

1 Applying and advancing the economic resource scarcity potential (ESP) method for 2 rare earth elements

3

4 Abstract

5 A number of studies have identified rare earth elements (REE) as critical metals due to their
6 high economic importance combined with a high risk of supply disruption (Du et al, 2011;
7 Nassar et al, 2015; Schneider et al, 2014). The current methods used to calculate resource
8 depletion in life cycle assessments (LCA) neglect socio-economic, regulatory and
9 geopolitical aspects, nor do they include functionalities such as material recycling or reuse
10 that control the supply of raw materials. These are important factors in determining criticality
11 and are the controlling factors on REE availability rather than geological availability. The
12 economic scarcity potential (ESP) method introduced by Schneider et al. (2014) provides a
13 framework to calculate criticality. This paper reviews the ESP method and advances the
14 method based on recent developments in material criticality. ESP criticality scores for 15
15 REE with the addition of Au, Cu, platinum-group metals (PGM), Fe and Li are measured.
16 The results highlight that Nd and Dy are the most critical REE, owing mainly to the high
17 demand growth forecast for these two elements. A pathway is presented for incorporating
18 these calculated scores into the ReCiPe life cycle impact assessment (LCIA) method of a
19 LCA.

20

21 1. Introduction

22 Life cycle assessment (LCA) is an important tool to quantify the environmental
23 performance of a product or a process such as rare earth element (REE) production. A LCA
24 can detail potential impacts that this process will have on human health, natural environment
25 and natural resources. However there are limitations and problems for assessing abiotic
26 resource depletion during a life cycle impact assessment (LCIA) (Drielsma et al, 2016).
27 Abiotic depletion potential has been used as an indicator, calculating future exhaustion of
28 resources based on current production levels. Advances were made to this approach by Vieira
29 et al. (2016) with the surplus cost potential method, which calculates the increased cost of
30 extracting raw materials due to depleting resources providing a cost per unit of metal
31 extracted in the future. Both methods are useful in understanding the long-term availability of
32 resources but fail to consider a range of factors which control the supply of critical raw
33 materials. In order to correctly assess the criticality of materials, it is necessary to have an

34 indicator that takes into account several impact categories for supply risk and economic
35 importance rather than just resource depletion. Otherwise, the assessment categorizes cerium
36 (which is as abundant in the crust as copper) as highly critical along with dysprosium,
37 praseodymium and the other heavy REE. This paper examines how an alternative method to
38 assess mineral resource inputs can be devised and used for critical metals such as the REE.

39 Rare earth elements include the lanthanides and the chemically similar elements yttrium
40 (Y) and scandium (Sc). The elements are often divided into two groups, the light rare earths
41 elements (LREE) and heavy rare earth elements (HREE). The LREE include La, Ce, Pr, Nd,
42 and Sm. The HREE include the elements from Eu to Lu in the Periodic Table as well as Y.
43 The REE have strategic importance, with uses in a number of emerging low-carbon
44 technologies. Specific physical properties of individual REE are necessary for efficient
45 electric vehicles, and direct drive wind turbines, such as Nd in NdFeB high strength magnets.
46 The addition of Dy is used to maintain the performance of these magnets at high
47 temperatures. Other REE such as La and Ce are used in catalysts for fluid catalytic cracking
48 of crude oil and production of transportation fuels; and Ce and La are used as emissions
49 catalysts in petrol fueled vehicles. Total industrial demand of REE, excluding Y, is small
50 with an estimated use of 159,500 tonnes in 2016 (USGS, 2016), but REE have a large
51 positive economic contribution to downstream industries. One of the major challenges of
52 REE supply is ‘the balance problem’; the misbalance between the economic market demand
53 and the supply of individual REE⁸. There is often high demand for REE that are minor
54 constituents of a REE ore (such as Pr), while the demand for the major constituents (such as
55 La and Ce) may be much lower.

56 The security of supply of REE has been a concern for import-dependent industrialized
57 countries with ambitions to advance their low-carbon economy. China currently dominates
58 the production of REE, excluding Y, accounting for 88% of total REE production in 2016
59 (USGS, 2016). There is a history of supply disruption of REE exports, this has fueled
60 increased attention into the future availability of such elements. From 2007 to 2009 China
61 reduced export quotas of REE by 25% (Binnemans et al, 2015). This resulted in significant
62 price increases following the export restrictions which were put in place by China (Mancheri,
63 2015). Concerns about the future supply of REE and the monopolistic nature of production
64 combined with the growing economic importance of downstream products has led to a
65 number of studies identifying individual REE, or REE as a single group, as critical materials
66 (NRC, 2008; Erdmann and Graedel, 2011; Nassar et al, 2015; BGS, 2015; Moss, 2013
67 Coulomb, 2015; Glöser et al, 2015).

68 A number of projects exist in various stages of development around the world that if
69 moved into production would diversify the supply of REE. For example mining projects are
70 in the prefeasibility or feasibility stage in Europe, with Sweden's Norra Kärr project; in
71 Africa with Malawi's Songwe Hill, Namibia's Lofdal, and South Africa's Zandkopsdrift; in
72 North America with Canada's Ashram, and Nechalacho, USA's Bear Lodge; Australia's
73 Nolans, Dubbo Zirconia project; South America has projects such as Araxá and Serra Verde,
74 both in Brazil. However, there are a number of barriers making production outside China
75 challenging. China currently possesses excess production capacity within the country,
76 suppressing prices and reducing the chances of projects outside China from accessing
77 funding. There is also a lack of proven processing technologies for the unconventional
78 mineralogy in some of the new prospects and a lack of efficient and clean technology for
79 separating and converting rare earth oxides to metals and alloys (USGS, 2018). These factors
80 mean that a large amount of time and capital are required to bring in new operations online
81 and diversify the supply.

82 Downstream uses of REE are often considered to have positive environmental
83 impacts when they are used in generating clean energy or replacing conventional combustion
84 engines in cars (Girardi, 2015). However, the mining, isolation and recovery of REE has a
85 number of environmental and social impacts throughout the life-cycle (Zaimes et al, 2015,
86 Koltun and Tharumarajah, 2014, Arshi et al, 2018, Du and Graedel, 2011, Haque et al, 2014,
87 Sprecher et al 2014).

88 REE production and processing requires a large amount of energy and chemicals, and
89 can produce greenhouse gas emissions, chemical pollutants, hazardous mine waste and
90 wastewater, which can contain radioactive material and can cause extensive land
91 transformation. Chemicals used in the refining process have been involved in REE
92 bioaccumulation and pathological changes in local residents (Li et al, 2013). Contaminants
93 associated with REE production, which include radionuclides and heavy metals, have been
94 identified as having negative impacts on human, plant and livestock health (Rim, 2016).

95 It is important to understand and manage the environmental and social costs
96 associated with REE production as we progress to a low-carbon economy and renewable
97 energy generation, which is likely to require more metal and mineral raw materials per unit
98 energy produced. When considering the sustainability of the raw materials that are produced
99 for the low-carbon economy, it is important to consider risks to supply disruption, which
100 could include market imbalances or governmental interventions such as export bans.

101 The aim of this paper is threefold. (i) To show that individual REE have unique
102 supply risks and economic importance and therefore different levels of criticality. (ii) To
103 provide a more appropriate impact category within LCIA for resource scarcity of critical
104 metals (iii) Explain how criticality can be included in LCA frameworks and see what results
105 would look like.

106

107 **2. Review of REE criticality studies.**

108 A variety of methodologies can be used to determine raw material criticality. The approaches
109 may vary but share a common aim to define the supply risk of a raw material and its relative
110 importance to the economy. The criticality calculation methodology typically contains an
111 evaluation of the level of supply risk and the impact of said supply risk in a two-dimensional
112 matrix (NRC, 2008; Erdmann and Graedel, 2011; Graedel et al, 2015). Environmental
113 impacts can be used to create a third axis (Graedel, 2015).

