen in the TiO₂ lattice might introduce some defects and oxygen vacancies (that could explain the slightly different properties of Cr/P 25-Ar catalyst, revealed by N₂ adsorption, XRD and XPS). All the other metals must be located either in the network interstices or adsorption, XRD and XPS). All the other metals must be located either in the network interstices or as aggregates on the TiO₂ surface. The latter is the most plausible option since no significant changes in in the crystalline structure of these M/P25-Ar samples compared to P25-Ar have been observed. the crystalline structure of these M/P25-Ar samples compared to P25-Ar have been observed.

As indicated in the introduction, electron transfer can occur from lower work function materials As indicated in the introduction, electron transfer can occur from lower work function materials to higher work function ones. In the case of the M/P25-Ar samples, the metal work function must be the Minimum ones. In the case of the M/P25-Ar samples, the metal work function must be higher than that of TiO₂ to achieve an efficient photocatalytic behaviour. The metallic species can, by the metallic species can, thus, behave as efficient traps for photogenerated electrons, preventing the e[−]/h⁺ recombination and, consequently, improving the photocatalytic activity [54]. consequently, improving the photocatalytic activity [\[54\]](#page-17-7).

Identified Metal Oxidation States ¹	Ionic Radius ² (Å)	Work Function 3 (eV)
Ti (IV)	0.75	5.4 ± 0.2
Cr (III)	0.76	5.0 ± 0.2
Cr (VI)	0.58	6.8 ± 0.2
Co (II)	0.88	4.6 ± 0.2
Ni (II)	0.83	6.3 ± 0.2
Ni (III)	0.74	4.9 ± 0.1
Cu (I)	0.91	4.9 ± 0.1
Cu (II)	0.87	5.9 ± 0.1

Table 5. Ionic radii and work function values for the predominant oxidation states of the transition metal ions in the photocatalysts.

¹ Data obtained from XPS. ² Obtained from reference [\[52\]](#page-17-5). ³ Obtained from reference [\[53\]](#page-17-6).

The activity results obtained show that $Cr/P25$ -Ar and $Co/P25$ -Ar are the samples with the lowest photocatalytic activity in both reactions. In the case of the Co catalyst, it can be due to the lower work function of Co (II) ions with respect to that of Ti (IV). This would not allow the Co (II) ions to capture electrons in an efficient way [\[19,](#page-15-16)[22,](#page-15-19)[53\]](#page-17-6). As the Cr species present in Cr/P25-Ar have a high work function, the low activity must be related to a different property. As commented before, it seems that the Cr (III) ions partially substitute Ti (IV) in Cr/P25-Ar and probably because of that, the catalyst loses activity. In contrast, Ni/P25-Ar and Cu/P25-Ar show the highest performances in both reactions. This can be explained considering that these samples contain, respectively, Ni (II) and Cu (II) ions, that have higher work function than Ti (IV). Another advantage of copper is that, as reported in the literature $[2,7,55,56]$ $[2,7,55,56]$ $[2,7,55,56]$ $[2,7,55,56]$, the conduction band (CB) edges of Cu₂O and CuO species are less negative than that of $TiO₂$, while the valence band (VB) edges of Cu₂O and CuO are less positive than that of TiO2. This implies that in the Cu-P25-Ar sample, Cu2O and CuO introduce energy levels close to both the CB and VB of TiO₂, which can act as effective e⁻ and h⁺ traps. Besides, the coexistence of the two Cu species generates electron transfer in cascade, from more positive energy values to less positive ones. This way of electron transfer may result in a better e⁻/h⁺ separation, associated to higher photoactivity.

