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Influence of glaciation on mechanisms of mineral weathering in two high Arctic catchments

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Abstract

In order to investigate the effect of glaciation on mineral weathering, the stream water chemistry and the bacterial community composition were analysed in two catchments containing nominally identical sedimentary formations but which differed in the extent of glaciation. The stream waters were analysed for major ions, δ^{34} S, $\delta^{18}O_{SO4}$ and $\delta^{18}O_{H2O}$ and associated stream sediments were analysed by 16S rRNA gene tagged sequencing.

Sulfate comprised 72-86% and 35-45% of the summer anion budget (in meq) in the unglaciated and glaciated catchments respectively. This indicates that sulfuric acid generated from pyrite weathering is a significant weathering agent in both catchments. Based on the relative proportions of cations, sulfate and bicarbonate, the stream water chemistry of the unglaciated catchment was found to be consistent with a sulfide oxidation coupled to silicate dissolution weathering process whereas in the glaciated catchment both carbonates and silicates weathered via both sulfuric and carbonic acids.

Stable isotope measurements of sulfate, together with inferences of metabolic processes catalysed by resident microbial communities, revealed that the pyrite oxidation reaction differed between the two catchments. No δ^{34} S fractionation relative to pyrite was observed in the unglaciated catchment and this was interpreted to reflect pyrite oxidation under oxic conditions. In contrast, δ^{34} S and $\delta^{18}O_{SO4}$ values were positively correlated in the glaciated catchment and were positively offset from pyrite. This was interpreted to reflect pyrite oxidation under anoxic conditions with loss of S intermediates.

This study suggests that glaciation may alter stream water chemistry and the mechanism of pyrite oxidation through an interplay of biological, physical and chemical factors.

Keywords:

chemical weathering, pyrite, sulfur isotopes, bacteria, biogeochemical cycles

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1 1. Introduction

The Arctic is currently experiencing a period of warming, resulting in the retreat of glaciers and an increase in the 2 active layer depth of permafrost (Vaughan et al., 2013). Major emphasis has been placed on quantifying the effects of 3 warming on nutrient fluxes, especially carbon (e.g. Schuur et al., 2009; Elberling et al., 2013), but major ion fluxes from both permafrost and glaciated areas (MacLean et al., 1999; Frey and McClelland, 2009; Pokrovsky et al., 2012; Nowak and Hodson, 2015) are also predicted to change with continued warming. Decreased permafrost cover is 6 expected to increase overall fluxes of solutes as the active layer deepens (MacLean et al., 1999; Frey and McClelland, 2009). Likewise, solute fluxes from glaciers are predicted to increase by virtue of the increased discharge as a result of 8 a longer melt season (Lafrenière and Sharp, 2005). However, it is unclear if, in addition to an increase in solute flux, q the composition of this flux will change and what affect these changes will have on the carbon cycle and the microbial 10 community which mediate many of the chemical reactions occurring in these environments (Skidmore et al., 2005; 11 Boyd et al., 2011, 2014). 12 Chemical weathering, which is a key part of the biogeochemical cycles of many elements, is assumed to mainly 13

occur by reaction with carbonic acid, formed by the dissolution of carbon dioxide (CO₂) in water. However, where 14 sulfide minerals e.g. pyrite (FeS_2) exist, sulfuric acid may form through the oxidation of sulfide, which can also 15 result in mineral dissolution (Holland, 1978). Understanding which agent is responsible for mineral dissolution is 16 important for understanding inputs to the sulfur biogeochemical cycle and for quantifying the contribution of sulfuric 17 acid weathering to global chemical weathering fluxes (Berner and Berner, 1996). Chemical weathering by sufuric 18 acid does not involve drawdown of atmospheric CO2 and can even be a net source of CO2 if carbonates are weathered. 19 Therefore, if this reaction were significant on a global scale then the weathering-climate negative feedback would be 20 weakened (Calmels et al., 2007; Li et al., 2008; Torres et al., 2014). 21

Sulfide oxidation coupled to carbonate dissolution (SOCD) is a key process determining stream water chemistry 22 in the high physical erosion environment of glaciated catchments (Fairchild et al., 1999; Tranter et al., 2002; Sharp 23 et al., 2002; Skidmore et al., 2005; Robinson et al., 2009; Wadham et al., 2010; Boyd et al., 2014). The high rates 24 of physical erosion expose both the carbonate and pyrite grains allowing them to weather rapidly. In non-glaciated 25 Arctic catchments, weathering is also strongly influenced by physical erosion processes such as frost shattering (Huh 26 and Edmond, 1999; Hall et al., 2002) and where sulfide is present, sulfide oxidation is a key chemical weathering 27 reaction (Elberling and Langdahl, 1998; Thorn et al., 2001; Calmels et al., 2007). Significant weathering by sulfuric 28 acid is also observed in other high erosion settings such as Taiwan and the Himalayas (Wolff-Boenisch et al., 2009; 29 Turchyn et al., 2013; Torres et al., 2014). 30

The stable isotopes of sulfate (δ^{34} S and $\delta^{18}O_{SO4}$) provide a useful tool with which to investigate the sources of sul-

se fate and the reactions leading to sulfate formation (e.g. Mandernack et al., 2003; Calmels et al., 2007; Wadham et al.,

33 2007; Turchyn et al., 2013). Based on stable isotope evidence, sulfate reduction was identified at Borup Glacier in

Ellesmere Island, Canada due to supra-glacial sulfur springs (Grasby et al., 2003) and in a sub-glacial upwelling from

³⁵ Finsterwalderbreen, Svalbard (Wadham et al., 2004) but not in the pro-glacial area of the same glacier (Wadham et al.,

³⁶ 2007). However, bacterial sequencing data from other catchments has failed to detect sulfate reducing bacteria where

water chemistry evidence suggested sulfate reduction was occurring (Skidmore et al., 2000, 2005). It is therefore not

clear how widespread this reaction is in glaciated environments.

The chemical reactions involving sulfide oxidation and sulfate reduction are microbially mediated and it is only in 39 the last 15 years that the diversity and functional importance of microorganisms in glacial landscapes has been recog-40 nised (e.g. Skidmore et al., 2000; Hodson et al., 2008). More recently, it has been shown that the microbial community 41 composition is strongly influenced by bedrock composition and that the microbial community strongly influences so-42 lute chemistry (Larouche et al., 2012; Montross et al., 2013; Mitchell et al., 2013). In particular, the presence of FeS₂ 43 was shown to be a dominant control on the composition of communities inhabiting the subglacial environment of 44 Robertson Glacier, Canada (Mitchell et al., 2013), which likely reflects the utilisation of energy derived from mineral 45 redox reactions to support primary productivity (Boyd et al., 2014). It is therefore probable that minerals which can 46 serve as electron donors and acceptors play a key role in determining the composition of microbial communities, and 47 by extension the chemical composition of solute fluxes, in other oligotrophic and obligately chemotrophic subglacial 48 environments. 49

This study focusses on two adjacent catchments with nominally identical lithology: sedimentary rocks known to contain pyrite. One catchment was glaciated and the other was unglaciated and underlain by permafrost. We assume that the extent of glacial cover is the primary cause of differences in hydrology, biology and chemistry between the two catchments. The paired catchment approach provides clues to long-term changes in weathering processes induced by deglaciation. The aim of the study was to utilise the combination of stream water chemistry, S and O isotopes of sulfate, and molecular analyses of microbial community composition to investigate the formation of and the role of sulfuric acid weathering in the two catchments.

57 2. Description of field area

Svalbard is located in the Arctic Ocean. The archipelago has an arctic climate with a mean annual air temperature of -5 °C and mean annual precipitation of 180 mm (measured at Longyearbyen airport, Humlum et al., 2003). Permafrost is continuous throughout the islands and can be up to 500 m thick (Humlum et al., 2003). The two studied catchments are situated next to each other (Fig. 1) in the Paleogene sedimentary Central Basin of Svalbard. The sedimentary formations exposed in the catchments are from the Van Mijenfjorden group which is Paleocene to Eocene in age (66 - 33.9 Ma) and contain sandstones, siltstones and shale (Fig. 2, Major et al., 2000).

⁶⁴ Dryadbreen has been retreating since the end of the Little Ice Age (~1890, Ziaja, 2001). The thermal regime ⁶⁵ of the glacier is expected to be cold-based with temperate patches, based on similar sized glaciers in the same area ⁶⁶ (Etzelmüller et al., 2000; Etzelmüller and Hagen, 2005). Between 1936 and 2006 the area of the glacier decreased ⁶⁷ from 2.59 to 0.91 km² leaving large terminal and lateral ice-cored moraines and a sandur in front of the glacier (Ziaja



Figure 1: Panoramic view of the two study catchments. Dryadbreen is on the left and Fardalen on the right. The red dashed line demarcates the catchment boundaries and the yellow crosses indicate the water sampling locations for each catchment. Note the large end moraine and sandur in the glaciated catchment. Photograph credit: Alix Guillot.

and Pipała, 2007). The sandur surface lowered 14 m between 2001 and 2006 due to the melting of dead ice (Ziaja and Pipała, 2007). The uppermost part of the catchment faces north-north-east and the valley then curves around such that at lower elevations (<500 m) the catchment faces south-east. The catchment area is 4.8 km² and ranges in elevation from 250 - 1031 m.a.s.l. The river in the sandur plain is braided, but the braids merge such that one stream drains the end moraine. This stream was sampled just before the confluence with the river in the main valley. In this paper 'Dryadbreen' will be used to refer to the whole catchment and not just the glacier.

