

1 **The challenge of habitat modelling for threatened low density species using**
 2 **heterogeneous data: the case of Cuvier's beaked whales in the Mediterranean**

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56 **ABSTRACT**

57 The Mediterranean population of Cuvier’s beaked whale (*Ziphius cavirostris*), a deep-diving
58 cetacean, is genetically distinct from the Atlantic, and subject to a number of conservation
59 threats, in particular underwater noise. It is also cryptic at the surface and relatively rare, so
60 obtain robust knowledge on distribution and abundance presents unique challenges. Here we
61 use multiplatform and multiyear survey data to analyse the distribution and abundance of this
62 species across the Mediterranean Sea. We use a novel approach combining heterogeneous
63 data gathered with different methods to obtain a single density index for the region. A total of
64 594,996 km of survey effort and 507 sightings of Cuvier’s beaked whales, from 1990 to 2016,
65 were pooled together from 24 different sources. Data were divided into twelve major groups
66 according to platform height, speed and sea state. Both availability bias and effective strip
67 width were calculated from the sightings with available perpendicular distance data. This was
68 extrapolated to the rest of the sightings for each of the twelve groups. Habitat preference
69 models were fitted into a GAM framework using counts of groups as a response variable with
70 the effective searched area as an offset. Depth, coefficient of variation of depth, longitude and
71 marine regions (as defined by the International Hydrographic Organization) were identified as
72 important predictors. Predicted abundance of groups per grid cell were multiplied by mean
73 group size to obtain a prediction of the abundance of animals. A total abundance of 5799
74 (CV=24.0%) animals was estimated for the whole Mediterranean basin. The Alborán Sea,
75 Ligurian Sea, Hellenic Trench, southern Adriatic Sea and eastern Ionian Sea were identified
76 as being the main hot spots in the region. It is important to urge that the relevant stakeholders
77 incorporate this information in the planning and execution of high risk activities in these high-
78 risk areas.

79
80 **KEYWORDS:** Cuvier’s beaked whales; abundance; distribution; conservation; density
81 surface modelling; correction factor; Mediterranean Sea

82 1. INTRODUCTION

83 The Cuvier's beaked whale (*Ziphius cavirostris*) is the only member of the Ziphiidae family
84 with a regular occurrence in the Mediterranean Sea, inhabiting both the western and eastern
85 basins (Notarbartolo di Sciara 2016; Podestà et al. 2016). Much of the early knowledge of this
86 species in the Mediterranean has come from stranding data (Figure S10 in Supplementary
87 Material). In total 316 animals were found between 1986 and 2003 (Podestà et al. 2006).
88 However, stranding data are potentially subject to severe bias because the location of the
89 strandings might be more related to the regional currents and the stranding place might be far
90 away from where the animals actually were, so they cannot be used alone to make strong
91 inferences about at-sea distribution (Peltier et al. 2014). The lack of more quantitative
92 distribution and abundance data has certainly contributed to the current 'Data Deficient'
93 IUCN listing for this species (Cañadas 2006), which means that there was insufficient
94 information available to assess the conservation status, and no Red List Category could be
95 assigned.

96 Cuvier's beaked whales seem to be relatively abundant in the eastern Ligurian Sea, off
97 southwestern Crete and in the Alborán Sea, especially over and around canyons (Cañadas and
98 Vázquez 2014; D'Amico et al. 2003; Frantzis et al. 2003). They appear to be regular
99 inhabitants of the western Ligurian Sea (Azzellino et al. 2008), the Hellenic Trench (Frantzis
100 et al. 2003), the southern Adriatic Sea (Holcer et al. 2007) and the eastern section of the
101 Alborán Sea (Cañadas et al. 2005; Cañadas and Vázquez 2014). They also occur in the central
102 Tyrrhenian Sea (Marini et al. 1992) and in Spanish Mediterranean waters (Raga and Pantoja
103 2004); M. Castellote, pers. comm.). However, survey effort and the efficiency of stranding
104 networks vary greatly across the region, with little or no effort to record sightings or to detect
105 strandings in some areas, particularly in the southern and eastern parts of the basin, except for
106 Syria and Israel (Aharoni 1944; Gonzalvo and Bearzi 2008; Kerem et al. 2012). In addition,
107 they are very difficult to detect reliably because of their long dive times (over 60 min; (Baird
108 et al. 2006; Baird et al. 2008; Cañadas and Vázquez 2014; Tyack et al. 2006) and usually
109 inconspicuous and brief appearances at the surface (Heyning 1989). As a result, knowledge of
110 the abundance and population trends in this population is severely limited. In the Gulf of
111 Genova (eastern Ligurian Sea) mark-recapture analysis (2002-2008) yielded estimates
112 between 95 (CV=9%) and 98 (CV=10%) using open population models (Podestà et al. 2016;
113 Rosso et al. 2009). In the Alborán Sea, off Southern Spain, spatial modelling of line transect
114 data (1992-2009) yielded an abundance estimate of 429 individuals (CV=22%, corrected for
115 availability bias; Cañadas and Vázquez 2014).

116 This species face multiple threats, of which the most significant are anthropogenic noise,
117 fishery interactions and shipping. Firstly, underwater acoustic pollution is recognized as a
118 threat for marine fauna, including deep diving species (Cox et al. 2006; Filadelfo et al. 2009).
119 Beaked whales appear especially vulnerable, with recorded cases of mortality as a
120 consequence of high-intensity noise in areas including the Mediterranean, Canary Islands,
121 United States, Bahamas and Japan, (Arbelo et al. 2008; Balcomb III and Claridge 2001;
122 Fernández et al. 2012; Frantzis 1998; Podestà et al. 2006). They have also shown behavioural
123 responses at sound levels well below those previously thought to affect this group (Cox et al.
124 2006; Fernández et al. 2012; Filadelfo et al. 2009; Pirotta et al. 2012; Tyack et al. 2011). The
125 numerous cases where mass-strandings of beaked whales followed (and where related to)
126 naval exercises (Balcomb III and Claridge 2001; Filadelfo et al. 2009; Frantzis 1998) have
127 resulted in these species becoming indicators for the effects of high intensity anthropogenic
128 noise.