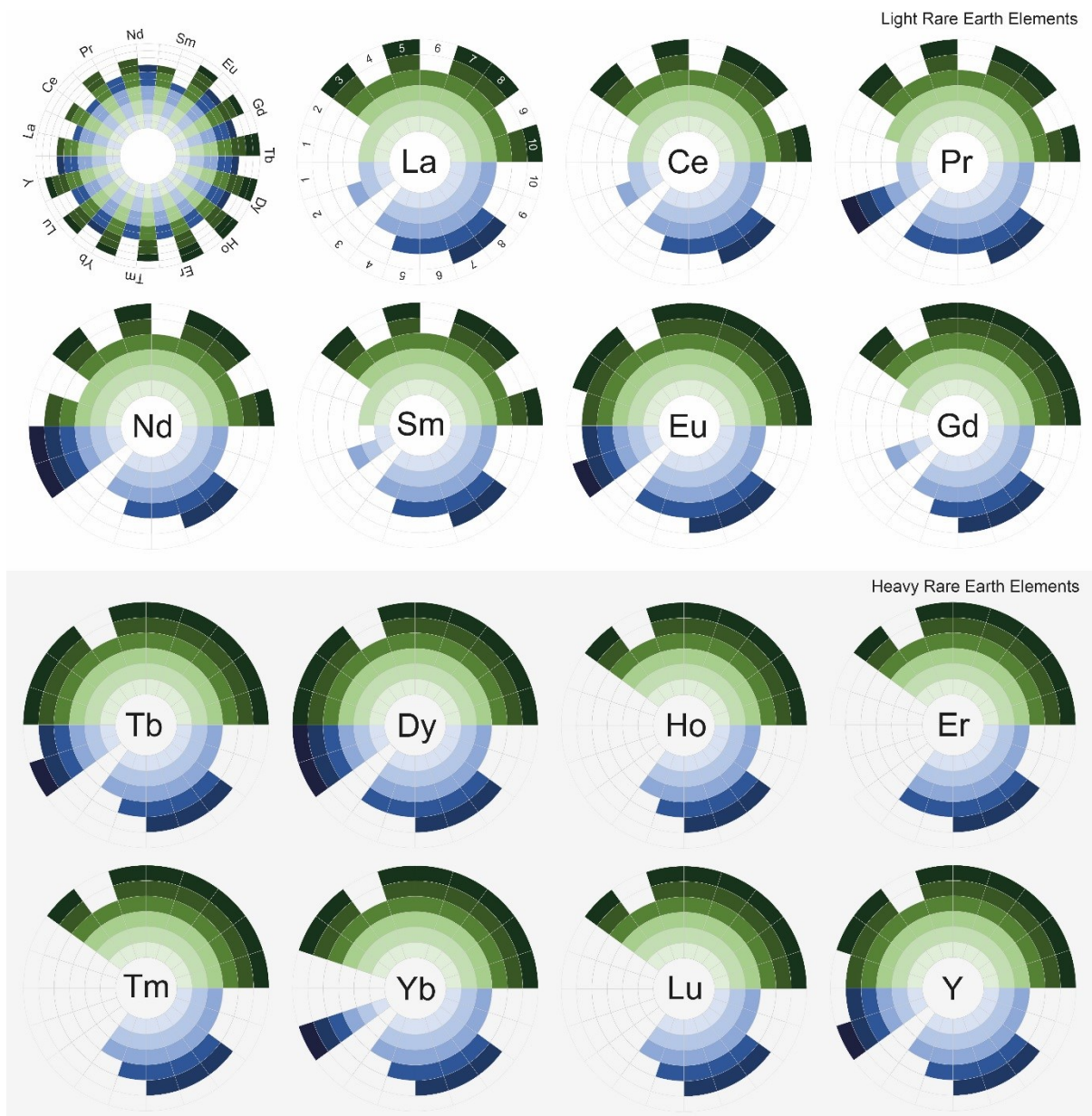
114 Criticality studies are context dependent and can be carried out on a range of scales
115 and for a range of stakeholders, which can be anything from a single company or technology,
116 to a national or multi-national economy (Graedel et al, 2012). For example, a criticality study
117 from the perspective of a country will be different from that of a company, and short-term
118 risk of raw material criticality may not be the same in the medium or long-term. Criticality
119 studies are connected to the concept of risk theory in a holistic way, including economic,
120 societal or environmental risk (Helbig et al, 2016; Frenzel et al, 2017). A wide variety of
121 factors are often considered in criticality assessments, including geological deposits,
122 geographical concentration of deposit or processing facilities, social issues, regulatory
123 structure, geopolitics, environmental issues, recycling potential, substitutability, and
124 sustainability (Achzet and Helbig, 2013; Erdmann and Graedel, 2011).

125 Eight studies that include criticality of REE have been reviewed (Figure 1). Each
126 study had a different context, with various spatial scales, from national to international and
127 looked at different areas of the economy. For example Nassar et al. (2015) looked at the
128 criticality of REE associated with the global economy, whilst Coulomb examined the
129 criticality of REE in the context of the low-carbon economy. Where possible the studies
130 looked at a medium-term time perspective of criticality.

131 All but one study (BGS, 2015) included two-dimensions typical of criticality studies
132 which could be translated into supply risk and economic importance. Figure 1 shows the
133 supply risk of the REE on the left hand side of each box and to the right shows economic

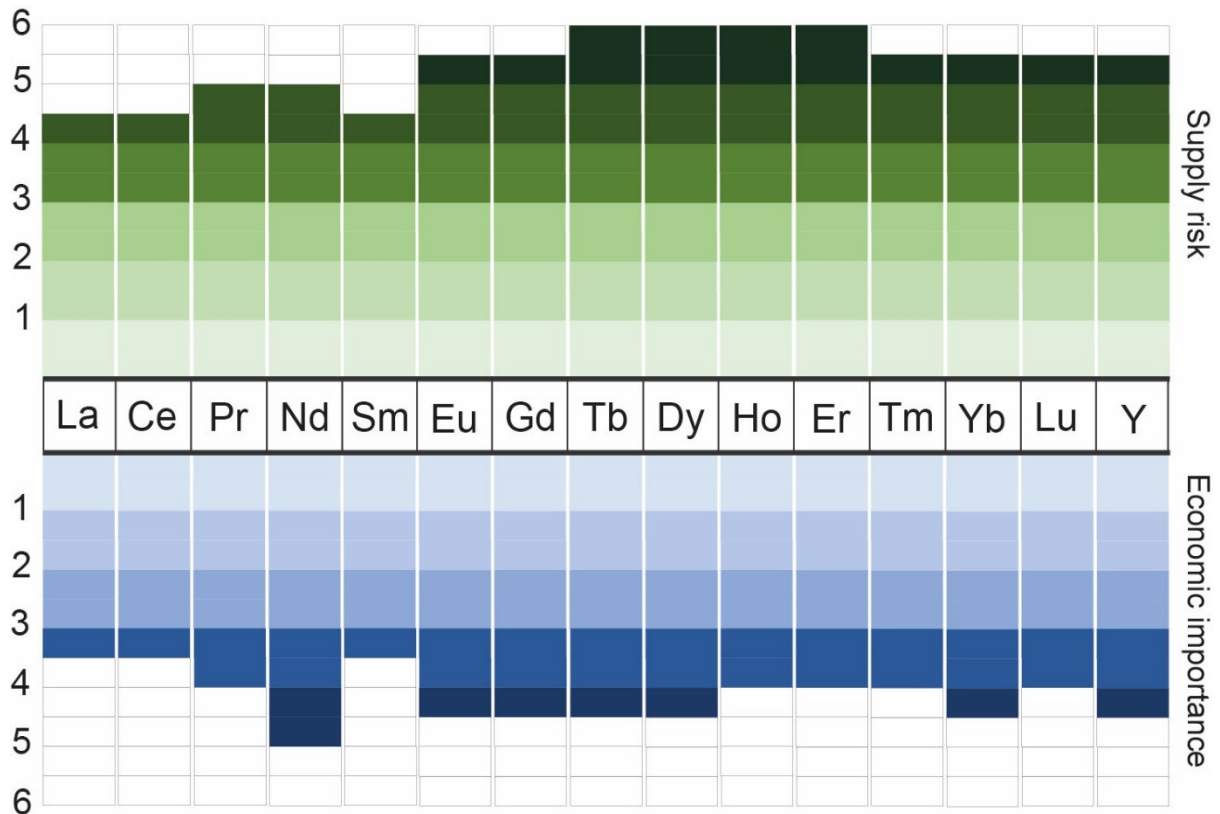
134 importance of the REE from these studies. The relative criticality scores are normalized and
 135 given a colour scale between 1 (non-critical) to 6 (extremely-critical). The terms used in the
 136 study also varied meaning that this approach includes subjective judgement of the criticality
 137 scores. The white categories indicate gaps in the criticality study.