Fardalen is a non-glaciated catchment at the head of a valley of the same name. In contrast to Dryadbreen, the whole catchment has a south-easterly aspect which contributes to the absence of present-day glaciation. The valley is currently underlain by continuous permafrost and is likely to have been unglaciated for at least the last 10 kyr (Svendsen and Mangerud, 1997). The catchment area is 3.4 km² and ranges in elevation from 250 - 1025 m.a.s.l. A first-order stream drains the catchment and it was sampled just before the confluence with the river in the main valley.

79 **3. Methods**

80 3.1. Hydrology

Water stand and water temperature were recorded every 10 minutes by a CS450 Campbell Scientific pressure transducer connected to a Campbell CR200X data logger. In Dryadbreen conductivity was recorded every 10 minutes using a Ponsel CE4 meter. Water stand was converted to discharge using discharge measurements obtained by salt tracing, which were performed using a point addition of 1-3 kg salt. The resulting change in conductivity ~70 m downstream was monitored by a Hobo U24 conductivity logger recording every second. The calibration of the



Figure 2: Geological map of the study area. The red dot in the inset shows the location of the study area in relation to the rest of Svalbard. The two stars indicate the location of the two catchments: Dryadbreen (glaciated) and Fardalen (unglaciated). The red square marks the location of Nybyen (part of the town of Longyearbyen) and the red circle shows from where the photograph in Figure 1 was taken. The map and the legend are from Major et al. (2000).

conductivity meter and conversion to discharge was done following the procedure outlined in Hudson and Fraser (2005). The amount of snow in May 2012 prevented installation of the loggers that early in the season, therefore high resolution hydrological data is only available for the period 25th July to 3rd August 2012. In mid-May there was no surface water and no sub-surface water was found by digging. Three weeks later, the landscape was still dominated

⁹⁰ by snow but both streams were flowing.

91 3.2. Collection of water samples

The Dryadbreen and Fardalen streams were sampled twice a day from 14th to 18th June 2012 and from 25th July 92 to 3rd August 2012. The number of days sampled corresponds to approximately 20% of the melt-season (Yde and 93 Knudsen, 2004). The two rivers were sampled just before they joined the main valley river. For Dryadbreen this 94 was approximately 1 km from the front of the glacier. Temperature and pH were measured in situ (Hanna HI 98160 95 pH meter). Water samples were filtered on the day of collection through 0.2 μ m nylon filters using a polycarbonate 96 vacuum filtration unit connected to a hand pump. A filtered water sample was titrated with 3.3 mM HCl within an 97 hour of collection and alkalinity was calculated from the titration curve using the Gran method (Stumm and Morgan, 98 1996). Assuming that alkalinity \approx [HCO₃⁻]+[CO₃²⁻] and using K₁, K₂, K_H values for 4 °C, then HCO₃⁻ comprises 99 more than 99.9% of alkalinity for all samples and we therefore assume that the measured alkalinity is equivalent to 100 the bicarbonate concentration. Filtered samples were stored in pre-cleaned HDPE bottles and those intended for the 101 analysis of cations were acidified to pH 2 with single-distilled concentrated HNO3. Sulfate was pre-concentrated by 102 passing 3 L of filtered water through a column filled with 5 mL Dowex 1x8, 100-200 mesh chloride form resin. The 103 resin was pre-cleaned by rinsing with 60mL 3M HCl followed by 60 mL 18.2 MΩ water. Three snow samples were 104 taken by filling a bucket with snow and allowing it to melt at room temperature. Two rain samples were collected from

¹⁰⁶ an HDPE bottle with a funnel. A supra-glacial stream sample was also collected. All of these samples were filtered ¹⁰⁷ as described for river water samples.

108 3.3. Analysis of water samples

For all water samples, major cations and Si were measured by inductively-coupled plasma optical emission spectrometry (ICP-OES, Perkin Elmer Optima 5300 DV at the University of Edinburgh) and anions by ion chromatography 110 (IC, Dionex DX 500 at Durham University). Measured cation concentrations of the water standards SLRS-5 (National 111 Research Council Canada) and BATTLE-02 (Environment Canada) were within 8% of the certified values. Measured 112 anion concentrations were within 5% of the certified values for LETHBRIDG-03 and BATTLE-02 (both Environment 113 Canada). External reproducibility, as measured by the mean normalized difference of 9 pairs of replicate field samples, 114 was <3% for cations and <5% for anions. Calculated charge balance errors (CBE) were <3% for the vast majority of 115 samples, confirming the accuracy of the anion and cation measurements (Table 1). 116 The oxygen isotopic composition ($\delta^{18}O_{H2O}$) of water samples was measured on 750 μ L samples of water which 117 had been equilibrated with a mixture of 0.3% CO₂ and He in septum capped vials. The CO₂/He mixture was measured 118

using a Gas Bench II (Thermo Scientific) connected to an isotope ratio mass spectrometer (IRMS, Delta PLUS XP,

120 Thermo Scientific at the University of St. Andrews). Measurements were calibrated with the international standards

121 SLAP-2, GISP and VSMOW-2. The results are reported in the conventional delta notation with respect to VSMOW

and sample standard deviation was less than 0.15% (2SD).

Sulfate from the anion resins was eluted with 2 M KCl, and the eluent acidified to pH 3 with HCl, heated to 123 sub-boiling, and barium sulfate precipitated by addition of BaCl₂. The barium sulfate was then recovered by four 124 cycles of centrifugation, discard of the supernatant, and washing in de-ionised water before oven drying at 80 °C. The 125 sulfate blank for the process is less than 0.5 mg BaSO₄ (limit of determination), which is less than 1% of the smallest 126 sample. ³⁴S/³²S ratios were determined by combustion to SO₂ with V₂O₅ in an EA-1120 elemental analyser on-line to 127 an IRMS (Delta+XL, ThermoFinnigan at the NERC Isotope Geosciences Laboratory), with ³⁴S/³²S ratios calculated 128 as δ^{34} S values versus CDT by comparison with standards IAEA SO6 and NBS-127. Analytical precision of replicates 129 was typically $\leq 0.2\%$ (1 SD). ¹⁸O/¹⁶O ratios were determined by thermal conversion to CO in a TC/EA on-line to an 130 IRMS (Delta+XL, ThermoFinnigan at the NERC Isotope Geosciences Laboratory), with ¹⁸O/¹⁶O ratios calculated as 131 $\delta^{18}O_{SO4}$ values versus SMOW by comparison with standards IAEA SO5 and SO6. Analytical precision of replicates 132 was typically <0.5% (1 SD). 133

134 3.4. Analysis of solid samples

Three rock samples were collected: R01 was collected from the Frysjaodden Formation (Fig. 2) in Fardalen, R02 was collected from the sandur in Dryadbreen and R04 was collected from the surface of the glacier. The rock samples, in addition to a sediment sample taken at the water sampling location in each catchment (O - Dryadbreen and L - Fardalen) were crushed and ground to fine powders. The rock samples were analysed by X-ray fluorescence

spectrometry (XRF, PANalytical Axios at the Norwegian Geological Survey) and all samples were analysed by X-139 ray diffraction (XRD). S contents of the rock samples were analysed by high temperature combustion followed by 140 infrared detection (Leco SC-444 at the Norwegian Geological Survey). XRD analysis was performed on a PANalytical 141 PW1050 X-ray diffractometer with a Hiltonbrooks DG2 X-ray generator (Co-Ka radiation) at the University of St. 142 Andrews. Data were collected between 5 and 70° 2θ with a step size of 0.02° and a counting time of 3 s per step. Semi-143 quantitative mineralogical abundances were obtained using the Siroquant software. The typical error on abundances is 144 estimated to be 5-10%. Carbon isotope analysis of calcite from R02 was performed by reaction of whole rock powder 145 with anhydrous phosphoric acid overnight at 25 °C and subsequent analysis of the liberated CO₂ in a MAT 253 mass 146 spectrometer (ThermoFinnigan at the NERC Isotope Geosciences Laboratory). ${}^{13}C/{}^{12}C$ ratios were converted to $\delta^{13}C$ 147 values versus VPDB by comparison with a laboratory calcite standard calibrated against NBS-19 and NBS-18. 148 Pyrite was separated from two shale samples by using a solution of lithium heteropolytungstate (LST). The sep-149 arated fraction contained both pyrite and magnetite and the pyrite was separated from magnetite using a magnet and 150 hand-picking. Pyrite separates were ground to a powder in an agate mortar and ³⁴S/³²S ratios determined by com-151 bustion to SO2 with V2O5 in an EA-1120 elemental analyser on-line to an IRMS (Delta+XL, ThermoFinnigan at the 152 NERC Isotope Geosciences Laboratory), with ${}^{34}S/{}^{32}S$ ratios calculated as $\delta^{34}S$ values versus CDT by comparison 153 with standards IAEA S-1 and S-2. Analytical precision of replicates was $\leq 0.4\%$ (1 SD). 154 Suspended sediment samples (>0.2 μ m, collected on nylon filter papers) were washed off the filter papers using 155 deionized water and freeze-dried. Depending on the amount of sediment recovered, part or all of the sample was 156 then reacted overnight with 1.5 M HCl to remove carbonates, washed free of acid, dried and homogenised. ${}^{13}C/{}^{12}C$ 157 ratios were determined by combustion to CO₂ in an EA-1120 elemental analyser on-line to an IRMS (Delta+XL, 158 ThermoFinnigan at the NERC Isotope Geosciences Laboratory), with ${}^{13}C/{}^{12}C$ ratios calculated as $\delta^{13}C$ values versus 159 VPDB by comparison with laboratory standards calibrated against NBS 19 and IAEA CH-7. Analytical reproducibil-160 ity was typically $\leq 0.2\%$ (1 SD).