129 Secondly, fishery interactions are a consistent threat to all Mediterranean cetaceans (Reeves
130 and Notarbartolo di Sciara 2006), and this includes Cuvier's beaked whales. Fourteen were

131 reported as having been captured incidentally between 1972 and 1982 (11 in French waters
132 and 3 in Spanish waters (Northridge 1984)) and two more in Italian waters in subsequent
133 years (Notarbartolo di Sciara 1990). Entanglement in fishing gear and other marine debris
134 have also been recorded (Cañadas and Vázquez 2014; Podestà et al. 2016), but actual
135 occurrence is unknown.

136 Finally, the Mediterranean is one of the busiest shipping regions in the world. Large cetaceans
137 are vulnerable to ship strikes and increased sea ambient noise. While there are no data on ship
138 strikes on Cuvier's beaked whales in the Mediterranean, Carrillo and Ritter (2010) reported
139 that 12% of the strandings with signs of ship strikes in the Canary Islands correspond to
140 beaked whales. Additionally, shipping increases ambient noise, with the potential to mask the
141 ultrasonic echolocation signals of beaked whales and thereby interfere with their sensory
142 biology (Aguilar Soto et al. 2006).

143 Increasing awareness of numerous and synergistic threats to cetaceans in the Mediterranean
144 Sea led, in part, to the creation of ACCOBAMS (Agreement for the Conservation of
145 Cetaceans in the Black Sea, Mediterranean Sea and Atlantic contiguous waters), under the
146 auspices of the Convention on migratory species. The Fourth meeting of the Scientific
147 Committee of ACCOBAMS (Monaco, November 2006) addressed the issue of the impact of
148 anthropogenic noise on marine mammals in the Mediterranean, and noted that in the specific
149 case of Cuvier's beaked whales, fundamental information on their distribution and habitat use
150 in the Mediterranean waters was scarce. The Committee agreed that information on the
151 distribution and habitat use of Cuvier's beaked whales in the region should be made available
152 to interested parties and stakeholders to prevent the production of high intensity noise in areas
153 of high density for this species. Given that appropriate data on distribution and relative (or
154 absolute) abundance of Cuvier's beaked whales in the Mediterranean were lacking, the
155 Committee recommended that a habitat modelling exercise should be attempted for the
156 Mediterranean Sea.

157 The use of multiplatform and multiyear survey data from multiple sources to estimate the
158 distribution and abundance of cetacean species is extremely challenging, but made necessary
159 by the paucity of data and large scale objectives of the study. For species which are
160 threatened, rare and difficult to detect, whose spatial range encompasses both international
161 and waters of multiple nations, pooling together all available information is the only option
162 for increasing knowledge. Heterogeneity in factors such as the data collection procedures,
163 height and speed of the platforms, observer experience, and so forth, can easily lead to biased
164 results (Jewell et al. 2012). Pooling together large amounts of multiplatform data to yield a
165 single result per species has been previously achieved using both line transect data (Jewell et
166 al. 2012; Roberts et al. 2016) and presence only data (Kaschner et al. 2006; Ready et al.
167 2010). Combining heterogeneous effort related data from both line transect data *and* non-line
168 transect data (i.e. with and without perpendicular distances) to obtain a single density index
169 has not however been done before to our knowledge. Here we present the results of an effort
170 to pool such data on Cuvier's beaked whales in the Mediterranean region. We adopted a novel
171 approach to combine heterogeneous data into a single habitat preference model. This was
172 based on stratification by platform type, extrapolation of perpendicular distance data
173 according to such stratification, and the application of correction factors to take into account
174 availability bias according to platform type.

175

176 2. METHODS

177 2.1 Data collection and compilation

178 Twenty four institutions contributed data, totalling 594,996 km of survey effort in good to
179 moderate visual conditions (sea state of Beaufort 3 or less). This survey effort yielded 507
180 sightings of Cuvier's beaked whales with a total of 1,166 individuals, covering a time span
181 from 1990 to 2016 (Table S1 in the Supplementary Material; Figure 1). These data are
182 divided by time period and platform type in the online supplementary material (Figures S1-6).

183 Areas with a low research effort and areas with no research effort were due to lack of funding
184 and/or lack of permits in some countries.

185 It was not possible to constrict the data used to a single platform type (e.g. ships vs airplanes,
186 large ships vs small ships) because none of them cover all the areas, so very large portions
187 would remain empty of effort and the purpose of this collaborative and integrating effort
188 would be meaningless. However, to minimise the potential bias created by using different
189 platforms, a correction factor is fundamental (see point 2.2.2 below).

