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Synergistic effects of combining ultrasound with the Fenton process in the degradation of Reactive Blue 19

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1	Synergistic effects of combining ultrasound with the Fenton process in the
2	degradation of Reactive Blue 19
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10	Modifications to manuscript after review in red
11	
12	Abstract
13	The decoloration of reactive dye C.I. Reactive Blue 19 (RB 19) using combined
14	ultrasound with the Fenton process has been investigated. The effect of varying the
15	concentrations of hydrogen peroxide and iron sulphate, initial pH, ultrasonic power,

initial dye concentration and dissolved gas on the decoloration and degradation 16 efficiencies was measured. Calibration of the ultrasound systems was performed using 17 18 calorimetric measurements and oxidative species monitoring using the Fricke dosimeter 19 and degradations were carried out with a 20 kHz probe type transducer at 2, 4, 6, 8 W cm⁻ ² of acoustic intensity at 15, 25, 50, 75 mg L⁻¹ initial dye concentrations. First order rate 20 21 kinetics was observed. It was found that while the degradation rate due to ultrasound 22 alone was slow, sonication significantly accelerated the Fenton reaction. While the results 23 were similar to those reported for other dyes, the effects occurred at lower concentrations.

24	The rate and extent of decoloration of RB 19 increased with rising hydrogen peroxide
25	concentration, ultrasonic powers and iron sulphate concentration but decreased with
26	increasing dye concentration. An optimum pH value of pH=3.5 was found. The rate of
27	decoloration was higher when dissolved oxygen was present as compared with nitrogen
28	and argon confirming the solution phase mechanism of the degradation.
29	
30	Keywords: Dye degradation; ultrasound promoted fenton process; ultrasonic cavitation;
31	dissolved gases; wastewater treatment
32	
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35	
36	1. Introduction
37	A variety of chemically different dyes are used for various industrial applications
38	such as textile dyeing, paper printing, leather, shoe polish, plastics, food coloring etc. A
39	significant amount of these dyes enter the environment as wastewater [1]. There are more
40	than 100,000 types of dyes commercially available and over 7 x 10^5 tonnes of dyestuff
41	are produced annually [2]. Reactive dyes are resistant to light, water and oxidizing agents
42	and are therefore difficult to degrade once released into aquatic systems. The presence of
43	very low concentrations of dyes in effluent can be highly visible and undesirable [3] on
44	aesthetic grounds. Their presence disturbs aquatic communities present in ecosystem by
45	obstructing light penetration and oxygen transfer into water bodies [4]. Moreover, they
46	can be toxic and carcinogenic [5-7].

A number of treatment techniques have been developed to remove dyes from the wastewaters. Among these, advanced oxidation processes (AOPs) are effective in degrading many reactive dyes. The Fenton process is a homogeneous advanced oxidation process using an acidic mixture of hydrogen peroxide and ferrous ions [8, 9] to produce highly oxidative hydroxyl radicals which react with dissolved species, removing colour and lowering chemical oxygen demand. The ·OH radicals (Eq .1), attack the unsaturated dye molecule and the chromophore of the dye molecule is destroyed and decolorized [9].

54 $\operatorname{Fe}^{+2} + \operatorname{H}_2\operatorname{O}_2 \to \operatorname{Fe}^{+3} + \operatorname{OH}^- + \operatorname{OH}$ (1)

The use of power ultrasound as an advanced oxidation process has been also 55 56 employed in the degradation of textile dyes [8, 10]. This is also generally based on the formation of short-lived radical species generated in violent cavitational bubble collapses. 57 These radical species can diffuse out of the bubble into the bulk fluid medium where they 58 59 are able to react with solute molecules. A steady-state concentration of reactive radical species in the liquid phase can be maintained by continuous irradiation with ultrasound. 60 61 Volatile solutes may evaporate into the bubble and be degraded by the harsh conditions 62 generated during cavitation. Non-volatile organic compounds present in the liquid phase can undergo degradation mainly by reaction with powerful oxidizing agents such as ·OH 63 64 radicals produced [11]. The sonochemical enhancement of several AOPs has been 65 reported. A number of studies have reported the use of ultrasound for the decoloration 66 and degradation of textile dyes, but it has often been difficult to completely mineralize 67 the dye stuff using ultrasound alone [12].

