



Zhang, Y., Jin, F., Shen, Z., Wang, F., Lynch, R. and Al-Tabbaa, A. (2019) Adsorption of Methyl tert-butyl ether (MTBE) onto ZSM-5 zeolite: fixed-bed column tests, breakthrough curve modelling and regeneration. *Chemosphere*, 220, pp. 422-431. (doi:[10.1016/j.chemosphere.2018.12.170](https://doi.org/10.1016/j.chemosphere.2018.12.170))

There may be differences between this version and the published version. You are advised to consult the publisher's version if you wish to cite from it.

<http://eprints.gla.ac.uk/176679/>

Deposited on: 3 January 2019

Enlighten – Research publications by members of the University of Glasgow
<http://eprints.gla.ac.uk>

Manuscript Number: CHEM57356R1

Title: Adsorption of Methyl tert-butyl ether (MTBE) onto ZSM-5 zeolite: Fixed-bed column tests, breakthrough curve modelling and regeneration

Article Type: Research paper

Section/Category: Treatment and Remediation

Keywords: MTBE, ZSM-5 zeolite, Fixed-bed column tests, Permeable reactive barriers, Regeneration

Corresponding Author: Miss Yunhui Zhang,

Corresponding Author's Institution: University of Cambridge

First Author: Yunhui Zhang

Order of Authors: Yunhui Zhang; Fei Jin, PhD; Zhengtao Shen, PhD; Fei Wang, PhD; Rod Lynch, PhD; Abir Al-Tabbaa, PhD

Abstract: ZSM-5, as a hydrophobic zeolite, has a good adsorption capacity for methyl tert-butyl ether (MTBE) in batch adsorption studies. This study explores the applicability of ZSM-5 as a reactive material in permeable reactive barriers (PRBs) to decontaminate the MTBE-containing groundwater. A series of laboratory scale fixed-bed column tests were carried out to determine the breakthrough curves and evaluate the adsorption performance of ZSM-5 towards MTBE under different operational conditions, including bed length, flow rate, initial MTBE concentration and ZSM-5 dosage, and regeneration tests were carried out at 80, 150 and 300°C for 24 h. Dose-Response model was found to best describe the breakthrough curves. MTBE was effectively removed by the fixed-bed column packed with a ZSM-5/sand mixture with an adsorption capacity of 31.85 mg·g⁻¹ at 6 cm bed length, 1 mL·min⁻¹ flow rate, 300 mg·L⁻¹ initial MTBE concentration and 5% ZSM-5 dosage. The maximum adsorption capacity increased with the increase of bed length and the decrease of flow rate and MTBE concentration. The estimated kinetic parameters can be used to predict the dynamic behaviour of column systems. In addition, regeneration study shows that the adsorption capacity of ZSM-5 remains satisfactory (>85%) after up to four regeneration cycles.

1 **Adsorption of Methyl tert-butyl ether (MTBE) onto ZSM-**
2 **5 zeolite: Fixed-bed column tests, breakthrough curve**
3 **modelling and regeneration**

4 *Yunhui Zhang^{a*}; Fei Jin^b; Zhengtao Shen^c; Fei Wang^d; Rod Lynch^a; Abir Al-Tabbaa^a*

5

6 ^aDepartment of Engineering, University of Cambridge, Cambridge, CB2 1PZ, United Kingdom

7 ^bSchool of Engineering, University of Glasgow, Glasgow, G12 8QQ, United Kingdom

8 ^cDepartment of Earth and Atmospheric Sciences, University of Alberta, Edmonton T6G 2E3, Canada

9 ^dInstitute of Geotechnical Engineering, School of Transportation, Southeast University, Nanjing, 210096,
10 China

11

12

13

14

15

16

17 **AUTHOR INFORMATION**

18 ***Corresponding Author**

19 Tel: +44- (0) 7821464199

20 E-mail address: yz485@cam.ac.uk.

21

22 **Abstract**

23 ZSM-5, as a hydrophobic zeolite, has a good adsorption capacity for methyl tert-butyl ether
24 (MTBE) in batch adsorption studies. This study explores the applicability of ZSM-5 as a
25 reactive material in permeable reactive barriers (PRBs) to decontaminate the MTBE-
26 containing groundwater. A series of laboratory scale fixed-bed column tests were carried out
27 to determine the breakthrough curves and evaluate the adsorption performance of ZSM-5
28 towards MTBE under different operational conditions, including bed length, flow rate, initial
29 MTBE concentration and ZSM-5 dosage, and regeneration tests were carried out at 80, 150
30 and 300 °C for 24 h. Dose-Response model was found to best describe the breakthrough
31 curves. MTBE was effectively removed by the fixed-bed column packed with a ZSM-5/sand
32 mixture with an adsorption capacity of 31.85 mg·g⁻¹ at 6 cm bed length, 1 mL·min⁻¹ flow rate,
33 300 mg·L⁻¹ initial MTBE concentration and 5% ZSM-5 dosage. The maximum adsorption
34 capacity increased with the increase of bed length and the decrease of flow rate and MTBE
35 concentration. The estimated kinetic parameters can be used to predict the dynamic behaviour
36 of column systems. In addition, regeneration study shows that the adsorption capacity of
37 ZSM-5 remains satisfactory (>85%) after up to four regeneration cycles.

38 **Key words:** MTBE, ZSM-5 zeolite, Fixed-bed column tests, Permeable reactive barriers,
39 Regeneration

40

41 **1. Introduction**

42 Gasoline spills from the accidental leakage of storage tanks, transfer pipes and boats are
43 typical pollution sources of soil, groundwater, surface water and the marine environment
44 (Reuter et al., 1998). Methyl tert-butyl ether (MTBE) was an extensively used gasoline
45 additive for fuel oxygenation. In spite of the bans in some countries, it is still the second most
46 common volatile organic compound in shallow groundwater (Levchuk et al., 2014). Due to
47 its genotoxicity, its hazard of causing skin and eye irritation and its unpleasant odour
48 (Mancini et al., 2002), the existence of MTBE in aquatic environments has raised
49 considerable public concerns.

50

51 Permeable reactive barriers (PRBs) is an effective in-situ technology for aquifer and
52 groundwater remediation (Hou et al., 2014). Due to the rapid migration (Levchuk et al., 2014)
53 and limited natural biodegradation potential of MTBE (Lindsey et al., 2017; Mohebbali, 2013),
54 using PRBs to mitigate/eliminate MTBE contamination holds much promise. As the key
55 component of PRBs, the reactive medium is selected primarily depending on the nature of
56 target contaminants and the hydro-geological conditions of field sites. **ZSM-5 as a reactive**
57 **medium in the PRBs can act as adsorbents due to its high adsorption capacity (Abu-Lail et al.,**
58 **2010; Martucci et al., 2015; Zhang et al., 2018b) and hydrogen form of ZSM-5 (HZSM-5)**
59 **may also catalyse the hydrolysis of MTBE to *t*-butyl alcohol (TBA) and methanol which are**
60 **more biodegradable (Centi et al., 2002; Knappe and Campos, 2005). These products can also**
61 **be adsorbed onto ZSM-5 and be released slowly with time which favours their**
62 **biodegradation by microorganisms growing on the barrier (Centi and Perathoner, 2003). The**
63 PRBs design requires a kinetic characterisation using fixed-bed columns as a simulation of
64 real PRBs to evaluate the dynamic removal of contaminants for the practical application
65 (Cruz Viggi et al., 2010; Gavaskar, 1999). Various theoretical models, such as Logit, Adams-

66 Bohart, Thomas, Yoon and Nelson, Dose and Response, and bed length/service time (BDST)
67 models, have been developed to fit the experimental data and obtain the breakthrough curves
68 and column kinetic parameters. These parameters can be employed to predict the adsorption
69 performance under new operational conditions and further facilitate the full-scale design of
70 fixed-bed column systems, e.g., PRBs.

