



1 TITLE: Influence of habitat on nest location and reproductive output of Montagu's
2 Harriers breeding in natural vegetation

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4 Running title: Habitat and breeding output of Montagu's Harrier

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25 **ABSTRACT**

26 We examined distribution and breeding success of semi-colonial Montagu's Harriers in
 27 relation to habitat in Castellon province (Eastern Spain). Breeding areas used by harriers
 28 at 1 km² scale were characterized by having intermediate percentages of scrub cover,
 29 their nesting habitat, and also had intermediate coverage of herbaceous crops and non-
 30 irrigated orchards. Out of all habitat variables considered, only percentage of
 31 herbaceous crops within 500 metres from individual nests had a positive and significant
 32 effect on breeding output of the species, suggesting that this habitat may be efficiently
 33 used by harriers to forage. Breeding output was also related to laying date and number
 34 of breeding neighbours within 500 metres around nests, with pairs laying later and
 35 having a higher number of breeding neighbours showing lower fledged brood sizes.
 36 Number of neighbours (but not laying date) was positively related to scrub cover within
 37 500 metres and to cover of herbaceous crops within 2000 metres. Conservation actions
 38 for Montagu's Harrier in the study area should be aimed to preserve areas of scrub with
 39 nearby presence of herbaceous crops or natural grasslands. However, habitat
 40 improvement for semi-colonial species such as Montagu's Harrier may not result in a
 41 change of species distribution area, and good habitat areas may remain unoccupied, as
 42 social factors like presence of conspecifics play an important role in breeding area
 43 selection for these species.

44
 45 **KEYWORDS:** *Circus pygargus*; colonial species; conservation; natural vegetation;
 46 Spain.

49 **Zusammenfassung**

50 **Der Einfluss von Habitat auf den Niststandort und Bruterfolg von Wiesenweihen** 51 **in natürlicher Vegetation**

52
 53 Wir untersuchten die Verteilung und den Bruterfolg von halb-kolonialen Wiesenweihen
 54 in Bezug zum Habitat in der Provinz Castellón im Osten Spaniens. Das Brutgebiet der
 55 Wiesenweihen war, auf einer 1 km²-Skala betrachtet, charakterisiert durch mittlere
 56 Bedeckung mit Buschwerk. Das Nesthabitat war zusätzlich bestimmt durch mittlere
 57 Bedeckung mit krautigen Pflanzen und nicht-bewässerten Obstgärten. Von allen
 58 Habitatvariablen, die wir betrachtet hatten, hatte nur der Prozentsatz der Bedeckung mit

59 Krautpflanzen innerhalb von 500 m um die individuellen Nester einen positiven und
60 signifikanten Effekt auf den Bruterfolg der Art. Dies lässt vermuten, dass dieses Habitat
61 effizient von Wiesenweihen zum Furagieren genutzt wird. Der Bruterfolg war auch mit
62 dem Legezeitpunkt korreliert, sowie mit der Anzahl von brütenden Nachbarn innerhalb
63 eines Umkreises von 500 m um das Nest. Spät legende Paare und die mit mehr
64 Nachbarn hatten einen niedrigeren Bruterfolg. Die Anzahl der Nachbarn (allerdings
65 nicht der Legezeitpunkt) war positiv korreliert mit dem Grad der Bedeckung mit
66 Buschwerk im 500 m Umkreis und mit dem Bedeckungsgrad der Krautpflanzen im
67 Umkreis von 2000 m um das Nest. Schutzmaßnahmen für Wiesenweihen sollten auf
68 Gebiete mit Buschbedeckung und Präsenz von Krautpflanzen oder natürlichen
69 Wiesengebieten abzielen. Jedoch sollte eine solche Habitatverbesserung für halb-
70 kolonial brütende Arten, wie die Wiesenweihe, nicht zu einer Veränderung der
71 Verbreitung führen. Auch können Gebiete mit gutem Habitat trotzdem nicht besiedelt
72 werden, da soziale Faktoren wie die Präsenz von Artgenossen eine wichtige Rolle für
73 die Wahl des Nistplatzes bei solchen Arten spielen.

74

75 INTRODUCTION

76
77 Declines of many bird populations have been linked to a reduction or degradation of
78 their preferred habitats (Browne et al. 2004; Fernández et al. 2004; Julliard et al. 2004;
79 Robinson et al. 2001). Both nesting and foraging habitats may play an important role in
80 limiting bird population numbers or distribution (e.g. Newton 1998). Thus, conservation
81 of bird species is frequently based on protection of their habitats, except in those cases
82 when direct intervention is necessary (e.g. rescue of nestlings of a given species,
83 supplementary food campaigns, captive breeding and release of birds in small
84 populations, etc.; Cade and Temple 1995; Oro et al. 2008). For example, protected areas
85 for birds (such as the Special Protection Areas - or SPAs □ designated under the EC
86 Birds Directive) usually consider financial incentives for sustainable management of the
87 land.

88
89 Like many birds, raptors are usually highly selective with respect to their habitats,
90 especially regarding the availability of suitable nesting areas, although foraging habitats
91 may also have an important effect at the time of choosing a site during the breeding
92 season (Newton 1998). Breeding habitat (which include both nesting and foraging
93 habitats) may limit species productivity or distribution (e.g. Benton et al. 2002; Soh et
94 al. 2006; Suárez et al. 2000). In these cases, increasing availability or suitability of
95 preferred habitats (e.g. restoring nesting habitats or increasing the availability of
96 foraging habitats) may potentially lead to increasing population sizes (Carrete et al.
97 2002; Hiraldo et al. 1996). Understanding the strength of the relationships between
98 habitat and species distribution or breeding success may be important to manage
99 protected areas and to predict how changes in habitat may influence population
100 dynamics, and thus contribute to the development of successful conservation
101 programmes (López-López et al. 2006, 2007; Suárez et al. 2000; Tapia et al. 2004;
102 Wilson et al. 2009).

103
104 However, the majority of raptor studies have been carried out on territorial species (but
105 see García-Ripollés et al. 2005; Poirazidis et al. 2004; Sergio et al. 2003, for semi-
106 colonial and colonial raptor species). Colonial or semi-colonial species may be atypical,
107 because habitat selection may play a relatively smaller role in breeding spatial
108 distribution for these species (e.g. Cornulier 2005; Cornulier and Bretagnolle 2006),

109 whereas factors like conspecific attraction may be more important (Cornulier 2005;
110 Sergio and Penteriani 2005, but see also Sergio et al. 2007). For these species,
111 increasing the availability of preferred habitats might be inefficient to ensure the
112 occupancy of given areas (see e.g. Reed and Dobson 1993).