68 The combination of ultrasound with other advanced oxidation process is a 69 convenient approach in degrading reactive dyes. There are a number of reports on the

3

70 combined use of ultrasound and Fenton process for the degradation of several textile dyes 71 [8, 9, 13, 14]. While Guimaraes *et al.* [15] showed that oxidation of Reactive Blue 19 can 72 be promoted by use of Fenton reagent and accelerated photochemically, the sono-Fenton 73 process has not previously been applied to the degradation of this dye. Given its 74 widespread use, [15, 16] the present study has focused on RB 19 dye to undergo 75 treatment with combined Fenton and ultrasonic processes. The effect of the various 76 experimental parameters including hydrogen peroxide and iron sulfate concentrations, pH, initial dye concentrations and dissolved gases on color removal were investigated. 77

78

79 **2. Experimental Procedure**

80 2.1. Materials

Reactive Blue 19 (RB 19) was purchased from Sigma Aldrich (UK) and was used as received; its structure is shown in Fig. 1. FeSO₄.7H₂O (analytical grade) was obtained from BDH Laboratory Supplies (England). Hydrogen peroxide (analytical grade) 35% w/v, H₂SO₄ and NaOH were obtained from Fisher Scientific (UK). All solutions were prepared with distilled water. Gases were obtained from BOC and were used without further purification.

~

87 2.2. Procedure

A stock solution of RB 19, prepared by dissolving 1 g of dye in 1 L of distilled water, was diluted to give initial concentrations for each experiment of 25 mg L^{-1} (4 × 10⁻⁵ mol L^{-1}). Sulphuric acid and sodium hydroxide were used to adjust the pH of the dye solution. 100 mL volumes from the stock solution were placed in the glass reactor, and the appropriate amount of hydrogen peroxide and iron sulphate added to the solution. The

glass reactor was equipped with a water circulating jacket for maintaining reaction 93 94 temperature to ± 1 °C. Sonication was performed with a Sonic processor L500-20 95 ultrasonic generator (20 kHz, 200W, Sonic Systems) equipped with titanium probe 96 transducer (23820T). The tip of the horn was 1 cm in diameter and was placed 1.5 cm 97 into the liquid layer (Fig. 2). At 5 min time intervals, samples were taken from the reactor 98 and analyzed by UV/visible spectrophotometry (Agilent 8453) using detection 99 wavelengths of 256 and 594 nm. Each experiment was performed in duplicate. Sonication was conducted in the presence of air with no added gas except when the effect of 100 101 dissolved gas was being investigated. In these experiments, the initial solution was 102 vigorously purged with Argon, O₂ or N₂ for 20 min prior to ultrasonic irradiation and a gas flow of *approx*. 1 ml min⁻¹ was maintained throughout the experiment. 103

104 The percentage (%) decolorization was found from equation (2)

105 Dye % decolorization =
$$(1-C_t/C_0) \times 100$$
 (2)

106 where C_t and C_0 are the concentrations (mg L⁻¹) of dye at reaction time t and prior to 107 sonication respectively [12]. The change in concentration in the solution was calculated 108 from the Beer-Lambert law [13].

$$109 \quad A = l \varepsilon C \tag{3}$$

110 where *A* is the absorbance, *l* is the path length (cm), ε is the molar extinction coefficient 111 (L mol⁻¹ cm⁻¹) and *C* is the dye concentration (mol L⁻¹).

112 2.3. Instrumentation

113 The samples were collected at different time intervals and the degradation 114 products were identified using Liquid chromatography/mass spectrometry (LC/MS). The 115 mass spectrometry was performed using a microTOF electrospray time of flight (ESI-

116 TOF) mass spectrometer (Bruker Daltonik GmbH, Bremen, Germany) coupled to an 117 Agilent 1200 LC system (Agilent Technologies, Waldbronn, Germany). The LC was 118 equipped with a Phenomex luna 5µ C18 (2) column with stationary phase mesh size of 119 100 Å and dimensions (50 mm x 2 mm), connected directly to the MS. The other 120 experimental conditions were: nitrogen supplied at a pressure of 1 bar as a nebulising gas and also used as the drying gas, with the flow rate of 8 L min⁻¹ at a temperature of 200°C, 121 water/acetonitrile was employed as mobile phase with a flow rate of 0.4 mL min⁻¹, 122 123 sample volume 1 µl, injection temperature 25 °C. For the detection of positive/negative 124 ions a capillary voltage of -4000 V / +4000 V was used.

125 2.4. Effects of acoustic intensity and sonolysis

The acoustic power dissipated by the horn into the solution in the reactor was measured using the usual calorimetric method [17] employing a digital thermometer that recorded the solution temperature every 5 min, over a 30 min interval. The overall power input was estimated according to Eq. (4).