71

72 To date, fixed-bed column tests have been widely applied to simulate PRBs towards various
73 contaminants, such as heavy metals and dyes (Calero et al., 2009; Han et al., 2008), with
74 different adsorbents such as activated carbon and zeolites (García-Mateos et al., 2015;
75 Ozdemir et al., 2009) etc. Nevertheless, to our best knowledge, limited studies exist on fixed-
76 bed column tests of using ZSM-5 for MTBE removal, especially regarding the influence of
77 operational conditions, such as the bed length, flow rate, inlet adsorbate concentrations and
78 the percentage of the adsorbent on the adsorption behaviour. Abu-Lail et al. (2010) studied
79 the removal of MTBE with three adsorbents including granular ZSM-5 in large and small
80 diameter fixed-bed columns, and evaluated the influence of bed length on the breakthrough
81 curves with the BDST model. It was shown that the granular ZSM-5 with a shorter bed length
82 reached the breakthrough point earlier due to the less mass of adsorbents in the column.
83 However, besides the bed length, other variables, such as flow rate, the MTBE concentration
84 and ZSM-5 dosage, also need to be considered in practical groundwater contamination
85 applications due to the fact that the groundwater flow rate and MTBE concentrations vary in
86 different regions. Therefore, this study discussed the influence of several operational
87 parameters (bed length, flow rate, initial MTBE concentration and ZSM-5 percentage) in
88 fixed-bed column tests. **The parameters obtained from modelling are crucial for PRB design
89 and can be used to guide the application of ZSM-5 as a reactive medium in the PRBs for the
90 MTBE-contaminated groundwater remediation.**

91

92 Reusability is considered as a key criterion to judge the feasibility of an adsorbent in practical
93 applications (Xin et al., 2017). The exhausted adsorbents are generally considered as
94 hazardous wastes and need to be incinerated, leading to secondary pollution, such as thermal
95 pollution and potential desorption of adsorbate in the atmosphere (Shah et al., 2014). The
96 regeneration of spent adsorbents can recover material resources, minimize the demands of
97 virgin adsorbents and avoid the generation of hazardous waste. Zeolites, including ZSM-5,
98 demonstrate good stability under a wide range of environmental conditions, such as acidic
99 (Pascoe, 1992) and high temperature environments (Anderson, 2000). They can be
100 regenerated by heat treatment, chemical treatment, such as Fenton oxidation (Wang and Zhu,
101 2006) and KCl (Katsou et al., 2011), and biological regeneration (Wei et al., 2011). However,
102 chemical or biological methods may lead to the generation of hazardous residues. **Although**
103 **HZSM-5 may adsorb and catalyse the hydrolysis of MTBE, and then release the adsorbed**
104 **reaction products (TBA and methanol) to achieve self-regeneration, this process takes a long**
105 **time (Centi and Perathoner, 2003) and our previous study showed that the desorption was**
106 **negligible after 3 days in batch tests (Zhang et al., 2018b). Further studies will investigate the**
107 **long term behaviour.** Thermal regeneration is effective and time-saving for adsorbents used
108 for volatile and semi-volatile organic compounds, including MTBE, due to its high vapour
109 pressure under normal temperatures and low boiling points. In this study, in order to avoid
110 excessive consumption of materials and secondary pollution, repeated thermal regeneration
111 was used for the regeneration of ZSM-5 to evaluate the stability of ZSM-5 after several
112 adsorption-desorption cycles.

113

114 This study aims to (1) analyse the effects of various operational conditions (flow rate, bed
115 length, initial MTBE concentration and ZSM-5 percentage) in fixed-bed column tests on the

116 MTBE adsorption onto ZSM-5; (2) find the most suitable model to describe the
117 breakthrough curve and obtain column parameters; (3) predict adsorption performance at a
118 new flow rate without further experimental runs with the BDST model and (4) examine the
119 recyclability of ZSM-5 with repeated thermal regeneration tests.

120

121 **2. Materials and methods**

122 2.1 Materials

123 MTBE was purchased from Fisher Scientific, and hydrogen form of ZSM-5 powder was
124 obtained from Acros Organics. ZSM-5 used in this study has a large surface area of $400 \text{ m}^2 \cdot \text{g}^{-1}$
125 and a high $\text{SiO}_2/\text{Al}_2\text{O}_3$ ratio of 469. Two pore systems, i.e. zig-zag channels and straight
126 channels, exist in the structure of ZSM-5 with pore sizes of $5.3 \times 5.6 \text{ \AA}$ and $5.1 \times 5.5 \text{ \AA}$,
127 respectively. The detailed physicochemical properties and framework structure of ZSM-5 can
128 be found in (Zhang et al., 2018b).

129

130 2.2 Fixed-bed column tests

131 A series of fixed-bed column tests were conducted in a Pyrex glass column (2 cm inner
132 diameter and 10 cm high) for the simulation of ZSM-5 containing PRBs for MTBE
133 adsorption. There is a layer of glass beads and a stainless steel mesh filter attached to each
134 end of the column to ensure the uniform flow of the solution. The schematic of the fixed-bed
135 column set-up is shown in Figure 1.



Figure 1 The schematic of the fixed-bed column set-up in this study

136

137

138

139 ZSM-5 was mixed with sand to increase the permeability due to the fine texture of ZSM-5

140 (Cappai et al., 2012). Columns were filled with a mixture of ZSM-5 (5% or 10% in w/w of

141 sand) and sand to produce different bed lengths (3, 6 and 9 cm). The initial water content of

142 the specimen was designated as 10% in w/w (of sand) and the bulk density was about $2 \text{ g}\cdot\text{cm}^{-3}$.

143 The values of hydraulic conductivity of the mixture in the column were measured as 6.32

144 $\times 10^{-6} \text{ m}\cdot\text{s}^{-1}$ (5% ZSM-5 and sand) and $1.21 \times 10^{-6} \text{ m}\cdot\text{s}^{-1}$ (10% ZSM-5 and sand). It was

145 assumed that the specific gravity values of sand and ZSM-5 were 2.65 and 2 respectively in

146 this study (Jha and Singh, 2016; Masad et al., 1996). It is noted that the seepage velocity

147 differs at different regions and different depths (Gavaskar et al., 2000). Therefore in this

148 study, pump rate (i.e., seepage rate) was selected based on a previous land remediation

149 project with sandy soils in the ground (Al-Tabbaa and Liska, 2012). The solutions with

150 different MTBE concentrations ($200, 300$ and $400 \text{ mg}\cdot\text{L}^{-1}$) were pumped upward at different

151 flow rates of $0.5 \text{ mL}\cdot\text{min}^{-1}$ (seepage velocity: $0.011 \text{ cm}\cdot\text{s}^{-1}$), $1 \text{ mL}\cdot\text{min}^{-1}$ (seepage velocity:

152 $0.022 \text{ cm}\cdot\text{s}^{-1}$) and $2 \text{ mL}\cdot\text{min}^{-1}$ (seepage velocity: $0.044 \text{ cm}\cdot\text{s}^{-1}$) controlled by a peristaltic pump.

153 Flow rates and MTBE concentrations in this study were higher than the actual conditions in

154 most regions to save the operational time in the lab. Also, the PRBs were generally installed
 155 near pollution sources with a high MTBE concentration.

156

157 The detailed operation variables are listed in Table 1. Where m_{ZSM-5} is the mass of ZSM-5 in
 158 the column (g). The effect of flow rate was studied by tests C, F0.5 and F2; the effect of bed
 159 length was examined by tests C, B3 and B9; tests C and Z10 were conducted to discuss the
 160 effect of ZSM-5 dosage and tests C, M200 and M400 ascertained the effect of initial MTBE
 161 concentration. The effluents at the outlet were collected at set intervals and the MTBE
 162 concentration was measured. The saturation time (t_s) was established when the effluent
 163 MTBE concentration exceeded 85% of inlet concentration. The breakthrough time (t_b) (Goel
 164 et al., 2005) is established when the effluent MTBE concentration reaches 5% of the inlet
 165 concentration ($C/C_0=0.05$) (García-Mateos et al., 2015).

166

167

Table 1 Operational variables for fixed-bed column tests

Test No.	Influencing factors	Flow rate (mL·min ⁻¹)	Bed length (cm)	m_{ZSM-5} (g)	ZSM-5 (%)	C_0 (MTBE) (mg·L ⁻¹)	Porosity
F0.5	Flow rate	0.5	6	2.05	5	300	0.25
C	Flow rate	1	6	2.05	5	300	0.24
F2	Flow rate	2	6	2.05	5	300	0.24
B3	Bed length	1	3	1.03	5	300	0.24
C	Bed length	1	6	2.05	5	300	0.24
B9	Bed length	1	9	3.08	5	300	0.24
C	ZSM-5 dosage	1	6	2.05	5	300	0.24
Z10	ZSM-5 dosage	1	6	4.50	10	300	0.23

M200	MTBE	1	6	2.07	5	200	0.24
	concentration						
C	MTBE	1	6	2.05	5	300	0.24
	concentration						
M400	MTBE	1	6	2.03	5	400	0.25
	concentration						

168

169 2.3 Regeneration cycles

170 The thermal regeneration tests were conducted to examine the recyclability of ZSM-5 at
 171 different temperatures based on batch adsorption tests. After MTBE adsorption in aqueous
 172 solution with ZSM-5, the saturated ZSM-5 was heated at 80, 150 or 300 °C for 24 h in a
 173 muffle furnace (Carbolite CWF 1200, UK), and then 0.1 g of regenerated ZSM-5 was added
 174 to 20 mL 300 mg·L⁻¹ of MTBE solution for adsorption for 24h. After each regeneration cycle,
 175 the MTBE removal percentage was determined and this process was repeated up to 6 times.

176

177 2.4 Analytical methods

178 MTBE concentration was analysed by an ambient headspace technique as described in our
 179 previous studies (Chan and Lynch, 2003; Zhang et al., 2018b) using a gas chromatograph
 180 (Agilent 6850 Series) with a flame ionisation detector (GC-FID). Each headspace sample was
 181 measured in triplicate. Data fitting and modelling was performed by OriginPro 8.5 software.
 182 The values of Akaike information criterion (AIC) and correlation coefficient (R^2) were used
 183 to compare different models. The lower AIC and higher R^2 values indicate a more suitable
 184 model.

