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Unravelling interactions between salt marsh evolution and sedimentary processes in the Wadden Sea (south-eastern North Sea)

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| Abstract: | <p>Salt marshes in the Wadden Sea constitute about 20% of all salt marshes along European coasts. They are of immense importance for coastal protection reasons and as habitat for coastal plant, bird, and invertebrate species. The Wadden Sea is a coastal sedimentary ecosystem in the south-eastern North Sea. Besides salt marshes, it is composed of tidal flats, high sands, and sandy shoals, dissected by (sub)tidal channels and located behind barrier islands. Accelerated global sea-level rise (SLR) and changes in storm climate have been identified as possible threats for the persistence of the Wadden Sea ecosystem including its salt marshes. Moreover, it is known that the amount and composition of the sediment available for salt marshes are the most important parameters influencing their ability to adapt to current and future SLR. Assessing these parameters requires a thorough understanding of the sedimentary system of the salt marshes and the adjacent tidal basins. In the present review, we investigate and unravel the interactions of sedimentary processes in the Wadden Sea with the processes taking place on salt marshes. We identify the most crucial processes and interactions influencing the morphological development of salt marshes in the Wadden Sea. A conceptual model is proposed, intended as a framework for improved understanding of salt marsh development and for incorporation into new salt marsh models. The proposed model may also be applicable to regions other than the Wadden Sea.</p> |

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16 **4 Abstract**
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11 22 model may also be applicable to regions other than the Wadden Sea.
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15 24 **Keywords**
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20 26 Salt marshes, Wadden Sea, sea level rise, sediment dynamics, tidal flats, modelling
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24 28 **I Introduction**
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28 30 The Wadden Sea ecosystem stretches over 450 km along the coast of The Netherlands,
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30 31 Germany, and Denmark and is characterized by a semidiurnal tidal regime (Fig. 1) (CWSS,
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32 32 2008). It is regarded as one of the world's largest unbroken wetland systems, consisting of
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34 33 barrier islands, sandy shoals, tidal sand and mud flats, (sub)tidal channels, and coastal salt
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36 34 marshes. The salt marshes are located at the interface between the tidal flats and the upland
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38 35 and cover a total area of about 400 km². Nearly 50% of them are foreland marshes, in most
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40 36 cases artificially created by the implementation of brushwood groynes in the tidal flats, and
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42 37 located in front of the dikes along the mainland coast (Esselink et al., 2009). In contrast, most
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44 38 natural salt marshes are found at the leeward side of the barrier islands (back-barrier marshes)
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46 39 (Bakker et al., 2005).

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49 40 Salt marshes in the Wadden Sea are considered important for coastal protection (Möller, 2006;
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51 41 Möller et al., 1999) and as habitat for coastal birds, invertebrates, and specialized plant species
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9 42 (Niedringhaus et al., 2008; van der Maarel and van der Maarel-Versluys, 1996). In addition, salt
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11 43 marshes have an important filter function with regard to nutrients and pollutants (e.g. heavy
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13 44 metals) (Reise et al., 2010) and act as a sink for fine-grained sediments (Andersen and Pejrup,
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15 45 2001; Pejrup et al., 1997).

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17 46 Growth and survival of salt marshes are primarily controlled by hydromorphological
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19 47 parameters such as inundation frequency and sediment availability (Pethick, 1981; van Wijnen
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21 48 and Bakker, 2001) as well as the influence of vegetation on the sedimentation process itself
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23 49 (Fagherazzi et al., 2012). Therefore, salt marshes are highly susceptible to changes of the
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25 50 hydromorphological regime, triggered, for example, by a rising sea level and/or increased
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27 51 storm activity (Morris et al., 2002; Mudd et al., 2004; Reed, 1995). Various modelling studies
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29 52 have shown that salt marshes may be able to accrete with roughly the same rate as sea level
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31 53 rises, if sediment availability is sufficient (French, 1993; D'Alpaos et al., 2011; Kirwan et al.,
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33 54 2010). However, as a consequence of recent climate change (including sea level rise (SLR)) and
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35 55 various anthropogenic pressures, salt marshes have been lost and/or are expected to be lost in
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37 56 the future due to drowning and/or lateral erosion in many parts of the world (Duarte et al.,
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39 57 2008; McFadden et al., 2007; Nicholls et al., 1999). In contrast to this global trend, salt marsh
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41 58 areas in the Wadden Sea have remained stable or have been expanding since the end of
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43 59 embanking about 25 years ago (Esselink et al., 2009; Wolff et al., 2010; Stock, 1998). Partly,
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45 60 this trend is due to the artificial creation of foreland marshes (Esselink et al., 2009). For the 21st
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47 61 century and beyond, accelerated SLR has been identified as a major threat for the salt marshes
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49 62 in the Wadden Sea, since not enough sediment may be available for their vertical growth (e.g.
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9 63 Andersen et al., 2011; van Wijnen and Bakker, 2001) and since artificial salt marsh creation is
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11 64 reduced for nature protection reasons (Esselink et al., 2009).

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13 65 For the future ability of the Wadden Sea salt marshes to adapt to SLR, it is assumed that the
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15 66 local availability of fine-grained sediments is the most important variable (Andersen et al.,
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17 67 2011; Schuerch et al., 2013). This implies that understanding the import of fine-grained
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19 68 sediments into the tidal basins as well as the sediment transport processes on the tidal flats
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21 69 and the salt marshes is crucial for estimating the future development of these marshes.

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24 70 However, the interactions between the sedimentary processes of the tidal basins, tidal flats,
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26 71 and salt marshes are insufficiently studied and have not yet been incorporated into predictions
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28 72 for future salt marsh development.

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30 73 With this review, we aim to identify knowledge gaps regarding the main processes that control
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32 74 morphodynamics within the tidal basin – tidal flat – salt marsh continuum and suggest how
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34 75 these gaps can be bridged in salt marsh modelling. For this purpose we have developed a
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36 76 conceptual model that helps to improve the formulation of salt marsh models that predict the
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38 77 ability of coastal salt marshes to adapt to future SLR. Rather than looking at salt marshes as an
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40 78 isolated landscape feature, a more comprehensive and broader-scale approach is taken for
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42 79 highlighting the role of salt marshes in the surrounding sedimentary system. The Wadden Sea
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44 80 ecosystem, in this context, serves as a case study, since it can be considered representative of
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46 81 many similar coastal systems around the world, and extensive research with respect to its
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48 82 morphological functioning has been conducted (Wang et al., 2012).

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9 84 **II Methodology**

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12 86 In the following sections, we outline the sedimentary processes and interactions relevant for
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14 87 the morphological development of salt marshes in the Wadden Sea. Where possible, existing
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16 88 process knowledge is reviewed based on studies directly referring to the Wadden Sea
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18 89 ecosystem. However, for site-independent processes driving the Wadden Sea sedimentary
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20 90 system, a more general review approach is taken referring to the most important literature in
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22 91 the field.

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26 92 We start with a synthesis of the processes governing the salt marsh itself and the identification
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28 93 of knowledge gaps. We continue with a description of the existing and missing knowledge on
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30 94 tidal flat dynamics and the processes taking place in the tidal basin. The importance of all these
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32 95 processes for the morphological development of salt marshes in the Wadden Sea is thereby
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34 96 highlighted. In the end, we integrate this knowledge in a conceptual model that will help salt
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36 97 marsh modellers to expand their models and improve the model performance with respect to
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38 98 short- and long-term temporal as well as spatial variations in salt marsh development.

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43 100 **III Morphodynamics of coastal salt marshes**

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47 102 *1 Vertical salt marsh development*