130 Energy input,
$$q = mc \, dT/dt$$
 (4)

131 where *m* is the mass of water; *c* is the heat capacity of water and dT/dt is the temperature 132 gradient over time. The acoustic intensity (W cm⁻²) was determined by dividing the 133 power input (*q*) by the horn surface area (*A*).

134 2.5. Oxidizing species determination

The homogeneous system used to monitor radical activity was Fricke dosimetry, in which iron II ions are oxidized (Eq. 5) by sonochemically generated species to iron III [17]. When the ultrasound is irradiated into a Fricke solution Fe^{+2} ions in the solution are oxidized to Fe^{+3} ions as follows:

139
$$\operatorname{Fe}^{+2} + \cdot \operatorname{OH} \to \operatorname{Fe}^{+3} + \operatorname{OH}^{-}$$
 (5)

140 The Fricke solution was prepared with 0.001 M (FeSO₄.7H₂O) and 0.005 M (H₂SO₄).

141 The concentration of Fe^{+3} formed during irradiation was measured by a UV/Visible

142 spectrophotometry at wavelength of 304 nm.

143

144 **3. Results and Discussion**

145 3.1. Decolorization of RB 19 in different systems

Experiments were performed using H₂O₂ alone, FeSO₄ alone, H₂O₂ /FeSO₄ and 146 147 ultrasound combined with H_2O_2 /FeSO₄ to observe independently the effects of different 148 parameters on RB 19 decoloration. Color removal over treatment times of 30 min was less than 5% when H_2O_2 alone, FeSO₄ alone or ultrasound alone was used. Due to the 149 non-volatile and hydrophilic nature of reactive dyes, decoloration would be expected to 150 151 occur mainly in the bulk solution by radical reaction rather than inside the bubbles by 152 pyrolytic reaction. At a frequency of 20 kHz, the concentration of hydroxyl radicals produced in the bulk solution by ultrasound alone was too low to affect the dye 153 decoloration. Also the limited oxidizing power of hydrogen peroxide ($E^0 = 1.78$ V) 154 means that no color removal could be achieved by hydrogen peroxide alone. However, as 155 156 shown in Fig. 3, reduction in color was observed when the combined H₂O₂ /FeSO₄ 157 system was used, resulting from the generation of hydroxyl (OH) radicals in the solution 158 as Equation (1). This contrasts with a related dye (CI Reactive Black 8) on which Zhang 159 and co-workers [13] showed ultrasound had no significant effect on decolourisation by 160 Fenton reagent although other systems have demonstrated a beneficial effect [8, 14, 15] 161 similar to that seen here.

162 Non-volatile organic compounds present in the liquid phase undergo degradation mainly by reaction with these \cdot OH radicals ($E^0 = 2.8$ V). Further, the combined 163 164 ultrasound/H₂O₂ /FeSO₄ system led to enhanced color removal, indicating an accelerating 165 effect due to cavitation [8]. The highest amount of decoloration of about 78% was found in the coupled ultrasound and H₂O₂ /FeSO₄ system after 30 min of treatment, while only 166 167 50% of decolorization was obtained with H_2O_2 /FeSO₄ in the absence of ultrasound. A 168 detailed comparison with other published systems is complicated by the use of different conditions but these results indicate that using ultrasound can enhance the action of the 169 170 H₂O₂ /FeSO₄ system so that each experimental parameter was studied to determine the 171 origin of the effects.

172 **3.2.** Effect of hydrogen peroxide concentration on RB 19 dye decolorization

The decoloration of RB 19 at different concentrations of hydrogen peroxide was 173 investigated (Fig. 4(a)) with a FeSO₄ concentration of 3 mg L^{-1} , pH of 3.5 and ultrasonic 174 intensity of 8 W cm⁻². As with previous studies using both homogeneous [13, 15] and 175 176 heterogeneous [14] sono-Fenton reactions, it was found that the rate of decoloration 177 increased with higher hydrogen peroxide concentration. It is notable that the 178 concentrations of hydrogen peroxide used here were lower than that used in previously 179 published studies and indicates that ultrasound may be applied at lower concentrations 180 than previously suggested and so useful in minimizing the amount of reagent necessary 181 for dye treatment. When hydrogen peroxide concentration was 0.05 mM the decoloration 182 efficiency was approx. 55 % after 30 min of reaction time.

