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1	Quantification and Genotyping of Aichi Virus 1 in Water Samples in the			
2	Kathmandu Valley, Nepal			
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4	Eiji HARAMOTO ^{1,*} and Masaaki KITAJIMA ²			
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6	¹ Interdisciplinary Center for River Basin Environment, Graduate Faculty of Interdisciplinary			
7	Research, University of Yamanashi, 4-3-11 Takeda, Kofu, Yamanashi 400-8511, Japan			
8	² Division of Environmental Engineering, Faculty of Engineering, Hokkaido University, North 13			
9	West 8, Kita-ku, Sapporo, Hokkaido 060-8628, Japan			
10				
11	*Corresponding author:			
12	Mailing address: Interdisciplinary Center for River Basin Environment, Graduate Faculty of			
13	Interdisciplinary Research, University of Yamanashi, 4-3-11 Takeda, Kofu, Yamanashi 400-8511,			
14	Japan			
15	Tel: +81-55-220-8725			
16	Fax: +81-55-220-8592			
17	E-mail address: eharamoto@yamanashi.ac.jp			
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19 ABSTRACT

21	Aichi virus 1 genomes were detected by quantitative PCR in groundwater from shallow dug
22	(10/22) and tube wells $(1/15)$, river water $(14/14)$, and sewage $(1/1)$, with the maximum
23	concentration of 4.0×10^9 copies/l. Nucleotide sequencing analysis demonstrated the prevalence of
24	genotype B in the virus positive samples.
25	
26	KEYWORDS
27	Aichi virus 1; groundwater; Kathmandu Valley; Picornaviridae
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30 **TEXT**

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Aichi virus 1 (AiV-1) is a single-stranded, positive-sense RNA virus that belongs to the genus *Kobuvirus* in the family *Picornaviridae* (Reuter et al. 2011). AiV-1 is considered as a potential causative agent of viral gastroenteritis in humans, primarily transmitted through contaminated food and water (Reuter et al. 2011).

36 Recent environmental studies have demonstrated a high prevalence of AiV-1 in various types 37 of water samples, such as raw and treated sewage, reclaimed water, and river water (Kitajima and 38 Gerba 2015). Quantitative PCR (qPCR) was employed in some of these studies, obtaining 39 quantitative data on the prevalence of AiV-1 in water, whereas nucleotide sequencing analysis 40 combined with conventional PCR was used to determine the distribution of AiV-1 genotypes 41 (Kitajima and Gerba 2015). Considering that a limited number of studies have been conducted so 42 far on the AiV-1 detection in water and that most of them were conducted in developed countries, it 43 is highly recommended to study the prevalence of AiV-1 in water samples in developing countries.

The Kathmandu Valley, the capital city area of Nepal, is well recognized as an area where waterborne diseases are one of the most serious public health concerns, partially because of low coverages of proper drinking water and wastewater treatment systems. People in the valley mainly depend on groundwater for their domestic water use. Since contamination of groundwater by waterborne pathogens could pose a health risk to humans, their prevalence in water needs to be surveyed prior to the risk estimation. However, a limited number of studies in Nepal have been reported on the prevalence of waterborne pathogens, such as pathogenic bacteria (Inoue et al. 2014; 51 Tanaka et al. 2012), protozoa (Haramoto et al. 2011; Ono et al. 2001, Shrestha et al. 2015; Shrestha 52 et al. 2016), and viruses (Haramoto et al. 2011), in water samples.

53 This study aimed to determine the prevalence of AiV-1 genomes in various types of water 54 samples, including groundwater, in the Kathmandu Valley. The first water sampling was conducted 55 in the wet season (August-September) of 2009 (Haramoto et al. 2011), followed by additional 56 sampling campaigns in both wet (August 2010) and dry seasons (May 2011). During the sampling 57 periods, a total of 53 water samples were collected from 9 shallow dug wells (n = 22), 6 shallow 58 tube wells (n = 15), 8 sites along the Bagmati River and its tributaries (n = 14), a tap in a house 59 supplied with tanker water (n = 1), and a sewage pipe (n = 1).

The water samples, except for the sewage sample, were concentrated to 12 ml by the 60 61 electronegative membrane-vortex method using a mixed cellulose ester membrane filter (pore size 62 of 0.45 µm and diameter of 47 mm; Merck Millipore, Billerica, MA, USA) (Haramoto et al. 2012), 63 as described previously (Haramoto et al. 2011). The volumes filtered were 50 or 100 ml for river 64 water, 3,000 ml for tap water, and 1,000 ml for groundwater, except for one sample that allowed 65 filtration of only 50 ml due to membrane clogging. No concentration step was done for the sewage 66 sample, but a supernatant after centrifugation was used for the subsequent analysis.

67 Viral RNA was extracted using a QIA amp Viral RNA Mini Kit (Qiagen, Hilden, Germany), 68 followed by reverse transcription (RT) using a High Capacity cDNA Reverse Transcription Kit 69 (Thermo Fisher Scientific, Waltham, MA, USA), as described previously (Haramoto et al. 2011). 70 Subsequently, 5-µl aliquot of the resulting cDNA was subjected to TaqMan MGB-based quantitative 71 PCR (qPCR) targeting approximately 101-bp VP1 region sequence of AiV-1 genomes (Kitajima et al. 2013). All water and standard samples (plasmid DNA containing the qPCR target sequence with known gene copy numbers) and negative controls were analyzed in duplicate using a Thermal Cycler Dice Real Time System TP800 (Takara Bio, Otsu, Japan). Prior to RNA extraction, murine norovirus (S7-PP3 strain), kindly provided by Dr. Y. Tohya (Nihon University, Fujisawa, Japan), was added to the sample as a molecular process control, resulting in high extraction-RT-qPCR efficiency values of $78.3 \pm 26.9\%$ (n = 53).