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Ē of water of Table 1: Sun

162 3.5. Sequencing of bacterial 16S rRNA genes

Four surface sediment samples were sequenced: two from Fardalen collected in spring at the water sampling 163 location, L (river sediment) and M (sediment by the side of the river resting on snow) and two from Dryadbreen 164 collected in summer, A (sediment from a pool of water in the sandur, not connected to main river) and O (sediment 165 adjacent to the river at water sampling location). Sediment samples were scooped directly into either sterile 300 mL 166 PVC containers or sterile 50 mL centrifuge tubes. The samples were stored at ambient temperature (<4 °C) until they 167 were transported to the lab where they were desiccated by drying at 40 °C (4 days). Desiccated samples were shipped 168 internationally to the USA where they were subjected to molecular analyses. 169 Nucleic Acid Extraction and Quantification. DNA extraction and purification were carried out with a Fast 170 DNA Spin Kit for Soil (MP Biomedicals, Solon, OH). DNA was extracted in triplicate from three independent 250 171 mg subsamples of sediment. Equal volumes of each replicate extract were pooled and the concentration of DNA 172 was determined using a Qubit dsDNA HS Assay kit (Molecular Probes, Eugene, OR) and a Qubit 2.0 Fluorometer 173

174 (Invitrogen, Carlsbad, CA).

PCR Amplification of bacterial and archaeal 16S rRNA Genes from Genomic DNA. Purified genomic DNA
 extracts were subjected to amplification of bacterial and archaeal 16S rDNA using primers

344F (5'-ACGGGGGYGCAGCAGGCGCGA-3') and 915R (5'-GTGCTCCCCCGCCAATTCCT-3') at an annealing 177 temperature of 61°C or 1100F (5'-YAACGAGCGCAACCC-3') and 1492R (5'-GGTTACCTTGTTACGACTT-3') at 178 an annealing temperature of 55 °C, respectively (Boyd et al., 2007). Approximately 1 ng of purified genomic DNA 179 was subjected to PCR in triplicate using the following reaction conditions: initial denaturation at 94 °C (4 min), 180 followed by 35 cycles denaturation at 94 °C (1 min), annealing at the optimal temperature for the primer pair (1 min), 181 primer extension at 72 °C (1.5 min), followed by a final extension step at 72 °C for 20 min. Reactions contained 2 mM 182 MgCl₂ (Invitrogen, Carlsbad, CA), 200 µM each deoxynucleotide triphosphate (Eppendorf, Hamburg, Germany), 0.5 183 μ M forward and reverse primer (Integrated DNA Technologies, Coralville, IA), 0.4 mg ml⁻¹ molecular-grade bovine 184 serum albumin (Roche, Indianapolis, IN), and 0.25 units Taq DNA polymerase (Invitrogen, Carlsbad, CA) in a final 185 reaction volume of 50 μ L. Positive control reactions were performed using genomic DNA from Azotobacter vinelandii 186 DJ and Sulfolobus solfataricus P2. Negative control reactions were performed in the absence of added genomic DNA. 187 PCR amplicons were only obtained from extracts when bacterial primer sets were applied. Archaeal 16S rRNA gene 188 amplicons were not recovered from any of the four sediment DNA extracts. 189

Sequencing and analysis of bacterial 16S rRNA genes. Bacterial 16S rDNA amplicons were sequenced by Molecular Research LC (Lubbock, TX). 16S rDNA from each site were barcoded and were sequenced using a 454 Genome Sequencer FLX System. Post sequence processing was performed using the Mothur (ver. 1.33.3) sequence analysis platform (Schloss et al., 2009). Raw libraries were trimmed to a minimum length of 250 bases and were subjected to a filtering step using the quality scores file to remove sequences with anomalous base calls. Unique sequences were aligned using SILVA databases and sequences were trimmed using a defined start and end sites and a maximum length of 250 bases. The resulting unique sequences were pre-clustered to remove amplification and sequencing errors

¹⁹⁷ and chimeras were identified and removed using UCHIME (Edgar et al., 2011). Operational taxonomic units (OTUs) ¹⁹⁸ were assigned at a sequence similarity of 0.03 using the furthest-neighbour method. The remaining sequences were ¹⁹⁹ randomly sub-sampled in order to normalize the total number of sequences in each library. Collectively, these steps ²⁰⁰ resulted in a normalized size of 1077 bacterial 16S rRNA gene sequences for each assemblage. Sequences were clas-²⁰¹ sified using the Bayesian classifier (Wang et al., 2007) and the RDP database, with manual verification using BLASTn ²⁰² (Supp. Table 1). Sequences representing each OTU have been deposited in the NCBI SRA database under accession ²⁰³ number SRR1562043.

Sequences representing each unique OTU (defined at 0.03% sequence identities) were compiled for each domain. 204 ClustalX (ver. 2.0.9, Larkin et al., 2007) was used to align nucleic acid sequences using default parameters and the 205 alignment was subjected to evolutionary model prediction via jModeltest (ver. 2.1.1, Darriba et al., 2012), Maximum-206 Likelihood phylogenetic reconstruction via PhyML (version 3.0, Guindon and Gascuel, 2003) specifying the general 207 time reversible model and gamma distributed rate variation with a proportion of invariable sites, and rate smoothing 208 using the multidimensional version of Rambaut's parameterization as implemented in PAUP (ver. 4.0, Swofford, 209 2001) as previously described (Meuser et al., 2013). Phylocom was used to calculate Rao's community phylogenetic 210 relatedness for the bacterial assemblages using relative sequence abundance weights and the rate-smoothed ultrameric 211 tree. PAST (ver. 1.72, Hammer et al., 2001) was used to generate cluster dendograms specifying single linkage and 212 Euclidean distances. Bootstrap values correspond to the frequency that each node was observed in a given position 213 out of 1000 replicates. 214

215 3.6. Precipitation correction

Snow is the primary source of precipitation to the two studied catchments, but the chemical composition of the water derived from melting snow varies temporally (e.g. Johannessen and Henriksen, 1978). Therefore instead of using the snow samples collected for this study, which were collected relatively late in the season as a measure of precipitation inputs, we compiled literature data on pre-melt Svalbard snow-pack chemistry (Hodgkins et al., 1997; Wynn et al., 2006; Tye and Heaton, 2007; Yde et al., 2008). Pre-melt snow-pack samples are typically taken in April and are assumed to represent "fresh" snow. From these data average *X*/Cl ratios were calculated, without any weighting, where *X* is a major cation or anion. By assuming that chloride is conservative and only derived from precipitation, the stream water data are corrected for snow inputs using the following formula:

$$X^* = X_{river} - (X/Cl)_{snow} \cdot Cl_{river}$$
(1)

where X^* and X_{river} denote precipitation-corrected and uncorrected concentrations respectively. The *X*/Cl ratios used for the precipitation correction were (±1SD, n=8-10): Ca/Cl 0.11±0.10, Mg/Cl 0.09±0.04, Na/Cl 0.85±0.09, K/Cl 0.02±0.01 and SO₄/Cl 0.11±0.04. The propagated error on the precipitation corrected values in summer was less than 7% (RSD) for K and Na and less than 4% (RSD) for Ca, Mg and SO₄. In spring, the propagated error was higher

Sample	SO ₄ in	8 ³⁴ S	δ ³⁴ S*	δ^{18} Oso4	δ ¹⁸ Oso4*						
bumpie	precipitation	0 0	0 0	0 0304	0 0304						
(YYYYMMDD)	%	‰	‰	‰	‰						
Dryadbreen (glacia	ted) - spring										
20120615D	21.3	-0.6	-4.5	-6.0	-10.2						
20120617D	29.0	0.7	-4.7	-5.4	-11.5						
Dryadbreen (glaciated) - summer											
20120725D	4.8	-2.0	-2.8	-7.7	-8.6						
20120727D	3.5	-2.3	-2.9	-8.1	-8.7						
20120729D	3.8	-2.3	-2.9	-8.0	-8.7						
20120731D	3.6	-2.6	-3.2	-9.7	-10.4						
20120802D	4.9	-2.3	-3.1	-8.5	-9.4						
Fardalen (unglaciat	ed) - spring										
20120614F	5.6	-3.6	-4.6	-8.1	-9.2						
20120616F	6.4	-3.7	-4.9	-8.1	-9.2						
20120618F	5.3	-3.4	-4.4	-8.4	-9.4						
Fardalen (unglaciat	ed) - summer										
20120726F	1.1	-6.7	-6.9	-9.7	-9.9						
20120728F	0.9	-7.1	-7.3	-9.5	-9.7						
20120730F	1.0	-7.0	-7.2	-9.7	-9.9						
20120801F	0.8	-7.7	-7.8	-10.1	-10.2						
20120803F	0.9	-7.1	-7.3	-10.1	-10.2						

Table 2: Precipitation-corrected δ^{34} S and $\delta^{18}O_{SO4}$ values ($\delta^{34}S^*$ and $\delta^{18}O_{SO4}^*$).