183 When the hydrogen peroxide concentration was increased to 0.5 mM the 184 decoloration efficiency increased due to the higher concentrations of generated hydroxyl

8

185 radicals [8, 9]. The decolorization rate increased as the H_2O_2 concentration increased 186 from 0.05 mM to 0.5 mM although further increase in concentration did not lead to the 187 further increase in decoloration. This can be explained since only a comparatively small 188 amount of additional hydrogen peroxide decomposed to generate hydroxyl radicals and 189 the undecomposed hydrogen peroxide may act as a scavenger for OH resulting in the 190 generation of hydroperoxy radicals (Eq. 6) that are less reactive than the hydroxyl 191 radicals. At these higher H₂O₂ concentrations, hydroxyl radicals react with the peroxide in preference to the RB19 so the degradation rate is reduced. 192 193 $H_2O_2 + \cdot OH \rightarrow HO_2 \cdot + H_2O$ (6) 194 Fig. 4(b) illustrates the kinetics of RB 19 decoloration and demonstrates that it fits a first order kinetic model as in Eq. (7). 195 196 $\ln (C_t / C_0) = -kt$ (7)

197 where k is the first order rate constant (min⁻¹) and t is reaction time in min. The 198 relationship between hydrogen peroxide concentration and first order rate constant is 199 shown in Fig. 4(c).

200 3.3. The effect of initial pH on the RB 19 dye decolorization

The pH values investigated were 2.5, 3.5, 4.5, 5.5 at a fixed FeSO₄ concentration of 3 mg L⁻¹, H₂O₂ concentration of 0.5 mM and ultrasonic intensity of 8 W cm⁻². It can be seen from Fig. 5(a) that the decoloration rate increased with decreasing pH, reaching a maximum at pH 3.5 after which the rate decreased. Zhang *et al.* [8, 13] reported that the rate of decolouration was relatively insensitive to pH below pH=6 although the reduction of chemical oxygen demand was maximized at pH=3. Similar results on other dyes have been reported [10]. The pH is an important parameter in Fenton reactions and it has often

been reported that that the optimum pH is around pH = 3 is usually optimum for Fenton oxidations [18, 19, 20]. Accordingly, the first order rate constant decreases linearly with the increase of pH (> 3.5) as shown in Fig. 5(b). This variation of reaction rate with pH arises from a complex mixture of factors. At low pH (< 2.5), the high concentration of H⁺ ions in solution dominates the reaction with \cdot OH (Eq. 8) [18].

$$213 \quad \cdot \mathrm{OH} + \mathrm{H}^{+} + \mathrm{e}^{-} \to \mathrm{H}_{2}\mathrm{O}$$

Under these conditions, the rate of production of hydroxyl radicals is relatively slow since iron mainly exists as $[Fe(H_2O)_2]^{+2}$. Iron is present in solution in catalytic amounts and a further limitation on the rate of \cdot OH production is the slow regeneration of Fe⁺² after reaction (1). [18].

These effects are less important as the pH rises and so the rate and extent of decolouration rises. However, above pH = 4, other effects come into play which reduce the rate [19-21]. The concentration of Fe^{+2+} in solution is reduced since Fe^{+3} species are more stable in solution. Solid oxyhydroxides such as $Fe(OH)_3$ can also precipitate. Thus, lower concentrations of •OH are generated and decolouration is less effective. These effects combine to give an optimum value of pH = 3.5 in this work as reported previously with other dyes.

225 3.4. The effect of iron sulphate concentration on the RB 19 dye decolorization

The effect of iron sulphate concentration on RB 19 dye decoloration was investigated with H_2O_2 concentration fixed at 0.5 mM, pH value at 3.5 and ultrasonic intensity of 8 W cm⁻². It was observed that the degradation of RB 19 increased with increasing Fe⁺² concentrations in solution (Fig. 6(a)) although the decoloration rate decreased when the iron sulphate concentration was above 3 mg L⁻¹. High concentrations

(8)

of iron sulphate would produce too many ferrous ions in solution, resulting in scavenging
of the ·OH radicals according to reaction (9) [9, 18] in preference to reaction with the
dye.

234
$$\operatorname{Fe}^{+2} + \cdot \operatorname{OH} \rightarrow \operatorname{Fe}^{+3} + \operatorname{OH}^{-1}$$

The relationship between iron sulphate concentration loading and first order rate constant is shown in Fig. 6(b) and indicates that the rate constant increases linearly with the increase of iron loading. It should be noted that the concentration of Fe^{+2} used here is lower than that used in other studies although where comparison is possible, the degradation rate does increase with rising iron concentration under comparable conditions in published studies.