190

191 2.2 Data organization

192 2.2.1 Sampling units

193 A grid of 7287 cells with a resolution of 0.2° (22.2 km) was built (with an average size of 494
194 km², ranging from 403 km² in the northern part of the area to 455 km² in the South). The size
195 of the grid was chosen as a trade-off between limiting the number of grid cells for
196 computational reasons and the resolution of the available covariates. A number of
197 geographical and environmental covariates were associated to each grid cell. These were of
198 three types: (a) Geographic: latitude and longitude, and Marine Region; (b) Fixed: depth,
199 distance from the 200, 1000 and 2000 m isobaths, coefficient of variation of depth, slope,
200 contour index ((max depth-min depth)*100/max depth), aspect (orientation of sea floor in
201 360°), factor with classification into three levels: Abyss, Slope and Shelf (Ab-Sl-Sh), factor
202 with classification into three levels: Canyon, Escarpment, or None (Cany-Escarp-None),
203 distance from the slope area (steep area between the continental shelf and the abyss plains),
204 from canyons and escarpments, and from sea mounts; (c) Dynamic: SST_All (mean annual
205 sea surface temperature 1990-2015) and SST.SD_All (Inter-annual standard deviation of the
206 annual sea surface temperature 1990-2015). The covariate 'Marine Regions' (see Figure S7
207 in supplementary material), is a subdivision of the Mediterranean basin into smaller areas,
208 obtained from the International Hydrographic Organization (IHO 1953). The large Libyan-
209 Levantine basin was subdivided into Libyan and Levantine according to the ICES ecoregions
210 (ICES 2004). The Hellenic Trench was added as a separate region (IHO 2016). Figure S11
211 shows the depth contours in the Mediterranean Sea.

212 Search effort was divided into segments fitting grid cells, with the tool *Identity* in ArcGis. In
213 this way, each segment of search effort track was assigned to a grid cell, and the covariates
214 associated with that grid cell were then associated to that segment, as well as the source (data
215 owner), type of survey (aerial, ferry, large research ship or small ship/boat), day and sea state.
216 This resulted in a total of 107,393 segments. These segments were aggregated in each grid
217 cell according to source and year, totalling 16,554 units of source-year-cell, which constituted
218 the sampling units, with total effort (in km), number of sightings, and number of animals
219 associated with unit. The total number of grid cells containing effort was 4449, representing
220 61.0% of the total Mediterranean Sea.

221 No stratification was possible by season or year (nor was the temporal aspect included as a
222 covariate) due to the high heterogeneity in coverage and platforms used among seasons and
223 among years. Areas with year-round effort, such as the Alborán Sea (Cañadas and Vázquez
224 2014) and Ligurian Sea (Rosso et al. 2011), have sightings of this species in the same areas in
225 all seasons, suggesting that major seasonal changes in distribution do not occur, although it
226 must be noted that these data pertain only to a sub-section of the study area.

227 *2.2.2 Correction for availability*

228 There was considerable heterogeneity in survey platforms (and therefore observer height and
229 platform speed). Platforms included aerial surveys (fast speed and pre-designed routes),
230 ferries (high observation point and speeds, usually around 30 km/h), research and whale
231 watching ships or boats (speed ranging between 6 and 14 km/h, and observer heights between
232 3 and 15 m). Platform speed was either provided directly or measured from the GPS data for
233 all segments. While in most cases the approximate height of the observation platform (an
234 approximation to the height of the observer's eye) was available, in some cases it was
235 assumed based on the characteristics of the vessel.

236 Density estimates from line transect surveys are usually subject to availability bias, due to
237 animals not always being available for detection (e.g. actually surfacing) while within
238 detectable range (Buckland et al. 2004), and perception bias due to observers failing to detect
239 animals even though they are available to be detected (Buckland and Elston 1993). For
240 beaked whales, both sources of bias are known to be important (Barlow 1999, 2006; Borchers
241 et al. 2013; Cañadas and Vázquez 2014). Correcting for perception bias typically requires
242 some form of double platform approach, and was not possible here because no such data were
243 available. However, we were able to take steps to mitigate the effect of availability bias.

244 As no radial or perpendicular distances were available for most datasets, abundance could not
245 be estimated with the distance sampling method (Buckland et al. 2001). However, such
246 distances were available for some of the datasets, allowing the estimation of an availability
247 bias. The availability bias was used as a correction factor to minimize the heterogeneity in
248 platforms and the large spatial differences in coverage by different platform types, which
249 could yield a bias in the density surface modelling. Laake et al. (1997) developed a correction
250 factor, \hat{a} , to correct estimates for availability bias. This factor takes into account the average
251 duration of the availability (animals present at surface) and unavailability (animals
252 underwater) and the time an animal is within a detectable range. The detectable range was
253 estimated by dividing the maximum forward distance at which animals are expected to be
254 detected by the platform's speed. The average duration of availability and unavailability was
255 estimated using data on focal follows of Cuvier's beaked whales collected during surveys in
256 the Alborán Sea in 2008 and 2009 (Cañadas and Vázquez 2014). For the datasets with
257 available radial distances, these were used to estimate the forward distances for the sightings.
258 Subsequently the particular correction factor for availability bias for a range of platform
259 speeds for those datasets were estimated, using a cut-off point of 80% of the forward
260 distances to avoid outliers (Cañadas and Vázquez 2014). The range of speeds used was
261 between 1 and 50 km/h (depending on the range of each platform, and at intervals of 0.1
262 km/h) and 185 km/h for aircraft. For other datasets without radial distance, the correction
263 factors of the platforms with similar attributes of type and height were assigned. Given that
264 the potential maximum radial distance of detection depends largely on the height of the
265 observation platform (as proxy to height of observer eye), data were divided into twelve major
266 groups according to the platform height, speed and sea state following Cañadas and Vázquez
267 (2014)(Table 1).

268 *2.2.3 Correction for effective searched area*

269 A similar procedure was used to estimate an effective strip width (esw) which was associated
 270 with all segments of effort. Using the known perpendicular distances where available, specific
 271 detection functions were created for all the platform groups. The particular esw for each
 272 platform type was estimated from their detection function and used for all platforms in that
 273 group. An effective search area was calculated for each segment (included in the models as
 274 offset), as $L*2*esw$ where L is the length of the segment (in kilometres). The mean speed for
 275 all segments of a particular platform and year was used to obtain a mean \hat{a} and esw for each
 276 platform/year. Finally, the calculated effective search area for each segment was multiplied by
 277 the appropriate mean \hat{a} to obtain the effective search area corrected for availability bias. This
 278 was then used as the final offset in the spatial models (Table 1).