²²⁸ due to the greater amount of snow-melt and the error was consistently higher in Dryadbreen. Errors were less than ²²⁹ 25% (RSD) except for Na in Dryadbreen which were closer to 50% (RSD). The δ^{34} S and $\delta^{18}O_{SO4}$ values were also ²³⁰ corrected for snow inputs (Table 2). Reported values for the δ^{34} S composition of snow in Svalbard vary from +11.5 ²³¹ to +18.0% (Tye and Heaton, 2007). To estimate a value of δ^{34} S for the snow pack we took the percentage of SO₄ ²³² derived from snow for the spring samples and extrapolated to a δ^{34} S value at 100% SO₄ from snow (Fig. 3). The value ²³³ estimated for snow using this approach was +14‰ which is within the range previously reported (Tye and Heaton, ²³⁴ 2007). Using this value, δ^{34} S values were corrected using the following formula:

$$\delta^{34}S^* = (\delta^{34}S_{river} - f \cdot \delta^{34}S_{snow})/(1 - f)$$
(2)

where *f* is the fraction of SO₄ from snowmelt. $\delta^{18}O_{SO4}$ values were corrected in an identical manner to that described above for $\delta^{34}S$ using a snow value of +9.5% based on a fresh snow sample from Svalbard (Tye and Heaton, 2007). Unless otherwise stated, precipitation-corrected values are used in all figures in this manuscript and are indicated by an asterisk.

239 4. Results

240 4.1. Water chemistry and hydrology

The water samples taken in spring were dominated by snow-melt as evinced by the high proportion of Cl and Na compared to the summer samples (Fig. 4, Table 1). The most abundant anion in Dryadbreen was HCO_3^- which comprised 48-58% of the major anions (in meq) in summer. This was in stark contrast to Fardalen where SO_4^{2-} was the major anion comprising 72-86% of the major anions (in meq) in summer (Fig. 4). The precipitation corrected



Figure 3: Stream water δ^{34} S values against the calculated %SO₄ derived from precipitation. The regression line is derived from the spring samples only. The δ^{34} S value when 100% SO₄ is derived from precipitation is +14‰ which is comparable with the +11.5 to +18.0‰ range reported for snow (fresh snow and snow-pack) from Svalbard by Tye and Heaton (2007). The δ^{34} S value when 0% SO₄ is derived from precipitation is -4.6±0.2‰ and is assumed to represent sulfate derived from a pyrite weathering source. The summer samples do not lie on the mixing line between snow and a pyrite source which either indicates additional pyrite sources with different δ^{34} S values or fractionation processes. Error bars smaller than symbol size are not depicted.

- ²⁴⁵ abundances of cations were similar in both catchments (Fig. 4). In Dryadbreen precipitation sources accounted for ²⁴⁶ 34-48% of the cation abundance (in meq) in spring and 10% in summer. In Fardalen, the precipitation contribution ²⁴⁷ to the cation abundance was slightly less: 19-29% in spring and 4% in summer. The abundances of Ca and Mg were ²⁴⁸ very similar in both catchments but in Fardalen Mg was consistently more abundant compared to Dryadbreen where ²⁴⁹ Ca was most abundant. The total dissolved load of Fardalen (53±7 mgL⁻¹, 1SD) was approximately double that of ²⁵⁰ Dryadbreen (25±4 mgL⁻¹, 1SD).
- The $\delta^{18}O_{H2O}$ value of the stream water varied from around -15‰ in spring to around -13‰ in summer (Table 1), reflecting the decrease in snow cover from spring to summer (Hindshaw et al., 2011). There is no significant difference in $\delta^{18}O_{H2O}$ values between the two catchments reflecting a common precipitation source. The sulfur isotopic composition of sulfate ($\delta^{34}S$) decreased from spring to summer: from +0.73 to -2.64‰ in Dryadbreen and from -3.40 to -7.66‰ in Fardalen (Table 1). Similar to $\delta^{34}S$ values, $\delta^{18}O_{SO4}$ values exhibited a seasonal shift towards lower values in summer with the lowest value (-10.1‰, Table 1) measured in Fardalen. The discharge of both streams at the time of sampling is given in Table 1. Both streams exhibit diurnal cycles in
- discharge and the range of discharge measured in both catchments was $0-0.5 \text{ m}^3 \text{s}^{-1}$ (not including periods of logger
- malfunction), but the median discharge over the period of data collection for Dryadbreen (0.40 m^3s^{-1}) was greater
- than for Fardalen $(0.22 \text{ m}^3 \text{s}^{-1})$.

Sample	Class ¹	Fm. ²	SiO ₂ wt%	Al ₂ O ₃ wt%	Fe ₂ O ₃ wt%	TiO ₂ wt%	MgO wt%	CaO wt%	Na ₂ O wt%	K ₂ O wt%	MnO wt%	P ₂ O ₅ wt%	LOI ³ wt%	Total wt%
R01	Shale	Fry.	63.2	16.3	7.1	0.8	1.4	0.3	1.0	2.6	0.1	0.3	7.5	101.0
R02	Wacke	Batt.	65.4	13.0	7.9	0.6	1.5	0.6	0.7	2.2	0.1	0.3	7.2	99.5
R04	Shale	Fry.	57.1	19.1	3.5	0.9	1.3	0.2	0.6	3.6	0.0	0.1	12.9	99.3

¹ Classification using SandClass for terrigenous sands and shales (Herron, 1988).

² Formation assignment based on Schlegel et al. (2013). Fry. = Frysjaodden Formation and Batt. = Battfjellet Formation.

³ LOI = Loss on Ignition.

261 4.2. Solid samples

Based on the SandClass system for terrigenous sand and shale samples (Herron, 1988), R01 and R04 were clas-262 sified as shales and R02 as a wacke (Table 3). Schlegel et al. (2013) have previously classified rock core samples from the Frysajodden Formation as shales and those from the Battfiellet and Aspelintoppen formations as wackes and 264 litharenites respectively. We therefore assume that R01 and R04 originated from the Frysajodden Formation and R02 265 from the Battfjellet Formation (Fig. 2). The main minerals in the bulk rock and sediment samples analysed by XRD 266 were quartz, plagioclase, chlorite, kaolinite, illite/mica and illite/smectite (Table 4). Clay minerals accounted for 44% 267 (sediment O) to 65% (R04) of the total composition. Calcite abundance was below 1% in all samples analysed. How-268 ever, calcite was detected in XRD analysis of orientated clay fractions from river sediments collected in the glaciated 269 catchment (Dryadbreen) but not in sediments collected from the unglaciated catchment (Fardalen). The low calcite 270 abundance is in agreement with previous studies which report <1-2% carbonate in core samples from these formations 27 (Dypvik et al., 2011; Schlegel et al., 2013). Calcite in rock sample R02 had a δ^{13} C value of -1.7% (Table 5). The 272 δ^{13} C values of suspended sediments show neither temporal nor spatial variation and the average value of -26% (Table 273 5) reflects carbon fixed by the Calvin-Benson-Bassham cycle (C3) which may be of plant (Kendall and Doctor, 2003) 274 or microbial origin (Havig et al., 2011). 275 Pyrite was detected in the bulk phase XRD analysis of R02 (Table 4) and was separated from both R01 and R02. 276 No gypsum was detected by XRD analysis and none was detected in thin sections from these same formations (Dypvik 277

et al., 2011; Schlegel et al., 2013). The S content of R01 and R02 was 0.02 and 0.90 wt% respectively. Assuming that all the S is from pyrite with a formula FeS₂, then this corresponds to 0.04 and 1.69 wt% pyrite in the two rocks samples, in agreement with the XRD data (Table 4). It is therefore highly likely that pyrite is the main S-bearing mineral phase in these rock samples. The δ^{34} S values of pyrite mineral separates from samples R01 and R02 were -5.4±0.4‰ and -7.6±0.2‰ (1SD) respectively (Table 5). The difference in values is likely due to the two samples being derived from different formations (Table 3).

284 4.3. Bacterial community composition

The phylogenetic affiliation of the 16S rRNA genes recovered from the sediment samples, as determined by BLASTn analysis, are depicted in Fig. 5.