185

186 2.5 Mathematical models for breakthrough curves

187 The operational parameters, such as the breakthrough time, saturation time, the shape of
188 breakthrough curves and the column adsorption capacity, play important roles in the
189 evaluation of the operational and adsorption performance of columns. They can be obtained
190 from a plot of C/C_0 against time (t) using the non-linear regression method. Several
191 mathematical models, such as Adams-Bohart model, the Logit method, Thomas model,
192 Yoon-Nelson model and Dose-Response model, have been developed and widely applied to
193 fit the experimental data of column tests to predict the concentration-time profiles and
194 breakthrough curves. Therefore, these models were used in this study to find the most
195 suitable model to describe the breakthrough curve and obtain maximum column capacity.
196 This will help avoid unnecessary investment and high operational costs in the design and
197 operation of a full-scale column caused by possible underutilized or oversaturated columns.

198 2.5.1 Adams-Bohart model

199 The Adams-Bohart model (Bohart and Adams, 1920) was developed based on the assumption
200 that adsorption rate is proportional to the adsorbent's residual capacity and the adsorbate's
201 concentration (Goel et al., 2005). It is generally used to describe the initial portion
202 ($C/C_0 < 0.15$) of the breakthrough curve and has been extensively applied in other various
203 systems (Calero et al., 2009; Sağ and Aktay, 2001). The expression is given as follows

$$204 \frac{C}{C_0} = \frac{e^{k_{AB}C_0t}}{e^{(k_{AB}N_0Z/v)-1} + e^{k_{AB}C_0t}} \quad (1)$$

205 where k_{AB} is the rate constant ($L \cdot mg^{-1} \cdot min^{-1}$) and N_0 is the volumetric adsorption capacity
206 ($mg \cdot L^{-1}$).

207 2.5.2 Bed depth/service time (BDST) model

208 BDST model (Oulman, 1980) is rearranged from the Adams-Bohart model by Hutchins
209 (Hutchins, 1973) to produce a linear relationship between the bed length (Z, cm) and service
210 time (t, min). It is based on the assumption that the moving speed of the adsorption zone in
211 the column is constant, and can be described as follows:

212 $t = \frac{N_0}{C_0 v} Z - \frac{1}{C_0 k_{AB}} \ln \left(\frac{C_0}{C} - 1 \right)$ (2)

213 $a = \frac{N_0}{C_0 v}$ (3)

214 $b = \frac{1}{C_0 k_{AB}} \ln \left(\frac{C_0}{C} - 1 \right)$ (4)

215 The values of N_0 and k_{AB} can be obtained from a plot of Z against t . The advantage of the
 216 BDST model is that only three column tests are required to collect the experimental data
 217 (Adak and Pal, 2006; Hutchins, 1973).

218 For a new operational condition, such as a new linear flow rate (v'), the new slope (a') and
 219 intercept (b') can be calculated directly by Equation (5) and (6), respectively.

220 $a' = a \frac{v}{v'}$ (5)

221 $b' = b$ (6)

222 2.5.3 Logit method

223 BDST model may cause errors if the service time at which the effluent exceeds the
 224 breakthrough criteria was selected. Therefore, Logit method was established to provide a
 225 rational basis for the fitting to the data and the reduction of errors (Oulman, 1980).

226 The equation of the Logit method (Oulman, 1980) can be written as

227 $\ln \left(\frac{\frac{C}{C_0}}{1 - \frac{C}{C_0}} \right) = K C_0 t - \frac{KNZ}{v}$ (7)

228 To apply it to describe the breakthrough curve, Equation (7) is rearranged as

229 $\frac{C}{C_0} = \frac{e^{(K C_0 t - KNZ/v)}}{1 + e^{(K C_0 t - KNZ/v)}}$ (8)

230 where v is the linear flow rate ($\text{cm} \cdot \text{min}^{-1}$), C is the solute concentration ($\text{mg} \cdot \text{L}^{-1}$), C_0 is the
 231 inlet MTBE concentration ($\text{mg} \cdot \text{L}^{-1}$), K is the adsorption rate coefficient ($\text{L} \cdot \text{mg}^{-1} \cdot \text{min}^{-1}$) and N
 232 is the adsorption capacity coefficient ($\text{mg} \cdot \text{L}^{-1}$).

233 2.5.4 Thomas model

234 Thomas model (Equation (9)) based on the mass-transfer theory and was used to calculate the
 235 maximum adsorption capacity (q_0 , $\text{mg}\cdot\text{g}^{-1}$) and the Thomas adsorption rate constant (K_{Th} ,
 236 $\text{L}\cdot\text{mg}^{-1}\cdot\text{min}^{-1}$) using experimental data from fixed-bed column tests (Thomas, 1944, 1948).

$$237 \quad \frac{C}{C_0} = \frac{1}{1 + e^{\frac{k_{\text{Th}}}{Q}(q_0 m - C_0 V)}} \quad (9)$$

238 where V is the effluent volume (L), m is the mass of adsorbent (g), and Q is the flow rate of
 239 the influent ($\text{L}\cdot\text{min}^{-1}$).

240 2.5.5 Yoon and Nelson model

241 The wide use of Yoon and Nelson model (Yoon, 1984) in single adsorbate systems is
 242 attributed to its simplicity since no detailed data is needed regarding the properties of
 243 adsorbate, adsorbent and the column. The equation is given by:

$$244 \quad \frac{C}{C_0} = \frac{1}{1 + e^{k_{\text{YN}}(\tau - t)}} \quad (10)$$

245 where τ is the time required for 50% adsorbate breakthrough (min) and k_{YN} is the rate
 246 constant (min^{-1}). This model assumes that the declining rate in the probability of adsorption is
 247 proportional to that of both adsorbate adsorption and adsorbate breakthrough on the adsorbent
 248 (Ayoob and Gupta, 2007).

249 2.5.6 Dose-Response model

250 Dose-Response model (Yan et al., 2001) is an empirical model and has been widely used to
 251 describe the column kinetics and behaviour, especially heavy metal removal (Dorado et al.,
 252 2014). The general equation is as follows:

$$253 \quad \frac{C}{C_0} = 1 - \frac{1}{1 + \left(\frac{C_0 V}{q_0 m}\right)^a} \quad (11)$$

$$254 \quad b = V_{(50\%)} = \frac{q_0 m}{C_0} \quad (12)$$

255 where a is the constant, b is equal to $V_{(50\%)}$, the concentration when 50% of the maximum
 256 response occurs (L).

257

258 **3. Results and discussion**

259 3.1 Breakthrough curve modelling

260 The concentration-time profiles were obtained after a series of fixed-bed column experiments.
261 Five models were applied to fit the experimental data to describe the fixed-bed column
262 behaviour. The empirical Dose-Response model best described the experimental data in
263 different column conditions ($R^2 > 0.95$ with the lowest AIC value), suggesting its suitability to
264 be used for the design and scale-up purpose. This model was also shown to reduce the errors
265 of two conventional mathematical models, i.e. Thomas model and Adams-Bohart model, for
266 the biosorption of heavy metals in a column (Yan et al., 2001). The fitting results of the
267 Dose-Response model are shown in Table 2 and those of other models are shown in Table S1
268 and Figure S1-S4 in the Appendix.

269
270 From Table 2, the values of q_0 increased with the increase of bed length and the decrease of
271 flow rate, ZSM-5 dosage and initial MTBE concentration. The adsorption capacity (q_0) was
272 calculated as $26.32 \text{ mg}\cdot\text{g}^{-1}$ at 6 cm bed length, $1 \text{ mL}\cdot\text{min}^{-1}$ flow rate, $300 \text{ mg}\cdot\text{L}^{-1}$ initial MTBE
273 concentration and 5% ZSM-5 dosage (Test No. C).

274

275 Table 2 Dose-Response model parameters for the MTBE adsorption on ZSM-5 under
276 different operational conditions

Variables	Test No.	a	b (mL)	q_0 ($\text{mg}\cdot\text{g}^{-1}$)	R^2
Flow rate	C	1.84	179.88	26.32	0.993
	F0.5	3.14	213.16	31.19	0.997
	F2	0.95	90.99	13.32	0.959
Bed length	C	1.84	179.88	26.32	0.993
	B3	1.06	43.46	12.66	0.997

	B9	3.14	294.63	28.70	0.991
ZSM-5 percentage	C	1.84	179.88	26.32	0.993
	Z10	1.45	280.78	18.72	0.971
Initial MTBE concentration	C	1.84	179.88	26.32	0.993
	M200	1.67	232.38	22.45	0.989
	M400	1.23	107.34	21.15	0.969

277

278 3.2 Column parameters calculation

279 The column adsorption capacity of the adsorbent is a critical indicator of column
280 performance and could be calculated from the breakthrough curve. Considering the best
281 fitting results of the Dose-Response model in Session 3.1, all the breakthrough parameters
282 under certain operational conditions were calculated based on the Dose-Response model
283 fitting and are listed in Table 3. Where MTZ is the length of the mass transfer zone (cm),
284 m_{adsorb} is the adsorbed amount of MTBE (mg), m_{total} is the total amount of MTBE through the
285 column (mg), q_e is the equilibrium MTBE uptake, also called column maximum separation
286 capacity ($\text{mg}\cdot\text{g}^{-1}$) (Gouran-Orimi et al., 2018), C_e is the equilibrium MTBE concentration
287 ($\text{mg}\cdot\text{L}^{-1}$), and R is the total MTBE removal percentage (%).