138



139

140 Figure 1. Criticality assessments for individual REE based on supply risk (green top half of each) and
 141 economic importance (blue bottom half) at various scales from national to global in a medium term
 142 time scale. White space means that the REE was not included in the criticality study (NRC, 2008;
 143 Erdmann and Graedel, 2011; Nassar et al, 2015; BGS, 2015; Moss, 2013; Coulomb, 2015;
 144 Glöser et al, 2015).



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146 Figure 2. Normalized average of the combined REE criticality studies from figure 1

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148 **2.1. Life cycle impact indicators for abiotic resource depletion.**

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The concept of the Area of Protection was founded in the early 1990s by the Society of Environmental Toxicology and Chemistry (Fava et al, 1993). It is used in the LCA community to identify classes of endpoint category indicators that society deems important to protect, and allows a linkage between damages because of environmental intervention and societal values. The Area of Protection are divided into the protection of: Human Health, the Natural Environment and Natural Resources (Finnveden, 1997; Udo de Haes et al, 1999). The ILCD handbook defines these natural resources and that challenge as;

“The concern of natural resources is the removal of resources from the environment (and their use) which results in a decrease in the availability of the total resource stock, as non-renewable (usually abiotic) resources are finite”

This definition and the depletion of abiotic resources is a much disputed category within LCA as it crosses the economy-environment system boundary in combination with the

163 fact that there are different ways to define the depletion problem, and there are different ways
164 of calculating these depletion definitions (Van Oers and Guinée, 2016).

165 For example Van Oers (2016) stated that the environmental impact of LCA should not
166 strive to take into account the different aspects of a criticality assessment due to the varying
167 temporal and spatial nature of each study. However this can be overcome with a clear
168 definition during the goal and scope phase of a LCA and matching the criticality calculation
169 to what is being measured. For example if the environmental performance of a mining project
170 is being measured, it is possible to complete the criticality calculation for the life of the
171 mining project with criticality scores in a global context.

172 Different approaches can be used to determine the decreasing availability of
173 resources. Different approaches have distinct visions or cultural perspectives for abiotic
174 resource depletion (De Schryver et al, 2018). The cultural perspective theory which has
175 categorised visions on resource depletion as either individualist, hierarchist and egalitarian is
176 explored is incorporated into different LCIA methodologies.

177 One approach to resource depletion which aims to remove the cultural perspective
178 from the process is through the use of entropy or exergy as a basis for characterization, which
179 considers the efficiency of extraction. A thermodynamic approach which can capture
180 resources is a useful approach as it has an established scientific basis. Exergy is a measure
181 of available energy, whilst entropy in this context refers to the dispersal of energy within a
182 system.

183 A common method that has been used and is considered individualist uses resource
184 scarcity for the basis of characterization. This method calculates the long-term depletion of
185 non-renewable resources. The depletion of resources is calculated and considers future
186 resource scarcity as a result of current consumption. The impact from resource use is then
187 calculated as an impact on human welfare due to reduced availability, increased competition,
188 and limited accessibility driven by social and geopolitical factors (Finnveden, 2005;
189 Sonnemann et al, 2015). These approaches have shortcomings. Firstly, calculations of
190 physical resource availability or 'depletion potential' used in LCIA rely on a fixed stock
191 paradigm, as described by Tilton (2002). The idea that there is a finite quantity of a resource,
192 often described as a crustal abundance, fails to calculate the reuse or recycling rate of these
193 materials and considers that materials are lost after use. There is also no clear definition for
194 undiscovered resources (Vieira et al, 2016). The alternative method used is the opportunity
195 cost paradigm, which states that if physical quantities reduce, or are more difficult to access,
196 prices will increase and innovations and alternatives to that material will be sought, reducing

197 demand. LCIA practitioners have used both methods which have very different views on
198 natural resources and can significantly alter LCIA results. In the fixed stock method, any use
199 of natural resources results in reduced availability for the future, whereas in the opportunity
200 cost view, natural resources are viewed as flows that need to be managed to meet human
201 demands (Drielsma et al, 2016).

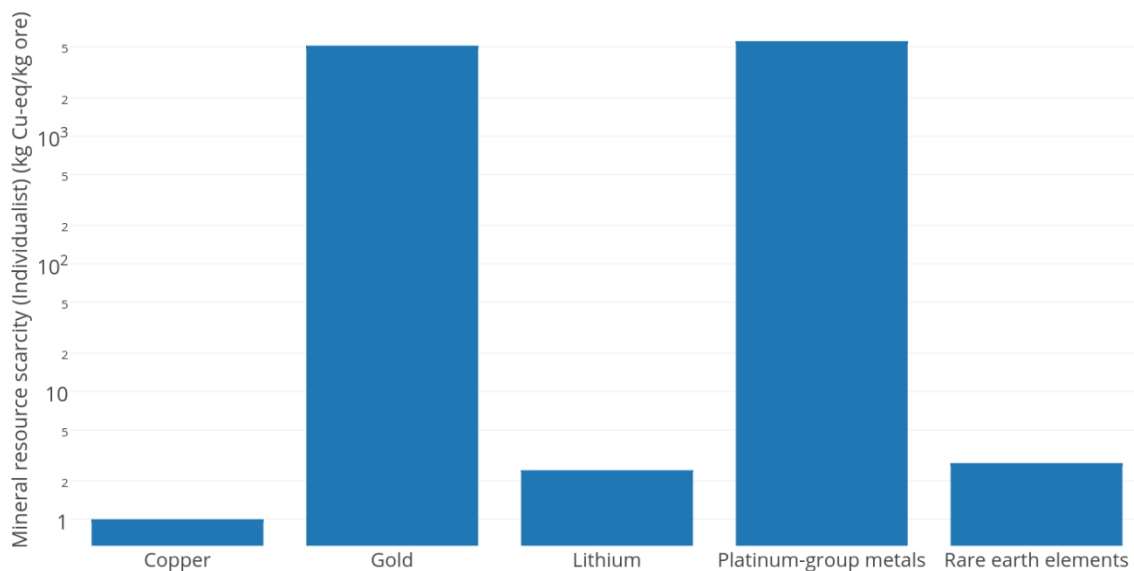
202 Different methods have different visions and methodologies. Many of these methods
203 that are currently employed to not consider the socio-economic, regulatory and geopolitical
204 aspects or functionalities such as material recycling or reuse.

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206 3. Materials and Methods

207 The abiotic depletion potential method (Van Oers and Guinée, 2016) and the surplus
208 cost potential method (Vieira et al, 2016) are used for comparison in this paper. The latter has
209 been integrated into the ReCiPe methodology (Huijbregts et al, 2016). This method to
210 calculate metal depletion provides scores for 75 mineral resources providing impact scores in
211 relation to 1kg of Cu. Figure 2 provides a comparison of five mineral resources and
212 categorizes rare earth elements as a single group.

213



214

215 Figure 2. Mineral resource scarcity results (individualist) using the surplus cost potential approach

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217 LCIA is a step in a LCA which translates data such as emissions or resource uses
218 from LCA studies to an easily understandable smaller number of impact scores. The method

219 of calculating these scores is referred to as characterization, and the results will produce an
220 environmental impact per unit of stressor (e.g. per kg of resource). Schneider et al. (2014)
221 identified that economic aspects of resource supply are neglected in current LCA
222 methodologies and attempted to overcome this by introducing the economic resource scarcity
223 potential (ESP) model.