Mineral		5	Sample		
	R01	R02	R04	O^1	L^1
Quartz	38	46	31	51	40
Plagioclase	7	3	3	5	6
K-Feldspar	<1	<1	1	<1	<1
Chlorite	9	8	11	12	5
Illite/Mica	26	24	28	17	32
Illite/Smectite	8	7	9	<3	<3
Kaolinite	12	9	17	15	17
Pyrite	<1	1	<1	<1	<1
Siderite	<1	3	<1	<1	<1
Calcite	<1	<1	<1	<1	<1
¹ O = Dryadbree	n sedim	ent, L = 1	Fardalen	sedime	ent

Table 4: Semi-quantitative mineral abundances (% total) in rock and sediment samples.

Table 5: Summary of C contents and C and S isotope data of solid samples.

Sample	C wt%	1SD*	δ ¹³ C ‰	1SD*	δ ³⁴ S ‰	1SD*
>0.2µm suspen	ded sed	iments				
Dryadbreen (gla	aciated)					
20120615D#	4.07		-26.1			
20120617D#	2.88		-26.3			
20120725D	2.17		-25.6			
20120725D#	2.33	0.10	-26.2	0.1		
20120727D#	2.17		-26.2			
20120729D	2.60		-25.7			
20120729D#	2.64	0.25	-26.2	0.0		
20120731D	2.46		-25.6			
20120731D#	2.30	0.10	-26.2	0.1		
20120802D#	2.54		-26.4			
Fardalen (ungla	iciated)					
20120616F#	2.43		-26.4			
20120618F#	2.02		-26.6			
20120726F#	2.59		-26.8			
20120728F#	2.63		-26.0			
20120730F#	2.49		-26.5			
Pyrite mineral	separat	es				
R01	-				-5.4	0.4
R02					-7.6	0.2
Calcite minera	l separa	te				
R02			-1.7			
*Errors reported for repeat measu # Denotes that s carbonates.	l where t urements ample w	here was as treated	enough sa l with 5%	mple mat HCl to re	erial move	

Clear differences were observed in the composition of sediment associated bacterial communities that were sub-287 merged in the stream versus those that were collected on the banks of the stream. For example, samples M (Fardalen, 288 sediment on snow adjacent to river) and O (Dryadbreen, sediment adjacent to river) both comprised large numbers 289 of firmicutes, whereas the two samples collected under water (A and L) contained very low abundances of these lin-290 eages (Fig. 5b). A (Dryadbreen) and L (Fardalen) were dominated by sequences affiliated with Proteobacteria (48 29 and 65% of the total bacterial community, respectively) with those affiliated with Betaproteobacteria representing the 292 most abundant proteobacterial class (21% and 31% respectively). Sequencing studies from glaciated environments 293 have found that Betaproteobacteria are the dominant phylogenetic group in subglacial and proglacial sediments where 294 they are likely to be involved in S and Fe cycling under oxic conditions (Foght et al., 2004; Skidmore et al., 2005; 295



Figure 4: Major ion water chemistry (in meq) shown in ternary diagrams. Uncorrected data are shown as open diamonds and precipitation corrected data are shown as filled-in circles. The data are further divided into samples collected in spring and those collected in summer. The striking difference between the two catchments is in their relative anion abundances with Dryadbreen (glaciated) dominated by HCO_3^- and Fardalen (unglaciated) dominated by SO_4^{2-} . Error bars are smaller than the symbols.

²⁹⁶ Mitchell et al., 2013).

Differences in the composition of the bacterial communities associated with sediments sampled from the two sub-297 merged sites A (Dryadbreen) and L (Fardalen) were also apparent. For example, sequences affiliated with the anaero-298 bic phylum Bacteriodetes represented the most abundant phylum (24%) in the community sampled from Dryadbreen 299 (A), whereas Bacteriodetes represented only a minor fraction of the community (3%) from Fardalen (L). This may 300 indicate a shift toward more anaerobic metabolisms in the proglacial sediment bacterial communities at Dryadbreen. 301 In the discussion we will focus on sediment samples A and L as representative of each catchment since they 302 were collected under water and therefore are most likely to be adapted to or responsible for the differences in water 303 chemistry observed between the two catchments. 304

305 5. Discussion

306 5.1. Overall weathering reactions

³⁰⁷ The most obvious difference between the two streams is in the anion composition: Dryadbreen is dominated by

 HCO_3^- , whereas Fardalen is dominated by SO_4^{2-} . This indicates that different weathering reactions are occurring

³⁰⁹ in the two catchments despite nominally identical lithology. A first indication of which weathering reactions are

³¹⁰ important for each catchment can be gained by looking at element ratios since the reactions of carbonates and silicates



Figure 5: (a) Dendrogram based on the Rao phylogenetic diversity index (Rao, 1982) which depicts the degree of phylogenetic dissimilarity between the bacterial communities associated with the different sediment samples. Samples A and O were collected in Dryadbreen (red letters) and L and M were collected in Fardalen (blue letters). Samples A and L were collected under water and form a cluster, indicating they are more similar in their bacterial phylogenetic composition as compared to samples O and M. Sample O was collected adjacent to the river and sample M was collected adjacent to the river but on top of snow. (b) Relative abundance of bacterial phyla or classes (Proteobacteria only) based on percent identities to their closest cultivated sequence, as determined by BLASTn analysis. The phyla Armatimonadetes, Chlamydiae, Gemmatimonadetes, Nitrospirae, Nitrospirae, Planctomycetes, TM7, Verrucomicrobia and WS3 are grouped together in 'Other'.

- with either carbonic or sulfuric acids will give distinct ratios of cations versus SO_4^{2-} and HCO_3^{-} (Fairchild et al.,
- 1994; Tranter et al., 2002; Wadham et al., 2010). Example reactions and their corresponding slopes in units of
- equivalents (eq) are presented below (Eqns. 3-6) and in Table 6. Feldspars are the primary silicate minerals in these
- rocks and although feldspars are a compositional solid solution between CaAl₂Si₂O₈, NaAlSi₃O₈ and KAlSi₃O₈, the
- variation in composition does not affect the ratio (in equivalents) of the product cation (Na^+ , K^+ , Ca^{2+}) to either the
- $_{316}$ HCO₃⁻ or the SO₄²⁻ produced (cf. Eqns 5a, 5b, 6a and 6b). Analogous equations can be written for the main Mg-
- ³¹⁷ bearing silicate phases: illite ((A1,Mg,Fe)₂(Si,Al)₄O₁₀OH₂) and chlorite ((Mg,Fe)₃(Si,Al)₄O₁₀) (Table 4). In addition
- to mineral weathering, the oxidation of organic matter will produce HCO_3^- , and this reaction, which is expected to
- ³¹⁹ occur in these permafrost-dominated catchments, is represented by equation 7.

Carbonate dissolution with carbonic acid (CDC)	
$Ca_{1-x}(Mg_x)CO_3(s) + CO_2(aq) + H_2O(l) \rightleftharpoons (1-x)Ca^{2+}(aq) + xMg^{2+}(aq) + 2HCO_3^{-}(aq)$	(3)
Aerobic sulfide oxidation coupled to carbonate dissolution (SOCD)	
$16Ca_{1-x}(Mg_x)CO_3(s) + 4FeS_2(s) + 15O_2(aq) + 14H_2O(l) \Longrightarrow$	
$16(1-x)Ca^{2+}(aq) + 16xMg^{2+}(aq) + 16HCO_{3}^{-}(aq) + 8SO_{4}^{2-}(aq) + 4Fe(OH)_{3}(s)$	(4a)
Anaerobic sulfide oxidation coupled to carbonate dissolution (SOCD)	
$4Ca_{1-x}(Mg_x)CO_3(s) + FeS_2(s) + 14Fe(OH)_3(s) \Longrightarrow$	
$4(1-x)Ca^{2+}(aq) + 4xMg^{2+}(aq) + 4HCO_{3}^{-}(aq) + 15Fe(OH)_{2}(s) + 2SO_{4}^{2-}(aq) + 4H_{2}O(l)$	(4b)
Silicate dissolution with carbonic acid, divalent cation (SDC)	
$2CaAl_2Si_2O_8(s) + 4CO_2(aq) + 6H_2O(l) \Longrightarrow$	
$2Ca^{2+}(aq) + 4HCO_3^{-}(aq) + Al_4Si_4O_{10}(OH)_8(s)$	(5a)
Silicate dissolution with carbonic acid, monovalent cation (SDC)	
$2NaAlSi_3O_8(s) + 2CO_2(aq) + 11H_2O(l) \Longrightarrow$	
$2Na^{+}(aq) + 2HCO_{3}^{-}(aq) + Al_{2}Si_{2}O_{5}(OH)_{4}(s) + 4H_{4}SiO_{4}$	(5b)
Sulfide oxidation coupled to silicate dissolution, divalent cation (SOSD)	
$2CaAl_2Si_2O_8(s) + FeS_2(aq) + \frac{15}{2}O_2(aq) + \frac{11}{2}H_2O(l) \Longrightarrow$	
$2Ca^{2+}(aq) + 2SO_4^{2-}(aq) + Al_4Si_4O_{10}(OH)_8(s) + Fe(OH)_3(s)$	(6a)
Sulfide oxidation coupled to silicate dissolution, monovalent cation (SOSD)	
$16NaSi_3O_8(s) + 4FeS_2(aq) + 15O_2(aq) + 86H_2O(l) \Longrightarrow$	
$16Na^{+}(aq) + 8SO_{4}^{2-}(aq) + 4Al_{4}Si_{4}O_{10}(OH)_{8}(s) + 4Fe(OH)_{3}(s) + 32H_{4}SiO_{4}(aq)$	(6b)
Organic matter oxidation (OMO)	
$CH_2O + O_2 \rightleftharpoons CO_2 + H_2O \rightleftharpoons HCO_3^- + H^+$	(7)