In the photodecomposition of acetic acid, Cu/P25-Ar is clearly more active than the rest of M/P25-Ar photocatalysts. Besides, as mentioned above, all the M/P25-Ar samples are more active than P25-Ar. On the other hand, in the gas phase oxidation of propene, Ni/P25-Ar and Cu/P25-Ar show a close behavior, being bare P25-Ar more active than the M/P25-Ar photocatalysts. This indicates that the course of the two reactions seems to depend differently on the catalysts properties. Previous works compared the photoactivity of $TiO₂$ systems in abatement of pollutants in liquid and gas phase and found that the photocatalysts' behaviour was different in the two phases [\[57,](#page-17-10)[58\]](#page-17-11). This can be related with either the own reaction mechanism, or to differences in experimental conditions, like the reaction media (liquid or gas phase), the reactor configuration or the substrate concentration. For example, as indicated in the literature [\[59](#page-17-12)[–62\]](#page-17-13), in the liquid phase, the solvent (usually water) helps to remove reaction intermediates and products from the catalyst surface, avoiding or hindering deactivation, whereas in fixed-bed reactors like the one used for propene oxidation, the light distribution becomes a limiting factor [\[63\]](#page-17-14). Probably, the gas phase photooxidation of propene is a rapid process occurring mainly on the titania surface. It seems that when the metallic species are present, the reaction rate decreases (i.e. by decrease of $TiO₂$ exposed surface) and this is also influenced, as mentioned above, by the metal's nature. In the liquid phase reaction, the diffusion limitations likely make the whole reaction slower and, thus, the pathway in which the supported metallic species are involved is not limiting the reaction. In this case, the positive effect of the metallic species as electron traps prevails.

5. Conclusions

In the present work, a series of $M/TiO₂$ photocatalysts (M = Cr, Co, Ni, Cu) were prepared by impregnation of the commercial P25 titania followed by heat treatment in Ar at 500 ◦C.

The characterization results have shown that the textural properties of the M/P25-Ar photocatalysts are very close to those of bare P25-Ar, and only sample Cr/P25-Ar shows a slightly higher surface area. The crystalline structure of P25-Ar is, as well, almost not affected by the metal incorporation, and XRD patterns due to metal oxides have not been detected, which allows us to conclude that the metallic species are highly dispersed on the $TiO₂$ surface. The crystal size of anatase and rutile was very similar in all M/P25-Ar samples, although in Cr/P25-Ar the average crystal size of rutile was the lowest. The band gap energies of the M/P25-Ar samples were a bit narrower than that for bare P25-Ar. The XPS analysis indicated that excepting Co, the rest of the metals are present in more than one oxidation state. The Ti 2p and O 1s spectra of Cr/P25-Ar show, compared to the rest of samples, a shift to lower binding energies, which suggests that Cr could have partially been incorporated in the $TiO₂$ structure in this sample, leading to some defects. This is possible because the atomic radii of Cr (III) and Cr (VI) ions are smaller than that of Ti (IV), and it could explain the differences in the textural properties and crystal size of Cr/P25 compared to the rest of M/P25 photocatalysts.

The photoactivity of the M/P25-Ar photocatalysts follows the same trend in the two studied reactions: Cu/P25-Ar > Ni/P25-Ar > Cr/P25-Ar \approx Co/P25-Ar. To explain the differences between catalysts, the work functions and the ionic radii of the incorporated metallic ions have been considered to have a role. It has been concluded that the low activity of Cr/P25-Ar sample is related with the partial incorporation of Cr species in the $TiO₂$ lattice, which results in lattice defects and oxygen vacancies that would facilitate recombination of electrons and holes. The low activity of Co/P25-Ar has been explained by the low work function of Co (II) compared to that of Ti (IV). The Cu/P25-Ar photocatalyst showed the best photocatalytic performance among the M/P25-Ar samples because Cu ions dispersed on $TiO₂$, present as Cu (II) and Cu (I), have suitable energy levels of CB and VB edge to trap e[−] and h⁺ and carry out an effective electronic transfer.

Although the relative activity order of the metal-containing photocatalysts is the same in both reactions, P25-Ar shows better activity than Cu/P25-Ar in propene oxidation, whereas Cu/P25-Ar is more active than P25-Ar for acetic acid degradation. This can be explained considering that the rate-determining steps are significantly different in gas and liquid media. In the gas phase, propene oxidation is a rapid process that happens mainly on the $TiO₂$ surface, and the supported metallic species reduce the exposed titania surface, decreasing the photocatalytic efficiency. In contrast, in the liquid phase, as the diffusion processes make the reaction slower, the positive effect of the metallic species as electron traps can be better identified.