241 3.5. Decay of UV/Visible absorption spectra

242 The change in UV/Visible absorption spectra for RB 19 decoloration as a function of time was recorded (Fig. 7), using a H_2O_2 concentration of 0.5 mM, FeSO₄ 243 concentration of 3 mg L^{-1} , pH 3.5 and ultrasonic intensity of 8 W cm⁻². As can be seen 244 from the spectra, before the oxidation, the absorption spectrum of RB 19 was 245 characterized by two main bands, in UV region (256 nm) and in visible region (594 nm). 246 247 The disappearance of the visible band was likely due to the fragmentation of 248 anthraquinone bond by oxidation and change in absorbance in UV region was considered 249 as evidence of aromatic fragment degradation in dye molecule and its degradation 250 products [8]. Initial radical attack occurs at the aromatic substituent on the anthraquinone 251 ring and at the carbonyl groups [22].

252

253

(9)

254 3.6. The effect of initial dye concentration on RB 19 decolorization

255 Fig. 8(a) illustrates the decoloration of RB 19 at different initial dye concentrations when H_2O_2 concentration is 0.5 mM, FeSO₄ concentration is 3 mg L⁻¹, pH 256 value is 3.5 and ultrasonic intensity of 8 W cm⁻². It can be seen that higher initial dye 257 258 concentrations led to decreased decolorization rates. This may be due to insufficient 259 hydroxyl radical concentration to react with higher amounts of dye in solution. A concentration of 75 mg L^{-1} corresponds to 0.12 mmol L^{-1} which is well in excess of the 260 expected generation of \cdot OH. Fig. 8(b) shows that the first order rate constant decreases 261 262 with the increase of initial dye concentration. This behavior is typical of such 263 sonochemical reactions.

264 A kinetic study using combined ultrasound and Fenton degradation processes was undertaken by measuring the rate constant (k) for first, second and third order reactions. 265 The values confirmed that the order of reaction within the current experimental 266 conditions was first order as shown in Fig. 8(c) at different initial concentrations [23]. 267 The correlation between $\ln (C_1/C_0)$ and irradiation time was linear (Fig. 8(c), the slopes 268 giving the apparent rate constant (k). The regression coefficient, R^2 , values ranged from 269 270 0.99 to 0.95, confirming that combined ultrasound and Fenton degradation process of RB 271 19 followed the first order reaction kinetics.

272 **3.7.** The effect of dissolved gases on RB 19 decolorization

273 Several authors have showed the dependence on the saturating gas of the 274 sonochemical treatment of organics in water [24, 25] although studies on dye systems are 275 rare. Zhang *et al.* showed [8] little difference in the degradation of CI Acid Orange 7 276 under nitrogen or oxygen. The nature of dissolved gas is an important parameter that

277 affects the sonochemical processes since it acts as nucleation sites for cavitation and also 278 influences the conditions such as final temperature achieved inside a collapsing cavitation 279 bubble [26]. Monatomic gases, with higher ratios of specific heat capacities, γ , such as argon, promote increased temperature of collapse and hence higher levels of pyrolysis of 280 281 substances inside the cavitation bubbles. The physical properties, such as specific heat 282 ratio, of oxygen and nitrogen are similar and so similar temperatures are generated inside collapsing bubbles. Any differences here are due to chemical effects. Thus, the gas 283 284 present in the solution plays a determinant role in the implosion conditions of the bubble 285 of cavitation and in the formation of radicals during the ultrasonic process.

Degradation of RB 19 was carried out under argon, nitrogen and oxygen with a 286 H₂O₂ concentration of 0.5 mM, FeSO₄ concentration of 3 mg L⁻¹, pH 3.5 and ultrasonic 287 intensity of 8 W cm⁻². The results showed that the dye decoloration was enhanced when 288 289 dissolved oxygen is present as compared with nitrogen or argon (Fig. 9). The lowest 290 degradation rate was measured under argon. In the presence of argon, the generation of 291 •OH is due only to the decomposition of vapor water (Eq. (10)). Evidence for this is the 292 generation of H_2O_2 under argon saturation compared with oxygen as dissolved gas [26]. $H_2O +))) \rightarrow \cdot H + \cdot OH$ 293 (10)

294 Under oxygen atmosphere the degradation rates were higher as compared to argon and 295 nitrogen, since oxygen promotes the formation of \cdot OH radicals as in Eq. (12).

$$296 \quad O_2 \rightarrow 20 \tag{11}$$

$$297 \qquad O + H_2 O \rightarrow 2 \cdot O H \tag{12}$$

298 These results are consistent with the reaction of RB 19 with hydroxyl radicals in solution

rather than through pyrolysis. If pyrolysis played a significant part in the degradation, a

300 faster rate under argon would be expected and this is not seen.