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9 104 *a Processes of vertical salt marsh growth*. Driven by regular tidal to episodic wind induced
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11 105 inundations, the growth of salt marshes is controlled by a continuous input of mainly mineral
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13 106 (and much less organic) sediment brought onto the salt marsh by the flooding water. Such
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15 107 allochthonous accretion is strongly related to the relative elevation of the salt marsh within
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17 108 the tidal frame, the inundation frequency, and the amount of external sediment supplied by
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19 109 the flooding water (Kirwan et al., 2010; French, 1993; van Wijnen and Bakker, 2001). Besides
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21 110 the allochthonous accretion, salt marshes accrete as a result of autochthonous growth, the
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23 111 accumulation of aboveground and belowground biomass including the growth of benthic
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25 112 microflora (e.g. cyanobacteria and eucaryotic algae, such as diatoms and *Vaucheria*) (Chmura,
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27 113 2013; Sullivan and Currin, 2002).
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29 114 For allochthonous marsh accretion, vegetation acts as a facilitating ecosystem-engineer. It
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31 115 reduces lateral and vertical marsh erosion and stabilizes the soil through rooting and other
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33 116 modifications of the soil properties (Feagin et al., 2009; Howes et al., 2010). Moreover, the
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35 117 aboveground plant structures reduce flow velocities and turbulence, enhancing the deposition
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37 118 of mineral and organic sediment particles by trapping suspended sediments directly and by
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39 119 increasing particle settling velocities in densely vegetated environments (Fig. 2) (Leonard and
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41 120 Croft, 2006; Mudd et al., 2010; Nepf, 1999; Stumpf, 1983). Spatial variability in the
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43 121 depositional pattern is introduced by the presence of vegetation as suspended sediment
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45 122 concentrations of the flooding water are depleted while travelling over the vegetated marsh
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47 123 surface (Christiansen et al., 2000; French and Spencer, 1993; Temmerman et al., 2003a).
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49 124 Depending on the marsh elevation, different vegetation types and densities establish on the
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9 125 marsh that vary with respect to their physical plant characteristics (e.g. stem diameter and
10 126 stiffness, leaf area, vertical biomass distribution) and their potential to reduce flow velocities
11 127 and turbulences as well as their ability of directly capturing sediments (Neumeier and Amos,
12 128 2006; Mudd et al., 2010; Marani et al., 2013). Additionally, sedimentation rates over vegetated
13 129 marsh surfaces vary spatially depending on the ratio of inundation height to vegetation height
14 130 (Temmerman et al., 2005) with smaller spatial variability, when the marsh vegetation is
15 131 completely submerged during an inundation event (Temmerman et al., 2005).
16 132 Marsh growth via vertical accretion is counteracted by autocompaction (Allen, 2000). This
17 133 process is generally assumed to be negligible on allochthonous marshes, but highly important
18 134 in autochthonous marshes (Cahoon et al., 1995; Allen, 2000). However, Bartholdy et al. (2010)
19 135 find autocompaction to be of major importance for a barrier-connected mineralogenic salt
20 136 marsh in the Danish Wadden Sea. Besides autocompaction, the salt marsh surface may also be
21 137 lowered by external factors, such as tectonic or human-induced soil subsidence (e.g. de Vlas,
22 138 2005).
23 139 If total marsh accretion exceeds the combined effect of soil subsidence, autocompaction, and
24 140 eustatic SLR, the marsh elevation increases relative to mean sea level (MSL). Inundation
25 141 frequency then decreases and sedimentation rates slow down when vegetation succession
26 142 proceeds from pioneer marsh over low marsh to high marsh vegetation (Adam, 1990;
27 143 Bockelmann et al., 2002; Leendertse et al., 1997; Olf et al., 1997).
28 144 *b Vertical salt marsh growth under the influence of SLR.* Salt marsh growth usually exists in a
29 145 quasi dynamic equilibrium with SLR (Allen, 1995; Allen, 2000; Morris et al., 2002). More
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9 146 frequent and higher inundation events enhance mineral sedimentation when sea level rises. A
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11 147 parallel increase of biomass is observed on the salt marsh up to a critical SLR rate (Morris et al.,
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13 148 2002). Given sufficient sediment supply, most marshes are likely to survive SLR (Kirwan and
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15 149 Temmerman, 2009). If sediment supply decreases, the salt marsh may not be able to grow fast
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17 150 enough (Andersen et al., 2011), leading to regressive succession (Warren and Niering, 1993) or
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19 151 drowning of the marsh (Fig. 3) (D'Alpaos et al., 2011; Kirwan et al., 2010). In the Danish
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21 152 Wadden Sea, Andersen et al. (2011) found indications for a decreasing sediment supply, either
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23 153 caused by a long-term regime shift or by short term variations in the sedimentary system of
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25 154 the Wadden Sea. Also in other parts of the Wadden Sea, trends of regressive succession are
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27 155 currently observed (Leendertse et al., 1997; Schröder et al., 2002; Stock, 2011), although this
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29 156 may partly be attributed to grazing activities (Bakker, 1985; Stock, 2011; Nolte et al., 2013).
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31 157 Apart from local sediment supply, the survival of salt marshes is crucially dependent on tidal
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33 158 range (Fig. 3) (Harrison and Bloom, 1977; French, 1993; D'Alpaos et al., 2011; Kirwan et al.,
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35 159 2010). In macro-tidal environments tidal currents are stronger, thereby enhancing sediment
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37 160 resuspension in the tidal basin and on tidal flats in particular, resulting in a higher sediment
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39 161 supply for the salt marsh (Temmerman et al., 2004a). Additionally, a large tidal range increases
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41 162 the ability of salt marshes to cope with SLR by allowing the marsh to cover a wider elevational
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43 163 range, thus surviving longer even if marsh elevation decreases relative to MSL (Kirwan and
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45 164 Guntenspergen, 2010).
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49 165 *c Vertical salt marsh growth and storm activity.* Apart from SLR and the prevailing tidal regime
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51 166 the occurrence of strong onshore wind and storm events crucially affects marsh development
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9 167 (Bartholdy et al., 2004; Schuerch et al., 2013). Positive effects of storm activity on vertical salt
10 168 marsh accretion rates are increased inundation frequencies and heights as well as increased
11 169 sediment supply (Stumpf, 1983; Bartholdy et al., 2004; Bellucci et al., 2007; Schuerch et al.,
12 170 2012). Sedimentation rates are generally assumed to gradually increase with higher inundation
13 171 depths, since sediment concentrations of the flooding water and the absolute amount of
14 172 sediment transported onto the marsh platform are higher (Temmerman et al., 2003b).
15 173 Nevertheless, it appears that largest sedimentation rates are associated with time periods in
16 174 which many strong wind or weak storm events occur, while the occurrence of extreme storm
17 175 events seems to be less important (Bartholdy et al., 2004). This may be explained simply by the
18 176 frequency of inundation events, but also by the high probability of inundations overtopping
19 177 the vegetation canopy. This is, in turn, associated with higher depth-averaged flow velocities, a
20 178 process, which has hardly been investigated, but is discussed in more detail later.
21 179 Consequently, sedimentation rates increase more slowly with increasing inundation height as
22 180 soon as the vegetation canopy is completely submerged during inundation (Schuerch et al.,
23 181 2012; Neumeier and Amos, 2006).
24 182 Generally, micro-tidal marshes have been shown to benefit more from increased storm activity
25 183 than macro-tidal marshes, since sediment dynamics in micro-tidal marshes are rather
26 184 controlled by wind induced processes than in macro-tidal marshes, which are more governed
27 185 by tidal currents (Stumpf, 1983; Kolker et al., 2009). Additionally, the effect of storm surges
28 186 with regard to inundation heights and frequencies as well as increased sediment supply tends
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9 187 to be relatively more important in micro-tidal marshes than in macro-tidal marshes (Stumpf,
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15 190 *2 Lateral marsh dynamics*

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19 192 *a Physical drivers for lateral marsh erosion.* Besides vertical growth, salt marsh development is

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21 193 subject to lateral marsh dynamics. Lateral salt marsh erosion and expansion are suggested to

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23 194 be part of a cyclic behaviour, where an erosive phase of salt marsh retreat is followed or

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25 195 accompanied by the re-establishment of pioneer vegetation in front of the marsh platform

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27 196 (Yapp et al., 1917; Redfield, 1972; Esselink et al., 2009; van de Koppel et al., 2005; Singh

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29 197 Chauhan, 2009). Salt marshes may emerge at upper tidal flats when and where sediment

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31 198 accretion outpaces SLR with inundation frequency and bottom shear stress gradually

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33 199 decreasing. The developing vegetation stabilizes the sediment and further enhances sediment

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35 200 accretion (Fig. 2) (Orson et al, 1985; van de Koppel et al., 2005). In case sediment supply is not

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37 201 sufficient for the tidal flat to adapt to SLR (e.g. due to decreased sediment availability or

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39 202 increased SLR), the tidal flat in front of the salt marsh is lagging behind the accretion rate on

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41 203 the salt marsh platform (Fagherazzi et al., 2006). A steepening scarp develops at the edge of

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43 204 the salt marsh, prone to wave attacks, and salt marsh retreat by lateral erosion is initiated (Fig.

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45 205 5) (Mariotti and Fagherazzi, 2010). This process may lead to catastrophic collapse of an entire

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47 206 salt marsh (Mariotti and Fagherazzi, 2013). Callaghan et al. (2010) showed that gently sloping

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49 207 and highly elevated tidal flats with high sediment stability most effectively attenuate wave