Supplementary Materials: The following are available online at [http://www.mdpi.com/1996-1944/12/1/40/s1,](http://www.mdpi.com/1996-1944/12/1/40/s1) Figure S1: scanning electron microscope (SEM) images of (**a**,**b**) P25-Ar and (**c**,**d**) Cu/P25-Ar samples with: 10 µm scale bar (**a**,**c**) and 1 µm scale bar (**b**,**d**). Figure S2: Scanning electron microscopy with energy dispersive X-ray (SEM-EDX) mapping for Cu/P25-Ar sample: Ti (red) and Cu (green).

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References

1. Hoffmann, M.R.; Martin, S.T.; Choi, W.; Bahnemann, D.W. Environmental Applications of Semiconductor Photocatalysis. *Chem. Rev.* **1995**, *95*, 69–96. [\[CrossRef\]](http://dx.doi.org/10.1021/cr00033a004)

- 2. Litter, M.I. Heterogeneous photocatalysis: Transition metal ions in photocatalytic systems. *Appl. Catal. B Environ.* **1999**, *23*, 89–114. [\[CrossRef\]](http://dx.doi.org/10.1016/S0926-3373(99)00069-7)
- 3. MiarAlipour, S.; Friedmann, D.; Scott, J.; Amal, R. TiO₂/porous adsorbents: Recent advances and novel applications. *J. Hazard. Mater.* **2018**, *341*, 404–423. [\[CrossRef\]](http://dx.doi.org/10.1016/j.jhazmat.2017.07.070) [\[PubMed\]](http://www.ncbi.nlm.nih.gov/pubmed/28806560)
- 4. Betts, L.M.; Dappozze, F.; Guillard, C. Understanding the photocatalytic degradation by P25 TiO₂ of acetic acid and propionic acid in the pursuit of alkane production. *Appl. Catal. A Gen.* **2018**, *554*, 35–43. [\[CrossRef\]](http://dx.doi.org/10.1016/j.apcata.2018.01.011)
- 5. Hassan, M.; Zhao, Y.; Xie, B. Employing TiO₂ photocatalysis to deal with landfill leachate: Current status and development. *Chem. Eng. J.* **2016**, *285*, 264–275. [\[CrossRef\]](http://dx.doi.org/10.1016/j.cej.2015.09.093)
- 6. Ding, C.; Sun, Y.; Lin, Y.; Sun, W.; Liu, H.; Zhu, X.; Dai, Y.; Luo, C. Magnetically separable functionalized TiO² nanotubes: Synthesis, characterization, and photocatalysis. *J. Photochem. Photobiol. A Chem.* **2018**, *356*, 123–131. [\[CrossRef\]](http://dx.doi.org/10.1016/j.jphotochem.2017.12.031)
- 7. Heciak, A.; Morawski, A.W.; Grzmil, B.; Mozia, S. Cu-modified TiO₂ photocatalysts for decomposition of acetic acid with simultaneous formation of C_1-C_3 hydrocarbons and hydrogen. *Appl. Catal. B Environ.* **2013**, *140*, 108–114. [\[CrossRef\]](http://dx.doi.org/10.1016/j.apcatb.2013.03.044)
- 8. Lillo-Ródenas, M.Á.; Bouazza, N.; Berenguer-Murcia, Á.; Linares-Salinas, J.J.; Soto, P.; Linares-Solano, Á. Photocatalytic oxidation of propene at low concentration. *Appl. Catal. B Environ.* **2007**, *71*, 298–309. [\[CrossRef\]](http://dx.doi.org/10.1016/j.apcatb.2006.10.004)
- 9. Bacsik, Z.; McGregor, J.; Mink, J. FTIR analysis of gaseous compounds in the mainstream smoke of regular and light cigarettes. *Food Chem. Toxicol.* **2007**, *45*, 266–271. [\[CrossRef\]](http://dx.doi.org/10.1016/j.fct.2006.08.018)