224 Various data that contribute to scarcity of resources are included, expanding the Area
225 of Protection for natural resources to include economic or socially derived scarcity. The
226 factors that are included in ESP include reserves, recycling, and country and company
227 concentration of mining activities, economic stability, demand growth, trade barriers, and
228 companion metal fraction. Drielsma et al. (2016) highlighted that this method assesses short
229 term availability of resources, and is a useful tool in identifying disruptions that may arise in
230 this timeframe. Drielsma et al. (2016) also argued that the Area of Protection for natural
231 resources is altered using this method as the ESP method aims to protect the product system
232 being measured rather than the resources themselves. For example, the protection of the value
233 that a resource has when being used rather than the resource itself.

234 Current LCIA methods, such as the ReCiPe approach only take into account geological
235 availability and the increased cost of accessing raw materials in the future. The surplus cost
236 potential method fails to take into account resource criticality. Additional methods, such as
237 the ESP approach, would be a useful step to incorporate criticality factors into the life cycle
238 sustainability assessment framework which would better represent impacts on the Area of
239 Protection for Natural Resources (Sonnemann et al, 2015). The ESP method put forward by
240 Schneider (2014) allows for a new characterization factor for resource use impact assessment.
241 Using these characterization factors and a framework to incorporate criticality into the life
242 cycle sustainability assessment context by Sonnemann et al. (2015) allows for integration of
243 the ESP method into the LCA.

244

245 **3.1. Methodology of ESP Calculations.**

246 The factors that impact resource availability were suggested by Schneider et al.
247 (2014) and have been highlighted in table 1. Equal weighting was used for all impact
248 categories initially replicating the method used by Schneider (2014). This was followed by a
249 comparison of results if the economic importance impact category was increased to represent
250 50 percent of the total ESP score. Production data were obtained by combining the USGS
251 data with other project scale information. Individual REE data were obtained from individual

252 companies, and when not possible were estimated from literature. All sources of information
 253 and origins of data used in the study are included in the supplementary information.

254

255 Table 1. Overview of impact categories, indicators and thresholds used in the ESP calculations
 256 (Thresholds are based and on data from Schneider et al (2014) DOJ and FDT (2010), The World Bank
 257 Group (2012),UNDP (2011), Rosenau-Tornow et al. (2009)

Impact category	Category indicators	Threshold
Supply risk		
Reserve availability	Reserve/Annual production	Low<0.4<high
Recycling	New material content (%)	Low<0.5<high
Mining country concentration reserves	HHI index	Low<0.15<high
Production bottleneck (country concentration)	HHI index	Low<0.15<high
Production bottleneck (company concentration)	HHI index	Low<0.15<high
Governance stability	WGI ¹	Low<0.25<high
Socioeconomic stability	HDI ²	Low<0.12<high
Trade barriers mine production	Share of mine production under trade barriers (%)	Low<0.25<high
Companion metal fraction	Production as companion metal (%)	Low<0.2<high
Trade		
Economic importance		
Average production and cost per kg	\$ per kg	Low<0.1<high

258

259 The data incorporate 10 impact categories and can be aggregated to provide a single ESP
 260 value (Equation 2). Each category has been described in a glossary in the supplementary
 261 information. This allows for the comparison of the 15 REE studied as well as providing a
 262 comparison with Au, Cu, PGM, Fe and Li. Other elements were selected because they offered
 263 a range of supply risk and economic importance scores in previous criticality studies. They
 264 are used for comparison with the REE and to give a context to how REE perform. The
 265 criticality in the context of this paper is within a “global economy” and so not specific to a
 266 particular technology or group. This also allows for integration within the ReCiPe LCIA as
 267 this is on a global scale. It should be noted that it is possible to adjust the context through
 268 weighting the results or changing the thresholds. Thresholds used in this study are shown in
 269 Table 1 with justification for their values.

270 The aggregation of the supply risk and economic importance impact factors is given equal
 271 weighting. Individual category indicator results (impact factor x LCI) give an indication for

272 the magnitude of the risk. However, the results only provide a comparison of the resources
273 studied. A greater number of resources used for this method will allow for a more
274 comprehensive estimation of supply risk and provide a better basis for decision making.

275 As noted by Schneider (2014), to produce a supply risk perspective for the resource
276 availability requires each category indicator to be placed in relation to a target. This method
277 is described in detail by the distance-to-target method by Frischknecht et al. (2008). The
278 resulting impact factors (I) provide a threshold, above which high risk of supply disruption is
279 expected. This was calculated for comparison for the 15 REE together with gold, copper,
280 platinum group metals (PGM), iron ore and lithium (i) and each impact category (j). The
281 ratio of current to critical flows is squared allowing large impact values (above the target
282 value) to be weighted above proportional (Frischknecht and Büsler Knöpfel, 2013; Drielsma
283 et al, 2014). The indicators are scaled from 0-1, with order being inverted when necessary to
284 ensure high score corresponds to high risk. All values below the value of “1” are deemed
285 uncritical and have no impacting score.

286

$$287 \quad I_{i,j} = \text{Max} \left\{ \left(\frac{\text{indicator value}_{i,j}}{\text{threshold}_{i,j}} \right)^2 ; 1 \right\}$$

288 Equation 1.

$$289 \quad ESP_i = \prod_j (I_{i,j})$$

290 Equation 2.

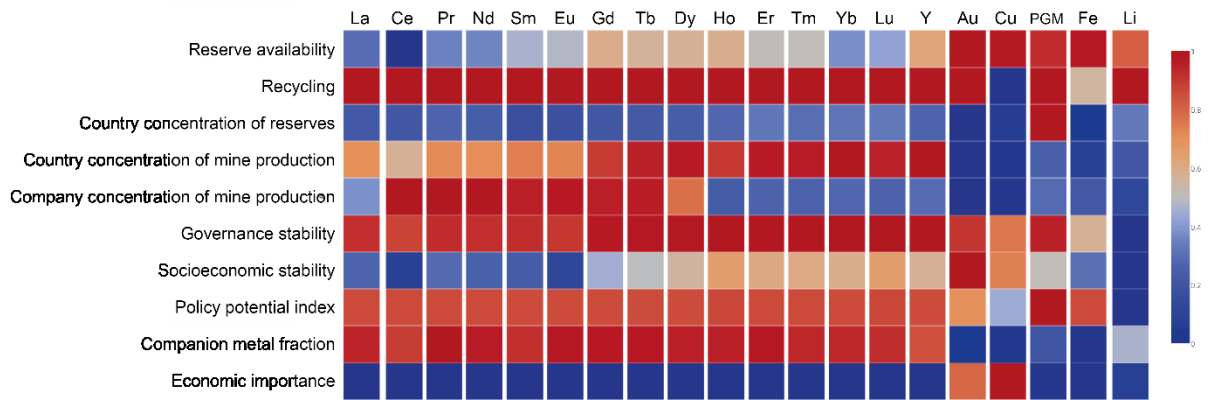
291

292 The resulting economic scarcity potential score for each element which includes the
293 impact categories from both supply risk and economic importance is a dimensionless quantity
294 determined by the ratio of the current indicator value to the determined threshold linked to
295 the LCI.

296

297 4. Results

298 The performance of individual REE compared to Au, Cu, PGM, Fe and Li has been
 299 calculated and highlighted in Figure 3.



300

301 Figure 3. Individual impact category scores for 10 categories. Data based on (Buijs et al, 2012; NRC,
 302 2016; Graedel et al, 2015; Nassar et al, 2015; Angerer et al, 2009).

303

304 *4.1. Reserve availability*

305 The 15 REE included in the study had a lower score for reserve availability than Au,
 306 Cu, PGM, Fe and Li. These other metals had higher impact scores because of their high level
 307 of production relative to REE; being produced in thousands or millions of tonnes per annum
 308 compared to REE which have a total production of the 126,000 tonnes in 2016. This,
 309 combined with the large reserves of REE, calculated as 120,000,000 (USGS, 2016) t based on
 310 their continued availability and typical metallurgical recoveries means the reserve availability
 311 of REE is higher than the other metals in the study leading to a low impact score. Of the
 312 REE, Y, Gd, Tb, Dy and Ho had the highest impact score whilst Ce and La had the lowest.
 313 These results can be explained by the fact that HREE are less abundant in the earth's crust
 314 and also less abundant in REE deposits, whilst consumption of some of these elements
 315 remains relatively high, such as Dy and Tb in permanent magnets. Er, Tm, Yb and Lu are not
 316 abundant in deposits but are exploited at very low rates leading to a moderate impact score.