288

289 It is obvious that both the breakthrough time and saturation time increased with the
290 decreasing flow rate and initial MTBE concentration. The same trend was shown when the
291 bed length or ZSM-5 dosage were increased. The maximum column separation capacity is
292 $31.85 \text{ mg}\cdot\text{g}^{-1}$ at 6 cm bed length, $1 \text{ mL}\cdot\text{min}^{-1}$ flow rate, $300 \text{ mg}\cdot\text{L}^{-1}$ initial MTBE
293 concentration and 5% ZSM-5 dosage (Test No. C) in this study. In comparison, the maximum
294 adsorption capacity in batch adsorption tests were calculated as $53.55 \text{ mg}\cdot\text{g}^{-1}$ in our previous
295 study (Zhang et al., 2018b), which almost doubled that in fixed-bed column tests (31.85

396 $\text{mg}\cdot\text{g}^{-1}$). This is mainly due to the insufficient contact time between ZSM-5 and MTBE in
 397 columns (461 min and 24 h for column tests and batch tests, respectively). It should be noted
 398 that both the adsorption capacity (q_0 in Table 2) and the maximum column separation
 399 capacity (q_e in Table 3) decreased with a higher ZSM-5 percentage in spite of a higher
 300 adsorbed amount of MTBE (m_{adsorb} in Table 3). This may be explained by the phenomenon
 301 that ZSM-5 was easier to run away with the MTBE flow with a higher ZSM-5 dosage,
 302 leading to an underestimate of the adsorption capacity, which is a limitation of this study.

303

304 Table 3 Parameters of breakthrough curves for MTBE adsorption on ZSM-5 in fixed-bed

305 columns under different operational conditions

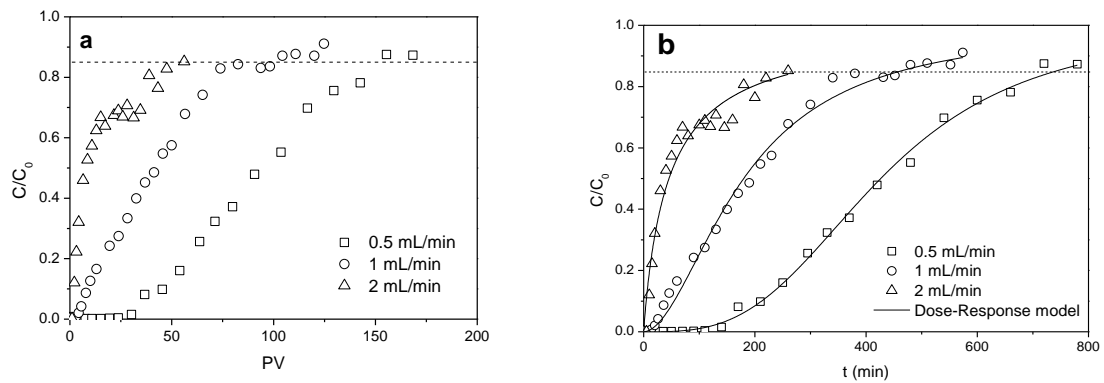
Test No.	C	F0.5	F2	B3	B9	Z10	M200	M400
t_b (min)	36.77	167.87	2.08	2.86	115.28	36.84	40.29	10.13
t_s (min)	460.81	740.18	260.00	220.00	512.25	919.75	655.79	442.04
$MTZ = \frac{Z(t_s - t_b)}{t_s}$ (cm)	5.52	4.64	5.95	2.96	6.97	5.76	5.63	5.86
$m_{\text{adsorb}} =$ $\frac{Q}{1000} \int_{t=0}^{t=t_{\text{total}}} (C_0 - C) dt$ (mg)	65.30	67.42	52.99	23.01	93.23	114.99	52.77	66.52
$m_{\text{total}} = \frac{C_0 Q t_{\text{total}}}{1000}$ (mg)	138.24	111.03	156.00	66.00	153.68	275.93	131.16	176.82
$q_e = \frac{m_{\text{adsorb}}}{m_{\text{ZSM-5}}}$ ($\text{mg}\cdot\text{g}^{-1}$)	31.85	32.89	25.85	22.34	30.27	25.55	25.49	32.77
$C_e = \frac{1000(m_{\text{total}} - m_{\text{adsorb}})}{Q t_{\text{total}}}$ ($\text{mg}\cdot\text{L}^{-1}$)	158.29	117.82	198.10	195.42	118.00	174.98	119.53	249.52
$R = \frac{100 m_{\text{adsorb}}}{m_{\text{total}}}$ (%)	47.24	60.73	33.97	34.86	60.67	41.67	40.23	37.62

306

307 3.3 Influence of operational conditions on MTBE removal

308 3.3.1 Effect of flow rate

309 Figure 2 shows the breakthrough curves at different flow rates of 0.5, 1 to 2 mL·min⁻¹ in
310 relation to pore volume and service time. As shown in Figure 2, the plots were closer to a
311 classic S-shaped breakthrough curve at a lower flow rate (0.5 mL·min⁻¹), indicating a slower
312 process and a higher adsorption capacity (32.89 mg·g⁻¹).



313

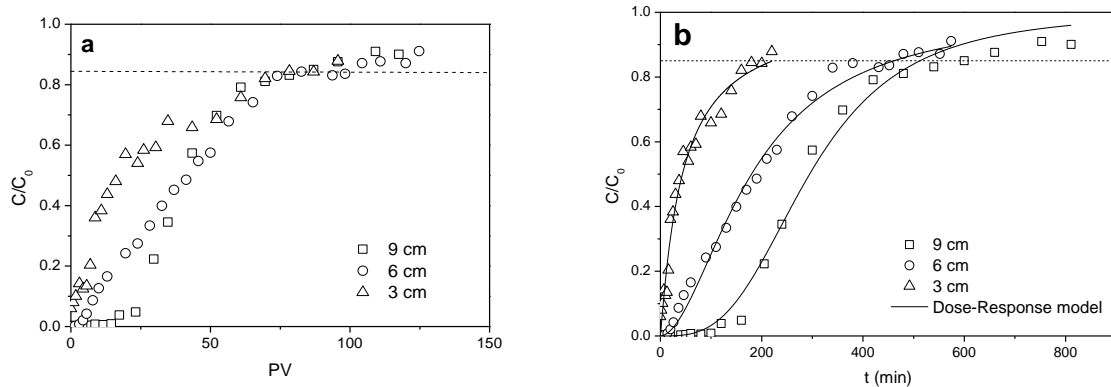
314 Figure 2 Breakthrough curves at different flow rates as a function of (a) pore volume (PV)
315 and (b) time (t). ($C_0=300$ mg·L⁻¹, bed length=6 cm, ZSM-5 dosage=5%)

316

317 As the flow rate increased from 0.5 mL·min⁻¹ to 2 mL·min⁻¹, the breakthrough time and
318 saturation time decreased from 167.87 min to 2.08 min and from 740.18 min to 260.00 min,
319 respectively. A lower column adsorption capacity was obtained at 25.85 mg·g⁻¹ as shown in
320 Table 3. This is due to the fact that the movement of MTBE is accelerated with an increase in
321 the flow rate, which could cause insufficient residence time of MTBE in the column
322 (Ozdemir et al., 2009; Salman et al., 2011). Similar agreement was found for the adsorption
323 of nitrate on bio-inspired polydopamine coated zeolite and was explained by low residency in
324 the column at high flow rate (Gouran-Orimi et al., 2018).