- 10. Centi, G.; Quadrelli, E.A.; Perathoner, S. Catalysis for CO₂ conversion: A key technology for rapid introduction of renewable energy in the value chain of chemical industries. *Energy Environ. Sci.* **2013**, *6*, 1711–1731. [\[CrossRef\]](http://dx.doi.org/10.1039/c3ee00056g)
- 11. Nakata, K.; Fujishima, A. TiO² photocatalysis: Design and applications. *J. Photochem. Photobiol. C Photochem. Rev.* **2012**, *13*, 169–189. [\[CrossRef\]](http://dx.doi.org/10.1016/j.jphotochemrev.2012.06.001)
- 12. Coronado, J.M.; Fresno, F.; Hernández-Alonso, M.D.; Portela, R. The Keys of Succes: TiO₂ as a Benchmark Photocatalyst. In *Design of Advanced Photocatalytic Materials for Energy and Environmental Applications*; Coronado, J.M., Hernández-Alonso, M.D., Fresno, F., Portela, R., Eds.; Springer: London, UK, 2013; pp. 85–102. ISBN 978-1-4471-5061-9.
- 13. Katsumata, H.; Sada, M.; Nakaoka, Y.; Kaneco, S.; Suzuki, T.; Ohta, K. Photocatalytic degradation of diuron in aqueous solution by platinized TiO² . *J. Hazard. Mater.* **2009**, *171*, 1081–1087. [\[CrossRef\]](http://dx.doi.org/10.1016/j.jhazmat.2009.06.110) [\[PubMed\]](http://www.ncbi.nlm.nih.gov/pubmed/19604634)
- 14. Kerkez-Kuyumcu, Ö.; Kibar, E.; Dayıoğlu, K.; Gedik, F.; Akın, A.N.; Özkara-Aydınoğlu, Ş. A comparative study for removal of different dyes over M/TiO $_2$ (M = Cu, Ni, Co, Fe, Mn and Cr) photocatalysts under visible light irradiation. *J. Photochem. Photobiol. A Chem.* **2015**, *311*, 176–185. [\[CrossRef\]](http://dx.doi.org/10.1016/j.jphotochem.2015.05.037)
- 15. Rauf, M.A.; Meetani, M.A.; Hisaindee, S. An overview on the photocatalytic degradation of azo dyes in the presence of TiO² doped with selective transition metals. *Desalination* **2011**, *276*, 13–27. [\[CrossRef\]](http://dx.doi.org/10.1016/j.desal.2011.03.071)
- 16. Inturi, S.N.R.; Boningari, T.; Suidan, M.; Smirniotis, P.G. Visible-light-induced photodegradation of gas phase acetonitrile using aerosol-made transition metal (V, Cr, Fe, Co, Mn, Mo, Ni, Cu, Y, Ce, and Zr) doped TiO₂. *Appl. Catal. B Environ.* **2014**, *144*, 333–342. [\[CrossRef\]](http://dx.doi.org/10.1016/j.apcatb.2013.07.032)
- 17. Sahu, M.; Biswas, P. Single-step processing of copper-doped titania nanomaterials in a flame aerosol reactor. *Nanoscale Res. Lett.* **2011**, *6*, 441. [\[CrossRef\]](http://dx.doi.org/10.1186/1556-276X-6-441) [\[PubMed\]](http://www.ncbi.nlm.nih.gov/pubmed/21733174)
- 18. Ashby, M.; Messler, R.; Asthana, R.; Furlani, E.; Smallman, R.E.; Ngan, A.H.W.; Crawford, R.J.; Mills, N. *Engineering Materials and Processes Desk Reference*, 1st ed.; Butterworth-Heinemann: Oxford, UK, 2009; ISBN 9780080878393.
- 19. Tayade, R.J.; Kulkarni, R.G.; Jasra, R.V. Transition Metal Ion Impregnated Mesoporous TiO₂ for Photocatalytic Degradation of Organic Contaminants in Water. *Ind. Eng. Chem. Res.* **2006**, *45*, 5231–5238. [\[CrossRef\]](http://dx.doi.org/10.1021/ie051362o)