113

114 The Montagu's Harrier (*Circus pygargus*) is a semi-colonial ground-nesting Palaearctic
115 raptor (Cramp and Simmons 1980). The species is considered vulnerable in France and
116 the Iberian Peninsula (Blanco and González 1992; Salamolard et al. 1999; SNPRCN
117 1990), which are the strongholds of its western European populations. In Western
118 Europe, this species mainly builds nests within cereal crops (Arroyo et al. 2002), but
119 some populations nest in natural vegetation (Cramp and Simmons 1980). The
120 importance of protecting populations breeding in natural vegetation has been
121 highlighted (Arroyo et al. 2002; Limiñana et al. 2006a), but most recent conservation
122 measures have been directed towards populations breeding in agricultural habitats. One
123 population nesting in natural vegetation is located in inland Castellon province in
124 eastern Spain (Limiñana et al. 2006a). This population has increased exponentially from
125 three pairs in early 80s to nearly 150 pairs in 2007, although population growth has
126 slowed down since 2002 (Limiñana et al. 2006a; Soutullo et al. 2006). Harrier breeding
127 sites in Castellon face an uncertain future, as the area is currently subject to social and
128 commercial developments, such as the recent building of an airport. A better
129 understanding of the relationship between breeding habitat availability, harrier
130 distribution and breeding performance would enable more effective conservation of
131 Montagu's Harrier in this area.

132 In this paper, we first examine the relationship between habitat and nest occurrence, to
133 assess the habitats preferred for breeding within the study area. Secondly, we examine
134 the relationship between habitat, timing of breeding, number of neighbours and
135 breeding output. We discuss the importance of habitat in explaining breeding
136 distribution or success, and its implications for conservation in this semi-colonial
137 species.

138

139 **METHODS**

140 **Study site and species**

141 The study was carried out in Castellon province (Eastern Spain, Fig. 1) where the
142 species breeds in the inland corridors between the mountain ranges. Montagu's Harriers

143 first bred in the study area in early 1980s. The only nesting habitat used in this area by
144 the species is Mediterranean scrub, dominated by Kermes Oak (*Quercus coccifera*).
145 Other vegetation types in the region are non-irrigated crops (including cereal fields and
146 orchards), as well as pine plantations. More details on the study area can be found in
147 Limiñana et al. (2006a, 2006b) and Soutullo et al. (2006). The main prey types for this
148 Montagu's Harrier population are passerines and insects (especially orthopterans and
149 small coleopterans), with other prey such as small mammals and lizards being taken less
150 frequently (Limiñana et al. 2008).

151

152 This study is based on data from 2005-2007, when positions of located nests were
153 recorded using a GPS. The area searched to locate Montagu's Harrier nests was *ca.*
154 1050 Km² (Fig. 1); field effort to locate the nests and monitor breeding performance
155 was kept constant through the study period. Population size in the study area was stable
156 between 2001 and 2005 at *ca.* 100 pairs (see Limiñana et al. 2006a and Soutullo et al.
157 2006). In 2006 and 2007 it increased to 129 and 145 breeding pairs, respectively. In
158 2005, a total of 80 Montagu's Harrier nests were located and positions of 76 of them
159 were recorded using a GPS. In 2006, 96 nests were located and positions of 85 of them
160 were recorded. In 2007, there were 101 located nests and positions of 86 were recorded.
161 Breeding data (number of eggs and nestlings) on these nests were recorded during nests
162 visits. All pairs in 2005 which nests were not located were inside known colonies, so
163 the fact that we are not including them in the analyses of harrier nesting occurrence is
164 not likely to strongly affect the results (as they were all in grid cells that were otherwise
165 occupied). Some nests in 2006 and 2007 appeared in new areas for which we did not
166 have accurate habitat data (see below), but study of aerial photographs confirmed that
167 these areas were similar to those included in these analyses in terms of habitat
168 composition.

169

170 **Habitat variables**

171 Habitat composition in the study area was determined using the 1:10 000 Land Use map
172 of the Comunidad Valenciana which is, in turn, based on recent aerial photographs
173 (taken at the time period of the study). We had access to aerial photographs (7x5 km)
174 and to the Land Use map corresponding to the areas where Montagu's Harriers nested
175 up to 2005. We imposed a 1 km² UTM grid over the study area and calculated the
176 proportion of each habitat type within each of the grid cells using a Geographic

177 Information System (ArcView 3.2). We also evaluated (using the GIS) whether each
178 grid cell was occupied or not by breeding harriers, and how many nests were located
179 within each one. This scale has been extensively used in other studies of harrier
180 occupancy (Arroyo et al. 2002, 2005; Tapia et al. 2004). To evaluate breeding
181 performance in relation to breeding habitat, we calculated habitat composition around
182 each nest at two different radii (500 metres and 2000 metres) using the GIS. The area of
183 a circle with 500 metres radius corresponds roughly to the area covered by a 1 km² grid
184 cell, and also corresponds roughly to the area that a female uses for hunting around the
185 nest, at least in Mediterranean areas (García and Arroyo 2005). We also used the 2000
186 metres radius since it includes a large part of the male core home range, also according
187 to studies in southern Europe (Arroyo et al. 2008; Cornulier 2005). By using both radii,
188 we can be sure that we are accounting for both nesting and foraging habitat in our
189 analyses on the effect of breeding habitat on reproductive performance.

190

191 Overall, habitat variables calculated (both for circles and grid cells) were the following:
192 percentage of scrub, percentage of forest (mainly pine plantations), percentage of
193 herbaceous crops (mainly cereal crops) and percentage of non-irrigated orchards
194 (mainly almond and olive). Across the study area, scrub covered 47% of the surface,
195 orchards covered 38%, herbaceous crops 8% and forest 6%.

196

197 **Statistical analyses**

198 Hierarchical partitioning (HP) was used to identify the most likely habitat variables
199 explaining the occurrence of Montagu's Harrier nests in the grid encompassing the
200 study area (Chevan and Sutherland 1991). Hierarchical partitioning computes all of the
201 possible hierarchical models that can be developed with a set of independent predictive
202 variables; this is to say that if U, V and W are variables, HP computes single-order (U,
203 V, W), second order (UV, UW, VW) and higher-order (UVW) models and tests whether
204 the addition of a given variable produces an improvement in goodness of fit. For each
205 independent variable, their explanatory power is segregated into the independent effect
206 'I' and the effects caused jointly with other variables 'J' (MacNally 2000). This analysis
207 was conducted in R (R Development Core Team 2009) with the 'hier.part' package
208 (Walsh and MacNally 2003), using logistic regression and log-likelihood as goodness-
209 of-fit measure. As suggested by MacNally (2002), significance of the individual
210 contribution of each variable included in the analysis was evaluated by a randomization

211 procedure based on 999 randomizations. Grid cell occupancy was defined as 1 =
212 occupied (if at least one nest was present in the cell in at least one year), and 0 =
213 unoccupied (if no nests were known in the cell). Only those cells for which we had
214 information on at least 75% of its surface for habitat variables were used for the
215 analyses. The initial model included percentage of each habitat variable: scrub, forest,
216 herbaceous crops and orchards, as well as their quadratic terms. Quadratic terms were
217 included since other studies of harriers have shown that percentage cover of certain
218 habitats may be optimal at intermediate levels (Arroyo et al. 2005).