317

318 *4.2. Recycling*

319 More work needs to be carried out to quantify the rate of recycling of different REE
 320 because the published data used for the calculations in this study does not represent the
 321 quantity of recycled material reentering the system.

322

323 *4.3. Country concentration of reserves*

324 The country concentration of reserves impact score was high for PGM compared to
325 the other raw material in this study. This is because of the dominance of South Africa in
326 holding the reserves of PGM. In contrast reserves of Au, Cu and Fe appear the most
327 widespread as they have the lowest score in this category. The REE had moderate scores in
328 this area with slightly increasing impact scores of the HREE because of the dominance of
329 China in holding much of the HREE reserves. The country concentration of reserves
330 indicated that although the reserves of rare earths are relatively widespread, there is a high
331 concentration of Sm and Eu in China, whilst Ho, Er, Tm, Yb and Lu in reserves is more
332 geographically widespread.

333

334 *4.4. Country concentration of production*

335 The impact score for the country concentration of production was high for all REE
336 compared to Au, Cu, PGM, Fe and Li owing to the dominance of REE production from
337 China. The HREE had the highest impact score for this section. Li was highest scoring in this
338 category for the non-REE.

339

340 *4.5. Company concentration of mine production*

341 The company concentration of mine production impact category displays the
342 dominance of Northern Rare Earth (Group) High-Tech Co., Ltd, China even when put in in
343 the context of other raw materials, with Ce, Pr, Nd, Sm, Eu, Gd, Tb having the highest scores
344 for this section. The lower impact score for the LREE can be explained by production from
345 Lovozerskiy GOK in Russia, Mount Weld in Australia and mineral sands in India, which are
346 all LREE-enriched deposits.

347

348 *4.6. Governance stability*

349 The impact scores were high for the REE, with highest scores being seen with the
350 HREE that are produced almost exclusively in China. Li had a low impact score in this
351 category is explained by its production in Australia and Chile. PGM and Au had high scores
352 in these categories highlighting that there are risks associated with the stability of
353 governments in regions where these materials are mined.

354

355 *4.7. Socioeconomic stability*

356 Au was the highest scoring element, followed by Cu and then the Ce, Pr, Nd, Sm, Eu,
357 Gd, Tb. The low socioeconomic stability of the countries producing Au are highlighted as
358 well as the moderate socioeconomic score of China. For REE the lowest impact scores were
359 Ce and Eu. This is owing to the combination of elevated levels of production of these
360 elements from Mt Weld, Australia and Australia's higher performance in government
361 stability and socioeconomic stability.

362

363 *4.8. Policy potential index*

364 The 15 REE studied had a high score for the policy potential index. However it is
365 PGM that had the highest score in this category, whilst Fe had a similar score to the REE.
366 The policy potential index impact score was the highest for Tb, whilst Ho had the lowest
367 score. Many of the REE received moderate scores in this impact category indicating that
368 there was only a small amount of variation in the impact scores for the REE.

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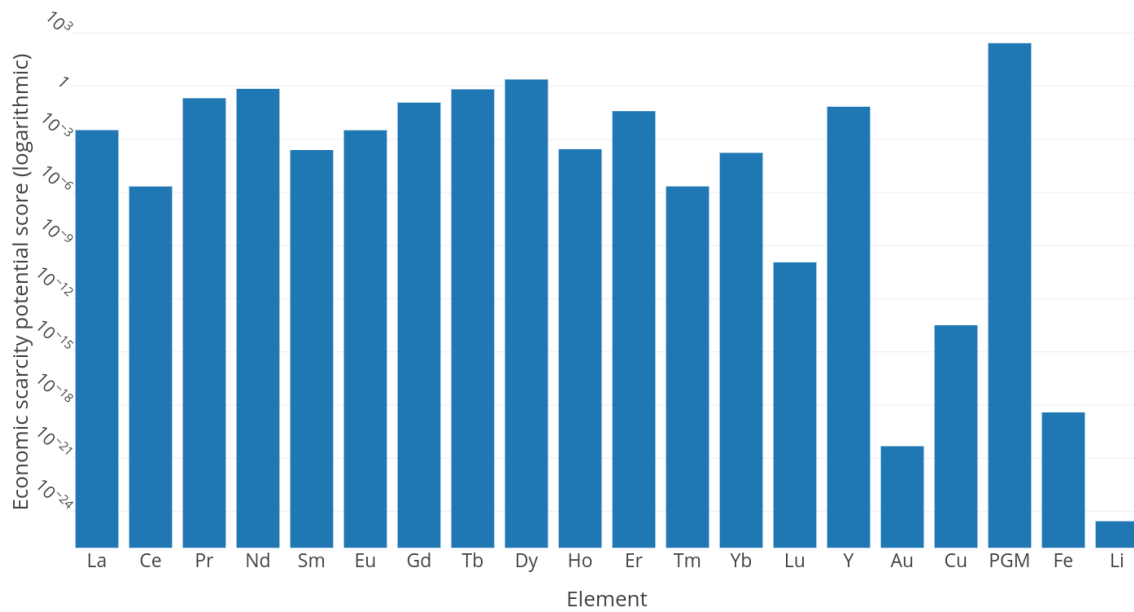
370 *4.9. Companion metal fraction*

371 REE have a high risk associated with the fact that they are commonly exploited as a
372 by-product of each other and of other raw materials (such as iron ore at Bayan Obo, China)
373 among others. The other raw materials used in comparison had low impact scores in this
374 category indicating that they are commonly extracted as the main component at a mine. The
375 companion metal fraction impact scores were relatively similar to each other. Pr had the
376 highest score whilst Y had the lowest.

377

378 *4.10. Economic importance*

379 In the economic importance category the REE have a low score. This category is
380 dominated by Cu and to a lesser extent Au. These are the two raw materials that have the
381 greatest economic importance during the raw material extraction phase. Of the REE, Nd had
382 a markedly higher economic importance impact score than the other REE. This is owing to
383 the use of Nd in NdFeB magnets, which are predicted to drive demand growth until 2022
384 (Roskill, 2016). Dy and Pr were calculated as having the next highest economic importance
385 scores. All other REE have low economic importance scores.



386

387 Figure 4. Individual economic scarcity potential scores for 10 categories, each of which has equal
 388 weighting

389

390 4.2. Overall ESP

391 The final ESP results are presented on a logarithmic scale to better display the relative
 392 performance of individual elements. The ESP scores displayed in Figure 4 show how the
 393 REE compared to Au, Cu, PGM, Fe and Li. Giving equal weighting to each category and
 394 using the methodology described above resulted in PGM having the highest ESP score and so
 395 these elements are considered the most critical in this context. The factors driving the PGM
 396 score up are the high policy potential index score, the high governance stability score as well
 397 as a high country concentration of reserves. Dy scores second highest for ESP. It is
 398 interesting to note that as a greater number of raw materials are included in the study, the
 399 relative performance of elements can change, as in this case where Dy has overtaken Nd in
 400 terms of relative ESP score. This is because the economic importance was an important factor
 401 in driving Nd's ESP score up in the REE comparison, but as more raw materials are added
 402 with a greater economic importance, this distinction becomes less important. Nd is the next
 403 highest scoring element, followed by Tb, Pr, Gd and Y. Au, Cu, Fe and Li all have lower ESP
 404 scores than the REE.

405 The economic scarcity potential approach used in this study provides results that
 406 greater reflect the reality of resource availability until 2021 when compared to the abiotic
 407 depletion potential or surplus cost potential approach, which are more suited to understanding

408 the long-term availability of resources. It considers socio-economic, regulatory and
409 geopolitical aspects or functionalities such as material recycling or reuse in the calculations
410 rather than geological availability. This is an area that is currently missing in the LCA
411 approach but has an impact on low-carbon technology development and proliferation. Nd and
412 Dy are the highest scoring REE using this approach, highlighting the need to broaden the
413 supply chain for these two elements in particular, whilst Ce has a low economic scarcity
414 potential score and is overproduced. New uses of Ce, which is cheap because of the
415 oversupply, would help to even up requirement for REE and help supply of Nd and Dy.