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9 208 energy, even at highly exposed sites, and thereby inhibit lateral erosion of salt marshes
10 209 (Callaghan et al., 2010). Locally, this can also result in a reduced sediment supply for the
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12 210 vertical growth of the salt marsh (Reed, 1988; van Leeuwen, 2008).
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14 211 Besides the morphology of the tidal flats, wave impacts on seaward marsh edges and on the
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16 212 adjacent tidal flats are influenced by local hydrodynamics. Strong wind or storm events
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18 213 producing waves within tidal basins may coincide with high tide, when the tidal flats are
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20 214 completely inundated. In such a case, larger fetch lengths, wave heights, and wave periods
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22 215 increase the probability of vertical erosion of higher elevated tidal flats (Fagherazzi and
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24 216 Wiberg, 2009). Elevated water levels may also amplify marsh edge erosion, since incoming
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26 217 waves could directly reach to the scarp of the salt marsh (Tonelli et al., 2010).
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28 218 *b Lateral marsh erosion in the Wadden Sea.* The process of lateral salt marsh erosion is of great
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30 219 importance in the Wadden Sea. Successive embankments have reduced salt marsh area by
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32 220 about 90% over the last millennium and often have shifted the coastline seaward (Reise,
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34 221 2005). As a consequence, tidal basins became truncated at the landward side and the loss of
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36 222 area increased the hydrodynamic energy due to an increased tidal range and enhanced storm
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38 223 tide wedge (Wang et al., 1995; Flemming and Nyandwi, 1994; Reise, 2005). Lateral erosion at
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40 224 the seaward edges of the foreland marshes was initiated (Reise et al., 2003). Brushwood
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42 225 groynes were set up on the tidal flats fronting sea walls to attenuate wave energy and to
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44 226 promote sediment accretion (Wolff et al., 2010). On these dike forelands, fast lateral
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46 227 expansion at the expense of mud flats is observed combined with very high vertical accretion
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48 228 rates of up to 18 mm/yr (Stock, 2011).
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9 229 Meanwhile, the presence of a continuous dike line along most of the mainland coast of the
10 230 Wadden Sea inhibits the inland migration of the foreland marshes, which would be their
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12 231 natural response to SLR and increased hydrodynamic energy. Wherever the marsh is laterally
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14 232 eroding or vertical accretion rates in the pioneer and low marsh zones are below current SLR
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16 233 rates, the marshes are squeezed in between the sea and the dike. This phenomenon, usually
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18 234 referred to as “coastal squeeze”, has been shown to intensify with accelerated sea level rise
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20 235 and stronger storm activity (Bartholomä and Flemming, 2007; Doody, 2004; Flemming and
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22 236 Nyandwi, 1994).
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28 238 *3 Knowledge gaps in understanding of salt marsh dynamics*

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32 240 Having evaluated the available literature related to vertical and lateral salt marsh dynamics,
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34 241 we conclude that the general mechanisms of vertical salt marsh accretion and lateral salt
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36 242 marsh development are well understood and described for homogenous salt marshes with a
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38 243 constant sediment supply. Existing salt marsh models (e.g. Randerson, 1979; Krone, 1987;
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40 244 Allen, 1990; French, 1993; D'Alpaos et al., 2007; D'Alpaos et al., 2011; Kirwan et al., 2010;
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42 245 Marani et al., 2010; Mariotti and Fagherazzi, 2010; Mudd et al., 2004; Fagherazzi et al., 2006)
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44 246 have proven their great potential for gaining process knowledge, since many of the involved
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46 247 processes are hard to measure (Nolte et al., 2013). Most of these models, however, exhibit
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48 248 shortcomings in the representation of spatial sedimentation patterns and with temporally
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9 249 varying environmental boundary conditions (Andersen et al., 2011). As the main reasons for
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11 250 these shortcomings we identify the following knowledge gaps:
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13 251 1) Bio-physical interactions between the marsh vegetation and the hydromorphological
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15 252 processes relevant for the vertical accretion processes on salt marshes are hardly investigated
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17 253 yet (Temmerman et al., 2005; Neumeier and Amos, 2006; Marani et al., 2013). These
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19 254 processes include the spatial influence of heterogeneous marsh vegetation on particle
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21 255 flocculation, hence particle sizes and net sediment flux affecting vertical accretion rates.
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24 256 Marani et al. (2013) have demonstrated the importance of such spatial heterogeneity for the
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26 257 marsh development and concluded that these may affect the resilience of marshes against sea
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28 258 level rise and sediment depletion.
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30 259 2) Very little attention has been given to the effect of vegetation height in relation to
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32 260 inundation height on sedimentation rates. While a few authors have compared the
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34 261 hydrodynamic influences of emerging and submerged marsh vegetation (Temmerman et al.,
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36 262 2005; Neumeier and Amos, 2006), nothing is known on how this ratio quantitatively influences
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38 263 long-term sedimentation rates. Assuming a significant increase of depth-averaged flow
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40 264 velocities once the marsh vegetation is completely submerged, one may expect a decrease of
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42 265 depth-averaged settling velocities. This would be associated with lower sedimentation rates
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44 266 than if a constant settling velocity is assumed. Specifically, this effect may be important if
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46 267 changes in the tidal range and/or changes in storm patterns in combination with SLR result in
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48 268 greater inundation depths.
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9 269 3) Even though, it is well understood that the morphological development of salt marshes
10 270 primarily depends on local sediment supply, a thorough understanding of how the temporal
11 271 variability of this parameter affects the short-term and long-term development of coastal salt
12 272 marshes is not available (Andersen et al., 2011). Such temporal variations could be induced by
13 273 increased hydrodynamic energy, e.g. as a consequence of an increased tidal range or a phase
14 274 of stronger storm activity, or by changes in the sediment dynamics of the surrounding coastal
15 275 environment, such as the sediment resuspension from tidal flats or the import of sediment
16 276 into the tidal basin. A range of conceptual numerical models is employed to investigate these
17 277 processes and their impacts on the lateral marsh development and the local sediment supply
18 278 of salt marshes (Marani et al., 2010; Mariotti and Fagherazzi, 2010; Tonelli et al., 2010).
19 279 However, none of these models has considered the large temporal variability in the stability of
20 280 tidal flat sediments.

21 281 A better understanding of these sedimentary processes could improve the reliability of
22 282 predictions for future salt marsh development and would help to evaluate whether observed
23 283 trends are a result of a long-term regime shift (e.g. triggered by SLR) or whether they are
24 284 induced by short-term variations in storm activity, for example (Schuerch et al., 2013;
25 285 Andersen et al., 2011).

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27 287 **IV Sediment dynamics on the tidal flats and in the tidal basins**

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29 289 *1 Sediment resuspension on tidal flats*

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11 291 Sediment resuspension on tidal flats positively influences vertical accretion rates on salt
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13 292 marshes through increased sediment availability (Callaghan et al., 2010). At the same time,
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15 293 lateral erosion may be caused as a consequence of vertically eroding tidal flats and hence
16
17 294 decreasing wave attenuation in front of the salt marsh (Mariotti and Fagherazzi, 2010).
18
19 295 Sediment resuspension on tidal flats is a function of the erodibility of the sediment and the
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21 296 prevailing shear stress (Mariotti and Fagherazzi, 2010). The process of sediment resuspension
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23 297 starts when the prevailing shear stress induced by tide- or wind-driven currents as well as
24
25 298 (breaking) waves exceeds the critical shear stress of the sediment surface (Miller et al., 1977;
26
27 299 van Rijn, 1993; Andersen et al., 2007; Mariotti and Fagherazzi, 2010). In most locations the
28
29 300 wind wave-induced shear stress is the dominant term for erosion on tidal flats (Le Hir et al.,
30
31 301 2000; French et al., 2000). The erosion rate thereby depends on the change of the critical
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33 302 shear stress within the sediment surface layer (Sanford and Maa, 2001).
34
35 303 *a Physical parameters affecting sediment stability.* The erodibility of sediments is directly
36
37 304 related to its grain size composition (Hjulström, 1955). For non-cohesive (sandy) sediments the
38
39 305 critical shear stress continuously increases with higher grain sizes (Soulsby and Whitehouse,
40
41 306 1997). For cohesive (muddy) sediments it increases with smaller particle sizes (Hjulström,
42
43 307 1955) and stronger consolidation (Postma, 1961).
44
45 308 In natural environments, sediments are mostly a mixture of non-cohesive sandy and cohesive
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47 309 muddy material (Andersen et al., 2010). Depending on the mineralogy and grain size
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49 310 composition of the sediment, the highest critical shear stress is found at mud contents of
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9 311 about 30-50% (Le Hir et al., 2007; Mitchener and Torfs, 1996; Grabowski et al., 2011). Based on
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11 312 laboratory experiments (Panagiotopoulos et al., 1997), Ahmad et al. (2011) show that the
12
13 313 critical shear stress moderately increases with higher mud contents up to a mud content of
14
15 314 about 50%; above 70% it dramatically decreases. Similarly, Panagiotopoulos et al., (1997)
16
17 315 suggests a maximum critical stress at a mud content of 50%.

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19 316 *b Grain size distribution in the tidal basins of the Wadden Sea.* As a consequence of land
20
21 317 reclamation and “coastal squeeze”, the finest grain size fraction in vicinity of the mainland has
22
23 318 been depleted in comparison to pre-embankment conditions (Flemming and Bartholomä,
24
25 319 1997; Flemming and Nyandwi, 1994; Mai and Bartholomä, 2000). This depletion is intensified
26
27 320 in narrow tidal basins and is assumed to continue as a consequence of increased SLR and
28
29 321 storm activity (Bartholomä and Flemming, 2007; Flemming and Nyandwi, 1994; Mai and
30
31 322 Bartholomä, 2000). Under storm conditions, wave-induced resuspension of tidal flat sediments
32
33 323 and the following export of primarily fine-grained material (Bartholomä and Flemming, 2007;
34
35 324 Lettmann et al., 2009) as well as an increasing amount of coarse-grained sediments imported
36
37 325 into the tidal basin via suspended load transport are responsible for this trend (Lettmann et al.,
38
39 326 2009; Santamarina Cuneo and Flemming, 2000).