219

220 To test whether frequency of use (the number of years that a given cell had been
221 occupied during the study) varied in relation to habitat, we used categorical modelling,
222 with the procedure CATMOD within SAS 9.1 (SAS Institute 1999).

223

224 Secondly, we evaluated the relationship between breeding habitat and breeding output
225 using Generalized Linear Mixed Models (GLMM). Laying date usually influences
226 breeding output in raptors, with pairs laying later having a lower productivity (e.g.
227 Newton and Marquiss 1984; Pietiäinen 1989). It is thus important to control for this
228 variable when analysing factors influencing breeding output. We calculated laying date
229 by backdating from nestling age, assuming an incubation period of 30 days (Cramp and
230 Simmons 1980). Nestling age was estimated from length of the eighth primary wing
231 feather following Arroyo (1995). Laying date was not known for nests that failed, since
232 most of them failed during the incubation period (65% of total nest losses in the period
233 2005-2007) or before the first visit in the nestling stage (so chick age could not be
234 assessed). To analyse breeding output in relation to habitat, we performed two separate
235 (and complementary) analyses. We first evaluated whether breeding success (production
236 of at least one fledged young from a nest) depended on habitat around the nest (at either
237 500 or 2000 m). This binomial response variable (coded as 1 = successful nest, 0 =
238 unsuccessful nest) was modelled using a binomial error distribution and a logit link
239 function, with year included as a random variable. Secondly, we evaluated whether
240 fledged brood size (number of fledglings per successful pair, modelled using a Poisson
241 error distribution and a log link function) depended on habitat within 500 or 2000 m,
242 when controlling for laying date, also including year as a random variable. In semi-
243 colonial species, local density (i.e. number of breeding neighbours) may also influence
244 breeding output because of density-dependent competition within colonies (Arroyo

245 1995). Hence, we also calculated number of breeding neighbours within 500 m from
 246 each nest to include this variable in the models as a covariate. We used 500 m because
 247 this approximates the maximum distance between nests of the same semi-colony
 248 (Arroyo 1995; Cornulier and Bretagnolle 2006; Limiñana 2004). Thus, initial models
 249 included laying date (for fledged brood size only), number of breeding neighbours, and
 250 the habitat variables that had a significant contribution to harrier occurrence in the HP
 251 analysis (proportion of scrub and its quadratic term, and quadratic terms of herbaceous
 252 crops and orchards, see Results). Backward-forward selection and AIC comparisons
 253 were used to identify the final models (those with lower AIC values).

254

255 Because of semi-coloniality, there could be spatial correlation in results according to the
 256 position of nests (i.e., there may be lower variance within than between colonies). In
 257 fact, the variable “colony” was almost significant in models of breeding output with
 258 only colony as explanatory variable. Hence, we also used the variable “colony” as a
 259 random variable in the GLMM, to explain breeding output in relation to habitat
 260 (although in final models, this random variable was not significant in any analyses). As
 261 specified above, semi-colonies were defined based on distances between nests, with
 262 distances between closest semi-colonies being larger than 1000 m (and thus, much
 263 larger than distances between nests within the same semi-colony).

264

265 Finally, we evaluated whether laying date or number of neighbours (assuming a normal
 266 error distribution and using an identity link function) of individual nests varied
 267 according to habitat around them. All the analyses related to breeding output were
 268 carried out using SAS 9.1.

269

270 **RESULTS**

271 **Breeding occurrence and habitat**

272 The habitat variables that best explained the occurrence of Montagu’s Harrier nests in
 273 the study area were the percentage of scrub and its quadratic term, which had each an
 274 independent contribution of more than 20% (Table 1). In addition, there was also a
 275 significant influence of the quadratic terms of percentage cover of herbaceous crops and
 276 non-irrigated orchards (Table 1). The quadratic term for scrub showed a preference for
 277 areas not entirely covered by this habitat, but with an intermediate degree of habitat
 278 diversity, as also illustrated by the significant contribution of the quadratic terms of

279 farmland variables. Overall, cells used in at least one year had a higher proportion of
 280 scrub than non-occupied cells, and a lower proportion of both herbaceous crops and
 281 non-irrigated orchards (Fig. 2).

282

283 Additionally, the frequency of use (the probability that cells were used only one year,
 284 two or three of the years of the study, or not at all) varied quadratically with scrub cover
 285 ($\chi^2_3 = 18.14$, $p = 0.0004$ for scrub cover, $\chi^2_3 = 16.15$, $p = 0.001$ for scrub cover squared),
 286 and with orchard cover ($\chi^2_3 = 7.69$, $p = 0.053$ for orchard cover, $\chi^2_3 = 9.66$, $p = 0.022$ for
 287 orchard cover squared). It was not significantly related to any other habitat variable (all
 288 $p > 0.20$). Grid cells occupied all three years were more likely to have between 40 and
 289 80% of scrub cover (Fig. 3a), and between 0.1 and 60% of orchards (Fig. 3b).

290

291 **Breeding parameters and habitat**

292 We found no relationship between breeding success and either number of neighbours or
 293 any habitat variable within 500 m or 2000 m (all $p > 0.2$).

294

295 When considering habitat within 500 m, fledged brood size was significantly related to
 296 laying date, number of neighbours and proportion of scrub and herbaceous crops ($F_{1,138}$
 297 $= 18.54$, $p < 0.0001$ for laying date; $F_{1,78} = 2.87$, $p = 0.095$ for number of neighbours;
 298 $F_{1,33} = 3.24$, $p = 0.081$ for scrub cover; $F_{1,53} = 5.15$, $p = 0.027$ for herbaceous crops
 299 cover squared). Fledged brood size decreased with laying date, number of neighbours
 300 and scrub cover, and increased with higher availability of herbaceous crops (parameter
 301 estimates: -0.01 ± 0.002 ; -0.015 ± 0.009 ; -0.003 ± 0.002 and 0.0005 ± 0.0002
 302 respectively).

303

304 When considering habitat within 2000 m, the only significant variables explaining
 305 fledged brood size were laying date and number of neighbours ($F_{1,150} = 21.78$, $p <$
 306 0.0001 for laying date; $F_{1,51} = 15.92$, $p = 0.0002$ for number of neighbours) (parameter
 307 estimates: -0.01 ± 0.002 ; and -0.03 ± 0.01 respectively). No habitat variable was
 308 retained in the final model.