279 We assumed that for similar platform type, height and speed, and similar sea state conditions,
 280 the mean availability bias and mean esw were similar. Other factors that might affect
 281 estimates of availability bias and esw include observer experience, the number of observers
 282 and searching protocols. However, as these could not rigorously be corrected for these factors,
 283 we assumed that the main sources of variability associated with platform height and speed
 284 were taken into account.

285

286 **2.3 Data analysis**

287 *2.3.1 Spatial models and abundance estimate*

288 The response variable used to formulate the spatial models of abundance of groups was the
 289 count of groups (N) in each sampling unit (Hedley et al. 1999). The abundance of groups was
 290 modelled using a Generalized Additive Model (GAM) with a logarithmic link function.
 291 Overdispersion was tested in models with a Poisson distribution using the Poisson Pearson
 292 residuals ($\sum \text{residuals}^2 / (N-p)$ where N is the sample size of effort and p is the number of
 293 parameters of the model). The results was 7.3, way above the acceptable limit of 1.5 for a
 294 Poisson distribution. Therefore, a Tweedie error distribution was used, with a parameter p of
 295 1.1, very close to a Poisson distribution but with some over-dispersion.

296 The general structure of the model was:

$$297 \quad \hat{N}_i = \exp \left[\ln(a_i) + \theta_0 + \sum_k f_k(z_{ik}) \right] \quad (2)$$

298 where the offset a_i is the search area for the i^{th} sampling unit (corrected for availability bias),
 299 θ_0 is the intercept, f_k are smoothed functions of the explanatory covariates, and z_{ik} is the value
 300 of the k^{th} explanatory covariate in the i^{th} segment.

301 Models were fitted using package ‘mgcv’ version 1.7-22 for R (Wood 2011). Model selection
 302 was done manually using three diagnostic indicators: (a) the GCV (Generalised Cross
 303 Validation score, an approximation to AIC; Wood 2000); (b) the percentage of deviance
 304 explained; and (c) the probability that each variable was included in the model by chance (p-
 305 value of the covariate in the model). Only one of the collinear covariates was used in each
 306 iteration of model selection, unless the collinearity was weak and the inclusion of the two
 307 covariates improved the model. Table S2 (Supplementary Material) shows the Pearson's product-
 308 moment correlation among pairs of all continuous covariates.

309 The model returned a prediction for the abundance of groups in each grid cell. A model for
 310 group size was attempted but there were no significant results, so we assumed there was no

311 systematic variation in group size across the study area. Therefore, we multiplied the
312 predicted number of groups in each grid cell by the mean group size of the Marine Region to
313 which the cell belonged (Figure S7 in Supplementary Material). The point estimate of total
314 abundance was then obtained by summing the abundance estimates of all grid cells over the
315 study area and plotted as a density surface map in ArcGis 10.0.

316 Finally, a non-parametric bootstrap with replacement with 400 iterations was used to generate
317 the model coefficient of variation (CV) and 95% confidence intervals for the resulting habitat
318 use prediction maps and abundance estimates. To obtain a total CV, the model CV was
319 combined with the overall *esw* CV and mean \hat{a} CV through the Delta method (Seber 1982).

320

321 3. RESULTS

322

323 All the group size records ranged between 1 and 8 individuals, with only one large group of
324 20 animals in the Alborán Sea. Mean group sizes ranged between 1.6 in the Libyan Sea and
325 2.5 in the Ionian Sea. Figure S11 (Supplementary Material) shows the detection functions for
326 all the combinations for which data were available, to obtain a measure of *esw*.

327 A total of 60 models were tried with different combinations of covariates. The best model for
328 abundance of groups, according to the diagnostics, included four covariates: depth, coefficient
329 of variation of depth, longitude and marine region, with a total deviance explained of 34%
330 (Table 2; Figure 2). All the other models either had smaller deviance explained, larger GCV,
331 non-significant covariates or edge-effect issues.

332 The total abundance estimate obtained through modelling, once the correction factor for the
333 effective searched area was applied, was 5799 animals in the whole Mediterranean (4261 when
334 excluding the area south of 34.3°N and the Aegean Sea), with a total CV of 24.0%
335 ($CV_{\text{model}}=11.5\%$; $CV_{\text{esw}}=14.7\%$; $CV_{\hat{a}}=15.0\%$) and a 95% CI of 4807 – 7254. This would equate
336 to an overall density of 0.00223 animals per km² for the whole Mediterranean.

337 Figure 2 shows the smoothed functions of the continuous covariates selected in the final model.
338 Cuvier's beaked whales show a highest density between 1000 and 1500m. Density declines
339 sharply in waters shallower than 1000m. There is also a preference for areas with medium to
340 high variability in bottom depth (CV of depth). However, the areas with highest CV of depth
341 are associated with low data density, so have a large prediction uncertainty and results for
342 these areas should therefore be interpreted with caution. The smooth term associated with
343 longitude has a lower density around 14°E-18°E, including the northern Adriatic, eastern
344 Tyrrhenian Sea and southeast of Sicily, and a much less pronounced area of low density
345 between 4°E-7°E (Figure 3) between France and Algeria.

346 The predicted abundance of Cuvier's beaked whales in the Mediterranean (Figure 3) shows
347 two areas marked with diagonal lines: the area south of 34.3°N and the Aegean Sea, where
348 reliability is low due to the very low effort (Figure 1). Figures S8 and S9 (Supplementary
349 Material) show the lower and upper 95% confidence intervals. Figure S10 (Supplementary
350 Material) shows the beaked whale sighting and stranding locations overlying this prediction.