416 A simplified calculation was used for economic importance, looking only at demand
417 growth, production volume and value of material produced. Improvements could be made to
418 this calculation. A novel empirical approach has been presented by Mayer and Gleich (2015)
419 which looks at risk associated with future price increases of raw materials. The approach
420 which uses a compounding framework to calculate net present values and volatility is a
421 potential avenue to include under these calculations which may provide more realistic
422 economic importance impact scores.

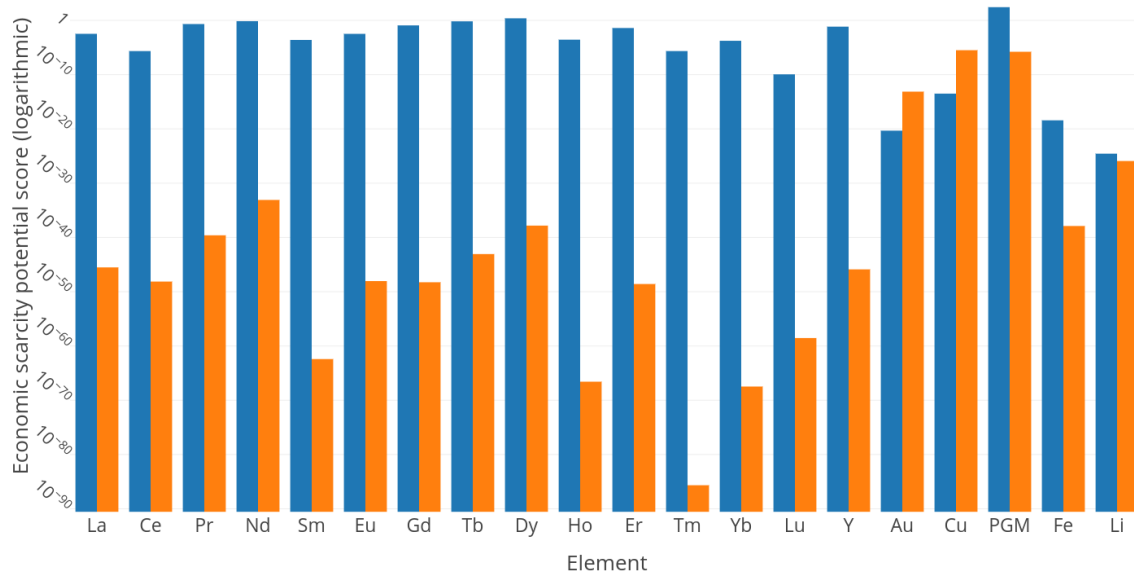
423 The method used in this study only looks at the impact categories associated with the
424 mining and dissolution phase and fails to consider the larger production chain of final
425 products which can be in a number of forms such as rare earth oxides, misch-metals or
426 separated metals and transport. Future work could look at the different processing stages and
427 see how this would alter the economic scarcity potential scores for different elements. Recent
428 work has examined the role of primary processing (first post-mining stage) in the supply risk
429 of critical metals (Nansai et al, 2017). Understanding the role of different processing stages in
430 raw material availability is an important area of research, especially for REE production
431 which has a long and complex production chain. Future work should cover all elements from
432 the periodic table using the economic scarcity potential approach to calculate scores for the
433 global economy for the short to medium term. Using improved economic importance
434 calculations would make the approach a useful addition to the LCIA results. Annual updates
435 on production would allow the method to be up to date and have practical use.

436

437 *4.3. Adjusting the weighting of economic importance*

438 Criticality studies are context dependent. The ESP results above use an equal
439 weighting for each impact category. However, it is possible to adjust the level of an impact

440 category or categories to represent a different context. Figure 5 shows this with the blue bars
 441 indicating the results of the ESP scores with equal weighting for the impact factors. The
 442 orange bars calculate the ESP score by giving all the supply risk impact categories (reserve
 443 availability, recycling, country concentration of reserves, country concentration of mine
 444 production, company concentration of mine production, governance stability, socioeconomic
 445 stability, trade barriers to mine production, companion metal fraction) equal weighting and
 446 giving the economic importance impact category the same weighting as the combined supply
 447 risk impact categories.



448

449 Figure 5. Economic scarcity potential scores for calculated using 10 categories for each
 450 individual element. Blue bars are ESP scores with equal weighting for the impact factors. The orange
 451 bars calculate the ESP score by giving all the supply risk impact categories (reserve availability,
 452 recycling, country concentration of reserves, country concentration of mine production, company
 453 concentration of mine production, governance stability, socioeconomic stability, trade barriers to mine
 454 production, companion metal fraction) equal weighting and giving the economic importance impact
 455 category the same weighting as the combined supply risk impact categories.

456 The results indicate an increased ESP score for Au and Cu, which is the highest
 457 scoring element in this context, because of their high economic importance score. A small
 458 reduction in the ESP score for PGM, which is the second highest scoring element, and Li
 459 which has a small reduction in ESP score. Fe has a large decrease. The REE have a
 460 substantial decrease in their ESP score owing to their relatively low economic importance

461 using the simple calculation in this study when compared to the other elements. Nd is highest
462 scoring of the REE, followed by Dy owing to their relative high economic importance
463 compared to other REE.

464 Increasing the weighting of economic importance (Figure 5) highlights the flexibility
465 of criticality studies. For example, giving equal weighting Cu was considered one of the
466 lowest scoring elements in comparison, but when economic importance was increased to 50%
467 of the total ESP score it became the highest scoring element in the study. Criticality studies
468 can be used to compare the relative levels of criticality of raw materials in different scenarios,
469 but these need to be clearly defined. This study used a global spatial scale for the whole
470 economy and used a medium term time scale, but it is possible to adjust the criteria for a
471 number of scenarios. The weighting of the impact categories will be different depending on
472 the context of the study. For example a study of the criticality of raw materials for the low-
473 carbon economy, would give a higher economic importance to the raw materials used in the
474 relevant technologies than has been given in this study. A valuable area of research would be
475 to develop understanding of appropriate weighting for the impact categories under different
476 scenarios. Understanding the importance of different processes of raw material availability
477 would be a useful step in developing a robust method and would be important in its
478 successful integration into the LCA approach.

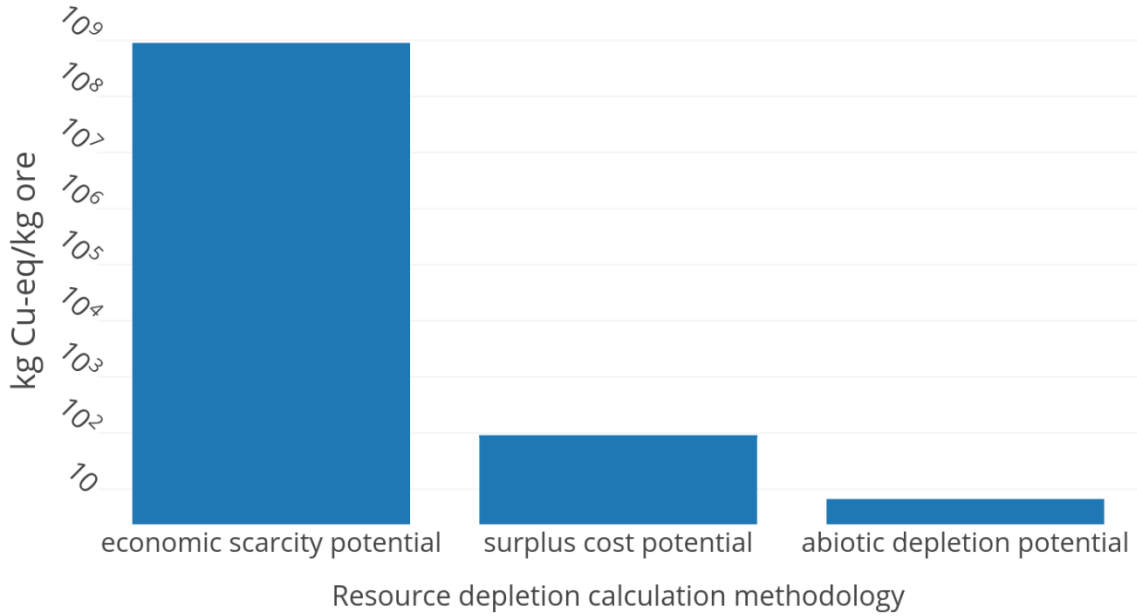
479 4.4. Integration into LCA

480 The scores of the individual elements will be calculated against the reference element
481 of copper. Figure 6 provides a simulation of resource depletion results using three different
482 calculation methodologies (economic scarcity potential, surplus cost potential, abiotic
483 depletion potential) with the example using a 1 kg NdFeB magnet. Simplified inventory data
484 were used (Jin et al, 2016), and is shown in table 2. A comparison of results is highlighted
485 results using the abiotic depletion potential approach, the surplus cost potential approach and
486 the economic scarcity potential approach.