40
41 327 A continuing depletion of fine-grained sediments could lead to higher sediment instability on
42
43 328 the tidal flats with less than 30% fine-grained sediments at present, and to higher sediment
44
45 329 stability on tidal flats with currently more than 50% fine-grained sediments (Le Hir et al., 2007;
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47 330 Mitchener and Torfs, 1996; Grabowski et al., 2011). Hence, it could affect the availability of
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49 331 fine-grained sediments for the salt marshes.
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9 332 *c Biological parameters influencing sediment stability.* The biotic activity in and on the
10 333 sediment is considered as a major factor mediating the mud content of the sediment via
11 334 deposition of micro-phytoplankton, production of micro-phytobenthos as well as pelletisation
12 335 of macro-zoobenthos (Andersen et al., 2010). Even more importantly, biological activity is
13 336 directly affecting the stability of the sediment (Reise, 2002).
14
15 337 Among the biological parameters influencing the stability of tidal flat sediments, particular
16 338 emphasis is given to the influence of benthic diatoms and cyanobacteria. In the Wadden Sea,
17 339 these have been shown to increase the sediment stability by the production of extracellular
18 340 polymeric substances (EPS) as well as colloidal carbohydrates and enhance local sedimentation
19 341 by direct trapping of suspended sediment (e.g. Andersen et al., 2010; Austen et al., 1999;
20 342 Lanuru et al., 2007; Paterson, 1989; Stal, 2010). The biomass of diatoms is restricted by the
21 343 availability of light and nutrients as well as by grazing and resuspension through the
22 344 macrozoobenthos. Highest concentrations are usually found in April and September (Andersen
23 345 et al., 2010) and on higher elevated tidal flats (Austen et al., 1999). It should, however, be
24 346 noted that benthic diatoms, occurring on the sediment surface only influence the critical shear
25 347 stress at the sediment surface, while they do not decrease the erosion rate, once the critical
26 348 shear stress has been exceeded (Andersen, 2001; Mariotti and Fagherazzi, 2010).
27
28 349 Another important biological influence on the erodibility of the tidal flats is the presence of
29 350 macrozoobenthos, either stabilizing or destabilizing the sediment surface (Knaapen et al.,
30 351 2003; Volkenborn et al., 2007). In the Wadden Sea, the lugworm (*Arenicola marina*), for
31 352 example, as one of the most prevalent macrobenthic animals, has been shown to increase the
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9 353 erodibility of the sediment by intense bioturbation. This activity also inhibits the colonization
10 354 of potentially sediment-stabilizing species, such as the tube-building polychaetes *Polydora*
11 355 *cornuta* and *Lanice conchilega* (Lanuru, 2004; Volkenborn et al., 2009). In any case, benthic
12
13 356 macrofauna tends to increase the surface roughness therefore enhancing the erodibility of the
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15 357 sediment (Lanuru, 2004). Some benthic macrofauna species, such as the mud snail *Hydrobia*
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17 358 *ulvae* additionally increase the erodibility of the sediment indirectly by grazing the biofilms of
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19 359 diatoms and producing fecal pellets that are easily erodible (Andersen, 2001; Andersen and
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21 360 Pejrup, 2002).
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28 362 2 Flocculation processes and floc settling velocity

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33 364 The settling velocity of suspended sediment is generally considered as a key factor for the
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35 365 sediment dynamics in an estuary or coastal lagoon (Mantovanelli, 2005; van Leussen, 1988;
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37 366 Winterwerp, 2002). While in theory, the settling velocity of non-cohesive single particles is
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39 367 shown to be a function of particle size only (Soulsby and Whitehouse, 1997), in reality it is
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41 368 controlled by the aggregation of cohesive sediment particles into flocs (Krone, 1962). Changes
42
43 369 of sediment composition and suspended sediment concentrations affect flocculation processes
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45 370 and hence settling velocities of the suspended sediment. While flocculation processes are
46
47 371 related to high spatial and temporal variability, a general approximation for the settling
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49 372 velocity of flocs (or single particles) as a function of the floc/particle diameter and its density is
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51 373 given by Stokes law (Rubey, 1933).
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9 374 According to their size and synthesis, Eisma (1986) distinguishes between micro-flocs (<125
10 375 μm) and macro-flocs (<3-4 mm), built through collision of micro-flocs. The latter are much
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12 376 larger, but less dense and more fragile than micro-flocs (Eisma, 1986). The organic content,
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15 377 salinity, size of the single grains determine the flocculation ability, a measure of the probability
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17 378 of particles to aggregate when colliding (Eisma, 1986; Kranck, 1973), while high sediment
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19 379 concentrations and turbulent shear stress increase the probability of particle collision.
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21 380 Meanwhile, the turbulent shear stress controls the maximum floc size by breaking up large
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23 381 flocs in highly turbulent environments (Fig. 5) (van Leussen, 1994; Manning, 2004; Burban et
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26 382 al., 1989; Winterwerp, 1998; Winterwerp et al., 2006). The turbulent shear stress threshold,
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28 383 resulting in the maximum floc size and the highest settling velocities, is controlled by the
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30 384 prevailing environmental conditions, such as, for example, the sediment composition
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32 385 (Manning et al., 2010). It ranges from 0.3 N/m² to 0.6 N/m² (Manning, 2004; Winterwerp et al.,
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34 386 2006; Manning and Dyer, 2007; Manning et al., 2010), whereas for microflocs it tends to be
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36 387 higher (Fig. 5) (Manning, 2004; Manning and Dyer, 2007). Given that the suspended sediment
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38 388 concentration is below the critical value where hindering effects for sediment settling occur
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40 389 (Winterwerp, 2002), the settling velocity of a floc can be approximated by an exponential
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42 390 relationship with the suspended sediment concentration (Burt, 1986).
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45 391 Bioaggregation, the flocculation process mediated by organic matter, is considered as the most
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47 392 important driver for flocculation in many coastal waters, including the Wadden Sea (Alldredge
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49 393 and Silver, 1988; van Straaten and Kuenen, 1957). It is mostly driven by the influence of
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51 394 microalgae, such as diatoms, binding mineral particles together, producing the so-called
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9 395 marine snow. High rates of bioaggregation or production of marine snow are observed during
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11 396 algae blooms with a corresponding increase in settling velocity of the suspended matter
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13 397 (Allredge and Gotschalk, 1989). Another important trigger for bioaggregation is the presence
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15 398 of filter-feeding organisms, such as *Mytilus edulis* and *Cerastoderma edule* (van Straaten and
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17 399 Kuenen, 1957; Verwey, 1952) or *Hydrobia ulvae* (Andersen and Pejrup, 2002). Through
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19 400 biodepositon, these filter-feeders enhance sediment settling, while producing faecal pellets
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21 401 and pseudo-faeces (Graf and Rosenberg, 1997; Kautsky and Evans, 1987). Therefore, particle
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23 402 sizes and settling velocities increase in areas, where filter-feeding organisms are abundant
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26 403 (Andersen and Pejrup, 2002).

27 28 404 29 30 405 *3 Flocculation processes on salt marshes*

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34 407 Flocculation processes are different on vegetated marsh platforms in comparison to
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36 408 unvegetated tidal flats (Graham and Manning, 2007). Due to a continuous decrease of flow
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38 409 velocity and turbulent energy towards the inner parts of the marsh, larger sediment particles
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40 410 and flocs are settling in the vicinity of the marsh edge and the tidal channels, while smaller
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42 411 sediment particles are transported further into the marsh (Fig. 2) (Christiansen et al., 2000;
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44 412 French and Spencer, 1993; Temmerman et al., 2004b; Temmerman et al., 2003a). A lower flow
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46 413 velocity and turbulent energy in the inner part facilitates sedimentation due to reduced
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48 414 vertical turbulent diffusion (Mudd et al., 2010; Neumeier and Amos, 2006; Shi et al., 1996), but
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50 415 may also inhibit flocculation of fine-grained sediments and reduce settling velocities, since the
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9 416 prevailing turbulent shear stress is usually lower than the critical turbulent shear stress for the
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11 417 maximum floc size (Fig. 2) (Neumeier and Amos, 2006; Fagherazzi et al., 2012).
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13 418 Furthermore, vertical flow gradients on vegetated salt marsh platforms (Leonard and Luther,
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15 419 1995) affect sediment deposition on salt marshes (Neumeier and Amos, 2006). Flow velocity
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17 420 and turbulent shear stress within the vegetation canopy are considerably lower than above the
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19 421 vegetation or on the bare tidal flats, thereby promoting the settlement of cohesive sediment.
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21 422 On the other hand, increasing flow velocity and turbulent shear stress above the vegetation
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23 423 canopy inhibit sediment settlement through high flow velocities (Shi et al., 1996), while
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25 424 simultaneously increasing the floc sizes through high turbulent shear stress (if those are not
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27 425 exceeding the critical value) (Manning, 2004). During storm tides, such large flocs are travelling
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29 426 over the marsh and reach higher marsh elevations or slowly sink into the vegetation canopy.
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31 427 Thus, events of increased sediment settling may be observed there. Generally, flocculation
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33 428 processes on salt marshes are controlled by the density and the morphology of marsh
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35 429 vegetation (Graham and Manning, 2007) but are still unknown to a large extent.
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41 431 *4 Knowledge gaps in understanding of tidal basin sediment dynamics*