In the unglaciated catchment, Fardalen, the slope of the data in a plot of total cations versus SO_4^{2-} (Fig. 6a) is 1.04±0.04 which is the same as the slope which would be expected by sulfide oxidation coupled to silicate weathering (SOSD, Eqn. 6). The intercept of this line is not zero which suggests that a portion of the cations are being produced by other weathering reactions such as silicate dissolution by carbonic acid (SDC, Eqn. 5). In Figs. 6b and 6d, the slopes greater than 1 (carbonic acid weathering of silicates and carbonates, Table 6), together with the greater scatter of the Fardalen points support the notion that sulfuric acid is the main weathering agent in this catchment. In summary, although it is not possible to quantify the relative importance of each of the reactions represented by equations 3 -

Reaction	Predicted slope (in eq)								
	Ca ²⁺ +Mg ²⁺	Ca ²⁺ +Mg ²⁺	Total Cations	Total Cations					
	vs SO4 ²⁻	vs HCO3 ⁻	vs SO4 ²⁻	vs HCO ₃ ⁻					
Carbonate + H_2CO_3 (Eqn. 3)	00	1	-	•					
Carbonate + H_2SO_4 (Eqn. 4a,b)	2	2							
Silicate + H ₂ CO ₃ (Eqn. 5a,b)	-	-	~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~	1					
Silicate + H ₂ SO ₄ (Eqn. 6a,b)	-	-	1	00					
		Observ	ved slope						
Dryadbreen	1.54 ± 0.14	1.48±0.29	1.84±0.18	2.04±0.42					
Fardalen	$0.88 {\pm} 0.04$	6.68 ± 2.10	1.04 ± 0.04	7.92 ± 2.50					

Table 6: Predicted and observed slopes of Ca+Mg and total cations versus either SO₄²⁻ or HCO₃⁻.

7, the data from Fardalen is consistent with weathering mainly by a sulfide oxidation coupled to silicate dissolution
 process.

³²⁹ If only silicate minerals were weathering in the glaciated catchment (Dryadbreen) then the slopes of the data in ³³⁰ Figs. 6a and 6b, which lie intermediate between SOSD (Eqn. 6) and SDC (Eqn. 5), would suggest that silicates ³³¹ are weathered by both carbonic and sulfuric acid. However, if it were assumed that all the Ca and Mg came from ³³² carbonates then the data points would also be consistent with carbonate weathering by sulfuric and carbonic acids ³³³ (Figs. 6c and 6d). In summary, the data from Dryadbreen suggests that all four reactions are occurring i.e. weathering ³³⁴ of both silicates and carbonates by both carbonic and sulfuric acid and further data from, for example, δ^{13} C-DIC and ³³⁵ ⁸⁷Sr/⁸⁶Sr, would be needed to constrain the relative proportions of the different reactions.

The high proportion of silicate weathering in the unglaciated catchment is likely due to the presence of the active 336 layer (the seasonally thawed top layer of soil above permafrost). If, over a long enough period, the active layer has 337 remained at a constant depth, this would mean that despite frost-shattering exposing fresh mineral surfaces, the vast 338 majority of carbonate phases will have already been leached, leaving behind the silicate minerals to weather. This 339 is in agreement with the lack of carbonate detected by XRD in the orientated clay fractions of solid samples from 340 Fardalen (Section 4.2). However, if the active layer deepens through a warming climate then fresh carbonates could 341 become exposed (Keller et al., 2010). In contrast, high rates of physical erosion still occur in the glaciated catchment 342 exposing carbonates to chemical weathering. 343

The nature of the weathering reactions in the glaciated catchment is in agreement with the conclusions of Tranter et al. (2002) who identified sulfide oxidation and Fe(III) reduction as key reactions in subglacial environments. The importance of sulfide oxidation reactions is also apparent in high-latitude unglaciated, permafrost dominated catchments (Fig. 6). This is in contrast to low-latitude, unglaciated catchments where weathering reactions are mainly driven by reaction with CO₂ either from respiration or from the atmosphere (Tranter et al., 2002).

Although the information in Fig. 6 provides an overview of the weathering reactions occurring in each catchment, further processes may modify these broad interpretations. Firstly, it is assumed that reactions have idealised stoichiometry and secondly, reactions removing solutes are not considered. The latter point is especially critical in subglacial environments where carbonate precipitation and sulfate reduction can occur (Wadham et al., 2004). Carbonate precipitation would decrease Ca^{2+} and HCO_3^- , whereas sulfate reduction would decrease SO_4^{2-} . All water

samples were undersaturated with respect to calcite and we therefore assume that carbonate precipitation did not have a significant impact on the stream water chemistry. In the following section we utilise the fact that the isotopes of sulfate (δ^{34} S and δ^{18} O_{SO4}) are fractionated during reduction and can therefore be used to assess whether sulfate reduction is occurring and under what conditions.

358 5.2. Oxic or anoxic weathering conditions?

Sulfate reduction only occurs under anoxic conditions and previous studies have distinguished between aerobic and anaerobic environments based on the source of O atoms in sulfate (e.g. Bottrell and Tranter, 2002; Wadham et al., 2004). There are two main sources of oxygen: atmospheric O_2 and aqueous oxygen (H₂O) and each has a distinct isotopic composition.

³⁶³ During the oxidation of sulfide to sulfate the exchange of oxygen atoms in the intermediate molecules is very ³⁶⁴ rapid but once sulfate has formed then exchange is negligible (Lloyd, 1968). Previously it has been assumed that if ³⁶⁵ more than 75% of the O atoms in sulfate derive from water then the pyrite was oxidised under anoxic conditions and ³⁶⁶ less than 75% implied oxidation under oxic conditions (Taylor et al., 1984; van Everdingen and Krouse, 1985). The ³⁶⁷ fraction of O atoms in sulfate derived from atmospheric O₂ or water can be calculated as follows (Balci et al., 2007):

$$\delta^{18}O *_{\text{SO4}} = x(\delta^{18}O_{\text{H2O}} + \varepsilon^{18}O_{\text{SO4-H2O}}) + (1 - x)(\delta^{18}O_{\text{O2}} + \varepsilon^{18}O_{\text{SO4-O2}})$$
(8)

where x is the fraction of O atoms derived from H₂O and ε are the fractionation factors. When this principle was 368 applied to glaciated catchments it was found that FeS2 oxidation in subglacial environments proceeded primarily 369 under anoxic conditions (Bottrell and Tranter, 2002; Wadham et al., 2004; Wynn et al., 2006; Wadham et al., 2007). 370 However, recent studies have questioned the utility of $\delta^{18}O_{SO4}$ isotopes in distinguishing between oxic and anoxic 371 conditions. These studies (Usher et al., 2004, 2005; Chandra and Gerson, 2011) found that even in solutions containing 372 dissolved O_2 the vast majority of the oxygen atoms in sulfate derive from water because water outcompetes O_2 for 373 adsorption sites. The percentage of oxygen atoms derived from water can therefore not be used to assess whether 374 anoxic conditions are present or not during the oxidation of sulfide. 375

The bacterial sequencing data can, however, be used to gain a rough idea of the redox status of the two catchments. 376 The samples from Dryadbreen were collected in summer whilst those from Fardalen were collected in spring and 377 therefore may not be directly comparable as a result of seasonal variation in the microbial community composition 378 (Crump et al., 2009; Schostag et al., 2015). However, a study on seasonal variations in the microbial community 379 composition in soils found that the relative abundance of most phyla was constant over a year (Schostag et al., 2015). 380 In addition, Arctic stream microbial community compositions were found to be strongly correlated with inorganic 381 stream water chemistry (Crump et al., 2009) and given that there is more variation in relative ion proportions between 382 the catchments than between the seasons (Fig. 6), we assume limited seasonal variability in the microbial community 383 compositions for the purposes of this discussion. 384

Phylum/Class	Order	Genera	А	0	L	М					
Reduces Fe(III)											
Acidobacteria	Holophagales	Geothrix	1.9	0.4	0.5	7.9					
Betaproteobacteria	Burkholderiales	Albidiferax	0.6	1.1	0.2	1.3					
Betaproteobacteria	Burkholderiales	Rhodoferax ferrireducens	2.9	4.1	0.4	1.3					
Deltaproteobacteria	Desulfuromonadales	Geobacter	8.1	1.8	0.0	3.9					
Oxidizes Fe(II)											
Betaproteobacteria	Ferritrophicales	Ferritrophicum radicicola	0.5	0.0	0.0	0.0					
Betaproteobacteria	Gallionellales	Sideroxydans lithotrophicus	0.6	1.1	3.4	0.1					
Betaproteobacteria	Hydrogenophilales	Thiobacillus	5.9	0.1	2.9	0.8					
Oxidizes reduced S compounds											
Betaproteobacteria	Burkholderiales	Thiobacter subterraneus	0.0	0.1	0.8	0.0					
Betaproteobacteria	Hydrogenophilales	Sulfuricella	0.9	0.0	0.8	0.4					
Betaproteobacteria	Hydrogenophilales	Thiobacillus	5.9	0.1	2.9	0.8					
Gammaproteobacteria	Chromatiales	Halothiobacillus neapolitanus	0.0	0.0	0.5	0.0					
Gammaproteobacteria	Chromatiales	Thioalkalivibrio versutus	0.0	0.0	0.1	0.0					
Gammaproteobacteria	Chromatiales	Thiocapsa machilipatnamensis	0.0	0.0	0.5	0.0					
1	R	educes sulfate									
Deltaproteobacteria	Desulfobacterales	Desulfatiferula	0.0	0.1	0.0	0.0					
Deltaproteobacteria	Desulfobacterales	Desulfobulbus	0.0	0.0	0.0	0.1					
Deltaproteobacteria	Desulfobacterales	unclassified	0.1	0.4	0.0	0.4					
Firmicutes	Clostridiales	Desulfosporosinus	0.3	0.6	0.0	5.2					

Table 7: Abundance (%) of bacteria associated with Fe and S cycling. Sites A and O are from Dryadbreen and site L and M are from Fardalen.