325 3.3.2 Effect of bed length

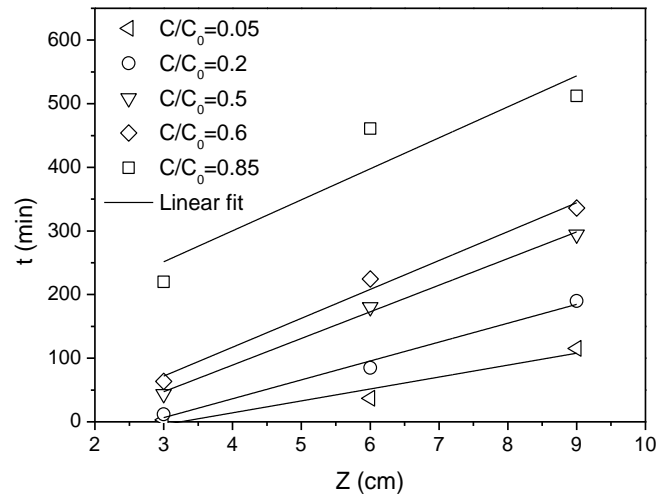
326 The breakthrough profiles at different bed lengths of 3 cm (1.03 g), 6 cm (2.05 g) and 9 cm
 327 (3.1 g) are shown in Figure 3. The decreasing bed length led to a faster breakthrough and
 328 saturation process, which resulted in earlier exhaustion of the bed. The increase in the
 329 breakthrough time could be attributed to the longer distance and moving time of the mass
 330 transfer zone between two ends of the column at a longer bed length (Salman et al., 2011),
 331 which was consistent with the calculated lengths of the mass transfer zone in Table 3. On the
 332 other hand, the increase in the bed length also led to the increasing mass of ZSM-5 and
 333 provided more adsorption sites for MTBE removal. It is noted that, as shown in Table 3, the
 334 increase in bed length gave rise to the increase in the total treated MTBE volume and
 335 saturation time in Figure 3b; however, the amounts of PVs through the column at saturation
 336 time were almost the same for various bed lengths in Figure 3a. This is due to that given the
 337 same flow rate and contaminant concentration, the adsorption capacity per unit bed length is
 338 constant.



339
 340 Figure 3 Breakthrough curves at different bed lengths as function of (a) pore volume (PV)
 341 and (b) time (t). (flow rate=1 mL·min⁻¹, C₀=300 mg·L⁻¹, ZSM-5 dosage=5%) (adapted from
 342 (Zhang et al., 2018a))

343
 344 In addition, the BDST model was applied to produce the plots of Z versus t in Figure 4 for
 345 5%, 20%, 50%, 60% and 85% saturation of the column with good linearity (R²>0.9). The

346 parameters are calculated and listed in Table 4. With the increase of C/C_0 values from 5%
 347 (breakthrough point) to 85% (saturation point), the values of N_0 increased from 1787.80
 348 $\text{mg}\cdot\text{L}^{-1}$ to 4646.93 $\text{mg}\cdot\text{L}^{-1}$, whereas those of K_{AB} decreased from 1.61×10^{-4} to 5.48×10^{-5} $\text{L}\cdot\text{mg}^{-1}$
 349 $\cdot\text{min}^{-1}$.



350
 351 Figure 4 BDST lines at C/C_0 of 0.05, 0.2, 0.5, 0.6, 0.85 with different bed lengths (flow
 352 rate= $1 \text{ mL}\cdot\text{min}^{-1}$, $C_0=300 \text{ mg}\cdot\text{L}^{-1}$, ZSM-5 dosage=5%)

353
 354 Table 4 Calculated parameters of the BDST model for MTBE adsorption on ZSM-5 in the
 355 fix-bed column tests

C/C_0	Equations	$N_0(\text{mg}\cdot\text{L}^{-1})$	$k_{AB}(\text{L}\cdot\text{mg}^{-1}\cdot\text{min}^{-1})$	R^2
0.05	$t=18.74Z-60.78$	1787.80	1.61×10^{-4}	0.900
0.2	$t=29.68Z-82.44$	2831.47	5.61×10^{-5}	0.979
0.5	$t=41.85Z-78.19$	3992.49	0	0.995
0.85	$t=48.71Z+105.44$	4646.93	5.48×10^{-5}	0.755

356
 357 The BDST model parameters are of great use for the scale-up of the adsorption process. For
 358 example, the groundwater velocities under natural gradient conditions are generally between

359 1 and 1000 m·year⁻¹ (0.002-2 cm·min⁻¹) (Mackay et al., 1985), far lower than the flow rates
 360 adopted in this study. According to Equation (12) and (13), the BDST model can be
 361 employed to predict the adsorption efficiency and column performance under other
 362 operational conditions without further experimental runs (Han et al., 2009a; Vijayaraghavan
 363 and Prabu, 2006). Table 5 lists the predicted breakthrough time for a new flow rate (0.01
 364 mL·min⁻¹ or 0.003 cm·min⁻¹). Where t_c is the predicted time and t_e is the observed time in the
 365 experiments.

366

367 Table 5 Breakthrough time prediction using BDST model at a new flow rate (ZSM-5
 368 percentage=5%)

Operational conditions	C/C ₀	New equations	t _c (min)	t _e (min)	RE ^a
Q'=0.5 mL·min ⁻¹	0.05	t'=37.48Z-60.78	164.1	167.87	2.25%
Z=6 cm	0.2	t'=59.36Z-82.44	273.72	274.83	0.40%
C ₀ '=300 mg·L ⁻¹	0.5	t'=83.70Z-78.19	424.01	427.09	0.72%
	0.85	t'=97.42Z+105.44	689.96	740.18	6.79%
Q'=0.01 mL·min ⁻¹	0.05	t'=1874Z-60.78	11183.22		
Z=6 cm	0.2	t'=2968Z-82.44	17725.56		
C ₀ =300 mg·L ⁻¹	0.5	t'=4185Z-78.19	25031.81		
	0.85	t'=4871Z+105.44	29331.44		

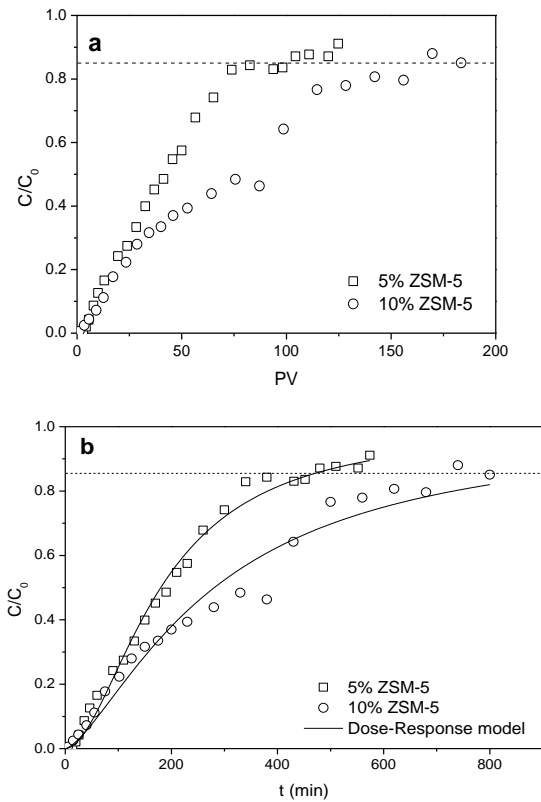
369 ^aRelative error

370

371 It was shown that the values of predicted time at a new flow rate were satisfactory with low
 372 relative errors. This indicates that the BDST model parameters in Table 4 can be employed to
 373 predict the column performance for the MTBE adsorption of ZSM-5 at different flow rates.

374 3.3.3 Effect of ZSM-5 dosage

375 The plots of effluent MTBE concentration versus PV and t at different ZSM-5 dosages are
376 shown in Figure 5a and 5b, respectively.



377

378

379 Figure 5 Breakthrough curves in fixed-bed columns with different ZSM-5 percentages as a
380 function of (a) pore volume (PV) and (b) time (t). ($C_0=300 \text{ mg}\cdot\text{L}^{-1}$, bed length=6 cm, flow
381 rate= $1 \text{ mL}\cdot\text{min}^{-1}$)

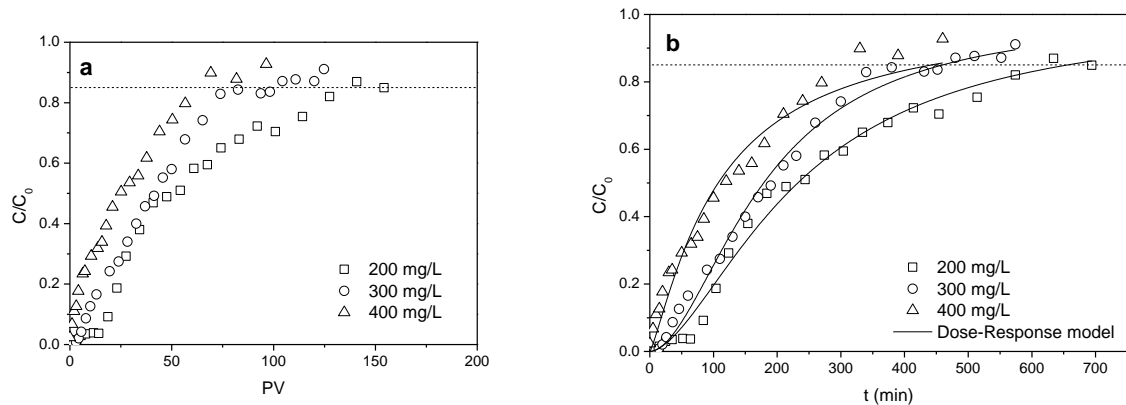
382

383 The saturation time of the column with a higher ZSM-5 percentage (10%) was significantly
384 longer and its breakthrough curve had a smaller gradient due to more available adsorption
385 sites for MTBE removal in the column. However, the breakthrough time was almost
386 unchanged with different ZSM-5 percentages.