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45 433 Sediment dynamics on tidal flats are extremely complex and not fully understood yet.
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47 434 Attempts to assess them in a spatio-temporal context are rare and usually subject to high
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49 435 uncertainties (Widdows et al., 2004; Rahbani, 2011). In particular, the factors that determine
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9 436 the grain size distribution on the tidal flats are not sufficiently known. This leads to the inability
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11 437 to accurately estimate sediment resuspension and settling velocities (Wang et al., 2012).
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13 438 Furthermore, the actual processes of sediment resuspension and flocculation are still highly
14
15 439 uncertain for sand-mud mixtures as well as the influence of biological parameters that show a
16
17 440 strong spatial and temporal variability. For modelling the morphological development of salt
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19 441 marshes, these uncertainties pose a serious challenge, particularly because the long-term
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21 442 sediment availability, its short-term temporal variations, and the sediment composition are
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23 443 often unknown (Andersen et al., 2011).
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27 28 445 **V The sedimentary system of the Wadden Sea**

29 30 446 31 32 447 *1 The concept of a sand-sharing system* 33 34 448

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36 449 The morphological elements of the Wadden Sea system can be described as a sand-sharing
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38 450 system (Fig. 6), which distributes the sediment within the system according to prevailing
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40 451 hydrodynamics and morphological conditions (CPSL, 2001). Through redistribution of the
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42 452 sediment within the sand-sharing system, sediment required for the tidal flats to adapt to SLR
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44 453 is removed from the ebb-tidal deltas and transported into the tidal basins by tide or wind
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46 454 induced currents, and is therefore potentially available for deposition on the salt marshes
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48 455 (Elias et al., 2007; Hofstede, 1999a). Wind and wave activity significantly modify the sediment
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50 456 exchange between the ebb-tidal delta and the tidal basin by reducing the size of the ebb-tidal
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9 457 delta through erosion and an increased shoreward sediment flux into the tidal basin (Hofstede,
10 458 1999b; Walton and Adams, 1976). The incoming sediment is distributed through (sub-)tidal
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12 459 channels, from where adjacent tidal flats and salt marshes are inundated and supplied with
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14 460 allochthonous sediment.
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19 462 *2 Fine and coarse-grained sediment transport into the tidal basins*
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23 464 As a consequence of the shore-normal energy gradient (i.e. decreasing current velocities
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25 465 towards the inner tidal basin) and the so-called settling and scour lags (van Straaten and
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27 466 Kuenen, 1957; Postma, 1961), fine-grained sediments are transported further into the tidal
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29 467 basins, whereas the coarse-grained sediments tend to settle in the vicinity of tidal inlets
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31 468 (Flemming and Nyandwi, 1994). The settling and scour lags are resulting from the transport of
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33 469 the fine-grained sediment during its settling phase and the higher current velocity needed to
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35 470 resuspend a sediment particle from the seabed than to deposit it (van Straaten and Kuenen,
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37 471 1957; Postma, 1961). These processes are considered as the most important ones for a net
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39 472 import of fine-grained sediments into the tidal basins of the Wadden Sea (van Straaten and
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41 473 Kuenen, 1958).
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45 474 Additionally, the distortion of the tidal wave in coastal areas and the density-driven currents
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47 475 between the North Sea and the Wadden Sea as well as the (de-)stabilizing effect of physical
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49 476 sediment properties and benthic organisms are controlling the net amount of sediment
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51 477 accumulating within the tidal basins (CPSL, 2010). When travelling in shallow water or along an
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9 478 estuary or tidal basin, the distortion of the tidal wave induces shorter but stronger flood
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11 479 currents and longer but weaker ebb currents (Dronkers, 1986; van der Spek, 1997). Stronger
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13 480 flood currents (in comparison to the respective ebb currents) promote the import of coarse-
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15 481 grained sediment via bed-load and/or suspended transport (Dronkers, 1986; van der Spek,
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17 482 1997; van Kreeke and Hibma, 2005; Wang et al., 2012). Meanwhile, an elongated high slack
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19 483 water period (in comparison to the previous and the following low slack water periods) may
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21 484 enhance the net import via suspended load transport and the sedimentation of fine-grained
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23 485 sediments (Dronkers, 1986; van Straaten and Kuenen, 1958; van Kreeke and Hibma, 2005). In
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25 486 enclosed, gently sloping tidal basins with most of the tidal flats located below MSL, the
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27 487 distortion of the tidal wave results in a larger import of fine-grained sediments compared to
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29 488 open-shaped, steeply sloping tidal basins with most of the tidal flats located above MSL
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31 489 (Dronkers, 1986; Pedersen and Bartholdy, 2006). The import of coarse-grained sediments is
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33 490 smaller or even negative in tidal basins with large tidal flat areas and larger in tidal basins with
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35 491 smaller tidal flat areas (Dronkers, 1986).
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37 492 Due to these different drivers for the import of fine and coarse grained sediments, contrasting
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39 493 fluxes may occur within a tidal basin during a single tidal cycle (van Kreeke and Hibma, 2005). A
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41 494 net export of coarse-grained sediment in tidal basins with large tidal flat areas, for example,
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43 495 may be balanced or exceeded by the import of suspended fine-grained sediments
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45 496 (Ridderinkhof, 1997). Usually (during calm weather conditions), the suspended load import of
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47 497 fine-grained sediment dominates over the net import of coarse-grained sediments (Wang et
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49 498 al., 2012). During strong-wind and storm events, the import of coarse-grained sediments via
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9 499 suspended load transport considerably increases, thereby exceeding the import of fine-grained
10 500 sediments (Santamarina Cuneo and Flemming, 2000; Van Goor et al., 2003).

11 501 Seasonal variations of the total sediment import into the tidal basin of the Wadden Sea are
12
13 502 induced by increased bioaggregation during spring and early summer algae blooms (Alldredge
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15 503 and Gotschalk, 1989) as well as through biodepositon of filter feeding organisms, which are
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17 504 more active during the warm summer months (Andersen and Pejrup, 2002). Density-driven
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19 505 currents, induced by higher variations of fresh water input into the tidal basins and larger
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21 506 seasonal temperature variations, may amplify the seasonal variations of the fine-grained
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23 507 sediment import, although its relevance for the Wadden Sea is under debate (van Beusekom et
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25 508 al., 2012; Wang et al., 2012).

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32 510 *3 Sediment sources*

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36 512 External sources of predominantly coarse-grained sediments other than the neighbouring tidal
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38 513 basins are found on the foreshore, the seaward beaches, and in the dunes of the barrier
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40 514 islands (CPSL, 2010). Most fine-grained sediments are supplied from riverine inputs such as the
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42 515 Rhine or the Elbe River as well as from soft-rock erosion on the English East Coast (Gayer et al.,
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44 516 2006) and presumably from ancient river valleys now located offshore (Dellwig et al., 2000).
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46 517 Other sources of fine-grained sediment are atmospheric deposition, primary production, direct
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48 518 fluvial input, and sediment from salt marsh erosion (Pedersen and Bartholdy, 2006; Pejrup et
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9 519 al., 1997). These sources strongly determine the availability of fine-grained sediments in the
10 520 coastal North Sea and the import into the tidal basins of the Wadden Sea.

11 521 For most tidal basins in the Wadden Sea, sediment import from the coastal North Sea is the
12 522 largest contribution to the total fine-grained sediment inventory (Dellwig et al., 2000).

13 523 Pedersen and Bartholdy (2006), for example, found the import to vary between 52 to 82% in
14 524 four Danish tidal basins. Estimated net imports of fine-grained sediment vary between 0.10 g
15 525 m⁻³ (List tidal basin) to 0.53 g m⁻³ (Grådyb basin) for an average tidal cycle within a distance of
16 526 only 60 km, for example (Pedersen and Bartholdy, 2006).

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18 528 *4 Trapping capacity of the tidal basins*

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20 530 An important parameter influencing the net sediment transport into the tidal basin is their
21 531 trapping capacity. It is generally controlled by bed roughness and sediment stability, which are,
22 532 to a large extent, determined by biological activity. Especially where vegetation canopies such
23 533 as seagrass beds or salt marshes are present, but also where mussel beds and other reef-
24 534 building epibenthic organisms are found, currents are slowed down, sediment resuspension is
25 535 inhibited, and sediment accretion is enhanced. At the same time the sediment stability is
26 536 increased due to the binding forces of the roots and the presence of biofilms inhibiting
27 537 sediment resuspension (CPSL, 2010; Ward et al., 1984). Increased sediment stability
28 538 additionally amplifies the scour lag effect and thereby contributes to the import of fine-grained
29 539 sediments (Vos and van Kesteren, 2000).