487 Table 2. Composition of virgin NdFeB magnet (Jin et al, 2016).

Element	Weight %
Fe	66.88
Nd	18.0

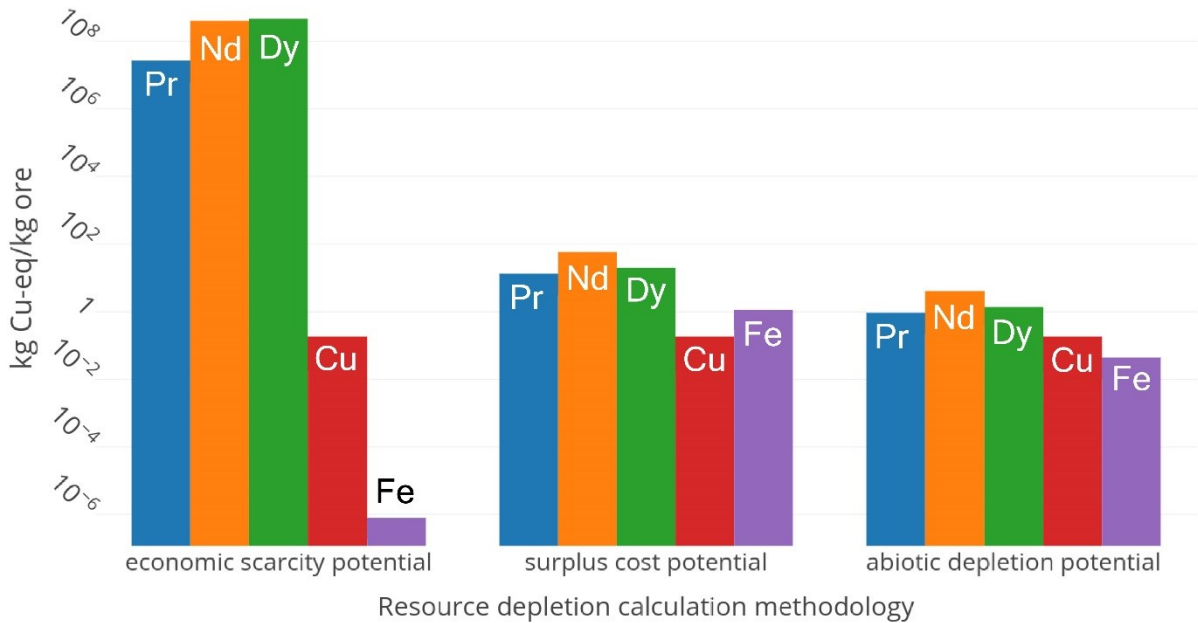
Dy	6.15
Pr	4.6
Cu	0.18



488

489 Figure 6. Comparison of resource depletion calculation methodology on the results for the
490 components of NdFeB magnet.

491



492

493 Figure 7. Elemental contribution to resource depletion calculation scores for economic

494 scarcity potential, surplus cost potential, and abiotic depletion potential for components of
495 NdFeB magnet.

496 The results show that there is an increased score (kg Cu-eq/kg ore for the economic
497 scarcity potential calculation method. This is because the REE components, Pr, Nd and Dy
498 have a high economic scarcity potential score as elements. Cu is the reference value for all
499 methods which explains the equal score with each method. Fe has a lower score using the
500 economic scarcity potential approach as it has been calculate to have low criticality. Figure 7
501 highlights how the economic scarcity potential approach places greater emphasis on elements
502 that have higher criticality scores and are more susceptible to supply disruption in the short to
503 medium term. This information could prove useful in comparative LCA when examining the
504 environmental performance of a product and process and provides an additional metric for
505 which to compare. Such a scenario could exist when comparing the environmental
506 performance of two mining operations. Results for environmental performance could be
507 included alongside criticality data for a better comparison.

508 **5. Conclusions**

509 The ESP approach is particularly useful when trying to understand the availability of
510 critical metals. This is important as they play a key role as raw materials for the low-carbon
511 economy. This is important as they play a key role as raw materials for the low-carbon
512 economy. This paper aimed to compare the performance of individual REE and put it in
513 context with other raw materials. The results indicate that REE need to be considered as
514 distinct elements with different criticality associated with each of them. For example Dy and
515 Nd had the highest economic scarcity potential scores, whilst Lu and Ce had the lowest of the
516 REE. One of the reasons for Ce having a low score is its overproduction. The excess
517 availability and low criticality means that companies have an opportunity to find new uses for
518 Ce. For example the Critical Materials Institute have developed aluminum-cerium alloys
519 (Sims et al., 2016). The high scores for Nd and Dy are due to the increase in demand of
520 NdFeB magnets in hybrid and electric vehicles until 2026 (Goodenough, 2017). Whilst
521 projections for Sm, Tm and Lu suggest that growth and production volume will remain low,
522 keeping the economic importance of these elements low. All REE have higher economic
523 scarcity potential scores than Au, Cu, Fe and Li, whilst PGM had the highest score of all the
524 elements included in the study. The high score for PGM was due to its concentration of
525 reserves and production in South Africa, which has a low score in the governance stability

526 and policy potential index. Although further work needs to be done and more elements need
527 to be included in the method before its integration into LCIA results, this study provides a
528 guideline for the approach.

529 A major challenge for this approach, as with all raw material studies is the availability of
530 data. An inconsistent amount of data are available for the calculations of the economic
531 scarcity potential impact categories. There is a lack of reliable production data for the REE,
532 and this would also be the case for other raw materials. USGS and BGS are useful sources of
533 data, and they are clear about the uncertainty of some production data. For example the high
534 level of illegal mining in REE in China has been ignored (Rao, 2016).

535 The development of economic and supply risk indicators that can fit alongside or within
536 LCA should be further explored and methods such as the approach shown here can be
537 considered complimentary to other resource depletion methods currently employed.

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544

545 7. References

546

- 547 1. Achzet, B.; Helbig, C. How to evaluate raw material supply risks - an overview.
548 *Resource. Policy*, **2013**, 38, 435-447.
- 549 2. Angerer G, Erdmann L, Marscheider-Weidemann F, ScharpM, Lüllmann A, Handke V,
550 Marwerde M (2009a) Rohstoffe für Zukunftstechnologien. ISI-Schriftenreihe
551 "Innovationspotenziale". Fraunhofer IRB Verlag, Stuttgart
- 552 3. *BGS World Mineral Statistics*, British Geological Survey, **2015**,
553 <http://www.bgs.ac.uk/mineralsuk/statistics/risklist.html>
- 554 4. Binnemans, K.; Jones, P. T. Rare earths and the balance problem, *Journal of Sustainable*
555 *Metallurgy*, **2015**, 1 (1) 29-38
- 556 5. Buijs, B.; Sievers, H.; Tercero Espinoza, L.A., Limits to the critical raw materials
557 approach. In: *Proceedings of the ICE e Waste and Resource Management*, **2012**, 201-208.