387

388 3.3.4 Effect of inlet MTBE concentration

389 The effect of the influent MTBE concentration at 200, 300 and 400 $\text{mg}\cdot\text{L}^{-1}$ on the
 390 breakthrough profiles was analysed (Figure 6). It was observed that both the breakthrough
 391 time and saturation time decreased, and the slope of breakthrough curves between the
 392 breakthrough and saturate points, i.e. mass transfer zone (García-Mateos et al., 2015), became
 393 slightly steeper with the increase in the influent MTBE concentration. The steeper curve at
 394 higher inlet concentrations was an indicator of a smaller effluent volume whereas the
 395 extended breakthrough curve at lower inlet MTBE concentrations indicated that more
 396 solution was treated (Salman et al., 2011). This is because the higher concentration gradient
 397 at higher inlet MTBE concentrations caused a stronger mass transfer driving force (Goel et al.,
 398 2005) and faster solute transport in the column, leading to the quicker saturation of the
 399 adsorption sites on the ZSM-5 surface. The results in Table 3 show that the highest column
 400 adsorption capacity of $27.33 \text{ mg}\cdot\text{g}^{-1}$ was obtained at the inlet MTBE concentration of 400
 401 $\text{mg}\cdot\text{L}^{-1}$. Column tests at a low MTBE level ($\mu\text{g}\cdot\text{L}^{-1}$ level) will be explored in future.



402
 403 Figure 6 Breakthrough curves at inlet MTBE concentrations of 200, 300 and 400 $\text{mg}\cdot\text{L}^{-1}$ as
 404 function of (a) pore volume (PV) and (b) time (t) (bed length=6 cm, flow rate= $1 \text{ mL}\cdot\text{min}^{-1}$,
 405 ZSM-5=5%)

406

407 **3.4 Predicted PRB thickness**

408 The flow through thickness of a PRB is a main factor for the PRB design and can be
409 calculated by Equation (13).

$$410 \quad b = v \times t_w$$

411 (13)

412 where v is the velocity in the flow direction and t_w is the residence time. The residence time
413 (half-life, t_w) was determined at 99.9% of the respective equilibrium MTBE removal using
414 the best-fitting pseudo-second-order model in our previous study (Zhang et al., 2018)
415 combined with the Solver function in MS Excel (Cai et al., 2018; Gavaskar et al., 2000). The
416 residence time at different initial MTBE concentrations and predicted PRB thicknesses at a
417 nominal groundwater velocity of $0.18 \text{ cm}\cdot\text{h}^{-1}$ (equivalent to $0.01 \text{ mL}\cdot\text{min}^{-1}$ pump rate in this
418 study) are listed in Table 6. For example, the predicted PRB flow through thickness was
419 found to be 114.85 cm for 99.9% MTBE removal at an inlet MTBE concentration of 300
420 $\text{mg}\cdot\text{L}^{-1}$.

421

422 Table 6 Predicted residence time (h) and PRB thickness (cm) ($v=0.18 \text{ cm}\cdot\text{h}^{-1}$)

Initial MTBE concentration ($\text{mg}\cdot\text{L}^{-1}$)	100	150	300	600
Residence time (h)	122.62	456.26	638.06	683.11
Thickness (cm)	22.07	82.13	114.85	122.96

423

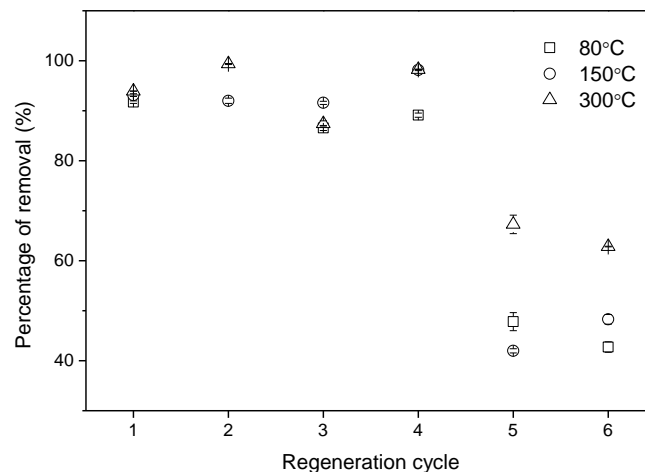
424 There are some limitations of this study, such as (i) the inaccuracy of using batch tests for
425 calculating residence time instead of calculating half-life of MTBE in the column tests; (ii)
426 the use of deionised water without considering the NOM (nature organic matter) and other
427 contaminants in the natural underground water, and (iii) the relatively high flow rate used.
428 More advance column design and selection of a wider range of flow rates will be conducted
429 in future studies to enable more accurate calculations.

430

431 Various remediation techniques have been applied to treat MTBE contaminated groundwater,
432 such as classical (Xu et al., 2004) and electrochemical Fenton treatment (Hong et al., 2007),
433 biodegradation by microorganism (Bradley et al., 1999), pump-and-treat, phytoremediation
434 (Hong et al., 2001), PRBs (Obiri-Nyarko et al., 2014), in-situ chemical oxidation (Krembs et
435 al., 2010), etc. The choice of remediation techniques depends on many factors, such as the
436 physiochemical properties of treating agents, site characterization, concentrations of MTBE
437 and other contaminants, and PRB is a promising in-situ groundwater remediation technique
438 due to its low-cost. The PRB treatment of MTBE contaminated groundwater with ZSM-5 as
439 the reactive medium is sustainable due to the adsorption of MTBE onto ZSM-5 without
440 precipitation which may cause clogging and reduce the permeability and removal efficiency
441 of PRBs (Zhou et al., 2014).

442

443 3.5 Regeneration study



444

445 Figure 7 The MTBE removal percentage by ZSM-5 after 6 regeneration cycles

446 In order to estimate the reusability of ZSM-5, the effect of repeated heat treatment at different
447 temperatures on the MTBE adsorption onto regenerated ZSM-5 was investigated and shown
448 in Figure 7. It was observed that there were no apparent changes in adsorption effects up to

449 four regeneration cycles at all temperatures and the regeneration at higher temperature
450 slightly increased the removal percentage. The abnormal value at the second cycle at 80 °C
451 was not included due to operating error. However, after 6 regeneration cycles, the removal
452 percentage decreased to ~67% at 300 °C compared with ~47% and ~52% for 80 °C and
453 150 °C, respectively. Therefore, ZSM-5 displays good regeneration potential compared with
454 modified activated carbon (~18% after 6 cycles) and iron oxide coated zeolites (<6% after 3
455 cycles) (Ania et al., 2004; Han et al., 2009b). It should be noted that sand or other medium in
456 PRBs should be heated with ZSM-5 in the practical application, and the vaporized MTBE
457 could be collected and treated to avoid secondary pollution.

458

459 **4. Conclusions**

460 Fixed-bed column tests were combined with breakthrough curve modelling to describe the
461 breakthrough curves and evaluate the adsorption performance under different operational
462 conditions. The regeneration characteristics of ZSM-5 were also discussed. The conclusions
463 are as follows:

464

- 465 (1) The results of both the regeneration tests and fixed-bed column tests show that ZSM-5 is
466 an effective reactive medium in PRBs for MTBE contaminated groundwater remediation.
- 467 (2) The Dose-Response model was found to best describe the breakthrough curves compared
468 with the Logit method, Adams-Bohart model, Thomas model and Yoon-Nelson model.
- 469 (3) The column adsorption capacity is ~31.85 mg·g⁻¹ at a 6 cm bed length, 1 mL·min⁻¹ flow
470 rate, 300 mg·L⁻¹ initial MTBE concentration and 5% ZSM-5 percentage.
- 471 (4) The maximum adsorption capacity increased with the increase of bed length and the
472 decrease of flow rate and MTBE concentration from the Dose-Response model, while the
473 adsorption capacity decreased with a higher ZSM-5 dosage due to the underestimate of

474 adsorption capacity caused by the fact that the ZSM-5 powder in the column may be more
475 likely to run away with the MTBE flow with a higher ZSM-5 dosage.

476 (5) The kinetic parameters obtained from the BDST model can be employed to predict the
477 dynamic behaviour of columns at new flow rates.

478 (6) The adsorption capacity of regenerated ZSM-5 remains satisfactory (>85%) after up to
479 four regeneration cycles at 80, 150 and 300 °C and regeneration at higher temperatures
480 performed slightly better.

481

482 **Acknowledgements**

483 The first author would like to thank China Scholarship Council (CSC) for providing the PhD
484 studentship, and the third author is grateful to the Killam Trusts of Canada for the Izaak
485 Walton Killam Memorial Postdoctoral Fellowship.