Some bacteria only live in anaerobic environments and therefore their presence can be used as an indication 385 that anoxic conditions are present. In both catchments, we found evidence for both anaerobic and aerobic bacteria 386 in the under-water sites (A and L) suggesting that anoxic micro-niches exist in the aqueous system. Sequences 387 affiliated with Thiobacillus, a facultative aerobe capable of oxidizing iron and sulfur compounds was represented in 388 both catchments (Table 7). It is likely that Thiobacillus is involved in the aerobic oxidation of pyrite and as a result 389 may help establish anoxic conditions. However, the Deltaproteobacteria class of bacteria, which consists of numerous 390 lithotrophic anaerobes (Kuever et al., 2005), was strongly represented in Dryadbreen (glaciated, Site A, 10.3%) but 391 was nearly absent in Fardalen (unglaciated, Site L, 0.1%) (Fig. 5), suggesting that anoxic conditions may be more 392 prevalent in the glaciated catchment and that iron reduction may be an important process. In support of this notion, 393 a number of genera capable of reducing iron compounds were identified (Table 7). For example, 8.1% and 2.9% 394 of the sequences from Dryadbreen (A) were closely affiliated with Geobacter (obligate anaerobe, Lovley, 2000) and 395 Rhodoferax (facultative anaerobe, Finneran et al., 2003) respectively (Table 7). Anoxia may be more prevalent in 396 the glaciated catchment due to the sandur, which is a large wet area covered by fine glacial flour, conditions which 397 would favour pyrite oxidation and the development of anoxia, analogous to marine sediments (e.g. Burdige, 1993). 398 Additionally, the sandur is underlain by ice (Ziaja and Pipała, 2007) which could be hydraulically linked to the 399 subglacial drainage system providing a pathway for anoxic subglacial meltwater to enter the stream downstream of 400 the apparent glacier front (Irvine-Fynn et al., 2011). In the unglaciated catchment, the aforementioned processes are 401 not relevant. Although the deeper permafrost areas will be anoxic, they may not have a strong hydraulic connection 402 to the stream, resulting in a stream which is dominantly oxic. 403

404 5.3. Sulfur isotope fractionation: oxidation or reduction?

Sulfate reduction only occurs in anaerobic environments and in low-temperature natural systems it is biologically mediated (Seal II, 2006). During sulfate reduction both δ^{34} S and $\delta^{18}O_{SO4}$ values in the remaining SO₄ will increase as the light isotope is the preferred reactant. A positive correlation between δ^{34} S and $\delta^{18}O_{SO4}$ has therefore been proposed as diagnostic of sulfate reduction (e.g. Mandernack et al., 2003).

The spring δ^{34} S values from both catchments have similar δ^{34} S values that are within error of one of the measured 409 pyrite mineral separates suggesting that no sulfate reduction has occurred (Fig. 7). The summer δ^{34} S values from 410 Fardalen are about 3^{\%} lighter in δ^{34} S compared to the spring points but have the same isotopic composition as one 411 of the measured pyrite mineral separates, again suggesting the absence of sulfate reduction. The apparent shift in the 412 inferred source pyrite δ^{34} S value could be due to differences in pyrite δ^{34} S values between the different formations 413 (Table 5): waters in spring could predominantly drain the Frysjaodden Formation whereas waters in summer could 414 access formations higher up in the catchment (e.g. Battfjellet) as the snow pack retreats. However, the summer 415 δ^{34} S values from Dryadbreen are heavier in δ^{34} S compared to both measured and inferred pyrite compositions and, 416 additionally, there is also a significant positive correlation between $\Delta^{18}O_{SO4-H2O}$ and $\delta^{34}S$ (R²=0.99, p<0.001, m=6±1) 417 which is suggestive of sulfate reduction (Mandernack et al., 2003; Turchyn et al., 2013). $\Delta^{18}O_{SO4-H2O}$ is used instead 418 of $\delta^{18}O_{SO4}$ in order to remove the effect on $\delta^{18}O_{SO4}$ of the temporal variation in $\delta^{18}O_{H2O}$. The gradient of the slope 419 between $\delta^{18}O_{SO4}$ and $\delta^{34}S$ is thought to give information on reaction pathways. Oxygen isotope variations during 420 reduction are thought to be controlled by a combination of intra-cellular isotope exchange between intermediate sulfur 421 compounds and ambient water (Brunner et al., 2005; Farguhar et al., 2008) and kinetic fractionation at the cell level 422 (Aharon and Fu, 2000; Mandernack et al., 2003). The balance between exchange and kinetic isotope fractionation 423 is dependent on the overall reaction rate. Thus, the exact slope of $\delta^{18}O_{SO4}$ versus $\delta^{34}S$ depends on which microbial 424 species mediate the reaction and the forward and backward reaction rates (Aharon and Fu, 2000; Mandernack et al., 425 2003; Kleikemper et al., 2004; Turchyn et al., 2010; Wankel et al., 2014). The slope of $\Delta^{18}O_{SO4-H2O}$ versus $\delta^{34}S$ from 426 Fig. 7 is 6 which would indicate a slow reaction rate (Brunner et al., 2005). 427

However, if significant sulfate reduction was occurring in Dryadbreen then appreciable amounts of sulfate reducing bacteria would have been expected to be detected using molecular analysis. In the sediment sample (A), only 0.4% of the bacteria were inferred to have the capacity to reduce sulfate. In addition, attempts to amplify the gene encoding a fragment of the alpha and beta subunit of the bisulphate reductase (*dsrAB*, a biomarker for sulfate reduction, Wagner et al., 1998) were not successful in any of the four sediment samples despite 40 cycles of PCR (data not shown). This supports the notion that the observed sulfur isotope fractionation was not due to sulfate reduction.

It is generally assumed that negligible sulfur isotope fractionation occurs during pyrite oxidation, however fractionation can occur under certain conditions. Several biotic and abiotic experiments under anoxic and oxic conditions have observed $\varepsilon_{SO4-pyrite}$ with values between -1.3 and +3.5% (Balci et al., 2007; Pisapia et al., 2007; Brunner et al., 2008). The difference between pyrite and sulfate in the Dryadbreen summer samples is $\Delta^{34}S_{SO4-sulfide} = +2.2$ to +4.8%, depending on which value for $\delta^{34}S_{sulfide}$ is used (Table 5). Positive values, of similar magnitude ($\varepsilon_{SO4-sulfide}$

= +3.5 and +0.4) have previously been reported for the initial, non-stoichiometric stages of pyrite oxidation (Pisapia 439 et al., 2007; Brunner et al., 2008). Non-stoichiometric reactions are expected to occur in areas of significant physical 440 erosion e.g. glaciated catchments, as material is removed before reactions are completed (kinetic limitation, Stallard 441 and Edmond, 1983). Positive values of $\varepsilon_{SO4-sulfide}$ have been attributed to heightened loss of SO₂ in the early stages 442 of the reaction and the breaking of thiosulfate S-S covalent bonds (Pisapia et al., 2007; Brunner et al., 2008). De-443 gassing of SO₂ is likely under acidic conditions, but the pH of this river is around 6.5 (Table 1), suggesting that this 444 mechanism of fractionation is likely to be minor. Similarly, the fractionation of O isotopes in sulfate can occur when 445 sulfite species are present, allowing oxygen isotope exchange with water, enriching sulfite, and ultimately sulfate, 446 with ¹⁸O (Brunner et al., 2008). The $\Delta^{18}O_{SO4-water}$ values from Dryadbreen summer samples (+3.0 to +5.4%) are 447 in agreement with $\varepsilon^{18}O_{SO4-water}$ data from pyrite oxidation experiments (+2.8 to +16%), Balci et al., 2007; Pisapia 448 et al., 2007; Brunner et al., 2008). In conclusion, sulfate enriched in both ³⁴S and ¹⁸O, as observed in Dryadbreen, 449 does not necessarily imply sulfate reduction and can be adequately explained as a result of non-stoichiometric reaction 450 pathways during the oxidation of pyrite. 451