351

352 4. DISCUSSION

353 Little or no data were available for large portions of the region, so it is necessarily the case that
354 the conclusions we draw here regarding distribution and abundance need to be taken with

355 caution. Therefore, the results presented here ideally need to be validated by a systematic and
356 region-wide survey of the Mediterranean Sea.

357 *4.1 Habitat preferences*

358 Cuvier's beaked whales show a clear habitat preference for areas with depths over 1000m,
359 and medium to high variability in bottom depth (CV of depth), which would usually include
360 escarpments, canyons and sea mounts. This coincides with previous descriptions of the
361 habitat of this species in the Mediterranean and the Northeast Atlantic as a predominantly
362 oceanic species often associated with steep slope habitat and a marked preference for
363 submarine canyons and escarpments (D'Amico et al. 2003; Frantzis et al. 20013; MacLeod
364 2005; Podestà et al. 2006; Azzellino et al. 2008). Also in the Eastern Tropical Pacific habitat
365 modelling on this species show a preference for depths over 1000m (Ferguson et al. 2006), as
366 does an habitat-cetacean relationship study in the Gulf of Mexico (Davis et al. 1998), among
367 other studies.. The lower density around 14°E-18°E detected by the smoothed term of
368 Longitud, coincides with shallower areas of the northern Adriatic and the southeast of Sicily.
369 Considering that there is generally good effort coverage in this region it suggests that this is a
370 genuine area of relatively low density. In contrast, there is little effort between France and
371 Algeria (4°E-7°E, less pronounced area of low density), especially in the south, so this
372 apparent gap in distribution should be treated with caution.

373 It is interesting to look at the effect of other covariates explored. The factor "Cany_Escarp",
374 with three levels: Canyon, Escarpment or None, explained 7% of the deviance and had a
375 positive effect (higher density) for Escarpment and negative for None, with respect to Canyon
376 (which was the intercept). Its associated covariate "Dist_c_e" (distance from canyons and
377 escarpments) explained 8.3% of the deviance and predicted higher numbers with declining
378 distances from canyons and escarpments. The distance from sea mounts (Dist_mounts
379 explained 9.2% of the deviance, and showed a strong positive effect at the closest distances,
380 and a second, smaller peak at long distances. Distance from the slope (Dist_Slope) explained
381 6% of the deviance and had a more positive effect at closer distances from the slope area. The
382 same happened with "Dist_1000", explaining 9% of the deviance. This information is
383 consistent with existing knowledge about habitat use by Cuvier's beaked whales (a preference
384 for deep waters and steep floors; e.g. Cañadas and Vazquez 2014; Arcangelli et al. 2016;
385 Podestà et al. 2016), suggesting that areas of high bathymetric relief are important for
386 Cuvier's beaked whales.

387 The main influence of the physical environment over cetacean distribution is most probably
388 the aggregation of prey species (Baumgartner 1997; Davis et al. 1998). For beaked whales
389 main prey species, cephalopods, sea floor physiography could play an indirect role through
390 mechanisms such as topographically induced up-welling of nutrients (Guerra 1992; Rubin
391 1997), increased primary production, and aggregation of zoo-plankton due to the enhanced
392 secondary production or convergence of surface waters (Rubin 1994). This would be in total
393 accordance with the patterns described above for Cuvier's beaked whales.

394 *4.2 High-use areas*

395 The best model highlighted six high density areas for beaked whales: Ligurian Sea, Alborán
396 Sea, Hellenic Trench, northern Ionian Sea, southern Adriatic Sea and northern Tyrrhenian Sea
397 (listed in decreasing order of density). These areas, particularly the first three, are supported by
398 a large proportion of the available sightings, giving more confidence that these are genuinely
399 high-use areas. All these areas are also well represented in the predicted map of lower 95%
400 confidence interval (Figure S8, Supplementary Material). This map is useful to show which
401 areas are the minimum hot spots for which we are certain at a 95% level of confidence. Most of
402 these areas, with the exception of the Levantine and Libyan basins, have previously been

403 reported as high-use areas by Cuvier's beaked whales (Arcangeli et al. 2016; Cañadas and
404 Vázquez 2014; Rosso et al. 2009).

405 Akkaya Bas et al. (2014) reported sightings of Cuvier's beaked whales in Antalya Bay, Turkey.
406 In this area, where a deep canyon and steep escarpment exist, there is also one stranding
407 (Podestà et al. 2016). Low to medium model predictions of density in this area, despite poor
408 information available for the model, suggests that further research effort may be worthwhile
409 here.

410 Much less confidence can be accorded to many areas of low predicted density because of
411 insufficient effort. These include the south-eastern Mediterranean, the Aegean Sea, the waters
412 north of Algeria and the Gulf of Lion. Additional survey effort should be made to assess the
413 occurrence of Cuvier's beaked whales in these regions. More generally, predictions in areas of
414 little or no effort are useful only in an exploratory region-wide context. This is why results for
415 the whole section south of 34.3°N and the Aegean Sea should be considered with caution
416 (Figure 3).

417 *4.3 Abundance estimate*

418 The lack of data on perpendicular distances from the trackline in most datasets meant that our
419 estimate of abundance relied heavily on the correction factors applied and the extrapolation of
420 the estimated e_{sw} from the available data according to the characteristics of the platforms.
421 However, we still consider it worthwhile to contribute an estimate of the population size of
422 Cuvier's beaked whales in the Mediterranean, given the concern regarding its conservation. The
423 abundance estimate provided here, of approximately 5800 individuals, should be taken with
424 caution as it only provides a tentative order-of-magnitude estimate for the population size of
425 Cuvier's beaked whales in the Mediterranean.