309

310 Number of breeding neighbours was significantly higher for nests placed in areas with
 311 higher scrub cover within 500 m ($F_{1,245} = 17.17$, $p < 0.0001$), or with intermediate scrub
 312 cover and relatively high herbaceous crops cover within 2000m ($F_{1,242} = 41.43$, $p <$

313 0.0001 for scrub cover, $F_{1,242} = 46.11$ $p < 0.0001$ for scrub cover squared, $F_{1,242} = 6.43$,
314 $p = 0.009$ for herbaceous crops cover, Fig. 4).

315

316 No significant relationship between laying date and any habitat variable at either at 500
317 or 2000 m was found (all $p > 0.2$).

318

319

320 **DISCUSSION**

321 Our results show that distribution of Montagu's Harrier nests within the occupied area
322 in inland Castellon province is influenced by habitat, as found in other populations of
323 the species or other raptors (see e.g. Tapia et al. 2004; Wilson et al. 2009; López-Iborra
324 et al. 2010). Probability of occurrence of harrier nests was significantly related to the
325 availability of scrub, their nesting habitat. This relationship was not linear, but
326 quadratic: probability and regularity of occurrence in a 1 km² area was greatest at
327 intermediate levels of scrub cover. Also, the areas used by harriers for breeding had
328 intermediate values of both herbaceous crops and non-irrigated orchards at the 1 km²
329 scale. A quadratic relationship between nesting habitat and probability of occupancy
330 was also found for Hen Harriers *Circus cyaneus* in Scotland. These birds use heather
331 *Calluna spp.* as nesting habitat, but forages primarily over a mosaic of heather and
332 grassland, so they favour areas with both foraging and nesting habitat to breed (Arroyo
333 et al. 2005). Similar results were also found in Galicia, NW Spain, where Montagu's
334 Harriers also breed in natural vegetation. There, plots occupied by harriers had a greater
335 extent of scrub (nesting habitat) than unoccupied squares, but also a higher degree of
336 pastureland (their preferred foraging habitat; Tapia et al. 2004). Our results thus indicate
337 that Montagu's Harriers in Castellon prefer areas with heterogeneous land uses, where
338 both scrub (nesting habitat) and farmland occur, and suggest that harriers probably use
339 these non-irrigated orchards and herbaceous crops for foraging in the study area. Indeed,
340 Montagu's Harriers in other parts of Spain often hunt in open areas with herbaceous
341 vegetation, including grasslands and arable fields, as well as open orchards (Martínez et
342 al. 1999; Guixé 2003; Arroyo et al. 2008). Also, Montagu's Harriers in the study area
343 have been observed hunting in open areas with herbaceous vegetation (e.g. grasslands
344 and cereal crops; Limiñana et al. 2008).

345

346 This result suggests that areas with intermediate scrub cover and nearby presence of
347 orchards or herbaceous crops are optimal, but that areas with too much or too little
348 scrub cover are suboptimal for harrier breeding, due to a lack of foraging and nesting
349 habitat, respectively. We could thus expect that breeding output would be higher in
350 areas with intermediate scrub and crop/orchard cover, where harriers probably assure
351 their nesting and foraging needs in a more profitable way.

352

353 We found no effect of any of the habitat variables considered on harrier breeding
354 success, and similarly habitat within 2000 m of the nest had not a significant effect on
355 fledged brood size. However, fledged brood size was positively related to percentage
356 cover of herbaceous crops within 500 m (and, concordantly, almost significantly
357 negatively related to the percentage cover of scrub within 500 m). It is noteworthy that
358 only herbaceous crops, not non-irrigated orchards, had a significant effect on breeding
359 output, suggesting that this habitat type may be better for foraging harriers, possibly
360 because prey are more easily captured in this habitat than in orchards (Martínez et al.
361 1999).

362

363 Overall, fledged brood size was mostly influenced by laying date and number of
364 neighbours, both of which had a negative effect on breeding output of the species. Pairs
365 laying later had a lower fledged brood size, a pattern common in many raptor species
366 that may be related to individual quality (e.g. Newton and Marquiss 1984). However,
367 there was no relationship between laying date and habitat, which suggests that optimal
368 breeding sites (in terms of habitat) are not necessarily occupied before less optimal sites.
369 This may indicate that breeding site occupancy in the species may be better explained
370 by the ideal free model rather than by the ideal despotic distribution (Fretwell and Lucas
371 1970). In any case, as we are only using data from three years, explanations of harrier
372 settlement and distribution in the study area derived from these models should be taken
373 with caution (see also Soutullo et al. 2006).

374

375 Fledged brood size was also related to number of breeding neighbours, with a higher
376 number of nearby breeding neighbours resulting in lower fledged brood sizes. This may
377 reflect local competition for food or other resources, or a higher amount of time spent in
378 conspecific interactions, in larger or denser nest clusters (Mougeot and Bretagnolle
379 2006). On the other hand, predation may also play a role in explaining this pattern, as a

380 higher density of harriers breeding in the same scrub patch could attract more the
381 attention of predators, which may result in lower fledged brood sizes in these areas.
382 However, predation rate of Montagu's Harrier nests in the study area is very low (e.g. a
383 maximum of 12 nests out of 80 located nest were predated in 2005, and a maximum of
384 7 out of 96 nests were predated in 2006), and no effect of either habitat or number of
385 neighbours was found on nesting success, which suggests that the effect observed refers
386 to partial brood reduction or clutch size differences. Interestingly, number of breeding
387 neighbours was related to habitat, being highest in areas with high scrub cover within
388 500 m or with intermediate cover of scrub and also a high percentage of herbaceous
389 crops cover within 2000 m. This suggests that the best habitat conditions to host large
390 colonies are areas where scrub is very abundant at a lower scale (*ca.* 500 m), but
391 intermixed with foraging areas (herbaceous crops) at a larger scale (*ca.* 2000 m).

392

393 Our results thus suggest that habitat influences nest distribution in this species.
394 However, in semi-colonial species, nest or colony location may be strongly influenced
395 by factors like presence of conspecifics and their breeding success (Boulinier and
396 Danchin 1997; Sergio and Penteriani 2005), which may have an even stronger effect
397 than habitat quality (Arroyo et al. 2002; Cornulier and Bretagnolle 2005). Quality of the
398 nesting patch (e.g. size of the scrub patch, availability of places to locate the nest or
399 availability of nearby good feeding areas) appears to be the factor that determines the
400 number of pairs in it (number of neighbours, thus colony size), and density within
401 nesting patches (i.e. local competition by interference) may determine when to settle in
402 a new patch rather than in an existing colony (Soutullo et al. 2006). Location of new
403 colonies might be chosen at random among patches of similar quality, or determined by
404 other factors, such as distance to occupied patches (Hanski 1999).