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10 541 *5 Knowledge gaps in understanding the sedimentary system of the Wadden Sea*

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15 543 Looking into the future development of the salt marshes in the Wadden Sea, we identify the

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17 544 availability of fine-grained sediments and their settling velocities on salt marshes to be

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19 545 crucially important. These parameters largely depend on the amount and composition of

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21 546 sediments being imported into the tidal basins and their erodibility, once deposited on the

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23 547 tidal flats. Given the projected SLR for the Wadden Sea region, insufficient knowledge is

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25 548 available about how the tidal flats will evolve in the future with respect to their elevation and

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27 549 composition. One of the main reasons for this is the current practice of assessing the sediment

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29 550 budgets of the different tidal basins per sediment fraction, rather than looking at the

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31 551 combined cohesive and non-cohesive sediment budgets (Wang et al., 2012). Future modelling

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33 552 approaches will need to concentrate on the total sediment budget, thereby integrating the

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35 553 cohesive and non-cohesive sediments and their grain sizes. They additionally need to account

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37 554 for the morphological differences between the various tidal basins and their trapping

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39 555 efficiencies that are strongly controlled by the variable biological activity.

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41 556 In this review, we studied the morphological processes within tidal basins consisting of

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43 557 channels, tidal flats and salt marshes (Fig. 6). However, it should be noted that in large tidal

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45 558 basins, subtidal flats comprise large areas (up to 60%). It is likely that the morphological

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47 559 behaviour in this deeper zone is not the same as described for the intertidal areas. Hardly any

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49 560 research has been performed on interactions between subtidal and intertidal flats on the

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9 561 sediment budgets of the tidal basins, but doing so could possibly further improve the
10 562 understanding of the sedimentary system of the Wadden Sea.

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15 564 **VI An integrated conceptual model for the Wadden Sea salt marshes**

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19 566 By evaluating the available literature on hydromorphological processes directly and indirectly
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21 567 affecting the development of the Wadden Sea salt marshes, we identify important interactions
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23 568 between sedimentary processes taking place on salt marshes as well as on tidal flats and the
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25 569 entire tidal basins. Knowledge gaps are identified that we consider responsible for the limited
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27 570 ability of salt marsh models to predict the future development of the Wadden Sea salt
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29 571 marshes. Partly, the existence of these knowledge gaps is due to the lack of reliable empirical
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31 572 data, but also due to the high degree of complexity of the involved processes. However, for
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33 573 estimating the future development of coastal salt marshes under the influence of projected
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35 574 global sea level rise, these processes have to be accounted for in salt marsh modelling, since
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37 575 they drive the temporal and spatial variability of sediment supplied to the salt marshes.

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39 576 In order to overcome this issue and bring forward the development of salt marsh models, we
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41 577 propose a conceptual model (Fig. 7) that integrates the most important interactions between
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43 578 the fringing salt marshes and the neighbouring tidal flats and tidal basins. This model aims to
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45 579 reduce the complexity of the system by identifying the processes that are directly responsible
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47 580 the quantity and quality of the sediment supplied to the salt marshes. It draws the sediment
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9 581 pathways that are relevant for the development of salt marshes and clarifies their role within
10 582 the morphological system of the Wadden Sea.

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15 584 *1 Description of the conceptual model*

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19 586 Sediment import during calm weather periods is dominated by the import of fine-grained
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21 587 sediments via suspended sediment load and a minor contribution of coarse-grained sediment
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23 588 import via bed-load and/or suspended load transport (Fig. 7, 2). When onshore winds increase
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25 589 (inducing set-up of the water level), more coarse-grained sediments are imported via
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27 590 suspended sediment load, exceeding the import of fine-grained sediments during storm events
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29 591 (Fig. 7, 1). While flooding the tidal flats, parts of these suspended sediments are deposited on
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31 592 the tidal flats (Fig. 7, 5). The net sediment accumulation is mainly determined by the
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33 593 topography and the bed roughness of the tidal flats, which strongly depend on the physical
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35 594 sediment characteristics as well as the epibenthic structures on the tidal flats (Fig. 7, 3).
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37 595 Similarly, the physical sediment characteristics of the tidal flats and the biological activity on
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39 596 them determine the erodibility of the tidal flat sediments (Fig. 7, 3+4). The eroded sediment
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41 597 from the tidal flats and from the marsh scarp is either exported from the tidal basin via the
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43 598 tidal channels (Fig. 7, 5) or transported onto the salt marsh during strong wind or storm events
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45 599 (Fig. 7, 6). The composition of this sediment is rather fine, but depends on the characteristics
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47 600 of the tidal flat and the prevailing hydrodynamic conditions. Meanwhile, a direct input of
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49 601 sediment from the tidal channel towards the salt marsh is predominantly occurring during
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9 602 strong wind and storm events (Fig. 7, 7), whereas the composition of this sediment may be
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11 603 considerably coarser than the sediment that is eroded from the tidal flats. Once transported
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13 604 onto the salt marsh, a spatial pattern regarding the suspended sediment characteristics is
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15 605 observed (Fig. 7, 8+9). More and coarser sediments are found towards the seaward part of the
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17 606 marsh or in vicinity of the marsh creeks (Fig. 7, 8) compared to the inner part of the marsh,
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19 607 where less and finer sediments are found in suspension (Fig. 7, 9). The differences in
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21 608 suspended sediment characteristics induce higher settling velocities in the seaward part of the
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23 609 marsh, whereas these may vary vertically. Due to the higher turbulence level above the
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25 610 vegetation, flocculation is promoted, while the higher flow velocities allow the settlement of
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27 611 only the largest flocs/particles. In contrast, the floc size within the vegetation is smaller due to
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29 612 the lower turbulence level, but strongly reduced flow velocities allow the small particles and
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31 613 flocs to settle (Fig. 7, 8+9).
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615 2 Suggested use of the proposed conceptual model

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617 The proposed conceptual model has been developed in order to reduce the complexity of
618 representing the involved processes in the (sub-)tidal sediment dynamics and thereby improve
619 the estimation of boundary conditions and model parameters for salt marsh models. An
620 implementation of this conceptual model includes the coupling of existing models that
621 specifically address the different processes as shown in figure 7. Such a coupled modelling
622 approach could be used to investigate how changes in the large-scale sedimentary system of

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9 623 the Wadden Sea would affect the development of the fringing salt marshes. Furthermore, the
10 624 model could be used to estimate the influence of long-term (SLR) as well as short-term sea
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12 625 level variations (storm activity) on the future survival of the salt marshes. Besides the region of
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14 626 the Wadden Sea, the model could also be applied to comparable sedimentary systems, such as
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16 627 coastal lagoons with active inlets behind a coastal barrier.
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18 628 Based on the proposed conceptual model, an integrated modelling approach could more
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20 629 efficiently capture the process interactions within the tidal basin - tidal flat - salt marsh
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22 630 continuum and thereby overcome some of the above identified knowledge gaps. For example,
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24 631 it could investigate and, to some extent quantify, the influence of a gradual increase of global
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26 632 and regional MSL in the coming century (Meehl et al., 2007; Vermeer and Rahmstorf, 2009;
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28 633 Wahl et al., 2010) on the sediment availability for coastal salt marshes as described in the
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30 634 following example:
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32 635 A gradual SLR increases the demand for fine and/or coarse-grained sediments on the tidal flats
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34 636 and the demand for fine-grained sediments for the salt marshes to keep pace with SLR. The
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36 637 fine-grained sediment budget will be affected by a modified distortion of the tidal wave within
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38 638 the tidal basin. Accelerated sea level rise with tidal flats lagging behind in net accretion would
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40 639 increase the proportion of shallow subtidal areas at the expanses of intertidal flats. A modified
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42 640 tidal wave distortion could favour the import of fine-grained sediments during calm weather
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44 641 conditions, increasing, in turn, the sediment availability for the fringing salt marshes. The
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46 642 single-grain settling velocities might decrease and the flocculation ability of the sediment
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48 643 particles might increase correspondingly.
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9 644 In contrast, increasing storm surge heights and a possible moderate increase of storm activity
10 645 (Weisse et al., 2006; Weisse et al., 2012; Woth et al., 2006), could lead to a stronger net export
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12 646 of fine-grained sediments. In the long-term, this could amplify the trend towards coarser
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14 647 sediments in the Wadden Sea and decrease the sediment availability for salt marshes. With a
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16 648 coarser sediment fraction available, the single-grain settling velocities would increase and the
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18 649 flocculation ability of the sediment would decrease.
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24 651 **VII Conclusions**

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28 653 The above conceptual model is proposed as a framework for salt marsh modellers in order to
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30 654 facilitate the integration of important interactions between processes taking place within tidal
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32 655 basins, tidal flats, and on salt marshes for estimating future salt marsh development. It
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34 656 graphically displays the following more generic conclusions that we draw from this study:

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36 657 1) Sediment availability for salt marshes depends on the morphology of tidal basins and on
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38 658 whether future SLR will be accompanied by increasing storm activity or not.

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40 659 2) Changing grain-size distribution in response to climate change potentially affects the salt
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42 660 marsh development by modified sediment availability and changes particle/floc settling
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44 661 velocities.