426 We were able to explore the reliability of our method by comparing with the only two available
427 abundance estimates of Cuvier's beaked whales in the Mediterranean: the Alborán Sea (Cañadas
428 and Vázquez 2014) and the Ligurian Sea (Rosso et al. 2009). When comparing the Alborán Sea,
429 by summing up the grid cells corresponding to the area for which an abundance estimate was
430 provided (Cañadas and Vázquez 2014), results are very similar. The original abundance
431 estimate of Cañadas and Vázquez (2014) was 429 individuals (CV=22%), in both cases taking
432 into account the correction factor for availability bias. For the same area, in the current
433 modelling exercise the estimate was 417 individuals. Similarly, when comparing the area of the
434 Ligurian Sea, by summing up the grid cells corresponding to the area for which an abundance
435 estimate from photo-identification exists (Rosso et al. 2009), the results are comparable. Rosso
436 et al. (2009) calculated the abundance estimate to be 95-98 (SD=9-10) individuals. For the same
437 area, in the current modelling exercise the estimate was 94 individuals.

438 Additionally, an abundance estimate was attempted with ISPRA-Tethys aerial surveys in the
439 Ligurian Sea and Central and South Tyrrhenian Seas from 2009 to 2014, with all seasons
440 pooled together. There were only nine sightings of Cuvier's beaked whales. Despite this, a
441 detection function could be fitted given the pattern of the distance data for this species with
442 good diagnostics of goodness of fit (this abundance estimate should only be considered in the
443 framework of this exploration, as sample size was too small). An abundance estimate of 59
444 individuals was obtained, which, corrected by the availability bias estimated for this survey
445 (0.078; see Table 1), yielded an estimate of 756 animals (CV=56.6%). When comparing the
446 area corresponding to this survey using the same methods as for the Alboran Sea and Ligurian
447 Sea results are once again similar. In the current modelling exercise the estimate was 755
448 individuals for the same area. Of course, the data from the surveys that generated these figures
449 were included in the present analysis, so it is not a genuinely independent test, but it does
450 indicate that the modelling approach we adopted is comparable to more standard approaches.

451 Given that our estimate was obtained through an unorthodox process, a full basin-wide survey
452 with line transect data collection is needed to obtain reliable estimates of abundance. Until then,
453 the preliminary information provided here could be used as a baseline. This analysis used a
454 compilation of 27 years of data, collected from a variety of survey platforms, by observers with
455 variable experience, with heterogeneous geographic coverage, under both good and moderate
456 sighting conditions. Little or no data were available for large portions of the region. Therefore,
457 the results presented here ideally need to be validated by a systematic and region-wide survey of
458 the Mediterranean Sea. Such a line transect survey would also confirm the validity or otherwise
459 of the approach used here for analysing multiplatform, multiyear, heterogeneous data covering
460 large areas for which no systematic surveys exist

461 *4.4 Strandings and mass strandings*

462 A further check of our results can be made by comparing with independent observations of
463 stranding events. Making inferences from strandings is problematic because carcasses may end
464 up stranding at a point on the coast which is actually distant from where the animal died.
465 Regardless, stranding records often compare well with sightings records (Maldini et al. 2005;
466 Peltier et al. 2014). Mass strandings can provide more useful information because these events
467 concern animals that strand alive or very fresh, potentially close to the area where they suffered
468 the stress that made them strand. Most mass stranding events reported by Podestà et al. (2016)
469 coincide with, or are very close to areas, where our model predicted higher densities of Cuvier's
470 beaked whales (Figure S10 in Supplementary Material).

471 The southern portion of the Mediterranean lacks stranding data. This does not, however, mean
472 that there are no strandings in that area, but rather that information is unavailable. Numerous
473 stranding records, including one mass stranding reported off the coast of Israel (Kerem et al.
474 2012; Podestà et al. 2016) suggest that these events may also occur in surrounding areas, but
475 remain unreported.

476 There have been a few mass strandings in the Balearic region, where the predicted density is not
477 particularly high. This corresponds with the fact that there are very few sightings in this region,
478 however, most of the surveys have been aerial, and the probability of detecting long divers like
479 Cuvier's beaked whales is rather low. Therefore, given the amount of strandings in this area,
480 coincident with the presence of some sightings and a medium density prediction, it would be
481 advisable to survey this region with a platform that allows for easier detection of deep divers.

482 *4.5 Implications for conservation and management*

483 Assuming the abundance estimate is on the correct order of magnitude, our results could
484 contribute toward an IUCN Red List assessment and upgrading of the Mediterranean
485 subpopulation of Cuvier's beaked whales, currently classified as Data deficient (Cañadas 2006).

486 The areas of predicted high density, together with the areas of concentration of atypical mass
487 strandings, constitute areas of concern for conservation of the Mediterranean Cuvier's beaked
488 whales population (Figures 3 and S10 in Supplementary Material). These maps concur with
489 long-held opinions of the scientific and regulatory community: that there are a number of
490 Mediterranean areas where Cuvier's beaked whales are often found and can be considered to be
491 at risk of exposure to high intensity anthropogenic noise, such as the Alboran Sea, the Ligurian
492 Sea and the Hellenic Trench. The other areas are not risk free, but rather of unknown risk,
493 where data are required to assess beaked whale presence prior to, and during, human activities
494 of potential impact (ACCOBAMS 2010; Kendra 2009). We know of multiple mass strandings
495 associated with intense anthropogenic noise production (Frantzis 1998; Podestà et al. 2016), but
496 mortality of Cuvier's beaked whales could be much higher considering that the probability of
497 finding a carcass of a deep diving species can be as low as 3% (Williams et al. 2011).