405

406 This means that, for this species, it may be difficult to predict the impact of small scale
407 habitat changes on population size or distribution at a larger scale. Improvement of
408 breeding habitat for colonial or semi-colonial species such as Montagu's Harrier may
409 result in higher nest numbers at the colony or semi-colony level (and even this may be
410 cupped up because of interspecific competition), but it may not necessarily result in the
411 creation of new colonies (or the occupancy of areas previously unoccupied). In these
412 cases, it may be important to evaluate which other factors (e.g. past occurrence or
413 productivity of breeding birds in that area, distance to other occupied sites, etc.) may

414 influence the distribution and probability of occurrence (Arroyo et al. 2002). If
415 protected/managed areas with good habitat are unoccupied, it may be possible to
416 consider whether some of those parameters may be manipulated for management
417 purposes (for example, by artificially increasing the productivity in certain areas
418 through hacking, or by using decoys to attract them there, as has been done with the
419 Montagu's Harrier (Pomarol 1994) or other semi-colonial species such as osprey
420 (Thibault et al. 1995).

421

422 On the other hand, our results also suggest that habitat (in particular, availability of
423 foraging areas nearby the nest) influences breeding output. Similar results were found in
424 Hen Harriers (Amar et al. 2008). This suggests that habitat management in areas that are
425 occupied regularly may have an impact, increasing local density (as seen above), as well
426 as the productivity of pairs breeding there (and thus, potentially, the likelihood of that
427 area being occupied in subsequent years). In our study area, this habitat seems to be the
428 herbaceous crops, being an open habitat where Montagu's Harriers can easily catch its
429 prey (e.g. Martínez et al. 1999). In that respect, it is important to note this habitat type is
430 one of the most restricted in the study area (only 8% of the surface of the study area).
431 Thus, an increase in its availability may enhance the suitability of the area for breeding
432 Montagu's Harriers. However, this should ideally not be done at the expense of scrub
433 availability, because the latter is important for harriers to choose an area to locate their
434 nests, and important to hold high local densities (which may enhance the suitability of
435 the area due to conspecific attraction). In the study area, the major land-use change
436 observed in recent years is the abandonment of traditional farming practices and non-
437 intensive agriculture for new intensive agricultural practices, mainly the transformation
438 of herbaceous crop fields into more lucrative irrigated orchards. As well as being less
439 well-suited to Montagu's Harrier hunting, the latter habitat may hold a lower density of
440 prey, due to a decrease in prey habitat suitability (especially for ground-nesting
441 passerines) and an increased use of pesticides. Hence, for species conservation in the
442 study area, it would be useful to develop measures to encourage farmers to stop the
443 transformation of herbaceous crops into orchards, or even encourage them to create new
444 herbaceous crops.

445

446 In conclusion, the results of this study suggest that conservation of semi-colonial
447 species, like Montagu's Harrier, should not be solely based on increasing availability of

448 nesting habitat (or on protecting only these habitats), without taking into account social
 449 factors (Reed and Dobson 1993) and the importance of foraging habitats (Sergio et al.
 450 2003; Amar et al. 2008; Arroyo et al. 2009). Habitat improvement may result in higher
 451 local densities and breeding success, but local actions aimed at preserving or enhancing
 452 nesting habitat in irregularly occupied areas for semi-colonial species may result in an
 453 inefficient investment of available conservation resources for these species. Indeed,
 454 habitat manipulation may be inefficient for changing species distribution (and thus,
 455 potentially, total breeding numbers at a regional scale). Thus, having a complete
 456 framework for a target species, including breeding habitat (both nesting and foraging),
 457 social factors and relationships between populations of the species would improve the
 458 effectiveness of conservation effort and investment. Also, protecting several core areas
 459 may be efficient for conservation of semi-colonial species (Sergio et al. 2003; Poirazidis
 460 et al. 2004), especially if such areas are regularly occupied, hold a large number of
 461 breeding pairs or are more productive.

462

463

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477 **REFERENCES**

- 478 Amar A, Arroyo B, Meek E, Redpath S, Riley H.(2008).Influence of habitat on
 479 breeding performance of Hen Harriers in Orkney. *Ibis* 150:400-404.
 480 Arroyo B (1995) Breeding ecology and nest dispersion of Montagu's harrier *Circus*
 481 *pygargus* in central Spain. PhD thesis. Oxford University.

- 482 Arroyo BE, García JT, Bretagnolle V (2002) Conservation of the Montagu's Harrier
483 (*Circus pygargus*) in agricultural areas. *Anim Conserv* 5:283–290.
- 484 Arroyo B, Leckie F, Amar A, Hamilton J, McCluskie A, Redpath S (2005) Habitat use
485 and range management on priority areas for hen harriers: 2004 report. Report to
486 Scottish Natural Heritage, Edinburgh, UK.
- 487 Arroyo B, Amar A, Leckie F, Buchannan G, Wilson J, Redpath S (2009) Hunting
488 habitat selection by hen harriers on moorland: implications for conservation. *Biol.*
489 *Cons* 142:586-596.
- 490 Arroyo B, Pinilla A, Mougeot F, Crystal F, Guerrero A. (2008) Estudio de la
491 ecología poblacional del aguilucho cenizo (*Circus pygargus*) en Extremadura.
492 Report to Servicio de Conservación de la Naturaleza y Espacios Naturales
493 Protegidos, Junta de Extremadura, Spain.
- 494 Benton TG, Bryant DM, Cole L, Crick HQP (2002) Linking agricultural practice to
495 insect and bird populations: a historical study over three decades. *J Appl Ecol*
496 39:673–687.
- 497 Blanco JC, González JL (1992) Libro Rojo de los Vertebrados de España. ICONA-
498 M.A.P.A., Madrid.
- 499 Boulinier T, Danchin E (1997) The use of conspecific reproductive success for breeding
500 patch selection in terrestrial migratory species. *Evol Ecol* 11:505–517.
- 501 Browne SJ, Aebischer NJ, Yfantis G, Marchant JH (2004) Habitat availability and use
502 by Turtle Doves *Streptopelia turtur* between 1965 and 1995: an analysis of
503 Common Birds Census data. *Bird Study* 51:1–11.
- 504 Cade TJ, Temple SA (1995) Management of threatened bird species: evaluation of the
505 hands-on approach. *Ibis* 137:S161–S172.
- 506 Carrete M, Sánchez-Zapata JA, Martínez JE, Calvo JF (2002) Predicting the
507 implications of conservation management: a territorial occupancy model of
508 Bonelli's eagle in Murcia, Spain. *Oryx* 36 (4):349–356.
- 509 Chevan A, Sutherland M (1991) Hierarchical partitioning. *Am Stat* 45: 90-96.
- 510 Cornulier T (2005) Composants de la distribution spatiale d'un prédateur : effets
511 respectifs de l'habitat, des ressources alimentaires et des interactions
512 comportementales. Analyse de processus ponctuels non homogènes. PhD thesis,
513 Université Claude Bernard, Lyon 1.