- 20. Devi, L.G.; Kottam, N.; Murthy, B.N.; Kumar, S.G. Enhanced photocatalytic activity of transition metal ions Mn^{2+} , Ni^{2+} and Zn^{2+} doped polycrystalline titania for the degradation of Aniline Blue under UV/solar light. *J. Mol. Catal. A Chem.* **2010**, *328*, 44–52. [\[CrossRef\]](http://dx.doi.org/10.1016/j.molcata.2010.05.021)
- 21. Halas, S. 100 years of work function. *Mater. Sci.* **2016**, *24*, 951–968.
- 22. Ola, O.; Maroto-Valer, M.M. Review of material design and reactor engineering on $TiO₂$ photocatalysis for CO² reduction. *J. Photochem. Photobiol. C Photochem. Rev.* **2015**, *24*, 16–42. [\[CrossRef\]](http://dx.doi.org/10.1016/j.jphotochemrev.2015.06.001)
- 23. Wang, L.; Egerton, T. The Effect of Transition Metal on the Optical Properties and Photoactivity of Nano-particulate Titanium Dioxide. *J. Mater. Sci. Res.* **2012**, *1*, 19–27. [\[CrossRef\]](http://dx.doi.org/10.5539/jmsr.v1n4p19)
- 24. Choi, W.; Termin, A.; Hoffmann, M.R. The role of metal ion dopants in quantum-sized TiO₂: Correlation between photoreactivity and charge carrier recombination dynamics. *J. Phys. Chem.* **1994**, *98*, 13669–13679. [\[CrossRef\]](http://dx.doi.org/10.1021/j100102a038)
- 25. Al-Azri, Z.H.N.; Chen, W.-T.; Chan, A.; Jovic, V.; Ina, T.; Idriss, H.; Waterhouse, G.I.N. The roles of metal co-catalysts and reaction media in photocatalytic hydrogen production: Performance evaluation of $M/TiO₂$ photocatalysts (M = Pd, Pt, Au) in different alcohol–water mixtures. *J. Catal.* **2015**, *329*, 355–367. [\[CrossRef\]](http://dx.doi.org/10.1016/j.jcat.2015.06.005)
- 26. Lin, H.; Shih, C. Efficient one-pot microwave-assisted hydrothermal synthesis of M (M = Cr, Ni, Cu, Nb) and nitrogen co-doped TiO² for hydrogen production by photocatalytic water splitting. *J. Mol. Catal. A Chem.* **2016**, *411*, 128–137. [\[CrossRef\]](http://dx.doi.org/10.1016/j.molcata.2015.10.026)
- 27. Di Paola, A.; Marcì, G.; Palmisano, L.; Schiavello, M.; Uosaki, K.; Ikeda, S.; Ohtani, B. Preparation of polycrystalline TiO₂ photocatalysts impregnated with various transition metal ions: Characterization and photocatalytic activity for the degradation of 4-nitrophenol. *J. Phys. Chem. B* **2002**, *106*, 637–645. [\[CrossRef\]](http://dx.doi.org/10.1021/jp013074l)
- 28. Devi, L.G.; Kavitha, R. A review on non metal ion doped titania for the photocatalytic degradation of organic pollutants under UV/solar light: Role of photogenerated charge carrier dynamics in enhancing the activity. *Appl. Catal. B Environ.* **2013**, *140–141*, 559–587. [\[CrossRef\]](http://dx.doi.org/10.1016/j.apcatb.2013.04.035)
- 29. Amorós-Pérez, A.; Cano-Casanova, L.; Lillo-Ródenas, M.Á.; Román-Martínez, M.C. Cu/TiO₂ photocatalysts for the conversion of acetic acid into biogas and hydrogen. *Catal. Today* **2016**. [\[CrossRef\]](http://dx.doi.org/10.1016/j.cattod.2016.09.009)