301 3.8. Effect of ultrasonic powers on RB 19 decolorization

302 The effect of ultrasonic power settings are shown in Fig. 10 (a) with optimized 303 conditions of other parameters. It can be seen that increase in ultrasonic intensities (0-8 W cm⁻²) increases the dye decoloration rate. This has commonly been reported in 304 sonochemical systems. The increase in ultrasonic powers would increase the mixing 305 306 intensity due to turbulence generated by cavitational bubble collapse as well as micro 307 jetting [8] in addition to yielding higher numbers of cavitation bubbles [17] and hence 308 higher yields of hydroxyl radicals. Fig. 10 (b) shows that the first order rate constant 309 increase with the increase of ultrasonic power settings.

By comparing the rate constant values of ultrasonic power settings with rate constant values of Fricke dosimetry, it was found that in combined ultrasound and Fenton process the production of ·OH radicals were almost 10 times more as compared to the ·OH radicals produced by alone ultrasound process (Table 1) in Fricke dosimetry.

314 3.9. RB 19 degradation studies

Attempts to identify the intermediates/end products after the treatment of RB 19 were made using LC/MS. Samples at different time intervals were collected during the combined ultrasound and Fenton process when initial RB 19 concentration was 200 mg L^{-1} , H₂O₂ concentration was 0.5 mM, FeSO₄ concentration was 3 mg L⁻¹, pH was 3.5 and ultrasonic intensities of 8 W cm⁻². From the results of this oxidation process, it can be expected that the 5 min treatment resulted in the disappearance of dye molecule with

decreased peak areas of dye residues. The initial LC/MS results could not assist to propose the degradation mechanism of dye therefore; further analysis will be required to determine the low molecular weight compounds.

4. Conclusions

325 This study showed the first results on the effect of combined ultrasound and 326 Fenton's process on RB 19 dye removal. In general they demonstrate similar effects to 327 other dyes that have been subjected to the process although there are some differences. In particular, significant decolourisation was detected at lower concentrations of iron and 328 329 peroxide than comparable studies on other dyes showing that the use of ultrasound can 330 save the cost and potential side effects of reagent use. Unlike some dye systems, a small 331 amount of degradation was detected using ultrasound alone although this was too slow to 332 be applied in practice. Using ultrasound accelerates the Fenton reaction by accelerating 333 the production of hydroxyl radicals. The decoloration of RB 19 dye was increased with the increase of hydrogen peroxide concentration, ultrasonic power and iron sulphate 334 335 concentration but decreased by increasing the initial dye concentration. The process was optimized at pH = 3.5. The decoloration of RB 19 follows first order rate constant. The 336 337 primary mechanism of reaction with hydroxyl radicals in solution rather than by pyrolysis 338 was confirmed since the rate of decoloration was higher when dissolved oxygen was 339 present as compared with nitrogen or argon.

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344 **References**

- 345 [1]. G. McKay, Waste colour removal from textile effluents, Am. Dyestuff Reporter.
 346 68 (1979) 29.
- V. Jaikumar, V. Ramamurthi, Effect of Biosorption Parameters Kinetics Isotherm
 and Thermodynamics for Acid Green Dye Biosorption from Aqueous Solution by
 Brewery Waste, Int. J. Chem. 1 (2009) 2.
- 350 [3]. S. Kim, C. Park, T.H. Kim, J. Lee, S.W. Kim, COD reduction and decolorization
 351 of textile effluent using a combined process, J. Biosci. Bioeng. 95 (2003) 102.
- 352 [4]. W.G. Kuo, Decolorizing dye wastewater with Fenton's reagent, Water Res. 26
 353 (1992) 881.
- J. S, Bae, H. S. Freeman, Aquatic toxicity evaluation of new direct dyes to the.
 Daplnia, Dye Pigment. 73 (2007) 81.
- R.D. Combes, R.B. Havelandsmith, A review of the genotoxicity of food, drug
 and cosmetic colors and other azo, triphenyl methane and xanthene dyes,
 Mutation. Res. 98 (1982) 101.
- 359 [7]. H.S. Rai, M.S. Bhattacharyya, J. Singh, T.K. Bansal, P. Vats, U.C. Banerjee,
 360 Removal of dyes from the effluent of Textile and dye stuff manufacturing
 361 Industry: A review of emerging techniques with reference to biological treatment,
 362 Crit. Rev. Env. Sci. Technol. 35 (2005) 219.
- 363 [8]. H. Zhang, J. Zhang, C. Zhang, F. Liu, D. Zhang, Degradation of C.I. Acid Orange
 364 7 by the advanced Fenton process in combination with ultrasonic irradiation,
 365 Ultrason. Sonochem. 16 (2009) 325.