- 514 Cornulier T, Bretagnolle V (2006) Assessing the influence of environmental
515 heterogeneity on bird spacing patterns: a case study with two raptors. *Ecography*
516 29:240–250.
- 517 Cramp S, Simmons KEL (1980) *The birds of the Western Palearctic*, Vol. 2. Oxford
518 University Press, Oxford.
- 519 Fernández GJ, Posse G, Ferretti V, Gabelli FM (2004) Bird-habitat relationship for the
520 declining Pampas meadowlark populations in the southern Pampas grasslands.
521 *Biol Conserv* 115:139–148.
- 522 Fretwell SD, Lucas HL (1970) On territorial behaviour and other factors influencing
523 habitat distribution in birds, theoretical development. *Acta Biotheor* 19:16–36.
- 524 García JT, Arroyo BE (2005) Food-niche differentiation in sympatric Hen and
525 Montagu's harriers. *Ibis* 147:144–154.
- 526 García-Ripollés C, López-López P, García-López F, Aguilar JM, Verdejo J
527 (2005) Modelling nesting habitat preferences of Eurasian Griffon vulture *Gyps*
528 *fulvus* in eastern Iberian Peninsula. *Ardeola* 52 (2):287–304.
- 529 Guixé D (2003) Caracterització de les àrees de nidificació de l'esperver cendrós a la
530 plana de Lleida. Resultats del radioseguiment i estudi d'alimentació i selecció de
531 l'hàbitat del nucli reproductor d'anglesola. Report to Departament De Medi
532 Ambient i Habitatge, Generalitat de Catalunya.
- 533 Hanski I (1999) *Metapopulation Ecology*. Oxford University Press, Oxford.
- 534 Hiraldo F, Negro JJ, Donazar JA, Gaona P (1996) A demographic model for a
535 population of the endangered lesser kestrel in southern Spain. *J Appl Ecol* 33
536 (5):1085–1093.
- 537 Julliard R, Jiguet F, Couvet D (2004) Common birds facing global changes: what makes
538 a species at risk? *Global Change Biol* 10:148–154.
- 539 Limiñana R (2004) *Biología reproductiva del aguilucho cenizo (Circus pygargus) en el*
540 *interior de la provincia de Castellón*. MSc Dissertation, Universidad de Alicante,
541 Alicante.
- 542 Limiñana R, Surroca M, Miralles S, Urios V, Jiménez J (2006a) Population trend and
543 breeding biology of Montagu's Harrier *Circus pygargus* in a natural vegetation
544 site in Northeast Spain. *Bird Study* 53:126–131.
- 545 Limiñana R, Soutullo A, Urios V, Surroca M (2006b) Vegetation height selection in
546 Montagu's Harriers *Circus pygargus* breeding in a natural habitat. *Ardea* 94
547 (2):280–284.

- 548 Limiñana R, Soria MC, Urios V (2008) Uso del habitat y dieta del aguilucho cenizo, *C.*
 549 *pygargus*, en el interior de la provincia de Castellón. Report. University of
 550 Alicante.
- 551 López-Iborra GM, Limiñana R, Pavón D, Martínez-Pérez JE (2010) Modelling the
 552 distribution of short-toed eagle (*Circaetus gallicus*) in semi-arid Mediterranean
 553 landscapes: identifying important explanatory variables and their implications for
 554 its conservation. Eur J Wildl Res. doi: 10.1007/s10344-010-0402-0
- 555 López-López P, García-Ripollés C, Aguilar JM, García-López F, Verdejo J (2006)
 556 Modelling breeding habitat preferences of Bonelli's eagle (*Hieraaetus fasciatus*)
 557 in relation to topography, disturbance, climate and land use at different spatial
 558 scales. J Ornithol 147 (1):97–107.
- 559 López-López P, García-Ripollés C, Soutullo A, Cadahía L, Urios V (2007) Identifying
 560 potentially suitable nesting habitat for golden eagles applied to Important Bird
 561 Areas design. Anim Conserv 10:208–218.
- 562 Martínez JA, López G, Falcó F, Campo A, de la Vega A (1999) Hábitat de caza y
 563 nidificación del aguilucho cenizo *Circus pygargus* en el Parque Natural de La
 564 Mata-Torre Vieja (Alicante, SE de España): Efectos de la estructura de la
 565 vegetación y de la densidad de presas. Ardeola 46 (2):205–212.
- 566 MacNally R (2000) Regression and model-building in conservation biology,
 567 biogeography and ecology: The distinction between -and reconciliation of-
 568 “predictive” and “explanatory” models. Biodivers Conserv 9:655-671.
- 569 MacNally R (2002) Multiple regression and inference in ecology and conservation
 570 biology: further comments on identifying important predictor variables. Biodivers
 571 Conserv 11:139-1401.
- 572 Mougeot F, Bretagnolle V (2006) Breeding biology of the Red kite *Milvus milvus* in
 573 Corsica. Ibis 148:436-448.
- 574 Newton I, Marquiss M (1984) Seasonal trend in the breeding performance of
 575 Sparrowhawks. J Anim Ecol 53:809–830.
- 576 Newton I (1998) Population limitation in birds. Academic Press, London.
- 577 Oro D, Margalida A, Carrete M, Heredia R, Donazar JA (2008) Testing the Goodness
 578 of Supplementary Feeding to Enhance Population Viability in an Endangered
 579 Vulture. PLoS ONE 3(12):e4084.
- 580 Pietiäinen H (1989) Seasonal and individual variation in the production of offspring in
 581 the ural owl *Strix uralensis*. J Anim Ecol 58:905–920.