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46 662 3) Flocculation processes on salt marshes are strongly influenced by the available sediment
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48 663 and its grain-size distribution as well as the structural dynamics of salt marsh vegetation and
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50 664 may considerably influence the spatial accretion patterns and the vertical accretion rates.
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9 665 Existing salt marsh models are well capable of identifying important processes dominating salt
10 666 marsh accretion. However, important processes in the tidal basins and on its tidal flats that
11 667 determine the local sediment availability and sediment characteristics are not yet sufficiently
12 668 incorporated.
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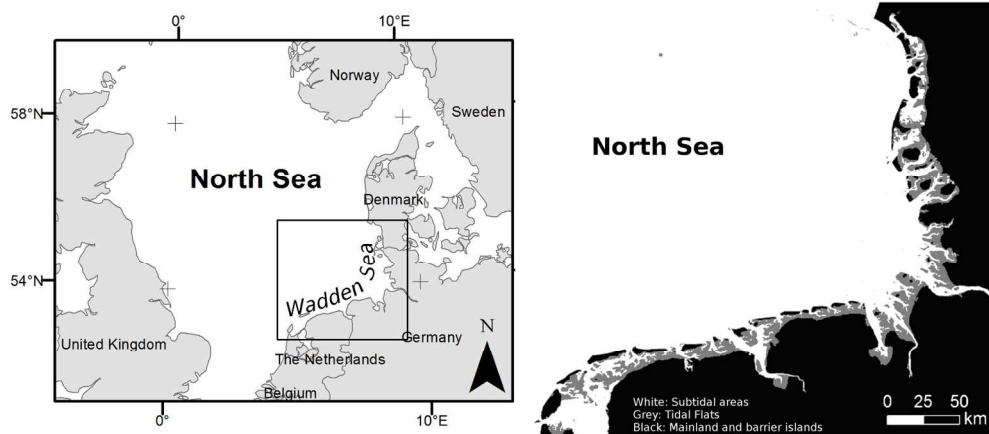


Figure 1: Location of the Wadden Sea along the south-eastern coast of the North Sea (left). The Wadden Sea area consists of the subtidal areas (white), the tidal flats (grey) and the barrier islands and mainland coasts (black) (right).

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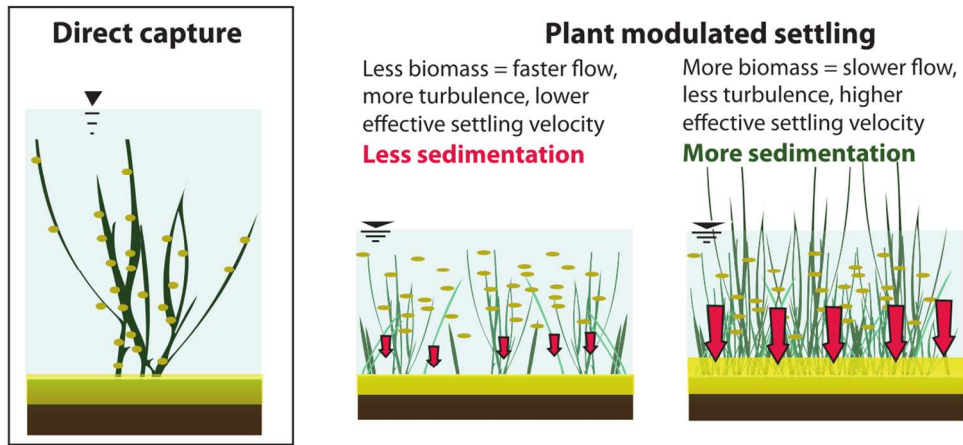


Figure 2: Vegetation effects leading to enhanced allochthonous marsh accretion. Source: Fagherazzi et al., 2012. 120x57mm (300 x 300 DPI)

Peer Review

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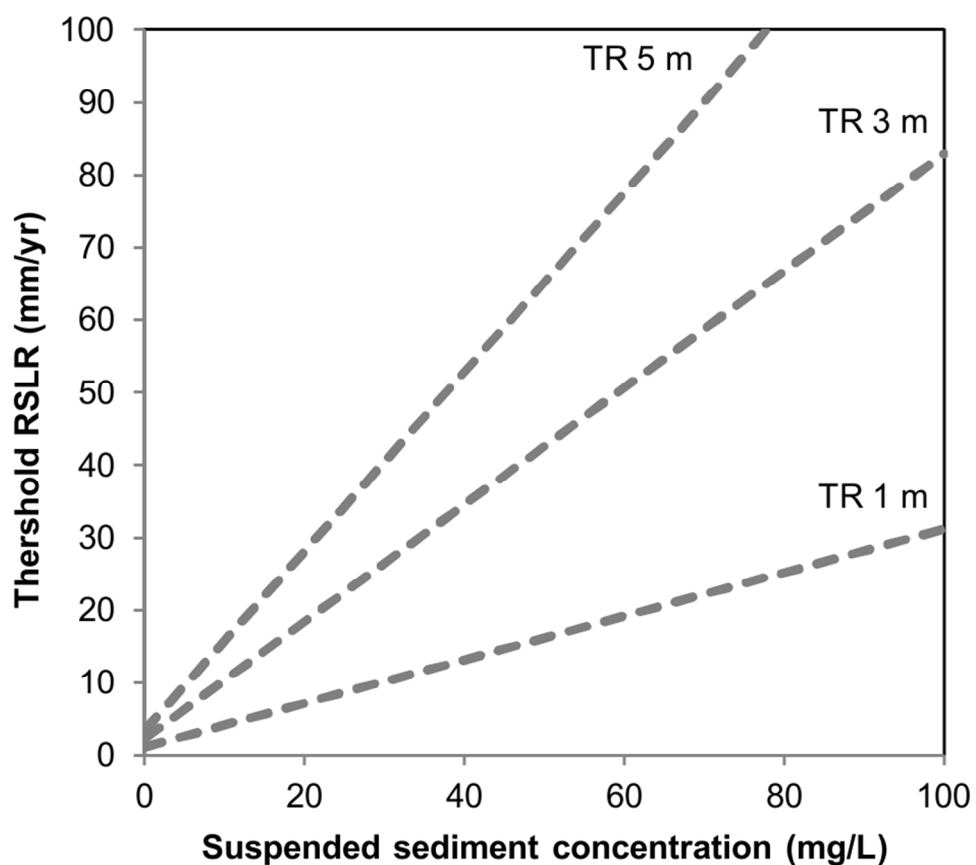


Figure 3: Threshold relative SLR rates for the survival of salt marshes as a function of tidal range (TR) and local sediment availability (suspended sediment concentration), modelled with five different salt marsh models. Dashed lines indicate three different tidal regimes (1 m, 3 m, and 5 m). Source: Modified after Kirwan et al., 2010.

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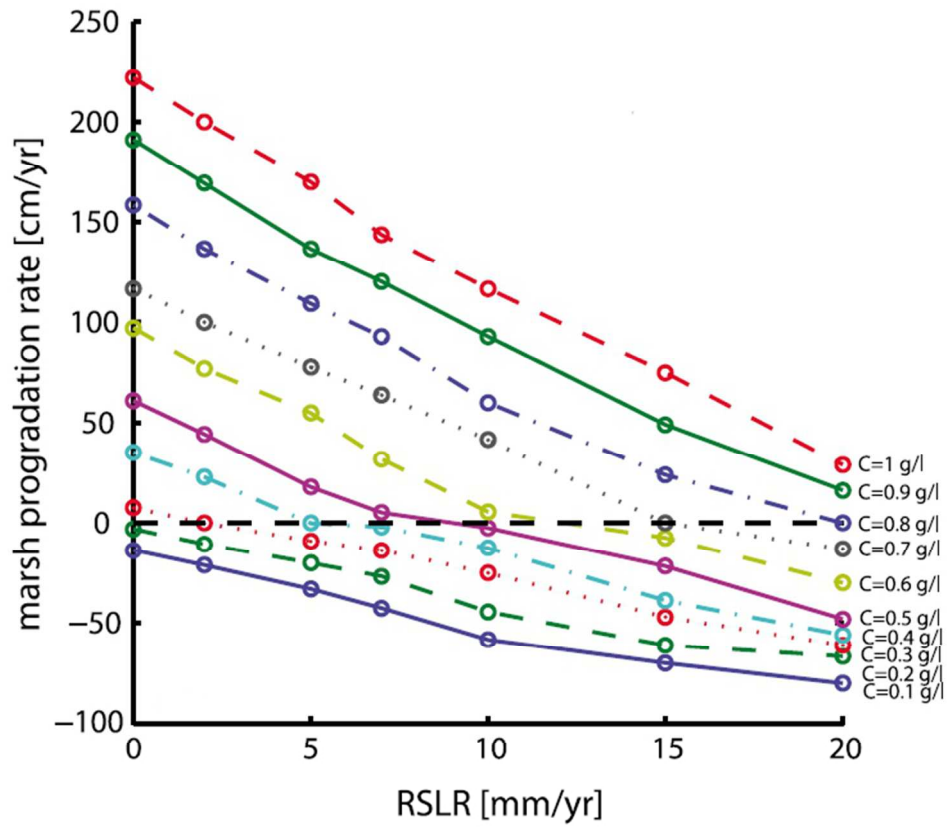