Furthermore, for genotype identification, samples positive for the AiV-1 qPCR were subjected to nested PCR targeting the 3C-3D junction region (Yamashita et al. 2000), followed by agarose gel electrophoresis, direct sequencing using a 3730x1 DNA Analyzer (Thermo Fisher Scientific), and phylogenetic analysis of the 224-bp amplified sequences, excluding primer sequences, using a Genetyx software version 9.1.0 (Genetyx, Tokyo, Japan).

AiV-1 genomes were detected by qPCR in 26 (49%) of the 53 samples tested. As shown in Figure 1, the highest concentration of AiV-1 genomes (4.0×10^9 copies/l) was obtained from the sewage sample. AiV-1 genomes were also detected in all the river water samples, with high concentrations ranging from 1.2×10^6 to 1.4×10^8 copies/l, which were greater than those in influent of wastewater treatment plants in Japan and United States (Kitajima et al. 2013, 2014). This can be explained by the situation where most untreated sewage is discharged directly to the river in the Kathmandu Valley.

90 Results of the detection of AiV-1 genomes in groundwater were quite different between the 91 types of the wells: frequency of AiV-1 genome detection was significantly higher in shallow dug 92 wells (10/22, 45%) than in shallow tube wells (1/15, 7%) (χ^2 -test, P < 0.05). More vulnerable

93 structure of dug wells, which are usually made of brick or stone, than tube wells may have 94 attributed to the difference in detection frequency. Concentrations of AiV-1 genomes in the positive samples ranged from 5.6×10^4 to 2.0×10^6 copies/l, with a geometric mean concentration of 2.3×10^6 95 10^5 copies/l (Fig. 1). AiV-1 genomes were detected in all the three samples collected at each of two 96 shallow dug wells (SG6 and SG37). The concentration of *Escherichia coli* at SG6 (1.1×10^2 – $2.6 \times$ 97 10^2 MPN/100 ml) was within the normal range found in shallow groundwater in the Kathmandu 98 Valley, whereas that at SG37 ($1.4 \times 10^3 - 1.1 \times 10^4$ MPN/100 ml) was relatively higher than average 99 100 (Shrestha et al. 2014).

101 Unlike groundwater samples, AiV-1 genomes were not detected in the tap water sample 102 although it was contaminated with *E. coli* $(1.1 \times 10^2 \text{ MPN}/100 \text{ ml})$. Since tanker water is one of the 103 important drinking water sources in the Kathmandu Valley, further studies need to be conducted to 104 evaluate the prevalence of AiV-1 genomes in taker water samples, as well as their water sources.

Fifteen (58%) of 26 AiV-1 qPCR-positive samples were also positive for AiV-1-specific nested PCR, and subsequent nucleotide sequencing analysis successfully classified all of these samples as genotype B (Fig. 2). According to previous studies conducted so far, genotype A has been identified more frequently in water and environmental samples than genotype B, but the trend was different among the regions studied (Kitajima and Gerba 2015).

Although AiV-1 sequences were found in an untreated sewage sample collected in Kathmandu by viral metagenomics (Ng et al. 2012), there has been no information available on the incidence of AiV-1 in human fecal specimens in the Kathmandu Valley. The results of this study indicate that AiV-1 of genotype B is circulating in human populations in the valley. Further studies are needed to

114	understand more deeply the prevalence and genotype distribution of AiV-1 in the environments.				
115	Nucleotide sequence accession numbers.				
116	The nucleotide sequences determined in this study have been deposited in the GenBank				
117	database under accession numbers LC200515–LC200529.				
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Fig. 1 Concentrations of AiV-1 genomes in water samples

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	AB092828_494/97(SIN)_Japan DQ145760_RN57_France FJ872497_Monastir27467/2003Tunisia_Tunisia EF079160_T-132/02_Thailand AB010145_A846/88_Japan	Genotype A	
	LC200524_2011-RW01 LC200525_2011-RW02 LC200526_2011-RW03 LC200523_2011-SG37 LC200528_2011-RW05 LC200518_2010-RW03 LC200518_2010-RW03 LC200518_2010-RW03		
	LC200517_2010-RW02 LC200521_2010-RW06 LC200522_2010-RW07 LC200515_2010-SG06 LC200520_2010-RW05 LC200516_2010-RW01 LC200519_2010-RW04 LC200529_2011-SW01 -FJ890516 Chshc5 China	Genotype B	
	DQ145759_RN48_France	Genotype C	Devine laskining AD004700 114

Fig. 2 Phylogenetic analysis of AiV-1 sequences identified in water samples. The tree was generated using the neighbor-joining method with 1,000 bootstrap replicates based on the 224-nt 3C-3D junction region sequences. The sequence of bovine kobuvirus (AB084788) was used as an outgroup. The scale bar represents the number of nucleotide substitutions per position. AiV-1 sequences obtained in this study are labeled with bold italics and indicate the GenBank accession number and sample identification.