- 366 [9]. F. Emami, A.R. Tehrani-Bagha, K. Gharanjig, F.M. Menger, Kinetic study of the
- 367 factors controlling Fenton-promoted destruction of a non-biodegradable dye,
- 368 Desalination. 257 (2010) 124.
- 369 [10]. C. Ozdemir, M.K. Oden, S. Sahinkaya, E. Kalipc, Color Removal from Synthetic
- 370 Textile Wastewater by Sono-Fenton Process, Clean Soil, Air, Water. 39 (2011)
 371 60.
- 372 [11]. E.A. Mamdouh, R.B. Peter, A study investigating the sonoelectrochemical
 373 degradation of an organic compound employing Fenton's reagent, Phys. Chem.
 374 Chem. Phys. 4 (2002) 5340.
- 375 [12]. J.P. Lorimer, T.J. Mason, M. Plattes, S.S. Phull, D.J. Walton, Degradation of dye
 376 effluent, Pure Appl. Chem. 73 (2001) 1957.
- 377 [13]. H. Zhang, Y. Zhang and D. Zhang, Decolorization and mineralization of CI
 378 Reactive Black 8 by the Fenton and ultrasound/Fenton methods, Color. Technol.
 379 123 (2007) 101.
- 380 [14]. H. Zhang, H. Fu, D. Zhang, Degradation of C.I. Acid Orange 7 by ultrasound
 381 enhanced heterogeneous Fenton-like process, J. Hazard. Mater. 172 (2009) 654.
- J.R. Guimaraes, M.G. Maniero, R.N. de Araújo A comparative study on the
 degradation of RB-19 dye in an aqueous medium by advanced oxidation
 processes, J. Environ. Manage. 110 (2012) 33.
- 385 [16]. S. Enes, M. E. Edecan, An optimization study using response surface methods on
 386 the decolorization of Reactive Blue 19 from aqueous solution by ultrasound,
 387 Ultrason. Sonochem. 15 (2008) 530.

- 388 [17]. G.J. Price, E.J. Lenz, The use of dosimeters to measure radical production in
 aqueous sonochemical systems, Ultrasonics. 31 (1993) 451.
- 390 [18]. G. E. Ustün, S.K. Solmaz, T. Morsünbül, H.S. Azak, Advanced oxidation and
 391 mineralization of 3-indole butyric acid (IBA) by Fenton and Fenton-like
 392 processes, J. Hazard Mater. 180 (2010) 508.
- 393 [19]. C. Duesterberg, S. Mylon, D. Waite pH Effects on Iron-Catalyzed Oxidation
 394 using Fenton's Reagent, Environ. Sci. Technol. 42 (2008) 8522.
- 395 [20]. P.R. Gogate, A.B. Pandit A review of imperative technologies for wastewater
 396 treatment I: oxidation technologies at ambient conditions, Adv Environ Res. 8
 397 (2004) 501.
- 398 [21]. S. Wang, A comparative study of Fenton and Fenton-like reaction kinetics in
 399 decolourisation of wastewater, Dyes Pigm. 76 (2008) 714.
- 400 [22]. M. Siddique, R Farooq, Z. M. Khan, Z. Khan, S. F. Shaukat, Enhanced
 401 decomposition of reactive blue 19 dye in ultrasound assisted electrochemical
 402 reactor Ultrason. Sonochem. 18 (2012) 190.
- 403 [23]. P.C. Fung, K.M. Sin, S.M. Tsui, Decolorization and degradation, kinetics of
 404 reactive dye wastewater by UV/ultrasonic / peroxide system, J. Soc. Dyers.
 405 Colour. 116 (2000) 170.
- 406 [24]. F. Mendez-Arriagad, R.A. Torres-Palmaa, C. Petriera, S. Esplugasd, J. Gimenezd,
 407 C. Pulgarinc, Ultrasonic treatment of water contaminated with ibuprofen, Water
 408 Res. 42 (2008) 4243.