In Dryadbreen (site A) 21.7% of the bacteria (collected in summer) are inferred to be involved in Fe and S cycling 452 reactions and 13.5% are inferred to have the ability to reduce Fe(III) e.g. Geobacter and Rhodoferax (Table 7). 453 Given that pyrite is likely the main source of iron in the system, the obligate anaerobic nature of Geobacter and 454 the facultative anaerobic nature of *Rhodoferax*, are suggestive of the presence and potential importance of anoxic 455 portions of the catchment where the 'anoxic' pyrite oxidation pathway may occur (Eqn 4b, note that this pathway can 456 also occur where oxygen is present e.g. Balci et al. (2007)). The pH of the water is circumneutral, likely due to the 457 buffering capacity of calcite dissolution, and at this pH Fe can be cyclically oxidised and reduced (Moses and Herman, 458 1991) accounting for the presence of bacteria associated with both Fe reduction and Fe oxidation. It is also possible 459 for other compounds to act as the electron acceptor in pyrite reduction such as NO₃ and MnO₂, and it is likely that all 460 will be involved in pyrite oxidation to some extent if they are available in the system (e.g. Burdige and Nealson, 1986; 461 Jørgensen et al., 2009). Indeed, an abundant component of the microbial community identified at Site A (Dryadbreen) 462 is Thiobacillus (Table 7) which can couple FeS₂ oxidation with NO₃⁻ reduction (Bosch et al., 2012). Therefore, 463 the presence of the above-mentioned bacteria indicates that the non-stoichiometric reaction pathway, identified as 464 being responsible for the observed change in δ^{34} S and $\delta^{18}O_{SO4}$ values, most likely occurs under anoxic conditions 465 and involves the reduction of Fe. In Fardalen (site L) the bacteria involved in Fe and S cycling are associated with oxidation of Fe and reduced S compounds, with a much lower fraction ($\sim 1\%$) inferred to be involved in Fe reduction. 467 This corroborates the inference of pyrite oxidation under oxic conditions and isotopic fractionation of S and O in 468 sulfate is not observed due to the complete conversion of sulfide to sulfate (Seal II, 2006). 469

470 6. Wider implications

The isotopic and concentration data suggest that in the unglaciated catchment, mineral dissolution occurs mainly 471 by reaction with sulfuric acid. In the glaciated catchment the balance between sulfuric acid and carbonic acid is 472 more even, but due to the non-stoichiometric conversion of pyrite to sulfate (loss of intermediate S compounds) as 473 indicated by the stable isotope data, weathering by sulfuric acid may be underestimated. This indicates that sulfuric 474 acid weathering is not only important in glaciated catchments (e.g. Wadham et al., 2010) but also in permafrost 475 dominated catchments (Nowak and Hodson, 2015) and could therefore be a widespread phenomenon throughout the 476 permafrost zone where shale exists (approximately 46% of the land draining into the Arctic ocean, Amiotte Suchet 477 et al., 2003). Indeed, data from the Mackenzie river, the fourth largest Arctic river by discharge and with extensive 478 shale and permafrost areas, demonstrates the significant contribution of sulfuric acid weathering (Calmels et al., 2007). 479 Quantification of the contribution of sulfuric acid weathering on a global scale is necessary because, if significant, 480 it weakens the climate-weathering negative feedback because if allows cations to be released without accompanying 48 bicarbonate in the case of silicates and the release of bicarbonate without uptake of atmospheric CO₂ in the case of 482 carbonates (Calmels et al., 2007; Li et al., 2008; Torres et al., 2014). 483

484 7. Conclusions

The presence of a glacier appeared to alter both the bacterial community composition and chemical weathering re-485 actions in a sedimentary catchment containing pyrite in Syalbard, compared to a neighbouring unglaciated catchment. 486 The dominant anion in the unglaciated catchment was SO_4^{2-} whereas in the glaciated catchment the dominant anion 487 was HCO₃⁻. The difference in major anion composition was attributed to differences in the chemical weathering re-488 actions occurring in both catchments: silicate weathering by sulfuric acid in the unglaciated catchment and carbonate 489 and silicate weathering by carbonic and sulfuric acids in the glaciated catchment. We speculate that the high erosion 490 rates in the glaciated catchment continually expose carbonate minerals whereas carbonate has already been leached 491 from the permafrost active layer in the unglaciated catchment. 492

Sulfide oxidation was a key process generating acidity in both catchments but this reaction appeared to occur by 493 different mechanisms in each catchment. In the unglaciated catchment, δ^{34} S values of stream water were identical to 494 those measured in pyrite and, together with the bacterial community composition, this suggested that pyrite oxidation 495 occurred under oxic conditions. A seasonal shift in the absolute δ^{34} S value of stream water was attributed to the 496 draining of different sedimentary formations. In the glaciated catchment, the summer δ^{34} S values were positively 497 correlated to $\delta^{18}O_{SO4}$, suggesting sulfate reduction. However, this process is microbially mediated and a biomarker for 498 sulfate reduction (dsrAB genes) was not detected in stream sediments. Instead, the bacterial composition is consistent 499 with fractionation due to sulfide oxidation under anoxic conditions with loss of sulfur intermediates, inducing isotopic 500 fractionation in S between sulfide and sulfate. The presence of bacteria associated with iron redox cycling suggests 501 the involvement of Fe³⁺ as an electron acceptor in this environment. 502

As the local environment changes due to deglaciation and permafrost thaw, the bacterial community is expected to change in response. As a result, microbial mediated reactions, such as those involving sulfur compounds, will be affected. Thus, changes in the solute composition of streams can be expected as a result of continued warming in the Arctic.

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519 9. References

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Figure 6: Plots of total cations (Ca+Mg+Na+K) versus SO_4^{2-} (a) and HCO_3^- (b) together with $Ca^{2+}+Mg^{2+}$ versus SO_4^{2-} (c) and HCO_3^- (d). Concentrations have been corrected for precipitation inputs apart from HCO_3^- for which the correction is negligible. Orthogonal distance regression lines (solid) for each catchment (summer and spring points combined) together with the theoretical slopes (dashed and red arrows) describing idealised reactions are shown. Figures (a) and (b) depict idealised silicate weathering reactions assuming all the cations are from silicates. The 1:1 line in (a) (long dash) arises from sulfide oxidation coupled to silicate dissolution (SOSD, Eqn. 6). The 1:1 line in (b) (dot-dash) arises from carbonation of silicate (SDC, Eqn. 5). The vertical arrows in (a) and (b) indicate SDC (Eqn. 5) and SOSD (Eqn. 6) respectively i.e. cations are produced without SO_4^{2-} (SDC, Eqn. 5) or HCO_3^- (SOSD, Eqn. 6). Figures (c) and (d) depict idealised carbonate weathering reactions assuming Ca and Mg derive from carbonate only. The 2:1 line (short dash) in both figures represents the expected slope if the only weathering reaction occurring was sulfide oxidation coupled to carbonate dissolution (SOCD, Eqn. 4). This reaction can occur aerobically (Eqn. 4a) and anaerobically (Eqn. 4b). It is not possible to distinguish between the aerobic (Eqn. 4a) or anaerobic (Eqn. 4b) reaction based on reaction stoichiometry alone (Table 6). Note that depending of the exact formula of the reactant and product Fe species (e.g. Fe^{3+} vs $Fe(OH)_3$), slopes in $Ca^{2+}+Mg^{2+}$ vs SO_4 space greater than 2 are possible for the anaerobic reaction without changing the slope in $Ca^{2+}+Mg^{2+}$ vs HCO_3^- space (Tranter et al., 2002). The 1:1 line in (c) (dot-dash) arises from carbonation of carbonate (CDC, Eqn. 3). The vertical arrow in (c) indicates CDC (Eqn. 3) i.e. Ca^{2+} and Mg^{2+} are produced without HCO_3^- . The horizontal arrows in (b) and (d) indicate the production of



Figure 7: Plot of $\Delta^{18}O_{SO4-H2O}$ versus $\delta^{34}S^*$. In order to remove the effect of temporal variations in $\delta^{18}O_{H2O}$, $\Delta^{18}O_{SO4-H2O}$ is used. In the key sp. is spring su. is summer, D is Dryadbreen and F is Fardalen. The vertical red lines indicate the $\delta^{34}S$ values of measured pyrite mineral separates (Table 5) and the grey shading indicates the analytical error (2 SD) on those measurements. The vertical black line is an inferred $\delta^{34}S$ value of SO₄ derived from pyrite oxidation (Fig. 3). The correlation (R²=0.99, p<0.001) between $\Delta^{18}O_{SO4-H2O}$ and $\delta^{34}S$ for Dryadbreen in summer (red points) suggests sulphate reduction and therefore an anoxic environment, however the magnitude of S isotope fractionation is not as great as would be expected for sulfate reduction (Seal II, 2006). Error bars are smaller than the symbols.

Highlights

Comparison of weathering processes in a glaciated and an unglaciated catchment Combination of major elements, δ^{34} S, δ^{18} O-SO₄, δ^{18} O-H₂O and 16S-RNA Mechanism of pyrite oxidation differed between the two catchments No evidence for sulfate reduction

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