- 30. Brunauer, S.; Emmett, P.H.; Teller, E. Adsorption of Gases in Multimolecular Layers. *J. Am. Chem. Soc.* **1938**, *60*, 309–319. [\[CrossRef\]](http://dx.doi.org/10.1021/ja01269a023)
- 31. Rodriguez-Reinoso, F.; Linares-Solano, A. *Microporous Structure of Activated Carbons as Revealed by Adsorption Methods*; Thrower, P.A., Ed.; Dekker, Marcel: New York, NY, USA, 1989; Volume 21, ISBN 0-8247-7939-8.
- 32. Joyner, L.G.; Barrett, E.P.; Skold, R. The Determination of Pore Volume and Area Distributions in Porous Substances. II. Comparison between Nitrogen Isotherm and Mercury Porosimeter Methods. *J. Am. Ceram. Soc.* **1951**, *73*, 3155–3158. [\[CrossRef\]](http://dx.doi.org/10.1021/ja01151a046)
- 33. Jenkins, R.; Snyder, R.L. *Introduction to X-ray Powder Diffractometry*; John Wiley and Sons: New York, NY, USA, 1996; ISBN 978-0-471-51339-1.
- 34. Jensen, H.; Joensen, K.D.; Jørgensen, J.E.; Pedersen, J.S.; Søgaard, E.G. Characterization of nanosized partly crystalline photocatalysts. *J. Nanopart. Res.* **2004**, *6*, 519–526. [\[CrossRef\]](http://dx.doi.org/10.1007/s11051-004-1714-3)
- 35. Olsen, E.D. *Modern Optical Methods of Analysis*; McGraw-Hill: New York, NY, USA, 1975.
- 36. Valencia, S.; Marin, J.M.; Restrepo, G. Study of the bandgap of synthesized titanium dioxide nanoparticules using the sol-gel method and a hydrothermal treatment. *Open Mater. Sci. J.* **2010**, *4*, 9–14. [\[CrossRef\]](http://dx.doi.org/10.2174/1874088X01004010009)
- 37. Biesinger, M.C.; Payne, B.P.; Grosvenor, A.P.; Lau, L.W.M.; Gerson, A.R.; Smart, R.S.C. Resolving surface chemical states in XPS analysis of first row transition metals, oxides and hydroxides: Cr, Mn, Fe, Co and Ni. *Appl. Surf. Sci.* **2011**, *257*, 2717–2730. [\[CrossRef\]](http://dx.doi.org/10.1016/j.apsusc.2010.10.051)
- 38. Bouazza, N.; Lillo-Ródenas, M.A.; Linares-Solano, A. Enhancement of the photocatalytic activity of pelletized TiO² for the oxidation of propene at low concentration. *Appl. Catal. B Environ.* **2008**, *77*, 284–293. [\[CrossRef\]](http://dx.doi.org/10.1016/j.apcatb.2007.07.027)
- 39. Sing, K.S.W.; Everett, D.H.; Haul, R.A.W.; Moscou, L.; Pierotti, R.A.; Rouquerol, J.; Siemieniewska, T. Reporting physisorption data for gas/solid systems with special reference to the determination of surface area and porosity. *Pure Appl. Chem.* **1985**, *57*, 603–619. [\[CrossRef\]](http://dx.doi.org/10.1351/pac198557040603)
- 40. The International Centre for Diffraction Data. Available online: <http://www.icdd.com/> (accessed on 30 October 2018).
- 41. Lopez, T.; Cuevas, J.L.; Ilharco, L.; Ramírez, P.; Rodríguez-Reinoso, F.; Rodríguez-Castellón, E. XPS characterization and E. Coli DNA degradation using functionalized Cu/TiO₂ nanobiocatalysts. *Mol. Catal.* **2018**, *449*, 62–71. [\[CrossRef\]](http://dx.doi.org/10.1016/j.mcat.2018.02.010)