- 582 Poirazidis K, Goutner V, Skartsi T, Stamou S (2004) Modelling nesting habitat as a
 583 conservation tool for the Eurasian black vulture (*Aegypius monachus*) in Dadia
 584 Nature Reserve, northeastern Greece. *Biol Conserv* 118:235–248.
- 585 Pomarol M (1994) Releasing Montagu's Harrier (*Circus pygargus*) by the method of
 586 hacking. *J Rapt Res* 28 (1):19-22.
- 587 R Development Core Team (2009). R: A language and environment for statistical
 588 computing. R Foundation for Statistical Computing, Vienna, Austria. ISBN 3-
 589 900051-07-0, URL <http://www.R-project.org>.
- 590 Reed JM, Dobson AP (1993) Behavioural constraints and conservation biology:
 591 conspecific attraction and recruitment. *Trends Ecol Evol* 8:253–256.
- 592 Robinson RA, Wilson JD, Crick HQP (2001) The importance of arable habitat for
 593 farmland birds in grassland landscapes. *J Appl Ecol* 38:1059–1069.
- 594 Salamolard M, Leroux ABA, Bretagnolle V (1999) Le Busard cendré. In: Rocamora G,
 595 Jarry G, Yeatman-Berthelot D (eds) Les oiseaux à statut de conservation
 596 défavorable ou fragile en France. Listes rouges et priorités nationales. S.E.O.F.,
 597 Paris.
- 598 SAS Institute (1999) SAS/STAT Users' Guide, Version 8. SAS Institute Inc, Cary, NC.
- 599 Sergio F, Blas J, Forero MG, Donázar JA, Hiraldo F (2007) Sequential settlement and
 600 site dependence in a migratory raptor. *Behav Ecol* 18 (5):811–821.
- 601 Sergio F, Pedrini P, Mareschi L (2003) Adaptive selection of foraging and nesting
 602 habitat by black kites (*Milvus migrans*) and its implications for conservation: a
 603 multi-scale approach. *Biol Conserv* 112:351–362.
- 604 Sergio F, Penteriani V (2005) Public information and territory establishment in a
 605 loosely colonial raptor. *Ecology* 86 (2):340–346.
- 606 SNPRCN (1990) Livro vermelho dos vertebrados de Portugal. Vol 1 - Mamíferos, Aves
 607 Répteis e Anfíbios. Secretaria de Estado do Ambiente e Defesa do Consumidor,
 608 Lisboa.
- 609 Soh MCK, Sodhi NS, Lim SLH (2006) High sensitivity of montane bird communities to
 610 habitat disturbance in Peninsular Malaysia. *Biol Conserv* 129:149–166.
- 611 Soutullo A, Limiñana R, Urios V, Surroca M, Gill JA (2006) Density-dependent
 612 regulation of population size in colonial breeders: Allee and buffer effects in the
 613 migratory Montagu's Harrier. *Oecologia* 149 (3):543–552.
- 614 Suárez S, Balbontin J, Ferrer M (2000) Nesting habitat selection by booted eagles
 615 *Hieraaetus pennatus* and implications for management. *J Appl Ecol* 37:215–223.

- 616 Tapia L, Domínguez J, Rodríguez L (2004) Modeling habitat use and distribution of
 617 Hen harriers (*Circus cyaneus*) and Montagu's harrier (*Circus pygargus*) in a
 618 mountainous area in Galicia, northwestern Spain. *J Raptor Res* 38:133–140.
- 619 Thibault JC, Bretagnolle V, Dominici JM (1995) Recovery of a resident population of
 620 osprey on Corsica. *J Raptor Res* 29:204–207
- 621 Walsh C, MacNally R (2003) Hierarchical partitioning. R Project for Statistical
 622 Computing. <http://cran.r-project.org/>
- 623 Wilson MW, Irwin S, Norriss DW, Newton SF, Collins K, Kelly TC, O'Halloran J
 624 (2009) The importance of pre-thicket conifer plantations for nesting Hen Harriers
 625 *Circus cyaneus* in Ireland. *Ibis* 151:332–343.

626 TABLES

627

628 Table 1. Results of hierarchical partitioning analysis performed to assess the importance
 629 of habitat on the occurrence of Montagu's Harrier nests in 1 km² scale in inland
 630 Castellon province (E Spain). I and J represent the independent and joint contribution of
 631 each habitat variable respectively. %I is the percentage of the total I accounted for each
 632 habitat variable. z-score is the randomization test for the independent contributions of
 633 each habitat variable calculated from 999 randomizations. *p<0.05.

634

	I	J	Total	%I	z-score
Forest	1.02	0.67	1.69	5.86	0.65
Non-irrigated orchards	1.49	-0.06	1.43	8.52	1.35
Herbaceous crops	1.15	0.82	1.97	6.57	0.81
Scrub	3.89	1.31	5.20	22.29	4.51*
Forest ²	1.27	0.91	2.17	7.26	1.01
Non-irrigated orchards ²	2.79	0.40	3.20	16.00	3.52*
Herbaceous crops ²	2.26	0.94	3.20	12.95	2.39*
Scrub ²	3.59	-1.33	2.26	20.55	4.03*

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636

637

638 **FIGURE LEGENDS**

639

640 **Fig. 1** Location of the study area (grey polygon) and overall distribution of nests.

641

642 **Fig. 2.** Mean (\pm SE) percentage of different habitats in 1 km² grid cells occupied (n =
643 53) and unoccupied (n = 985) by breeding harriers in the study area.

644

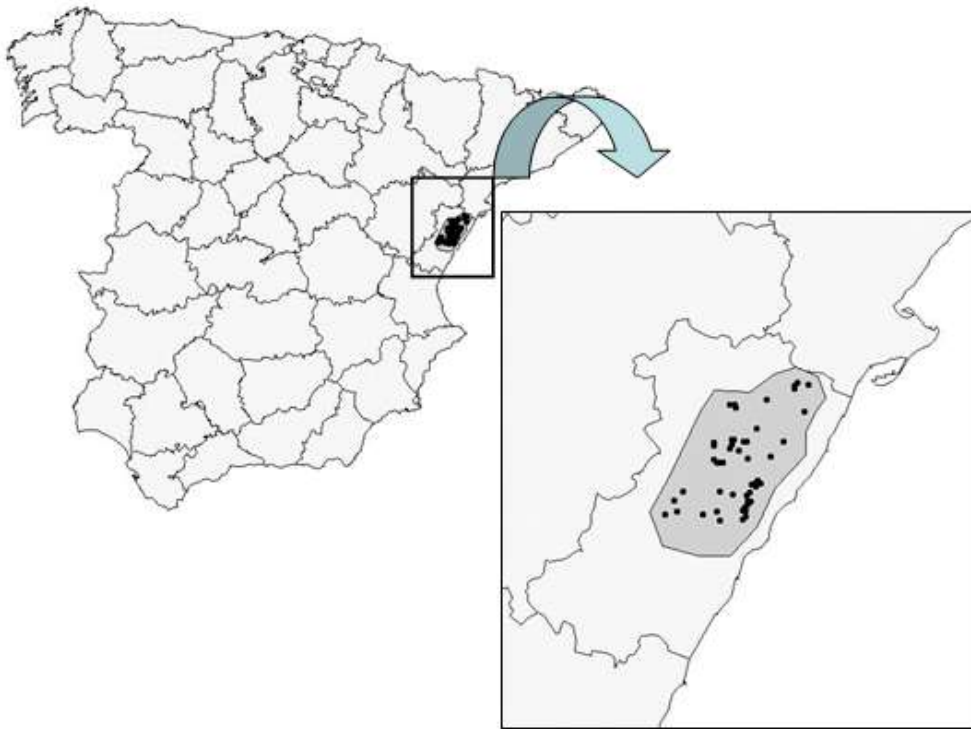
645 **Fig. 3.** Mean (\pm SE) frequency of occupancy (i.e., 0, 1, 2 or 3 years) of 1 km² grid cells
646 in the period 2005 – 2007, according to scrub coverage (above) and orchard cover
647 (below).

648

649 **Fig. 4.** Mean (\pm SE) number of neighbours within 500 m in relation to scrub cover
650 within 500 m of the nest (above) or within scrub cover and herbaceous crops within
651 2000 m of the nest (below).