- 409 R. Torres, C. Petrier, E. Combet, M. Carrier, C. Pulgarin, Ultrasonic cavitation [25]. 410 applied to the treatment of bisphenol A. Effect of sonochemical parameters and 411 analysis of BPA byproducts, Ultrason. Sonochem. 15 (2008) 605. K. Yasui, T. Tuziuti, T. Kozuka, A. Towata, Y. Iida, Relationship between the 412 [26]. ar MANDER 413 bubble temperature and main oxidant created inside an air bubble under 414
- 415

Figure Captions

- Fig. 1. Structure of RB 19 dye
- Fig. 2. Experimental setup
- Fig. 3. Decolorization of RB 19 by with/without H_2O_2 , FeSO₄ by sonolysis ($C_{0=}$ 25 mg L⁻¹, H_2O_2 conc. 0.5 mM, FeSO₄ conc. 3 mg L⁻¹, pH 3.5, US power =8 W cm⁻²)
- Fig. 4 (a). The effect of hydrogen peroxide conc. on the decolorization of RB 19 ($C_{0=}$ 25 mg L⁻¹, FeSO₄ conc. 3 mg L⁻¹, pH 3.5, US power =8 W cm⁻²)
- Fig. 4 (b). First order kinetic plot of RB 19 decoloration by different hydrogen peroxide conc. ($C_{0=}$ 25 mg L⁻¹, FeSO₄ conc. 3 mg L⁻¹, pH 3.5, US power = 8 W cm⁻²)
- Fig. 4 (c). The effect of hydrogen peroxide concentration on the rate constant
- Fig. 5 (a). The effect of pH on the decolorization of RB 19 ($C_{0=}$ 25 mg L⁻¹, FeSO₄ conc. 3 mg L⁻¹, H₂O₂ conc. 0.5 mM, US power = 8 W cm⁻²)
- Fig. 5 (b). The effect of different pH on the rate constant
- Fig. 6 (a). The effect of iron sulphate addition on the decolorization of RB 19 ($C_{0=}$ 25 mg L⁻

¹, H₂O₂ conc. 0.5 mM, pH 3.5, US power = 8 W cm⁻²)

- Fig. 6 (b). The effect of iron sulphate addition on the rate constant
- Fig. 7. UV/Visible absorption changes with reaction time ($C_{0=}$ 25 mg L⁻¹, H₂O₂ conc. 0.5 mM, FeSO₄ conc. 3 mg L⁻¹, pH 3.5, US power = 8 W cm⁻²)
- Fig. 8 (a). The effect of initial dye concentration on the decolorization of RB 19 (H_2O_2 conc. 0.5 mM, FeSO₄ conc. 3 mg L⁻¹, pH 3.5, US power = 8 W cm⁻²)
- Fig. 8 (b). The effect of initial dye concentration on the rate constant
- Fig. 8 (c). First order kinetic plot of RB 19 degradation by combined ultrasound and Fenton process at different dye concentrations (H_2O_2 conc. 0.5 mM, FeSO₄ conc. 3 mg L^{-1} , pH 3.5, US power = 8 W cm⁻²)

- Fig. 9. The effect of different gases on the decolorization of RB 19 ($C_{0=}$ 25 mg L⁻¹, H₂O₂ conc. 0.5 mM, FeSO₄ conc. 3 mg L^{-1} , pH 3.5, US power =8 W cm⁻²)
- Fig. 10 (a). The effect of ultrasonic powers on the decolorization of RB 19 ($C_{0=}$ 25 mg L⁻¹, H_2O_2 conc. 0.5 mM, FeSO₄ conc. 3 mg L⁻¹, pH 3.5)
- Fig. 10 (b). The effect of ultrasonic power settings on the rate constant

Acceleration







Fig. 4 (a).



Fig. 4 (c).



Fig. 5 (b).



Fig. 6 (b).



Fig. 8 (a).



Fig. 8 (c).



Fig. 10 (a).



Ultrasonic power settings (W cm ⁻²)	k min ⁻¹ (US)	k min ⁻¹ (US) - k min ⁻¹ (no US)	k min ⁻¹ (Fricke)
0	0.0204		
2	0.0253	0.0041	0.00044
4	0.0345	0.0141	0.00105
6	0.0389	0.0185	0.0018
8	0.0442	0.0238	0.00235

Table.1: Comparison of rate constant (min	¹) values between	US power	settings
$(W \text{ cm}^{-2})$ and Frick	ke dosimetry		