- 42. Tseng, L.-T.; Luo, X.; Bao, N.; Ding, J.; Li, S.; Yi, J. Structures and properties of transition-metal-doped TiO₂ nanorods. *Mater. Lett.* **2016**, *170*, 142–146. [\[CrossRef\]](http://dx.doi.org/10.1016/j.matlet.2016.02.021)
- 43. Behzad, H.; Ghodsi, F.E.; Peksu, E.; Karaagac, H. The effect of Cu content on structural, optical and photo-electrical properties of sol-gel derived CuxCo³ -xO⁴ thin films. *J. Alloys Compd.* **2018**, *744*, 470–480. [\[CrossRef\]](http://dx.doi.org/10.1016/j.jallcom.2018.02.114)
- 44. López, R.; Gómez, R.; Oros-Ruiz, S. Photophysical and photocatalytic properties of TiO² -Cr sol–gel prepared semiconductors. *Catal. Today* **2011**, *166*, 159–165. [\[CrossRef\]](http://dx.doi.org/10.1016/j.cattod.2011.01.010)
- 45. Dong, Z.; Ding, D.; Li, T.; Ning, C. Ni-doped TiO² nanotubes photoanode for enhanced photoelectrochemical water splitting. *Appl. Surf. Sci.* **2018**, *443*, 321–328. [\[CrossRef\]](http://dx.doi.org/10.1016/j.apsusc.2018.03.031)
- 46. Fu, Z.; Hu, J.; Hu, W.; Yang, S.; Luo, Y. Quantitative analysis of Ni^{2+}/Ni^{3+} in Li[NixMnyCoz]O₂ cathode materials: Non-linear least-squares fitting of XPS spectra. *Appl. Surf. Sci.* **2018**, *441*, 1048–1056. [\[CrossRef\]](http://dx.doi.org/10.1016/j.apsusc.2018.02.114)
- 47. Biesinger, M.C.; Lau, L.W.M.; Gerson, A.R.; Smart, R.S.C. Resolving surface chemical states in XPS analysis of first row transition metals, oxides and hydroxides: Sc, Ti, V, Cu and Zn. *Appl. Surf. Sci.* **2010**, *257*, 887–898. [\[CrossRef\]](http://dx.doi.org/10.1016/j.apsusc.2010.07.086)
- 48. Mozia, S.; Heciak, A.; Morawski, A.W. Photocatalytic acetic acid decomposition leading to the production of hydrocarbons and hydrogen on Fe-modified TiO² . *Catal. Today* **2011**, *161*, 189–195. [\[CrossRef\]](http://dx.doi.org/10.1016/j.cattod.2010.09.020)
- 49. Zhang, H.; Zhou, P.; Ji, H.; Ma, W.; Chen, C.; Zhao, J. Enhancement of photocatalytic decarboxylation on TiO² by water-induced change in adsorption-mode. *Appl. Catal. B Environ.* **2018**, *224*, 376–382. [\[CrossRef\]](http://dx.doi.org/10.1016/j.apcatb.2017.10.020)
- 50. Ngo, S.; Betts, L.M.; Dappozze, F.; Ponczek, M.; George, C.; Guillard, C. Kinetics and mechanism of the photocatalytic degradation of acetic acid in absence or presence of O² . *J. Photochem. Photobiol. A Chem.* **2017**, *339*, 80–88. [\[CrossRef\]](http://dx.doi.org/10.1016/j.jphotochem.2017.02.019)
- 51. Ouzzine, M.; Lillo-Ródenas, M.A.; Linares-Solano, A. Photocatalytic oxidation of propene in gas phase at low concentration by optimized TiO² nanoparticles. *Appl. Catal. B Environ.* **2013**, *134–135*, 333–343. [\[CrossRef\]](http://dx.doi.org/10.1016/j.apcatb.2013.01.035)
- 52. Shannon, R.D. Revised effective ionic radii and systematic studies of interatomic distances in halides and chalcogenides. *Acta Crystallogr. Sect. A* **1976**, *32*, 751–767. [\[CrossRef\]](http://dx.doi.org/10.1107/S0567739476001551)