- 558 6. Coulomb, R.; Dietz, S.; Godunova, M.; Bliggard Nielson, T. Critical minerals today and
559 in 2030: an analysis of OECD countries, *ESRC Centre for Climate Change Economics*
560 *and Policy, Grantham Research Institute on Climate Change and the Environment*, **2015**
- 561 7. Drielsma J.; Russell-Vaccari A.; Drnek T.; Brady T.; Weihed P.; Mistry M.; Perez
562 Simbor L Mineral resources in life cycle impact assessment – defining the path forward,
563 *Int. J. Life Cycle Assess.* **2016**, 21:85-105
- 564 8. Du, X.; Graedel, T. E. Global In-Use Stocks of the Rare Earth Elements : A First
565 Estimate. *Environ. Sci. Technol.* **2011**, 4096–4101.
- 566 9. Erdmann, L.; Graedel, T. E. Criticality of non-fuel minerals: a review of major
567 approaches and analyses. *Environ Sci Technol*, **2011**, 45(18): 7620–7630
- 568 10. Fava, J. A.; Consoh, F.; Demson, R.; Dickson, K.; Mohin, T.; Vigon, B. A conceptual
569 framework for life cycle impact assessment. *Society for Environmental Toxicology and*
570 *Chemistry*, **1993**, Pensacola
- 571 11. Finnveden, G., Valuation methods within LCA - where are the values? *Int. J. Life Cycle*
572 *Assess.* **1997**, 2, 163-169.
- 573 12. Finnveden, G. The resource debate needs to continue. *Int J Life Cycle Assess*, **2005**,
574 10(5):372
- 575 13. Frenzel, J.; Kullik, J.; Reuter, M. A.; Gutzmer, J. Raw material ‘criticality’ – sense or
576 nonsense? *Journal of Physics D: Applied Physics*, **2017**, 50 123002
- 577 14. Frischknecht, R.; Büsser Knöpfel, S. Swiss Eco-Factors 2013 according to the Ecological
578 Scarcity Method. *Methodological fundamentals and their application in Switzerland.*
579 *Environmental studies no. 1330.* **2013** Federal Office for the Environment, Bern, 254 pp
- 580 15. Glöser, S.; Tercero Espinoza, L.; Grandenberger, C. Faulstich, M. Raw material criticality
581 in the context of classical risk assessment, *Resour. Policy*, **2015**, 44 35–46
- 582 16. Goodenough, K. M.; Wall, F.; Merriman, D. The Rare Earth Elements: Demand, Global
583 Resources, and Challenges for Resourcing Future Generations, *Natural Resources*
584 *Research*, **2017**, 1-16 <https://doi.org/10.1007/s11053-017-9336-5>
- 585 17. Graedel, T. E.; Harper, E. M.; Nassar, N. T.; Nuss, P.; Reck, B, K. Criticality of the
586 metals and metalloids, *PNAS*, **2015**, 112 (14) 4257-4262, doi: 10.1073/pnas.1500415112
- 587 18. Graedel, T. E.; Barr, R.; Chandler, C.; Chase, T.; Choi, T.; Christofferson, L.;
588 Friedlander, E.; Henly, C.; Jun, C.; Nassar, N. T.; Schechner, D.; Warne, S.; Yang, M.;
589 Zhu, C. Methodology of metal criticality determination. *Environ Sci Technol* **2012**
590 46:1063–1070
- 591 19. Helbig, C.; Wietschel, L.; Thorenz, A.; Tuma, A. How to evaluate raw material
592 vulnerability—an overview, *Resour. Policy*, **2016**, 48 13–24
- 593 20. Huijbregts, M. A. J.; Steinmann, Z. J. N.; Elshout, P.M. F.; Stam, G.; Verones, F.; Vieira,
594 M. D. M.; Hollander, A.; Zijp, M.; van Zelm, R. ReCiPe 2016 : A harmonized life cycle

- 595 impact assessment method at midpoint and endpoint level Report I: Characterization,
596 *RIVM Report 2016-0104*, **2016**
- 597 21. Jin, H.; Afiuny, P.; McIntyre, T.; Yih, Y.; Sutherland, J. W. Comparative Life Cycle
598 Assessment of NdFeB Magnets: Virging Production versus Magnet-to-Magnet Recycling,
599 *23rd CIRP Conference on Life Cycle Engineering*, **2016**, 48 45-50
- 600 22. Li, X.; Chen, Z.; Chen, Z.; Zhang, Y. A human health risk assessment of rare earth
601 elements in soil and vegetables from a mining area in Fujian Province, Southeast China,
602 *Chemosphere*, **2013**, 93 (6) 1240-1246
- 603 23. Mancheri, N. World trade in rare earths, Chinese export restrictions, and implications,
604 *Resources Policy*, **2015**, 46 262-271
- 605 24. Mayer, H.; Gleich, B. Measuring Criticality of Raw Materials: An Empirical Approach
606 Assessing the Supply Risk Dimension of Commodity Criticality, *Natural Resources*,
607 **2015**, 6 56-78
- 608 25. Moss, R. L. Critical Metals in the Path Towards the Decarbonisation of the EU Energy
609 Sector (Luxembourg: Publications Office of the EU) **2013**
- 610 26. Nansai, K.; Nakajima, K.; Suh, S.; Kagawa, S.; Kondo, Y.; Takayanagi, W. The role of
611 primary processing in the supply risk of critical metals, *Economic Systems Research*,
612 **2017**, 1-22,
- 613 27. Nassar, N. T.; Du, X.; Graedel, T. E. Criticality of the Rare Earth Elements. *Journal of*
614 *Industrial Ecology*. **2015**, 19(6). <http://doi.org/10.1111/jiec.12237>
- 615 28. *National Research Council*, Minerals, critical minerals, and the U.S. economy. National
616 Academies Press, Washington, DC, **2008**
- 617 29. *NSTC*. Assessment of Critical Minerals: Screening Methodology and Initial Application;
618 Subcommittee on Critical and Strategic Mineral Supply Chains of the Committee on
619 Environment, Natural Resources, and Sustainability of the National Science and
620 Technology Council; Executive Office of the President, National Science and Technology
621 Council (NSTC) **2016**
- 622 30. Rao, Z. Consolidating policies on Chinese rare earth resources, *Mineral Economics*, **2016**,
623 29 (1) 23–28
- 624 31. Rim, K. T. Effects of rare earth elements on the environment and human health: A
625 literature review, *Toxicology and Environmental Health Sciences*, **2016**, 8 (3) 189-200
- 626 32. Roskill. *Rare earths: Global industry, markets and outlook* 2016 (16th ed.). London, UK:
627 Roskill.
- 628 33. Schneider, L.; Berger, M.; Schüler-hainsch, E.; Knöfel, S.; Ruhland, K.; Mosig, J;
629 Finkbeiner, M. The economic resource scarcity potential (ESP) for evaluating resource
630 use based on life cycle assessment, **2014**, 601–610. [http://doi.org/10.1007/s11367-013-](http://doi.org/10.1007/s11367-013-0666-1)
631 [0666-1](http://doi.org/10.1007/s11367-013-0666-1)

- 632 34. Sonnemann, G.; Gemechu, E.D.; Adibi, N.; De Bruille, V. Bulle, C. From a critical
633 review to a conceptual framework for integrating the criticality of resources into life cycle
634 sustainability assessment. *J Clean Prod*, **2015**, 94:20–34
- 635 35. Tilton, J.E.; Skinner, B.J. The meaning of resources. **1987** In: Skinner BJ, McLaren DJ
636 (eds) *Resources and world development*. Wiley, New York, pp 13–27
- 637 36. Udo de Haes, H.A.; Joliet, O.; Finnveden, G.; Hauschild, M.; Krewitt, W.; Müller-
638 Wenk, R. Best available practice regarding impact categories and category indicators in
639 life cycle impact assessment. *Int. J. Life Cycle Assess.* **1999**, 4, 167-174.
- 640 37. *USGS Mineral commodity summaries*. U.S. Geological Survey, Department of the
641 Interior, Reston, **2016**, https://minerals.usgs.gov/minerals/pubs/commodity/rare_earths/
- 642 38. Van Oers, L.; Guinée, J. The Abiotic Depletion Potential: Background, Updates, and
643 Future, *Resources*, **2016**, 5 (1) 16. doi:10.3390/resources5010016
- 644 39. Vieira, M. D. M.; Ponsioen, T. C.; Goedkoop, M.; Huijbregts, M. A. J. Surplus cost
645 potential as a life cycle impact indicator for metal extraction. *Resources* **2016**, 5:1–12