486 **References**

- 487 Abu-Lail, L., Bergendahl, J.A., Thompson, R.W., 2010. Adsorption of methyl tertiary butyl
488 ether on granular zeolites: Batch and column studies. *J. Hazard. Mater.* 178, 363–369.
489 <https://doi.org/10.1016/j.jhazmat.2010.01.088>
- 490 Adak, A., Pal, A., 2006. Removal of phenol from aquatic environment by SDS-modified
491 alumina: Batch and fixed bed studies. *Sep. Purif. Technol.* 50, 256–262.
492 <https://doi.org/10.1016/J.SEPPUR.2005.11.033>
- 493 Al-Tabbaa, A., Liska, M., 2012. Soil mix remediation technology (SMiRT) final technical
494 summary report (confidential).
- 495 Anderson, M.A., 2000. Removal of MTBE and other organic contaminants from water by
496 sorption to high silica zeolites. *Environ. Sci. Technol.* 34, 725–727.
497 <https://doi.org/10.1021/es990390t>
- 498 Ania, C.O., Menéndez, J.A., Parra, J.B., Pis, J.J., 2004. Microwave-induced regeneration of
499 activated carbons polluted with phenol. A comparison with conventional thermal
500 regeneration, in: *Carbon*. <https://doi.org/10.1016/j.carbon.2004.01.010>
- 501 Ayooob, S., Gupta, A.K., 2007. Sorptive response profile of an adsorbent in the defluoridation
502 of drinking water. *Chem. Eng. J.* <https://doi.org/10.1016/j.cej.2007.02.013>
- 503 Bohart, G.S., Adams, E.Q., 1920. Some aspects of the behavior of charcoal with respect to
504 chlorine. *J. Am. Chem. Soc.* 42, 523–544. <https://doi.org/10.1021/ja01448a018>
- 505 Bradley, P.M., Landmeyer, J.E., Chapelle, F.H., 1999. Aerobic mineralization of MTBE and
506 tert-butyl alcohol by stream-bed sediment microorganisms. *Environ. Sci. Technol.*
507 <https://doi.org/10.1021/es990062t>
- 508 Cai, Q., Turner, B.D., Sheng, D., Sloan, S., 2018. Application of kinetic models to the design
509 of a calcite permeable reactive barrier (PRB) for fluoride remediation. *Water Res.*
510 <https://doi.org/10.1016/j.watres.2017.11.046>

511 Calero, M., Hernáinz, F., Blázquez, G., Tenorio, G., Martín-Lara, M.A., 2009. Study of Cr
512 (III) biosorption in a fixed-bed column. *J. Hazard. Mater.* 171, 886–893.
513 <https://doi.org/10.1016/j.jhazmat.2009.06.082>

514 Cappai, G., De Gioannis, G., Muntoni, A., Spiga, D., Zijlstra, J.J.P., 2012. Combined use of a
515 transformed red mud reactive barrier and electrokinetics for remediation of Cr/As
516 contaminated soil. *Chemosphere* 86, 400–408.
517 <https://doi.org/10.1016/J.CHEMOSPHERE.2011.10.053>

518 Centi, G., Grande, A., Perathoner, S., 2002. Catalytic conversion of MTBE to biodegradable
519 chemicals in contaminated water, in: *Catalysis Today*. <https://doi.org/10.1016/S0920->
520 [5861\(02\)00046-9](https://doi.org/10.1016/S0920-5861(02)00046-9)

521 Centi, G., Perathoner, S., 2003. Remediation of water contamination using catalytic
522 technologies. *Appl. Catal. B Environ.* [https://doi.org/10.1016/S0926-3373\(02\)00198-4](https://doi.org/10.1016/S0926-3373(02)00198-4)

523 Chan, M.S.M., Lynch, R.J., 2003. Photocatalytic degradation of aqueous methyl-tert-butyl-
524 ether (MTBE) in a supported-catalyst reactor. *Environ. Chem. Lett.* 1, 157–160.
525 <https://doi.org/10.1007/s10311-003-0037-4>

526 Cruz Viggì, C., Pagnanelli, F., Cibati, A., Uccelletti, D., Palleschi, C., Toro, L., 2010.
527 Biotreatment and bioassessment of heavy metal removal by sulphate reducing bacteria
528 in fixed bed reactors. *Water Res.* 44, 151–158.
529 <https://doi.org/10.1016/j.watres.2009.09.013>

530 Dorado, A.D., Gamisans, X., Valderrama, C., Solé, M., Lao, C., 2014. Cr(III) removal from
531 aqueous solutions: A straightforward model approaching of the adsorption in a fixed-bed
532 column. *J. Environ. Sci. Heal. - Part A Toxic/Hazardous Subst. Environ. Eng.* 49, 179–
533 186. <https://doi.org/10.1080/10934529.2013.838855>

534 García-Mateos, F.J., Ruiz-Rosas, R., Marqués, M.D., Cotoruelo, L.M., Rodríguez-Mirasol, J.,
535 Cordero, T., 2015. Removal of paracetamol on biomass-derived activated carbon:

536 Modeling the fixed bed breakthrough curves using batch adsorption experiments. Chem.
537 Eng. J. 279, 18–30. <https://doi.org/10.1016/j.cej.2015.04.144>

538 Gavaskar, A., Gupta, N., Sass, B., Janosy, R., Hicks, J., 2000. Design guidance for
539 application of Permeable Reactive Barriers for groundwater remediation.

540 Gavaskar, A.R., 1999. Design and construction techniques for permeable reactive barriers. J.
541 Hazard. Mater. 68, 41–71. [https://doi.org/10.1016/S0304-3894\(99\)00031-X](https://doi.org/10.1016/S0304-3894(99)00031-X)

542 Goel, J., Kadirvelu, K., Rajagopal, C., Kumar Garg, V., 2005. Removal of lead(II) by
543 adsorption using treated granular activated carbon: Batch and column studies. J. Hazard.
544 Mater. 125, 211–220. <https://doi.org/10.1016/j.jhazmat.2005.05.032>

545 Gouran-Orimi, R., Mirzayi, B., Nematollahzadeh, A., Tardast, A., 2018. Competitive
546 adsorption of nitrate in fixed-bed column packed with bio-inspired polydopamine coated
547 zeolite. J. Environ. Chem. Eng. 6, 2232–2240. <https://doi.org/10.1016/j.jece.2018.01.049>

548 Han, R., Ding, D., Xu, Y., Zou, W., Wang, Y., Li, Y., Zou, L., 2008. Use of rice husk for the
549 adsorption of congo red from aqueous solution in column mode. Bioresour. Technol. 99,
550 2938–2946. <https://doi.org/10.1016/j.biortech.2007.06.027>

551 Han, R., Wang, Y., Zhao, X., Wang, Y., Xie, F., Cheng, J., Tang, M., 2009a. Adsorption of
552 methylene blue by phoenix tree leaf powder in a fixed-bed column: Experiments and
553 prediction of breakthrough curves. Desalination 245, 284–297.
554 <https://doi.org/10.1016/j.desal.2008.07.013>

555 Han, R., Zou, L., Zhao, X., Xu, Y., Xu, F., Li, Y., Wang, Y., 2009b. Characterization and
556 properties of iron oxide-coated zeolite as adsorbent for removal of copper(II) from
557 solution in fixed bed column. Chem. Eng. J. 149, 123–131.
558 <https://doi.org/10.1016/j.cej.2008.10.015>

559 Hong, M.S., Farmayan, W.F., Dortch, I.J., Chiang, C.Y., McMillan, S.K., Schnoor, J.L., 2001.
560 Phytoremediation of MTBE from a groundwater plume. Environ. Sci. Technol.

561 <https://doi.org/10.1021/es001911b>

562 Hong, S., Zhang, H., Duttweiler, C.M., Lemley, A.T., 2007. Degradation of methyl tertiary-
563 butyl ether (MTBE) by anodic Fenton treatment. *J. Hazard. Mater.*
564 <https://doi.org/10.1016/j.jhazmat.2006.12.030>

565 Hou, D., Al-Tabbaa, A., Luo, J., 2014. Assessing effects of site characteristics on remediation
566 secondary life cycle impact with a generalised framework. *J. Environ. Plan. Manag.* 57,
567 1083–1100. <https://doi.org/10.1080/09640568.2013.863754>

568 Hutchins, R., 1973. New method simplifies design of activated-carbon systems. *Chem. Eng.*
569 80, 133–138.

570 Jha, B., Singh, D.N., 2016. Fly ash zeolites, *Advanced Structured Materials*. Springer
571 Singapore, Singapore. <https://doi.org/10.1007/978-981-10-1404-8>

572 Katsou, E., Malamis, S., Tzanoudaki, M., Haralambous, K.J., Loizidou, M., 2011.
573 Regeneration of natural zeolite polluted by lead and zinc in wastewater treatment
574 systems. *J. Hazard. Mater.* 189, 773–786. <https://doi.org/10.1016/j.jhazmat.2010.12.061>

575 Krembs, F.J., Siegrist, R.L., Crimi, M.L., Furrer, R.F., Petri, B.G., 2010. ISCO for
576 groundwater remediation: Analysis of field applications and performance. *Gr. Water*
577 *Monit. Remediat.* <https://doi.org/10.1111/j.1745-6592.2010.01312.x>