498 Therefore, it is important to recommend caution in these high-risk areas of the Mediterranean,
499 and urge that the relevant bodies incorporate this information in the planning and execution of
500 high risk activities, such as naval exercises and seismic surveys.

501 Avoiding the production of high levels of noise within the areas with predicted higher density of
502 Cuvier's beaked whales identified here (Figure 3) will undoubtedly reduce the risk of exposure
503 and consequent mortalities for a significant part of the Mediterranean population of this species.
504 Mitigation should include dedicated surveys and monitoring efforts. Additionally, mitigation
505 requirements should be incorporated into national regulations and incorporated into the
506 planning, consultation and permitting processes whenever the use of high-intensity noise is
507 planned in the Mediterranean.

508

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516

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703 **Table 1.** Mean speed (km/hr), associated mean correction factor for availability bias (\hat{a}), and
704 estimated esw (km) per group of platform type/height/sea state, total track length (km) total
705 area searched before correction ($L*2*esw$, km^2), and total area searched after correction
706 ($L*2*esw*\hat{a}$, km^2). Large ships of more than 15m platform height used BigEyes binoculars
707 (usually more than 20x magnification), while large or medium ships of more than 10m
708 platform height did not use BigEyes binoculars. Small ships could either use crow's nest
709 platform (10-12 m height), deck (3-4.5 m) or both/undefined (3-12m). Sea state "0-3" means
710 it was undefined but less than 4 Beaufort.

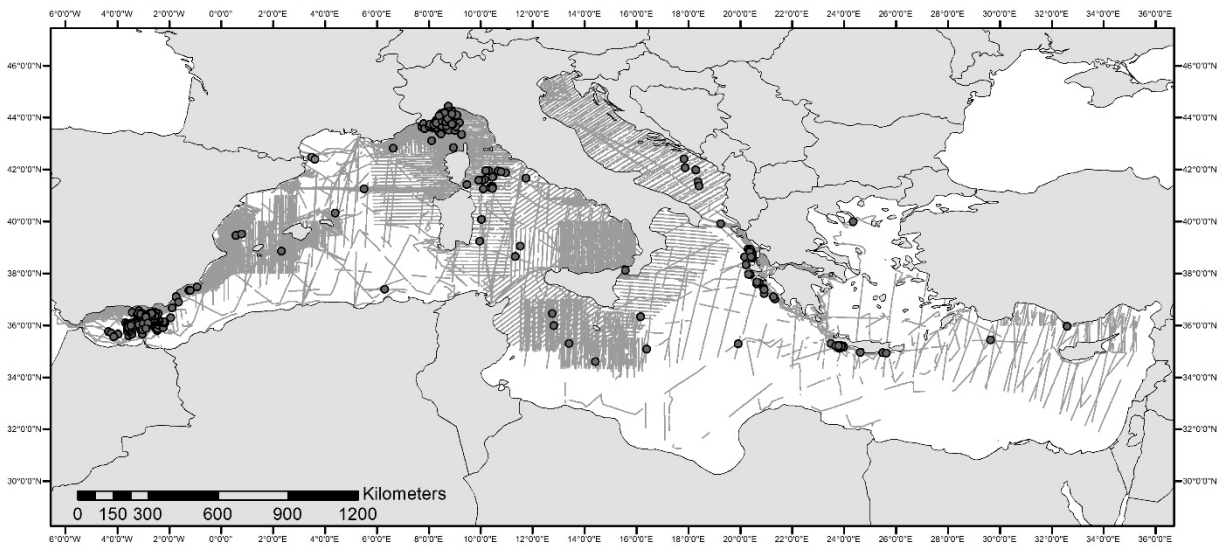
Platform type	Platform height (m)	Sea state	Mean speed	Mean \hat{a}	Estimated esw	Track length	Search area (not corrected)	Search area (corrected)
Large ship	>15	0-1	10.15	0.8677	2.280	1134	5173	4496
		2-3	10.02	0.7778	1.930	2676	10320	8055
Large or medium ship	>10	0-1	25.92	0.6582	1.410	7497	21141	10376
		2-3	38.26	0.4053	1.440	15296	44051	15048
		0-3	26.08	0.6710	1.460	17176	50153	32046
Small ship	10 - 12	0-1	8.77	0.6715	1.080	30313	65476	43911
		0-3	9.12	0.4654	0.480	24440	23462	10602
	3 - 4.5	0-1	13.05	0.3388	0.350	204190	142933	51076
		0-3	11.71	0.4519	0.980	19240	37711	17100
	3 - 12	2-3	10.31	0.2521	0.250	61391	30696	7688
		0-3	9.67	0.4392	0.780	18478	28862	12807
Aircraft		0-3	185	0.0781	0.615	193168	237597	18622
TOTAL			63.43	0.3016	0.573	594996	697538	231826

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712 **Table 2.** Covariates selected in the model, their estimated degrees of freedom (approximately
713 number of knots in the smoothed function - 1) and their p-value (probability that their inclusion
714 in the model is by chance).