- 53. Greiner, M.T.; Chai, L.; Helander, M.G.; Tang, W.M.; Lu, Z.H. Transition metal oxide work functions: The influence of cation oxidation state and oxygen vacancies. *Adv. Funct. Mater.* **2012**, *22*, 4557–4568. [\[CrossRef\]](http://dx.doi.org/10.1002/adfm.201200615)
- 54. Hinojosa-Reyes, M.; Camposeco-Solís, R.; Zanella, R.; Rodríguez González, V. Hydrogen production by tailoring the brookite and Cu2O ratio of sol-gel Cu-TiO² photocatalysts. *Chemosphere* **2017**, *184*, 992–1002. [\[CrossRef\]](http://dx.doi.org/10.1016/j.chemosphere.2017.06.066)
- 55. Ni, M.; Leung, M.K.H.; Leung, D.Y.C.; Sumathy, K. A review and recent developments in photocatalytic water-splitting using TiO₂ for hydrogen production. *Renew. Sustain. Energy Rev.* **2007**, 11, 401–425. [\[CrossRef\]](http://dx.doi.org/10.1016/j.rser.2005.01.009)
- 56. Helaïli, N.; Bessekhouad, Y.; Bouguelia, A.; Trari, M. Visible light degradation of Orange II using xCuyOz/TiO² heterojunctions. *J. Hazard. Mater.* **2009**, *168*, 484–492. [\[CrossRef\]](http://dx.doi.org/10.1016/j.jhazmat.2009.02.066)
- 57. Rimoldi, L.; Giordana, A.; Cerrato, G.; Falletta, E.; Meroni, D. Insights on the photocatalytic degradation processes supported by TiO2/WO³ systems. The case of ethanol and tetracycline. *Catal. Today* **2018**. [\[CrossRef\]](http://dx.doi.org/10.1016/j.cattod.2018.11.035)
- 58. Sabatini, V.; Rimoldi, L.; Tripaldi, L.; Meroni, D.; Farina, H.; Ortenzi, M.; Ardizzone, S. TiO₂-SiO₂-PMMA Terpolymer Floating Device for the Photocatalytic Remediation of Water and Gas Phase Pollutants. *Catalysts* **2018**, *8*, 568. [\[CrossRef\]](http://dx.doi.org/10.3390/catal8110568)
- 59. Piera, E.; Ayllón, J.; Doménech, X.; Peral, J. TiO² deactivation during gas-phase photocatalytic oxidation of ethanol. *Catal. Today* **2002**, *76*, 259–270. [\[CrossRef\]](http://dx.doi.org/10.1016/S0920-5861(02)00224-9)
- 60. Peral, J.; Domènech, X.; Ollis, D.F. Review Heterogeneous Photocatal y sis for Purification, Decontamination and Deodorization of Air. *J. Chem. Technol. Biotechnol.* **1997**, *70*, 117–140. [\[CrossRef\]](http://dx.doi.org/10.1002/(SICI)1097-4660(199710)70:2<117::AID-JCTB746>3.0.CO;2-F)
- 61. Alberici, R.M.; Jardim, W.F. Photocatalytic destruction of VOCs in the gas-phase using titanium dioxide. *Appl. Catal. B Environ.* **1997**, *14*, 55–68. [\[CrossRef\]](http://dx.doi.org/10.1016/S0926-3373(97)00012-X)
- 62. Dibble, L.A.; Raupp, G.B. Fluidized-Bed Photocatalytic Oxidation of Trichloroethylene in Contaminated Air Streams. *Environ. Sci. Technol.* **1992**, *26*, 492–495. [\[CrossRef\]](http://dx.doi.org/10.1021/es00027a006)
- 63. Howe, R.F. Recent Developments in Photocatalysis. *Dev. Chem. Eng. Miner. Process.* **1998**, *6*, 55–84. [\[CrossRef\]](http://dx.doi.org/10.1002/apj.5500060105)

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