652

653 **Fig. 1**

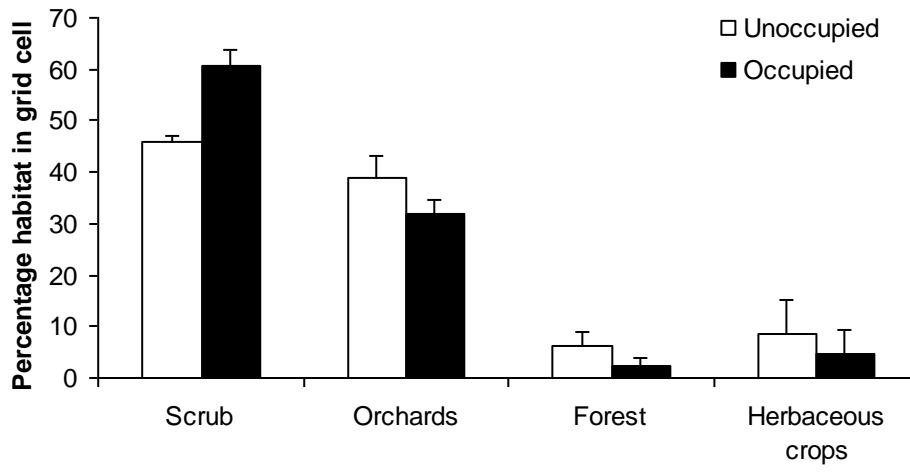


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655 **Fig. 2**

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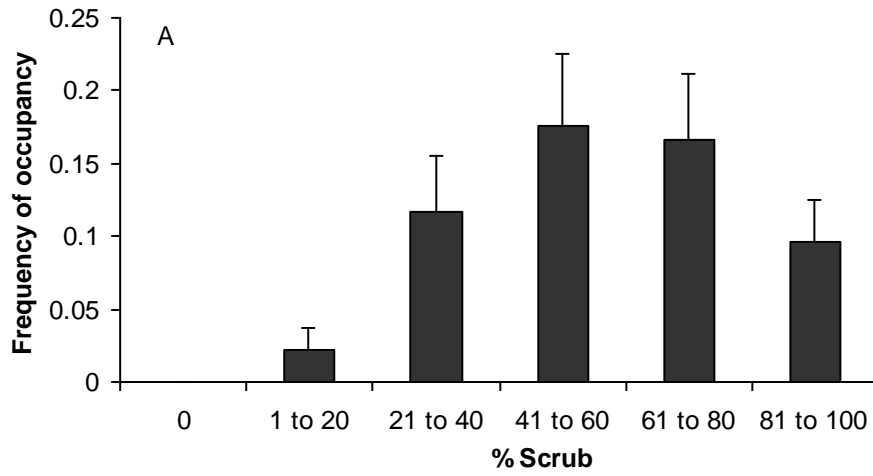


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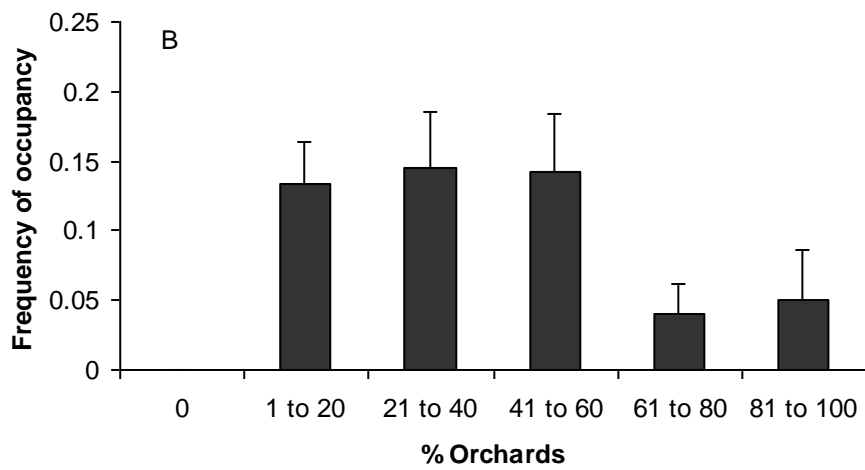
660 **Fig. 3**

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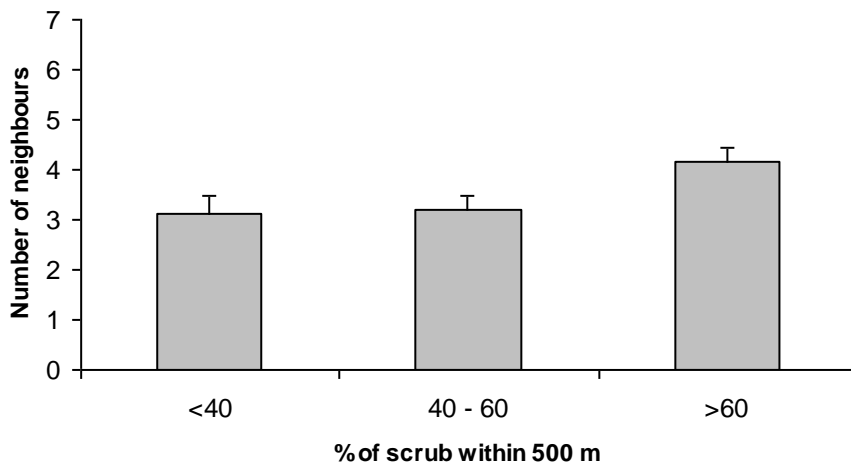
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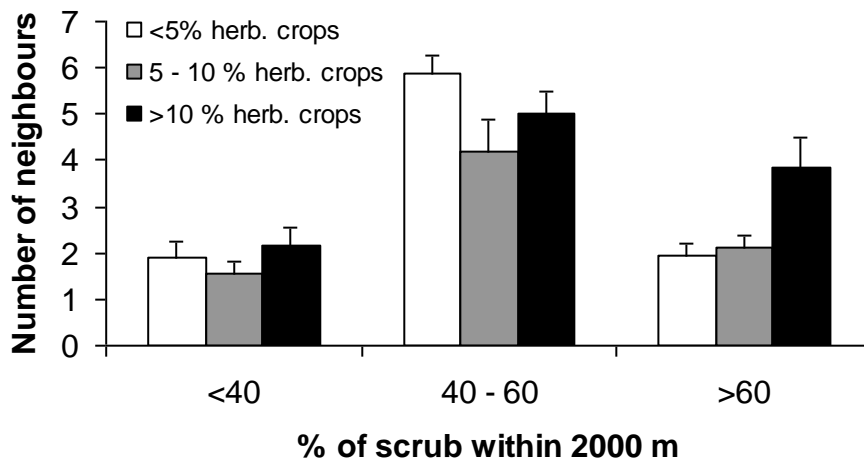
668 **Fig. 4**

669



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