Figure 4: Lateral marsh dynamics in response to RSLR and sediment availability (C). Source: Mariotti and Fagherazzi, 2010.
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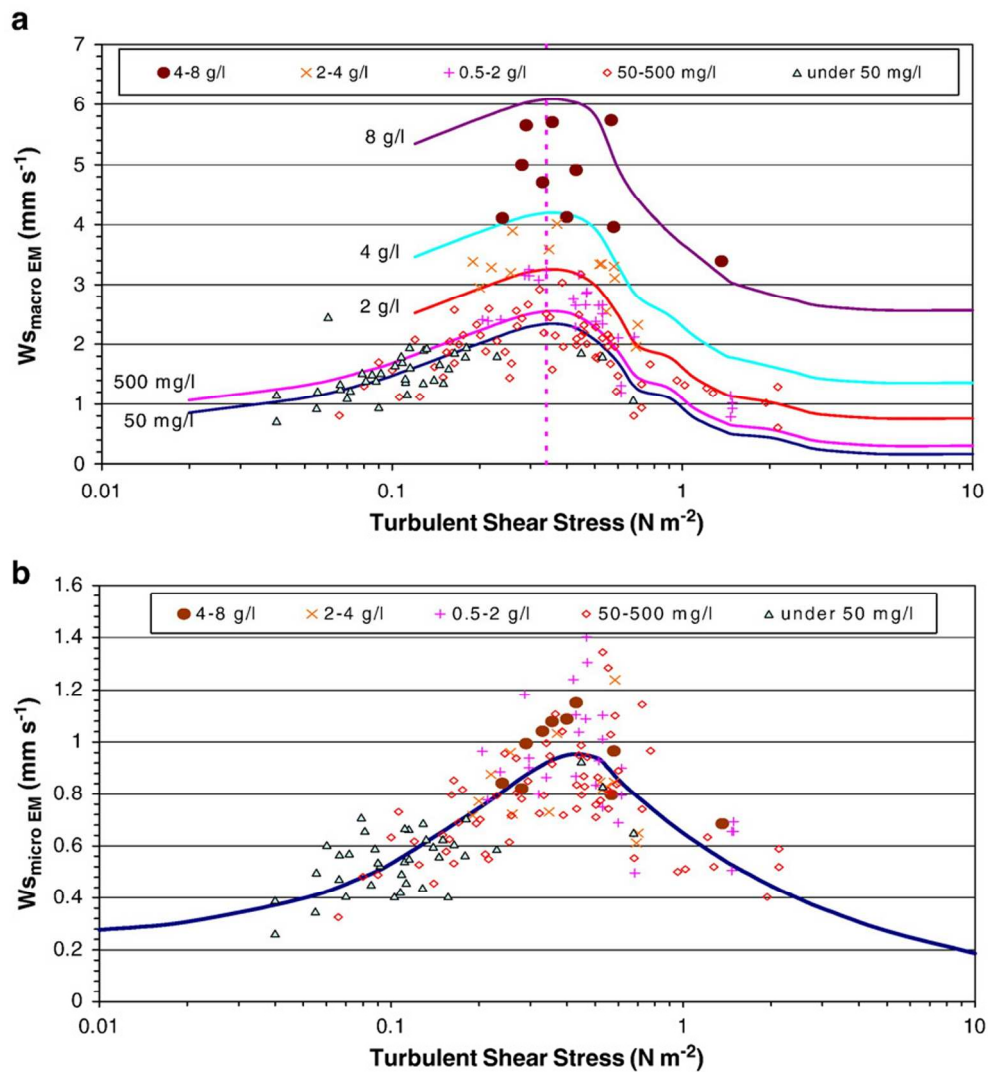


Figure 5: Floc settling velocity (w_s) as a function of turbulent shear stress and suspended sediment concentrations for macro-flocs (a) and micro-flocs (b) assessed for several European estuaries. Source: Manning and Dyer, 2007.
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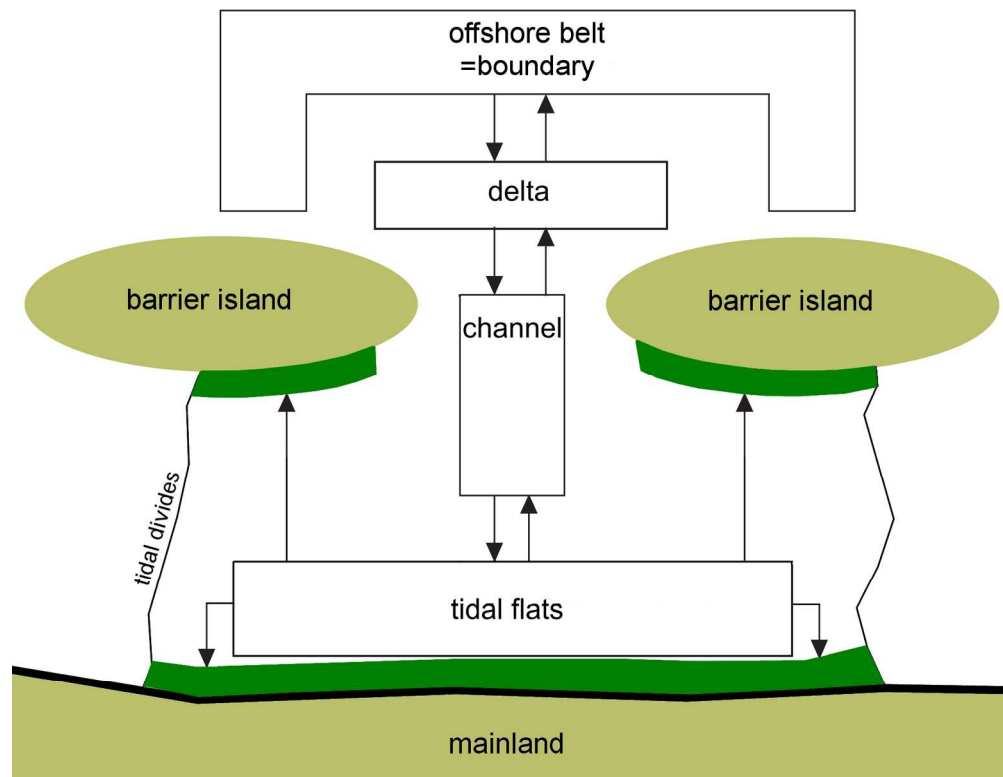
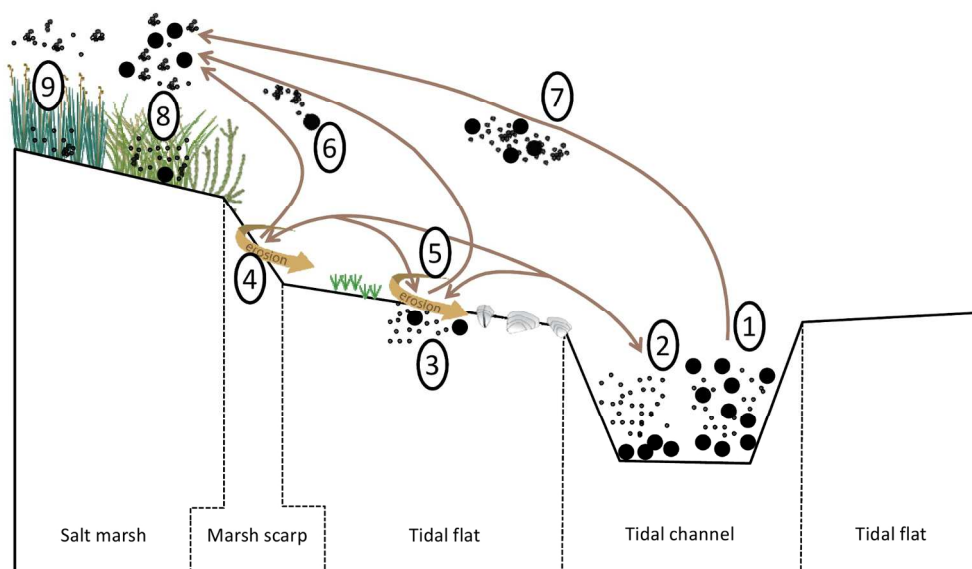


Figure 6 (modified after Kragtwijk et al., 2004): Morphological elements of the sand-sharing system for one tidal basin, as schematized by the "Aggregate scale morphodynamic model of integrated coastal systems" (ASMITA) model. The fringing salt marshes are added as an additional morphological element that act as a sediment sink for the tidal basin. Source: Modified after Kragtwijk et al., 2004.
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- 1) Sediment influx through tidal channel during storm events (coarser sediment)
- 2) Sediment influx through tidal channel during calm periods (finer sediment)
- 3) Sediment resuspension on tidal flats affected by physical (e.g. grain size) and biological (benthic flora and fauna) sediment properties.
- 4) Sediment resuspension at marsh edge (controlled by physical and biological parameters as well as the slope and height of the scarp)
- 5) Sediment exchange between tidal channel, tidal flat, and salt marsh
- 6) Transport of sediments resuspended from the tidal flats and the marsh edge towards the salt marsh
- 7) Direct sediment transport from tidal inlet to salt marsh during storm events
- 8) Marsh accretion in seaward part of the marsh (more and coarser sediment)
- 9) Marsh accretion in landward part of the marsh (less and finer sediment)

Figure 7: Conceptual model of sediment dynamics between the different morphological units in the Wadden Sea (tidal channel, tidal flats, salt marshes; subtidal flats have been omitted because of gap in knowledge).
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