<i>Covariates</i>	<i>Estimated degrees of freedom</i>	<i>P value</i>
Depth	4.87	<<0.0001
Depth CV	4.99	<<0.0001
Longitude	8.83	<<0.0001
Marine Regions (factor)	Coefficient	P value
(Intercept – Adriatic Sea)	-3.4714	0.0079
Aegean Sea	-3.7951	0.0188
Alborán Sea	-8.3304	0.0033
Balearic Sea	-9.4726	<<0.0001
Hellenic Trench	-1.8803	0.0417
Ionian Sea	-1.2692	0.0732
Levantine Basin	-3.4277	0.0822
Libian Basin	-1.5717	0.1255
Ligurian Sea	-5.5045	0.0005
NorthWestern Basin	-8.5522	<<0.0001
SouthWestern Basin	-10.9357	<<0.0001
Tyrrhenian Sea	-4.5613	0.0014

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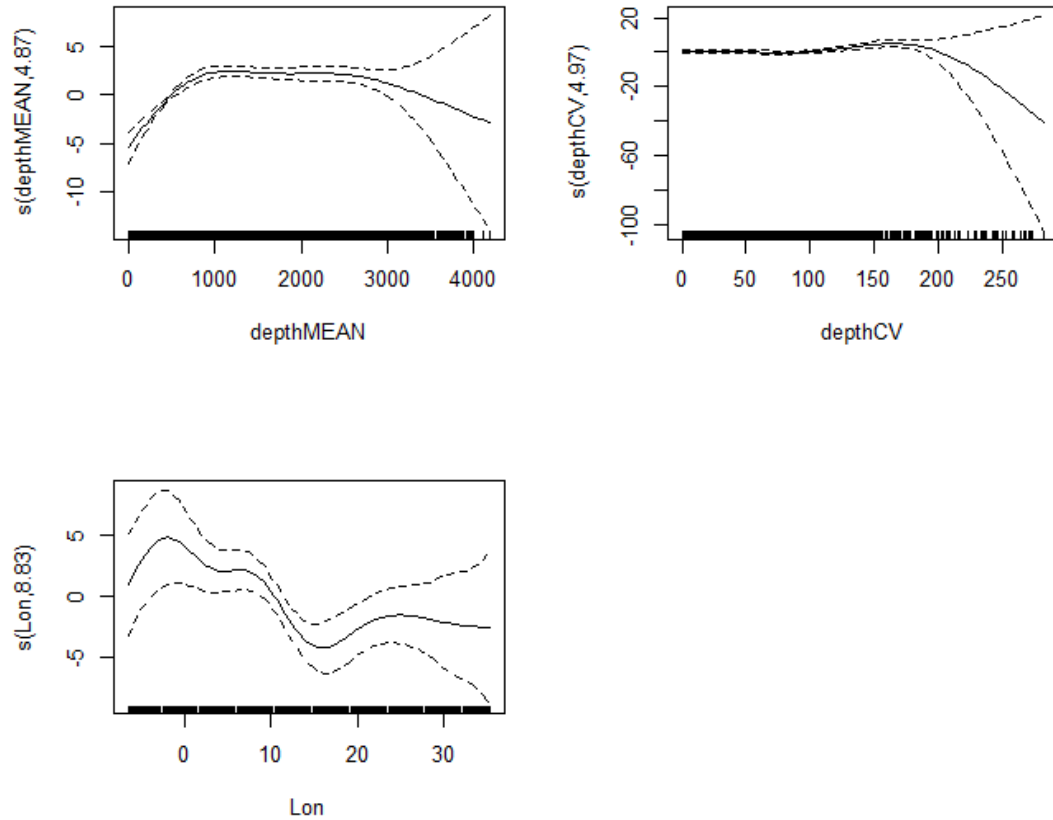


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Figure 1. Searching effort (track lines) and sightings of beaked whales from 1990 to 2016.

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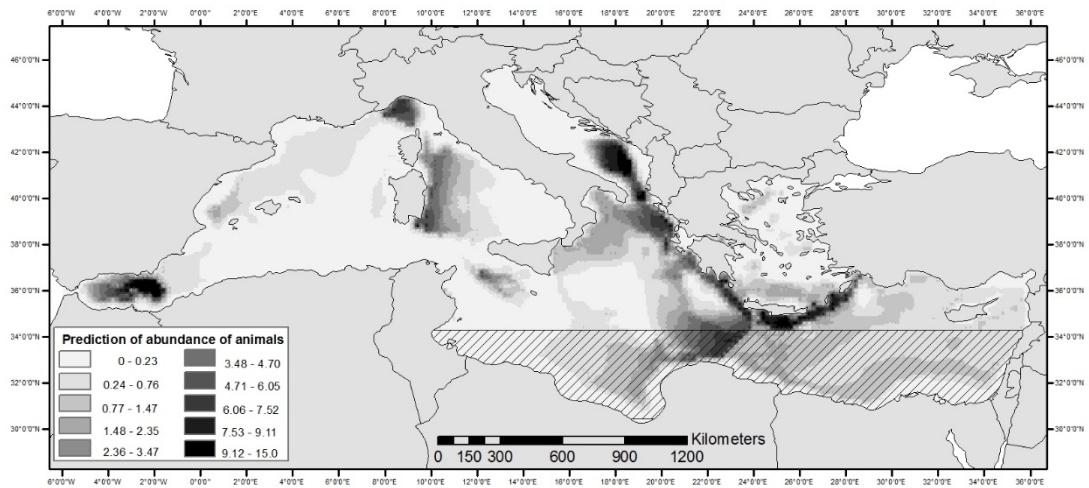
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Figure 2. Smoothed functions of the continuous covariates selected in the final model of abundance of groups: depth, depth CV and longitude. The ticks on the x axis show the distribution of the sample data used in the model for each covariate. The dashed lines represent ± 1 se. The y-axis represents an index of relative density. When the fitted line of the smooth function is greater than 0, the covariate has a positive effect and *vice versa*.



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Figure 3. Predicted abundance of beaked whales in the whole Mediterranean (the grey scale represent the number of animals predicted in each grid cell). Results in striped areas (Aegean Sea and South-eastern Mediterranean) are not very reliable due to very small sample size.