578 Levchuk, I., Bhatnagar, A., Sillanpää, M., 2014. Overview of technologies for removal of
579 methyl tert-butyl ether (MTBE) from water. *Sci. Total Environ.*
580 <https://doi.org/10.1016/j.scitotenv.2014.01.037>

581 Lindsey, B.D., Ayotte, J.D., Jurgens, B.C., Desimone, L.A., 2017. Using groundwater age
582 distributions to understand changes in methyl tert-butyl ether (MtBE) concentrations in
583 ambient groundwater, northeastern United States. *Sci. Total Environ.* 579, 579–587.
584 <https://doi.org/10.1016/j.scitotenv.2016.11.058>

585 Mackay, D.M., Roberts, P. V., Cherry, J.A., 1985. Transport of organic contaminants in

586 groundwater. *Environ. Sci. Technol.* 19, 384–392. <https://doi.org/10.1021/es00135a001>

587 Mancini, E.R., Steen, A., Rausina, G.A., Wong, D.C.L., Arnold, W.R., Gostomski, F.E.,
588 Davies, T., Hockett, J.R., Stubblefield, W.A., Drottar, K.R., Springer, T.A., Errico, P.,
589 2002. MTBE ambient water quality criteria development: A public/private partnership.
590 *Environ. Sci. Technol.* 36, 125–129. <https://doi.org/10.1021/es002059b>

591 Martucci, A., Braschi, I., Bisio, C., Sarti, E., Rodeghero, E., Bagatin, R., Pasti, L., 2015.
592 Influence of water on the retention of methyl tertiary-butyl ether by high silica ZSM-5
593 and Y zeolites: a multidisciplinary study on the adsorption from liquid and gas phase.
594 *RSC Adv.* 5, 86997–87006. <https://doi.org/10.1039/C5RA15201A>

595 Masad, E., Taha, R., Ho, C., Papagiannakis, T., 1996. Engineering properties of tire/soil
596 mixtures as a lightweight fill material. *Geotech. Test. Journal*, 19, 297–304.
597 <https://doi.org/10.1520/GTJ10355J>

598 Mohebbali, S., 2013. Degradation of methyl t-butyl ether (MTBE) by photochemical process
599 in nanocrystalline TiO₂ slurry: Mechanism, by-products and carbonate ion effect. *J.*
600 *Environ. Chem. Eng.* 1, 1070–1078. <https://doi.org/10.1016/j.jece.2013.08.022>

601 Obiri-Nyarko, F., Grajales-Mesa, S.J., Malina, G., 2014. An overview of permeable reactive
602 barriers for in situ sustainable groundwater remediation. *Chemosphere.*
603 <https://doi.org/10.1016/j.chemosphere.2014.03.112>

604 Oulman, C., 1980. The logistic curve as a model for carbon bed design. *J. AWWA* 72, 50–53.

605 Ozdemir, O., Turan, M., Turan, A.Z., Faki, A., Engin, A.B., 2009. Feasibility analysis of
606 color removal from textile dyeing wastewater in a fixed-bed column system by
607 surfactant-modified zeolite (SMZ). *J. Hazard. Mater.* 166, 647–654.
608 <https://doi.org/10.1016/j.jhazmat.2008.11.123>

609 Pascoe, W.E., 1992. *Catalysis of organic reactions*. Marcel Dekker, New York.

610 Reuter, J.E., Allen, B.C., Richards, R.C., Pankow, J.F., Goldman, C.R., Roger L. Schol, A.,

611 Seyfried, J.S., 1998. Concentrations, sources, and fate of the gasoline oxygenate Methyl
612 tert-Butyl Ether (MTBE) in a multiple-use lake. *Environ. Sci. Technol.* 32, 3666–3672.
613 <https://doi.org/10.1021/ES9805223>

614 Sağ, Y., Aktay, Y., 2001. Application of equilibrium and mass transfer models to dynamic
615 removal of Cr(VI) ions by Chitin in packed column reactor. *Process Biochem.* 36, 1187–
616 1197. [https://doi.org/10.1016/S0032-9592\(01\)00150-9](https://doi.org/10.1016/S0032-9592(01)00150-9)

617 Salman, J., Njoku, V., Hameed, B., 2011. Batch and fixed-bed adsorption of 2, 4-
618 dichlorophenoxyacetic acid onto oil palm frond activated carbon. *Chem. Eng. J.* 174,
619 33–40.

620 Shah, I.K., Pre, P., Alappat, B.J., 2014. Effect of thermal regeneration of spent activated
621 carbon on volatile organic compound adsorption performances. *J. Taiwan Inst. Chem.*
622 *Eng.* 45, 1733–1738. <https://doi.org/10.1016/J.JTICE.2014.01.006>

623 Thomas, H., 1944. Heterogeneous ion exchange in a flowing system. *J. Am. Chem. Soc.* 66,
624 1664–1666.

625 Thomas, H.C., 1948. Chromatography: A problem in kinetics. *Ann. N. Y. Acad. Sci.* 49,
626 161–182. <https://doi.org/10.1111/j.1749-6632.1948.tb35248.x>

627 Vijayaraghavan, K., Prabu, D., 2006. Potential of *Sargassum wightii* biomass for copper(II)
628 removal from aqueous solutions: Application of different mathematical models to batch
629 and continuous biosorption data. *J. Hazard. Mater.* 137, 558–564.
630 <https://doi.org/10.1016/j.jhazmat.2006.02.030>

631 Wang, S., Zhu, Z.H., 2006. Characterisation and environmental application of an Australian
632 natural zeolite for basic dye removal from aqueous solution. *J. Hazard. Mater.* 136, 946–
633 952. <https://doi.org/10.1016/j.jhazmat.2006.01.038>

634 Wei, Y.X., Ye, Z.F., Wang, Y.L., Ma, M.G., Li, Y.F., 2011. Enhanced ammonia nitrogen
635 removal using consistent ammonium exchange of modified zeolite and biological

636 regeneration in a sequencing batch reactor process. *Environ. Technol.* 32, 1337–1343.
637 <https://doi.org/10.1080/09593330.2010.536784>

638 Xin, S., Zeng, Z., Zhou, X., Luo, W., Shi, X., Wang, Q., Deng, H., Du, Y., 2017. Recyclable
639 *Saccharomyces cerevisiae* loaded nanofibrous mats with sandwich structure constructing
640 via bio-electrospraying for heavy metal removal. *J. Hazard. Mater.* 324, 365–372.
641 <https://doi.org/10.1016/j.jhazmat.2016.10.070>

642 Xu, X.R., Zhao, Z.Y., Li, X.Y., Gu, J.D., 2004. Chemical oxidative degradation of methyl
643 tert-butyl ether in aqueous solution by Fenton's reagent. *Chemosphere.*
644 <https://doi.org/10.1016/j.chemosphere.2003.11.017>

645 Yan, G., Viraraghavan, T., Chen, M., 2001. A new model for heavy metal removal in a
646 biosorption column. *Adsorpt. Sci. Technol.* 19, 25–43.
647 <https://doi.org/10.1260/0263617011493953>

648 Yoon, Y.H., 1984. Application of gas adsorption kinetics. *A Theor. Model Respir. Cart. Serv.*
649 *life* 45, 509–516.

650 Zhang, Y., Jin, F., Lynch, R., Al-Tabbaa, A., 2018a. Breakthrough curve modelling of ZSM-
651 5 zeolite packed fixed-bed columns for the removal of MTBE, in: *The 8th International*
652 *Congress on Environmental Geotechnics (Forthcoming).*

653 Zhang, Y., Jin, F., Shen, Z., Lynch, R., Al-Tabbaa, A., 2018b. Kinetic and equilibrium
654 modelling of MTBE (methyl tert-butyl ether) adsorption on ZSM-5 zeolite: Batch and
655 column studies. *J. Hazard. Mater.* 347, 461–469.
656 <https://doi.org/10.1016/J.JHAZMAT.2018.01.007>

657 Zhou, D., Li, Y., Zhang, Y., Zhang, C., Li, X., Chen, Z., Huang, J., Li, X., Flores, G., Kamon,
658 M., 2014. Column test-based optimization of the permeable reactive barrier (PRB)
659 technique for remediating groundwater contaminated by landfill leachates. *J. Contam.*
660 *Hydrol.* <https://doi.org/10.1016/j.jconhyd.2014.09.003>

Supplementary Material

[Click here to download Supplementary Material: Supplementary Material.docx](#)

Highlights

- ZSM-5 is evaluated on its removal for MTBE in fixed-bed column tests.
- Dose-Response model was found to best describe the breakthrough curves.
- The removal capacity is $\sim 31.85 \text{ mg}\cdot\text{g}^{-1}$ in fixed-bed column tests.
- Parameters from BDST model can predict breakthrough curves at a new flow rate.
- ZSM-5 is effective and recyclable for MTBE contaminated groundwater remediation.